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#### Article:

Wang, J., Wang, H., Zhang, C. et al. (18 more authors) (2025) Ozone Production Underestimation Over the Tibetan Plateau: The Role of NOx and OVOCs Modeling Uncertainties. Journal of Geophysical Research: Atmospheres, 130 (10). e2025JD043321. ISSN 2169-897X

https://doi.org/10.1029/2025jd043321

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#### Ozone Production Underestimation over the Tibetan Plateau: The Role of NO<sub>x</sub> and **OVOCs Modeling Uncertainties** 3

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#### **Key Points:** 29

- Field observations serve as benchmarks for evaluating uncertainties in GEOS-Chem's 30 simulation of the O<sub>3</sub> chemical budget. 31
- GEOS-Chem underestimates  $O_3$  chemical production by 60% compared to observations. 32 •
- Model biases in  $O_3$  chemistry are largely driven by model underestimations of  $NO_x$ , with 33 • significant contributions from OVOCs. 34

35

### 36 Abstract

- 37 Chemical production and destruction are key components of the tropospheric ozone (O<sub>3</sub>) budget,
- 38 yet significant model spreads exist among various global models in estimating O<sub>3</sub> chemical
- 39 budgets. Observations in the background atmospheres play a crucial role in validating and
- 40 evaluating model uncertainties in O<sub>3</sub> chemistry. Here we conducted comprehensive observations
- $_{41}$  of the O<sub>3</sub> chemical budget at a remote site on the Tibetan Plateau. By comparing the observed O<sub>3</sub>
- 42 chemical budgets with those simulated by GEOS-Chem, a global chemical transport model, we
- 43 identified a 60% underestimation in the O<sub>3</sub> production rate by GEOS-Chem. Detailed budget
- analysis of  $O_3$  and radicals revealed that this discrepancy primarily arose from GEOS-Chem's inadequate representation of nitrogen oxides (NO<sub>x</sub>), followed by oxygenated volatile organic
- inadequate representation of nitrogen oxides  $(NO_x)$ , followed by oxygenated volatile organic compounds (OVOCs). Comparisons with previous studies indicate a widespread model bias in
- simulating O3 precursors and their chemical budgets. These findings underscore the challenges
- in understanding  $O_3$  chemistry linked to our imperfect understanding of the atmospheric fate of
- 49 NO<sub>x</sub> and OVOCs.

## 50 Plain Language Summary

51 Tropospheric ozone (O<sub>3</sub>) plays a crucial role in air quality, climate change, and ecosystem health.

52 The estimate and prediction for tropospheric  $O_3$  levels and variabilities are heavily dependent on

 $_{53}$  our understanding of the O<sub>3</sub> chemical budgets. However, significant simulation spreads of

- tropospheric  $O_3$  photochemistry exist among various chemical transport models, indicating an
- incomplete scientific understanding of  $O_3$  chemistry. This study compares the observed  $O_3$
- chemical budget with predictions from the GEOS-Chem model, and analyzes the source of  $\Omega_{12}$  and  $\Omega_{12}$
- modeling uncertainties. A significant underestimation of  $O_3$  chemical production by GEOS-Chem is derived, which is primarily due to uncertainties in simulating nitrogen oxides (NO<sub>x</sub>) and
- $_{59}$  oxygenated volatile organic compounds (OVOCs). Future efforts should particularly focus on
- refining the representation of  $NO_x$  and OVOCs to improve our understanding of  $O_3$  chemistry
- and enhance predictive capabilities for  $O_3$  changes and their environmental impacts.

# 62 Keywords

63 Tropospheric ozone, atmospheric photochemistry, reactive nitrogen, carbonyl compounds,

- 64 chemical transport model
- 65

# 66 **1 Introduction**

67 Tropospheric ozone  $(O_3)$  stands as a critical factor influencing air quality, climate change, 68 human health, and ecosystem productivity (*Monks et al.*, 2015). It plays a central role in the 69 photochemical oxidation of air pollutants, directly reacting with alkenes and generating OH

radicals, the key oxidant in the atmosphere, through photolysis (*Levy*, 1971). Ranked as the third-

largest greenhouse gas after carbon dioxide and methane, tropospheric  $O_3$  significantly

- influences the Earth's radiative balance, with a direct radiative forcing of  $\sim 0.4$  W m<sup>-2</sup> according
- to the sixth assessment report of IPCC (Intergovernmental Panel on Climate Change, 2023). In
- addition, O<sub>3</sub> exposure poses adverse effects on human health, agricultural yields, and ecosystem

productivity (*Ainsworth et al.*, 2012; *Arjomandi et al.*, 2018). The significance drives a

- <sup>76</sup> substantial interest among researchers and policymakers to understand the distribution and
- variability of tropospheric O<sub>3</sub>.

The distribution and variability of tropospheric O<sub>3</sub> are determined by an interplay of 78 79 several physical and chemical processes, known as O<sub>3</sub> budgets. Physical budgets of O<sub>3</sub> involve processes such as stratosphere-to-troposphere transport and surface deposition, while chemical 80 budgets include chemical production via the oxidation reactions of precursors like nitrogen 81 oxides (NO<sub>x</sub> = NO + NO<sub>2</sub>) and volatile organic compounds (VOCs) under ultraviolet light, 82 alongside its chemical destruction through photolysis and reactions with radicals and alkenes. 83 Chemical production and destruction constitute the major terms in the global tropospheric  $O_3$ 84 budget, estimated at  $\sim$ 5000 Tg yr<sup>-1</sup> on average, significantly outweighing physical budget terms 85 (Young et al., 2018). Notably, regions with high surface O<sub>3</sub> concentration usually coincide with 86 areas of elevated production rate of O<sub>3</sub>, such as polluted regions. Moreover, the long-term trends 87 of net O<sub>3</sub> production, i.e., O<sub>3</sub> chemical production minus its chemical destruction, closely align 88 with the trends of O<sub>3</sub> concentration over the last decades (Archibald et al., 2020), emphasizing 89 the pivotal role of chemical budgets in shaping O<sub>3</sub> distribution and trends. 90

Despite incorporating the state-of-the-art understanding of O<sub>3</sub> chemistry, global models 91 exhibit large spreads in simulating O<sub>3</sub> chemical budgets. As assessed in a recent model 92 intercomparison project, the Atmospheric Chemistry-Climate Model Intercomparison Project 93 (ACCMIP), models' estimates of  $O_3$  chemical budgets vary by a factor of 2 (Young et al., 2018). 94 Additionally, chemistry-climate models have been found to underestimate  $O_3$  change rates by 95 96 over 50% over the past 5-6 decades (Parrish et al., 2014). These model uncertainties indicate an incomplete understanding of the chemical processes, which hinders our ability to estimate the 97 historic O<sub>3</sub> trends and predict its future changes and environmental impacts, including radiative 98 99 forcing (Iglesias-Suarez et al., 2018).

