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1 **Interspecific competition negates CO₂ benefits for most plant species**

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Elevated CO₂ (eCO₂) frequently stimulates plant growth and is widely assumed to enhance productivity in multispecies communities, yet species responses to eCO₂ vary widely with biotic and abiotic context. Whether eCO₂ relaxes competition or amplifies existing competitive hierarchies, shaping how growth benefits are distributed among species, remains poorly resolved. We synthesize data from 19 CO₂ enrichment experiments (glasshouse, chamber, and field studies), encompassing 97 plant species from grass-dominated ecosystems to quantify how competitive context alters species- and mixture-level responses. Although most species respond positively to eCO₂ without neighbours, over half show neutral or negative responses in mixtures. Despite this, dominant species often drive modest increases in total mixture biomass under eCO₂. Trait analyses indicate that acquisitive strategies, rather than intrinsic CO₂ sensitivity alone, predict success under competition. Together, our results indicate that under competition, eCO₂ redistributes growth benefits unevenly among species, favouring dominant species while many subdominants experience neutral or negative responses.

Rising atmospheric CO₂ is altering the productivity and composition of plant communities globally. Elevated CO₂ (eCO₂) often enhances plant photosynthesis and growth¹, but the magnitude of CO₂ fertilization varies widely among species and across biotic and abiotic contexts²⁻⁶. Within plant communities, individual responses emerge from interactions with neighbours, including potential positive facilitation as well as negative resource competition. For example, 24-year biomass responses of different herbaceous functional groups to eCO₂ were amplified in mixtures and further dependent on level of N availability⁷. However, whether interspecific competition consistently modifies CO₂-fertilization in species mixtures remains unresolved.

Here, we address this gap by synthesizing species-level responses to eCO₂ across contrasting ecological neighbourhoods (i.e. no neighbours, conspecific or interspecific neighbours) and examine whether functional traits shape these responses. Variation in plant responses to eCO₂ reflects traits that influence carbon use, growth, and competitive ability. For example, C₃ species often show stronger physiological responses to eCO₂ than C₄ species, whose photosynthesis is less carbon-limited⁸, while growth form shapes how additional carbon is allocated to structure, water use, or species turnover⁹. However, consistent trait-based patterns remain elusive^{1,10}.

Evidence suggests that eCO₂ responses differ between isolated individuals and those subject to competition^{1,4,11}, and that the form of competition—whether within (monoculture) or among (mixture) species—can further shape species-level responses¹². To synthesise these patterns, we adopt a conceptual framework that separates intrinsic growth responsiveness to eCO₂ from neighbour effects on plant performance, allowing species- and mixture-level outcomes to be considered jointly. In this framework (Fig. 1), species are positioned by their eCO₂ responsiveness and vulnerability to neighbours, with mixture-level responses emerging from their relative performance and dominance. Because CO₂ is well mixed in the atmosphere, plants are not expected to compete directly for CO₂; instead, eCO₂ relaxes carbon limitation to intensify competition for other limiting resources, particularly light, water, and nutrients^{2,6,7,12}. eCO₂-driven biomass enhancement can therefore strengthen asymmetric competition for light, amplifying differences in growth and coexistence among species¹³. Because the axes in Fig. 1 are independent, realised eCO₂ effects reflect the joint effects of intrinsic responsiveness (i.e. responses when no limitations exist) and neighbour-mediated constraints, so growth stimulation in isolation may not translate to positive responses in mixtures (Fig. 1).

To examine these dynamics empirically, we extend prior syntheses of species-level CO₂ responses under competition¹ using a larger, more taxonomically diverse dataset (97 species across 19 studies) from grass-dominated CO₂ enrichment experiments, spanning functional groups (herbaceous vs. woody; C₃ vs. C₄) and neighbour contexts. We explicitly link species- and mixture-level CO₂ responses using a multilevel meta-analytic framework integrating plant trait data. Given limited comparable data from mature forests, grass-dominated ecosystems, which cover 37% of the world's vegetated land and support key ecosystem services¹⁴, provide the strongest experimental basis for evaluating CO₂ × competition effects, where physiological responsiveness and competitive interactions jointly shape functional composition¹¹ and CO₂-driven growth responses.

We first quantify how competition from neighbours modifies species-level biomass responses to eCO₂ across competitive contexts (e.g. isolated vs. monoculture or multi-species mixture). Across studies, eCO₂ significantly enhanced plant biomass when species were grown in isolation (log eCO₂ response ratio (RR) intercept = 0.10, 95% CI: [−0.00, 0.20], $p < 0.06$), with 81% of species showing positive responses (Fig. 2a; Table S8a). As expected¹, this stimulation was markedly reduced when grown with neighbours, particularly in inter-specific mixtures (log eCO₂ RR Δ = −0.14, 95% CI: [−0.23, −0.06], $p < 0.001$), with no significant difference between monoculture and isolation. When competing for resources with other species, 56% of species exhibited negative biomass responses to eCO₂; i.e. roughly half as many (44%) responded positively as seen in isolation (81%). These effects were consistent across functional types (ICC = 0.00), suggesting that competition itself, rather than niche overlap within functional types, constrains eCO₂ responsiveness. Although we cannot directly test this, constraints on eCO₂ responses likely reflect limited availability of key resources such as nitrogen, phosphorus, and

water⁶, which can suppress physiological benefits of eCO₂ and may be further shaped by mycorrhizal associations¹⁵ and herbivore regimes¹⁶.

We next examined whether species with lower vulnerability to neighbours retain stronger eCO₂ stimulation (log RCV response ratio (RR); Fig. 2B; Table S8B). On average, species showed significantly negative RCV values (log RCV RR intercept = -0.45, 95% CI: [-0.86, -0.03], $p < 0.05$), indicating that most species lost biomass under competition regardless of CO₂ treatment ($p = 0.60$) or competition type (monoculture or mixture; $p = 0.97$; Fig. 2b). These responses show that CO₂ enrichment does not inherently alleviate competitive suppression by boosting growth, contradicting earlier suggestions^{17,18}. Notably, species that exhibited strong eCO₂ stimulation also tended to be less suppressed by competition, as shown by the strong positive correlation between eCO₂ RR and RCV RR (Fig. 2c; Table S8c; isolated vs. mixture: slope = 1.08, $R^2 = 0.75$; monoculture vs. mixture: slope = 0.81, $R^2 = 0.76$; both $p < 0.001$) and by a strong negative correlation between competitive suppression (log(mono/mix)) and eCO₂ responses in mixtures (Extended Data Fig. 1, Extended Data Table 1, slope = -0.65, $R^2 = 0.53$, $p < 0.001$). Although the 19 studies ranged from controlled chambers to field experiments, precision weighting and random study effects ($\tau^2 = 0.016$) accounted for contextual differences, revealing a consistent ecological signal. This pattern highlights a subset of species (e.g., the top 5 log eCO₂ RR under mixtures: *Vachellia karroo*, *Poa pratensis*, *Lupinus perenis*, *Panicum virgatum*, *Andropogon gerardi*) with dual advantages (high eCO₂ responsiveness and low competitive vulnerability) that may disproportionately benefit from eCO₂ in competitive environments, consistent with trait-based modeling¹⁸ and species-level experimental outcomes¹¹. Although species maintain their relative positions across CO₂ treatments, eCO₂ increases the performance gap between dominant and subordinate species, strengthening existing competitive hierarchies. This pattern does not require

rank reversals to generate strong mixture-level effects. Accordingly, positive mixture-level responses may arise as dominant species combine higher intrinsic responsiveness to eCO₂ with tolerance of competitive suppression, rather than because competition is relaxed or responses are uniform.

Shifting from species-level responses to mixture-level outcomes, despite negative average responses at the species level in mixtures (Fig. 2), plot-level biomass showed modest, non-significant, increases under eCO₂ in inter-specific mixtures (intercept = 0.04, 95% CI: [-0.01, 0.10], $p = 0.13$; Fig. 3a; Table S9a). Monocultures, by contrast, showed a significant positive response ($\Delta = 0.06$, 95% CI: [0.02, 0.10], $p < 0.01$), mirroring the direction but not the magnitude of the stronger stimulation seen in isolated plants. However, the magnitude and direction of responses varied substantially, underscoring that the effect of eCO₂ is not uniform but contingent on competitive, environmental and trait contexts. Overall, eCO₂ effects on productivity were modest at the mixture scale (+8% in monocultures and +7% in mixtures, the latter not differing significantly from zero) and substantially lower than assumed in many ecosystem and Earth system models, which often assume uniform or plant function type specific productivity gains under eCO₂^{19,20}, indicating that competition dampens the CO₂ fertilization effect.

Modest mixture-level stimulation reflects dominant species maintaining performance under eCO₂ compensating for declines in subordinate species¹¹. To test this interpretation, we examined species-level eCO₂ responses as a function of their relative dominance within mixtures (Fig. 3b; Table S9b). Species were grouped into quantile-based dominance categories (rare, subdominant, and dominant) based on adjusted biomass contribution, reflecting Grime's mass ratio theory that dominant species disproportionately influence ecosystem function²¹. Dominant species had the highest responses, subdominant species the second, and rare species the lowest: with both

dominant and subdominant species showing significantly stronger eCO₂ responses than rare species ($\Delta = 0.82$, $p < 0.001$) ($\Delta = 0.54$, $p < 0.05$). Mixture-level eCO₂ biomass responses are driven by dominant species, with eCO₂ selectively disadvantaging rarer species, likely through asymmetric resource capture of light, water and nutrients. Although mixture-level responses to eCO₂ were generally positive, variability among studies (Fig. 3) likely reflects interactions among dominance structure, functional composition, resource availability, and experimental duration; all factors that modulate CO₂ fertilisation and community reorganisation.

Building on the conceptual framework in Fig. 1, which highlights the dual roles of CO₂ responsiveness and competitive ability in shaping mixture outcomes, a key remaining question is which traits enable species to succeed in mixtures under eCO₂. To explore this, we grouped species into four “success” categories based on their combined eCO₂ responsiveness and competitive vulnerability: “Highly Successful”, “CO₂ Responsive but Vulnerable”, “Stable but Unresponsive”, and “Less successful” (see Fig. 1), with most species falling into the two intermediate categories. These four groups differed significantly in multivariate trait space (Fig. 4a; Table S10a; Pillai’s trace = 6.00, $p < 0.0001$), with “Highly Successful” species strongly separated along the first linear discriminant axis (LD1), largely due to acquisitive traits such as rapid growth rate, greater height, and lower specific stem density, with some showing reduced drought tolerance. We then tested which individual traits predicted success components (Fig. 4b; Table S10b). No individual trait significantly predicted eCO₂ responsiveness, whereas competitive vulnerability was consistently associated with traits linked to competitive performance under eCO₂, even after accounting for multiple testing (Table S10B). In isolated versus mixture settings, higher growth rate and lower drought tolerance were weakly associated with reduced vulnerability to competition (growth rate: $p < 0.02$, $p_{\text{adj}} = 0.10$; drought tolerance: $p < 0.05$, $p_{\text{adj}} = 0.10$). In monoculture versus mixture

comparisons, reduced vulnerability was most strongly associated with lower stem density ($p < 0.01$, $p_{\text{adj}} = 0.01$) and greater plant height ($p < 0.05$, $p_{\text{adj}} = 0.04$), with invasive status showing a similar association ($p < 0.05$, $p_{\text{adj}} = 0.04$). Leaf area and leaf carbon content showed weaker, non-significant trends in the same direction. These results suggest that acquisitive, and in some cases invasive, species are more likely to maintain high performance under $e\text{CO}_2$ and competition, indicating trait-based consistency among successful species rather than strict trait-based filtering. Such trait syndromes define the ‘Highly Successful’ group. Importantly, the association with invasive status likely reflects the relevance of traits linked to invasiveness under conditions of conspecific crowding, rather than a general effect of invasiveness on CO_2 responsiveness.

