



Cite this: DOI: 10.1039/d4ea00141a

Key role of short-lived halogens on global atmospheric oxidation during historical periods†

Adriana Bossolasco,^{ab} Rafael P. Fernandez,^{cd} Qinyi Li,^e Anoop S. Mahajan,^f Julián Villamayor,^a Javier A. Barrera,^{gh} Dwayne E. Heard,^{id i} Carlos A. Cuevas,^{id a} Cyril Caram,^j Sophie Szopa^j and Alfonso Saiz-Lopez^{id *a}

Atmospheric oxidation largely determines the abundance and lifetime of short-lived climate forcers like methane, ozone and aerosols, as well as the removal of pollutants from the atmosphere. Hydroxyl, nitrate and chlorine radicals (OH, NO₃ and Cl), together with ozone (O₃), are the main atmospheric oxidants. Short-lived halogens (SLH) affect the concentrations of these oxidants, either through direct chemical reactions or indirectly by perturbing their main sources and sinks. However, the effect of SLH on the combined abundance of global oxidants during historical periods remains unquantified and is not accounted for in air quality and climate models. Here, we employ a state-of-the-art chemistry–climate model to comprehensively assess the role of SLH on atmospheric oxidation under both pre-industrial (PI) and present-day (PD) conditions. Our results show a substantial reduction in present-day atmospheric oxidation caused by the SLH-driven combined reduction in the global boundary layer levels of OH (16%), NO₃ (38%) and ozone (26%), which is not compensated by the pronounced increase in Cl (2632%). These global differences in atmospheric oxidants show large spatial heterogeneity due to the variability in SLH emissions and their nonlinear chemical interactions with anthropogenic pollution. Remarkably, we find that the effect of SLH was more pronounced in the pristine PI atmosphere, where a quarter (OH: –25%) and half (NO₃: –49%) of the boundary layer concentration of the main daytime and nighttime atmospheric oxidants, respectively, were controlled by SLH chemistry. The lack of inclusion of the substantial SLH-mediated reduction in global atmospheric oxidation in models may lead to significant errors in calculations of atmospheric oxidation capacity, and the concentrations and trends of short-lived climate forcers and pollutants, both historically and at present.

Received 15th October 2024
Accepted 16th February 2025

DOI: 10.1039/d4ea00141a

rsc.li/esatmospheres

Environmental significance

In this work, we used a state-of-the-art chemistry–climate model to evaluate the effect of short-lived halogens (SLH) on the combined abundance of global atmospheric oxidants in the pre-industrial (PI) and present-day (PD) atmospheres. The results show a significant global reduction in the PD levels of hydroxyl (OH), nitrate (NO₃), and ozone (O₃) concentrations. The simulations also show that the role of halogens on atmospheric oxidation was more prominent in the pristine pre-industrial atmosphere where a quarter of OH and half of NO₃ concentrations in the boundary layer are accounted for by SLH. We conclude that atmospheric oxidation, particularly in the preindustrial atmosphere, cannot be fully understood without consideration of SLH emissions and chemistry.

1 Introduction

The global abundance of atmospheric oxidants determines the lifetime of natural and anthropogenic chemicals emitted into

the atmosphere. Hence, understanding the burden and budget of atmospheric oxidants is critical for tackling air quality and climate change.^{1,2} Indeed, the interplay of key atmospheric oxidants and other important reactive compounds such as

^aDepartment of Atmospheric Chemistry and Climate, Institute of Physical Chemistry Blas Cabrera, CSIC, Madrid, Spain. E-mail: a.saiz@csic.es

^bPhysics Institute of Northwest Argentina, National Research Council (INFINOA-CONICET), National University of Tucumán (UNT), Tucumán, Argentina

^cInstitute for Interdisciplinary Science, National Research Council (ICB-CONICET), Mendoza, Argentina

^dSchool of Natural Sciences (FCEN), National University of Cuyo (UNCuyo), Mendoza, Argentina

^eEnvironment Research Institute, Shandong University, Qingdao, China

^fIndian Institute of Tropical Meteorology, Ministry of Earth Sciences, Pune, India

^gResearch Institute for Physical Chemistry of Córdoba, National Research Council (INFIQC-CONICET), Córdoba, Argentina

^hDepartment of Physical Chemistry, School of Chemical Sciences, National University of Córdoba (UNC), Córdoba, Argentina

ⁱSchool of Chemistry, University of Leeds, Leeds, LS2 9JT, UK

^jLaboratoire des Sciences du Climat et de l'Environnement, LSCE/IPSL, CEA-CNRS-UVSQ, Université Paris-Saclay, Gif-sur-Yvette, France

† Electronic supplementary information (ESI) available. See DOI: <https://doi.org/10.1039/d4ea00141a>



carbon monoxide (CO), methane (CH₄), and volatile organic compounds (VOCs) determines a large fraction of the tropospheric composition.^{3,4} The hydroxyl radical (OH) is the principal oxidant in the atmosphere.⁵ However, other tropospheric oxidants like ozone (O₃), nitrate (NO₃), and chlorine radicals (Cl) also influence the concentrations of crucial atmospheric components.^{5–10} By improving our quantitative understanding of the chemical processes regulating the sources and sinks of the main atmospheric oxidants, uncertainties in long-term trends and interannual variability of other compounds should thus decrease.^{11,12}

Accurately estimating the levels of these oxidants remains a significant challenge due to their short lifetime and the many factors influencing their production and loss; indeed, measuring these oxidants directly is difficult at large scales.¹³ Therefore, chemistry–climate models (CCMs) are valuable tools for understanding their temporal trends and spatial variability, although several uncertainties and shortcomings remain.^{14–16} This highlights the need to improve our understanding of the main chemical processes affecting the levels of these oxidants in the atmosphere.^{11,17,18} One factor attracting growing attention is the complex role of short-lived halogens (SLH) emitted from both natural and anthropogenic sources, as they play a critical role in tropospheric chemistry.^{8,19–24} SLH change the lifetime of methane, the HO_x (hydrogen oxides – OH and HO₂) and NO_x (nitrogen oxides – NO and NO₂) ratios, atmospheric mercury oxidation, and aerosol formation.^{8,20,25,26} SLH also lead to the depletion of tropospheric O₃,^{21,27} which consequently impacts the primary production of OH and NO₃.^{28,29} SLH represent a large source of Cl radicals in the atmosphere.³⁰ Consequently, SLH alter the levels of atmospheric oxidants either directly (for example, through catalytic destruction of O₃) or indirectly (for example, by changing the primary and secondary production of OH). The effects of SLH can be highly variable due to the spatial heterogeneity in emissions and the complex nonlinear atmospheric chemistry of reactive halogens. For instance, they can reduce OH levels in clean marine environments^{31,32} but increase OH production in polluted regions with high NO_x and VOC emissions.^{33,34} Indeed, previous studies have reported global decreases in OH and O₃,^{23,35} while others suggest regional enhancements of OH, O₃, and Cl,^{21,34,36} highlighting the importance of quantifying the counteracting effects of SLH on the production and loss pathways of the main oxidants.

Measurements of Arctic and Alpine ice cores, as well as tree rings in Tibet, show that the emissions of natural iodine-containing SLH have increased by a factor of two to three since pre-industrial times^{37–39} due to a positive feedback mechanism between anthropogenic pollution and oceanic SLH emissions.^{40–42} Following past climate changes, the emissions of SLH have also shown considerable variability on paleo-climate timescales^{43,44} and ice cores also show an increase in bromine and chlorine during the industrial era.^{45,46}

While previous studies have highlighted the role of SLH in specific regions or processes, a quantitative global evaluation of the impact of SLH on the combined abundance of atmospheric oxidants during historical periods has not yet been performed, which adds to the uncertainty in model projections of past and

present climate. To address this critical gap, this study employs a global CCM – the Community Earth System Model (CESM), with a state-of-the-art SLH chemistry mechanism implemented in the Community Atmospheric Model with Chemistry (CAM-Chem) (see Methods). We conducted global simulations for both pre-industrial (PI, representative of the year 1850) and present-day (PD, representative of the year 2000) conditions with (wSLH) and without (noSLH) halogen chemistry and emissions to assess the role played by halogens on altering atmospheric oxidants levels. The results reveal that SLH substantially reduce atmospheric oxidation at present and even more in the pristine pre-industrial atmosphere. The reduction in atmospheric oxidant levels in turn results in an increase in the lifetime and loading of key air quality and climate-relevant chemical compounds in the atmosphere. These findings underscore the importance of considering the impact of SLH on atmospheric oxidants, and their associated implications for past and present air quality and climate, both in pristine and polluted regions, on regional and global scales.

2 Methods

2.1 CESM (CAM-Chem) model set-up and experiments design

The Community Earth System Model (CESM) version 1.1.1, with the Community Atmospheric Model with interactive chemistry (CAM-Chem) version 4.0,⁴⁷ was used to explore the combined impact of SLH on global atmospheric oxidation at pre-industrial (PI) times and present-day (PD). The model extends from the surface to approximately 40 km (3.5 hPa in the upper stratosphere), following a hybrid sigma pressure coordinate with 26 vertical levels and has a horizontal resolution of 1.9° latitude × 2.5° longitude (96 × 144 grid points, respectively). This version of CAM-Chem contains a detailed representation of the tropospheric and stratospheric chemistry, which includes 169 species with both gas-phase and heterogeneous reactions coupled to the radiation module.⁴⁸ Updates for the chemical processing of SLH include a state-of-the-art chemical mechanism for halogens in the troposphere and the stratosphere, which has been described in detail in previous studies.^{20,42,49–51} Further, we updated CAM-Chem with an improved chlorine emission mechanism from the lofting of iron-bearing mineral dust aerosols from the Sahara with sea spray aerosols (van Herpen *et al.*, 2023).³⁰ When these dust aerosols descend into the marine boundary layer (MBL) over the Atlantic, they mix with sea spray aerosol to form Mineral Dust-Sea Spray Aerosols (MDSA). Within MDSA, iron transforms into Fe(III) ions under specific atmospheric conditions, which then react with abundant chloride from sea spray to form Fe(III) ³⁰ chlorides. These Fe(III) chlorides can be activated by sunlight, releasing Cl atoms that can combine to form Cl₂ gas. Additionally, the iron component is regenerated, making the process catalytic.