100 Efforts have been made to elucidate the sources of the uncertainties in  $O_3$  chemistry simulations. Sensitivity analyses of global models have revealed a significant sensitivity of  $O_3$  to 101 its precursor emissions and various chemical factors such as mechanisms, kinetics, photolysis 102 frequencies, and atmospheric humidity (Christian et al., 2017; Ridley et al., 2017; Wild et al., 103 104 2020). The ACCMIP evaluation highlights that significant uncertainties stem from poor models' representation of the chemical and physical processes of O<sub>3</sub> and its precursors (Young et al., 105 2018). Observations in the background atmosphere provide crucial constraints of tropospheric O<sub>3</sub> 106 chemistry, which serve as benchmarks for model validation and evaluation, aiding in identifying 107 108 the uncertainty sources in the model. For example, observation-to-model comparisons of  $O_3$ chemistry in the tropical marine boundary layer have highlighted that the omission of halogen 109 chemistry significantly contributes to modeling errors in O<sub>3</sub> chemical destruction (Read et al., 110 2008). Aircraft measurements and subsequent comparisons to model results have uncovered 111 consistent model underestimation of  $NO_x$ , the chemical precursor of  $O_3$ , in the troposphere by 112 global models (Nicely et al., 2016; Ye et al., 2023). In remote forested regions, emissions of 113 biogenic VOCs, notably isoprene, together with NO<sub>x</sub>, are attributed to the discrepancies in the O<sub>3</sub> 114 chemical budget and O<sub>3</sub> concentration between observations and chemical transport model 115 predictions (Pike et al., 2010). 116

The in-depth study of atmospheric chemistry over the Tibetan Plateau from spring to summer of 2019, referred to as the @Tibet 2019, offers an opportunity to comprehensively observe the O<sub>3</sub> chemical budget in the background atmosphere. As the third pole of the world, with an average altitude exceeding 4000m, the Tibetan Plateau represents the pristine tropospheric conditions. Our study presents the detailed O<sub>3</sub> chemical budget observations during the @Tibet 2019 field campaign and compares them with those derived from the global model

- 123 GEOS-Chem. We reveal an alarming modeling bias in O<sub>3</sub> production by GEOS-Chem, with a 60%
- underestimation of O<sub>3</sub> production rates compared to our observations. Modeling uncertainties in
- $NO_x$  and oxygenated volatile organic compounds (OVOCs) emerge as major contributors to the
- underestimation, with our observations serving as a benchmark for model evaluation.

# 127 **2 Methods**

# 128 **2.1 Measurements of O<sub>3</sub> chemistry**

The @Tibet 2019 field campaign took place at the Nam Co Multisphere Observation and Research Station (NMC,  $30^{\circ}46.44'$ N,  $90^{\circ}59.31'$ E, 4730 m a.s.l.) from April 28 to July 10, 2019. Located on the southeastern shore of Nam Co Lake and approximately 15 km north of the Nyainqentanglha Mountains, the NMC represents a pristine alpine site. Limited local anthropogenic emissions primarily stemmed from sporadic cooking activities and vehicle use at the station. A comprehensive dataset on O<sub>3</sub> chemistry was collected from May 1 to May 21 during the @Tibet 2019 campaign.

136  $O_3$  measurements were conducted using a commercial UV photometric analyzer (49C, 137 Thermo Electronics, USA) with a time resolution of 1 min (*Xu et al.*, 2023). The limit of 138 detection for the O<sub>3</sub> analyzer is determined to be 1 ppbv, and the precision was 2% of the 139 instrument response as evaluated by multiple calibrations during the campaign. The O<sub>3</sub> 140 instrument was usually calibrated with an O<sub>3</sub> primary standard calibrator (49i-PS, Thermo 141 Scientific, USA). A calibration range of 0-200 ppbv was applied to fully cover our measurement

142 range of ambient  $O_3$  concentration during the campaign.

NO<sub>x</sub> was measured using a customized two-channel broadband cavity-enhanced 143 absorption spectrometer (BBCEAS) as described in *Fang et al.*, (2017). For this study, the 144 optical cavities were extended from 42cm to 1 meter to improve the instrument response 145 146 sensitivity. Instrument performance was evaluated by supplying cylinder zero air (purity>0.99.999%) through the sampling line for 7 hours (Figure S1a). During this period, the 147 instrument responses were  $8.4\pm32.4$  (1 $\sigma$ ) pptv for channel one and  $-4.3\pm17.7$  pptv for channel 148 two, respectively (Figure S1b). The precision of the BBCEAS instrument for NO<sub>2</sub> was 149 150 determined to be  $\pm 10$  pptv. Allan deviation analysis indicated a detection limit (1 $\sigma$ ) of 16 and 25 pptv for NO<sub>2</sub> in channel one and channel two, respectively, at a time resolution of 36 s (Figure 151 152 S1c). In this study, 1-min NO<sub>2</sub> measurement data were collected, and 1-h averaged NO<sub>2</sub> measurements were shown and used as observational constraints for the chemical box model. 153 Throughout the campaign, the BBCEAS uncertainty was determined to be 4%. 154

During the field measurements, ambient air was directly sampled into channel two for NO<sub>2</sub> measurements. For NO<sub>x</sub> measurements, ambient air was mixed with O<sub>3</sub> (5 ppmv) for NO titration before being sampled into channel one. However, NO titration in channel one was unsuccessful due to an O<sub>3</sub> generator power failure. Therefore, only NO<sub>2</sub> measurements from channel two were used for further analysis.

