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Article:

Bixby, RB, Cooper, SD, Gresswell, RD et al. (3 more authors) (2015) Fire effects on aquatic ecosystems: an assessment of the current state of science. *Freshwater Science*, 34 (4). 1340 - 1350. ISSN 2161-9549

<https://doi.org/10.1086/684073>

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1 RH: Fire effects on aquatic ecosystems

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3 Fire effects on aquatic ecosystems: an assessment of the current state-of-the-science

4

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Abstract

40 Fire is a prevalent feature of many landscapes with numerous and complex effects on
41 geological, hydrological, ecological, and economic systems. In some regions, the frequency and
42 intensity of wildfire have increased in recent years and are projected to escalate with predicted
43 climatic and land use changes. Further, prescribed burns continue to be used in many parts of
44 the world to clear vegetation for development projects, encourage desired vegetation, and reduce
45 fuel loads. Given the prevalence of fire on the landscape, this special issue examines the
46 complexities of fire as a disturbance in shaping freshwater ecosystems and highlights the state-
47 of-the-science through 16 research papers. These papers cover key aspects of fire effects on
48 vegetation loss and recovery in catchments to impacts on hydrology and water quality with
49 consequences for communities (from algae to fish), food webs, and ecosystem processes (e.g.,
50 organic matter subsidies, nutrient cycling) across a range of scales. The results presented in this
51 issue expand our knowledge of fire effects in different biomes, water bodies, and geographic
52 regions, encompassing aquatic population, community, and ecosystem responses. Each paper has
53 been summarized in this overview with an emphasis on each paper's contributions to knowledge
54 on fire ecology and freshwater ecosystems. This overview concludes with a list of research
55 needs to further our knowledge of fire impacts on aquatic ecosystems, including research: 1) on
56 additional biomes and geographic regions; 2) on additional habitats, including wetlands and
57 lacustrine ecosystems; 3) on different fire severities, sizes, and spatial configurations; 4) on
58 additional response variables (e.g., ecosystem processes); 5) over longer (> 5 years) time scales;
59 6) with more rigorous study designs and data analyses; and 7) considering the impacts of fire
60 management practices and policies.

61 Keywords: Wildfire, aquatic ecosystems, streams, rivers, wetlands, ecosystem, biota, prescribed

62 burns

63

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Introduction

65
66 Fires are natural disturbances and agents of landscape change that have a diversity of
67 effects across a variety of spatial scales. Perceptions of the consequences of fire are closely tied
68 to human values (Langston 1995). For example, the use of fire distinguishes humans from other
69 animal species, enhances food nutritional value, and promotes the expansion of valued plant and
70 animal resources. Fire also was an integral driver of the invention and adoption of tools, other
71 technological innovations, and, ultimately, the industrialization and urbanization of human
72 societies, creating the modern world we know today (Pyne 2012). In contrast, humans generally
73 view uncontrolled fire as harmful, destroying natural vegetation, property, and life. From an
74 ecological perspective, however, many ecosystems have co-evolved with fire, with resilient
75 successional trajectories (Pyne et al. 1996, Gresswell 1999, Bowman et al. 2009). Although fire
76 management and policy tend to be focused on protecting human property and life and on
77 protecting or salvaging the economic value of terrestrial resources, such as timber, fire also
78 affects freshwater resources, habitats, and biodiversity. Given the critical importance of water
79 resources to human populations and natural communities globally, a thorough understanding of
80 fire effects on water resources is increasingly important for guiding fire management practices
81 and policy decisions. Although some short-term effects of fire on freshwater ecosystems can be
82 similar to the effects of land use changes (e.g., agricultural and urban development and logging),
83 fire is a pulsed disturbance, with the duration of its effects on freshwater ecosystems depending
84 on terrestrial ecosystem recovery. In contrast, land use changes constitute a press disturbance
85 with more permanent effects (Allan 2004, Wootton 2012, Verkaik et al. 2013). The purpose of
86 this special issue is to illustrate the importance and complexities of fire as a prime driver of
87 change in the physical, chemical, and biological characteristics of freshwater habitats in different

88 geographic regions and biomes (Figure 1). Given the projected effects of climate change on fire
89 frequency and intensity (Knowles et al. 2006, Seager et al. 2007, Pausas and Fernández-Muñoz
90 2011, Westerling et al. 2011), we argue that our focus on the effects of fire on freshwater
91 ecosystems is timely.