The MDSA mechanism was included in our model configuration with an extended vertical range beyond 900 hPa (van Herpen *et al.*, 2023).³⁰ Our model showed a MDSA-induced Cl₂ production peaking around 850–900 hPa, with 2.2×10^{-13} kg s⁻¹, decreasing significantly ($<1.5 \times 10^{-13}$ kg s⁻¹) above 800



hPa. This indicates that the MDSA mechanism is most active in the lower troposphere, and its contribution to Cl_2 production decreases significantly at higher altitudes.

Globally, our model results show a boundary layer (BL, 1000–850 hPa) MDSA-induced Cl burden of 15 Tg per year, which is comparable to the 13 Tg per year reported by van Herpen *et al.* (2023).³⁰ This agreement suggests that our extended MDSA parameterization provides a reasonable estimate of global Cl production from this mechanism. While the MDSA mechanism can potentially operate at higher altitudes, the limited availability of observational data and the complexities of atmospheric chemistry at these altitudes limit our ability to accurately parameterize the process. Future research on atmospheric halogens should include improving the representation of MDSA in chemistry climate models.

The natural biogenic sources include nine halocarbons (CHBr_3 , CH_2Br_2 , CH_2BrCl , CHBr_2Cl , CHBrCl_2 , CH_3I , CH_2I_2 , CH_2IBr and CH_2ICl) which are the result of micro- and macroalgae as well as phytoplankton metabolism coupled to photochemistry at the ocean's surface.⁵⁰ Inorganic iodine (HOI and I_2) is directly emitted from the ocean surface following O_3 deposition on seawater and its reactions with aqueous iodide.^{40–42} The current chemical scheme includes the heterogeneous recycling of inorganic halogens reservoirs on SSA (sea-salt aerosols), for example, HOX and XONO_2 ($\text{X} = \text{I}, \text{Br}, \text{Cl}$), which constitute a net source of chlorine and bromine.

Anthropogenic SLH sources are included following an emission inventory of the two dominant organic chlorine species (CH_2Cl_2 and C_2Cl_4),⁵² complemented by lower boundary conditions of other anthropogenic chlorinated substances (CHCl_3 , $\text{C}_2\text{H}_4\text{Cl}_2$ and C_2HCl_3). In the PI simulation set-up, long-lived halocarbon lower boundary conditions (LBC) are zeroed, except for CH_3Cl and CH_3Br for which we assume the same natural contribution as for PD. Further we utilized a contemporary anthropogenic global emission inventory of reactive inorganic halogen species for 2014, adjusted to present-day conditions. This inventory encompassed inorganic chlorine (HCl and particulate chloride) from coal, biomass, and waste combustion, as well as inorganic bromine (HBr and Br_2) and iodine (HI and I_2) from coal combustion, following the methodology established in Saiz-Lopez *et al.*, 2023.²⁰

The standard anthropogenic emissions and biomass burning emissions for pollutants developed for the Chemistry–Climate Intercomparison Project (CCMI) (IPCC 2021) have been used here following Tilmes *et al.*⁴⁸ The aircraft emissions of black carbon and nitrogen dioxide, as well as volcanic emissions of sulfur and sulfate, are vertically distributed. For the present-day CH_4 emission, we have included an emissions-driven approach for CH_4 instead of applying the standard lower boundary surface mixing ratios for long-lived species. The main CH_4 sources include agriculture, landfill, fossil fuel industry, biomass and biofuel burning, and natural emissions from wetlands (see Li *et al.*, 2022 (ref. 25) for further details) representative of year 2020. Biogenic emissions are calculated online throughout the land module using the Model of Emissions of Gases and Aerosols from Nature (MEGAN) version 2.1.⁵³

The CAM-Chem simulations were conducted using a Specified Dynamics (SD) configuration,⁴⁷ nudging the model towards high-frequency meteorological fields from a common year (nominally 2000) cyclically repeated across 20 years (obtained from a previous simulation that omitted the contribution of SLH⁵⁴), ensuring consistent meteorological physics (*e.g.*, transport, hydrology and BL fluxes) between the PI and PD simulations. This approach allows us to isolate the impact of changes in atmospheric composition and chemistry, particularly SLHs, while minimizing the influence of meteorological variability. In addition, the PI and PD simulations include different SST and sea ice conditions representative of each time period, in order to allow the model estimate changes in surface emissions that depend on these parameters. These conditions, also modulate MDSA-induced chlorine (Cl) production and the chemical partitioning between reactive and reservoir chlorine species within the online MDSA mechanism. Despite this, the resulting MDSA-induced Cl production is comparable in both the PI (15 079 Gg) and PD (15 444 Gg) periods.

All experiments were initialized from a previous simulation after allowing 40 years of spin-up to ensure all chemical species, particularly CH_4 , were stabilized. In both PI and PD conditions the long-lived halogens species (*e.g.*, halons, CFCs, *etc.*) are included as LBC based on the A1 halogens scenario from the Scientific Assessment on Ozone Depletion Report.⁵⁵

We conduct four sets of simulations with and without SLH for both periods PI and PD, Table 1 shows the configuration for each case together with the changes in surface flux emissions for the SLH^{Cl}, SLH^I, SLH^{Br} and VOCs, CH_4 and CO. It is important to note that the anthropogenic and biogenic emissions other than SLH are identical within the noSLH, therefore the changes in atmospheric composition (CH_4 , NO_3 , O_3 , OH, *etc.*) between wSLH and noSLH represent the effect of SLH. All the simulations are running for a period of 5 years, 1850–1854 for PI and 2000–2004 for PD. We then used the average values of these 5 years for the analyses of our results. To isolate the effect of SLH emissions between PD and PI, we have used the same chemical mechanisms in both scenarios, therefore the observed differences in the results are primarily attributed to variations in SLH emission and other pollutant (like NO_x) emissions in the different periods, which lead to distinct distributions and transformations of halogenated compounds in each scenario, such as the stronger impact of SLH in PI due to the lower NO_x concentration than in PD.

The percentage of changes of each scenario (wSLH_PI or wSLH_PD) is presented relatively to the corresponding noSLH scenario as follow:

$$\% \Delta(\text{wSLH-noSLH}) = \frac{(\text{wSLH}_X - \text{noSLH}_X)/\text{noSLH}_X}{1} \times 100\%.$$

The global tropospheric OH, O_3 , NO_3 and Cl burden from the different sources and sinks is computed from the surface to the tropopause considering all latitudes (90° N–90° S) and longitudes (0°–360°).



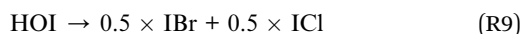
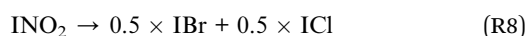
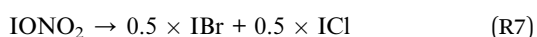
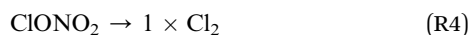
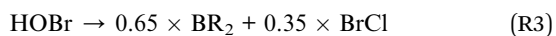
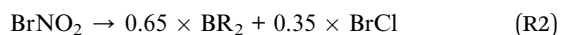
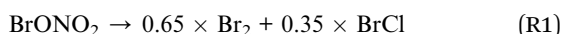
Table 1 Configuration of the simulation scenarios and associated emissions. Configuration of different simulation scenarios, including the surface emissions of SLH and VOCs. Each scenario is identified by a unique case name, and the corresponding emissions are listed in units of gigagrams per year (Gg per year) for SLH and teragrams per year (Tg per year) for VOCs. Additionally, the table is followed by information on boundary layer flux halocarbons due to sea salt recycling, chlorine production from acid displacement and from MDSA

Cases	SLH emissions (Gg per year)							VOCs emission fluxes (Tg per year)			
	Surface flux halocarbons			Sea salt recycling* inorganic halogens		Oceanic emissions	Acid-displacement*	Cl from MDSA*	VOCs	CH ₄	CO
	SLH ^{CL}	SLH ^{Br}	SLH ^I	SLH ^{CL}	SLH ^{Br}	I ₂ /HOI	SLH ^{CL}	Cl ₂			
noSLH_PI ^a	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	782	189	459.5
wSLH_PI ^b	60.9	595.7	586.4	2483	1578	898	2090	15 079	782	189	459.5
noSLH_PD ^c	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	828	499	1142
wSLH_PD ^d	849.7	616.9	591.7	2136	2750	1875	14 214	15 444	828	499	1142

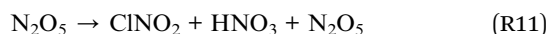
^a noSLH_PI, standard chemical scheme without SLH sources. ^b wSLH_PI, only natural SLH emissions for pre-industrial. ^c noSLH_PD standard chemical scheme without SLH sources. ^d wSLH_PD, natural plus anthropogenic SLH sources from present-day conditions.

The global annual mean surface flux (SF) source strength for SLH is computed as follows: $SLH^{CL} = CH_2Cl_2 + C_2Cl_4 + CH_2BrCl + CH_2ICl + CHBrCl_2 + CHBr_2Cl$; $SLH^{Br} = CHBr_3 + CH_2Br_2 + CH_2BrCl + CH_2IBr + CHBrCl_2 + CHBr_2Cl$; $SLH^I = CH_3I + CH_2I_2 + CH_2ICl + CH_2IBr$. Oceanic emission of inorganic iodine (HOI/I₂) are computed online and depends on near-surface O₃ concentration, wind speed and SST (Prados-Roman *et al.*, 2015 (ref. 42)). Surface flux for the VOCs = ISOPRENE + C₁₀H₁₆ + CH₃OH + C₂H₅OH + CH₂O + CH₃CHO + CH₃COOH + CH₃COCH₃ + HCOOH + C₂H₂ + C₂H₄ + C₂H₆ + C₃H₈ + C₃H₆ + BIGALK + BIGENE + MEK + TOLUENE, including its corresponding biogenic emissions from MEGAN for ISOPRENE + C₁₀H₁₆ + CH₃OH + C₂H₅OH + CH₂O + CH₃CHO + CH₃COOH + CH₃COCH₃ + HCOOH + CO + C₂H₄ + C₂H₆ + C₃H₈ + C₃H₆ + BIGALK + BIGENE + MEK + TOLUENE.

• Halogen heterogeneous recycling at the boundary layer on Seas Salt Aerosols (SSA) are calculated from the following reactions:



• Acid-displacement heterogeneous reactions on SSA:



The total mass of inorganic halogens emitted by each source is computed by considering the halogen atomicity of each species and the halogen mass.