Nonmethane VOCs (NMVOCs) and OVOCs were analyzed using an online gas
 chromatography system equipped with mass spectrometry and a flame ionization detector (GC MS/FID, TH-PKU 300B, Wuhan Tianhong Instrument Co. Ltd, China). Ambient VOCs were
 concentrated in a cryogen-free pre-concentration system with a temperature of -150 °C before
 C2-C5 hydrocarbons were quantified by the flame ionization detector, while C5-C12
 hydrocarbons, OVOCs, and halocarbons were quantified by the mass spectrometer (*Wang et al.*,

166 2014). The time resolution of the instrument was 1 h. Multipoint calibrations were performed at

the beginning and the end of the campaign, and daily single-point calibrations at 2 ppbv were

168 conducted throughout the campaign. Photochemical Assessment Monitoring Station (PAMS)

standard mixture and TO-15 standard mixture (100 ppbv, Spectra Gases Inc., New Jersey, USA) were used as the calibration source. The instrument uncertainties were within  $\pm 10\%$  for all

calibrated species. The limit of detections for single VOC species were characterized to be 0.01 -

171 calibrated s<sub>1</sub> 172 0.05 ppbv.

172 0.05 ppbv. 173 Measurements of other parameters related to  $O_3$  chemistry were briefly described as

follows. HONO levels were observed using a commercial LOng-Path Absorption Photometer

(LOPAP-03, QUMA, Germany), based on wet chemical sampling and photometric detection
 techniques (*Heland et al.*, 2001; *Wang et al.*, 2023). The detection limit of the LOPAP was 5

pptv at a time resolution of 3min. Carbon monoxide (CO), methane(CH<sub>4</sub>), and water vapor were

measured using a commercial wavelength-scanned cavity ring-down spectroscopy (model G2401,

179 PICARRO). Formaldehyde could not be detected by the GC-MS/FID instrument, so a

180 commercial MAX-DOAS instrument was employed (Xing et al., 2021). The MAX-DOAS

instrument captured the vertical distribution of HCHO at altitudes of 0–4 km, and average

182 formaldehyde concentration in the lowest level (0–100 m) was applied in this study. Photolysis

183 frequencies were measured using a spectroradiometer (CCD-Spectrograph, Meteorologie

184 Consult GmbH, Germany). Meteorological parameters, such as temperature and pressure, were

185 measured by an automatic weather station. Refer to Table S1 for a comprehensive list of

186 parameters described above.

# 187 2.2 GEOS-Chem model description

The GEOS-Chem (version 13.3.4, DOI: https://zenodo.org/record/7254256) was utilized 188 to simulate the O<sub>3</sub> photochemistry of the @Tibet observation period (1 to 21 May 2019). The 189 configuration of the GEOS-Chem model was briefly described as follows. The GEOS-Chem 190 model was driven by meteorological data obtained from the Modern-Era Retrospective Analysis 191 for Research and Applications, version 2 (MERRA-2), which was retrieved from the NASA 192 Global Modeling and Assimilation Office (GMAO) at the Goddard Earth Observing System 193 (GEOS) (Gelaro et al., 2017). The MERRA-2 data has a horizontal resolution of 194  $0.625^{\circ}(\text{longitude}) \times 0.5^{\circ}(\text{latitude})$  and was regrided to  $5^{\circ}(\text{longitude}) \times 4^{\circ}(\text{latitude})$  for input 195 into the global nested-grid simulation. The simulation over the East Asian region  $(60^\circ - 120^\circ \text{ E},$ 196  $10^{\circ} - 50^{\circ}$  N) was operated at a higher resolution of  $0.625^{\circ} \times 0.5^{\circ}$ . The vertical grid structure 197 comprised 47 hybrid sigma/pressure levels spanning from the surface to the mesosphere, with the 198 199 lowest layers about 100 m in height.

200 Emissions were calculated using the Harvard-NASA Emissions Component (HEMCO,

version 3.2.2) (*Keller et al.*, 2014). Global anthropogenic emissions of SO<sub>2</sub>, NO<sub>x</sub>, NH<sub>3</sub>, CO,

202 VOCs, dust, primary organic carbon, and black carbon were obtained from the Community

Emissions Data System (CEDS) (*Hoesly et al.*, 2018) and then overwritten over mainland China by the Multi-resolution Emission Inventory (MEIC) (*Li et al.*, 2017; *Zheng et al.*, 2018).

Biogenic VOC emissions were based on the Model of Emissions of Gases and Aerosol from

Nature version 2.1 (MEGAN2.1) framework (*Guenther et al.*, 2012). Soil NO<sub>x</sub> emissions were

calculated using the Berkeley-Dalhousie Soil  $NO_x$  Parameterization (BDSNP) scheme (*Hudman* 

208 *et al.*, 2012).

NO<sub>x</sub> + O<sub>x</sub> + Br + Cl + I + aerosols chemical mechanisms in the troposphere (fullchem) were included in the GEOS-Chem simulation (*Jacob*, 2000; *Mao et al.*, 2013). Photolysis frequencies were calculated with the FAST-JX v7.0 photolysis mechanism(*Eastham et al.*, 2014; *Wild et al.*, 2000). As for O<sub>3</sub> production, the reactions of NO with HO<sub>2</sub> radicals and 78 RO<sub>2</sub> species were included. As for O<sub>3</sub> loss, the reaction of O<sub>3</sub> with OH radicals, O<sub>3</sub> with HO<sub>2</sub> radicals, the O<sup>1</sup>D atom (the product of O<sub>3</sub> photolysis) with H<sub>2</sub>O, and NO<sub>2</sub> with OH radicals were included. O<sub>3</sub> chemical budgets from the NMC grid (centered at 31°N, 91.25°E, 557 hPa) modeled

by GEOS-Chem were extracted from the model. Additionally, modeled concentrations for  $O_3$ and its precursors during the same period as the field campaign were extracted, compared with observations, and utilized for box model constraints for examination of chemical mechanisms related to  $O_3$  chemistry in the GEOS-Chem. See Table S2 for a list of the parameters extracted from the GEOS-Chem.