Our findings update, expand upon, and extend those of Poorter and Navas¹, who emphasized the role of acquisitive strategies in competition but found no consistent link between competitiveness and $e\text{CO}_2$ responsiveness. Here, we show that species with traits promoting rapid growth and resource capture often gain a dual advantage; reduced vulnerability to the presence of neighbours, which tends to have knock on positive effects on responses to $e\text{CO}_2$. This highlights how modern multilevel synthesis can uncover cross-scale patterns that were not detectable previously. By disproportionately increasing the abundance of dominant or invasive species, rising CO_2 may widen performance gaps within grass-dominated systems and promote biotic homogenization. Importantly, this amplification of dominance does not require changes in species rank order, but instead reflects strengthened competitive asymmetries that disproportionately suppress subdominant species under $e\text{CO}_2$. Since our synthesis focuses on grass-dominated systems, many functional groups and biomes remain underrepresented. Nevertheless, the consistent directional responses observed across diverse grass-dominated glasshouse, chamber, and field studies indicate that the patterns we report for the grassland biome are robust despite

contextual variation, a conclusion that is qualitatively supported by a small set of long-term, intact multispecies field experiments reporting changes in relative species or functional group abundance under eCO₂ (Table S11). More broadly, a lack of long-term, multispecies experiments explicitly designed to quantify competitive interactions under eCO₂ remains a major knowledge gap, limiting direct inference about how CO₂-driven changes in competition scale up to biodiversity change in natural ecosystems.

While slower-growing species may respond positively to eCO₂ under low-competition or over longer timescales, our results emphasize that competitive context and trait variation can jointly shape eCO₂ outcomes. Importantly, the responses synthesized here are necessarily constrained to experimental timescales, and competitive hierarchies may continue to reorganize over decades as species turnover, soil feedbacks, and disturbance regimes unfold, potentially modifying or even reversing short-term patterns. Incorporating these filters into models and predictions, where species-level or trait-based dynamics are of interest, will be critical for capturing the context-dependent nature of vegetation change under future climate scenarios. This shift in interpretation is consequential: mixture-level eCO₂ responses may be weak or neutral as eCO₂ amplifies dominance by a subset of species, with implications for biodiversity loss, invasion dynamics, and trait filtering under global change.

Methods

Data selection and extraction

A systematic literature search was conducted in Web of Science and Google Scholar (March 2023) using the search string: ('elevated CO₂' OR 'high CO₂' OR 'rising CO₂') AND ('competition' OR 'mixture') AND ('grassland' OR 'savanna' OR 'woodland' OR 'shrubland') AND ('experiment'

OR 'trial' OR 'FACE' OR 'free air CO₂ enrichment' OR 'OTC' OR 'open-top chamber' OR
'glasshouse' OR 'growth chamber' OR 'convirion'). This yielded 183 studies from Web of Science
and 146 from Google Scholar. Titles, abstracts, and full texts were manually screened to identify
studies reporting aboveground biomass for at least one competitive and one non-competitive
treatment under both ambient and eCO₂ conditions. Only studies on wild species were retained;
those focused on crops or involving fully mature woody plants were excluded. For studies that
included additional treatments (e.g. water, nutrient, or light manipulations), analyses were
restricted to treatment levels that allowed assessment of CO₂ and competition effects only. All
selected studies provided species-level aboveground biomass data (standardized to g·m⁻² when
necessary). Ambient CO₂ (aCO₂) treatments ranged from 320–400 ppm, and eCO₂ treatments from
500–1000 ppm; studies using eCO₂ levels above 1000 ppm were excluded based on projected 2080
atmospheric scenarios. In total, 19 peer-reviewed studies were included, along with one additional
unpublished dataset known to the authors (full list in Table S1). The final dataset represents
experiments from five continents (see Fig. S1). Included studies comprised glasshouse and growth-
chamber experiments, open-top chamber studies, and FACE field experiments, and incorporated
both monocultures and experimentally imposed or naturally assembled species mixtures under
competitive and non-competitive conditions.

Species-level aboveground biomass data (as a measure of plant performance) were
extracted from text, tables, or figures using WebPlotDigitizer²². When biomass data were not
directly reported but inferred to exist based on the methods, corresponding authors were contacted
to request data. One dataset was obtained directly from the authors of this manuscript (see Table
S2). For each species within each study, mean and standard error of biomass were recorded for
each CO₂ and competition treatment. Species were assigned to one of three functional types based

on photosynthetic pathway and growth form: C₃ woody, C₃ herbaceous, or C₄ herbaceous. Competition context was classified into four categories: (1) Isolated (plants grown individually), (2) Monoculture (intraspecific competition), (3) Mixtures (interspecific competition). Pairwise experiments (those including two competing species) were classified within the “mixture” category to maintain consistency with broader multispecies comparisons. All biomass data were converted to standardized units of g m⁻² to ensure comparability across studies. Mixture biomass was defined as the summed aboveground biomass of all species within a given plot or treatment, and comparisons were conducted separately for isolation versus inter-specific mixture and monoculture versus mixture designs to account for differences in experimental structure. Additional metadata were extracted to account for sources of heterogeneity in plant responses: experimental method, duration, CO₂ treatment difference, and soil volume. In total, biomass data from 97 species were compiled, including 25 young C₃ woody species, 49 C₃ herbaceous species, and 23 C₄ herbaceous species.

Effect size calculations

Species- and mixture-level responses to eCO₂ were calculated using log-transformed response ratios in R version 4.5.0²³. For each species in each study and competition treatment, the eCO₂ response ratio (eCO₂ RR; Eq. 1) was calculated as the natural logarithm of the ratio of mean biomass under eCO₂ versus ambient CO₂ (aCO₂). To assess the effects of competitive context, relative competition vulnerability (RCV) was quantified as the log ratio of biomass in mixture versus either isolation or monoculture, depending on the study design (Eq. 2 and Eq. 3). These were computed separately under each CO₂ treatment to assess interaction effects. We interpret the log relative competitive vulnerability (RCV RR) as an operational measure of neighbour effect.

242 Less-negative (or more positive) RCV indicates higher competition tolerance (i.e., smaller
 243 proportional biomass loss in mixtures relative to isolation/monoculture), whereas more-negative
 244 RCV indicates higher competition vulnerability. As we excluded experimental categories that
 245 manipulated other resources (e.g., nutrient or water addition, imposed shade), the competition
 246 captured here reflects ambient resource limitation, most commonly light and soil nutrients.
 247 Mixture-level responses were assessed by summing aboveground biomass across all species within
 248 a plot and calculating the log-transformed response ratio between eCO₂ and aCO₂ treatments (Eq.
 249 4). Variance for each log response ratio was estimated using the delta method (Eq. 5). Within
 250 mixture treatments, species dominance was quantified as the ratio of each species' biomass
 251 contribution to the total mixture biomass. Dominance ratio = biomass of focal species in mixture
 252 / total biomass of mixture. Species were binned into three quantile-based dominance groups (rare,
 253 subdominant, dominant) based on dominance ratio adjusted by species richness. Values are
 254 bounded in (0,1).

255 **Eq. 1** $eCO_2 RR = \log\left(\frac{biomass_{eCO_2}}{biomass_{aCO_2}}\right)$

256 **Eq. 2** $RCV RR = \log\left(\frac{biomass_{mixture}}{biomass_{isolated}}\right)$

257 **Eq. 3** $RCV RR = \log\left(\frac{biomass_{mixture}}{biomass_{monoculture}}\right)$

258 **Eq. 4** $Mixture eCO_2 RR = \log\left(\frac{community\ biomass_{eCO_2}}{community\ biomass_{aCO_2}}\right)$

259 **Eq. 5** $Var RR = \left(\frac{SE_{eCO_2}}{Biomass_{eCO_2}}\right) + \left(\frac{SE_{aCO_2}}{Biomass_{aCO_2}}\right) \text{ or } \left(\frac{SE_{mixture}}{Biomass_{mixture}}\right) + \left(\frac{SE_{iso\ or\ mono}}{Biomass_{iso\ or\ mono}}\right)$

260 Effect sizes were synthesized using multilevel meta-analysis models (rma.mv in the
 261 *metafor* R package²⁴), with random intercepts for study and functional type. Species- and mixture-

level eCO₂ responses were analyzed with competition treatment as a fixed effect (levels: Isolated, Monoculture, Mixture). RCV responses were analyzed using factorial models with CO₂ treatment and competition context (isolated vs. mixture; monoculture vs. mixture) as moderators. Estimates were weighted by the inverse of their sampling variance. Models were fit using restricted maximum likelihood (REML), and precision-weighted mean responses with 95% confidence intervals were extracted for each group. Individual effect sizes and fitted model estimates were visualized using forest-style plots. This hierarchical structure explicitly accounts for between-study and between-functional-type heterogeneity, ensuring that mean effects reflect consistent directional trends rather than artefacts of averaging across diverse experimental designs²⁵. Variance components reported in Supplementary Table S8 show that most of the modest heterogeneity occurred among studies ($\tau^2 = 0.0156$), with negligible variance among functional-type groupings ($\tau^2 \approx 0.0000$). Thus, between-study differences were present but small, indicating that observed patterns are not driven by methodological inconsistency.

To assess potential publication bias and small-study effects, we followed the recommendations of Nakagawa et al.²⁵. We visually inspected funnel plots of log response ratios (y_i) against study precision ($1/SE$), with pseudo 95% confidence intervals centered on the model intercept (Figure S2). An Egger-style multilevel meta-regression tested for small-study effects by including the square root of the sampling variance ($\sqrt{v_i}$) as a moderator. To evaluate robustness, we refitted the model after excluding low-precision studies ($SE > 1$); results were consistent with the original model (Table S2). We also conducted a leave-one-out influence analysis to assess the impact of individual studies. In all cases, changes in the intercept were minimal (maximum $\Delta = 0.075$), and the CO₂ effect remained consistently positive (Table S3, Figure S3), confirming that no single study disproportionately influenced the meta-analytic outcome. Finally, we performed a

trim-and-fill sensitivity analysis by excluding studies with extreme effect sizes (± 2.5 SD from the mean). The resulting model (Table S4) showed slightly attenuated effect sizes but preserved the direction and significance of the mixture competition effect, supporting the robustness of the overall findings. We applied the same procedures to assess potential bias and robustness in the RCV meta-analysis. Funnel plots (Figure S4), Egger-style meta-regression, and exclusion of low-precision studies (Table S5) indicated no evidence of small-study effects. A leave-one-out analysis confirmed that no single study disproportionately influenced the RCV estimates (Figure S6, Table S6), and trimming outliers (± 2.5 SD) had minimal effect on model outcomes (Table S7).

Trait data extraction and analysis

To determine which plant traits predict CO₂ responsiveness and competitive vulnerability, we compiled species-level trait data for all species in the meta-analysis. Continuous and categorical traits were extracted from the TRY²⁶ and AusTraits²⁷ databases, and supplemented with trait values from the original study manuscripts. The final dataset included the following traits: leaf area, leaf nitrogen content (Leaf N), leaf dry matter content (LDMC), specific stem density (SSD), specific leaf area (SLA), leaf carbon content (Leaf C), leaf C:N ratio (Leaf C:N), root:shoot ratio, maximum plant height, life strategy (annual, perennial, and deciduous), growth rate, lifespan, Grime strategy, resprouting ability, shade, fire, cold and drought tolerances (low, moderate, and high), and invasive status (manual assessment of reports of invasion occurrence using the CABI Invasive Species Compendium²⁸). To assess potential multicollinearity among trait predictors, we calculated variance inflation factors (VIFs) for each model. All VIF values were below 3.3, indicating that collinearity was minimal and did not affect model inference.

Trait values were log-transformed (with offsets to avoid zeroes) and imputed using PCA-based imputation (`missMDA` package in R²⁹) to complete missing values across the dataset. All

traits were standardized prior to analysis. Traits were then merged with species-level CO₂ response and competitive vulnerability metrics, and each species was classified into one of four success categories based on its position relative to median CO₂ response and RCV values: “Highly Successful,” “eCO₂ Responsive but Vulnerable to Competition,” “Stable but Unresponsive to eCO₂,” and “Less successful”. Discriminant Analysis of Principal Components (DAPC) was used to examine whether success categories occupy distinct multivariate trait spaces (Fig. 4A). Trait contributions to discriminant axes were used to identify key axes of variation, and significance of trait space separation was assessed via MANOVA and pairwise comparisons.