• Cl production from Mineral Dust-Sea Spray Aerosols (MDSA) at the boundary layer.

Note that MDSA and acid displacement only release chlorine. In addition, during PD, values from Cl Acid-displacement are equivalent to values from MDSA. During PI, acid displacement is smaller and MDSA Cl production becomes more important.

2.2 OH recycling probability

The so-called recycling probability “*r*” represents the probability that OH, once formed, is recycled. The calculation of *r* follows the definition of Lelieveld *et al.*, 2002:⁵⁶

$$r = 1 - \frac{P}{P + S} \times 100\% \quad (1)$$

In eqn (1), the sum of primary (*P*) and secondary (*S*) sources of OH (see Table 2) is the “oxidation power”, or the time rate at which OH is produced (gross OH formation (*G*)) *P* and *S* (Tmol per year) was calculated as follows:

$$P \text{ or } S (\text{Tmol per year}) = \text{rate} \times 1 \times 10^{-3} \times \left(\frac{M_w}{\rho \times Na} \right) \times 1 \times 10^3 \times \left(\frac{dp}{\text{grav}} \right) \times \text{factor } 3.15 \times 10^7 \times 1a \times 10^9 \times \frac{1}{M_w} \quad (2)$$

where: rate = is the rate constant in cm³ per molecule s, *M_w* = molecular weight (g mol⁻¹), *ρ* = air density (m³ kg⁻¹), *Na* = 6.022 × 10⁻²³ Avogadro's number, *dp* = pressure layer thicknesses (Pa (kg m⁻¹ s⁻²)). *grav* = gravity (m² s⁻¹). Factor = 2



Table 2 Main reactions leading to the tropospheric production of OH and their relative contribution. The values are shown for both pre-industrial (PI) and present-day (PD) conditions, with and without short-lived halogens (wSLH and noSLH). The reactions are divided into primary (P) and secondary (S) sources of OH. G (gross OH formation) is the sum of P and S. The recycling probability (*r*) in % is calculated as $1 - (P/P + S) \times 100\%$. The inclusion of SLH leads to an increase in secondary OH production, particularly through the photolysis of HOX

Sources (Tmol per year)	noSLH_PD	wSLH_PD	noSLH_PI	wSLH_PI
Primary				
O ¹ D + H ₂ O	87 (47%)	71 (41%)	53 (50%)	41 (43%)
Secondary				
NO _x + HO ₂	59 (31%)	56 (32%)	31 (29%)	28 (29%)
O ₃ + HO ₂	22 (12%)	17 (10%)	11 (11%)	8 (8.5%)
H ₂ O ₂ + <i>hν</i>	14 (8%)	14 (8%)	8 (8%)	8 (8%)
VOCs, ROOH + <i>hν</i>	5.0 (2.7%)	5.3 (3.0%)	3.4 (3.1%)	3.5 (3.6%)
HOX + <i>hν</i>	—	11.0 (6%)	—	7.5 (8%)
Others	0.18 (0.1%)	0.2 (0.09%)	0.08 (0.07%)	0.07 (0.07%)
Total secondary	100 (53%)	103 (59%)	54 (50%)	56 (57%)
G (P + S)	187.2	173.7	107.3	97.02
Recycling probability, <i>r</i> (%)	53	59	50	57

(2 OH radicals from the reaction $\text{O}^1\text{D} + \text{H}_2\text{O} \rightarrow 2\text{OH}$). 3.15×10^7 conversion from seconds to year.

2.3 CESM model evaluation

Our model configuration is based on the well-established CESM1.2 (CAM-Chem) with SLH framework, which has been extensively validated against a wide range of ground, ship, aircraft and satellite-based observations over the past 20 years, particularly in the context of halogen chemistry and key atmospheric species such as O₃, CH₄, and HCl. This is revised in Saiz-Lopez *et al.* (2023)²⁰ based on numerous previous studies of the group.^{21,22,49,50,57,58} Specifically, our model configuration is based in the configuration of Saiz-Lopez *et al.*, 2023 (ref. 20) and incorporates the latest advancements in halogen chemistry, including the MDSA mechanism from van Herpen *et al.*, 2023.³⁰ This allows us to provide a more comprehensive representation of halogen-mediated atmospheric processes and their impact on climate.

Briefly, Ordóñez *et al.* 2012,⁵⁰ validate the natural oceanic sources of short-lived halocarbons within CESM1 (CAM-Chem) with data collected from near-surface and aircraft campaigns conducted in extra-polar regions, incorporating additional observations of reactive short-lived halogens and ozone in tropical regions to further validate the model. Furthermore, Fernandez *et al.*, 2014,⁴⁹ Fernandez *et al.*, 2017 and Saiz-Lopez and Fernandez 2016 (ref. 57 and 59) demonstrated the model's accurate representation of bromine transport and its improvement in the total O₃ column and the Antarctic ozone hole. Barrera *et al.*, 2020 (ref. 54) showed the model's ability to reproduce observed stratospheric halogen levels. Additionally, the implementation of iodine chemistry in CESM (CAM-Chem)⁵¹ allowed reproducing aircraft observations in the tropical upper tropopause, suggesting the occurrence of iodine-driven stratospheric O₃ depletion, which was later confirmed by Koenig *et al.*, 2020.⁶⁰ Prados Roman *et al.*, 2015 (ref. 42) demonstrated the need to consider an inorganic iodine source from the ocean surface to accurately reproduce the observed iodine oxide mixing ratios over the open ocean. Cuevas *et al.*, 2022 (ref. 58) reported the improved model performance of

stratospheric O₃ by including reactive iodine chemistry into the CESM/WACCM4-SD configuration. Furthermore, Li *et al.*, 2022 (ref. 25) validated the model's representation of key atmospheric species, including ozone, methane, as well as for the global sea-salt aerosol abundance in the marine boundary layer compared with global observational results. To summarize, these studies, provide strong evidence for the model's ability to accurately simulate halogen chemistry.

In addition to these core validations, our model incorporates the latest advancements in halogen chemistry, including the Cl atoms' photochemical production from Sahara dust (MDSA mechanism) from van Herpen *et al.*, 2023.³⁰ This mechanism has been initially validated using ¹³C depletion in CO in air samples from Barbados.⁶¹ In the ESI,[†] we added more details about the contribution of Cl production from MDSA and comparison to observation (see Section 1 in the ESI[†]).

Here we provide further evaluation of the model results for OH burdens production (Table S4[†]). OH is a short-lived species, making its direct measurement challenging. However, our model simulations of OH burden are in the range of the reported values using other model simulations.^{9,56} Indeed, our simulated global tropospheric OH concentration in the noSLH_PD and wSLH_PD cases are 1.06×10^6 molec. per cm³ and 9.28×10^5 molec. per cm³, respectively, and are in agreement with the averaged value reported by observation and previous modeling studies $\sim 1.0 \times 10^6$ molec. per cm³ for the 21st century.^{9,12,18} This agreement provides confidence in the model's ability to represent the complex interactions between OH and other atmospheric species. ESI Table S5[†] shows that our tropospheric O₃ and CH₄ budgets in PI and PD with and without reactive halogens are in agreement with the values reported by other model studies.^{27,62}

3 Results and discussion

3.1 Changes in global atmospheric oxidants due to halogens

SLH lead to the release of halogen radicals (*i.e.*, Cl, Br and I atoms and ClO, BrO and IO oxides) on photolysis, which can



react with numerous species in the atmosphere, leading to a change in the O_3 , OH, NO_3 and Cl burdens and their distribution, which in turn affect the lifetime of many key atmospheric constituents. Bromine and iodine predominantly lead to the destruction of O_3 , and this reduction is more pronounced over the oceans, which are the largest bromine- and iodine-SLH sources. Chlorine can also lead to ozone destruction; however, in the presence of VOCs and NO_x , it can also lead to the formation of O_3 . Our results show that SLH reduce tropospheric O_3 by 20% in PD and 21% in PI, respectively (Fig. 1a). The simulations show a larger reduction in ozone in the PD (20%) compared to previous studies^{22,25,27} (12–16%). This difference results from the recent inclusion of additional chlorine sources from dust³⁰ in our model (see Methods for details). During PD, the net integrated tropospheric O_3 burden is reduced from 330 Tg (noSLH) to 260 Tg (wSLH). Despite the total ozone burden being much lower in the PI period (noSLH: 230 Tg), the relative

reduction due to SLH chemistry (wSLH: 180 Tg, 21%) is similar to PD.

O_3 photolysis is the main source of OH in the troposphere⁵⁶ and hence the SLH-driven ozone destruction indirectly reduces the tropospheric OH burden by 13% in the PD (wSLH: 181 Mg and noSLH: 210 Mg; Fig. 1b). This value is larger than the 8.2% decrease reported by Sherwen *et al.*²⁸ and the 4.5% decrease reported by Stone *et al.*³⁵ This is because these studies did not account for sea-salt debromination and dust sources of Cl, which represent a significant fraction of the tropospheric halogen burden (see Table 1). Additionally, the study by Stone *et al.*³⁵ did not include iodine chemistry which is the dominant halogen leading to tropospheric O_3 reduction.²³ The reduction in global OH due to SLH increases from 13% during PD to 17% during PI (*i.e.*, reducing the tropospheric burden from 213 Mg (noSLH) to 177 Mg (wSLH)). The production and loss pathways controlling the OH levels in PI and PD are detailed in the next section.