#### 221 **2.3 Chemical box model description**

A photochemical box model, based on the Framework for 0-D Atmospheric Modeling 222 223 (F0AM, v4.0) (Wolfe et al., 2016), was employed to calculate the chemical budget of O<sub>3</sub> and investigate the factors contributing to the bias in O<sub>3</sub> chemistry modeling. The chemical 224 mechanisms of the box model were based on the Master Chemical Mechanism (MCM, v3.3.1), 225 with a subset extracted from the MCM website (https://mcm.york.ac.uk/MCM/), comprising 226 11152 gas-phase reactions relevant to our VOC and OVOC measurement (Jenkin et al., 2015; 227 Saunders et al., 2003). Gas-phase reaction rate constants were determined using equations from 228 229 the MCM and constrained by our temperature and pressure observations. The photolysis frequencies were calculated using the parameterization method provided by the MCM (Eq. 1), 230 which accounted for the solar zenith angles (SZA) based on the location of the site and the period 231 of the campaign, and the species-specific coefficients (l, m, and n) extracted from the MCM 232 website. Physical processes such as dilution and dry deposition were not incorporated into the 233 chemical box model. The model timespan ranged from 1 to 21 May 2019, with a time step of 1h. 234  $j = l \cdot \cos(SZA)^m \cdot \exp(-n \cdot \sec(SZA))$ 235 Eq. 1

In Eq.1, *j* represented the photolysis frequency; *SZA* represented the solar zenith angles; *l*, *m*, and *n* were species-specific coefficients.

238 Several model runs were designed to examine the  $O_3$  chemistry and its influencing 239 factors. The setup and configuration for each model run were described as follows.

240 A base model (referred to as Box\_S0) was established to calculate the chemical budgets of  $O_3$  constrained by our measurements. In the model, time series data for the parameters 241 outlined in Table S1 were constrained with a time resolution of 1h. Due to a lack of NO 242 measurements, simulation of NO was conducted in the model of Obs. where VOCs, NO<sub>2</sub>, O<sub>3</sub>, 243 and HONO, along with the photolysis frequencies for NO<sub>2</sub> ( $j_{NO2}$ ) and HONO ( $j_{HONO}$ ) were 244 constrained. Concentrations of radicals (OH, HO<sub>2</sub>, and RO<sub>2</sub>) were also predicted. Comparisons 245 between the predictions of radicals and field observations were conducted to validate the model. 246 As shown in Figure S2, the averaged measurement/model ratio of 0.88, 0.99, and 1.22 for OH, 247 HO<sub>2</sub>, and RO<sub>2</sub>, respectively. The discrepancy between model prediction and observation of 248 radical concentrations fell within the measurement uncertainty of  $\pm 25\%$ . These evaluations 249 validated the model's satisfactory representation of O<sub>3</sub> chemistry observation. 250

A GEOS-Chem model scenario (referred to as Box GC) was designed to simulate the  $O_3$ 251 photochemistry constrained by GEOS-Chem extractions with the MCM chemical mechanism. 252 All parameters listed in Table S2 were constrained in Box\_GC with a time resolution of 1h. 253 Compared with Obs. constraints, Box GC was constrained with fewer NMVOCs and OVOCs 254 species. The NMVOCs and OVOCs species extracted from the GEOS-Chem accounted for 80% 255 of total OH reactivity and further contributed substantially to O<sub>3</sub> chemistry, while the species not 256 extracted contributed little, as estimated by Box GC. Moreover, the concentrations of the 257 remaining NMVOCs and OVOCs species were nearly null as simulated in GEOS-Chem. 258

Sensitivity model runs (referred to as Box S1-S3) were designed to investigate the 259 magnitude of  $O_3$  chemistry change in response to its influencing factors and quantify the 260 contribution of influencing factors to the modeling bias. The influencing factors included 261 NMVOCs, OVOCs, and NO<sub>x</sub>. Influencing parameters were substituted with measurements in the 262 sensitivity model runs to serve as constraints, replacing the GEOS-Chem extractions. 263 Specifically, in model S2, NMVOCs were constrained with our measurements in place of that 264 extracted from GEOS-Chem, while other settings remained identical to model S1. In model S3, 265 OVOCs were constrained with our measurements in place of those extracted from GEOS-Chem, 266 while other settings remained identical to model S2. The measured concentration of  $NO_2$  and 267 modeled NO in model S0 was constrained in place of the NO<sub>x</sub> extracted from GEOS-Chem in 268 model S4, while other settings remained identical to model S3. By replacing and adding the 269 measurement constraints of the influencing parameters instead of replacing them individually, 270 the synthesized effects of these factors on O<sub>3</sub> chemistry were derived, considering that the factors 271 272 were chemically coupled with each other.

For each model run described above, the  $O_3$  production rate ( $P_{03}$ ) and the  $O_3$  destruction rate ( $D_{03}$ ) were extracted from the model according to the equations below. Finally, the net chemical production rate of  $O_3$  ( $Net_{03}$ ) is calculated by subtracting  $D_{03}$  from  $P_{03}$ .

276 
$$P_{03} = k_1[NO][HO_2] + \sum k_i[NO][RO_2]_i \quad \text{Eq. 2}$$
  
277 
$$D_{03} = f \times j_{0^1D}[O_3] + k_2[O_3][HO_2] + k_3[O_3][OH] + k_4[NO_2][OH] \quad \text{Eq. 3}$$
  
278 
$$Net_{03} = P_{03} - D_{03} \quad \text{Eq.4}$$

In the equation above, the [X] represented the concentration of species X;  $j_{O^1D}$  was the photolysis frequency of O<sub>3</sub>; *f* was the fraction of O<sup>1</sup>D radicals reacting with H<sub>2</sub>O to produce OH radical, relative to the total sink strength of O<sup>1</sup>D radicals, which was calculated with H<sub>2</sub>O concentration, temperature, and pressure.

