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**Article:**

Gordon, L.E., San Nicolas, R. and Provis, J.L. [orcid.org/0000-0003-3372-8922](https://orcid.org/0000-0003-3372-8922) (2014) Chemical characterisation of metakaolin and fly ash based geopolymers during exposure to solvents used in carbon capture. *International Journal of Greenhouse Gas Control* , 27. pp. 255-266. ISSN 1750-5836

<https://doi.org/10.1016/j.ijggc.2014.06.005>

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31 exposure to  $K_2CO_3$ . The limited curing duration of the specimens tested here is certainly  
32 contributing to the degradation taking place under  $K_2CO_3$  exposure, whereas the low water  
33 activity in the MEA solutions used means that bond hydrolysis in the aluminosilicate  
34 geopolymer framework is restricted, and the materials perform much better than in a more  
35 water-rich environment.

36

37 **Keywords:** Construction materials; process equipment; concrete; geopolymer; solvent  
38 processes

39

## 40 **1. Introduction**

41 Traditional cementing binders are formed via the hydration reactions of calcium silicate  
42 phases, whereas geopolymer concretes are synthesised by the alkali activation of  
43 aluminosilicates to form the binder. There is currently widespread research into the  
44 development of geopolymers and alkali activated systems (Provis 2014; Provis and van  
45 Deventer 2014), motivated not only by the beneficial technical properties of the material,  
46 such as thermal and chemical resistance, but also by the greatly reduced carbon emissions in  
47 geopolymer production relative to traditional cementing binders (McLellan et al. 2012). It is  
48 estimated that 5-8% of global anthropogenic carbon emissions are due to cement production  
49 (Olivier et al. 2012), driving research towards forming alternative low emissions binders as a  
50 viable commercial option (van Deventer et al. 2010; Juenger et al. 2011; van Deventer et al.  
51 2012).

52

53 As a means of reducing the CO<sub>2</sub> emissions profiles of the global energy and cement  
54 industries, which are large point sources of CO<sub>2</sub>, carbon capture and storage (CCS) processes  
55 and facilities are currently being trialed (Jenkins et al. 2012), with viability depending  
56 strongly on the costs involved. Post-combustion carbon capture solvent columns are generally  
57 constructed from stainless steel, as good mechanical performance and a high degree of  
58 chemical resistance are required from a column construction material. Geopolymer concretes  
59 are proposed to provide a durable, cost saving ‘green cement’ alternative structural material  
60 for the construction of large CCS facilities, with the aluminosilicate geopolymer chemistry  
61 suggested to provide the possibility of withstanding the harsh chemical environments found  
62 within these systems (Gordon et al. 2011) as the chemical resistance of geopolymers has long  
63 been promoted as a benefit of these materials (Davidovits 1991), and has been highlighted in  
64 laboratory testing over a number of years (Davidovits 1991; Bakharev 2005; Fernández-  
65 Jiménez et al. 2006; Duxson et al. 2007a; Sindhunata et al. 2008; Fernández-Jiménez and  
66 Palomo 2009; Temuujin et al. 2011).

67

68 There are many appealing aspects of geopolymer concretes which may make them suitable  
69 for use as a construction material in carbon capture facilities. Relative to traditional  
70 concreting systems based on ordinary Portland cement (OPC), geopolymers can offer as  
71 much as 80% reduction in CO<sub>2</sub> emissions (Duxson et al. 2007b) These materials also utilise  
72 industrial wastes such as fly ash as an integral binder component; fly ash is currently  
73 landfilled at a rate of several million tonnes p.a. in Australia alone, and hundreds of millions  
74 of tonnes p.a. worldwide, bringing appealing synergies if the material was to be used in  
75 carbon capture applications associated with coal-fired electricity generation.

76

77 Fly ash is a by-product from the combustion of coal in thermal power plants, whereas  
78 metakaolin is an industrial mineral product formed by the calcination of kaolinite clay.  
79 However, the use of metakaolin as the aluminosilicate source in geopolymer synthesis  
80 provides a relatively simpler model system for the study of geopolymers when compared to  
81 the multiphase and highly variable nature of industrial aluminosilicate wastes such as fly ash,  
82 In particular, metakaolin-based systems provide a more straightforward system to aid in  
83 understanding the mechanisms of reaction leading to binder formation, as well as resistance  
84 to chemical attack and consequently durability in service. This study builds new  
85 understanding of these properties by analysing the effects of exposure to the solvents used in  
86 post combustion carbon capture, on both fly ash and metakaolin based geopolymer systems.  
87 It is unlikely that metakaolin-based geopolymer concretes would be seriously considered for  
88 large-scale use in this type of application due to their generally high porosity, but they are  
89 valuable as a means of understanding the influence of the solvents on the aluminosilicate gel  
90 structure.

91

92 The chemical durability problems of traditional ordinary Portland cement based systems are  
93 generally caused by the degradative changes that occur in the calcium containing phases  
94 (Taylor 1997). Previous work in the analysis of the potential use of different types of concrete  
95 in CCS applications considered the exposure of OPC and OPC-based composite systems to  
96 the same carbon capture solvents that are analysed here (Gordon et al. 2011). Decalcification  
97 occurred upon exposure to MEA, and significant carbonation occurred upon exposure to  
98  $K_2CO_3$ . However, the mechanisms controlling chemical resistance in an aluminosilicate  
99 geopolymeric gel will be expected to be different, based on the very different degree of cross-  
100 linking in the silicate binder gels (Duxson et al. 2007a; Abora et al. 2014).

101 This paper will describe the chemical characterisation of geopolymer paste samples during  
102 exposure to lean carbon capture solvents, through elemental analysis of the leaching solutions  
103 and X-ray diffractometry (XRD) of the pristine and leached materials. Lean solvents have  
104 been selected for analysis, rather than CO<sub>2</sub>-loaded solvents, because it is more likely that  
105 concrete construction would be considered initially for use in the less chemically-aggressive  
106 parts of a carbon capture process, rather than in the more challenging environment of  
107 exposure to a rich solvent. Compressive strength data are presented to provide some insight  
108 into the influence of the leaching process on mechanical performance.

109

## 110 **2. Materials and Methods**

### 111 **2.1. Sample formulations**

112 The metakaolin used was sold under the brand name Metastar 402 by Imerys Minerals, UK.  
113 It has a BET surface area of 12.7 m<sup>2</sup>/g and a mean particle size d<sub>50</sub> of 1.58 µm (Duxson et al.  
114 2006). The fly ash was from the Gladstone power station, Queensland, Australia and the  
115 ground granulated blast furnace slag (GGBFS) was obtained from Independent Cement &  
116 Lime, Australia. Table 1 presents the component oxide ratios for these materials.

117

118 A series of 32 geopolymer formulations were examined (Table 2). Of these, 15 were based on  
119 a mixture of metakaolin and ground granulated blast furnace slag (MK:GGBFS mass ratio  
120 3:1) and 15 based on fly ash mixed with ground granulated blast furnace slag (FA:GGBFS  
121 mass ratio 2:1), along with a pure metakaolin-based binder and a pure fly ash-based binder.  
122 All mixes were activated with liquid sodium silicate activators, made by combining a  
123 commercial sodium silicate solution (Grade N®, PQ Australia, composition (mass basis):

124 SiO<sub>2</sub> 28.7%, Na<sub>2</sub>O 8.9%, H<sub>2</sub>O 62.4%) with NaOH and Milli-Q water, to give the oxide ratios  
125 shown in Table 2. Table 3 summarises the parameters which were varied in each sample set  
126 in Table 2.

127

## 128 **2.2 Analytical methodology**

129 Leaching analysis was carried out using cylindrical sections of paste samples, which were  
130 cured under sealed conditions for 48 hours at 40°C and a further 5 days at room temperature  
131 (23±2°C) prior to exposure to aggressive conditions. A solid section of ~15 g weight was  
132 submerged in 200 mL of solvent: either 98% monoethanolamine (MEA), 2.5 M K<sub>2</sub>CO<sub>3</sub>, or  
133 Milli-Q grade purified water. An aliquot of 1 mL of solution was removed after 1, 3, 7, 14, 28  
134 and 90 days. The volume of leachate removed was replenished with fresh solvent and a pH  
135 measurement taken at each sampling. The leachate samples were diluted 10× with 10 wt.%  
136 HCl solution for analysis by inductively coupled plasma-optical emission spectroscopy (ICP-  
137 OES) to determine the concentrations of Al, Si, Na and Ca present.

138

139 For compressive strength testing, mortar specimens of each formulation were made with the  
140 addition of sand at a 3:1 volume ratio of sand to aluminosilicate solids. These were sealed in  
141 50 mm cubic moulds at 40 °C for 48 hours, then held at room temperature and maintained  
142 sealed, to match the curing regime used for the paste samples. After 7 and 28 days, they were  
143 analysed for compressive strength. Three replicate specimens were tested for each sample,  
144 and the mean results are reported.

145

146 Mortar cube samples were also submerged in the solvents (98% MEA, 2.5M K<sub>2</sub>CO<sub>3</sub> and  
147 Milli-Q water) for 28 days, without replenishment of the solvent, before being analysed for  
148 compressive strength; again there were three replicates for each sample and the mean result is  
149 reported.

150

151 X-ray diffraction analysis was carried out on crushed paste samples after 90 days of  
152 exposure, on a Bruker D8 Advance using Cu K $\alpha$  radiation,  $\lambda = 1.5405 \text{ \AA}$ . Traces were  
153 measured at a scan speed of 5s/step and a step size of 0.02°, from 5 to 55° 2 $\theta$ .

154

### 155 **3. Results and discussion**

#### 156 **3.1. Leaching and mechanical strength**

157 Figures 1 and 2 show the measured extent of leaching of both Si (Figure 1) and Al (Figure 2)  
158 from all of the metakaolin-based and fly ash-based geopolymer formulations in each of the  
159 three solvents. These plots do not show sample-specific data, but instead give an overall view  
160 of the leaching resistance of each binder as a function of the solvent used. There was high  
161 resistance to MEA, with generally <1% Si and Al leaching in this solvent during the test  
162 duration. However, there was notably increased leaching of matrix components induced by  
163 alkali attack on the matrix in both K<sub>2</sub>CO<sub>3</sub> and H<sub>2</sub>O. The pH value exceeded 12 after one day  
164 of exposure in all solvents, due to release of alkalis from the pore solution of the samples, and  
165 remained at this level throughout the duration of the study. The leaching of matrix  
166 components was rapid, with the majority of the leaching taking place within the first day of  
167 exposure to all solvents. There was then a further, very gradual release of Si and Al with  
168 increased solvent ingress into the binders over time, and this effect was more marked in the  
169 fly ash formulations than the metakaolin binders, probably due to the lower w/b ratios

170 achievable in a mix design using spherical fly ash particles rather than plate-shaped  
171 metakaolin particles as the aluminosilicate source (Provis et al. 2010).

172

173 The main reason for the high degree of leaching observed here was that the pastes were quite  
174 immature at the point in time when they were first immersed in the solvents (i.e., 7 days of  
175 curing under sealed conditions at near-ambient temperature). This was designed to enable a  
176 comparative analysis between the dissolution of binder components within a realistic  
177 laboratory test time frame, as more mature geopolymer binders have been shown to display a  
178 much lower degree of leaching of binder components during extended exposure to alkaline  
179 solutions (Sindhunata et al. 2008; Temuujin et al. 2011). This is necessary in order to develop  
180 a scientific understanding of the influence of synthesis parameters on binder performance  
181 under these conditions, as more mature binders would be expected to show relatively less  
182 variation in nanostructure and microstructure after leaching due to the reduced influence of  
183 the solvent environment, and would thus be less instructive than the less-mature specimens  
184 investigated here.

185

186 An intermediate level of Ca content was provided to the binders studied here by the addition  
187 of granulated blast furnace slag (GGBFS). This is known to enhance the impermeability  
188 properties of the binder by the formation of a void filling, low-Ca, Al-substituted calcium  
189 silicate hydrate (C-(A)-S-H) gel in coexistence with the alkali aluminosilicate gel resulting  
190 from geopolymerisation (Kumar et al. 2010; Provis et al. 2012). The relatively low content  
191 and accessibility of calcium in the geopolymer binders is predicted to prevent the degradative  
192 effects of solvent-induced carbonation. Carbonation by gaseous CO<sub>2</sub>-rich environments has  
193 been shown to be reasonably rapid in some alkali-activated binder systems (Puertas et al.

194 2006; Bernal et al. 2010), but the mechanism of carbonation in the fully saturated  
195 environments studied here, and its influence on the binder structure, are expected to differ  
196 significantly.

197

198 Previous studies of acid and alkali resistance in geopolymers found that mass and strength  
199 loss occurred through leaching, correlating with the increasing porosity of the binders  
200 through (Temuujin et al. 2011; Lloyd et al. 2012; McLellan et al. 2012) or alkali (Sindhunata  
201 et al. 2008; Temuujin et al. 2011) attack. The leaching of Fe, Al and Na was high in acidic  
202 solutions, and Si leaching dominated in alkaline conditions.

203

204 The trends displayed in Figure 1 show that the leaching of Si in  $K_2CO_3$  in the fly ash-based  
205 systems occurred more readily relative to the metakaolin-based systems, while Al leaching  
206 (Figure 2) is significantly greater in water than in either of the carbon capture solvents used.  
207 Perera et al. (2006) showed that the release of alkali metal cations from both fly ash and  
208 metakaolin based geopolymers was high in distilled water when accompanied by hydroxide  
209 leaching, and the pH, both in that study and in our experiments, was found to be greater than  
210 12 throughout the test period. It is possible that the presence of carbonate at this pH is  
211 suppressing the solubility of  $Al(OH)_4^-$ , most likely by increasing the ionic strength of the  
212 solution.

213

214 Figure 3 shows the total Si, Al, Na and Ca leached, expressed as a percentage of the total  
215 inventory of each element in the samples, after 90 days of exposure.

216

217 If each oxide ratio series is viewed separately (Table 3), Si leaching increases through the  
218 series with increasing  $\text{SiO}_2/\text{Al}_2\text{O}_3$  ratio (2Na-4Na for MK/GGBFS and 18-22Na for  
219 FA/GGBFS) in all solvents. This is also accompanied by a loss in compressive strength  
220 (Figure 4). The strength loss is attributed to the reduced availability of Al for reaction, which  
221 is known to control the strength properties of geopolymers (Weng et al. 2005; Fernández-  
222 Jiménez et al. 2006; Duxson et al. 2007a). Aly *et al.* (2008) proposed that a lower degree of  
223 polymerisation occurs at higher Si/Al ratios, and supported this proposal using  $^{29}\text{Si}$  and  $^{27}\text{Al}$   
224 magic angle spinning nuclear magnetic resonance (MAS NMR) spectroscopy: a reduced  
225 influence of Al on the average Si coordination sphere was observed, and XRD and infrared  
226 spectroscopic data also indicated a structural shift towards amorphous silica with increasing  
227 Si/Al ratio. This indicates that the correlation between reduced strength and increased  
228 leaching of Si is due to the reduced extent of geopolymerisation reactions within these  
229 immature binders, as the extent of incorporation of Si into the binder itself is not yet high.

230

231 However, the correlation between more marked strength loss and increased leaching does not  
232 follow for all of the binders. For example, the leaching extents observed in samples 7Na and  
233 23Na are the lowest in their respective series (Figure 3), but unexposed strength was low in  
234 7Na (~13 MPa) after 28 days of aging, and exposure to MEA resulted in catastrophic strength  
235 loss (Figure 4). The same trend is observed in sample 23Na (strengths of <5 MPa, no residual  
236 strength in water or  $\text{K}_2\text{CO}_3$ ). This may be related to the low extent of reaction of the  
237 aluminosilicate precursors, as both of these specimens are formulated specifically to test the  
238 effect of a reduction in the  $\text{Na}_2\text{O}/\text{SiO}_2$  ratio; i.e., a reduction in activator alkalinity, and thus  
239 reduced extent of interaction with the aluminosilicate particles during the geopolymerisation  
240 reaction. It is therefore necessary to consider this process to bring understanding of the  
241 formation, and subsequent degradation by leaching, of geopolymer binders.

242

243 Previous studies have considered the rate and extent of dissolution of fly ash in alkaline  
244 media, which is central to this discussion. Li *et al.* (2011) showed that the extent of fly ash  
245 dissolution in NaOH solutions was relatively low, even at high NaOH concentrations (6.0-  
246 13.4 M). In that study, the dissolution of Al occurred more readily than Si, shown by the  
247 variation of Si/Al ratio over time (Li *et al.* 2011) but in an earlier study (Pietersen *et al.*  
248 1989), congruent dissolution was observed. It appears that the details of this process depend  
249 on the specific chemistry of the fly ash under analysis, and also the leaching environment  
250 used; Phair and van Deventer (2001) tested the leaching of Si and Al from a fly ash sourced  
251 from the same power station as the material used here, and found that the degree of  
252 incongruency of dissolution (favouring more rapid Al release) generally decreased with  
253 increasing NaOH concentration. Here, at the lowest alkali concentrations tested (sample 7Na  
254 in the fly ash system), it appears that the binder after 7 days of curing is sufficiently immature  
255 that the concentration of gel (which is the part of the material susceptible to solvent leaching)  
256 is low enough to give an apparently low extent of overall leaching, in parallel with poor  
257 strength performance. For the case of the metakaolin samples (particularly 13Na), the  
258 situation is broadly similar. Additionally, previous leaching studies have observed  
259 reprecipitation of aluminosilicate gels during leaching of geopolymers in alkaline solutions  
260 (Sindhunata *et al.* 2008), and this may also partially explain the low leaching rate found here  
261 when high strength loss and structural damage have occurred.

262

263 Where the Na<sub>2</sub>O/SiO<sub>2</sub> ratio is increased (samples 7Na-9Na for MK/GGBFS and 23Na-27Na  
264 for FA/GGBFS), variation in water content was required to hold H<sub>2</sub>O/Na<sub>2</sub>O constant with the  
265 increase in alkalinity. As noted above, Sindhunata *et al.* (2008) observed reprecipitation of

266 aluminosilicate species when geopolymers were exposed to highly alkaline solutions, but this  
267 effect was reduced marked at very high alkalinity. There is increased leaching of Si when  
268 increasing the  $\text{Na}_2\text{O}/\text{SiO}_2$  ratio in the fly ash sample series (23Na-27Na, Figure 3e), which is  
269 assigned in part to the increased formation of geopolymer gel at higher alkalinity, and in part  
270 to the higher water content, which would be expected to increase porosity and permeability  
271 and thus increase alkali attack.

272

273 Correspondingly, the samples where the water content was the sole parameter varied (13Na-  
274 16Na and 28Na-32Na) do show a trend towards reduction in strength at increased water  
275 content (Figure 4). The generally increased leaching of Si follows in the fly ash formulations  
276 (28Na-32Na), Figure 3e. However, leaching remains low and follows no particularly apparent  
277 trend with respect to water content in the metakaolin binders; reprecipitation of  
278 aluminosilicate gels is likely to provide an explanation for this observation. The leaching of  
279 Al is significantly higher in water than in  $\text{K}_2\text{CO}_3$ . The samples with very low leaching of Al  
280 (28Na-31Na, Figure 3f) are those that have both high unexposed and residual strengths.