92 Most previous work on the effects of fire on freshwater ecosystems has concentrated on
93 wildfire effects on hydrology, sediment transport, geomorphology, water quality, aquatic
94 macroinvertebrate communities, and fish populations in forested, montane streams in the western
95 U.S. (Gresswell 1999, Rieman et al. 2003). This issue of *Freshwater Science* expands on these
96 topics by considering fire effects on a diversity of organisms (ranging from algae and riparian
97 vegetation to spiders and fish) and processes (including micro-climate, hydrology, and
98 biogeochemistry; nutrient inputs, uptake, and limitation; and subsidies between terrestrial/aquatic
99 habitats and tributary/main stem systems). These organismal and process studies were done
100 across a wide array of geographic areas (North America, Europe, Australia, Asia), biomes
101 (boreal forest, Mediterranean shrublands, tropical savanna, temperate, tropical, and semi-tropical
102 wetlands and forests), and habitats (rivers, riparian zones, lakes, wetlands). Although prior work
103 has focused on the effects of fire on state variables, many papers in this issue concentrate on
104 effects of fire on ecosystem processes or rate variables, including nutrient uptake (Diemer et al.,
105 this issue), nutrient limitation (Klose et al., this issue), leaf decomposition (Rodriguez-Lozano et
106 al., this issue), subsidies from river tributaries to river main stems (Harris et al., this issue) and
107 subsidies from streams to riparian zones (Jackson et al., this issue).

108 This special issue was developed in conjunction with a special symposium on the same
109 topic that was held at the Joint Aquatic Sciences Meeting in Portland, Oregon in May 2014. The
110 papers in this issue collectively emphasize the pervasive influence of fire on the structure and

111 function of aquatic ecosystems throughout the world, underscoring the importance of considering
112 fire on freshwater systems in furthering our knowledge of drivers of ecosystem change and in
113 guiding and developing effective natural resource management practices and policies.

114 Building on a series of research needs identified by Verkaik et al. (2013), we evaluate
115 how the papers published in this issue address some of the knowledge gaps in the literature on
116 fire effects on aquatic ecosystems. Specifically, we focus on key aspects of fire effects on
117 riparian and wetland vegetation, microclimate and hydrology, water quality, organic matter
118 subsidies, and stream biota. We conclude with a list of the most critical research needs. The
119 research advances that are reported in this special issue can provide a foundation and springboard
120 for future research studies, leading to the formulation of effective fire management practices and
121 policies that sustain better freshwater resources, habitats, and biodiversity.

122 Riparian and wetland vegetation

123 When terrestrial vegetation is consumed by fire, nutrients are mobilized, runoff and
124 erosion increases, and soils may be altered. Habitat changes occur that favor some species and
125 impede others. Although there is an extensive literature on the responses and recovery of upland
126 vegetation to fire, information on fire effects on riparian and wetland vegetation is limited
127 (Dwire and Kauffman 2003, Pettit and Naiman 2007). Because of differences in the
128 microclimate, foliar moisture, structure, composition, and life histories of riparian/wetland and
129 upland plant species, these plant communities often show very different responses to fire (Van de
130 Water and North 2011). Although basin-wide effects of fire on sediment and nutrient inputs
131 have been studied extensively, the specific effects of riparian or wetland burning on freshwater

132 ecosystems, including organic matter loading, biogeochemical cycles, light and temperature
133 levels, and, ultimately, the aquatic biota, have rarely been delineated (Cooper et al. 2015).

134 In this issue, Douglas et al. examine the effects of annual intensely managed fires on the
135 composition and structure of riparian vegetation in Australia's savannas. In an experiment
136 conducted in whole catchments, they compared riparian vegetation characteristics in burned and
137 unburned watersheds. Vegetation sampling was conducted one year after three years of
138 sequential annual burning. The application of prescribed burning significantly reduced woody
139 species richness, total species abundance, total basal area, the abundance of small trees, canopy
140 cover and the richness and cover of vines, but increased grass cover. Results of this study
141 identified riparian plant species that appeared to be adapted to low frequency, low intensity fires
142 and others adapted to frequent high-intensity fires. This research showed that riparian areas are
143 considerably more sensitive to fire than the surrounding savanna.

144 The floodplain shifting habitat mosaic concept proposes that habitat patch dynamics are
145 driven by flood pulses that alter the geomorphology of channels, banks, and floodplains, thus
146 creating new habitats and changing existing habitats (Stanford et al. 2005). Kleindl et al. (this
147 issue) extend the shifting habitat mosaic concept to examine the effects of multiple, different
148 disturbances, including floods and fire, on the composition of vegetation along the riparian
149 corridor of the Flathead River (British Columbia and Montana). By applying a combination of
150 path and graphical analysis to 22 years of data, they examined relationships among hydrology,
151 fire, land use, geomorphic position, and floodplain habitat patch dynamics. Results suggest that
152 three factors (fire, stream power, and geomorphic position) collectively explained much of the
153 variation in floodplain vegetation patch composition across study reaches and years with wildfire
154 having the strongest total effect. Long-term investigation of disturbance and recovery pathways

155 in a floodplain allow the authors to expand the shifting habitat mosaic concept from one driven
156 only by major hydrologic events to one incorporating the influences of other riverscape and
157 landscape disturbances, particularly fire.