To quantify the contribution of each trait to CO₂ responsiveness and competitive vulnerability, linear mixed-effects models (LMMs; `missMDA` package in R³⁰) were fitted with standardized trait predictors and random intercepts for study and functional type. Separate models were fitted for log-transformed eCO₂ RR, RCV RR from mixture versus isolation, and RCV RR from mixture versus monoculture. Standardized model coefficients and 95% confidence intervals were extracted and summarized in forest plots. All trait analyses were performed in R version 4.5.0²³.

Data availability

The data supporting the results of this study are available on DRYAD (<https://doi.org/10.5061/dryad.8sf7m0d2q>). This includes the raw biomass and trait data.

Code availability

The code supporting the results of this study are available on DRYAD (<https://doi.org/10.5061/dryad.8sf7m0d2q>). This includes the R scripts used for statistical analysis and trait imputation. The code is fully documented to ensure reproducibility of the analyses.

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Author contributions

SLR developed the idea, approach, and design for this study, with assistance from PBR. SLR conducted data analyses and wrote the initial draft of the manuscript. SLR, KJS, BSR, CPO, and PBR all contributed unpublished data. All authors contributed to revisions of the manuscript. We thank Dr. Thiago Gonçalves Souza for his advice on the manuscript.

Competing interests statement

There are no competing interests.

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Figure legends

Figure 1. Conceptual framework illustrating species- and mixture-level responses to elevated CO₂ (eCO₂) in competitive plant mixtures. (a) At the species level, success under eCO₂ depends on both a species' intrinsic responsiveness to CO₂ (i.e., responses in the absence of neighbors) and its ability to tolerate neighbours. Here, tolerance refers to a species' ability to maintain biomass with neighbours, while vulnerability is the proportional biomass loss when grown with neighbours. As studies/data altering other resources were excluded, "competition from neighbours" primarily reflects ambient growth limitation by light, nutrients and water. Species can be categorized into four quadrants: (1) *Less successful* (low eCO₂ responsiveness and low neighbour tolerance), (2) *CO₂ responsive but vulnerable* (high eCO₂ responsiveness but low neighbour tolerance), (3) *CO₂ unresponsive but tolerant* (low eCO₂ responsiveness but high neighbour tolerance), and (4) *Highly successful* (both high eCO₂ responsiveness and neighbour tolerance). The example species used in Fig 1A is a grass, but this could be any growth form. (b) At the mixture level, the overall biomass response to eCO₂ increases with the proportion of species classified as individually successful. However, this relationship is shaped by species' dominance and identity, meaning that mixture-level gains may occur even when many species respond negatively to eCO₂.

Figure 2. Effects of competition context on plant responses to elevated CO₂ (eCO₂) relative to ambient CO₂ (aCO₂) across functional types and experimental designs. (a) Mean log response ratios to eCO₂ (eCO₂ RR; biomass at eCO₂ / biomass at aCO₂) for plant biomass under three competition treatments: isolated (n = 41), monoculture (n = 47), and mixture (n = 96). Coloured points represent individual species-level responses, with point colour indicating functional type (young C₃ woody, C₃ herbaceous, C₄ herbaceous) and point size scaled by precision (1/SE). Black points and horizontal error bars show model-predicted group means and 95% confidence intervals from a multivariate meta-analysis. Model estimates: intercept = 0.10 (95% CI: -0.00 to 0.20, p = 0.06); mixture effect relative to isolation, $\Delta = -0.14$ (95% CI: -0.23 to -0.06, p < 0.001). (b) Mean log relative competitive vulnerability response ratios (RCV RR; biomass in mixture / biomass in isolated or monoculture), separated by CO₂ treatment: ambient (circles) and elevated (triangles). Coloured points show species-level responses with colours indicating functional type; black points and error bars indicate group means and 95% confidence intervals from a multivariate meta-analysis. High CO₂: diversity effect (n = 58), competition effect (n = 43); low CO₂: diversity effect (n = 57), competition effect (n =

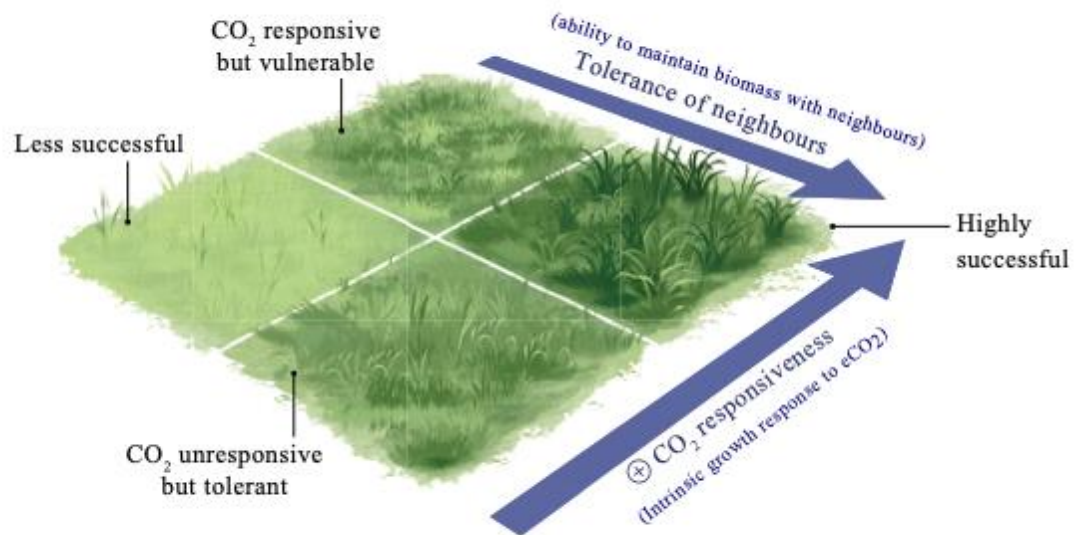
42). Model estimates: intercept = -0.45 (95% CI: -0.86 to -0.03 , $p < 0.05$); CO_2 treatment effect, $p = 0.60$; competition-type effect, $p = 0.97$. **(c)** Correlation between species' eCO_2 RR (y-axis) and RCV RR (x-axis) ($n = 98$). Solid and dashed lines represent fitted regression models for different competition comparisons (isolated vs. mixture: solid; monoculture vs. mixture: dashed). Shaded regions represent 95% confidence intervals around fitted linear regression lines. Point colours indicate functional type; regression equations and R^2 values are provided in the figure. Model slopes were 1.08 ($R^2 = 0.75$, $p < 0.001$) for isolated-versus-mixture comparisons and 0.81 ($R^2 = 0.76$, $p < 0.001$) for monoculture-versus-mixture comparisons.

Figure 3. Mixture level responses to elevated CO_2 (eCO_2) across competition context. **(a)** Meta-analysis of mixture-level eCO_2 response ratios (log-transformed) comparing monocultures ($n = 51$) and mixed-species ($n = 35$) competition treatments. Black points and error bars represent model-predicted means and 95% confidence intervals from a multilevel meta-analysis including random intercepts by study. Grey points show individual plot responses from each study, with size reflecting the inverse standard error (precision) of individual study estimates. The dashed line at $x = 0$ indicates no response to eCO_2 . Model estimates: mixture intercept = 0.04 (95% CI: -0.01 to 0.10 , $p = 0.13$); monoculture effect relative to mixtures, $\Delta = 0.06$ (95% CI: 0.02 to 0.10 , $p < 0.01$). **(b)** Species-level responses to eCO_2 grouped by relative dominance within mixtures. Species were binned into three quantile-based dominance groups [rare ($n = 31$), subdominant ($n = 30$), dominant ($n = 30$)] based on dominance ratio adjusted by species richness. Points show individual species-level log response ratios, while large points and vertical error bars denote group means $\pm$ 95% confidence intervals. Sample sizes for each bin are indicated above. The dashed line at $y = 0$ indicates no response to eCO_2 . Dominant and subdominant species exhibited stronger eCO_2 responses than rare species (dominant vs. rare: $\Delta = 0.82$, $p < 0.001$; subdominant vs. rare: $\Delta = 0.54$, $p < 0.05$).

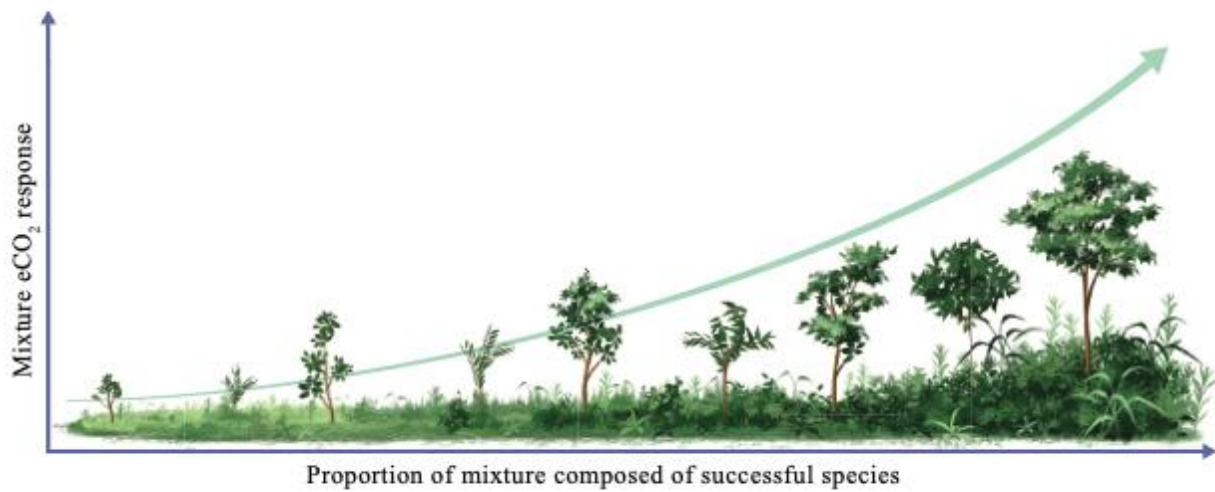
Figure 4. Trait-based separation of plant success groups under elevated CO_2 and competition. **(a)** Discriminant Analysis of Principal Components (DAPC) shows separation among species classified into four success groups based on their CO_2 responsiveness and competitive vulnerability. Groups are plotted by the first two linear discriminants (LD1 and LD2), with ellipses representing 68% confidence intervals. Arrows indicate trait loadings, scaled by

correlation strength with the DAPC axes. Success groups differed significantly in multivariate trait space (Pillai's trace = 6.00, $p < 0.0001$). **(b)** Standardized coefficients ($\pm 95\%$ CI) from linear mixed-effects models showing how individual traits predict species' CO₂ response (left) and competitive vulnerability (right). Significant predictors ($P < 0.05$) are shown in red. Trait effects are shown for eCO₂ response ratios ($n = 99$), relative competitive vulnerability response ratios for isolated versus mixture comparisons ($n = 88$), and relative competitive vulnerability response ratios for monoculture versus mixture comparisons ($n = 110$). Error bars indicate 95% confidence intervals around standardized coefficient estimates. No individual trait significantly predicted eCO₂ responsiveness. In isolated-versus-mixture comparisons, higher growth rate ($p < 0.02$, $p\text{-adj} = 0.10$) and lower drought tolerance ($p < 0.05$, $p\text{-adj} = 0.10$) were associated with reduced competitive vulnerability. In monoculture-versus-mixture comparisons, reduced vulnerability was associated with lower stem density ($p < 0.01$, $p\text{-adj} = 0.01$), greater plant height ($p < 0.05$, $p\text{-adj} = 0.04$), and invasive status ($p < 0.05$, $p\text{-adj} = 0.04$). **Trait abbreviations:** Cold tol. – cold tolerance (ordinal: 1 = low to 10 = high); Drought tol. – drought tolerance (ordinal: 1 = low to 10 = high); Fire tol. – fire tolerance (ordinal: 1 = low to 10 = high); Grime strat. – Grime's strategy classification (e.g., C, S, R types); Growth form – life strategy (annual, perennial, deciduous); Growth rate – growth speed (fast- or slow-growing); Height – maximum vegetative height (cm); Invasive – invasive status (classified as invasive: yes/no); LA – leaf area (mm²); Leaf C – leaf carbon content (%); Leaf C:N – leaf carbon-to-nitrogen ratio (unitless); Leaf N – leaf nitrogen content (%); LDMC – leaf dry matter content (g/g); Lifespan – estimated lifespan (years); LMA – leaf mass per area (g/m²); Resprouting – resprouting ability (yes/no); Root:Shoot – root-to-shoot biomass ratio (unitless); Shade tol. – shade tolerance (ordinal: 1 = low to 10 = high); SLA – specific leaf area (mm²/mg); SSD – stem specific density (g/cm³).

a. Species-level Success



b. Mixture-level Success



494
495 Figure 1.