281

282 The amount of Na leached is higher than all the other ions for every formulation, which is  
283 related to the high mobility rate of those ions compared to the other ones studied, as Na is  
284 present largely in the pore solution of the materials, and in weakly associated charge-  
285 balancing sites (Lloyd et al. 2010). Further, the leaching of Na is significantly lower in MEA  
286 compared to  $\text{H}_2\text{O}$  and  $\text{K}_2\text{CO}_3$ , as MEA is a less effective solvent than aqueous environments  
287 for alkali hydroxides. For the fly ash formulations, no clear trends can be observed in the  
288 MEA solution. However in water and  $\text{K}_2\text{CO}_3$  there is decreased leaching of Na with  
289 increased  $\text{Na}_2\text{O}/\text{SiO}_2$  ratio in the fly ash sample series (23Na-27Na) which is contrary to the

290 behaviour of Si and Al (Figure 3g) where increased leaching is attributed to the higher water  
291 content and assumed higher porosity through this series. The higher alkalinity of these  
292 systems results in greater production of geopolymer gel and so possibly the Na is chemically  
293 bound within the pore structure through the electrostatic interactions with the geopolymer  
294 matrix. Na is increasingly leached when the  $H_2O/SiO_2$  ratio increases (28Na-32Na), due to  
295 the higher water content.

296 The significant observed extent of Ca leaching (Figure 3d,h) shows that a significant  
297 proportion of the Ca present is available for leaching by the solvent solutions. Previous work  
298 has shown that significant decalcification occurs in OPC, OPC/GGBFS and OPC/FA binders  
299 on exposure to MEA (Gordon et al. 2011), so these data show that the Ca-containing gels  
300 formed in these geopolymer systems seem to release Ca in a similar manner. X-ray  
301 diffraction data, presented below, show that in addition to the Ca present as dissolved species  
302 in the leaching solution, additional Ca is removed from the gel binder and precipitated in the  
303 formation of calcite ( $CaCO_3$ ) on exposure to  $K_2CO_3$ , and also due to atmospheric carbonation  
304 in a few samples where the seal on the leaching vessel had been breached during sampling  
305 throughout the testing period.

306

307

### 308 **3.2. X-ray diffraction**

309

310 Figures 5-6, and the Supporting Information, show XRD traces for the metakaolin/GGBFS  
311 based binders, with separate figures focusing on each series of oxide ratios and their exposure  
312 under different conditions (Table 2), and compared to the binders of the same formulation  
313 which were aged at room temperature under sealed conditions for durations corresponding to  
314 the solvent exposure tests.

315

316 Figure 5 presents data for samples 2Na-4Na. The muscovite ( $\text{KAl}_2(\text{Si,Al})_4\text{O}_{10}(\text{OH})_2$ ,  
317 PDF#58-2035) was present as an impurity in the metakaolin used, and reacts only slightly  
318 during geopolymer formation (Yip et al. 2008). The mixed carbonate - bicarbonate phase  
319 trona  $\text{Na}_3(\text{CO}_3)(\text{HCO}_3)\cdot 2\text{H}_2\text{O}$  formed on the surface of the pure metakaolin sample 1Na and  
320 also on the surfaces of samples 1Na-4Na through atmospheric exposure, possibly in the initial  
321 curing stage when the samples were held at 40 °C for 48 hours as this phase is  
322 thermodynamically more stable at temperatures slightly higher than room temperature  
323 (Bernal et al. 2012). Pirssonite  $\text{CaNa}_2(\text{CO}_3)\cdot 2(\text{H}_2\text{O})$ , formed by atmospheric carbonation, was  
324 also detected in samples 3Na (and thus is shown in the corresponding plots as 9Na and 14Na)  
325 and also in sample 12Na, but was not detected in these samples following solvent exposure.  
326 However, it has formed in sample 15Na following exposure to MEA, although this may also  
327 have been due to some atmospheric carbonation prior to or during XRD analysis.

328

329 A zeolite phase, faujasite (approximately  $\text{Na}_2\text{Al}_2\text{Si}_{3.3}\text{O}_{10}\cdot 7\text{H}_2\text{O}$ , PDF# 12-0228) has formed in  
330 all samples. At the highest  $\text{SiO}_2/\text{Al}_2\text{O}_3$  ratio a Na-chabazite zeolite phase (approximately  
331  $\text{NaAlSi}_2\text{O}_6\cdot 3\text{H}_2\text{O}$ , PDF# 19-1178) has also formed. Very little zeolite was detected in all  
332 samples following solvent exposure, with only small faujasite peaks detected in sample 3Na  
333 following exposure to  $\text{H}_2\text{O}$ . This appears to be related to the limited extent of zeolite  
334 formation observed after 7 days of curing (where the residual activator in the pores was  
335 washed out during immersion in water, effectively stopping the geopolymerisation/zeolite  
336 growth reaction), rather than via the removal of zeolitic phases during immersion. A study of  
337 accelerated aging of geopolymers (Lloyd 2009) showed that at 23°C there was no zeolite  
338 crystallisation in a metakaolin geopolymer for up to 3 months of aging. That study also

339 considered the relationship between zeolite formation and strength retention in geopolymeric  
340 systems, concluding that the strongest correlation is between pore network stability and  
341 strength retention, rather than observing a direct link between zeolite evolution and  
342 geopolymer strength. Palomo et al. (1999) also exposed immature alkali activated metakaolin  
343 geopolymer to aggressive solutions, where the binders had been subjected to accelerated  
344 curing (85 °C) for just 2 days prior to exposure, and also found faujasite across a wide range  
345 of pH conditions, but not until at least 3 months of exposure in the samples immersed at high  
346 pH.

347

348 Zeolite formation within geopolymer binders is a complex process (Provis et al. 2005) and  
349 not yet fully understood. However, the formation of Na-chabazite along with faujasite in only  
350 sample 4Na (Figure 5a) is attributed to the increased Si/Al ratio of this sample. The high  
351 Si/Al ratio, reduced extent of geopolymerisation and alkali-rich pore solution conditions lead  
352 to the coexistence of these zeolitic phases (faujasite and chabazite), which is not observed in  
353 the formulations with less Si.

354

355 Solvent exposure in general inhibits the formation of zeolites compared to the unexposed  
356 samples, most likely due to the removal or dilution of the alkaline pore fluids required for  
357 ongoing reaction within the geopolymer gel. Exposure to MEA (Figure 5b) left muscovite as  
358 the only crystalline phase observed in the samples, while the amorphous 'hump' characteristic  
359 of the disordered aluminosilicate geopolymer framework (Provis et al. 2005) remains intact.  
360 This is in good agreement with the low level of leaching of both Si and Al found on exposure  
361 to MEA in all samples.

362

363 Exposure to water (Figure 5c) has also hindered the development of zeolitic phases through  
364 the dilution of the pore solution, and some carbonation, evident via the formation of  
365 crystalline calcite ( $\text{CaCO}_3$ , PDF #05-0586) occurred in samples 2Na and 3Na after exposure  
366 to water and  $\text{K}_2\text{CO}_3$ , showing that calcium is available for reaction with the atmosphere. This  
367 is in good agreement with the high leachability of Ca as observed in all solvents (Figure 3d).  
368 No zeolitic phases form during exposure to  $\text{K}_2\text{CO}_3$  (Figure 5d), apart from a small amount of  
369 faujasite detected in samples 2Na and 3Na, which also forms in these formulations following  
370 exposure to water (Figure 5c).

371

372 Figure 6, which shows XRD data for the sample series with increasing  $\text{Na}_2\text{O}/\text{SiO}_2$  ratio,  
373 demonstrates the formation of faujasite in samples 8Na and 9Na. No zeolitic phases have  
374 formed in sample 7Na (Figure 6a), which has the lowest  $\text{Na}_2\text{O}$  content; the lower extent of  
375 reaction of the aluminosilicate precursor phases in this system will affect the availability of  
376 species for zeolite crystallisation. Additionally, in this sample there was also a reduction in  
377 water content in order to achieve the  $\text{Na}_2\text{O}/\text{SiO}_2$  ratio required, and this too may be  
378 significant in inhibiting the formation of zeolites, as steric restrictions may reduce the growth  
379 of germ nuclei which lead to the formation of crystalline zeolites (Barrer 1981; Provis et al.  
380 2005). Sample 7Na also shows a small degree of atmospheric carbonation, seen through the  
381 formation of aragonite, a metastable polymorph of  $\text{CaCO}_3$ .

382

383 Following exposure to solvents, a significant quantity of faujasite develops only in 8Na after  
384 exposure to MEA (Figure 6b) and water (Figure 6c). A smaller quantity of faujasite is seen in  
385 the 9Na formulation exposed to water, and slight traces in 9Na following exposure to  $\text{K}_2\text{CO}_3$   
386 (Figure 6d). Significant calcite formation is again observed during  $\text{K}_2\text{CO}_3$  exposure (Figure

387 6d) through decalcification, and kalcinite ( $\text{KHCO}_3$ , PDF # 12-0297) has also formed in all  
388  $\text{K}_2\text{CO}_3$  exposed samples through the reaction of  $\text{K}_2\text{CO}_3$  with  $\text{CO}_2$  and  $\text{H}_2\text{O}$ . Corresponding  
389 data and additional discussion related to the metakaolin sample series with increasing  
390  $\text{H}_2\text{O}/\text{SiO}_2$  ratios are presented as Supporting Information (Figure S1 and associated text).  
391 These data show the formation of zeolites in each sample, although their prevalence is  
392 reduced by  $\text{K}_2\text{CO}_3$  exposure, which appears to hinder the structural evolution and  
393 crystallisation of the zeolites.

394

395 The XRD data for the fly ash-based formulations are presented in Figures 7-8 and in the  
396 Supporting Information, again divided by oxide ratio series as outlined in Table 2. In these  
397 samples, a basic Na-sodalite phase (approximately,  $1.08\text{Na}_2\text{O}\cdot\text{Al}_2\text{O}_3\cdot 1.68\text{SiO}_2\cdot 1.8\text{H}_2\text{O}$ , PDF#  
398 31-1271) was formed; this is a feldspathoid structure, closely related to zeolites, and appears  
399 to be present in all samples regardless of solvent exposure. Sample 17Na, the pure fly ash  
400 binder, also contains trona ( $\text{Na}_3(\text{CO}_3)(\text{HCO}_3)\cdot 2\text{H}_2\text{O}$ ) as a product of slight atmospheric  
401 carbonation. Also present in all of the fly ash based formulations were the unreactive phases  
402 supplied by the fly ash, being quartz and mullite.

403

404 Figure 7 shows the data for the sample series with increasing  $\text{SiO}_2/\text{Al}_2\text{O}_3$  ratio. The formation  
405 of basic sodalite appears to reduce as the Si content increases, likely due to the reduction in  
406 Al available for crystallisation, as this phase forms with  $\text{SiO}_2/\text{Al}_2\text{O}_3 = 2$  and so requires high  
407 Al availability for crystallisation. The sodalite appears to be largely unaffected by exposure to  
408 MEA (Figure 7b). Si and Al leaching was also low for the fly ash formulations immersed in  
409 MEA (Figures 1 and 2), and residual strength on exposure to solvents was high (Figure 4). It  
410 therefore seems that for the fly ash formulations, 28 days of exposure to MEA has not

411 significantly altered the expected path of microstructural development, with the slight losses  
412 in strength attributed to alkali attack on the framework.

413

414 The basic sodalite phase also develops in almost all samples following exposure to water  
415 (Figure 7c), but in concentrations again decreasing though the sample series. However,  
416 sample 18Na showed instead the formation of a faujasite phase as the main zeolitic product,  
417 similar to the fly ash-only sample 17Na. Sample 18Na had the lowest Si/Al ratio of the  
418 samples in this series. However, faujasite was not formed in the 18Na sample subjected to  
419 other immersion conditions, so its formation here may be due to the effectively increased  
420 water content on immersion, and the effect of the subsequent diffusion of the contents of the  
421 pore solution out into the leaching solution. Exposure to water is also accompanied by a  
422 significant amount of carbonation, with the formation of calcite in all samples of the fly ash-  
423 slag series.

424

425 A small amount of sodalite was also formed on exposure to  $K_2CO_3$  (Figure 7d). Carbonation  
426 also occurred in all samples exposed to  $K_2CO_3$  with the formation of calcite  $CaCO_3$  (PDF#  
427 47-1743) through decalcification and kalcinite,  $KHCO_3$  (PDF# 12-0292) through reaction of  
428  $K_2CO_3$ ,  $CO_2$  and  $H_2O$ . Some crystalline  $K_2CO_3$  (PDF# 70-0292) also precipitated in all  
429 samples. The high leachability of Na and Si in  $K_2CO_3$  (Figure 3) resulted in total strength loss  
430 after 28 days of solvent exposure (Figure 4), and this is consistent with the notable reduction  
431 in the amorphous 'hump' corresponding to the geopolymer gel in this leaching environment  
432 in particular (Figure 7d). The microstructural development has been significantly affected  
433 following 28 days of exposure to  $K_2CO_3$ , but despite this the mechanism of zeolite formation

434 found in the unexposed samples still persists, although to a lesser extent due to the leaching  
435 processes taking place.

436

437 The XRD data for the series with increasing  $\text{Na}_2\text{O}/\text{SiO}_2$  ratio are given in Figure 8. In the  
438 unexposed samples (Figure 8a), basic sodalite is more prominent with increasing Na content,  
439 with the exception of sample 27Na. In this formulation, the hydroxide and water contents are  
440 highest, and there is no trace of sodalite, but faujasite has formed instead. In sample 23Na,  
441 where hydroxide and water contents are low, no zeolitic phase has formed. Again, there is  
442 very little effect on the microstructure induced by exposure to MEA (Figure 8b) or water  
443 (Figure 8c). The same zeolites as in the unexposed samples, faujasite (in 27Na) and basic  
444 sodalite (in 24Na-26Na), have again developed (but to a lesser extent than in the unexposed  
445 samples) after exposure to  $\text{K}_2\text{CO}_3$  (Figure 8d).

446

447 The XRD data for the series with increasing  $\text{H}_2\text{O}/\text{SiO}_2$  ratio are shown in the Supporting  
448 Information. Basic sodalite was formed in all of the unexposed samples, and there appears to  
449 be very little difference in the crystalline phases formed, other than calcite formation in all  
450 samples, and slight suppression of crystallisation by  $\text{K}_2\text{CO}_3$  immersion. This series had the  
451 highest unexposed and also residual strengths upon exposure after 28 days of immersion, and  
452 it appears that these five samples have the least alteration in crystallographic structure among  
453 those studied.

454

### 455 **3.3 Implications**

456

457 The specimens tested here show notable degradation under exposure to carbon capture  
458 solvents, and as such would not be directly useful in applications where they are brought into  
459 immediate contact with these solvents in service. However, this was a part of the design of  
460 this study, which is in effect an accelerated test to simulate a worst-case leaching scenario.  
461 The resistance of geopolymers to alkali attack and carbonation would certainly be higher if  
462 more mature binders were tested, but the aim of this study has been to elucidate the  
463 degradation mechanisms through induction of artificially rapid degradation processes. The  
464 key to geopolymer concrete durability in carbon capture applications will be the development  
465 of low porosity and high tortuosity of the pore structure, in order to reduce the accessibility of  
466 the key binder phases. A forthcoming study will directly address the pore network geometry  
467 of these binder systems, and provide further insight into this important issue. In any case, the  
468 results of this study show that the resistance of geopolymers to attack by aqueous  $K_2CO_3$  is  
469 not as high as was predicted based upon nanostructural arguments, and this solvent does  
470 cause significant damage to the aluminosilicate geopolymer framework.

471

472

#### 473 **4. Conclusions**

474 Alkali activated geopolymer concretes may provide a cost saving alternative construction  
475 material for use in place of large stainless steel vessels in carbon capture facilities. This study  
476 was based around the study of immature and relatively porous binders, in order to best  
477 understand the mechanism of chemical attack on geopolymers during exposure to carbon  
478 capture solvents. Leaching occurred rapidly, within one day of immersion, due to the release  
479 of alkalis from the pore solution and attack on loosely bound Si and Al species within these  
480 immature binders. Resistance to MEA is highest among the solvents tested, with low  
481 solubility of Na in this solvent.

482

483 The microstructural development of the metakaolin/slag geopolymer formulations was more  
484 significantly affected by solvent immersion than was the structure of the fly ash/slag  
485 geopolymers. Solvent immersion seems to suppress the formation of faujasite in the  
486 metakaolin/slag binders, and the residual strength of these binders is relatively low. However,  
487 the formation of basic sodalite, which is the predominant crystalline phase observed in the fly  
488 ash/slag formulations, is much less influenced by solvent immersion, particularly in the case  
489 of immersion in MEA or water, where the 28 day residual strength is high. The reduced  
490 extent of binder development observed on exposure to  $K_2CO_3$  (evidenced by the lower degree  
491 of zeolite formation and also the dissolution of much of the amorphous geopolymer gel  
492 phase), along with the effects of significant carbonation does, in the majority of cases, caused  
493 catastrophic strength failure of the matrix on exposure to  $K_2CO_3$ .

494

495

## 496 **5. Acknowledgements**

497 The authors acknowledge the funding provided by the Australian Government through its  
498 CRC Program for this CO2CRC research project, and the salary support of JLP provided  
499 through the Discovery Grants program via APD and DORA Fellowships.

500

## 501 **6. References**

502 Abora, K., Beleña, I., Bernal, S.A., Dunster, A., Nixon, P.A., Provis, J.L., Tagnit-Hamou, A.  
503 and Winnefeld, F. (2014). Durability and testing - Chemical matrix degradation  
504 processes. in: Alkali-Activated Materials: State-of-the-Art Report, RILEM TC 224-  
505 AAM. Eds. J.L. Provis and J.S.J. van Deventer. Dordrecht, Springer/RILEM: 177-  
506 221.

- 507 Aly, Z., Vance, E.R., Perera, D.S., Hanna, J.V., Griffith, C.S., Davis, J. and Durce, D. (2008).  
 508 "Aqueous leachability of metakaolin-based geopolymers with molar ratios of  
 509 Si/Al=1.5-4." Journal of Nuclear Materials **378**(2): 172-179.
- 510 Bakharev, T. (2005). "Resistance of geopolymer materials to acid attack." Cement and  
 511 Concrete Research **35**(4): 658-670.
- 512 Barrer, R.M. (1981). "Zeolites and their synthesis." Zeolites **1**(3): 130-140.
- 513 Bernal, S.A., Mejía de Gutiérrez, R., Provis, J.L. and Rose, V. (2010). "Effect of silicate  
 514 modulus and metakaolin incorporation on the carbonation of alkali silicate-activated  
 515 slags." Cement and Concrete Research **40**(6): 898-907.
- 516 Bernal, S.A., Provis, J.L., Brice, D.G., Kilcullen, A., Duxson, P. and van Deventer, J.S.J.  
 517 (2012). "Accelerated carbonation testing of alkali-activated binders significantly  
 518 underestimates service life: The role of pore solution chemistry." Cement and  
 519 Concrete Research **42**(10): 1317-1326.
- 520 Davidovits, J. (1991). "Geopolymers - Inorganic polymeric new materials." Journal of  
 521 Thermal Analysis **37**(8): 1633-1656.
- 522 Duxson, P., Fernandez-Jimenez, A., Provis, J.L., Lukey, G.C., Palomo, A. and van Deventer,  
 523 J.S.J. (2007a). "Geopolymer technology: the current state of the art." Journal of  
 524 Materials Science **42**(9): 2917-2933.
- 525 Duxson, P., Lukey, G.C. and van Deventer, J.S.J. (2006). "Thermal evolution of metakaolin  
 526 geopolymers: Part 1 - Physical evolution." Journal of Non-Crystalline Solids **352**(52-  
 527 54): 5541-5555.
- 528 Duxson, P., Provis, J.L., Lukey, G.C. and Van Deventer, J.S.J. (2007b). "The role of  
 529 inorganic polymer technology in the development of 'green concrete'." Cement and  
 530 Concrete Research **37**(12): 1590-1597.
- 531 Fernández-Jiménez, A. and Palomo, A. (2009). Chemical durability of geopolymers. in:  
 532 Geopolymers: Structure, Processing, Properties and Industrial Applications. Eds. J.L.  
 533 Provis and J.S.J. van Deventer. Cambridge, UK, Woodhead: 167-193.
- 534 Fernández-Jiménez, A., Palomo, A., Sobrados, I. and Sanz, J. (2006). "The role played by the  
 535 reactive alumina content in the alkaline activation of fly ashes." Microporous and  
 536 Mesoporous Materials **91**(1-3): 111-119.
- 537 Gordon, L.E., Provis, J.L. and van Deventer, J.S.J. (2011). "Durability of fly ash/GGBFS  
 538 based geopolymers exposed to carbon capture solvents." Advances in Applied  
 539 Ceramics **110**(8): 446-452.
- 540 Jenkins, C.R., Cook, P.J., Ennis-King, J., Undershultz, J., Boreham, C., Dance, T., de Caritat,  
 541 P., Etheridge, D.M., Freifeld, B.M., Hortle, A., Kirste, D., Paterson, L., Pevzner, R.,  
 542 Schacht, U., Sharma, S., Stalker, L. and Urosevic, M. (2012). "Safe storage and  
 543 effective monitoring of CO<sub>2</sub> in depleted gas fields." Proceedings of the National  
 544 Academy of Sciences of the United States of America **109**(2): E35-E41.