158 Microclimate and hydrology

159 Fire effects on terrestrial and wetland vegetation, and on soils, in turn, influence aquatic
160 ecosystems by altering microclimatic regimes, increasing runoff and river discharge, and
161 enhancing erosion and sediment inputs, transport, and deposition (Gresswell 1999, Benda et al.
162 2003, Coombs and Melack 2013). As a consequence, fire effects on aquatic ecosystems
163 represent compounded effects of two types of disturbances, including post-fire seasonal or
164 interannual increases in runoff and erosion associated with storms or snowmelt (Gresswell et al.
165 2004, Gresswell et al. 2006) superimposed on less frequent changes in vegetation driven by fire.

166 Fire also can affect the physical characteristics of ecotones, including transitions from riparian
167 and wetland areas to uplands. Two years following wildfire, Watts and Kobziar (this issue)
168 compared air temperature, relative humidity (RH), and vapor pressure deficit (VPD) within
169 patches of pond cypress and adjacent grasslands in south Florida, USA. Increasing differences in
170 air temperature, RH, and VPD were observed with distance from the dome centers into savanna
171 habitats but, surprisingly, microclimates were either similar or, in some cases cooler or more
172 humid, in burned compared to unburned domes. The authors attribute this response to vigorous
173 vegetative regrowth following fire. This study increases our understanding of interactions
174 between cypress domes and ecotonal microclimates, thus increasing the ability of resource
175 managers to maintain these unique plant communities under predicted scenarios of greater
176 variability in climate and fire regimes.

177 Given that the ecological effects of smoldering fires are largely unknown, Watts et al. (this issue)
178 develop the first conceptual model of smoldering fires in wetlands, focused on relationships
179 among fire, wetland hydrology, and carbon dynamics. This model underscores the complex and
180 integrated feedbacks between burn depths and extent of smoldering fires on local and regional
181 hydrology, with increased burn depths and extended hydroperiods reducing initiation and
182 frequency of fire in these habitats.

183 Covering approximately 17% of the land surface area in the United Kingdom, peatlands
184 are distributed broadly across the headwater areas of most major river catchments. Brown et al.
185 (this issue) synthesize current knowledge about how rivers in peatlands respond to both wildfires
186 and prescribed burns. The hydrologic response of peatland streams to fire is complex; peak
187 flows are lower during many precipitation events, but peak flows are actually greater during the
188 largest rainfall events. Further, concentrations of dissolved organic carbon (DOC) in surface
189 waters are higher in burned than unburned basins. The authors present a conceptual model that
190 illustrates linkages and feedbacks among the hydrological, chemical, and biological properties
191 and processes of watersheds following fire. This model provides a framework for identifying
192 knowledge gaps and for forecasting changes in peatland streams related to the removal of
193 vegetation by wildfire or prescribed burning.

194 Water quality

195 Fire effects on water quality are of particular concern to water resource managers because
196 of potential effects on water supply systems and aquatic communities. Advances in technology
197 and instrumentation (e.g., sondes) allow the collection of continuous water quality data to
198 monitor changes related to complex disturbances such as wildfires. Chemical datasets with high
199 temporal and spatial resolution document hydrochemical responses to fire, and subsequent floods

200 and debris flows, that are often non-linear and rapid (Krause et al. 2015). For example, water
201 quality data analyzed from a network of sondes in the Rio Grande watershed, New Mexico,
202 documented dramatic decreases in dissolved oxygen and pH as debris pulses moved downstream
203 into a large river system following a large wildfire in headwater areas (Dahm et al. 2015).

204 Reale et al. (this issue) show the value of collecting high resolution, continuous data from
205 networks of water quality sensors and streamflow gages to assess initial and long-term effects of
206 wildfire on the water quality of 2nd and 4th order streams in the Jemez Mountains and Rio
207 Grande in New Mexico. Although there was no difference in precipitation before versus after the
208 fire, episodic post-fire storms results in significantly elevated turbidity and specific conductance
209 (SC) (linked to soil, sediment, rock and ash debris, and solutes entrained from burned catchment
210 areas). There is also greater variability in dissolved oxygen concentrations, in a second-order
211 stream, with more muted responses downstream, in a 4th order river. An additional study of four
212 sites over four months encompassing the wildfire also shows stronger fire effects on turbidity
213 and SC in 1st and 2nd order streams than in higher order downstream sites, implying that flow
214 pathways, geomorphology, and biogeochemical processes moderate fire effects on water quality
215 along the river continuum.

216 Because fires kill or damage vegetation and alter soil chemistry, thereby reducing uptake,
217 nutrients, such as nitrogen and phosphorous, are often mobilized by fire, resulting in increased
218 loading to stream and river ecosystems (Sherson et al. 2015). These post-fire nutrient pulses,
219 which are usually associated with floods, can increase nutrient concentrations many fold.