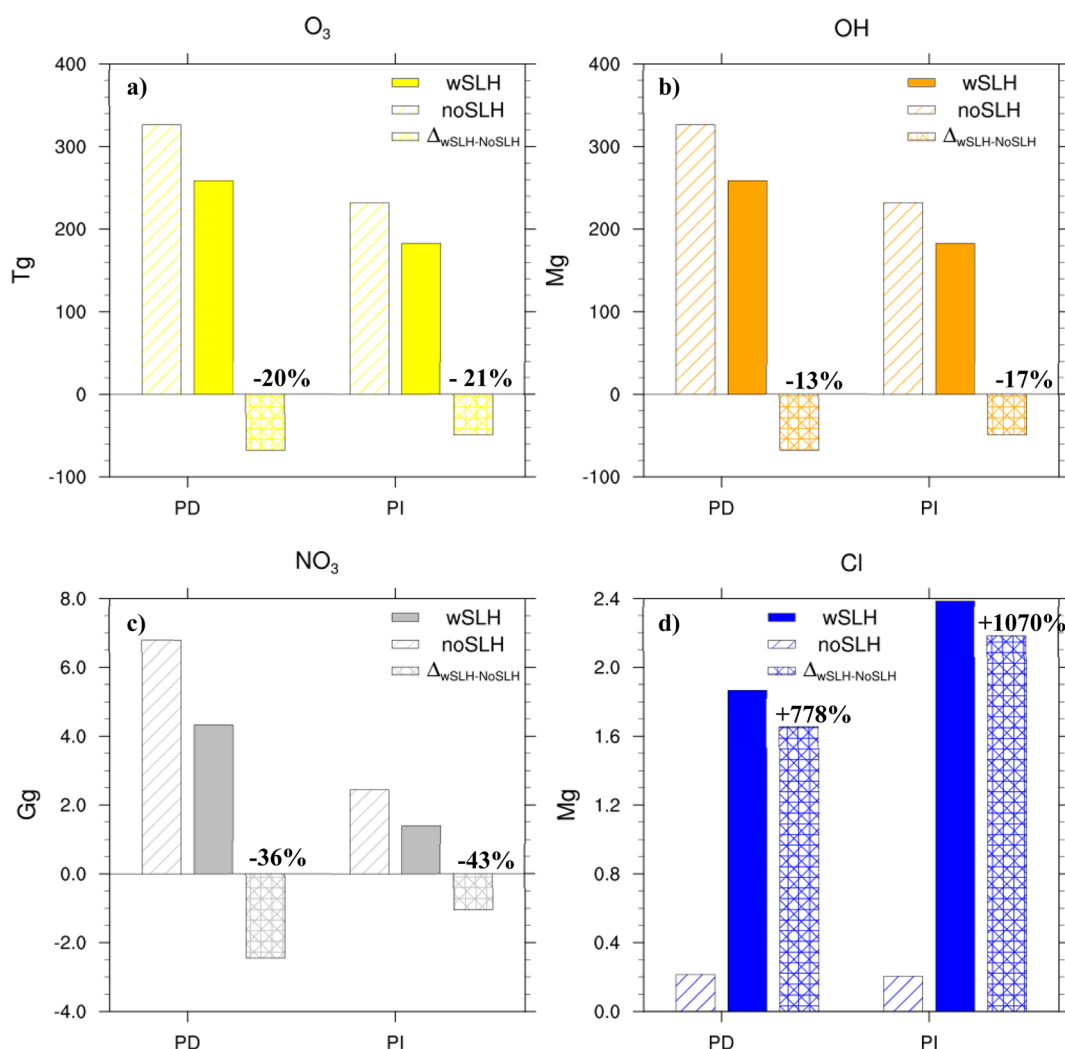
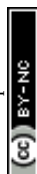


Fig. 1 Global integrated tropospheric burden of main oxidants (OH, O_3 , NO_3 , and Cl) and their relative changes due to the inclusion of SLH in the present-day (PD) and pre-industrial (PI) atmospheres. Filled bars represent the wSLH case, hatched bars the noSLH case, and crosshatch bars the absolute differences between them. The relative percentage differences for PD and PI times are indicated above the bars in bold and are calculated as follow: $\% \Delta_{\text{oxidant}} = [\text{oxidant}]_{\text{wSLH}} - [\text{oxidant}]_{\text{noSLH}} / [\text{oxidant}]_{\text{noSLH}} \times 100\%$. (a) Changes for O_3 (Tg), (b) changes for OH (Mg), (c) changes for NO_3 (Gg) and (d) changes for Cl (Mg).



NO_3 radicals, the main atmospheric oxidant at night, are mainly formed by reaction of ozone with NO_2 , therefore SLH reduce NO_3 formation primarily by depleting ozone and titrating NO_2 to form reservoir halogen nitrates and nitrites.⁸ Similar to O_3 and OH, NO_3 reduction caused by SLH is more pronounced during the PI than in the PD (Fig. 1c), primarily due to slightly larger ozone depletion and lower NO_2 abundance in the pristine preindustrial atmosphere, reducing the titration of reactive SLH to reservoir species. During PD, halogen chemistry reduces the net integrated tropospheric NO_3 burden from 6.78 Gg (noSLH) to 4.33 Gg (wSLH; 36% reduction). While the NO_3 levels are much lower in the PI period (noSLH: 2.44 Gg), as expected due to the absence of anthropogenic NO_x emissions, the relative reduction due to SLH chemistry is larger (wSLH: 1.4 Gg; 43% reduction).

Considering that SLH are a direct source of Cl atoms, an increase in the Cl burden is expected in both PD and PI (Fig. 1d). Integrated throughout the troposphere, the Cl burden increases from 0.21 Mg (noSLH) to 1.87 Mg (wSLH; 778% increase) in the PD, while the increase is much higher in PI going from 0.20 Mg (noSLH) to 2.38 Mg (wSLH; 1070% increase). Note that the small fractions of chlorine present in the noSLH arise from the slow degradation of long-lived chlorocarbons, as some of them such as methyl chloride have natural sources. However, if we consider the total chlorine family (Cl_y), which includes all reactive and reservoir chlorine species (see Methods), the Cl_y tropospheric abundance is larger during PD (from 32 Gg (noSLH) to 302 Gg (wSLH); 833% increase) compared to PI (from 17 Gg (noSLH) to 200 Gg (wSLH); 1047% increase). While both Cl and Cl_y exhibit significant increases in both periods, the relative increase in Cl is larger in the PI compared to the PD simulation. This is due to the coupled interplay between direct emissions of SLHs and the chemical partitioning of chlorine species. This shift in partitioning, driven by higher NO_x and HO_x levels in the PD from anthropogenic emissions⁵⁴ reducing the X/Y ratios ($\text{X} = \text{Cl}, \text{Br}, \text{I}$) (Table S1†), leads to a greater fraction of chlorine being stored in unreactive reservoir species, reducing the available Cl atoms. Thus, despite the larger increase in total Cl emissions in the PD compared to PI (see Table 1), the net effect on Cl abundance is more significant in the PI simulation.

Note that the SLH-mediated percentage changes in oxidant burdens within the BL (1000–850 hPa) are higher than in the integrated troposphere (Fig. S2†). The reductions in oxidants in the BL are OH: –16% and –25%; O_3 : –26% and –32%; NO_3 : –38% and –49% for PD and PI, respectively, while an increase of Cl radicals of 2632% and 3114%, while an increase of Cl radicals of 2632% and 3114% are observed in the PD and PI, respectively.

3.2 Spatial heterogeneity in the influence of SLH on oxidants

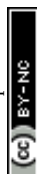
The regional impact of SLH on atmospheric oxidants is highly dependent on natural and anthropogenic emissions and their chemical processing, which depend on the local environment. SLH can lead to both the lowering of OH and O_3 in remote, clean environments and an increase in OH production in polluted

continental regions by increasing O_3 formation through the oxidation of VOCs and forming OH through the photolysis of hypohalous acids (HOX ; $\text{X} = \text{Cl}, \text{Br}, \text{I}$). At a global scale, SLH lead to large decreases in OH, O_3 and NO_3 over the oceans (Fig. 2). Present-day reductions of up to 40% in OH are observed over the Atlantic Ocean and in the Southern Ocean boundary layer. The average OH change over the global oceans is –19% at present. Comparatively, some locations over land show increased OH production due to the SLH reaction with pollutants and HOX photolysis. Therefore, although OH is globally reduced due to the catalytic loss of O_3 by SLH, the decrease is smaller integrated over the land (PD: –7%; Table S2†). This is clearly observed in the northern mid-latitudes, where in some regions, anthropogenic sources of SLH also cause an increase in the secondary production of OH, leading to an increase in OH over the continents (Fig. 2a–c).

The effect of SLH on OH is even more pronounced in the pre-industrial atmosphere. The overall reduction in the global BL over oceans is –29% in the PI (compared to –19% in the PD). This is also true over the land, with a change of –13% (–7% in the PD). SLH lead to the catalytic loss of O_3 throughout the global BL. O_3 losses of up to 45% are observed over the northern Atlantic Ocean in the PD and are more than 50% in certain areas during the PI period (Fig. 2). Averaged in the global BL, the O_3 reduction over the oceans (PD: –29% and PI: –34%) is larger as compared to over the land (PD: –20% and PI: –28%) (Table S2†). The smaller reduction in O_3 over land compared to the ocean can be attributed to several factors, such as higher NO_x emissions, specially over land areas. These relatively higher NO_x levels can titrate reactive halogen species, limiting their potential to deplete ozone. On the other hand, as discussed in Barrera *et al.*, 2023,²⁷ the iodine-ozone feedback mechanism, particularly in the Western Pacific Warm Pool, plays a relevant role in oceanic ozone depletion. This mechanism is less prominent over land due to much lower iodine emissions.

The open, clean ocean regions also show the largest NO_3 relative changes (up to 90%). The reduction is more pronounced in the northern Atlantic and the Southern Oceans in PD, while during the PI period, the largest changes are seen in the Southern, Atlantic and Northern Oceans. The change over the oceans (PD: –44% and PI: –59%) is larger than over the land (PD: –22% and PI: –38%) (Table S2†), similar to OH and O_3 . These results show that the levels and distribution of global NO_3 radicals, cannot be understood without consideration of the crucial impact of SLH chemistry.

In the case of Cl, as expected, the inclusion of SLHs leads to significant increases in Cl concentrations in both the PI and PD simulations, particularly in regions with abundant dust and sea salt aerosols, such as the Atlantic Ocean. Continental regions also exhibit an increase in Cl radical, which is attributed to anthropogenic SLH^{Cl} emissions (Fig. 2j and k). However, while Cl concentrations increase in both the PI and PD periods, the relative increase of atomic Cl is more pronounced in PI. This behavior is primarily attributed to the interplay between SLH emissions and NO_x concentrations. In the PI simulation, lower NO_x levels favor the formation of reactive chlorine species, leading to a more pronounced Cl increase than in PD.