- 545 Juenger, M.C.G., Winnefeld, F., Provis, J.L. and Ideker, J.H. (2011). "Advances in alternative  
546 cementitious binders." Cement and Concrete Research **41**(12): 1232-1243.
- 547 Kumar, S., Kumar, R. and Mehrotra, S.P. (2010). "Influence of granulated blast furnace slag  
548 on the reaction, structure and properties of fly ash based geopolymer." Journal of  
549 Materials Science **45**(3): 607-615.
- 550 Li, C., Li, Y., Sun, H. and Li, L. (2011). "The composition of fly ash glass phase and its  
551 dissolution properties applying to geopolymeric materials." Journal of the American  
552 Ceramic Society **94**(6): 1773-1778.
- 553 Lloyd, R.R. (2009). Accelerated ageing of geopolymers. in: Geopolymers: Structure,  
554 Processing, Properties and Industrial Applications. Eds. J.L. Provis and J.S.J. van  
555 Deventer. Cambridge, UK, Woodhead: 139-166.
- 556 Lloyd, R.R., Provis, J.L. and van Deventer, J.S.J. (2010). "Pore solution composition and  
557 alkali diffusion in inorganic polymer cement." Cement and Concrete Research **40**(9):  
558 1386-1392.
- 559 Lloyd, R.R., Provis, J.L. and van Deventer, J.S.J. (2012). "Acid resistance of inorganic  
560 polymer binders. 1. Corrosion rate." Materials and Structures **45**(1-2): 1-14.
- 561 McLellan, B.C., Corder, G.D., Giurco, D.P. and Ishihara, K.N. (2012). "Renewable energy in  
562 the minerals industry: a review of global potential." Journal of Cleaner Production **32**:  
563 32-44.
- 564 Olivier, J.G.J., Janssens-Maenhout, G. and Peters, J.A.H.W. (2012). Trends in global CO<sub>2</sub>  
565 emissions; 2012 Report. The Hague, Netherlands, PBL Netherlands Environmental  
566 Assessment Agency: 40 pp.
- 567 Palomo, A., Blanco-Varela, M.T., Granizo, M.L., Puertas, F., Vazquez, T. and Grutzeck,  
568 M.W. (1999). "Chemical stability of cementitious materials based on metakaolin."  
569 Cement and Concrete Research **29**(7): 997-1004.
- 570 Perera, D.S., Vance, E.R., Aly, Z., Davis, J. and Nicholson, C.L. (2006). Immobilization of  
571 Cs and Sr in geopolymers with Si/Al molar ratio of ~2. in: Environmental Issues and  
572 Waste Management Technologies in the Ceramic and Nuclear Industries XI (Ceramic  
573 Transactions Vol. 176), John Wiley & Sons: 91-96.
- 574 Phair, J.W. and van Deventer, J.S.J. (2001). "Effect of silicate activator pH on the leaching  
575 and material characteristics of waste-based inorganic polymers." Minerals  
576 Engineering **14**(3): 289-304.
- 577 Pietersen, H.S., Fraay, A.L.A. and Bijen, J.M. (1989). Reactivity of fly ash at high pH. 1989  
578 Materials Research Society Fall Meeting (MRS Proceedings Vol. 178), Boston, MA,  
579 Cambridge University Press.
- 580 Provis, J.L. (2014). "Geopolymers and other alkali activated materials - Why, how, and  
581 what?" Materials and Structures **47**(1): 11-25.

- 582 Provis, J.L., Duxson, P. and van Deventer, J.S.J. (2010). "The role of particle technology in  
583 developing sustainable construction materials." Advanced Powder Technology **21**(1):  
584 2-7.
- 585 Provis, J.L., Lukey, G.C. and van Deventer, J.S.J. (2005). "Do geopolymers actually contain  
586 nanocrystalline zeolites? A reexamination of existing results." Chemistry of Materials  
587 **17**(12): 3075-3085.
- 588 Provis, J.L., Myers, R.J., White, C.E., Rose, V. and van Deventer, J.S.J. (2012). "X-ray  
589 microtomography shows pore structure and tortuosity in alkali-activated binders."  
590 Cement and Concrete Research **42**(6): 855-864.
- 591 Provis, J.L. and van Deventer, J.S.J., Eds. (2014). Alkali-Activated Materials: State-of-the-  
592 Art Report, RILEM TC 224-AAM. Dordrecht, Springer/RILEM.
- 593 Puertas, F., Palacios, M. and Vazquez, T. (2006). "Carbonation process of alkali-activated  
594 slag mortars." Journal of Materials Science **41**(10): 3071-3082.
- 595 Sindhunata, Provis, J.L., Lukey, G.C., Xu, H. and van Deventer, J.S.J. (2008). "Structural  
596 evolution of fly ash based geopolymers in alkaline environments." Industrial &  
597 Engineering Chemistry Research **47**(9): 2991-2999.
- 598 Taylor, H.F.W. (1997). Cement Chemistry, 2nd Ed. London, UK, Thomas Telford.
- 599 Temuujin, J., Minjigmaa, A., Lee, M., Chen-Tan, N. and van Riessen, A. (2011).  
600 "Characterisation of class F fly ash geopolymer pastes immersed in acid and alkaline  
601 solutions." Cement & Concrete Composites **33**(10): 1086-1091.
- 602 van Deventer, J.S.J., Provis, J.L. and Duxson, P. (2012). "Technical and commercial progress  
603 in the adoption of geopolymer cement." Minerals Engineering **29**: 89-104.
- 604 van Deventer, J.S.J., Provis, J.L., Duxson, P. and Brice, D.G. (2010). "Chemical research and  
605 climate change as drivers in the commercial adoption of alkali activated materials."  
606 Waste and Biomass Valorization **1**(1): 145-155.
- 607 Weng, L.Q., Sagoe-Crentsil, K., Brown, T. and Song, S.H. (2005). "Effects of aluminates on  
608 the formation of geopolymers." Materials Science and Engineering B-Solid State  
609 Materials for Advanced Technology **117**(2): 163-168.
- 610 Yip, C.K., Provis, J.L., Lukey, G.C. and van Deventer, J.S.J. (2008). "Carbonate mineral  
611 addition to metakaolin-based geopolymers." Cement & Concrete Composites **30**(10):  
612 979-985.
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- 615

616 Table 1. Compositions of aluminosilicate sources; wt.% as oxides, determined by X-ray  
617 fluorescence. LOI is loss on ignition at 1000 °C.

	<b>Metakaolin</b>	<b>Fly ash</b>	<b>GGBFS</b>
SiO <sub>2</sub>	52.8	45.5	32.9
Al <sub>2</sub> O <sub>3</sub>	39.2	27.8	13.2
CaO	0.1	5.6	40.1
Na <sub>2</sub> O	0.0	0.3	0.3
K <sub>2</sub> O	2.7	0.5	0.3
Fe <sub>2</sub> O <sub>3</sub>	0.8	11.2	0.3
MgO	0.2	1.4	6.0
P <sub>2</sub> O <sub>5</sub>	0.1	0.5	0.0
SO <sub>3</sub>	0.0	0.2	3.5
TiO <sub>2</sub>	0.0	1.4	0.7
MnO	0.0	0.2	0.0
LOI	1.2	2.7	1.2

618

619

620 Table 2. Component molar oxide ratios, NaOH concentrations and water/binder ratios (w/b;  
 621 'binder' is defined as the solid aluminosilicate precursor) of the geopolymer formulations  
 622 tested.

<b>Solid aluminosilicate</b>	<b>ID</b>	<b>SiO<sub>2</sub>/Al<sub>2</sub>O<sub>3</sub></b>	<b>Na<sub>2</sub>O/SiO<sub>2</sub></b>	<b>H<sub>2</sub>O/Na<sub>2</sub>O</b>	<b>NaOH (M)</b>	<b>w/b</b>
MK	1Na	3.0	0.5	13.0	6.9	0.44
MK:GGBFS 3:1	2Na	3.0	0.5	13.0	7.2	0.44
MK:GGBFS 3:1	3Na	3.3	0.5	13.0	6.9	0.44
MK:GGBFS 3:1	4Na	3.6	0.5	13.0	6.6	0.45
MK:GGBFS 3:1	7Na	3.3	0.33	13.0	6.0	0.37
MK:GGBFS 3:1	8Na	3.3	0.4	13.0	6.5	0.40
MK:GGBFS 3:1	9Na	3.3	0.5	13.0	6.9	0.44
MK:GGBFS 3:1	12Na	3.3	0.5	11.0	8.2	0.40
MK:GGBFS 3:1	13Na	3.3	0.5	12.0	7.5	0.42
MK:GGBFS 3:1	14Na	3.3	0.5	13.0	6.9	0.44
MK:GGBFS 3:1	15Na	3.3	0.5	14.0	6.4	0.46
MK:GGBFS 3:1	16Na	3.3	0.5	15.0	6.0	0.48
FA	17Na	3.33	0.45	8.0	10.6	0.32
FA:GGBFS 2:1	18Na	3.1	0.45	8.0	12.2	0.32
FA:GGBFS 2:1	19Na	3.2	0.45	8.0	12.0	0.32
FA:GGBFS 2:1	20Na	3.33	0.45	8.0	11.7	0.33
FA:GGBFS 2:1	21Na	3.4	0.45	8.0	11.6	0.33
FA:GGBFS 2:1	22Na	3.5	0.45	8.0	11.4	0.33
FA:GGBFS 2:1	23Na	3.33	0.30	8.0	10.3	0.24
FA:GGBFS 2:1	24Na	3.33	0.36	8.0	10.9	0.27
FA:GGBFS 2:1	25Na	3.33	0.45	8.0	11.7	0.33
FA:GGBFS 2:1	26Na	3.33	0.6	8.0	12.1	0.36
FA:GGBFS 2:1	27Na	3.33	0.9	8.0	12.7	0.42
FA:GGBFS 2:1	28Na	3.33	0.45	7.0	13.4	0.27
FA:GGBFS 2:1	29Na	3.33	0.45	8.0	11.7	0.33
FA:GGBFS 2:1	30Na	3.33	0.45	9.0	10.4	0.33
FA:GGBFS 2:1	31Na	3.33	0.45	10.0	9.4	0.35
FA:GGBFS 2:1	32Na	3.33	0.45	11.0	8.5	0.37

623 <sup>a</sup> Samples shaded in grey are the same mix (one in the metakaolin-based sample set and one in the fly ash-based  
 624 sample set), replicated and renumbered each time in the sample listing to place this composition in the correct  
 625 position in the series in which each of the synthesis parameters was varied.

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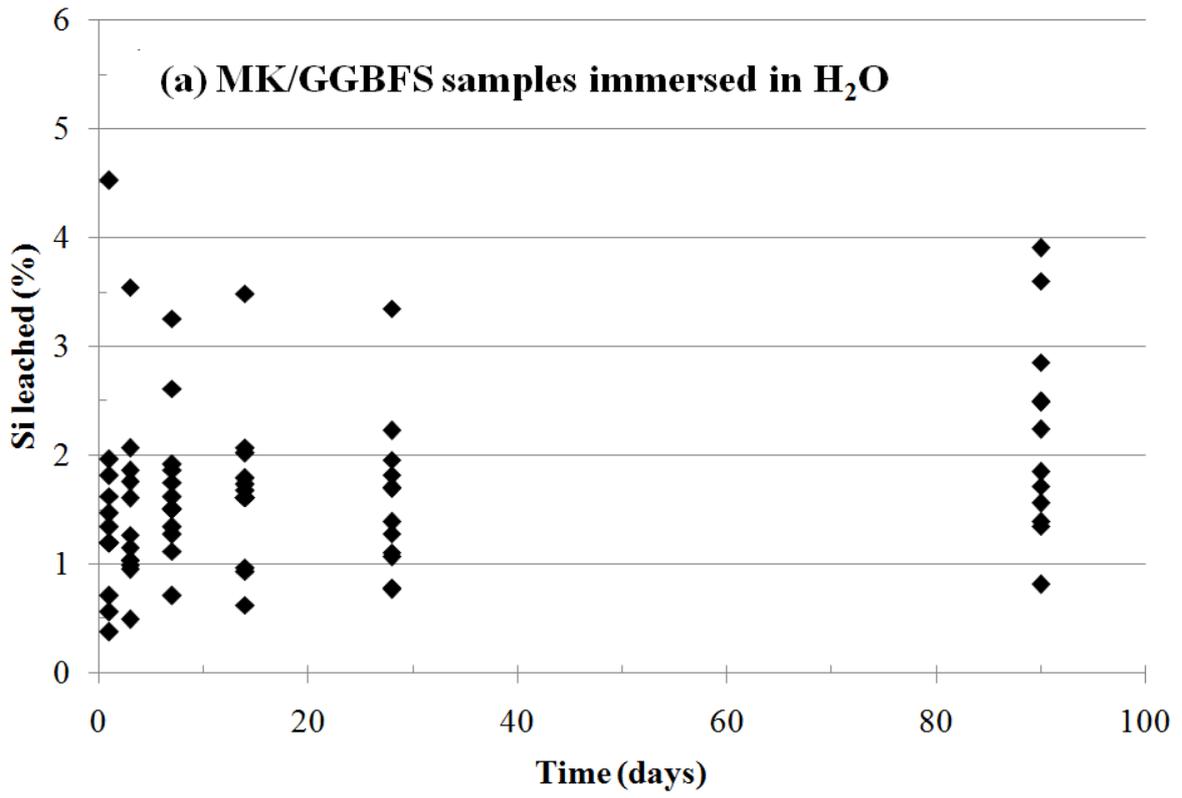
629 Table 3. Summary of sample series and the parameters varied in each.

Series	Aluminosilicate source	Parameter
2Na-4Na	Metakaolin-slag	Increasing $\text{SiO}_2/\text{Al}_2\text{O}_3$
7Na-9Na	Metakaolin-slag	Increasing $\text{Na}_2\text{O}/\text{SiO}_2$ (and w/b)
12Na-16Na	Metakaolin-slag	Increasing w/b
18Na-22Na	Fly ash-slag	Increasing $\text{SiO}_2/\text{Al}_2\text{O}_3$
23Na-27Na	Fly ash-slag	Increasing $\text{Na}_2\text{O}/\text{SiO}_2$ (and w/b)
28Na-32Na	Fly ash-slag	Increasing w/b

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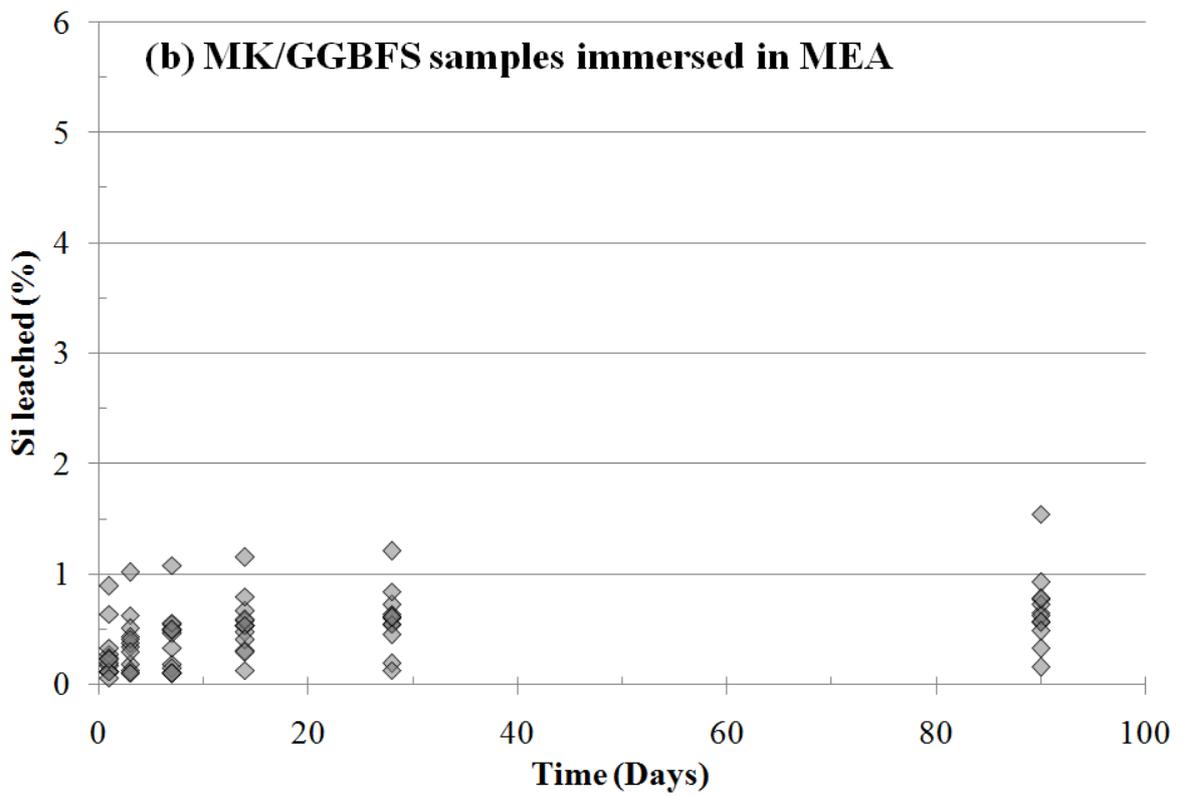
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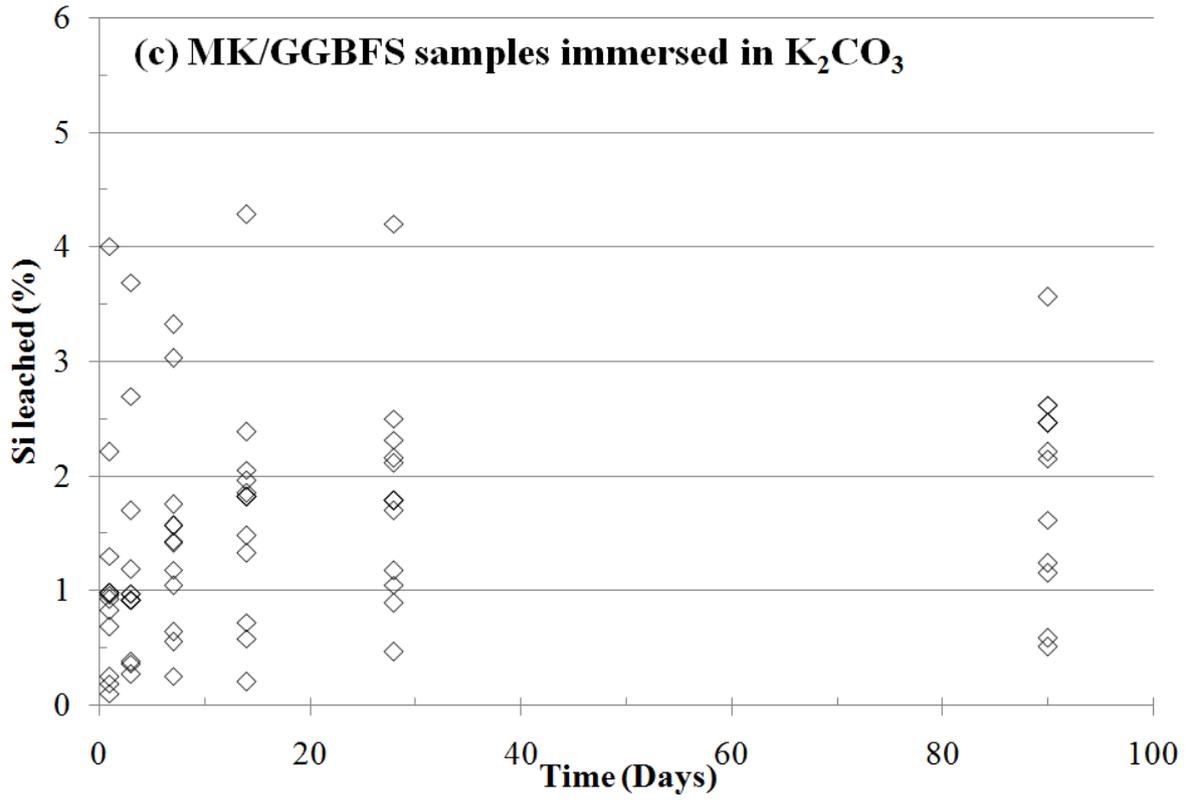