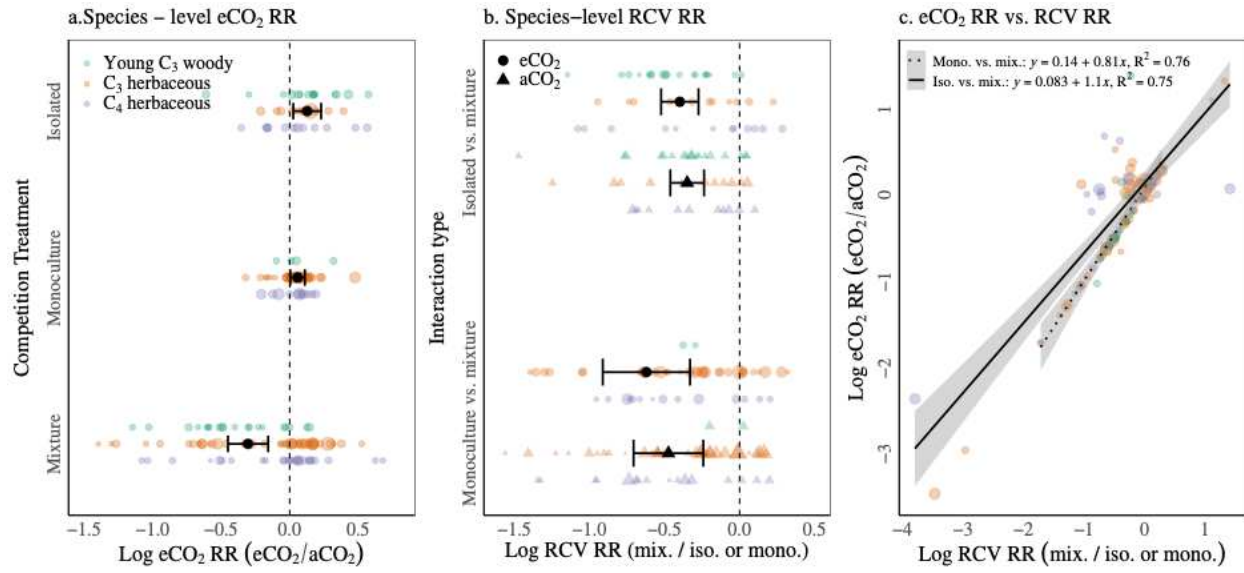
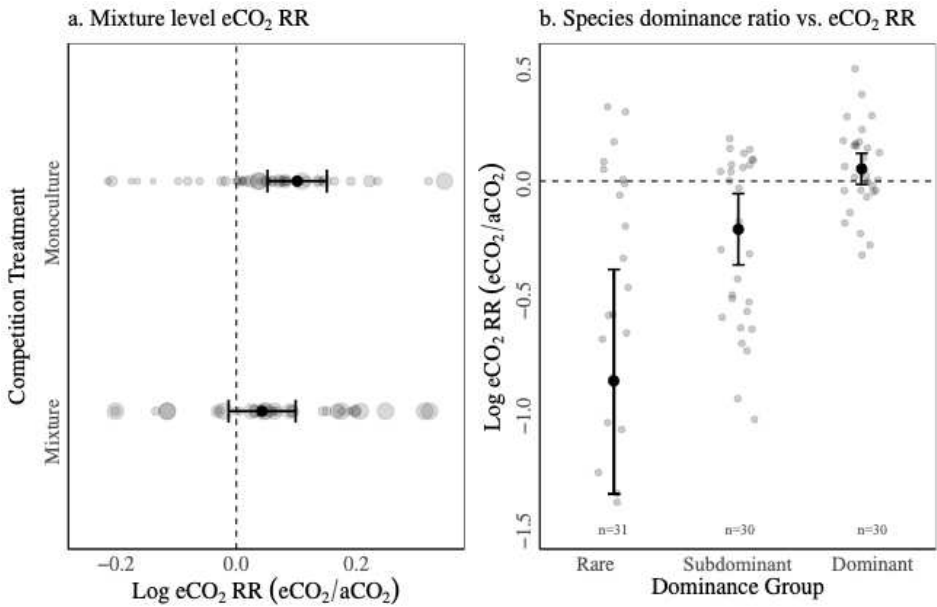


Figure 2.

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502 Figure 3.

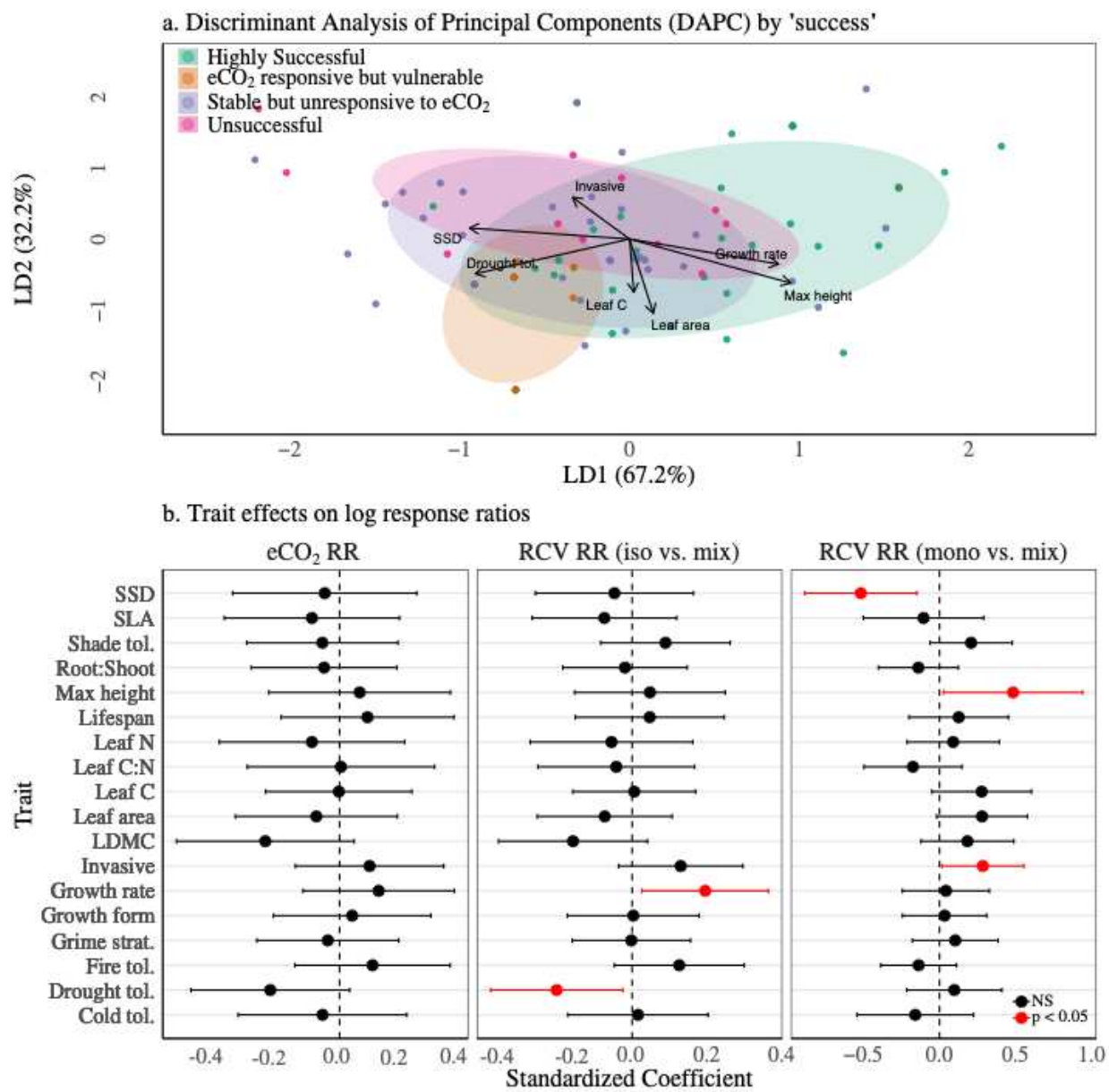
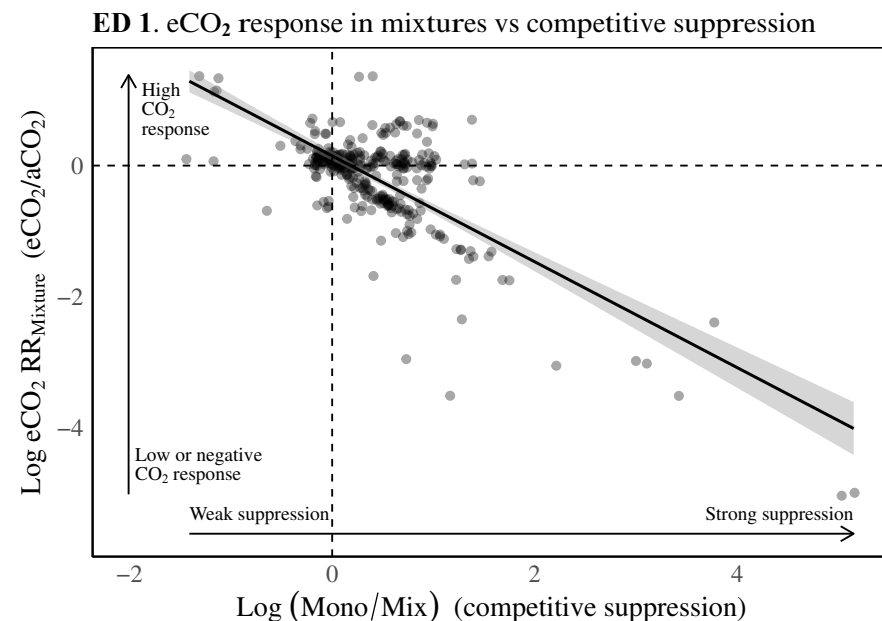


Figure 4.

Extended Data



Extended Data Figure 1. CO₂ responsiveness in mixtures as a function of competitive suppression. Species that experience stronger biomass suppression in mixtures relative to monocultures (higher log mono/mix values) show weaker or negative biomass responses to elevated CO₂ (log eCO₂ RR). Competitive suppression was negatively associated with eCO₂ responses in mixtures (slope = -0.65 , $R^2 = 0.53$, $p < 0.001$). Species that are less suppressed by neighbours retain more positive eCO₂ responses. This pattern indicates that competitive context strongly constrains the expression of intrinsic CO₂ responsiveness.

Extended Data Table 1. Linear model results testing the correlation between species' CO₂ responsiveness in mixtures and competitive suppression. The model relates species' log-transformed biomass response ratios to elevated CO₂ in mixtures (eCO₂ RR_{mix}) to their competitive suppression (log(mono/mix)). The table reports the intercept and slope for eCO₂ RR_{mix}, along with associated standard errors, 95% confidence intervals, and p-values. Model fit is summarized using R-squared (R^2).

ED Table 1. CO ₂ responsiveness in mixtures vs. competitive suppression					
	Term	Estimate	SE	X95..CI	p.value
Mixture correlation model	(Intercept)	0.304	0.028	[0.249, 0.359]	< 0.001
	X	-0.652	0.035	[-0.720, -0.583]	< 0.001
Model Fit	R-squared	0.526			

1 **Supplementary materials**

S1. Studies included in meta-analysis



2

3 **Fig. S1.** Geographic distribution of studies included in the meta-analysis. Points indicate study locations,
4 with symbol size scaled by the number of species assessed in each study. The dataset includes 19 studies
5 spanning temperate, tropical, and Mediterranean regions across six continents. Coordinates and species
6 richness were extracted from original publications or supplementary materials.