220 Diemer et al. (this issue) extend our knowledge of long-term fire effects on nutrient dynamics in
221 streams to the boreal forests of central Siberia. Boreal forest streams and their ecosystems are
222 highly susceptible to the effects of climate change, including the intensity, frequency, duration,

223 and extent of forest fires. Diemer et al. show that forest fires in boreal forests alter stream
224 chemistry for many years, affecting the retention and export of nitrogen and phosphorus in these
225 stream networks. Streams within catchments that burned within the last 4-10 years in Central
226 Siberia had lower DOC and higher nitrate (NO₃) concentrations, differing from nutrient
227 responses to fire in boreal regions of North America.

228 Organic matter subsidies

229 By damaging or killing upland vegetation, fires modify the inputs of dissolved and
230 particulate (e.g., as ash and charcoal) organic matter into streams (Earl and Blinn 2003). Where
231 riparian or wetland vegetation is destroyed or damaged by fire, the canopy opens, decreasing
232 allochthonous inputs and increasing light and temperature levels, which promote autochthonous
233 production, with repercussions for aquatic communities and food webs (Beakes et al. 2014,
234 Cooper et al. 2015). In some cases, there can be a pulse of leaf and woody debris from damaged
235 vegetation after riparian fires. Allochthonous inputs often decrease subsequently to the loss of
236 riparian vegetation but organic inputs eventually rebound as riparian vegetation recovers (Britton
237 1990). Further, post-fire hydrological conditions can greatly affect the biomass of organic matter
238 on stream bottoms with floods often mobilizing and transporting organic matter to downstream
239 areas. Riparian trees damaged by fire may not fall into or across streams until years after the fire,
240 usually associated with wind throw or flood events (Robinson et al. 2005, Bendix and Cowell
241 2010).

242 After four post-fire years, Harris et al. (this issue) compare watersheds that were burned
243 then affected by subsequent debris flows to watersheds that had not burned or had been burned
244 without subsequent debris flows. They document a major increase in sediment export during
245 spring runoff in the burned, but not unburned, catchments. Furthermore, stream DOC

246 concentrations are 75% greater in drainages with fires and debris flows than in unburned
247 watersheds, but concentrations of chlorophyll a and the chlorophyll a:organic matter ratio are
248 higher in unburned watersheds. Macroinvertebrate export from tributary streams to the main
249 stem is dominated by r-strategist taxa (Chironomidae, Baetidae, and Simuliidae) in streams that
250 were burned, and the export of invertebrate biomass is greater from streams in burned basins
251 with debris flows than from streams draining unburned basins (Harris et al., this issue).

252 Vaz et al. (this issue) review changes in large wood inputs, distributions, characteristics,
253 and related effects on invertebrate communities, based primarily on their research in Portuguese
254 streams. In a separate study described in the same paper, they also examine the effects of
255 wildfire on large wood subsidies to a lake in northern Minnesota. Their results extend our
256 knowledge of the effects of wildfire on large wood inputs to streams and lakes, suggesting that
257 fire may simplify the structure of wood in streams while resulting in increased habitat
258 complexity in lakes.

259 Although Rodriguez-Lozano et al. (this issue) reported that stream macroinvertebrate
260 functional feeding groups recover quickly, within one or two years, after wildfire, they find that
261 leaf litter inputs were reduced and leaf litter breakdown rates were higher in a stream draining a
262 burned basin than in a stream in an unburned basin 8 years post-fire. The results suggest that
263 microbially mediated leaf decomposition rates are enhanced by increased temperatures
264 engendered by the opening of the riparian canopy by fire and that total (microbial + shredder)
265 leaf breakdown rates were increased by shredder aggregation in coarse-mesh leaf bags in the
266 burned stream where leaf litter inputs are low. These results contribute to a very limited
267 literature on fire effects on detrital dynamics and leaf breakdown rates (Koetsier et al. 2010,
268 Jackson et al. 2012) and results in both Vaz et al. and Rodriguez-Lozano et al. suggest that fire

269 effects on detrital dynamics can be long-lived (exceeding 5 years) (also see Robinson et al.
270 2005).

271 Stream biota

272 Although immediate effects of fire on the stream biota may be muted, stream biological
273 communities usually change radically with post-fire floods, which scour stream substrates and
274 remove most organisms (Gresswell 1999, Minshall 2003). Further, effects on aquatic
275 communities can be modified by pre- or post-fire drought (Rugenski and Minshall 2014). The
276 responses of different types of organisms to fire and floods or droughts and related to life cycles,
277 dispersal abilities, and the availability and distribution of refugia, with short-lived, fast-
278 colonizing species often dominating after fires and floods or droughts (Minshall 2003, Grace
279 2006, Malison and Baxter 2010a).