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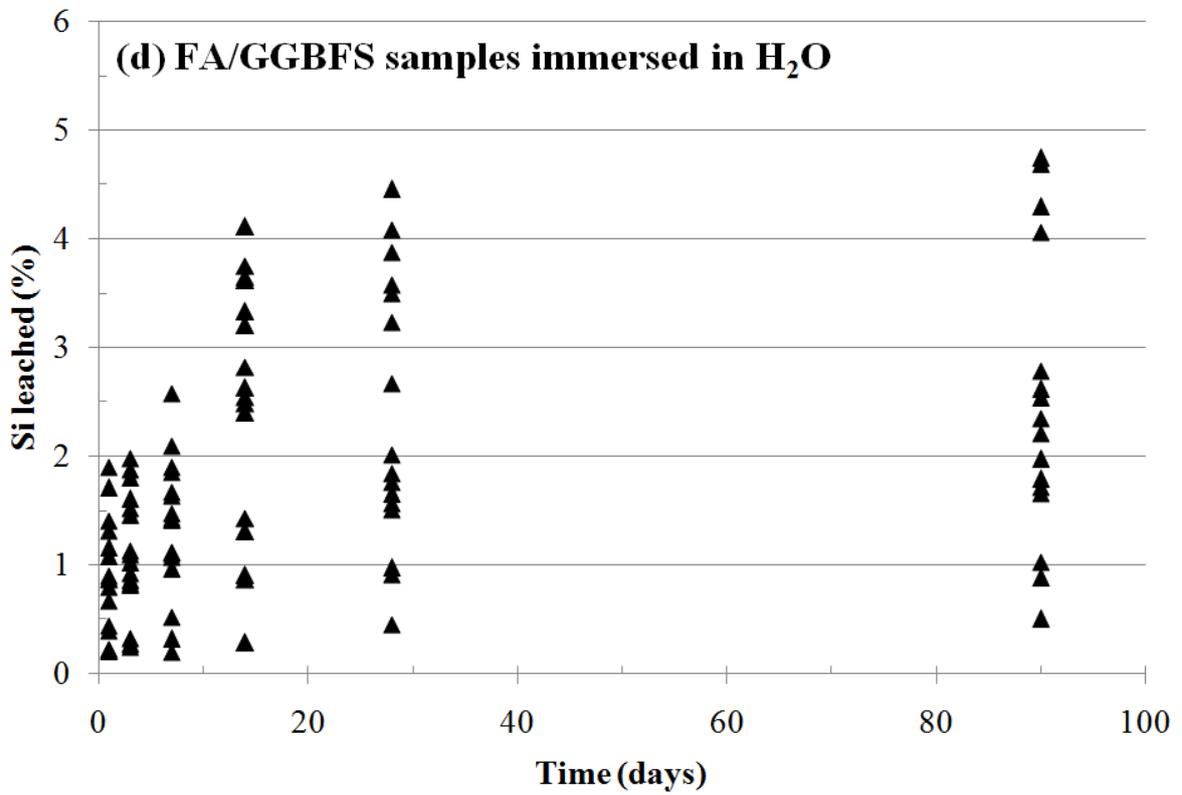


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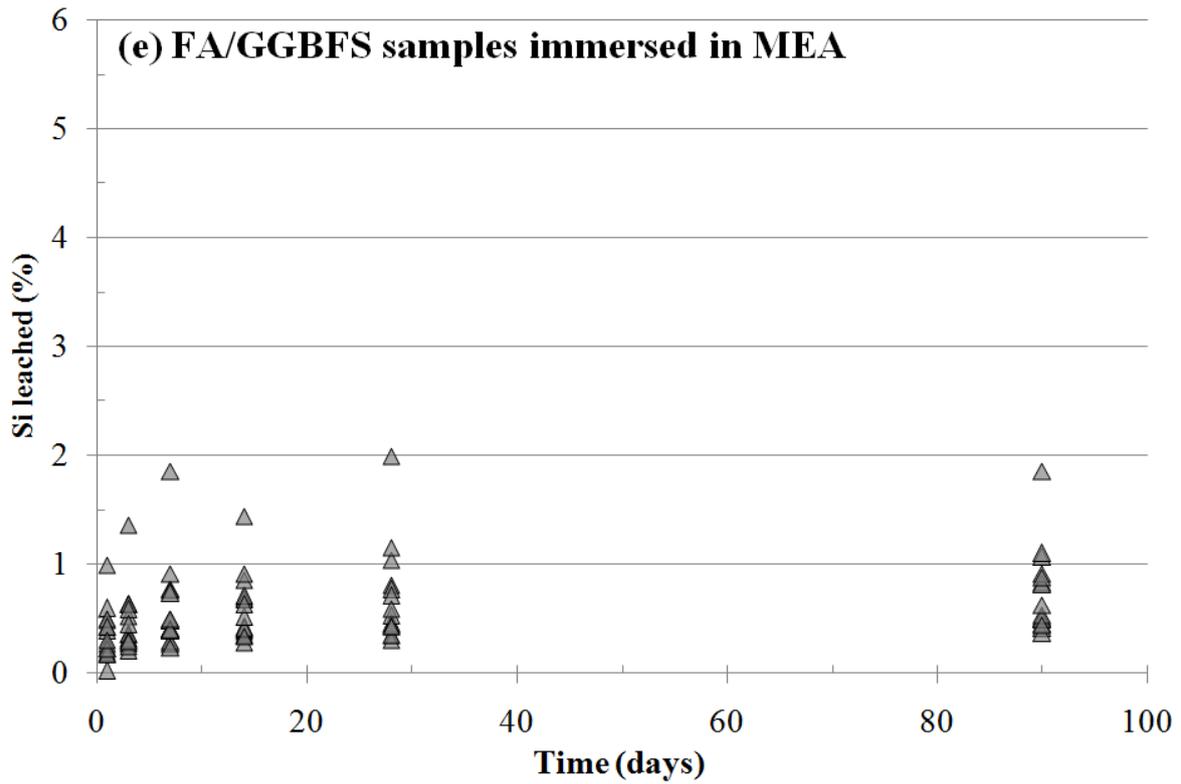


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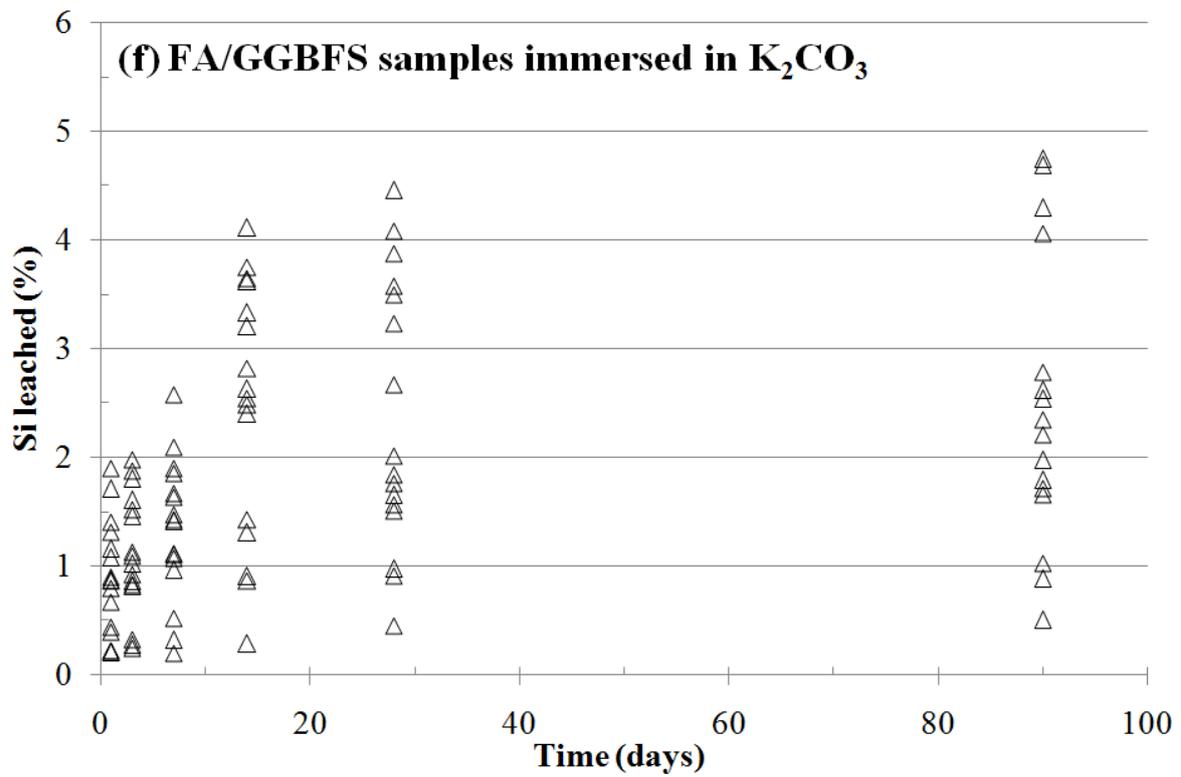
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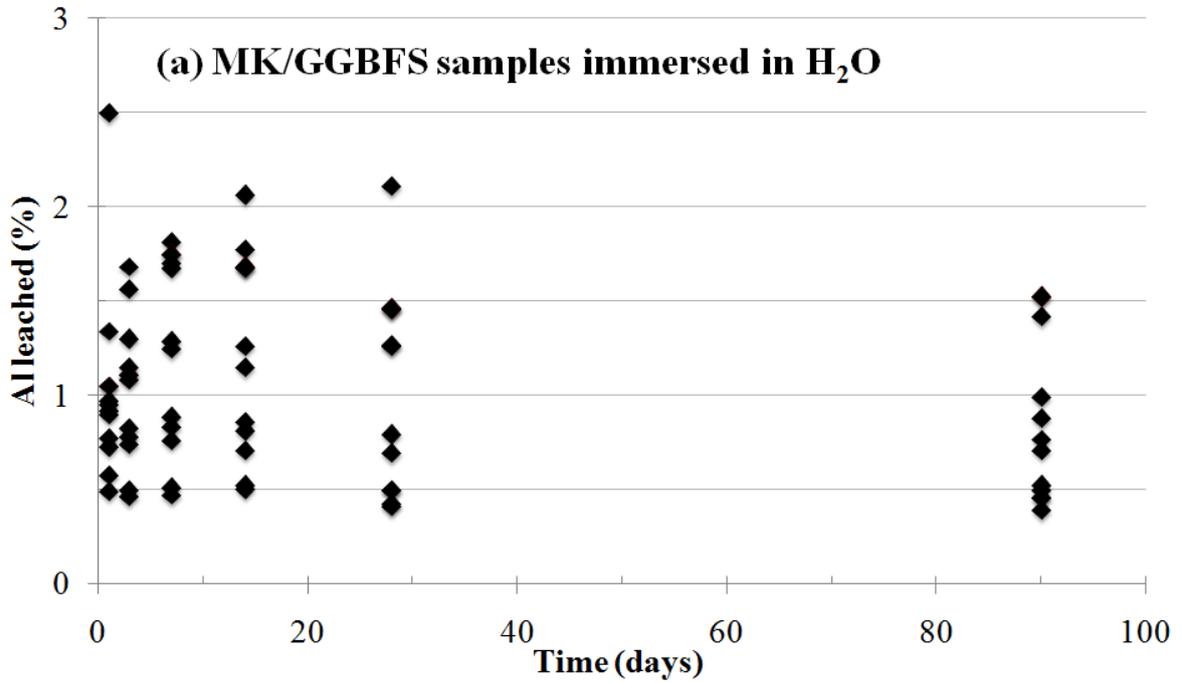


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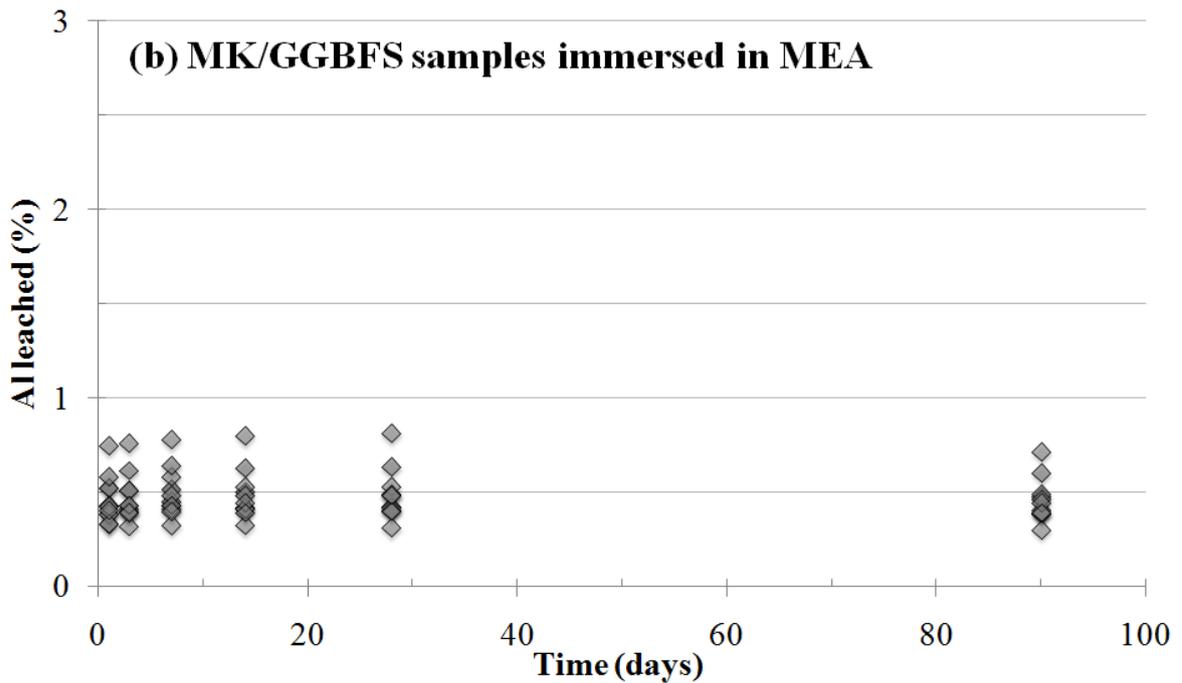


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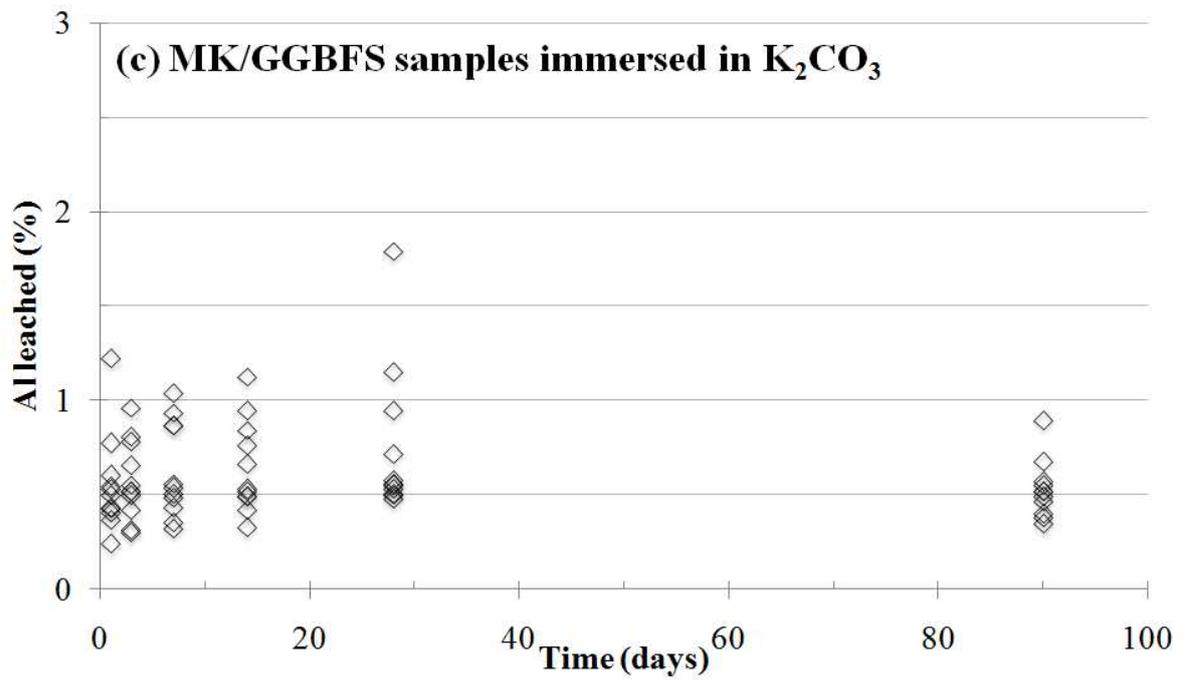
641 Figure 1. Leaching of Si from (a, b, c) the MK/GGBFS binders and (d, e, f) FA/GGBFS  
642 binders during 90 days of solvent exposure, (a, d) H<sub>2</sub>O, (b, e) MEA, (c, f) K<sub>2</sub>CO<sub>3</sub>