7 **Table S1.** Summary of studies used in the meta-analysis

Table S1.						
Data source	Location	Experimental conditions	CO ₂ treatment levels	Duration of experiment (days)	Species	Functional type
Bloor <i>et al.</i> , 2009 ¹	Orsay, France	Glasshouse growth chambers (CO ₂ -enriched); experimentally sown, potted tree seedlings grown alone or with grass competitors	380, 645	105	<i>Dactylis glomerata</i>	C ₃ Herbaceous
				75	<i>Fraxinus excelsior</i>	C ₃ Woody
Collins <i>et al.</i> , 2018 ²	Sydney, Australia	Field (FACE); naturally assembled temperate eucalypt woodland with transplanted woody seedlings grown with or without herbaceous neighbours	400, 550	1825	<i>Eucalyptus tereticornis</i>	C ₃ Woody
					<i>Hakea sericea</i>	C ₃ Woody
Dijkstra <i>et al.</i> , 2010 ³	Central Plains Experimental Range, Colorado	Glasshouse (CO ₂ -enriched); experimentally sown, potted monocultures and five-species mixtures	400, 780	85	<i>Artemisia frigida</i>	C ₃ Herbaceous
					<i>Bouteloua gracilis</i>	C ₄ Herbaceous
					<i>Hesperostipa comata</i>	C ₃ Herbaceous
					<i>Linaria dalmatica</i>	C ₃ Herbaceous
					<i>Pascopyrum smithii</i>	C ₃ Herbaceous

Hager <i>et al.</i> , 2020 ⁴	Guelph, Canada	Glasshouse CO ₂ -controlled chambers; experimentally sown, potted grass monocultures and native–invasive species mixtures	390, 700	98	<i>Andropogon gerardii</i>	C ₄ Herbaceous
					<i>Elymus virginicus</i>	C ₃ Herbaceous
					<i>Koeleria macrantha</i>	C ₃ Herbaceous
					<i>Panicum virgatum</i> cv. <i>Cave-in-Rock</i>	C ₄ Herbaceous
Goncalves de Oliveira <i>et al.</i> 2021 ⁵	Federal University of Viçosa, Campus Florestal, Brazil	Open-top chambers; experimentally sown; isolated or 2 species mixture	350, 100	270	<i>Melinis minutiflora</i>	C ₄ Herbaceous
					<i>Stylosanthes capitata</i>	C ₃ Herbaceous
Jongen & Jones, 1997 ⁶	Dublin, Ireland	Open-top chambers; experimentally sown, potted grass monocultures (density manipulation) and four-species mixtures	354, 700	80	<i>Cynosurus cristatus</i>	C ₃ Herbaceous
					<i>Holcus lanatus</i>	C ₃ Herbaceous
					<i>Lolium perenne</i>	C ₃ Herbaceous
Lau <i>et al.</i> , 2010 ⁷	Cedar Creek, Minnesota, USA	Field (FACE); potted plants grown with experimentally imposed competition (none, intraspecific, C ₃ or C ₄ grass competitors)	368, 560	35	<i>Arabidopsis thaliana</i>	C ₃ Herbaceous
Manea & Leishman, 2011 ⁸	Sydney, Australia	Glasshouse (CO ₂ -enriched); experimentally sown, potted native–invasive species pairs grown singly or with heterospecific neighbours. Additional data that is not in Manea and Leishman (2011) was provided by the authors on request.	380, 675	84	<i>Acacia implexa</i>	C ₃ Woody
					<i>Acacia parramattensis</i>	C ₃ Woody
					<i>Acetosa sagittata</i>	C ₃ Herbaceous
					<i>Araujia sericifera</i>	C ₃ Herbaceous
					<i>Atriplex semibaccata</i>	C ₃ Herbaceous
					<i>Bidens pilosa</i>	C ₃ Herbaceous
					<i>Bothriochloa macra</i>	C ₄ Herbaceous
					<i>Chloris gayana</i>	C ₄ Herbaceous
					<i>Chloris truncata</i>	C ₄ Herbaceous
					<i>Convolvulus erubescens</i>	C ₃ Herbaceous
					<i>Conyza bonariensis</i>	C ₃ Herbaceous
					<i>Dichanthium sericeum</i>	C ₄ Herbaceous
					<i>Digitaria sanguinalis</i>	C ₄ Herbaceous
					<i>Einadia hastata</i>	C ₃ Herbaceous
					<i>Eragostis brownii</i>	C ₄ Herbaceous
					<i>Eragrostis curvula</i>	C ₄ Herbaceous
					<i>Glycine microphylla</i>	C ₃ Herbaceous
					<i>Ligustrum sinense</i>	C ₃ Woody
					<i>Olea europaea</i> subsp. <i>cuspidata</i>	C ₃ Woody
					<i>Paspalum dilatatum</i>	C ₄ Herbaceous
					<i>Phytolacca octandra</i>	C ₃ Herbaceous
					<i>Plectranthus parviflorus</i>	C ₃ Herbaceous
					<i>Sida rhombifolia</i>	C ₃ Herbaceous
					<i>Sigesbeckia orientalis</i>	C ₃ Herbaceous
					<i>Sorghum halepense</i>	C ₄ Herbaceous

					<i>Sporobolus creber</i>	C ₄ Herbaceous
					<i>Sporobolus indicus</i>	C ₄ Herbaceous
					<i>Themeda australis</i>	C ₄ Herbaceous
Melo <i>et al.</i> , 2018 ⁹	Viçosa, Brazil	Field (open-top chambers); experimentally established tree seedlings grown in soil with density-manipulated herbaceous neighbours	380, 700	320	<i>Hymenaea stigonocarpa</i>	C ₃ Woody
					<i>Melinis minutiflora</i>	C ₄ Herbaceous
Miranda-Apodaca <i>et al.</i> , 2015 ¹⁰	Leioa, Spain	Growth chamber (CO ₂ -controlled); experimentally sown, potted monocultures and interspecific mixtures	370, 740	38	<i>Agrostis capillaris</i>	C ₃ Herbaceous
					<i>Festuca rubra</i>	C ₃ Herbaceous
					<i>Trifolium pratense</i>	C ₃ Herbaceous
					<i>Trifolium repens</i>	C ₃ Herbaceous
Moore & Field, 2006 ¹¹	Stanford, California, USA	Field (open-top chambers); experimentally sown grassland microcosms grown as monocultures or two-species mixtures	400, 700	365	<i>Avena barbata</i>	C ₃ Herbaceous
					<i>Hemizonia congesta</i>	C ₃ Herbaceous
Raubenheimer & Ripley, 2022 ¹²	Grahamstown/Makhandha, South Africa	Field (open-top chambers); potted tree seedlings grown with density-manipulated grass competition	400, 800	180	<i>Vachellia karroo</i>	C ₃ Woody
					<i>Senegalia burkeii</i>	C ₃ Woody
					<i>Senegalia caffra</i>	C ₃ Woody
					<i>Senegalia nigrescens</i>	C ₃ Woody
					<i>Vachellia erioloba</i>	C ₃ Woody
					<i>Vachellia exuvialis</i>	C ₃ Woody
					<i>Vachellia gracilis</i>	C ₃ Woody
					<i>Vachellia karroo</i>	C ₃ Woody
					<i>Vachellia nilotica</i>	C ₃ Woody
					<i>Vachellia robusta</i>	C ₃ Woody
					<i>Vachellia sieberiana</i>	C ₃ Woody
					<i>Vachellia swazica</i>	C ₃ Woody
					<i>Vachellia tortilis</i>	C ₃ Woody
					<i>Achillea millefolium</i>	C ₃ Herbaceous
					<i>Agropyron repens</i>	C ₃ Herbaceous
					<i>Amorpha canescens</i>	C ₃ Woody
					<i>Andropogon gerardii</i>	C ₄ Herbaceous
					<i>Anemone cylindrica</i>	C ₃ Herbaceous
					<i>Asclepias tuberosa</i>	C ₃ Herbaceous
					<i>Bouteloua gracilis</i>	C ₄ Herbaceous
					<i>Bromus inermis</i>	C ₃ Herbaceous
					<i>Koeleria cristata</i>	C ₃ Herbaceous
					<i>Lespedeza capitata</i>	C ₃ Woody
					<i>Lupinus perennis</i>	C ₃ Woody
					<i>Petalostemum villosum</i>	C ₃ Woody
					<i>Poa pratensis</i>	C ₃ Herbaceous
					<i>Schizachyrium scoparium</i>	C ₄ Herbaceous
					<i>Solidago rigida</i>	C ₃ Herbaceous
Reich <i>et al.</i> , 2024 ¹³	Cedar Creek, Minnesota, USA	Field (FACE); plot-level experimentally sown monocultures and 16-species mixtures	415, 550	8395		

					<i>Sorghastrum nutans</i>	C ₄ Herbaceous
Tom-Dery <i>et al.</i> , 2019 ¹⁴	Tolon, Ghana (experiment carried out in Germany)	Glasshouse growth chambers (CO ₂ -controlled); potted tree seedlings grown alone or with grass competitors	400, 950	112	<i>Vitellaria paradoxa</i>	C ₃ Woody
Van De Velde <i>et al.</i> , 2019 ¹⁵	Wilrijk, Belgium	CO ₂ - and climate-controlled chambers; experimentally sown, potted monocultures and interspecific mixtures	400, 620	90	<i>Lolium perenne</i>	C ₃ Herbaceous
					<i>Plantago lanceolata</i>	C ₃ Herbaceous
Van Kleunen <i>et al.</i> , 2006 ¹⁶	Basel, Switzerland	Field (FACE); experimentally assembled monocultures and two-species mixtures across planting-density gradients	400, 600	365	<i>Bromus erectus</i>	C ₃ Herbaceous
					<i>Carex flacca</i>	C ₃ Herbaceous
Wray & Strain, 1987 ¹⁷	Durham, North Carolina, USA	Glasshouse (CO ₂ -controlled); experimentally sown, potted monocultures and two-species replacement-series mixtures	350, 500	63	<i>Andropogon virginicus</i>	C ₄ Herbaceous
					<i>Aster pilosus</i>	C ₃ Herbaceous
Yu <i>et al.</i> , 2019 ¹⁸	Durham, North Carolina, USA	CO ₂ -controlled growth chambers; experimentally sown, potted monocultures and two-species mixtures	400, 800	120	<i>Bouteloua eriopoda</i>	C ₄ Herbaceous

Table S2. Description of unpublished data used in the meta-analysis

Unpublished study	Description
RUECF NERC Experiment	The unpublished data are from the Rhodes University Elevated CO ₂ Facility NERC (National Ecological Research Council.) experiment which is a collaboration between Rhodes University, South Africa, Sheffield and Edinburgh Universities, UK. This study was conducted at the Rhodes University Elevated CO ₂ Facility in Open-Top Chambers (OTCs) as described by Raubenheimer and Ripley ¹² . In this experiment ¹⁹ , 12 C ₃ woody species were grown either individually or in competition with the C ₄ grass species <i>Themeda triandra</i> at two CO ₂ treatment levels (400 or 550 ppm) for a total of 360 days in 37.L pots.

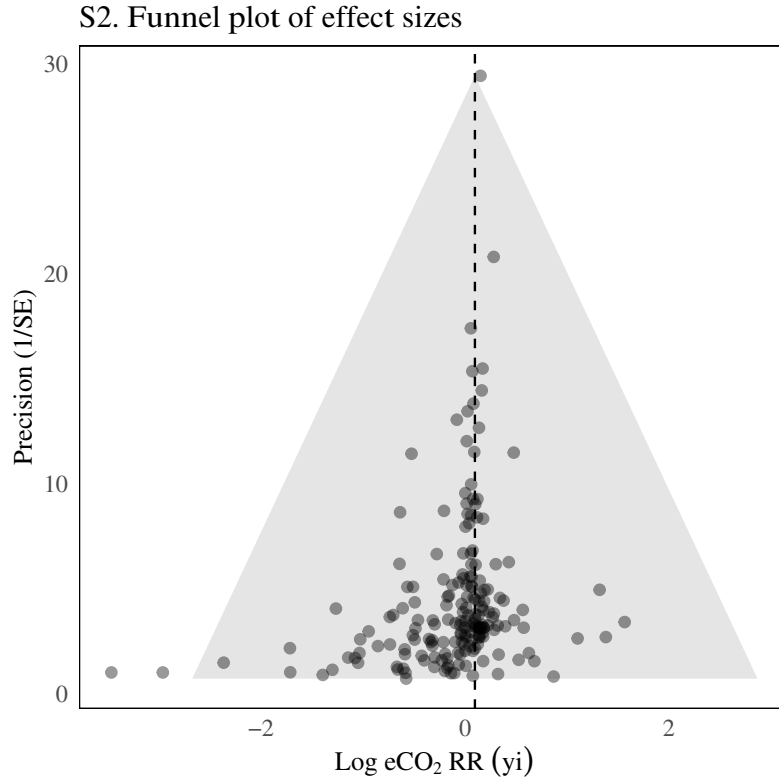


Figure S2. Funnel plot showing the distribution of log elevated CO₂ response ratios (yi) against study precision (1/SE) for all species-level estimates included in the meta-analysis. The dashed vertical line indicates the overall estimated effect size under ambient competition conditions, and the shaded triangle represents the expected 95% pseudo confidence region in the absence of publication bias. Asymmetry in the plot suggests the presence of small-study effects, motivating the sensitivity analysis presented in Table S2. n = 97.