# 283 **3 Results and Discussion**

# 284 **3.1 Meteorological and chemical conditions**

The time series of the meteorological and chemical parameters observed during the campaign are shown in Figure S3. Throughout the campaign, temperature varied from -6.4 °C to 13.7 °C with an average of  $3.1\pm3.8(1\sigma)$  °C. Precipitations rarely occur during the campaign, limited to three small snow events. Wind patterns showed distinct diurnal cycles, with westerly winds reaching speeds of ~4-7 m s<sup>-1</sup> in the daytime, shifting to the southerly winds accompanied by lower speeds at night. Solar radiation was notably intense, with  $j_{0^1D}$  reaching approximately  $5.5\times10^{-5}$  s<sup>-1</sup> at noon, twice the level observed at the same latitude at the sea level (*Yang et al.*, 2021). These meteorological observations align with previous multi-year data from this site (*Yin* 

*et al.*, 2017), indicating stable and representative late-spring conditions on the plateau,

characterized by cold and dry weather influenced by high altitude and westerlies.

The background nature of this site is further supported by observations of emission 295 tracers. For instance, observed NO<sub>2</sub> concentrations averaged at  $153\pm115(1\sigma)$  pptv. NO levels, 296 297 calculated with Box S0, averaged at  $20\pm28(1\sigma)$  pptv, reaching ~50 pptv in the daytime (Figure S3d). CO and CH<sub>4</sub> were on the levels of  $115\pm17(1\sigma)$  ppbv and  $1.89\pm0.02(1\sigma)$  ppmv, respectively. 298 Alkanes contributed the most to the NMVOC concentrations with a mean concentration of 299  $2.67\pm0.65(1\sigma)$  ppbv (Figure S3f). The mixing ratios of NO<sub>x</sub> and NMVOCs are similar to those 300 measured at Waliguan global atmospheric watch station on the Tibetan Plateau (Xue et al., 301 2013). Occasional spikes in NO2 and NMVOC concentrations indicated the limited influence of 302 anthropogenic emissions at the NMC station, such as cooking and vehicle emissions (Figure 303 S3d&g). The biogenic sources were also relatively weak, with concentrations of alkenes on the 304 order of 0.2 ppby. It was worth noting that substantial levels of OVOCs were observed at this 305 background site, with a mean value of 3.02±1.15(1sd) ppbv, primarily composed of 306 formaldehyde, acetaldehyde, and acetone (Figure S3h). 307

308 Overall, observations of the meteorological and chemical parameters confirm the alpine background nature of the NMC site, characterized by minimal anthropogenic and biogenic 309 emissions, representing the regional background atmospheric condition of the Tibetan Plateau. 310 The meteorological and chemical conditions at NMC are comparable to those reported at other 311 312 remote sites across the Tibetan Plateau, which indicate a stable and regionally representative atmospheric condition during our campaign. This makes NMC observations a suitable basis for 313 model evaluation and comparison, even at the relatively coarse resolution of 0.5 degrees in the 314 GEOS-Chem. 315

The observed  $O_3$  concentration averaged at  $61\pm14(1\sigma)$  ppbv, displaying a distinct bridge-316 shaped diurnal pattern. O<sub>3</sub> levels measured are in accordance with previous multi-year O<sub>3</sub> 317 measurements at this site, which reported a concentration of  $58.6\pm12.2(1\sigma)$  ppbv in May (Yin et 318 al., 2017). The  $O_3$  levels are comparable with those measured at other remote sites over the 319 320 Tibetan Plateau, such as the Waliguan Global Atmospheric Watch Station (Xu et al., 2016), but higher than typical alpine background sites in the midlatitudes of the Northern Hemisphere, such 321 as Jungfraujoch and Hohenpeissenberg (Parrish et al., 2021; Parrish et al., 2014). Compared to 322 GEOS-Chem simulations, the observed O<sub>3</sub> concentrations were reasonably compared, especially 323 during the daytime. Nevertheless, the GEOS-Chem failed to reproduce the distinct diurnal 324 patterns observed (Figure S4). Previous research has shown that vertical mixing and downward 325 transport of stratospheric and free-tropospheric air are the primary contributors to the enhanced 326 O<sub>3</sub> concentration in the daytime over the Tibetan Plateau (Xu et al., 2018; Xu et al., 2023). 327 GEOS-Chem's underestimation of the impact of deep stratospheric ozone intrusions at such 328 high-altitude sites could partially explain the discrepancies in O<sub>3</sub> (Fiore et al., 2014; Zhang et al., 329 2020). Additionally, uncertainties in chemical budget terms also play a crucial role in influencing 330 O<sub>3</sub> modeling accuracy (Archibald et al., 2020). However, such discussions and comparisons 331 remain limited due to the scarcity of observational constraints of O<sub>3</sub> chemistry, especially in 332 remote atmospheres. In this study, the comprehensive suite of observational parameters 333 regarding O<sub>3</sub> chemistry allowed for the detailed calculation and analysis of the O<sub>3</sub> chemical 334 budget at Nam Co, enabling the model comparison and analysis of modeling uncertainties. 335

**336 3.2 GEOS-Chem's underestimation of observed ozone chemical budget** 

The diurnal pattern of the  $O_3$  chemical budget is shown in Figure 1a. Generally,  $P_{O3}$ 337 exceeded the  $D_{03}$  during daytime hours (8:00-20:00, Beijing Time), resulting in net O<sub>3</sub> chemical 338 production. The average daytime  $P_{03}$  and  $D_{03}$  were at  $0.8\pm0.3(1\sigma)$  and  $0.4\pm0.2(1\sigma)$ , respectively. 339 The average daytime  $Net_{03}$  reached  $0.4\pm0.2(1\sigma)$  ppbv h<sup>-1</sup>, with an accumulative  $Net_{03}$ 340 averaging at  $3.5\pm3.2(1\sigma)$  ppbv day<sup>-1</sup> over the campaign period. The net ozone production result 341 aligns with the conclusions drawn from other continental background sites, and the  $Net_{03}$ 342 derived in this study is comparable with previous measurements (Fischer et al., 2003; Xue et al., 343 2013; Zanis et al., 2000). For instance, at Waliguan, spring and summer daytime Net<sub>03</sub> were 344

reported as 0.31 and 0.26 ppbv h<sup>-1</sup>, respectively (*Xue et al.*, 2013).  $Net_{O3}$  was in the range of 0.1 to 0.3 ppbv h<sup>-1</sup> at Mt. Cimone in summer (*Fischer et al.*, 2003), and in the range of 0.13 to 0.27

ppbv  $h^{-1}$  at the Mt. Jungfraujoch in spring and summer (*Zanis et al.*, 2000).