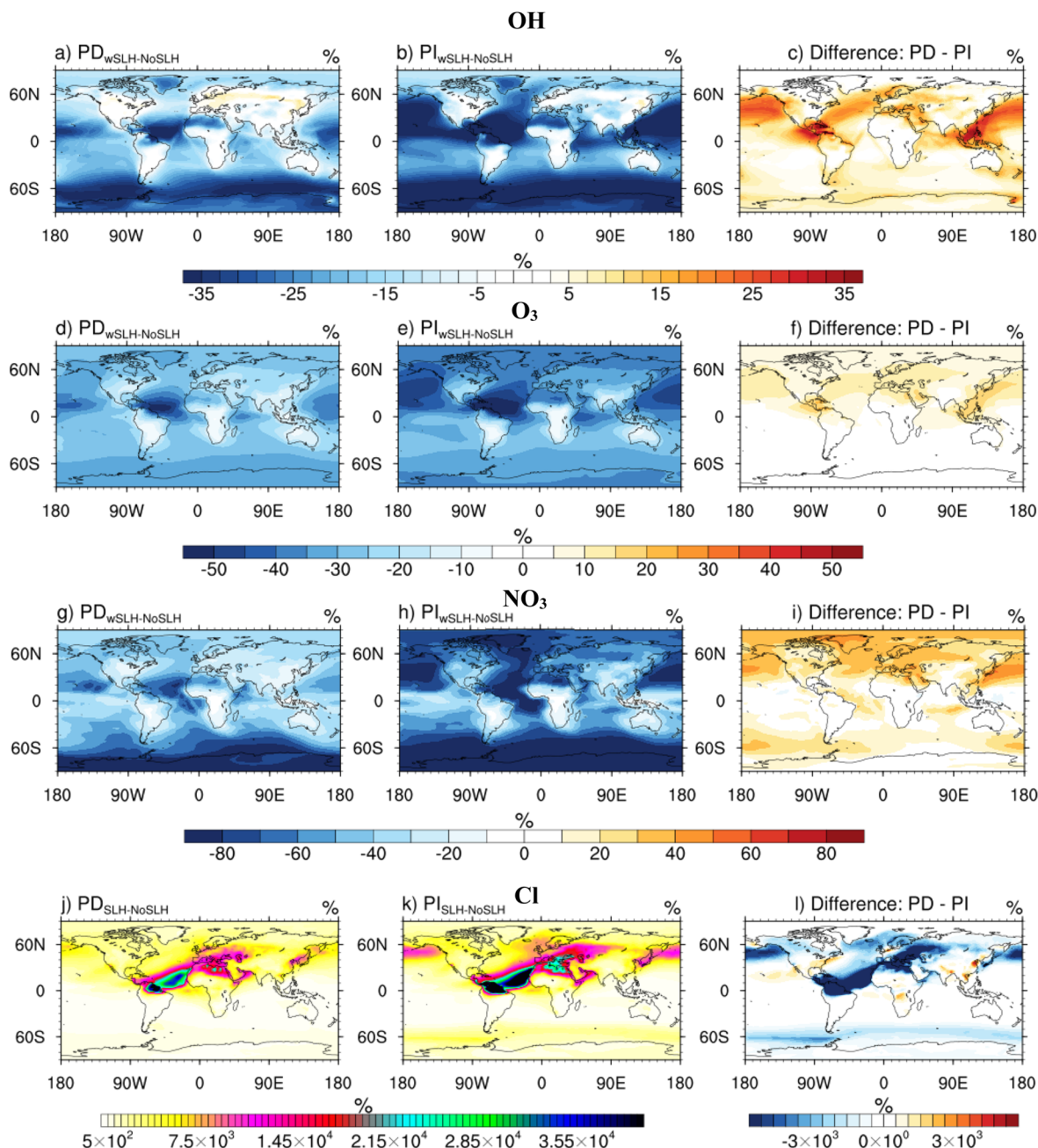


Fig. 2 Spatial distribution of relative changes in oxidant concentrations (OH, O₃, NO₃, and Cl) due to the inclusion of SLH in the boundary layer (1000–850 hPa) for the present-day (PD) and pre-industrial (PI) times. The color scale indicates the percentage change in each oxidant relative to the noSLH case. (a)–(c) Show the relative differences for OH in the PD, PI and the changes from PI to PD (wSLH scenarios); (d)–(f) show the changes for O₃ in the PD, PI and the changes from PI to PD (wSLH scenarios); (g)–(i) show the relative changes for NO₃ in the PD, PI and the changes from PI to PD (wSLH scenarios); and (j)–(l) show the relative changes for Cl radicals in the PD, PI due to SLH and the changes from PI to PD (wSLH scenarios), respectively; note that the color scale for the change in Cl is different due to the percentage changes for the Cl radical are positive and greater than 100%. (Panel l) Shows that the change from PI to PD are negative, indicating that Cl radicals are higher in pre-industrial periods.

Conversely, higher NO_x levels in the PD simulation promote the formation of reservoir species like ClNO₂ and ClONO₂, reducing the active Cl budget. Consequently, despite the potential for increased Cl production due to SLH emissions, the dominant NO_x-driven titration reactions ultimately result in lower Cl concentrations in the PD compared to the PI (Table S1†).

Vertically, the effect of SLH in OH, O₃ and NO₃ is highest in the BL and reduces gradually up to 300 hPa (Fig. 3). Large effects are seen in the polar regions, where emissions of SLH peak, particularly in springtime. Above 300 hPa, the effects of SLH are more marked in the equatorial regions, where large-scale convective systems transport SLH to the upper troposphere



(Fig. 3). The effect of SLH on Cl differs from OH, O₃ and NO₃, peaking in the lower tropical troposphere, which arise mainly from dust-induced Cl production,³⁰ aerosol recycling,⁶³ and biomass burning.^{64,65} In the upper troposphere, Cl is elevated by the photodegradation of relatively longer-lived anthropogenic SLH^{Cl}, like dichloromethane and chloroform^{24,52} (Fig. 3).

Due to the large spatial heterogeneity in the effect of SLH on oxidants, the resulting atmospheric oxidation capacity shows differences on regional scales. Li *et al.*³⁴ have reported an increase in atmospheric oxidation due to SLH over China, particularly during the daytime due to enhanced OH and Cl

radical production. This is also consistent with Dai *et al.*,⁶⁶ who found that OH reactions with VOCs and CO account for most daytime oxidation capacity and the study by Chen *et al.*,⁶⁷ who reported an increase in oxidation due to halogens in the Yangtze River Delta, with OH increasing by up to 16% and O₃ increasing by 2%. Our results are consistent with this spatial heterogeneity while showing a global decrease in the combined concentration of atmospheric oxidants.

These findings underscore the spatial heterogeneity in the halogen-induced oxidation changes, showing that while regional models in polluted areas consistently indicate oxidation enhancements, the effects can be contrasting between different pristine, semi-polluted and polluted environments. This highlights the complexity and the need to incorporate detailed halogen emissions and chemistry in climate and air quality models to more accurately capture atmospheric oxidation from the regional to the global scale.

3.3 Influence of halogens on the chemical pathways for OH formation

Given OH dominates the oxidation of VOCs and pollutants in the troposphere, now we turn to quantify the SLH-driven changes on its major sources and sinks. The photolysis of O₃ at wavelengths lower than 330 nm leads to the formation of O¹D, which then reacts with water vapour to form OH. This is known as the primary production of OH (*P*). In addition, several reactions lead to the formation of OH through subsequent reactions involving VOC_s and NO_x, known as secondary sources of OH (*S*).⁹ Secondary production can in turn be subdivided into contributions by the so-called NO_x mechanism (NO_x + HO₂), the O_x mechanism (O₃ + HO₂ and H₂O₂ + *hν*), the recycling of OH by the chemistry of VOCs, OVOCs and ROOH, as well as the OH production due to the photolysis of HOX (Table 2).

In the absence of SLH, present-day *P* accounts for 87 Tmol per year (noSLH) of the OH production, similar to Lelieveld *et al.*⁹ (84 Tmol per year). In the absence of anthropogenic ozone pollution, *P* accounts for 53 Tmol per year (noSLH) in the PI atmosphere. When SLH are included, tropospheric *P* decreases by 18% in the PD and 22% in the PI, while total *S* remains almost identical (+3% and +4%, respectively; see Table 2).

Regionally, the most substantial reductions in *P* due to SLH occur in the tropics, where *P* dominates the total OH production (Fig. 4a and c). *P* is also the dominant OH source up to approximately 500 hPa in the absence of SLH. When SLH are included, *P* still remains dominant up to 600 hPa, beyond which *S* becomes more important (Fig. S3†). Fig. S4† presents an expanded analysis of changes in *P* and *S* across the selected regions (Polluted Regions: China (20–40° N, 100–125° E); Semi-Polluted Regions: South America (10–30° S, 70–40° W); Clean Regions: Open Ocean (0–30° N, 180–120° W)), showing that in clean environments, the reaction O¹D + H₂O account for ~74% of the OH formation.

S slightly increases from 100 Tmol per year (noSLH) to 103 Tmol per year (wSLH) in the PD, presenting an almost identical vertical profile for the different sensitivities (Fig. S2†). The photolysis of HOX contributes approximately 6% to boundary

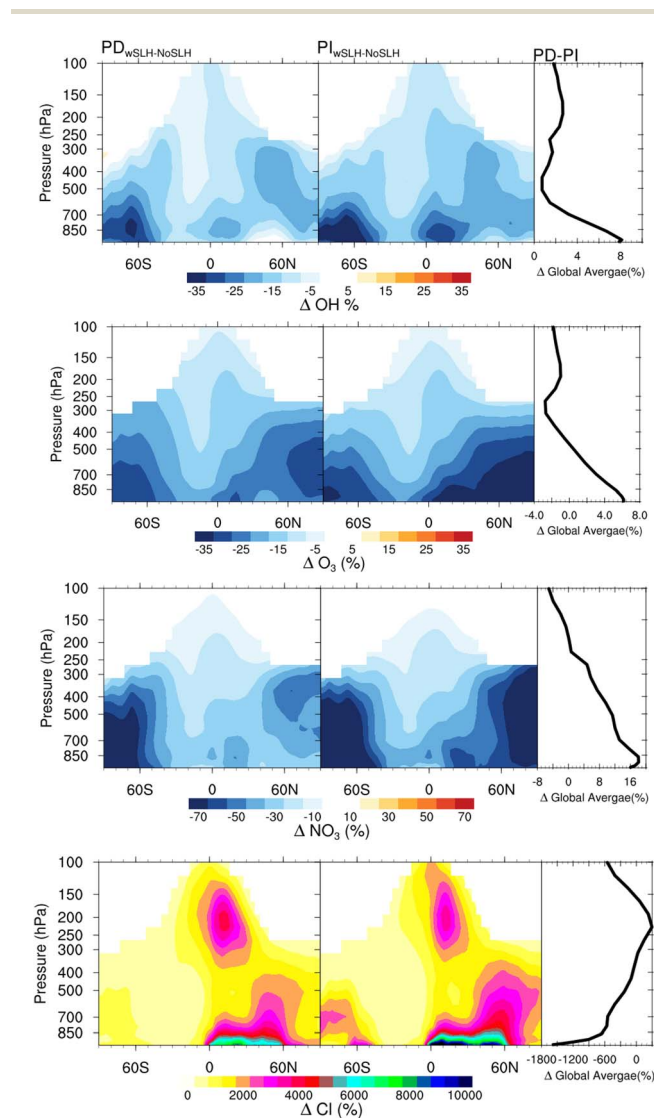


Fig. 3 Vertical distribution of relative changes in oxidant concentrations (OH, O₃, NO₃, and Cl) due to the inclusion of SLH in the present-day (PD) and pre-industrial (PI) times. The color scale indicates the percentage change in each oxidant relative to the noSLH case. Panels from top to bottom shows the relative change for OH, O₃, NO₃ and Cl. Decreases are observed for OH, O₃ and NO₃, but Cl radicals show a large increase throughout the vertical column. Note that the color scale applied for ΔCl (%) is different. The vertical profile Δglobal average (%) for the last panel is calculate as the difference between the global average in the PD_{wSLH-noSLH} and PI_{wSLH-noSLH}.