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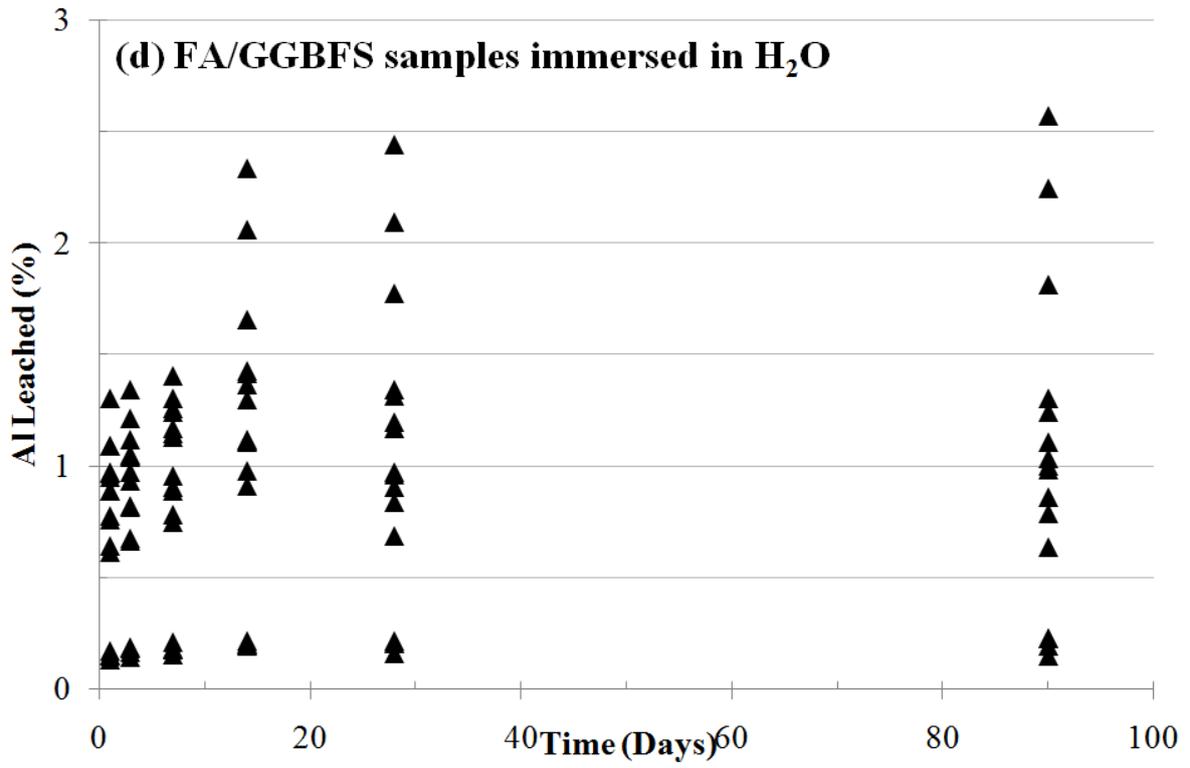
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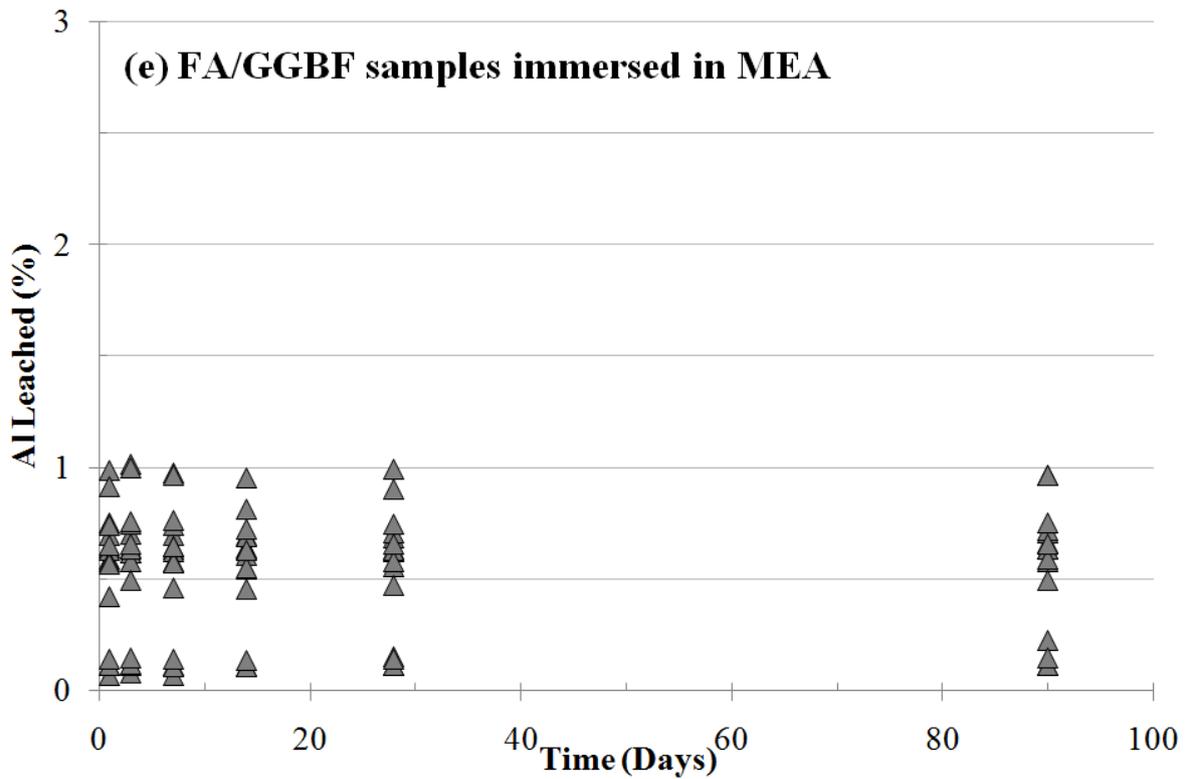
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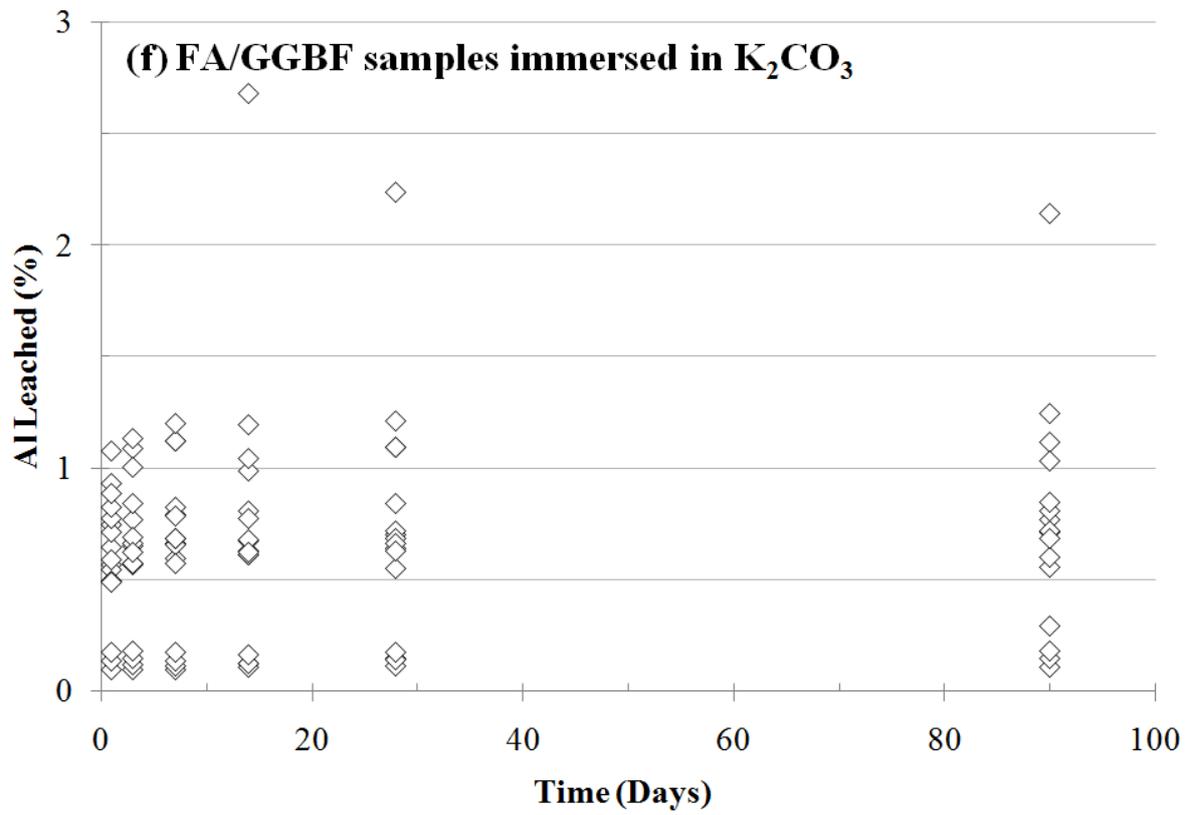
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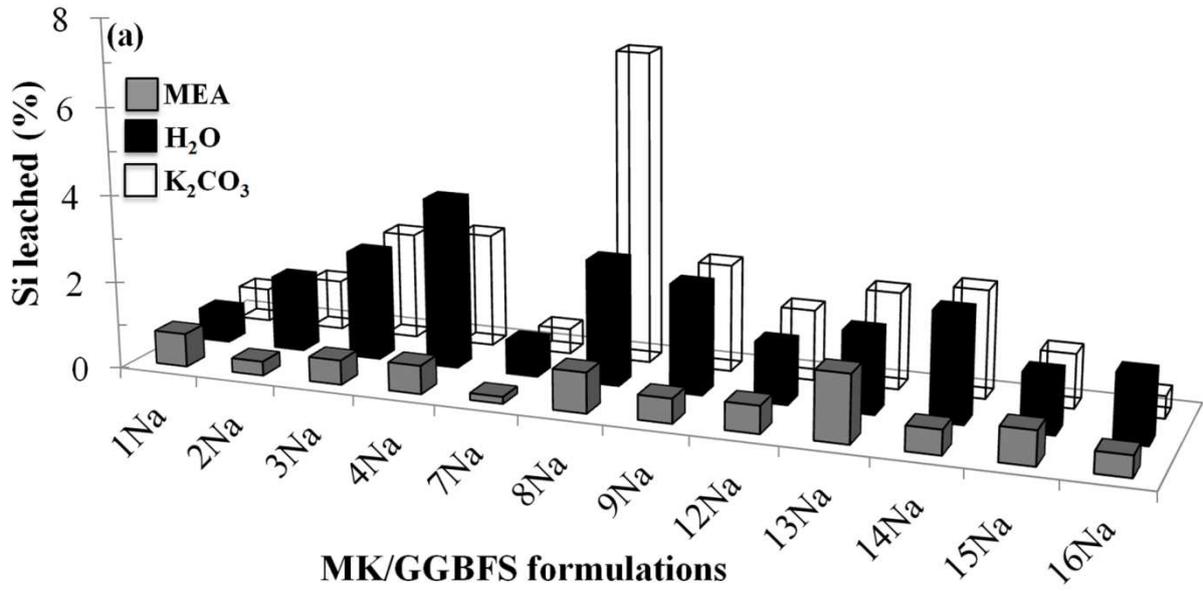


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651 Figure 2. Leaching of Al from (a, b, c) the MK/GGBFS binders, and (d, e, f) the FA/GGBFS  
 652 binders during 90 days of solvent exposure, (a, d)  $H_2O$ , (b, e) MEA, (c, f)  $K_2CO_3$