Table S2. Comparison of meta-analysis results including all studies (Original Model) and excluding low-precision studies (Sensitivity Model; SE > 1). Estimates represent the log response ratio to elevated CO₂ across competition treatments. The similarity of estimates between models indicates that the main findings are robust to the influence of low-precision studies.

Table S2. Comparison of meta-analysis results with and without low-precision studies (SE > 1)									
Term	Estimate	SE	95% CI	P-value	Estimate	SE	95% CI	P-value	Δ Estimate
intrcpt	0.10	0.05	[-0.004, 0.196]	0.059	0.10	0.05	[-0.001, 0.198]	0.051	0.003
Monoculture vs Isolated	-0.03	0.05	[-0.127, 0.067]	0.543	-0.03	0.05	[-0.130, 0.063]	0.494	-0.004
Mixture vs Isolated	-0.14	0.04	[-0.227, -0.057]	0.001	-0.14	0.04	[-0.225, -0.055]	0.001	0.002

S3. Influence of individual studies on eCO₂ RR estimate

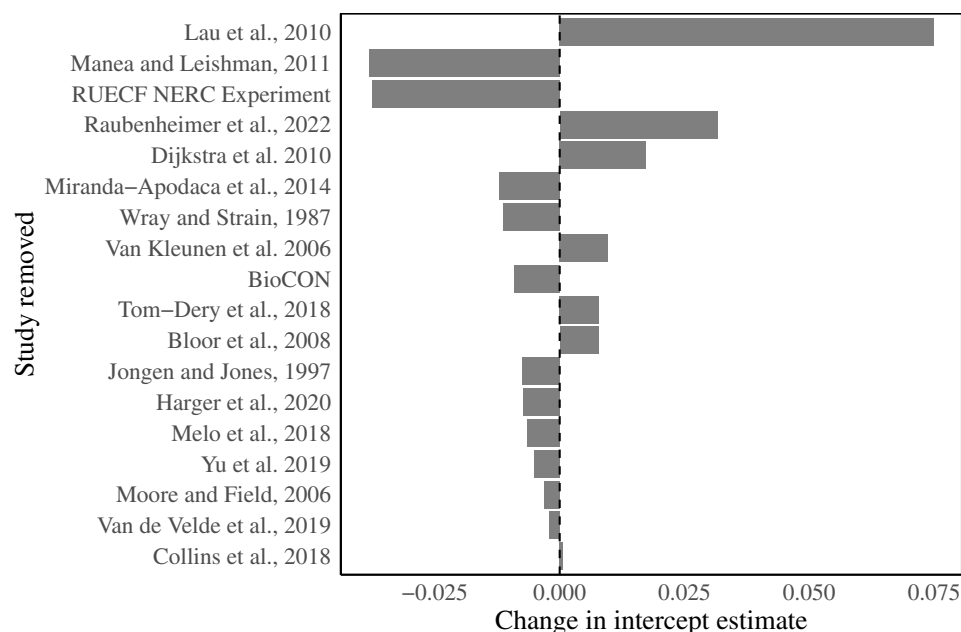


Figure S3. Influence diagnostics based on leave-one-out meta-analysis. Each bar represents the change in the overall intercept (CO₂ response estimate under isolated conditions) when a single study is removed from the model. The dashed line indicates the full-model estimate. While a few studies notably shift the intercept (e.g., Lau et al., 2010; Raubenheimer et al., 2022), the overall effect direction remains consistent, suggesting that findings are not driven by any single influential study.

Table S3. Results of leave-one-out influence analysis for the meta-analysis of plant biomass response to elevated CO₂ under isolated conditions. Each row shows the estimated intercept (CO₂ effect) when a single study is excluded from the model. The difference from the full-model estimate (Δ Intercept) and the absolute value of that difference ($|\Delta|$ Intercept) indicate the relative influence of each study. Larger values suggest greater influence on the overall result, though no single study qualitatively altered the direction or significance of the effect.

Table S3. Leave-one-out influence analysis: effect of excluding each study on the estimated CO₂ response under isolated competition.

Study (Paper)	Intercept (LOO)	Δ Intercept	$ \Delta $ Intercept
Lau et al., 2010	0.171	0.075	0.075
Manea and Leishman, 2011	0.058	-0.038	0.038
RUECF NERC Experiment	0.059	-0.037	0.037
Raubenheimer et al., 2022	0.128	0.032	0.032
Dijkstra et al. 2010	0.113	0.017	0.017
Miranda-Apodaca et al., 2014	0.084	-0.012	0.012
Wray and Strain, 1987	0.085	-0.011	0.011
Van Kleunen et al. 2006	0.106	0.010	0.010
BioCON	0.087	-0.009	0.009
Tom-Dery et al., 2018	0.104	0.008	0.008
Bloor et al., 2008	0.104	0.008	0.008
Jongen and Jones, 1997	0.089	-0.007	0.007

Table S3. Leave-one-out influence analysis: effect of excluding each study on the estimated CO₂ response under isolated competition.

Study (Paper)	Intercept (LOO)	Δ Intercept	$ \Delta $ Intercept
Harger et al., 2020	0.089	-0.007	0.007
Melo et al., 2018	0.089	-0.007	0.007
Yu et al. 2019	0.091	-0.005	0.005
Moore and Field, 2006	0.093	-0.003	0.003
Van de Velde et al., 2019	0.094	-0.002	0.002
Collins et al., 2018	0.097	0.001	0.001

Table S4. Comparison of meta-analysis results before and after excluding small-study outliers (trim-and-fill adjustment). Estimates represent the log response ratio to elevated CO₂ under different competition treatments. While the trimmed model yielded slightly smaller absolute effect sizes, the direction and significance of the mixture competition effect remained consistent, indicating that the main results are robust to the influence of outlier studies.

Table S4. Comparison of meta-analysis results before and after trim-and-fill adjustment

Term	Estimate (Orig)	SE (Orig)	95% CI (Orig)	p-value (Orig)	Estimate (Trimmed)	SE (Trimmed)	95% CI (Trimmed)	p-value (Trimmed)	Estimate Diff
Intercept	0.10	0.05	[-0.00, 0.20]	0.059	0.08	0.05	[-0.01, 0.18]	0.093	-0.01
Monoculture vs Isolated	-0.03	0.05	[-0.13, 0.07]	0.543	-0.02	0.05	[-0.12, 0.07]	0.624	0.01
Mixture vs Isolated	-0.14	0.04	[-0.23, -0.06]	0.001	-0.13	0.04	[-0.21, -0.04]	0.003	0.01

S4. Funnel plot of RCV effect sizes

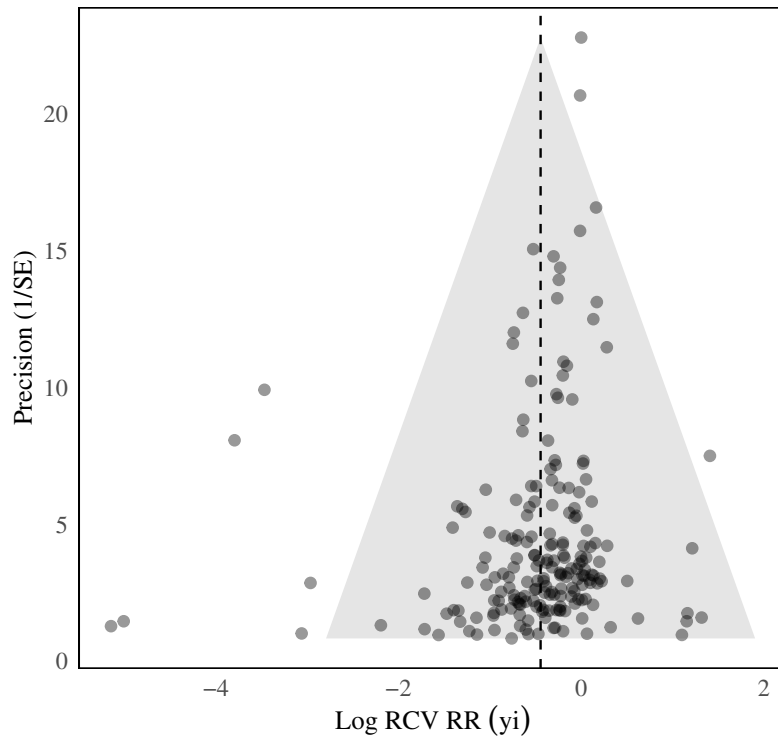


Figure S4. Funnel plot showing log response ratios of relative competitive vulnerability (RCV RR; y_i) plotted against study precision ($1/SE$). The shaded triangle represents pseudo 95% confidence limits centered on the model intercept (dashed line). Points represent individual effect sizes; their vertical spread reflects variation in study precision. The symmetrical shape suggests limited evidence for publication bias in the RCV data. $n = 97$.

Table S5. Comparison of meta-analysis estimates for relative competitive vulnerability (RCV) with and without low-precision studies ($SE > 1$). Values represent model estimates, standard errors, 95% confidence intervals, and p-values for each fixed effect term. The final column shows the difference in estimates between the original and sensitivity analyses. Minimal changes suggest robust results.

Table S5. Comparison of RCV meta-analysis with and without low-precision studies ($SE > 1$)

Term	Estimate	SE	95% CI	P-value	Estimate	SE	95% CI	P-value	Δ Estimate
intrept	-0.44	0.21	[-0.860, -0.030]	0.036	-0.45	0.21	[-0.864, -0.026]	0.037	-0.01
Low CO ₂ vs High CO ₂	0.03	0.06	[-0.093, 0.160]	0.602	0.03	0.06	[-0.093, 0.161]	0.601	0.00
Diversity vs Competition	-0.01	0.26	[-0.510, 0.492]	0.972	-0.01	0.26	[-0.515, 0.497]	0.971	0.00
Interaction: Low CO ₂ × Diversity	0.03	0.07	[-0.107, 0.169]	0.657	0.03	0.07	[-0.109, 0.168]	0.676	0.00

S6. Influence of individual studies on RCV estimate

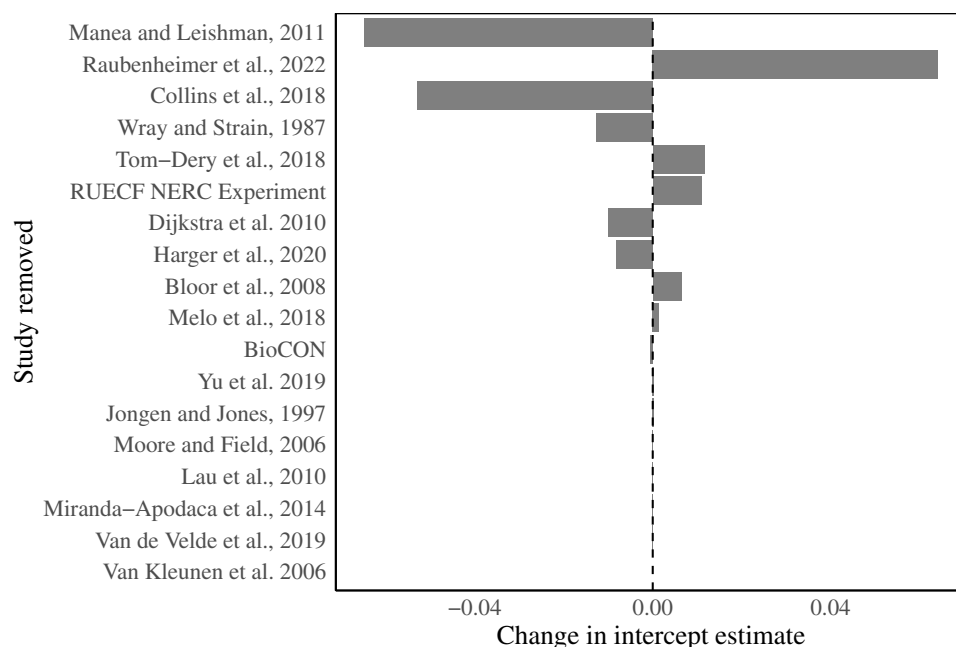


Figure S6. Leave-one-out influence analysis of individual studies on the RCV meta-analysis intercept estimate. Each bar shows the change in the estimated intercept when the corresponding study is removed from the model. The dashed line represents the full-model intercept. The modest magnitude of all changes ($|\Delta| < 0.05$) indicates that no single study unduly influenced the overall estimate of relative competitive vulnerability.