280 Working in southern California, Klose et al. (this issue) studied the impacts of wildfire
281 and post-fire flooding on algal abundance, community composition, and nutrient limitation
282 (using nutrient diffusing substrata) in stream reaches in unburned and burned catchments. They
283 also consider reaches where riparian vegetation did and did not burn. Results suggest that algal
284 responses (e.g., density, biovolume, chlorophyll a, and species composition) to fire and nutrient
285 enrichment are primarily driven by fire effects on riparian canopy cover, and associated light and
286 temperature levels, flood disturbance intensities, and nutrient concentrations. Decreased riparian
287 cover mediated faster algal recovery post-fire. The results provide insights into processes that
288 create and maintain habitat heterogeneity in riparian and stream habitats.

289 Most information on wildfire effects on stream and river ecosystems is derived from
290 studies of single wildfire events in cooler headwater systems. In contrast, Whitney et al. (this
291 issue) quantify changes in riverine habitat, benthic algal chlorophyll a concentration, and both

292 warm- and cold-water invertebrate and fish communities following consecutive fires that
293 covered >100 km² in southwestern New Mexico, USA. Cumulative fire effects, fire size, and
294 post-wildfire rainfall are strongly associated with siltation of river beds, decreases in chlorophyll
295 a concentration, and decreases in the biomass of most insect taxa and 6 out of 7 native fish
296 species. Among native fish species, the headwater chub *Gila nigra* (100%) and spikedace *Meda*
297 *fulgida* are lost from streams in burned basins for up to two years post-fire. Fish kills are thought
298 to have resulted from hypoxia, and elevated concentrations of ammonium, trace metals, and
299 ferrocyanides, generated by wildfires. Non-native warm-water fish, crayfish, and tadpoles are
300 less affected by fire, suggesting that fires threaten native taxa more strongly than invasive taxa.

301 Verkaik et al. (this issue) consider how stream macroinvertebrate community responses
302 to fire are mediated by interactions with preceding droughts or subsequent flood events. This
303 global-scale, multi-site analysis includes data from central Idaho, USA, northeastern Spain, and
304 Victoria, Australia. Macroinvertebrate community responses to wildfire after 9-11 months were
305 similar across all three regions (i.e., lower taxonomic richness, higher total macroinvertebrate
306 abundance and high percentages of Chironomidae, Simuliidae and Baetidae), but the magnitude
307 of the response differs between among regions. The greatest differences in stream
308 macroinvertebrate communities between burned and unburned basins are found in Australia,
309 where fire is accompanied by ongoing drought and persistent low flows. In contrast, stream
310 macroinvertebrate recovery was faster in the cold-temperate climate of Idaho and the
311 Mediterranean climate of northeastern Spain, where postfire floods may have acted to re-
312 establish or reset biotic colonization processes. These interactions between hydrological and fire
313 events are likely to become more pronounced with climate change.

314 These effects of wildfire and hydrological disturbances on stream invertebrates also can
315 affect subsidies of emerging stream insects to riparian zones, altering the availability of food
316 resources for riparian predators (Malison and Baxter 2010b). Jackson et al. (this issue)
317 investigated the effects of fire on linked aquatic and terrestrial habitats in the Mediterranean
318 climate of California, which is characterized by high interannual variability in precipitation and
319 frequent high-severity wildfires. More specifically, they assessed the effects of wildfire on
320 stream geomorphology; the density and community composition of aquatic benthic
321 macroinvertebrates; and the densities, tissue mercury concentrations, trophic position, and food
322 sources of riparian spiders (Family Tetragnathidae) in Yosemite National Park. Although
323 differences in spider responses between paired burned and unburned study sections are were not
324 statistically significant, modelling suggests that variability in benthic invertebrate density,
325 catchment-scale fire frequency, and precipitation are important predictors of tetragnathid spider
326 density and trophic position. Perhaps most importantly, precipitation is related to multiple spider
327 responses, a relationship suggesting that climate variability could have greater effects on the
328 aquatic-terrestrial ecological linkages than the influence of fire alone.

329 Effects of fire on physical and chemical conditions, and on biological communities can
330 affect populations of apex predators in streams, such as fish (Rieman et al. 2003, Sestrich et al.
331 2011, Beakes et al. 2014). Although wildfires and subsequent floods have been observed to kill
332 or remove fish in isolated, small, headwater streams, fish populations appear to recover quickly,
333 provided there are no barriers to fish immigration (Gresswell 1999). Sedell et al. (this issue) use
334 a qualitative, heuristic model to map the predicted distribution of post-fire debris slides in the
335 Colorado Rocky Mountains. They compare these maps to the distribution of Colorado River
336 cutthroat trout populations. The results indicate that interconnected trout populations would be

337 resilient to wildfire-induced debris flows. Surprisingly they also show that trout populations in
338 headwater streams and lakes likely act as refuge populations for the recolonization of lower
339 stream reaches that are at much higher risk from debris flows.