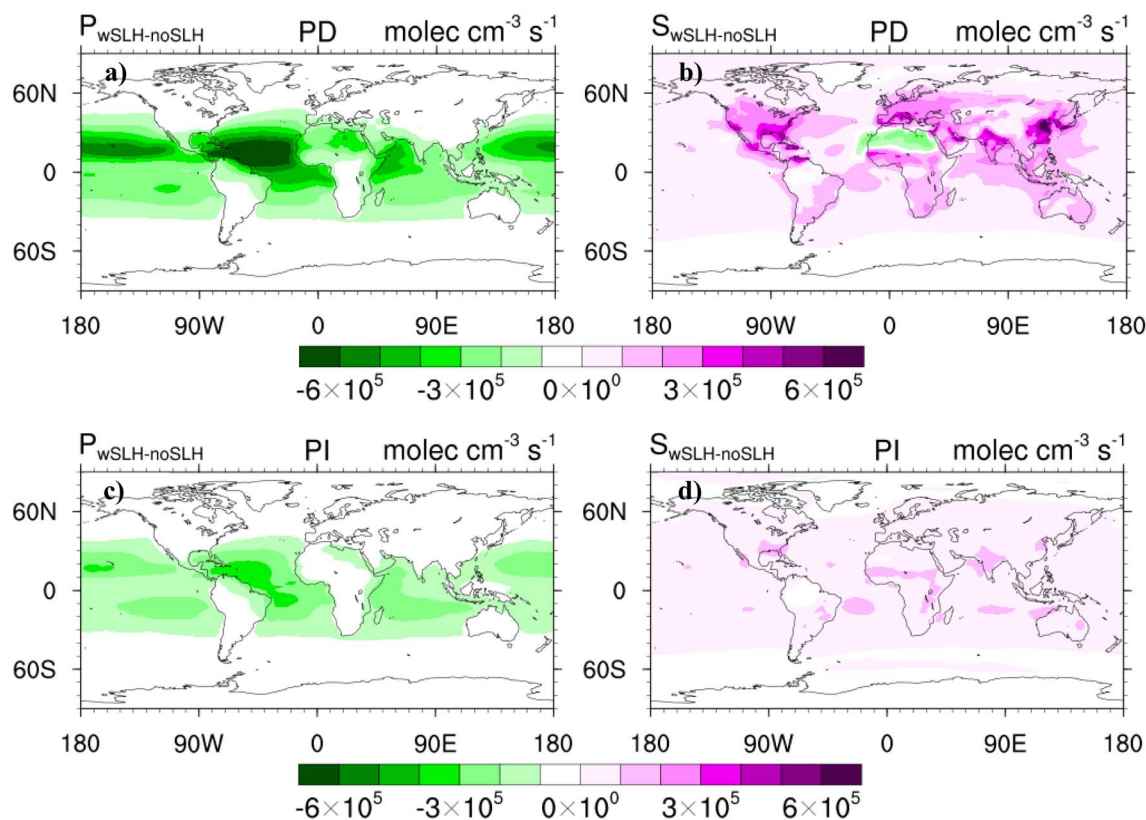


Fig. 4 Spatial distribution of changes in primary (P) and secondary (S) OH production (molecules $\text{cm}^{-3} \text{s}^{-1}$) due to the inclusion of SLH in the boundary layer (1000–850 hPa) for the present-day (PD) and pre-industrial (PI) times. (a) and (b) Shows the changes in the primary and secondary OH production due to SLH for present day, (c) and (d) shows the changes in the primary and secondary OH production due to SLH for pre-industrial times. $P_{\text{wSLH-noSLH}} = P(\text{wSLH})^{\text{PD/PI}} - P(\text{noSLH})^{\text{PD/PI}}$. $S_{\text{wSLH-noSLH}} = S(\text{wSLH})^{\text{PD/PI}} - S(\text{noSLH})^{\text{PD/PI}}$. Note that SLH lead to a decrease in primary OH production in both PD and PI, with more pronounced effects in the tropical and subtropical regions, while increase secondary OH production, particularly through the photolysis of HOx.

layer S in the PD (Table 2), increasing to 14% in the upper troposphere (Fig. S5†). S values in the noSLH case (100 Tmol per year) are in agreement with a previous study by Lelieveld *et al.*⁵⁶ (96 Tmol per year) but lower than in a later study by Lelieveld *et al.*⁹ (167.2 Tmol per year), who used higher VOC emissions and a faster VOC degradation mechanism (Table S2†). Fig. S4† shows that in these clean regions, SLHs contribute 12% to total secondary OH sources. In polluted regions like China, with high NO_x concentrations, the impact of SLHs on secondary OH production, and therefore on OH recycling, is less pronounced. The primary mechanism contributing to OH formation is $\text{NO}_x + \text{HO}_2$. Because NO_x can titrate reactive halogen species, their contribution to OH production is limited in these regions. However, SLHs can still modulate OH recycling by influencing the balance between primary and secondary OH sources.

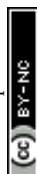
In the pre-industrial atmosphere, S increases from 54 Tmol per year (noSLH) to 56 Tmol per year, with a similar contribution from HOx photolysis of 8%. Note that several models with different levels of complexity have been used to study how halogen chemistry influences HOx. For instance, Whalley *et al.*⁶⁸ showed that HOI and HOBr photolysis contributed roughly 13% to instantaneous OH formation in the tropical Atlantic BL, while Sommariva *et al.*⁶⁹ calculated an increase of up to 15% in OH due to HOI photolysis, which are comparable to our simulations.

The spatial differences in P and S due to SLH in the boundary layer show that P mostly decreases in the tropical and extra-tropical marine regions. At the same time, increases in S are seen predominantly over the continents (Fig. 4b and d). The differences due to the inclusion of SLH are larger in the PD, showing that inclusion of halogen chemistry reduces the dependence of OH abundance on primary OH production, increasing the relative contribution of secondary sources. At present, S consistently exceeds P near the surface in the presence of SLH, but in the pre-industrial atmosphere, S shows lower values due to lower NO_x and lack of anthropogenic SLH emissions (Fig. S3†).

Overall, our results highlight the complex interplay between SLHs, NO_x , and other atmospheric constituents in shaping the oxidizing capacity of the atmosphere, which presents different behaviors between polluted and pristine environments. By understanding the regional variations in these interactions, we can better assess the impact of SLHs on climate and air quality.

3.4 Role of halogen chemistry on OH recycling in the atmosphere

The OH recycling probability (r)⁵⁶ refers to the ability of OH to be regenerated from secondary sources. It is calculated as $r = (1 - P/$



$(P + S) \times 100$. When secondary production exceeds primary production ($S > P$), the recycling probability surpasses 50%. Lelieveld *et al.*⁹ suggested that the atmospheric chemical system is buffered when r approaches 60%. Our calculations show that SLH enhance tropospheric OH recycling, with global r increasing from 53% to 59% at present (50% to 57% in PI) (Table 2).

Fig. 5 shows that substantial regional differences in r are observed, with the tropical BL showing values of $r < 50\%$, as this region strongly depends on P (more sensitive to O_3 photolysis and water vapour) (see also Fig. S2 and S4† for vertical distribution). When SLH are incorporated, r increases (due to a significant reduction of P). In the northern mid-latitudes, halogens increase r values $> 60\%$ near the surface (Fig. S6†), with r values reaching up to 60% over the North Atlantic Ocean, where the photocatalytic chlorine production from Saharan dust is highest (Fig. 2). Southern mid-latitudes have lower r values due to lower NO_x and SLH levels. Consequently, r values are below 60% (Fig. 5 and S6†), with both P and S lower as compared to the Northern Hemisphere (Fig. S3†).

SLH significantly enhance the global OH recycling from 4–6% in continental region to 12–16% in the open ocean (Fig. 5). Pre-industrial r over oceans increases by 14% when SLH are included (Fig. S7†).

The impact of SLH on OH recycling is more pronounced in the PI, especially in the tropics (Fig. 5) due to the lower contribution of other secondary sources besides the HOX production channel (Table 2). This demonstrates the significant but largely heterogeneous role of SLH in modulating OH recycling efficiency and, therefore, modulating the chemical pathways for OH formation in the atmosphere.

Our model results show that the reduction in P , as well as the corresponding enhancement of S are more pronounced in PD

than in PI, as a consequence of nonlinear chemical interactions with anthropogenic pollution. Therefore, SLH not only reduce OH levels but also increase the r , and as a consequence SLH not only reduces the oxidation capacity of the global atmosphere but also represents an additional coupling route between different oxidants that buffer the response of main oxidants to the changes in the background composition of the troposphere. In other words, SLH increase the natural resilience of the atmosphere to external perturbations affecting the chemical composition of the atmosphere.

3.5 Impact of SLH on the abundance of key organic compounds

The overall SLH-driven reduction of oxidant levels has important implications for the atmospheric degradation of many organic compounds. While SLH generally decreases OH, it also increases Cl. Therefore, since many VOCs have different reaction rates towards oxidation by OH and Cl atoms, we investigated how the addition of SLH along with the difference in the reaction rates between VOCs with OH and Cl leads to positive/negative changes in the burdens of different VOCs.