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349 Figure 1. Diel pattern of  $O_3$  chemical budget as calculated from (a) the box model constrained with observations at Nam Co (Box S0) and (b) the GEOS-Chem model. In both 350 panels, the light blue and dark blue bars represent the reaction rate of NO with HO<sub>2</sub> radicals and 351  $RO_2$  radicals, respectively. The dark orange bars represent the reaction rate of the O<sup>1</sup>D atom, the 352 product of O<sub>3</sub> photolysis, with water vapor. The light orange bars represent the reaction rate of 353  $O_3$  with OH radicals. The dark purple bars represent the reaction rate of  $O_3$  with HO<sub>2</sub> radicals, 354 355 and the purple bars represent the reaction rate of NO<sub>2</sub> with OH radicals, which competes with  $NO_2$  photolysis and is considered an  $O_3$  destruction pathway. The line shows the net  $O_3$ 356 production rate over the diel cycle. 357

In contrast to the Box\_S0 simulation, the GEOS-Chem model predicted net O<sub>3</sub> chemical 358 destruction in the daytime (Figure 1b). The average daytime  $Net_{03}$  simulated by the GEOS-359 Chem was  $-0.05\pm0.06(1\sigma)$  ppbv h<sup>-1</sup>. The  $P_{03}$  simulated by GEOS-Chem was approximately 360  $0.3\pm0.1(1\sigma)$  ppbv h<sup>-1</sup> during the daytime, representing a 60% underestimation compared to 361 observations. The  $D_{03}$  was predicted at  $0.3\pm0.2(1\sigma)$  ppbv h<sup>-1</sup> in GEOS-Chem, which was slightly 362 underestimated ( $\sim 25\%$ ) relative to observations. As for specific terms in the O<sub>3</sub> chemical budget, 363 364 reactions of NO with RO<sub>2</sub> and HO<sub>2</sub> radicals are both notably higher in the observations than in the GEOS-Chem simulation. For O<sub>3</sub> destruction, GEOS-Chem remarkably underestimated the 365

rate of  $O_3$  photolysis, while destruction budget terms were similar between observations and GEOS-Chem simulations.

Our finding regarding the significant underestimation of  $P_{03}$  in global model simulations 368 are consistent with results from previous research. For instance, in the marine troposphere, global 369 model underestimation of  $O_3$  precursors has been validated through comparisons with aircraft 370 and satellite measurements, which result in similar biases in  $P_{03}$  (Nicely et al., 2016; Shah et al., 371 2023). Similarly, studies over polar snowpacks have reported a 50% underestimation of  $NO_x$  by 372 global models such as GEOS-Chem, which would further lead to  $P_{03}$  underestimation (Huang et 373 al., 2017). Both our study and these previous investigations focus on remote atmospheres, which 374 375 are more representative of the tropospheric atmosphere than studies conducted in polluted regions. These comparisons suggest that the model biases in  $O_3$  chemical budgets identified in 376 these studies may reflect a broader, potentially global issue, underscoring the importance of 377 further attribution of this problem. 378

#### 379

## 3.3 Factors controlling the ozone production underestimation

The discrepancies underscore GEOS-Chem's limitations in accurately modeling the  $O_3$ chemical budget. The modeling uncertainties may come from either the different treatments of chemical mechanisms regarding  $O_3$  chemistry between the box model and the GEOS-Chem, or the GEOS-Chem's bias in predicting these  $O_3$  precursors or chemical parameters such as photolysis frequencies. In this section, we aim to attribute the discrepancies in the ozone chemical budget and identify the major sources of the modeling bias.

386 To evaluate the treatment of chemical mechanisms related to  $O_3$  chemistry in GEOS-Chem, the GEOS-Chem modeled  $O_3$  chemical budget was compared with a box model 387 simulation constrained with GEOS-Chem extractions (Box GC, Method). The box model was 388 389 incorporated with the near-explicit chemical mechanism, Master Chemical Mechanism, and was validated to be reasonably satisfying for the radicals and thus O<sub>3</sub> chemical budget modeling at 390 this remote site (Figure S2). The O<sub>3</sub> chemical budget simulated by Box\_GC is shown in Figure 391 S6. The average daytime  $P_{03}$  and  $D_{03}$  was  $0.3\pm0.1(1\sigma)$  and  $0.3\pm0.2(1\sigma)$  pptv h<sup>-1</sup>, respectively, 392 resulting the daytime  $Net_{03}$  of -0.02±0.06(1 $\sigma$ ).  $P_{03}$ ,  $D_{03}$ , and  $Net_{03}$  simulated by Box\_GC were 393 therefore comparable to GEOS-Chem model (Figure S6), excluding the treatment of chemical 394 395 mechanism in the GEOS-Chem as the major reason for the observation to model discrepancies in 396 O<sub>3</sub> chemical budget.

397 Simulation of O<sub>3</sub> precursor by GEOS-Chem was investigated and compared with our observations at Nam Co to further attribute the observation to model discrepancies. In 398 background atmospheres, O<sub>3</sub> chemical production was strongly NO<sub>x</sub>-limited, with NO reacting 399 with peroxy radicals outcompeting the self-reaction of peroxy radicals, thereby resulting in  $O_3$ 400 chemical production (Kleinman et al., 2001). We first compared the NO<sub>x</sub> concentrations 401 simulated by GEOS-Chem with observations at NMC and found an average underestimation of 402 NO<sub>x</sub> concentrations by 35% in GEOS-Chem (Figure 2a). As for diurnal patterns, GEOS-Chem 403 underpredicted NO<sub>x</sub> concentrations all day, with severe underestimation during the daytime 404 (Figure 2c), posing a severer influence on  $O_3$  production. Given the linear sensitivity of  $O_3$ 405 chemical production to NO<sub>x</sub> concentration, a 35% underestimation of NO<sub>x</sub> would lead to  $P_{O3}$  of a 406 similar magnitude. However, the 60% underestimation of  $P_{03}$  derived from our comparison 407 could not be fully explained by uncertainties in NO<sub>x</sub> simulations. The GEOS-Chem model 408

- underpredicts the reaction rate of NO with  $RO_2$  radicals and  $HO_2$  radicals by 70% and 40%,
- 410 respectively (Figure 1), suggesting a model underestimation of other O<sub>3</sub> precursors.