340 Rosenberger et al. (this issue) documented that rainbow trout are present throughout
341 streams in burned basins after a decade following fires and debris flows, but that individuals in
342 older age classes are least abundant in streams in burned basins with debris flows and most
343 abundant in streams in unburned basins. Rainbow trout from burned watersheds also are
344 characterized by fast growth, low lipid content, and early maturity compared to those in
345 unburned watersheds. Gresswell (2004) reported that stream temperatures were higher in burned
346 basins with debris flows than in unburned basins and burned basins without debris flows.
347 Rosenberger et al. (this issue) developed models whose output suggests suggested that moderate
348 warming, associated with wildfire and channel disturbance history, associated with faster
349 individual trout growth, exacerbating competition for limited food resulting in decreases in trout
350 densities.

351 Future Research Recommendations

352 The papers included in this issue expand our knowledge of the effects of fire on aquatic
353 ecosystems to different geographic regions, biomes, habitats, and response variables, including
354 both rate and state variables. The research presented here emphasizes the importance of fire
355 'type' [wildfire versus prescribed fire, different prescribed burn approaches (e.g., large forest
356 burns, strips to mitigate fire spread, patches to create mosaics)], fire effects on riparian and
357 wetland vegetation, and pre- and post-fire hydrological events on riparian-stream subsidies,
358 stream and wetland communities, and ecosystem processes. All of these topics have
359 implications for the effective management of aquatic resources. Fire effects on aquatic

360 ecosystems are inherently complex. Impacts depend on the characteristics (e.g., extent, intensity,
361 severity, timing, frequency) of fires and the previous or subsequent hydrological events (e.g.,
362 drought and floods). Impacts also depend on features of catchments (e.g., slopes, soils, and
363 vegetation) and receiving waters (e.g., lentic or lotic, discharge, geomorphology, and biota).
364 Future research on fire effects on aquatic systems requires increased focus on a wider array of
365 combinations of fire, hydrology, catchment geomorphology, and aquatic conditions, and models
366 integrating fire effects and natural resource management. As a consequence, we propose that
367 future research be expanded to:

368 1. Additional geographic areas and biomes. Although fire is regularly used to manage savannas
369 and to clear rainforests or wetlands for agricultural activities, very little information exists on the
370 effects of fire on aquatic ecosystems in the tropics (e.g., tropical South America, Africa, Asia,
371 Australia) (Malmer 2004, Townsend and Douglas 2004, Cochrane 2010). Furthermore, the
372 incidence of fire has increased in many additional regions and biomes where fire effects have
373 been little-studied (e.g., arctic and boreal areas, temperate rainforests, grasslands, and semi-arid
374 savannas) (Jacobs et al. 2007, Betts and Jones 2009, Larson et al. 2013, Veach et al. 2014). With
375 the enhanced availability of data from different biomes and regions, it should be possible to
376 undertake more detailed meta-analyses of fire effects (e.g., Verkaik et al., this issue) to look for
377 generalities in the responses of the aquatic biota and ecosystem processes in different types of
378 ecosystems to fire (Brown et al. 2013).

379 2. Other aquatic habitats. Most literature on fire effects on aquatic systems focuses on streams,
380 with few data on fire effects on lakes, ponds, and wetlands (Prepas et al. 2009, Kotze 2013,
381 Lewis et al. 2014). Like the addition of different biomes mentioned above, the inclusion of other

382 aquatic habitats support generalizations (or conversely unique characteristics) that describe fire
383 effects on a large variety of aquatic ecosystems.

384 3. Fires with different characteristics. To date most research has concentrated on the effects of
385 severe or large fires on stream ecosystems; however, many fires across a landscape are small and
386 seemingly inconsequential, but these are underrepresented in research programs. Apparently,
387 prehistorical and historical fire practices concentrated on frequent, small, and low intensity fires,
388 but current fire regimes have been greatly altered by human population expansion, increased
389 ignition sources, and, in some areas, fuel management and fire suppression practices (Stephens
390 et al. 2007). Increased research on the effects of fires differing in severity, extent, and frequency
391 could guide the formulation of fire management practices that better sustain water-associated
392 resources. Even within a given fire perimeter, research is often focused on the most severely
393 and extensively burned areas, and more subtle fire effects on aquatic systems are often ignored.
394 Finally, there have been no landscape or regional quantitative assessments of fire effects on
395 aquatic ecosystems over a complete fire season or across years, including no analyses of
396 cumulative fire effects on the regional distributions and abundances of the aquatic biota. Such
397 assessments will require a combination of extensive and intensive sampling across the landscape
398 using a probabilistic sampling design.