Table S6. Leave-one-out influence analysis for RCV meta-analysis: effect of excluding each study on the estimated log response ratio under high CO₂ and competitive conditions. The table shows the model intercept when each study is excluded (Intercept (LOO)), the change from the full model estimate (Δ Intercept), and the absolute magnitude of this change ($|\Delta|$ Intercept).

Table S6. Leave-one-out influence analysis: effect of excluding each study on the estimated RCV under high CO₂ × competition.

Study (Paper)	Intercept (LOO)	Δ Intercept	$ \Delta $ Intercept
Manea and Leishman, 2011	-0.51	-0.06	0.06
Raubenheimer et al., 2022	-0.38	0.06	0.06
Collins et al., 2018	-0.50	-0.05	0.05
Wray and Strain, 1987	-0.46	-0.01	0.01
Tom-Dery et al., 2018	-0.43	0.01	0.01
RUECF NERC Experiment	-0.43	0.01	0.01
Dijkstra et al. 2010	-0.45	-0.01	0.01
Harger et al., 2020	-0.45	-0.01	0.01
Bloor et al., 2008	-0.44	0.01	0.01
Melo et al., 2018	-0.44	0.00	0.00
BioCON	-0.45	0.00	0.00
Yu et al. 2019	-0.44	0.00	0.00
Jongen and Jones, 1997	-0.44	0.00	0.00
Moore and Field, 2006	-0.44	0.00	0.00

Table S6. Leave-one-out influence analysis: effect of excluding each study on the estimated RCV under high CO₂ × competition.

Study (Paper)	Intercept (LOO)	Δ Intercept	Δ Intercept
Lau et al., 2010	-0.44	0.00	0.00
Miranda-Apodaca et al., 2014	-0.44	0.00	0.00
Van de Velde et al., 2019	-0.44	0.00	0.00
Van Kleunen et al. 2006	-0.44	0.00	0.00

Table S7. Comparison of RCV meta-analysis results before and after outlier trimming. Meta-analysis fixed effects estimates for log(RCV) vulnerability ratios under high CO₂ × competition conditions, comparing the original model (all studies) to a trimmed model excluding extreme outliers (yi values beyond ±2.5 SD from the mean). Estimates, standard errors (SE), confidence intervals (95% CI), and p-values are reported for each model, along with the difference in effect estimates (Δ Estimate). Trimming had minimal influence on effect estimates, indicating model robustness.

Table S7. Comparison of RCV meta-analysis results before and after outlier trimming (±2.5 SD)

Term	Estimate (Orig)	SE (Orig)	95% CI (Orig)	p-value (Orig)	Estimate (Trimmed)	SE (Trimmed)	95% CI (Trimmed)	p-value (Trimmed)	Estimate Diff
intrcpt	-0.44	0.21	[-0.86, -0.03]	0.036	-0.45	0.19	[-0.81, -0.08]	0.017	-0.00
Low CO ₂ vs High CO ₂	0.03	0.06	[-0.09, 0.16]	0.602	0.03	0.06	[-0.10, 0.16]	0.636	-0.00
Diversity vs Competition	-0.01	0.26	[-0.51, 0.49]	0.972	0.23	0.22	[-0.21, 0.67]	0.298	0.24
Interaction: Low CO ₂ × Diversity	0.03	0.07	[-0.11, 0.17]	0.657	-0.05	0.07	[-0.19, 0.09]	0.501	-0.08

S7. Funnel plot of community-level CO₂ effect sizes

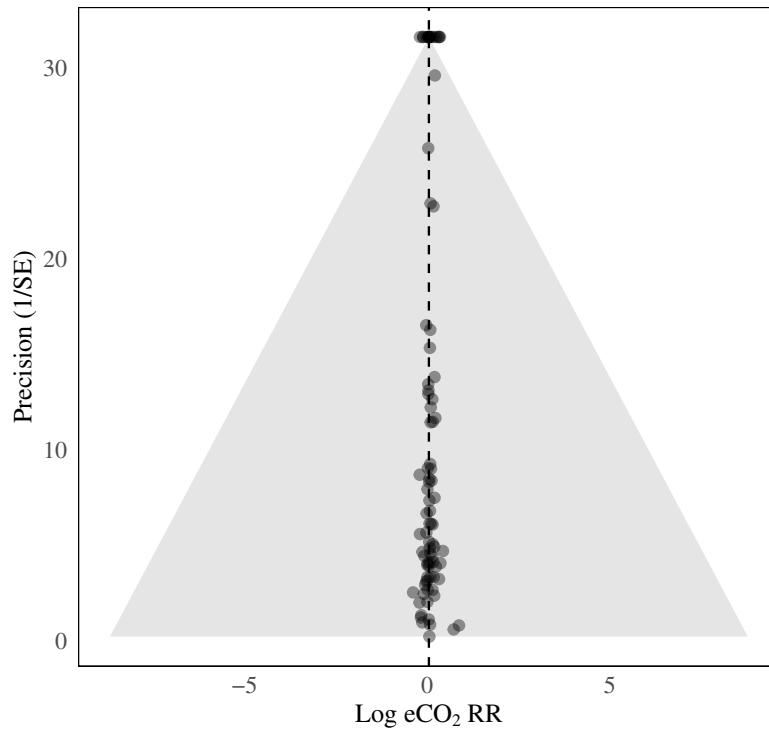


Figure S7. Funnel plot of mixture-level CO₂ effect sizes (log response ratios, y_i) plotted against study precision (1/SE). The shaded triangle represents pseudo 95% confidence limits centered on the model intercept (dashed line). Each point corresponds to an individual site-level estimate. The overall symmetry of the plot and lack of extreme outliers suggest minimal evidence of small-study effects or publication bias in the mixture-level dataset.

Table S8a. Meta-analysis results for species-level log-transformed biomass response ratios to elevated CO₂ across competition treatments (log eCO₂ RR). Fixed effects represent estimated differences in eCO₂ response between monoculture or mixture treatments relative to isolated plants. The model was fit using a multivariate random-effects meta-analysis (*rma.mv*) with restricted maximum likelihood (REML), including random intercepts for study (*Study*) and functional type (*FT*). Variance components, model fit metrics (AIC and log-likelihood), and intra-class correlation coefficients (ICC) are reported to quantify between-study and between-functional-type heterogeneity.

Table S8a: Species-level meta-analysis results – log(eCO₂ response ratio)

	Term	Estimate	SE	X95..CI	p.value
Fixed Effects	intrcpt	0.096	0.051	[-0.004, 0.196]	0.059
Fixed Effects	Monoculture vs Isolated	-0.030	0.049	[-0.127, 0.067]	0.543
Fixed Effects	Mixture vs Isolated	-0.142	0.043	[-0.227, -0.057]	0.001
Model Fit	AIC	321.52			
Model Fit	Log Likelihood	-155.76			
Random Effects	Study	0.015622	0.124987		
Random Effects	Functional type	0.000000	0.000005		
Variance Partitioning (ICC)	Study	1.000			
Variance Partitioning (ICC)	Functional type	0.000			

Table S8b. Meta-analysis results for species-level log-transformed relative competitive vulnerability ratios (RCV RR) across CO₂ treatments and competition contexts. Fixed effects represent differences in RCV RR between ambient and elevated CO₂ conditions (Low CO₂ vs High CO₂), between diversity-based and competition-based contrasts (Diversity vs Competition), and their interaction. The model was fit using a multivariate random-effects meta-analysis (*rma.mv*) with restricted maximum likelihood (REML), including random intercepts for study (*Study*) and functional type (*FT*). Variance components, model fit metrics (AIC and log-likelihood), and intra-class correlation coefficients (ICC) are reported to quantify between-study and methodological heterogeneity.

Table S8b: Meta-analysis Results – log(RCV vulnerability ratio)

	Term	Estimate	SE	X95..CI	p.value
Fixed Effects	intrept	-0.445	0.212	[-0.860, -0.030]	0.036
Fixed Effects	Low CO ₂ vs High CO ₂	0.034	0.065	[-0.093, 0.160]	0.602
Fixed Effects	Diversity vs Competition	-0.009	0.256	[-0.510, 0.492]	0.972
Fixed Effects	Interaction: Low CO ₂ × Diversity	0.031	0.070	[-0.107, 0.169]	0.657
Model Fit	AIC	1704.42			
Model Fit	Log Likelihood	-846.21			
Random Effects	Paper	0.015622	0.124987		
Random Effects	Functional type	0.000000	0.000005		
Variance Partitioning (ICC)	Paper	1.000			
Variance Partitioning (ICC)	Functional type	0.000			

Table S8c. Linear model results testing the correlation between species' log-transformed biomass response ratios to elevated CO₂ (eCO₂ RR) and their relative competitive vulnerability (RCV RR). Separate models are presented for diversity-based (isolated vs. mixture) and competition-based (monoculture vs. mixture) contrasts. Each model includes an intercept and slope for RCV RR, along with associated standard errors, 95% confidence intervals, and p-values. Model fit is summarized using R-squared (R²) for each contrast.

Table S8c: Correlation between eCO₂ response and competitive vulnerability

	Term	Estimate	SE	X95..CI	p.value
Isolated vs. mixture	(Intercept)	0.083	0.056	[-0.030, 0.196]	0.147
	yi_RCV	1.077	0.098	[0.880, 1.274]	< 0.001
Monoculture vs. mixture	(Intercept)	0.138	0.065	[0.007, 0.269]	0.039
	yi_RCV	0.812	0.063	[0.685, 0.939]	< 0.001
Model Fit	R-squared (Mono v. mix)	0.757			
Model Fit	R-squared (Iso v. mix)	0.748			

Table S9a. Meta-analysis results for mixture-level log-transformed biomass response ratios to elevated CO₂ (log eCO₂ RR). Fixed effects represent differences in eCO₂ response between monoculture and mixture treatments, with the intercept corresponding to mixed-species communities. The model was fit using a multivariate random-effects meta-analysis (*rma.mv*) with restricted maximum likelihood (REML), including a random intercept for study (*Study*). Variance components, model fit metrics (AIC and log-likelihood), and intra-class correlation coefficients (ICC) are reported to quantify between-study heterogeneity.

Table S9a: Meta-analysis Results - mixture log(eCO₂ response ratio)

	Term	Estimate	SE	X95..CI	p.value
Fixed Effects	Intercept (Mixture)	0.043	0.028	[-0.013, 0.099]	0.131
	Monoculture vs Mixture	0.058	0.021	[0.018, 0.099]	0.005
Model Fit	AIC	165.30			
	Log Likelihood	-79.65			
Random Effects	Study	0.007284	0.085346		
Variance Partitioning (ICC)	Study	1.000			

Table S9b. Mixed-effects model summary of species-level biomass responses to elevated CO₂ (eCO₂ RR) as a function of relative dominance within communities. Fixed effects represent differences in eCO₂ response across dominance categories, with the intercept corresponding to the “Rare” group. The model was fit using a linear mixed-effects model (`lmer`) with random intercepts for study (`Study`) and functional type (`FT`). The table reports fixed effects, random standard deviations, model metrics (R², AIC, sample sizes), and pairwise comparisons between dominance categories based on estimated marginal means.