412 Figure 2. Comparison of the observation (red) with the GEOS-Chem simulation (blue) of

413 **O<sub>3</sub> precursors.** Panels a-b compare the statistical distribution of NO<sub>x</sub> and OVOC, respectively, 414 and panels c-d compare the diurnal pattern of NO<sub>x</sub> and OVOC, respectively. OVOC represents

415 the sum concentration of formaldehyde, acetaldehyde, and acetone.

We then compared other  $O_3$  precursors including NMVOCs, OVOCs, and CO between 416 our observations and GEOS-Chem simulations. As shown in Figure 2b, GEOS-Chem 417 underestimated OVOC concentrations by 60%. As for specific species, we found a 90% GEOS-418 Chem underestimation of acetaldehyde, followed by formaldehyde by 60% and acetone by 40% 419 (Figure S2 & Figure S3). Moreover, the a larger discrepancy of OVOCs in the daytime than at 420 nighttime. GEOS-Chem representation for NMVOCs seemed more accurate than OVOCs, with 421 an underestimation of 40% (Figure S7b). In contrast, GEOS-Chem generally predicted the CO 422 levels at Nam Co, with a discrepancy of 20% between simulations and observations (Figure S7a). 423 A detailed budget of  $RO_x$  radicals (= OH, HO<sub>2</sub>, and  $RO_2$ ) was calculated using the Box\_S0 424 model (box model constrained by our measurements) to examine the role of NMVOC and 425 426 OVOCs in O<sub>3</sub> production. As shown in Figure S8, the reaction between OH radicals and OVOCs accounted for 40% of the total OH loss rate, leading to the production of RO<sub>2</sub> radicals, while the 427 reaction between OH and NMVOCs contributed only around 10%. In addition to promoting 428 429 secondary peroxy radical production via oxidation by OH radicals, photolysis of OVOCs directly generated HO<sub>2</sub> and RO<sub>2</sub> radicals, which contributed to 15% of the total RO<sub>x</sub> radicals budget on 430 the daytime average. Overall, our observations indicated that OVOCs greatly facilitated the 431 primary production and chain propagation of radicals at this site, further promoting O<sub>3</sub> chemical 432 production. Therefore, underestimation of OVOCs was deduced as another cause of  $P_{03}$ 433 underestimation in GEOS-Chem. 434

435 Other parameters influencing the  $O_3$  production were also examined in this study.  $O_3$ 436 photolysis contributed to ~70% of RO<sub>x</sub> radical sources (Figure S8) and further promoted  $O_3$ 437 chemical production. In addition, it presents as the major  $O_3$  destruction pathway. Parameters

- 438 affecting the  $O_3$  photolysis rates, including  $O_3$  concentration, photolysis frequency, and  $H_2O$
- 439 concentrations, were assessed. As shown in Figure S5&S7, GEOS-Chem overestimates  $O_3$
- 440 concentrations by 15% on a daily average, while it overestimates daytime  $O_3$  by about 7%. These 441 discrepancies may reflect an underrepresentation of background  $O_3$  levels and transport-
- 441 discrepancies may reflect an underrepresentation of background O<sub>3</sub> revers and transport 442 chemistry interactions in the model. However, differences in O<sub>3</sub> concentrations cannot account
- 443 for the 60% underestimation in  $O_3$  chemical production compared to observations.  $j_{O1D}$  were
- reasonably represented in the model, with an observation-model discrepancy of 20%, which fell
- within the combined uncertainties of both measurements and modeling. A similar magnitude of
- underestimation, around 20%, was predicted for H<sub>2</sub>O concentrations in the GEOS-Chem. These
- slight model discrepancies in  $O_3$ ,  $j_{O1D}$ , and  $H_2O$  resulted in overall comparable  $O_3$  photolysis rates between observations and GEOS-Chem simulations, with a minor discrepancy of around
- 449 25%.

450 Sensitivity model runs were conducted to examine the combined effects of NMVOCs, OVOCs, and NO<sub>x</sub> on O<sub>3</sub> chemistry (Methods). The  $P_{O3}$ ,  $D_{O3}$  and  $Net_{O3}$  calculated in each 451 sensitivity model were compared with Box\_S0 (box model scenario constrained with 452 observations) and Box GC (box model scenario constrained with GEOS-Chem extractions). In 453 Box\_S1, where observational constraints on NMVOCs were added, little change in  $P_{03}$  was 454 found, compared to Box\_GC, and  $P_{03}$  was still underestimated by the same magnitude (Figure 3). 455 This could be explained by the limited contribution of NMVOCs to peroxy radical production 456 and therefore  $O_3$  chemistry (Figure S8). In contrast, since OVOCs contributed heavily to the 457 peroxy radical production, adding observational constraints on OVOCs (Box\_S2) leads to the 458  $P_{O3}$  underestimation decreased to 50% (Figure 3a) compared to Box\_S0, and net O<sub>3</sub> production 459 is predicted (Figure 3c). When observational constraints on NO<sub>x</sub> were further added (Box\_S3), 460  $P_{03}$  underestimation decreased to 20% (Figure 3a). Comparing Box\_S3 to Box\_S0,  $P_{03}$  and  $D_{03}$ 461 are underestimated by 20%, which could be explained by either the model-observation 462 discrepancy between  $O_3$ ,  $j_{OID}$ , and water vapor as illustrated above. Overall, the sensitivity 463 models revealed that NO<sub>x</sub> simulation bias was the major contributor to the difference between 464 GEOS-Chem predicted and observed  $O_3$  chemical budgets, followed by OVOCs. 465



466

Figure 3. Daytime O<sub>3</sub> chemical budget calculated in each model scenario. (a) O<sub>3</sub> production 467 rate  $(P_{03})$ . (b) O<sub>3</sub> destruction rate  $(D_{03})$ . (c) Net O<sub>3</sub> production rate  $(Net_{03})$ . In each panel, Obs. 468 represents the box model constrained with observations at Nam Co; Box GC represents the box 469 model constrained with GEOS-Chem model-extracted parameters related to O<sub>3</sub> chemistry; Box 470 S1-S3 represent the sensitivity model run for NMVOCs (Box S1), OVOCs (Box S2), and NO<sub>x</sub> 471 (Box S3), respectively. The horizontal solid line in panels (a) and (b) represents the mean value 472 of  $P_{0,3}$  and  $D_{0,3}$  in Obs., respectively. The horizontal solid line in panel (c) represents the zero net 473 ozone chemical production. 474