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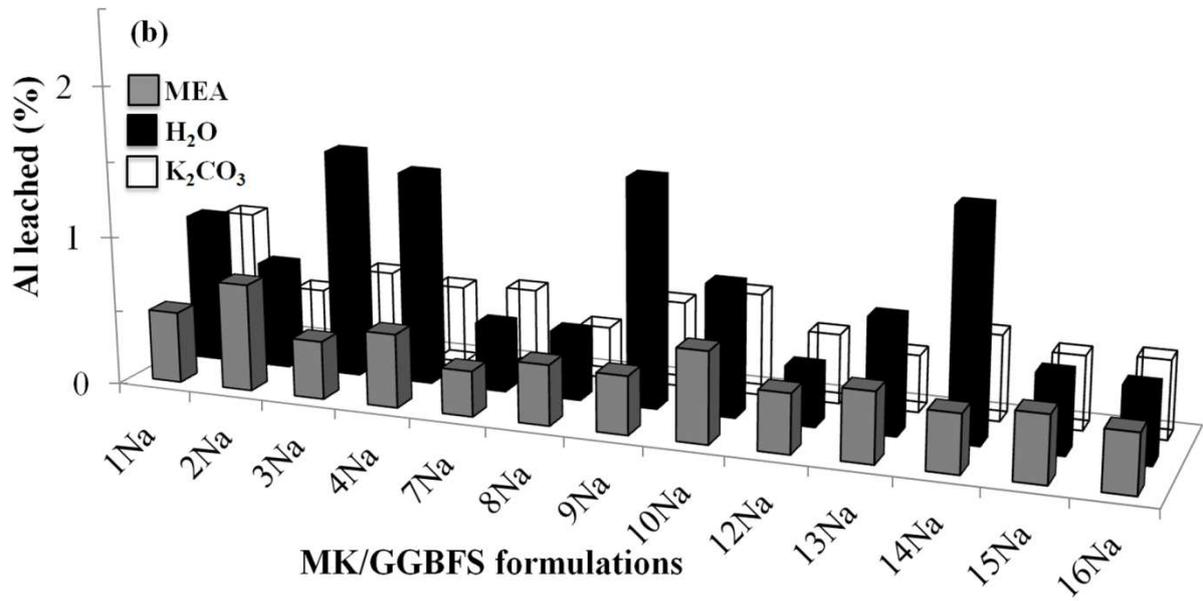
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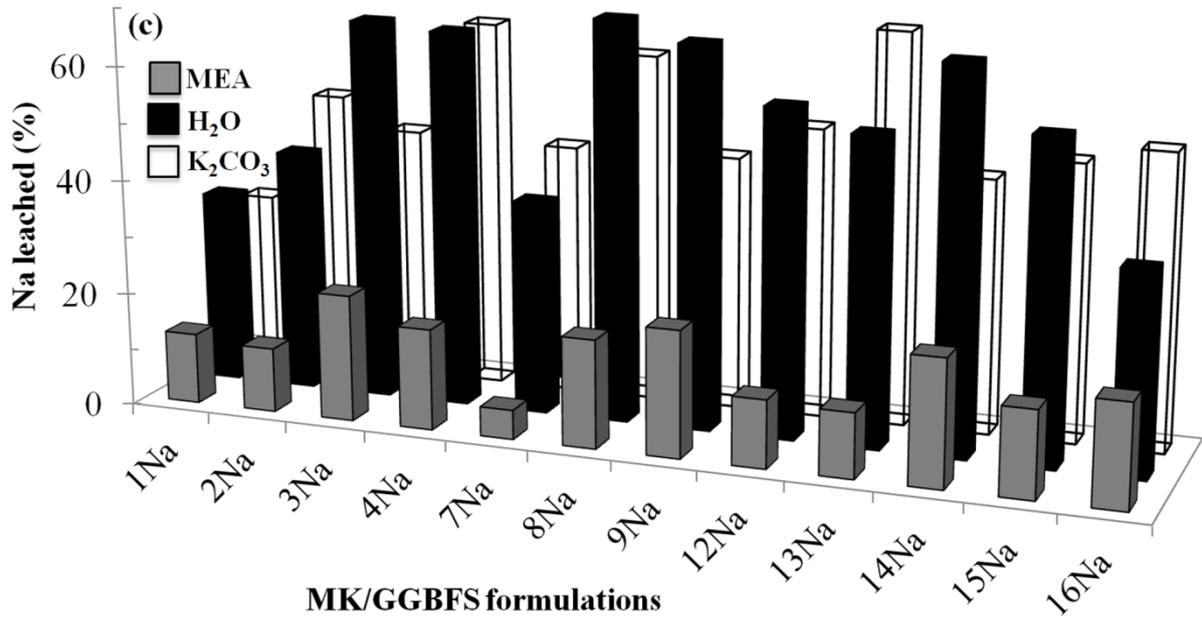
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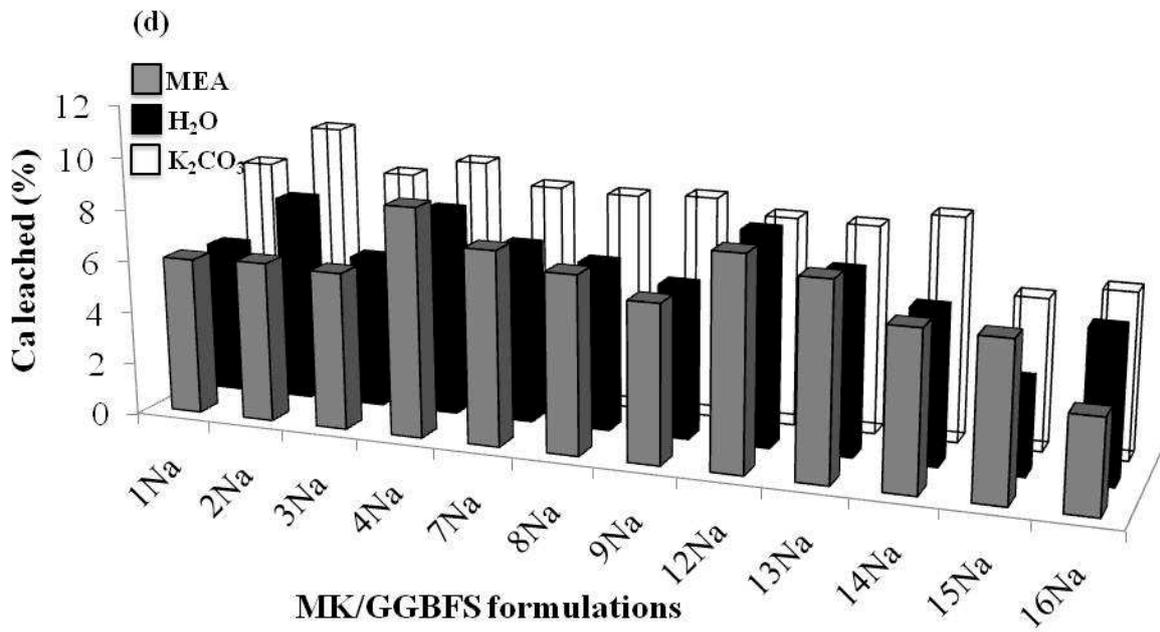
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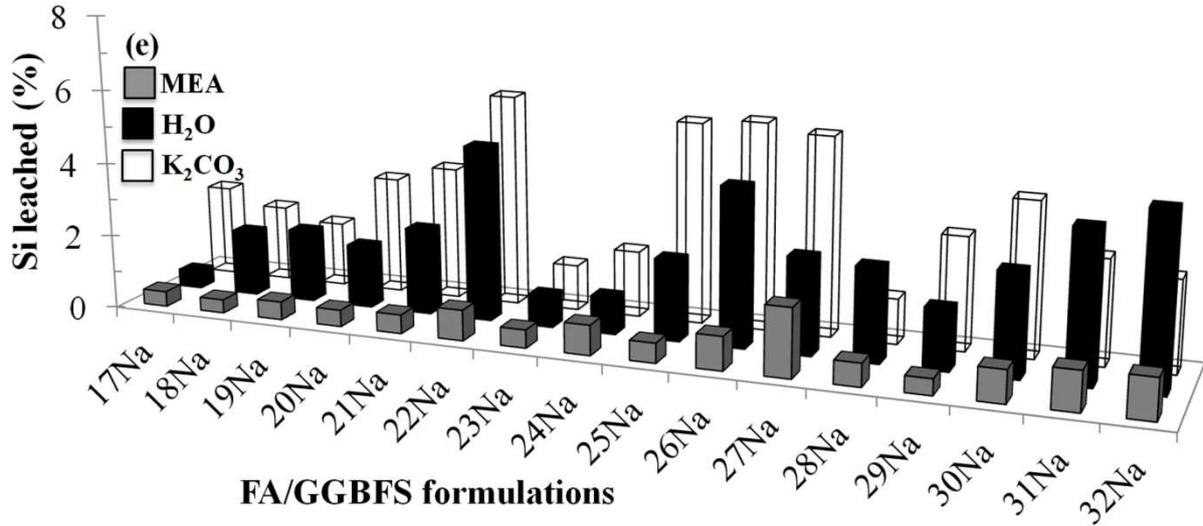
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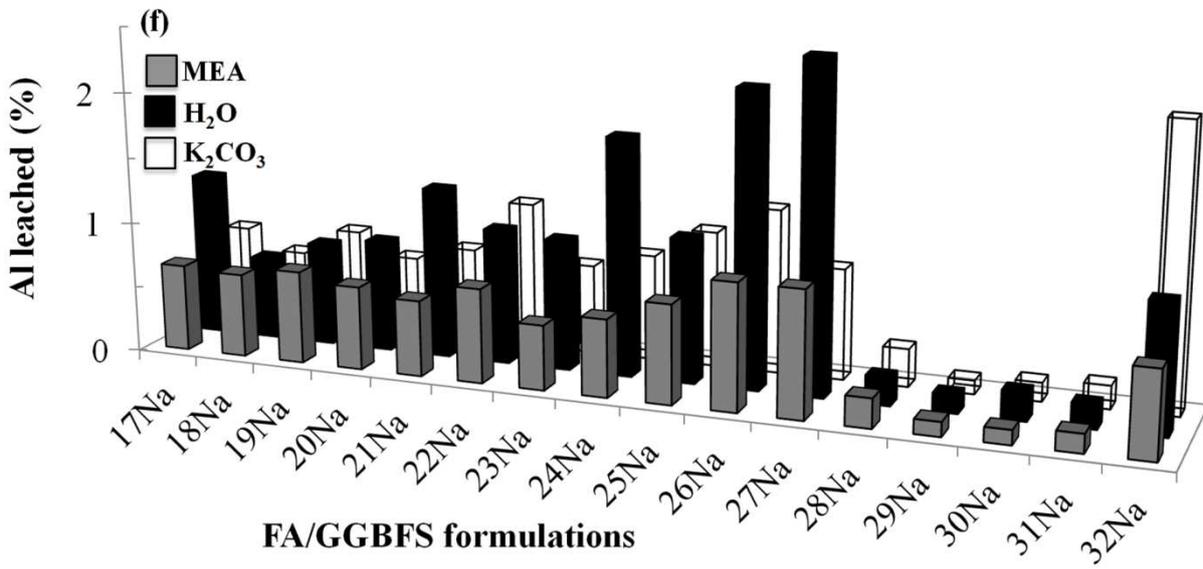
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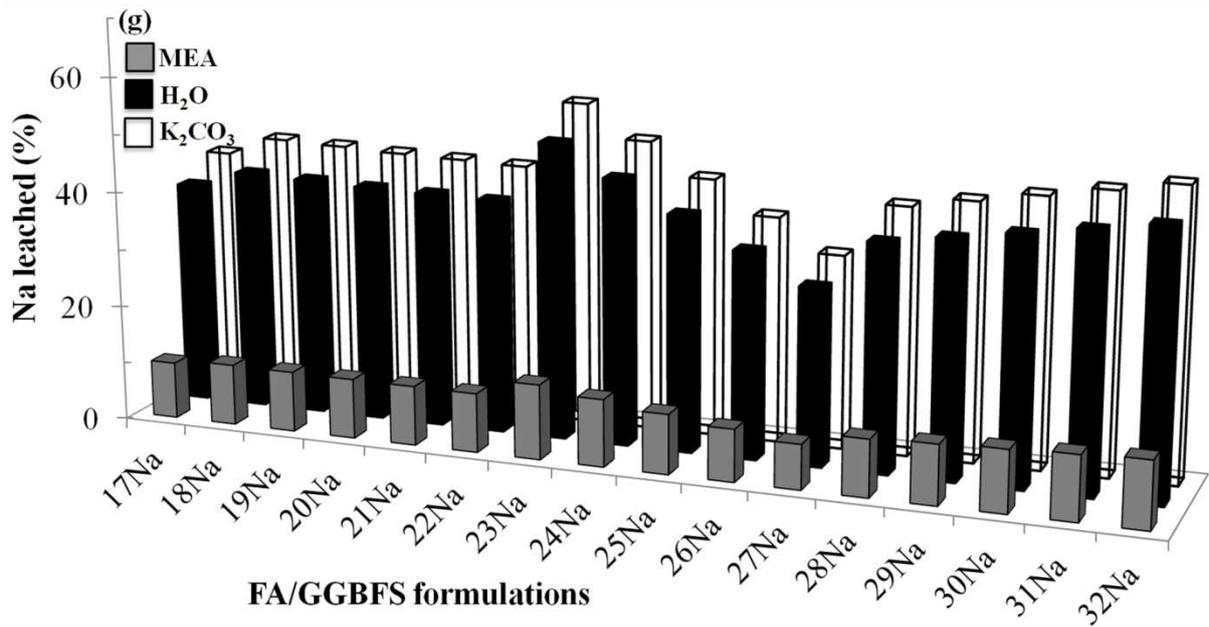
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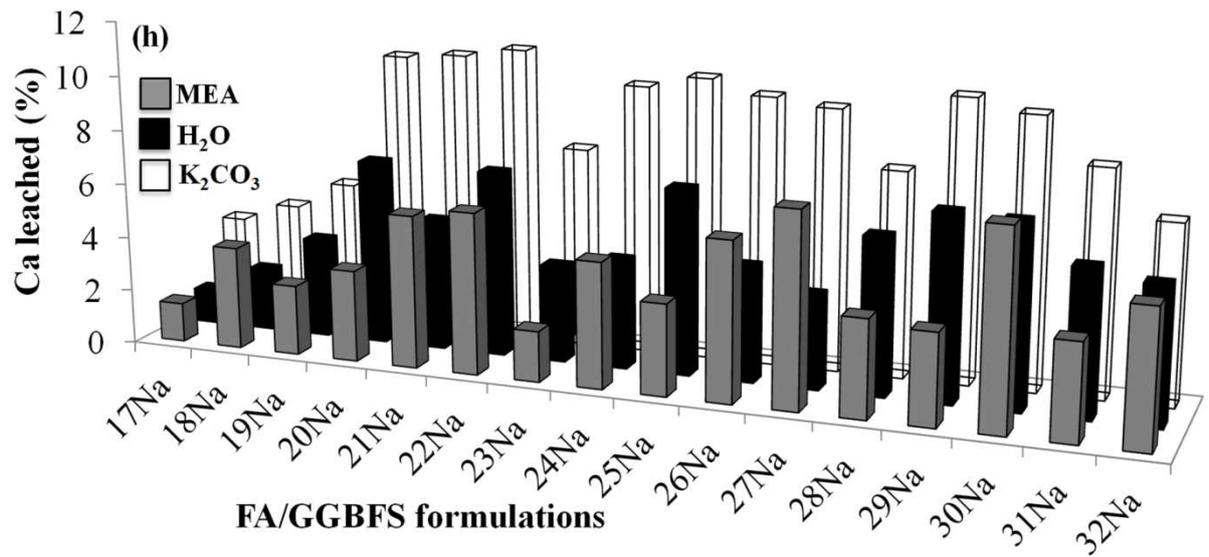
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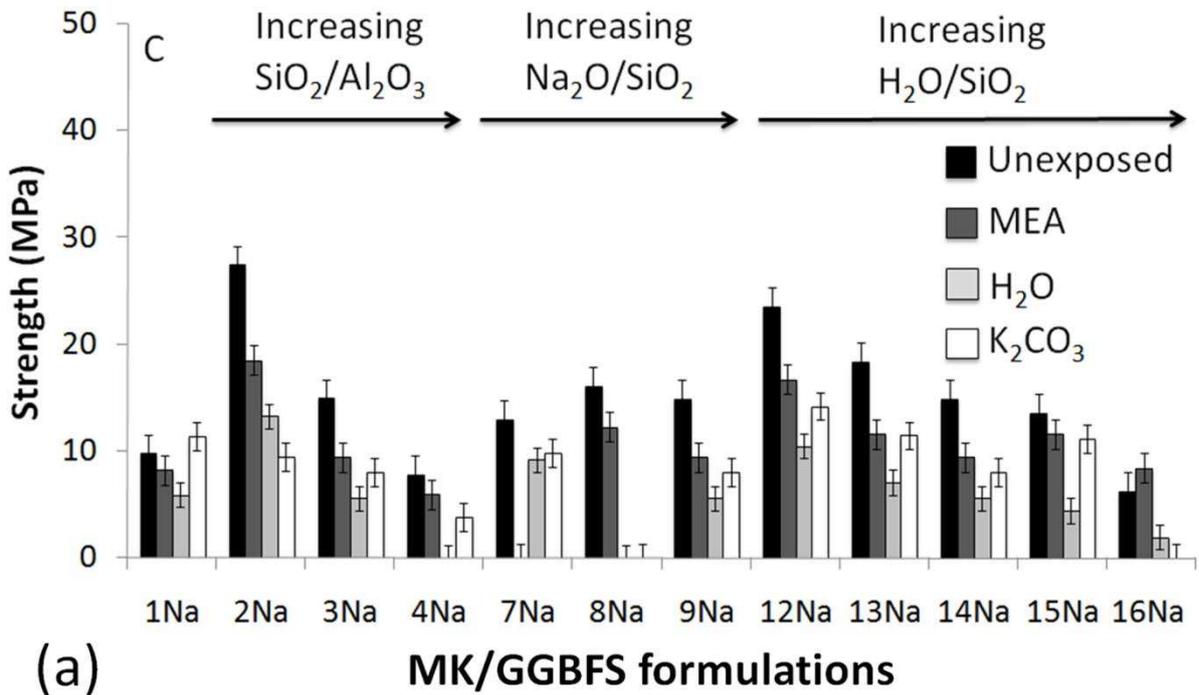
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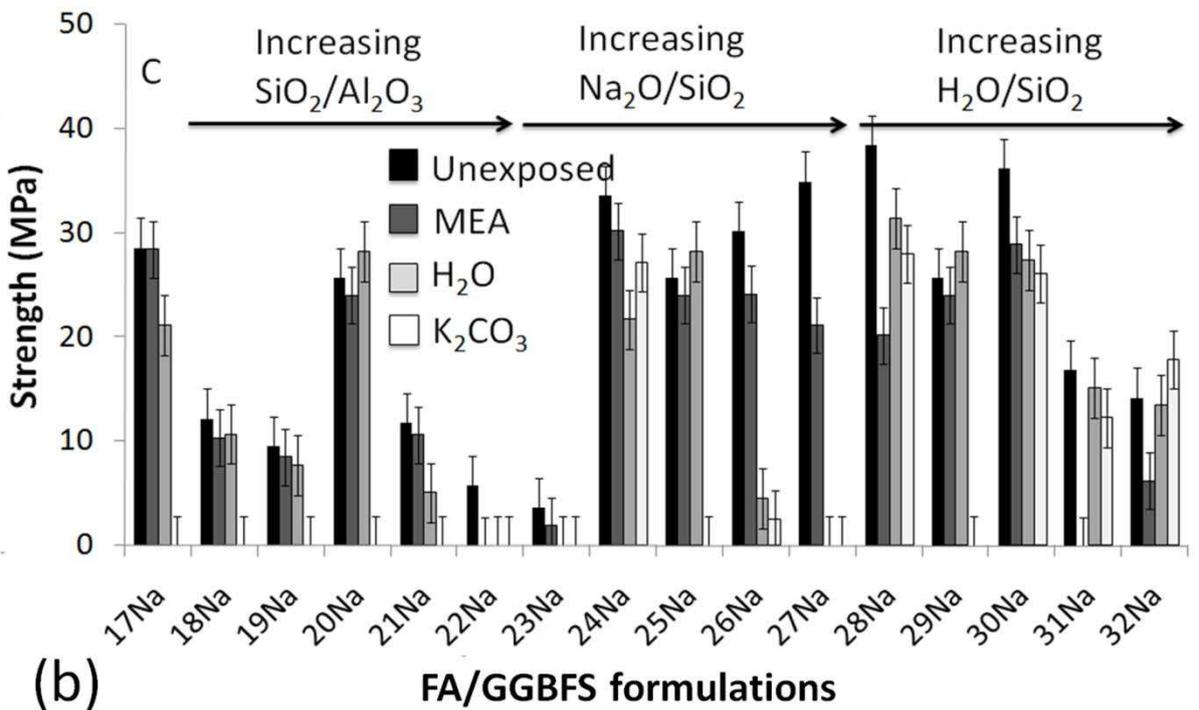
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665 Figure 3. Leaching of (a) Si (b) Al (c) Na and (d) Ca from each of the MK/GGBFS  
 666 geopolymer formulations, and of (e) Si, (f) Al, (g) Na and (h) Ca from each of the  
 667 FA/GGBFS geopolymer formulations, following 90 days of solvent exposure.

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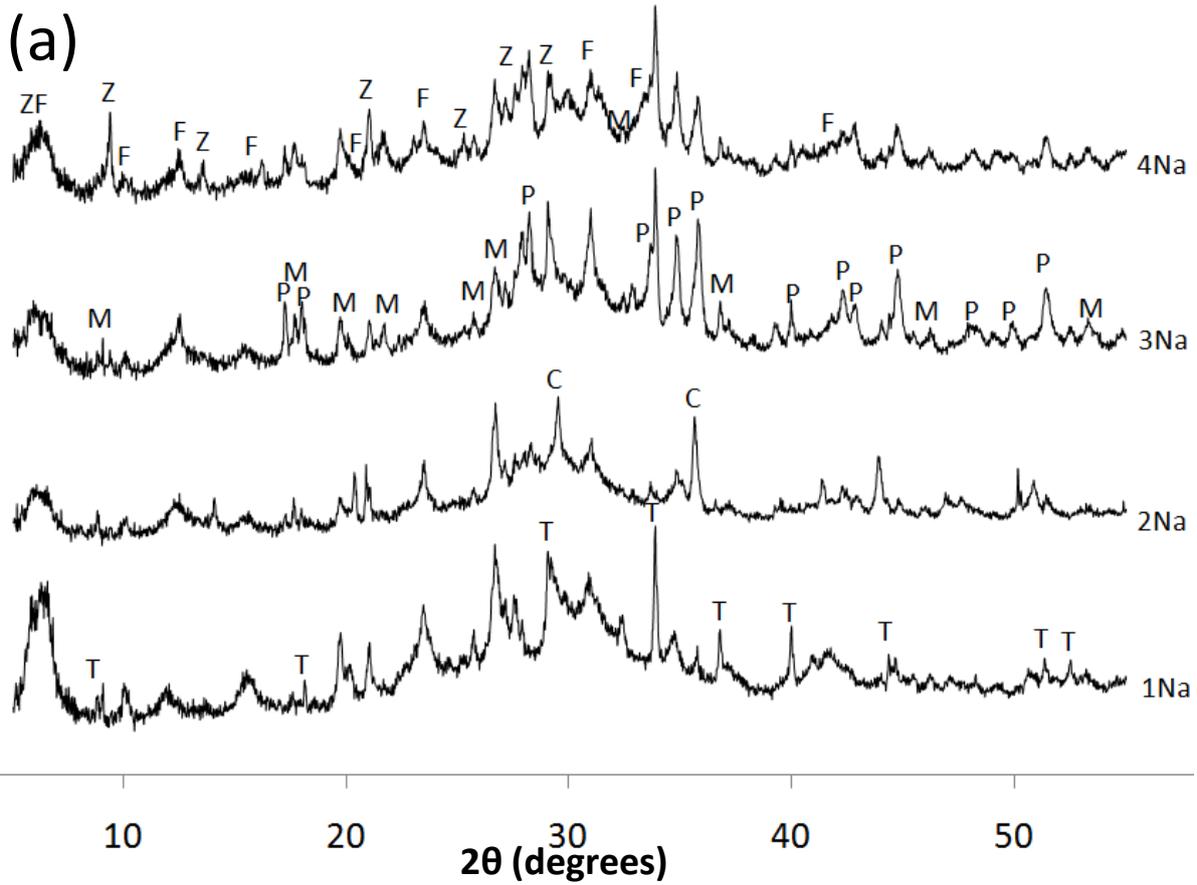
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672 Figure 4. Compressive strengths of the (a) MK/GGBFS and (b) FA/GBFS formulations after  
 673 28days of unexposed aging and 28 days of solvent exposure. The formulation trends in each  
 674 series are marked with arrows, and c denotes the control sample (without GGBFS) in each  
 675 set.

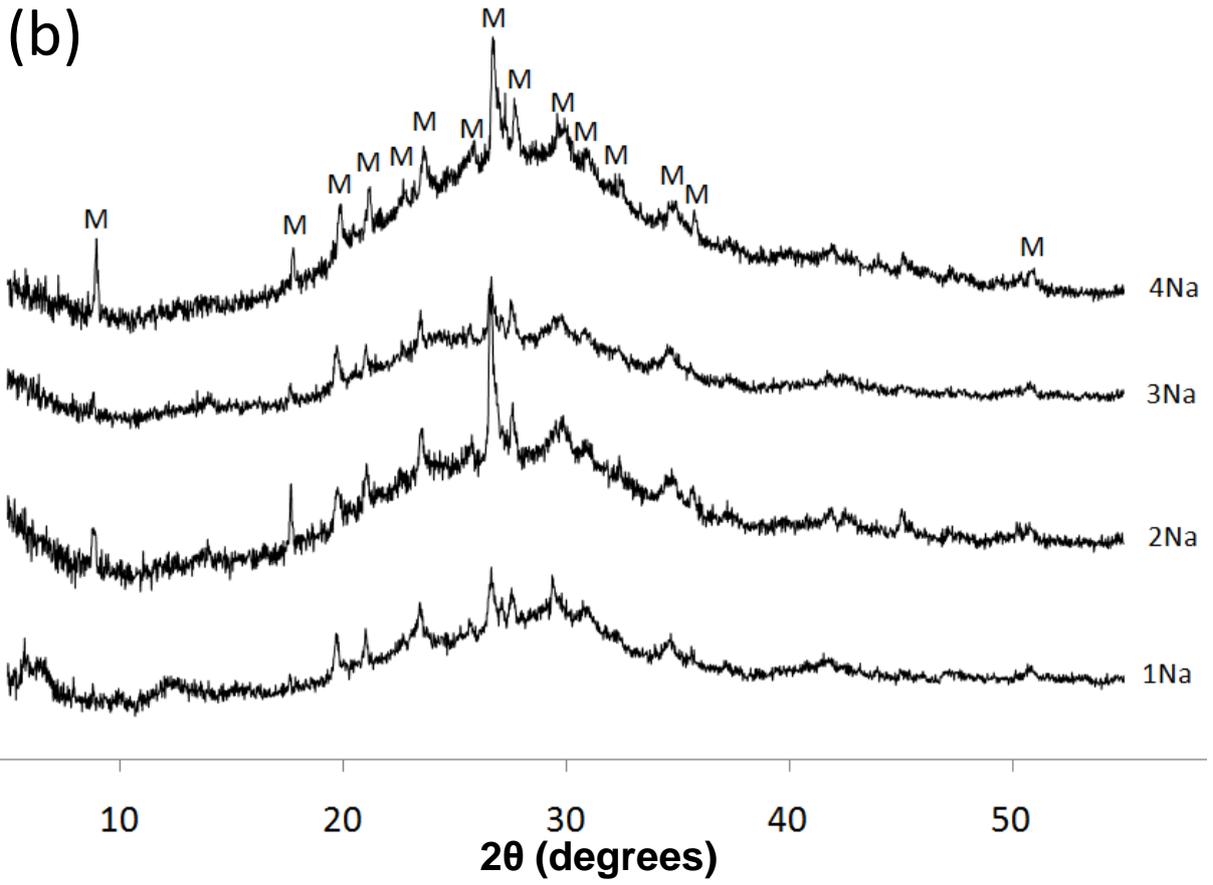
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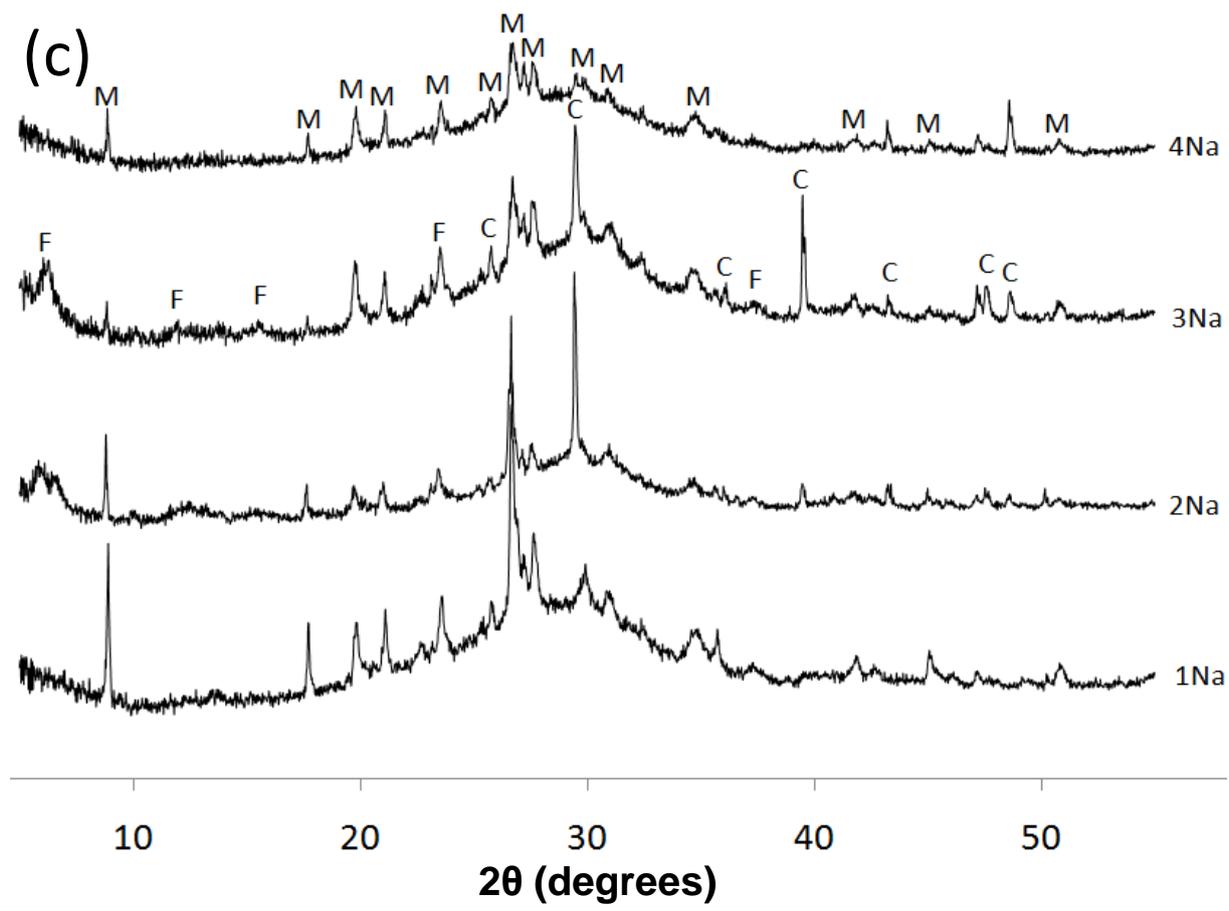
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(b)