399 4. Additional response variables investigated. Most studies have concentrated on documenting
400 changes in the abundance and biomass of aquatic organisms, with little attention to more subtle
401 or indirect biological responses to fire. For example, indirect, sublethal effects of fire on fish
402 distributions, food availability, growth, reproductive potential, and population structure have
403 received little attention (Gresswell 2004, Beakes et al. 2014). Although this issue has provided
404 some data on fire effects on stream ecosystem processes, such as nutrient uptake and limitation

405 and leaf inputs and decomposition rates, research on these and related topics (e.g., nutrient
406 spiraling, microbial activity, primary and secondary production, stream metabolism) are
407 promising avenues for research on the effects of fire on aquatic ecosystems. Also, this issue's
408 studies and related literature dealing with fire effects on cross-habitat subsidies could enhance
409 our knowledge of drivers of community change in both aquatic and riparian habitats.

410 5. Longer time frames. Although there is a substantial literature on short-term (< 5 years)
411 stream responses to fire (Gresswell 1999, Verkaik et al. 2013), the longer term effects of fire on
412 aquatic ecosystems are largely unknown. Although some stream variables recover quickly after
413 fire, Rodriguez-Lozano et al. (this issue) and Kleindl et al. (this issue) report longer term fire
414 effects on vegetation and detritus (see also Robinson et al. 2005 for detritus, Rugenski and
415 Minshall 2014 for algae). Although limited results indicate some fire effects can be long-lived,
416 much longer time series of data are needed to evaluate the legacy effects of fire. Furthermore,
417 long-term monitoring of a number of systems in a given area will increase the probability that at
418 least one will burn by wildfire (see Jackson et al, this issue), increasing the strength of our
419 inferences by incorporating both pre-fire and post-fire data (Verkaik et al. 2013).

420 6. More rigorous study designs and analyses. Effects of fire on aquatic ecosystems may depend
421 on the spatial pattern of burning. Statistical inferences could be strengthened by greater attention
422 to site selection, which is often opportunistic or based on logistical considerations. In most
423 cases, sites are not selected probabilistically (Hankin and Reeves 1988, Gresswell et al. 2004) or
424 in a manner that addresses issues related to spatial pattern (Ganio et al. 2005, Gresswell et al.
425 2006). Studies that compare changes through time within and among watersheds are rare, but
426 such studies could greatly increase the scope of our conclusions. Because fire effects on aquatic
427 ecosystems are mediated through linkages from vegetation and soils to hydrological,

428 geomorphological, and chemical responses to, ultimately, biotic and ecosystem process
429 responses (e.g., Brown et al., this issue), causal pathway analysis (structural equation modeling)
430 may strengthen inferences regarding the mechanistic routes leading from fire to stream responses
431 (Figure 1, Grace 2006).

432 7. Numerous management practices have been employed before, during, and after fires, but
433 studies of the effects of these practices on freshwater ecosystems are limited despite the
434 important ecosystem services and high biodiversity provided by these critical habitats. Of
435 particular current interest are aquatic responses to: the use of fire retardant to contain fire spread,
436 the construction and maintenance of in-stream structures (e.g., debris dams) to intercept post-fire
437 sediment and debris, applications that stabilize hillslopes (e.g., hydromulch, reseeded), and pre-
438 fire and postfire vegetation removal (e.g., via prescribed burns, mechanical removal, salvage
439 logging) (Karr et al. 2004, Reeves et al. 2006). Most studies have shown muted and short-lived
440 stream ecological responses to prescribed burns (Britton 1991a, Britton 1991b, Bêche et al. 2005,
441 Arkle and Pilliod 2010). Yet, some responses have been more substantial (e.g., see Douglas et
442 al. and Brown et al., this issue) and there has been little investigation of the effects of different
443 prescribed fire severities, extent, and spatial configurations on aquatic ecosystems. The
444 management of fire and fuel loads in riparian areas presents especially difficult challenges
445 (Beschta et al. 2004, Stone et al. 2010, McDaniel 2015), particularly where dominated by
446 flammable exotic taxa [e.g., *Acacia* (acacia), *Arundo* (giant reed), *Tamarix* (salt cedar)] (Lambert
447 et al. 2010, Le Maitre et al. 2011, Drus et al. 2013). During fire-fighting activities, nutrients from
448 fire retardants can increase stream nutrient concentrations (Tobin et al. 2015), have apparently
449 caused fish kills (NMFS 2008), and, when coupled with drought, have had synergistic, negative
450 effects on organisms in mesocosm experiments (Martin et al. 2014). Finally, wildfires in many

451 countries are started by humans and the incidence of wildfire increases with the encroachment of
452 human activities into wildland areas (Syphard et al. 2007, McMorrow et al. 2009), emphasizing
453 the importance of evaluating effects of roads, building construction, and land use regulations
454 (e.g., zoning) on stream community structure and ecosystem processes at the wildland-developed
455 land interface.