Several factors may account for the model bias in  $NO_x$  and OVOCs, including poorly 475 characterized emission inventories, incomplete model representation of their chemical and 476 physical processes, such as inaccurate parameterization of soil-to-atmosphere exchange, and 477 incomplete chemical mechanisms. Emission inventory handling was ruled out as the primary 478 source of GEOS-Chem's underestimation of these O<sub>3</sub> precursors. Observations of low 479 concentrations of anthropogenic and biogenic emission tracers confirm minimal local emissions 480 at Nam Co, though emissions from upwind regions such as South Asia could potentially exert an 481 influence. To investigate this, we conducted a sensitivity analysis by setting the  $NO_x$  emissions 482 in South Asia to zero. As shown in Figure S9, this adjustment caused no significant change in 483  $NO_x$  output for the Nam Co grid. An underlying reason is that the reactive  $O_3$  precursors are 484 unlikely to survive the long-range transport from South Asia to the central Tibetan Plateau, 485 compared to less reactive ones such as CO and water vapor (Fu et al., 2006). These results 486 indicate the model-to-observation discrepancies arose from something inherently wrong in the 487 model, rather than simply a result of poorly characterized emissions. 488

The inaccurate soil emission parameterization scheme in the GEOS-Chem could be a 489 possible reason for the model underestimation. Previous comparisons of soil NO<sub>x</sub> emissions 490 estimated by the BDSNP scheme and observations indicated the parametrization scheme tended 491 to underestimate the soil NO<sub>x</sub> emissions (Lu et al., 2021). Soil emissions of HONO are another 492 projected source of NO<sub>x</sub>, which is not included in the current model scheme of soil emissions 493 (Hudman et al., 2012). Emitted HONO quickly photolyzes to produce NO<sub>x</sub> via the internal cycle 494 mechanism between HONO and NO<sub>x</sub> in the atmosphere (*Wang et al.*, 2023). The comparison of 495 modeled and observed HONO concentrations suggests GEOS-Chem underestimated HONO 496 concentrations by around one order of magnitude (Figure S7 f), which supports missing HONO 497 sources in GEOS-Chem such as soil emissions. As for OVOCs, missing emission sources from 498 surface or plants might be attributed to the model bias of acetaldehyde and formaldehyde (Travis 499 et al., 2020; Wang et al., 2019). In NMC observations, soil emission fluxes of formaldehyde 500 were observed with distinct diurnal patterns following those in concentrations (not shown). 501 Nevertheless, the OVOC emission included in GEOS-Chem for the Nam Co is negligible, which 502 might account for the model underestimation of the OVOCs (Guenther et al., 2012). For 503 chemical mechanisms, omitting the photolysis of surface-absorbed nitrates also serves as a 504 505 feasible reason for the NO<sub>x</sub> underestimation in the GEOS-Chem (Ye et al., 2023). Production from the photolysis of organic aerosols may provide a source for OVOCs (Wang et al., 2019). 506 Measurement constraints are necessary for optimizing these parameterizations and chemical 507 508 mechanisms and their kinetics.

509 **4 Conclusions and outlook** 

This study evaluated the simulation of  $O_3$  chemical budgets in the GEOS-Chem model, 510 using observations from a remote site on the Tibetan Plateau as a benchmark. Significant 511 discrepancies were identified between observations and the GEOS-Chem simulations, with the 512 model underestimating O<sub>3</sub> production rates by 60%. This model bias was primarily attributed to 513 the inadequate representation of  $O_3$  precursors in the models, rather than inaccuracies in 514 chemical mechanisms regarding the  $O_3$  production. Specifically, both NO<sub>x</sub> and OVOCs were 515 underestimated by GEOS-Chem, with biases ranging from 40% to 60%, consequently leading to 516 517 the model underestimation of O<sub>3</sub> chemical production.

Model underestimations in  $NO_x$  and OVOCs have been widely reported across various 518 background environments, including the marine troposphere and polar regions(Huang et al., 519 2017; Nicely et al., 2017; Shah et al., 2023). Consistent with these studies, our results highlight 520 that chemical transport models generally struggle to capture these O<sub>3</sub> precursors, which results in 521 522 systematic biases in  $O_3$  chemical production. These findings underscore the current knowledge gap in the understanding of the atmospheric fate of NO<sub>x</sub> and OVOCs and their contributions to 523 the tropospheric  $O_3$  chemical budget. We hope this study motivates further observational 524 constraints of the O<sub>3</sub> budget in remote regions, incorporating comprehensive budget closure 525 analysis of reactive nitrogen and OVOCs. 526

# 527 Acknowledgments

The authors appreciate the assistance and contributions of all the researchers and staff involved in the @Tibet 2019 field campaign. This work was supported by the Ministry of Science and Technology of China (Grant No. 2022YFC3701104), Natural Science Foundation of Tianjin City

- 531 (Grant No. 22JCYBJC01330), and the National Natural Science Foundation of China (Grant No.
- 42375088). The authors declare no financial conflicts of interest.

#### 533 **Open Research**

- 534 Observational data and model results used in this study are available at the Peking University
- Open Research Data website (*Wang*, 2024). F0AM box model framework and codes are
- available at (*Wolfe et al.*, 2016) and <u>https://sites.google.com/site/wolfegm/models</u>. MCM
- 537 mechanisms are available via https://mcm.york.ac.uk/MCM/. The GEOS-Chem model used in
- this study is available via Zenodo (*Yantosca et al.*, 2022).
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Figure 1.





Figure 2.









Figure 3.

	2			
Pog (pbbv h <sup>-1</sup> )			(a)	
	1			
	0			
D <sub>03</sub> (ppbv h <sup>-1</sup> )	1.5		(b)	
	1.0			
	0.5			
	0.0			
(L'H Nd Q	2		(c)	
	1			
et G C	0			
Ž	-1			

![](_page_24_Picture_3.jpeg)