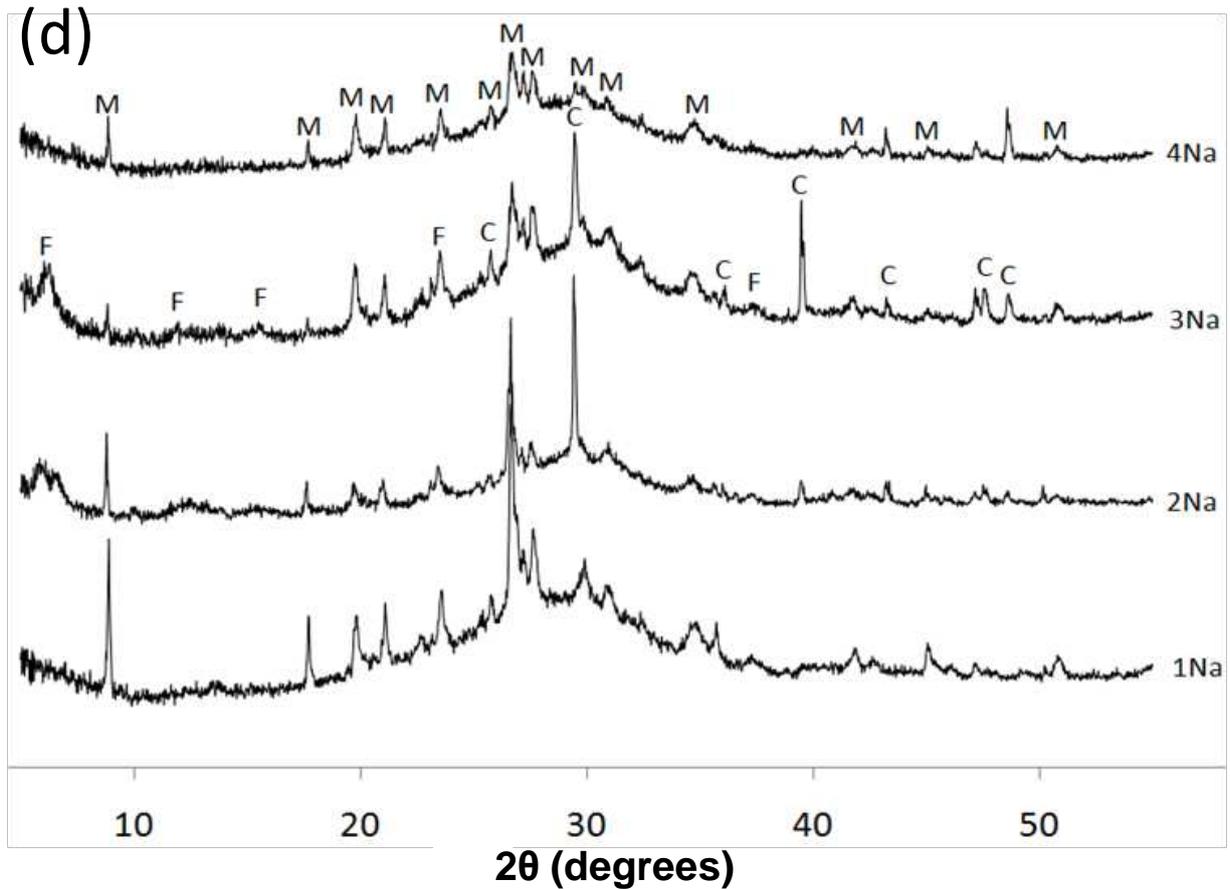


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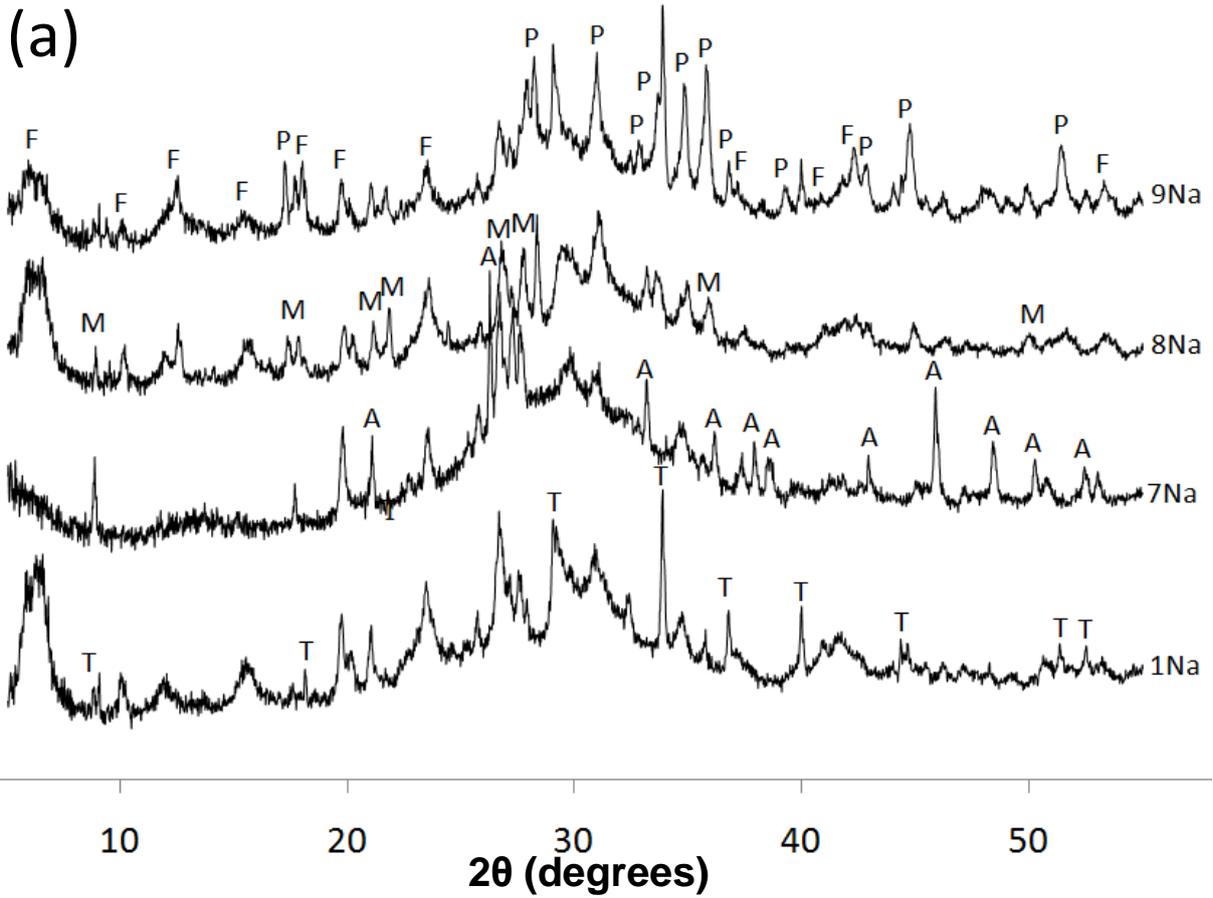


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687 Figure 5. XRD traces of MK/GGBFS samples 1Na-4Na after (a) 7 days of curing and 90 days  
688 of aging, or 7 days of curing and 90 days of exposure to solvents: (b) MEA; (c) Milli-Q H<sub>2</sub>O;  
689 (d) K<sub>2</sub>CO<sub>3</sub>. Phases marked are C: calcite, F: faujasite, M: muscovite, P: pirssonite, T: trona,  
690 Z: chabazite-Na.

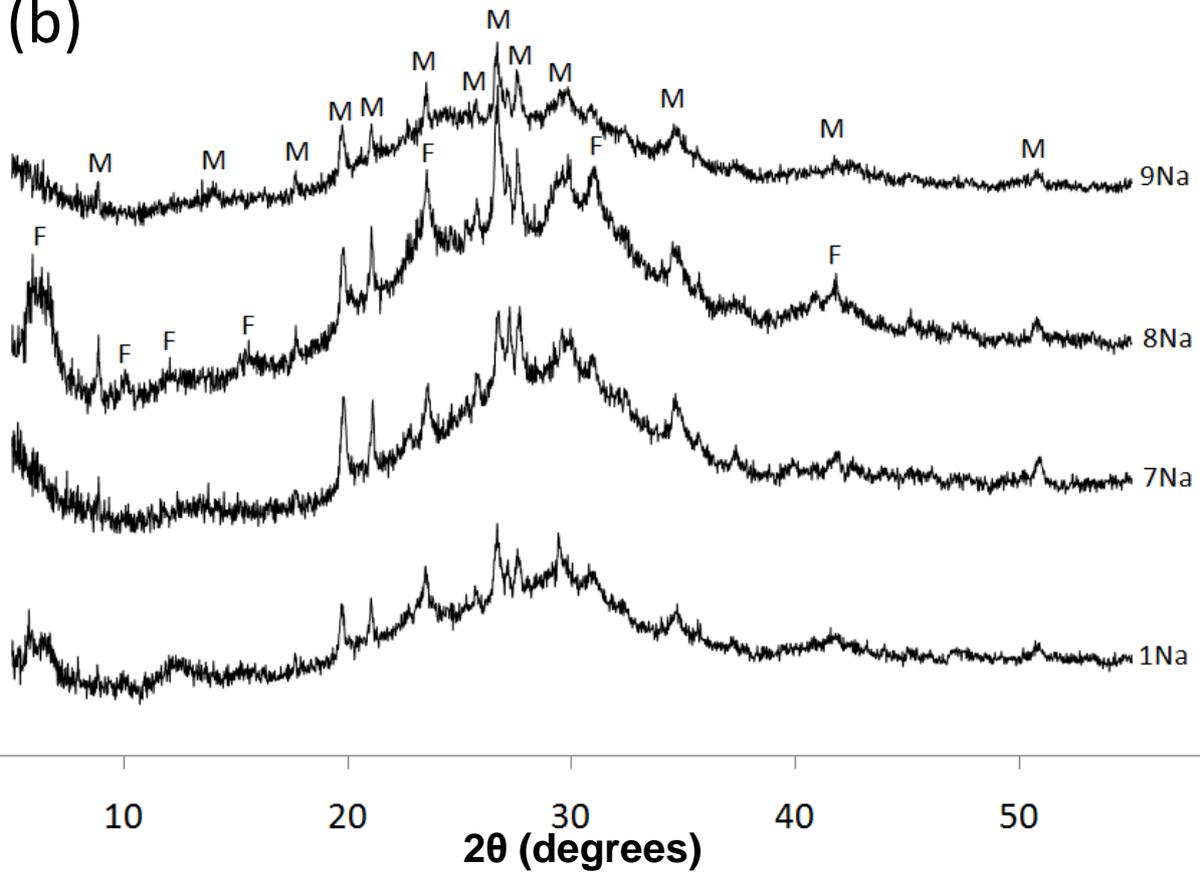
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(a)



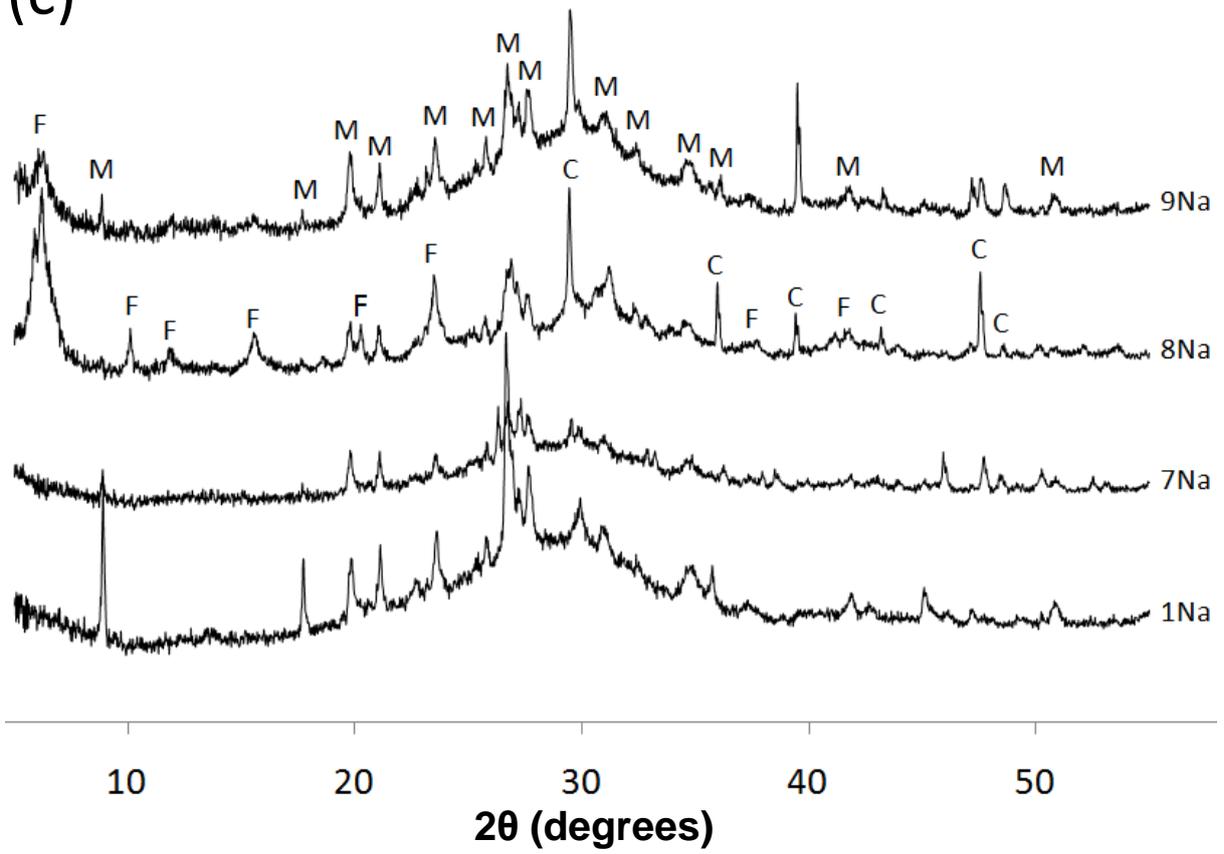
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(b)



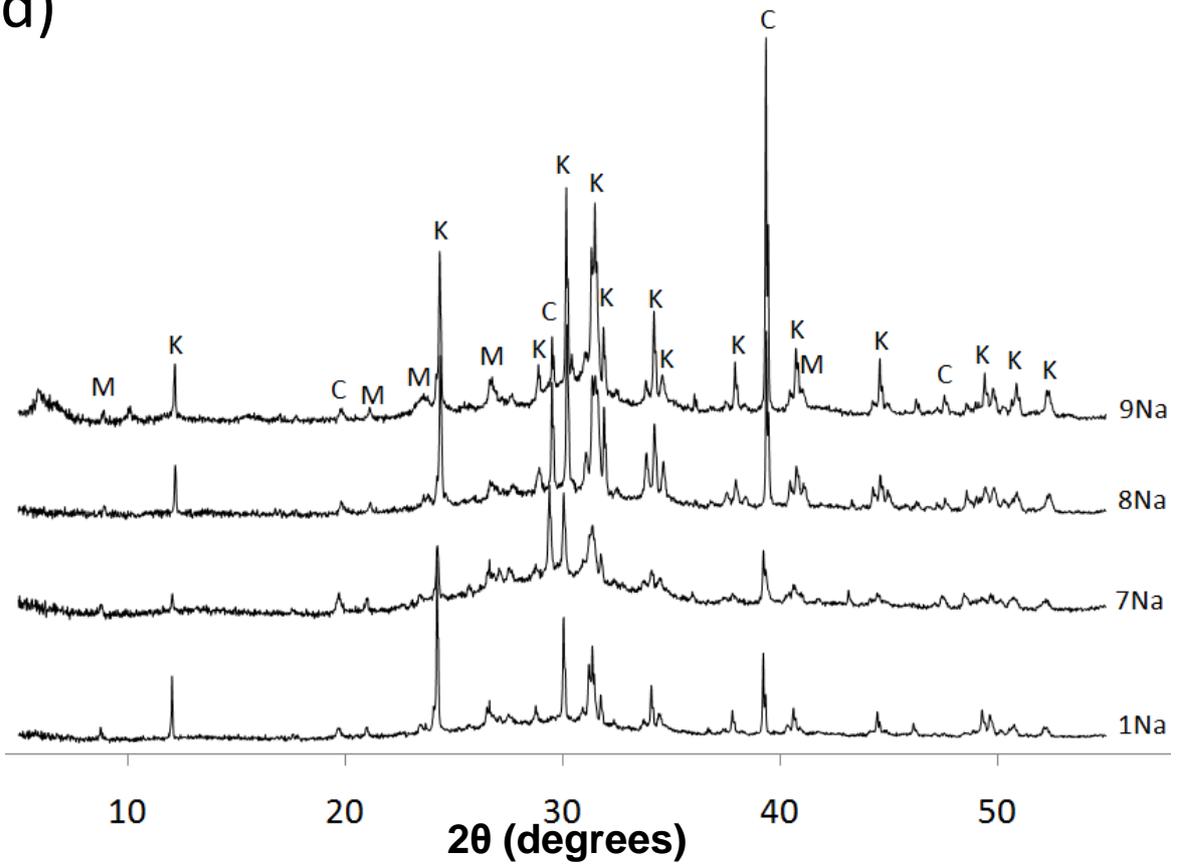
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(c)