456 Conclusions

457 In many regions, fires are becoming more severe and frequent, associated with effects of global
458 climate and land use changes. Both wildfires and prescribed fires affect terrestrial and aquatic
459 ecosystems in numerous and complex ways. This special issue expands our knowledge of fire as
460 a primary driver of hydrological, geochemical, and biological changes in riparian, wetland, and
461 aquatic habitats. In some cases, this is through research into unexplored habitats, biomes, and
462 response variables. Novel approaches, including continuous monitoring, modelling, and
463 probabilistic sampling designs, aid our abilities to generalize and predict outcomes from fire.
464 Many of the studies in this issue also highlight the multifaceted nature of aquatic ecosystem
465 responses to fire; i.e., the interaction of fire with climatic variables (temperature, precipitation),
466 which drive diverse interactions among hydrological, geomorphological, hydrochemical,
467 biological, and ecosystem processes. Finally, we recommend key research needs including the
468 expansion to additional geographic regions, biomes, habitats, and response variables; larger
469 spatial and temporal scales; and fires with different characteristics. We also emphasize the
470 critical need for research on the effects of fire management practices and policies on aquatic
471 ecosystems and for the consideration of aquatic ecosystems when making fire management and
472 policy decisions.

473 Acknowledgements

474 We thank Sheila Wiseman for drafting the figure. R. Bixby and C. Dahm acknowledge funding
475 through the New Mexico Experimental Program to Stimulate Competitive Research (NM
476 EPSCoR) (National Science Foundation). S. Cooper was supported by funds from the National
477 Science Foundation's Rapid Response and Long-Term Ecological Research programs. L.
478 Brown's contribution was supported via the EMBER (Effects of Moorland Burning on the
479 Ecohydrology of River basins) project funded by the UK's Natural Environment Research
480 Council (NE/G00224X/1). Any use of trade, firm, or product names is for descriptive purposes
481 only and does not imply endorsement by the U.S. Government.

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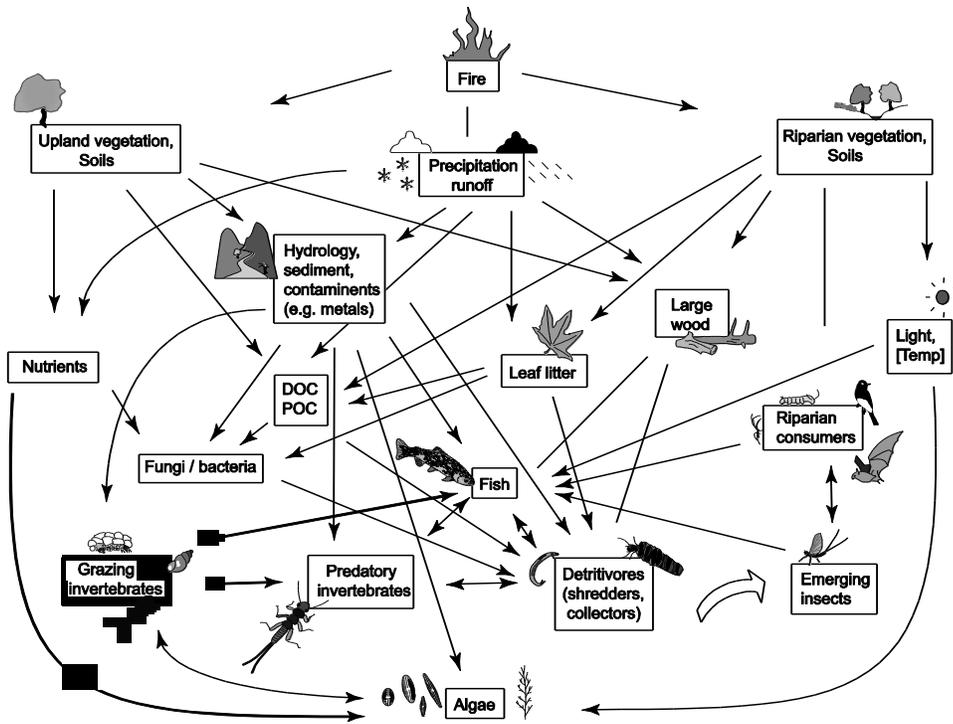
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729 Figure 1: Path diagram showing probable cause-effect relationships leading from fire to stream
730 communities. Lines without arrows indicate factors that are associated with each other,
731 unidirectional arrows point from driving to response variables, and double-headed arrows
732 indicate consumer-resource interactions where consumers both depress, and benefit from the
733 consumption of, their resources.



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