Our results show that the general reduction in OH levels caused by SLH increases the boundary layer burden of CO by 14% and 21%, CH_4 by 9% and 13%, isoprene by 7% and 9%, propene (C_3H_6) by 4% and 17% and limonene ($C_{10}H_{16}$) by 11 and 13%, during PD and PI, respectively (Fig. 6a and S6a†). These changes in organic compound burdens have important implications for air quality and climate. For instance, the increased burden of methane, a potent greenhouse gas, contributes to enhanced radiative forcing. Additionally, the

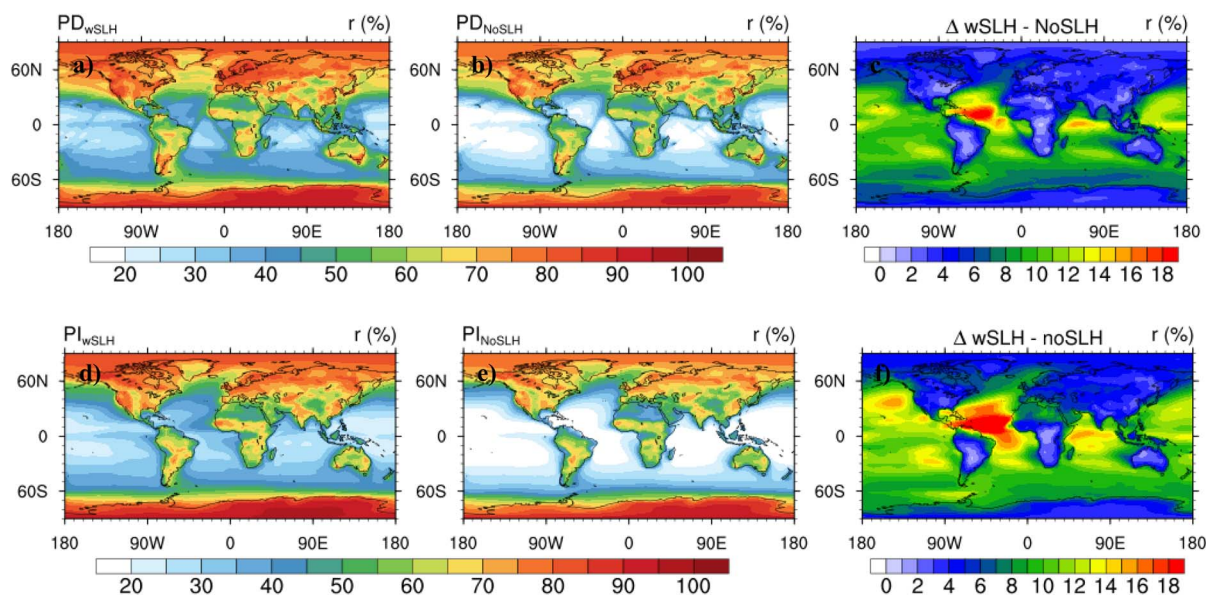


Fig. 5 Spatial distribution of the recycling probability (r , %) for OH average at the boundary layer (BL, 1000–850 hPa) in both scenarios wSLH and noSLH for present-day (PD) and pre-industrial (PI) times. r was calculated as $(1 - P/P + S) \times 100\%$, where P and S refers to the primary and secondary OH production. (a) r values for wSLH case at PD, (b) r values for noSLH case at PD (c) shows the differences between panel (a) and (b), i.e. the increase in r due to SLH. panel (d)–(f) are equivalent to panel (a)–(c) but for PI.



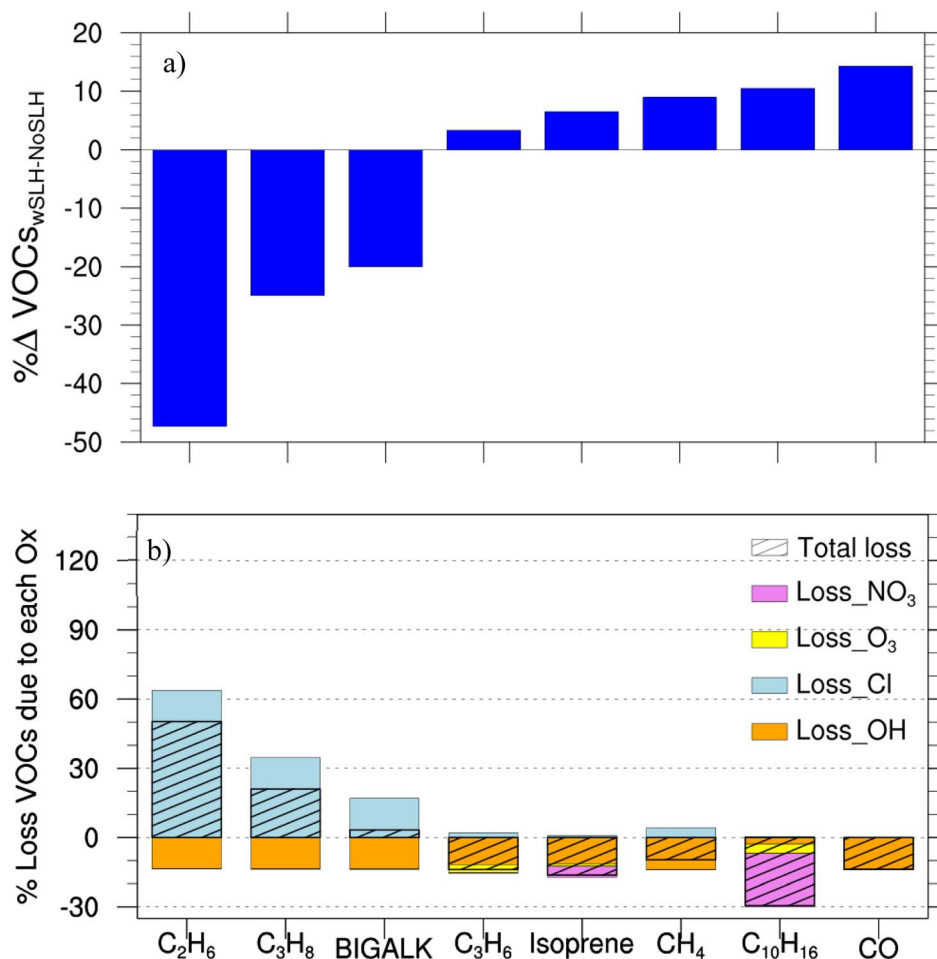


Fig. 6 Percentage of changes in the concentration of some principal VOCs due to the inclusion of SLH for present-day (PD) and the percentage contribution of each oxidant to VOC loss at the boundary layer (1000–850 hPa). (a) Percentage changes in VOC concentrations due to SLH. (b) The stacked bar chart shows the percentage contribution of each oxidant (OH, Cl, O₃ and NO₃) due to SLH to the total loss of VOCs. The calculations are based on the rate constants for the reaction of each VOC with the different oxidants and its relative change due to SLH, and was calculated as follows.

increased burden of isoprene can lead to increased secondary organic aerosol formation, affecting air quality and climate.

The halogen-mediated increase in CH₄ burden is primarily driven by the indirect effect of reduced OH concentrations, which outweighs the direct impact of Cl-atom oxidation. This finding is consistent with previous studies,^{22,25} which have also highlighted the importance of halogen chemistry in regulating atmospheric methane levels. Our results further quantify the magnitude of this effect, showing that wSLHs increasing CH₄ burden by approximately 350 Tg (9%) in the PD and 200 Tg (13%) in the PI, respectively (ESI Table S5†), which leads to increasing the CH₄ lifetime (~6–9% (Li *et al.* 2022)²⁵).

By contrast, SLH reduce the boundary layer burden of ethane (C₂H₆) by 47% and 49%, propane (C₃H₈) by 25% and 23% and larger alkanes (BIGALK > C₅) by 20% and 18%, during PD and PI, respectively (Fig. 6a and S6a†). This contrasting effect of SLH on the abundance and lifetime of VOCs results from the faster oxidation of the longer-chained VOCs by Cl than by OH (Fig. 6b). Indeed, in the presence of SLH, reaction with Cl is the main loss pathway for C₂H₆, C₃H₈ and bigalkanes and, as a consequence, their burden decreases in the presence of halogen chemistry, despite the decrease in OH levels (Fig. 6a).

$$\Delta\% \text{ loss VOCs Tot}_{\text{wSLH-noSLH}} =$$

$$\left(\frac{(k_{\text{OH}} \cdot [\text{OH}] + k_{\text{Cl}} \cdot [\text{Cl}] + k_{\text{O}_3} \cdot [\text{O}_3] + k_{\text{NO}_3} \cdot [\text{NO}_3])_{\text{wSLH}} - (k_{\text{OH}} \cdot [\text{OH}] + k_{\text{Cl}} \cdot [\text{Cl}] + k_{\text{O}_3} \cdot [\text{O}_3] + k_{\text{NO}_3} \cdot [\text{NO}_3])_{\text{noSLH}}}{(k_{\text{OH}} \cdot [\text{OH}] + k_{\text{Cl}} \cdot [\text{Cl}] + k_{\text{O}_3} \cdot [\text{O}_3] + k_{\text{NO}_3} \cdot [\text{NO}_3])_{\text{noSLH}}} \right) \times 100$$



$$\% \text{ contribution of loss by OH} = \frac{k_{\text{OH}} \cdot [\text{OH}]_{\text{wSLH}} - k_{\text{OH}} \cdot [\text{OH}]_{\text{noSLH}}}{(k_{\text{OH}} \cdot [\text{OH}] + k_{\text{Cl}} \cdot [\text{Cl}] + k_{\text{O}_3} \cdot [\text{O}_3] + k_{\text{NO}_3} \cdot [\text{NO}_3])_{\text{noSLH}}} \times 100$$

$$\% \text{ contribution of loss by Cl} = \frac{k_{\text{Cl}} \cdot [\text{Cl}]_{\text{wSLH}} - k_{\text{Cl}} \cdot [\text{Cl}]_{\text{noSLH}}}{(k_{\text{OH}} \cdot [\text{OH}] + k_{\text{Cl}} \cdot [\text{Cl}] + k_{\text{O}_3} \cdot [\text{O}_3] + k_{\text{NO}_3} \cdot [\text{NO}_3])_{\text{noSLH}}} \times 100$$

$$\% \text{ contribution of loss by O}_3 = \frac{k_{\text{O}_3} \cdot [\text{O}_3]_{\text{wSLH}} - k_{\text{O}_3} \cdot [\text{O}_3]_{\text{noSLH}}}{(k_{\text{OH}} \cdot [\text{OH}] + k_{\text{Cl}} \cdot [\text{Cl}] + k_{\text{O}_3} \cdot [\text{O}_3] + k_{\text{NO}_3} \cdot [\text{NO}_3])_{\text{noSLH}}} \times 100$$

$$\% \text{ contribution of loss by NO}_3 = \frac{k_{\text{NO}_3} \cdot [\text{NO}_3]_{\text{wSLH}} - k_{\text{NO}_3} \cdot [\text{NO}_3]_{\text{noSLH}}}{(k_{\text{OH}} \cdot [\text{OH}] + k_{\text{Cl}} \cdot [\text{Cl}] + k_{\text{O}_3} \cdot [\text{O}_3] + k_{\text{NO}_3} \cdot [\text{NO}_3])_{\text{noSLH}}} \times 100$$

where k_{OH} is the rate constant for the loss of VOCs with OH and k_{Cl} is the rate constant for the loss of VOCs with Cl. In the case of Isoprene, C_3H_6 and $\text{C}_{10}\text{H}_{16}$ we have also taken into account the additional loss due to O_3 and NO_3 .