Table S9b: Mixed-effects model summary: fixed effects, random SDs, model metrics, and pairwise comparisons.

Term	Estimate	Std. Error	df	t value	Pr(> t)
(Intercept)	-0.79	0.19	32.00	-4.26	0.000
Dominance: Subordinate	0.54	0.23	87.36	2.39	0.019
Dominance: Dominant	0.82	0.23	80.91	3.53	0.001
Random SD: Study	0.28				
Random SD: FT	0.00				
Random SD: Residual	0.81				
Marginal R ²	0.15				
Conditional R ²					
AIC	242.90				
n (total obs)	92.00				
n (Study groups)	16.00				
n (FT groups)	3.00				
Contrast: Rare vs. Subordinate	-0.54	0.24		-2.24	0.070
Contrast: Rare vs. Dominant	-0.82	0.25		-3.32	0.004
Contrast: Subordinate vs. Dominant	-0.28	0.22		-1.25	0.429

Table S10a.1. MANOVA results testing for trait space separation across species success groups using four multivariate test statistics (Pillai’s Trace, Wilks’ Lambda, Hotelling–Lawley Trace, and Roy’s Largest Root). The analysis was performed on the first two discriminant axes from a discriminant analysis of principal components (DAPC). Residual sums of squares (SS), residual degrees of freedom, and standard errors (SE) are reported for LD1 and LD2.

Table S10a.2. Pairwise comparisons among species success groups along the first two DAPC axes (LD1 and LD2). P-values are reported both as raw values and after Benjamini–Hochberg (BH) correction for multiple testing. Comparisons were conducted using Tukey’s HSD test following ANOVA on each discriminant axis.

Table S10a.1: MANOVA results for trait space separation across success groups

Test	F	Num.df	Den.df	p-value
Pillai	5.63	6	190	0.0000

Table S10a.1: MANOVA results for trait space separation across success groups

Test	F	Num.df	Den.df	p-value
Wilks	5.63	6	188	0.0000
Hotelling-Lawley	5.62	6	186	0.0000
Roy	7.77	3	95	0.0001
Residual SS (LD1)	95.00			
Residual SS (LD2)	95.00			
Residual df	95.00			
Residual SE (LD1)	1.00			
Residual SE (LD2)	1.00			

Table S10a.2: Pairwise comparisons of success groups across DAPC axes (BH-adjusted)

Comparison	Raw.p	Adj..p..BH.	Axis
Highly Successful-CO ₂ Responsive but Vulnerable	0.0143665582	0.034933689	LD1
Stable but Unresponsive-CO ₂ Responsive but Vulnerable	0.9943892211	0.999941837	LD1
Less Successful-CO ₂ Responsive but Vulnerable	0.9999418373	0.999941837	LD1
Stable but Unresponsive-Highly Successful	0.0003122387	0.001873432	LD1
Less Successful-Highly Successful	0.0174668445	0.034933689	LD1
Less Successful-Stable but Unresponsive	0.9977874607	0.999941837	LD1
Highly Successful-CO ₂ Responsive but Vulnerable	0.0661265945	0.132253189	LD2
Stable but Unresponsive-CO ₂ Responsive but Vulnerable	0.0058786531	0.019811111	LD2
Less Successful-CO ₂ Responsive but Vulnerable	0.0145175015	0.064185091	LD2
Stable but Unresponsive-Highly Successful	0.9108633281	0.910863328	LD2
Less Successful-Highly Successful	0.5431766976	0.814765046	LD2
Less Successful-Stable but Unresponsive	0.8100422308	0.910863328	LD2

Table S10b. Standardized coefficients from linear mixed-effects models (*lmer*) predicting log-transformed species responses to elevated CO₂ (eCO₂ RR) and competitive vulnerability (RCV RR) from plant functional traits. Separate models were fit for each response: eCO₂ RR, RCV RR (isolated vs. mixture), and RCV RR (monoculture vs. mixture). Models included random intercepts for study (*Study*) and functional type (*FT*). Coefficients, standard errors (SE), 95% confidence intervals (CI), and p-values are reported for each trait predictor. Traits were log-transformed prior to analysis using safe offsets, and model parameters were standardized by refitting with scaled predictors.

Table S10b: Standardized coefficients (LMM) predicting CO₂ and competition responses from plant traits. Raw p-values and Benjamini–Hochberg FDR-adjusted p-values are shown; adjustment applied within biologically defined trait groups.

Response	Trait	Coefficient	SE	CI low	CI high	p	p adj
RCV RR (iso vs. mix)	Growth rate	0.20	0.09	0.03	0.36	0.02	0.10
	Drought tol.	-0.20	0.09	-0.38	-0.03	0.03	0.10
	LDMC	-0.16	0.10	-0.36	0.04	0.12	0.72
	Invasive	0.13	0.08	-0.04	0.30	0.13	0.13
	Fire tol.	0.13	0.09	-0.05	0.30	0.15	0.31
	Shade tol.	0.09	0.09	-0.08	0.26	0.31	0.41
	Leaf area	-0.07	0.09	-0.25	0.11	0.42	0.83
	SLA	-0.07	0.10	-0.27	0.12	0.45	0.83
	Leaf N	-0.06	0.11	-0.27	0.16	0.62	0.83
	Max height	0.05	0.10	-0.15	0.25	0.64	0.66

Table S10b: Standardized coefficients (LMM) predicting CO₂ and competition responses from plant traits. Raw p-values and Benjamini–Hochberg FDR-adjusted p-values are shown; adjustment applied within biologically defined trait groups.

Response	Trait	Coefficient	SE	CI low	CI high	p	p adj
	Lifespan	0.05	0.10	-0.15	0.25	0.64	0.98
	SSD	-0.05	0.11	-0.26	0.16	0.66	0.66
	Leaf C:N	-0.04	0.11	-0.25	0.17	0.69	0.83
	Root:Shoot	-0.02	0.08	-0.19	0.15	0.82	0.82
	Cold tol.	0.02	0.09	-0.17	0.20	0.87	0.87
	Leaf C	0.01	0.08	-0.16	0.17	0.95	0.95
	Growth form	0.00	0.09	-0.17	0.18	0.97	0.98
	Grime strat.	0.00	0.08	-0.16	0.16	0.98	0.98
RCV RR (mono vs. mix)	SSD	-0.52	0.18	-0.88	-0.15	0.01	0.01
	Max height	0.48	0.23	0.03	0.94	0.04	0.04
	Invasive	0.28	0.13	0.02	0.55	0.04	0.04
	Leaf area	0.28	0.15	-0.02	0.58	0.07	0.29
	Leaf C	0.28	0.16	-0.05	0.60	0.10	0.29
	Shade tol.	0.21	0.14	-0.06	0.48	0.13	0.52
	LDMC	0.18	0.15	-0.12	0.49	0.23	0.43
	Leaf C:N	-0.17	0.16	-0.50	0.15	0.28	0.43
	Fire tol.	-0.14	0.12	-0.39	0.11	0.28	0.54
	Root:Shoot	-0.14	0.13	-0.40	0.12	0.29	0.29
	Cold tol.	-0.16	0.19	-0.54	0.22	0.41	0.54
	Lifespan	0.13	0.16	-0.20	0.45	0.44	0.81
	Grime strat.	0.10	0.14	-0.18	0.38	0.46	0.81
	Drought tol.	0.10	0.16	-0.21	0.41	0.54	0.54
	Leaf N	0.09	0.15	-0.21	0.39	0.56	0.60
	SLA	-0.10	0.20	-0.50	0.29	0.60	0.60
	Growth rate	0.04	0.14	-0.24	0.33	0.77	0.81
	Growth form	0.03	0.14	-0.24	0.31	0.81	0.81
eCO ₂ RR	Drought tol.	-0.22	0.13	-0.47	0.03	0.09	0.35
	LDMC	-0.24	0.14	-0.52	0.05	0.10	0.60
	Growth rate	0.12	0.12	-0.12	0.36	0.31	0.75
	Fire tol.	0.10	0.12	-0.14	0.35	0.40	0.69
	Invasive	0.10	0.12	-0.14	0.33	0.42	0.42
	Lifespan	0.09	0.14	-0.19	0.36	0.52	0.75
	SLA	-0.09	0.14	-0.37	0.19	0.53	0.85
	Leaf N	-0.09	0.15	-0.38	0.21	0.55	0.85
	Leaf area	-0.07	0.13	-0.33	0.18	0.57	0.85
	Shade tol.	-0.05	0.12	-0.29	0.19	0.65	0.69
	Max height	0.06	0.14	-0.22	0.35	0.66	0.75
	Root:Shoot	-0.05	0.12	-0.28	0.18	0.67	0.67
	Cold tol.	-0.05	0.13	-0.32	0.21	0.69	0.69
	Grime strat.	-0.04	0.11	-0.26	0.19	0.74	0.75
	SSD	-0.05	0.15	-0.34	0.24	0.75	0.75
	Growth form	0.04	0.13	-0.21	0.29	0.75	0.75
	Leaf C	0.00	0.12	-0.23	0.23	0.98	0.98
	Leaf C:N	0.00	0.15	-0.29	0.30	0.98	0.98

133

134 **Table S11.** Summary of field experiments exposing intact, multi-species plant communities to
135 elevated CO₂ (eCO₂), showing ecosystem type, experimental approach, and whether eCO₂ was
136 associated with changes in total community biomass and/or relative species abundance. This table
137 is included to support the idea that community-level responses to eCO₂ rarely reflect uniform
138 stimulation across all species; instead, increases in community biomass, where observed, typically

139 arise from heterogeneous species responses in which some taxa respond positively while others
 140 show weak, neutral, or negative responses. By documenting whether relative abundance or
 141 dominance shifts were reported across systems, the table provides qualitative evidence that
 142 competitive interactions under eCO₂ can structure community outcomes.

Study (site)	Ecosystem / vegetation type	Experimental design (FACE / OTC / in situ)	Reported change in total community biomass under eCO ₂	Change in relative species abundance under eCO ₂ ?
Morgan et al. 2011 ²⁰ (CPER, CO, USA)	Semi-arid shortgrass steppe grassland	OTC	Variable; precipitation- dependent	Yes — C ₃ grasses favoured under eCO ₂ ; warming favoured C ₄ grasses
Nowak et al. 2004 ²¹ (multiple sites)	Forests, grasslands, deserts	FACE (review of multiple experiments)	Generally positive but variable across systems	Indirect — review synthesizes species- specific and dominance-driven responses in intact communities but does not quantify abundance shifts
Seiler et al. 2009 ²² (Merritt Island, FL, USA)	Fire-regenerated scrub- oak ecosystem	OTC	Increase	Yes — community- level response driven primarily by dominant species
Hovenden et al. 2006 ²³ (TasFACE, Tasmania, Australia)	Mixed native grassland (C ₃ /C ₄)	FACE	Increase	No — productivity increased under eCO ₂ without sustained changes in community composition
Owensby et al. 1999 ²⁴ (Konza Prairie, KS, USA)	Tallgrass prairie (C ₄ - dominated)	OTC	Increase in dry years; little effect in wet years	Yes — functional group shifts (decline of C ₃ grasses; increase of C ₃ forbs and sedges)
Niklaus & Körner 2004 ²⁵ (Nenzlingen, Switzerland)	Temperate calcareous grassland	In situ CO ₂ enrichment	Increase	Yes (subtle) — increased evenness via positive responses of subdominant sedges and bryophytes
Mueller et al. 2016 ²⁶ (PHACE, WY, USA)	Semi-arid northern mixed-grass prairie	FACE	eCO ₂ alone: variable; eCO ₂ + warming: consistent increase	Yes (strong) — long- term reversal of dominance from C ₄ grasses to C ₃ graminoids

Dijkstra et al., 2002 ²⁷ (Merritt Island, FL, USA)	Fire-regenerated scrub- oak ecosystem	OTC	Increase	Yes — biomass gains were species-specific and dominated by <i>Quercus myrtifolia</i> ; subdominant species showed weak or no response
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