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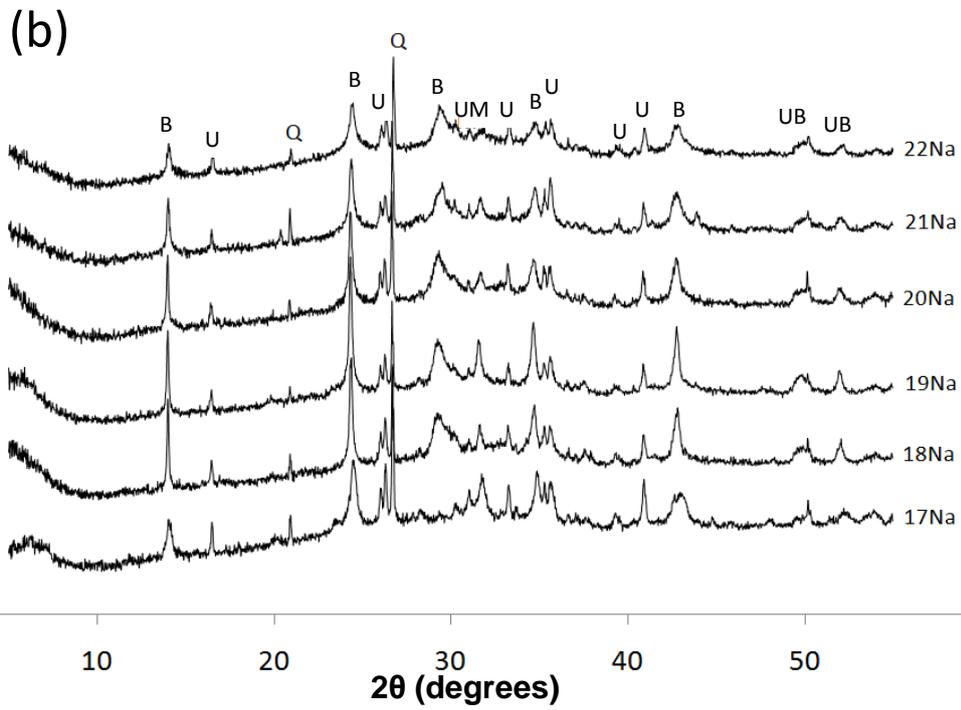
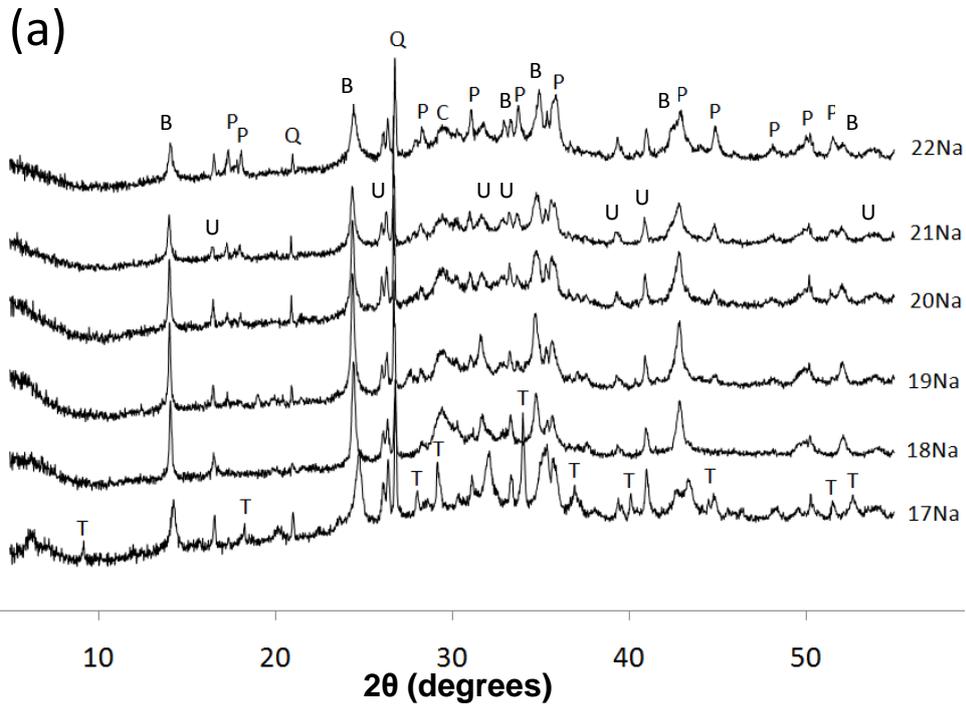
(d)

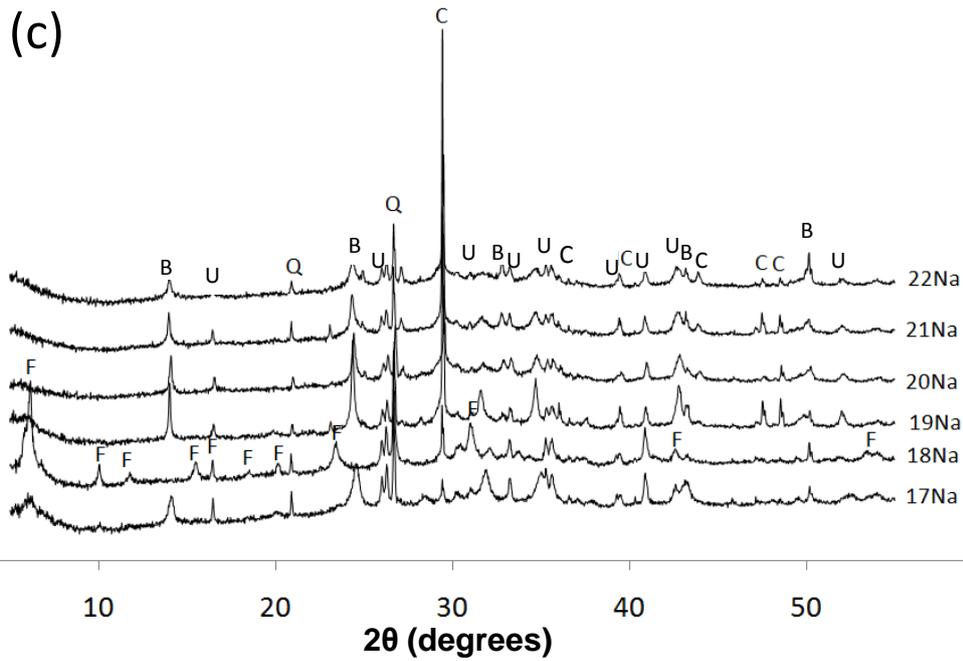


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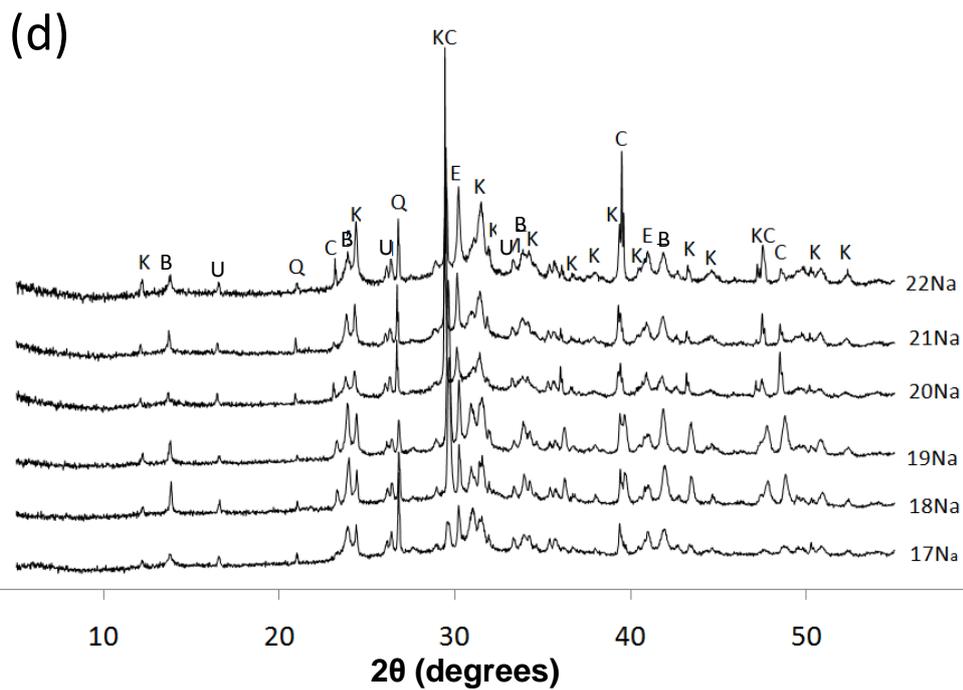
700 Figure 6. XRD traces of MK/GGBFS samples 1Na and 7Na-9Na after (a) 7 days of curing  
701 and 90 days of aging, or 7 days of curing and 90 days of exposure to solvents: (b) MEA; (c)  
702 Milli-Q H<sub>2</sub>O; (d) K<sub>2</sub>CO<sub>3</sub>. Phases marked are A: aragonite, C: calcite, F: faujasite, K:  
703 calcinite, M: muscovite, P: pirssonite, T: trona, Z: chabazite-Na.

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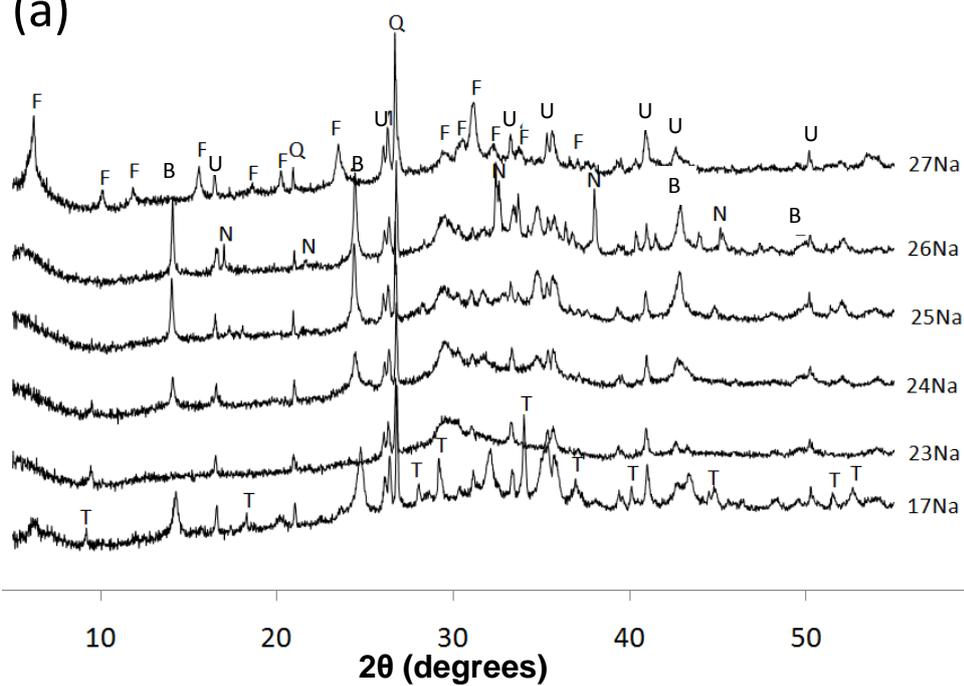


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710 Figure 7. XRD traces of FA/GGBFS samples 17Na-22Na after (a) 7 days of curing and 90  
 711 days of aging, or 7 days of curing and 90 days of exposure to solvents: (b) MEA; (c) Milli-Q  
 712 H<sub>2</sub>O; (d) K<sub>2</sub>CO<sub>3</sub>. Phases marked are B: basic sodalite, U: mullite, Q: quartz, P: pirssonite, C:  
 713 calcite, T: trona, F: faujasite, K: calcinite, E: potassium carbonate.

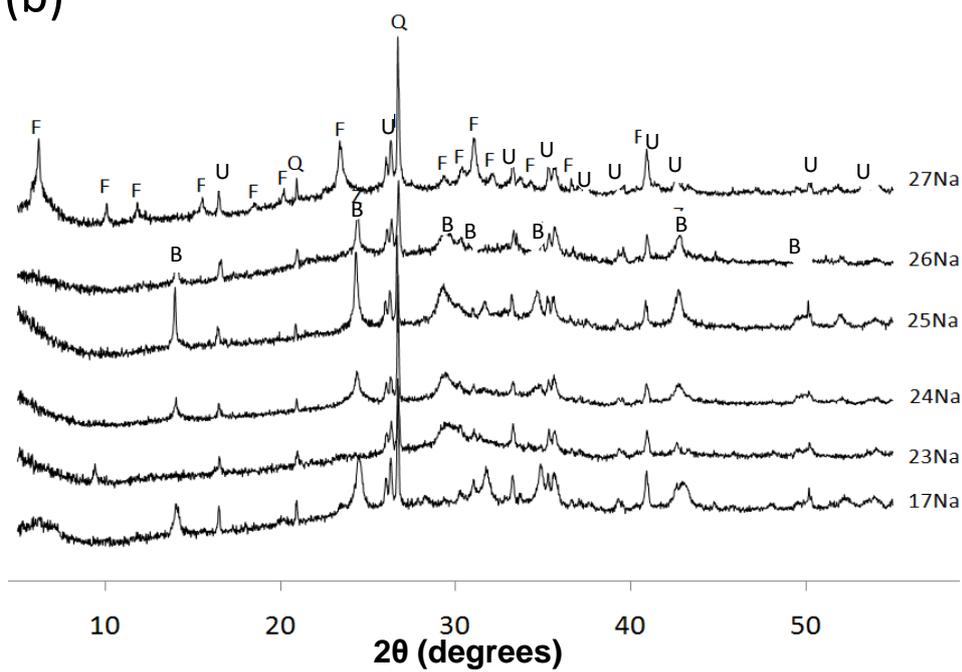
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(a)

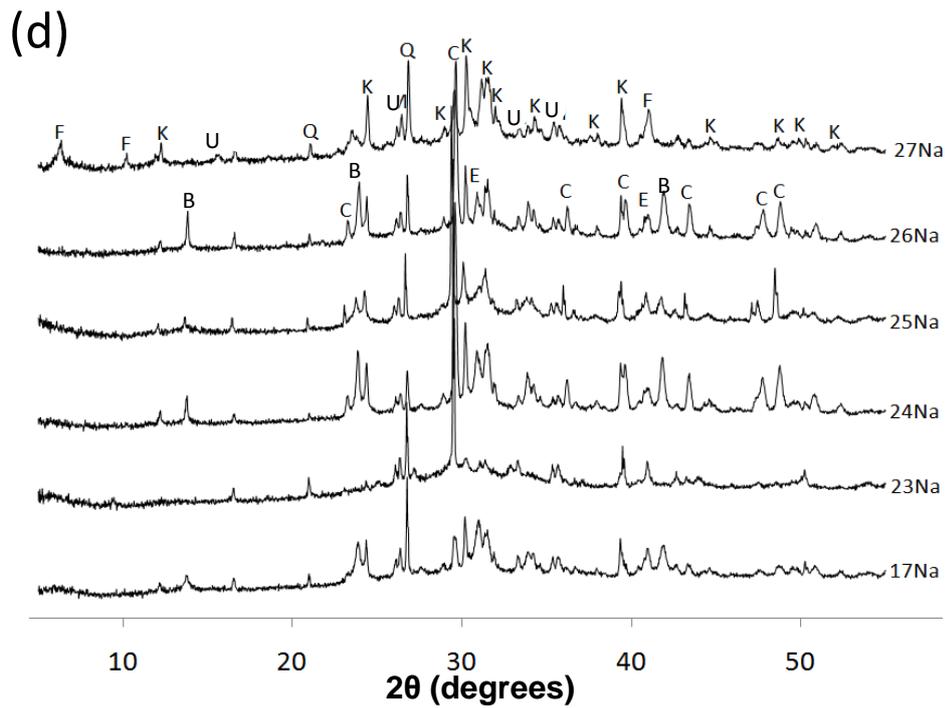
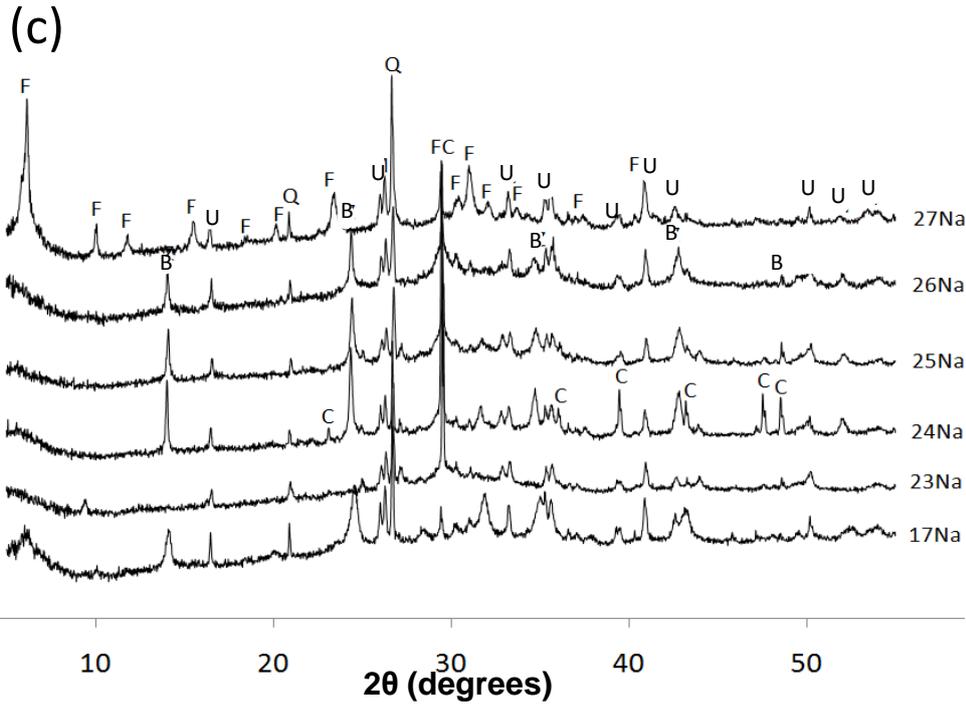


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(b)



716



719 Figure 8. XRD traces of FA/GGBFS samples 17Na and 23Na-27Na after (a) 7 days of curing  
 720 and 90 days of aging, or 7 days of curing and 90 days of exposure to solvents: (b) MEA; (c)  
 721 Milli-Q H<sub>2</sub>O; (d) K<sub>2</sub>CO<sub>3</sub>. Phases marked are B: basic sodalite, F: faujasite, Q: quartz, U:  
 722 mullite, N: thermonatrite, T: trona, C: calcite, K: kalcinite, E: potassium carbonate.

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