4 Summary and conclusions

This study presents the complex interplay between SLH and atmospheric oxidants. We show that in the present-day atmosphere, SLH substantially decrease atmospheric oxidation through a reduction effect on the combined concentration of the main atmospheric oxidants. Importantly, the effect of SLH on atmospheric oxidation has changed over time due to SLH emission changes and their chemical interaction with anthropogenic pollution, and is expected to vary further due to changing climate, therefore their effects should not be assumed constant in models. Remarkably, we find that the role of halogens on atmospheric oxidation was more prominent in the pristine pre-industrial atmosphere where a quarter of OH and half of NO_3 concentrations in the boundary layer are accounted for by SLH. This result implies that the pre-industrial atmospheric oxidation capacity cannot be understood without the consideration of SLH emissions and chemistry. Critically, our analysis shows that while MDSA represents a significant source of chlorine in the boundary layer, its impact on global tropospheric OH reduction is relatively modest, increasing the total OH decrease by approximately 3% and 8% during the present day (PD) and pre-industrial (PI) periods, respectively, compared to simulations without MDSA-derived Cl. This limited impact underscores the dominant role of iodine and bromine in regulating tropospheric O_3 and, consequently, OH.

The results also show, particularly for the pre-industrial atmosphere, that SLH reduces the OH burden dependence on primary production and, in turn, increases OH recycling efficiency, which enhances the natural resilience of the atmosphere to external perturbations altering the OH budget. Furthermore, the interplay between oxidants, SLH, and VOCs ultimately

results in differing but substantially changes in the burden of key atmospheric organic compounds, with important implications for air quality and aerosol formation.

Our findings reveal that the complex and multidirectional chemical interactions between SLH and atmospheric oxidants, currently unaccounted for in air quality and climate assessments, substantially reduce global atmospheric oxidation across the pre-industrial and present-day atmospheres, and thus influence the concentrations and trends of pollutants and short-lived climate forcers. These reductions exhibit spatial heterogeneity, with the most significant impacts observed over oceanic regions, which can be considered pristine environments. In contrast, the influence of SLHs on OH production is less pronounced in polluted continental regions, such as China, where high NO_x concentrations limit the impact of halogen chemistry on overall OH production, although SLHs can still modulate OH recycling. Semi-polluted regions, such as South America, represent an intermediate case. This regional variation underscores the importance of considering local environmental conditions, including pollution levels, when assessing the impact of SLHs.

Our results reveal the dominant role of SLHs in the global oxidation state of the pre-industrial atmosphere. We therefore suggest incorporating detailed SLH emissions and chemistry into air quality and climate models to improve their predictive skills on atmospheric oxidation in past, present, and future climates.

Data availability

The CESM software used in this work is publicly available for download at <https://www.cesm.ucar.edu/models/>. Simulations outputs for each scenario analyzed in this manuscript are available in <https://zenodo.org/records/13933571>.

Author contributions

A. S.-L. devised the research. A. B. performed the CESM simulations with the help of R. P. F. and Q. L. All authors discussed the findings and commented on the manuscript. A. B., A. S. M., and A. S.-L. wrote the manuscript with contributions from all authors.

Conflicts of interest

The authors declare no competing interests.

Acknowledgements

This study has received funding from the European Research Council Executive Agency under the European Union's Horizon 2020 Research and Innovation Programme (Project ERC-2016-COG 726349 CLIMAHAL). The CESM project, supported primarily by the NSF (National Science Foundation), is led by the National Center for Atmospheric Research (NCAR), a major facility sponsored by the NSF under cooperative agreement 1852977. The Climate Simulation Laboratory at NCAR's



Computational and Information Systems Laboratory, sponsored by the NSF, provided computing resources, support and data storage.

References

- 1 IPCC, *Climate Change 2021: the Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change*, 2021.
- 2 A. M. Fiore, L. J. Mickley, Q. Zhu and C. B. Baublitz, Climate and Tropospheric Oxidizing Capacity, *Annu. Rev. Earth Planet. Sci.*, 2024, **52**, 321–349. Available from: <https://www.annualreviews.org/content/journals/10.1146/annurev-earth-032320-090307>.
- 3 D. J. Jacob, *Introduction to Atmospheric Chemistry*, Princeton University Press, 1999.
- 4 B. J. Finlayson-Pitts and J. N. Pitts, *Atmospheric Chemistry: Fundamentals and Experimental Techniques*, John Wiley, New York, 1986.
- 5 R. P. Wayne, I. Barnes, P. Biggs, J. P. Burrows, C. E. Canosamas, J. Hjorth, *et al.*, The Nitrate Radical – Physics, Chemistry, and the Atmosphere, *Atmos. Environ. Part A Gen. Top.*, 1991, **25**(1), 1–203.
- 6 O. C. Zafiriou, Photochemistry of halogens in the marine atmosphere, *J. Geophys. Res.*, 1974, **79**(18), 2730–2732.
- 7 R. von Glasow and P. J. Crutzen, Tropospheric halogen chemistry, in *Treatise on Geochemistry*, The Atmosphere, ed. R. Keeling, Elsevier-Pergamon, Oxford, 2007, pp. 1–76.
- 8 A. Saiz-Lopez and R. von Glasow, Reactive halogen chemistry in the troposphere, *Chem. Soc. Rev.*, 2012, **41**, 6448–6472.
- 9 J. Lelieveld, S. Gromov, A. Pozzer and D. Taraborrelli, Global tropospheric hydroxyl distribution, budget and reactivity, *Atmos. Chem. Phys.*, 2016, **16**(19), 12477–12493. Available from: <https://acp.copernicus.org/articles/16/12477/2016/>.
- 10 J. Lelieveld, F. J. Dentener, W. Peters and M. C. Krol, On the role of hydroxyl radicals in the self-cleansing capacity of the troposphere, *Atmos. Chem. Phys.*, 2004, **4**(9/10), 2337–2344. Available from: <https://acp.copernicus.org/articles/4/2337/2004/>.
- 11 C. D. Holmes, M. J. Prather, O. A. Søvde and G. Myhre, Future methane, hydroxyl, and their uncertainties: key climate and emission parameters for future predictions, *Atmos. Chem. Phys.*, 2013, **13**(1), 285–302.
- 12 D. Stone, L. K. Whalley and D. E. Heard, Tropospheric OH and HO₂ radicals: field measurements and model comparisons, *Chem. Soc. Rev.*, 2012, **41**(19), 6348–6404, DOI: [10.1039/C2CS35140D](https://doi.org/10.1039/C2CS35140D).
- 13 D. E. Heard, Field Measurements of Atmospheric Composition, in *Analytical Techniques for Atmospheric Measurement*, John Wiley & Sons, Ltd, 2006, pp. 1–71, DOI: [10.1002/9780470988510.ch1](https://doi.org/10.1002/9780470988510.ch1).
- 14 D. S. Stevenson, A. Zhao, V. Naik, F. M. O'Connor, S. Tilmes, G. Zeng, *et al.*, Trends in global tropospheric hydroxyl radical and methane lifetime since 1850 from AerChemMIP, *Atmos. Chem. Phys.*, 2020, **20**(21), 12905–12920.
- 15 V. Naik, A. Voulgarakis, A. M. Fiore, L. W. Horowitz, J. F. Lamarque, M. Lin, *et al.*, Preindustrial to present-day changes in tropospheric hydroxyl radical and methane lifetime from the Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP), *Atmos. Chem. Phys.*, 2013, **13**(10), 5277–5298.
- 16 O. Wild, A. Voulgarakis, F. O'Connor, J. F. Lamarque, E. M. Ryan and L. Lee, Global sensitivity analysis of chemistry-climate model budgets of tropospheric ozone and OH: Exploring model diversity, *Atmos. Chem. Phys.*, 2020, **20**(7), 4047–4058.
- 17 G. Myhre, D. Shindell, F. M. Bréon, W. Collins, J. Fuglestad, J. Huang, *et al.*, Anthropogenic and Natural Radiative Forcing, in *Climate Change 2013: the Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2013.
- 18 A. Voulgarakis, V. Naik, J. F. Lamarque, D. T. Shindell, P. J. Young, M. J. Prather, *et al.*, Analysis of present day and future OH and methane lifetime in the ACCMIP simulations, *Atmos. Chem. Phys.*, 2013, **13**(5), 2563–2587.
- 19 A. Badia, C. E. Reeves, A. R. Baker, A. Saiz-Lopez, R. Volkamer, T. K. Koenig, *et al.*, Importance of reactive halogens in the tropical marine atmosphere: A regional modelling study using WRF-Chem, *Atmos. Chem. Phys.*, 2019, **19**, 3161–3189.
- 20 A. Saiz-Lopez, R. P. Fernandez, Q. Li, C. A. Cuevas, X. Fu, D. E. Kinnison, *et al.*, Natural short-lived halogens exert an indirect cooling effect on climate, *Nature*, 2023, **618**(7967), 967–973.
- 21 A. Saiz-Lopez, J. F. Lamarque, D. E. Kinnison, S. Tilmes, C. Ordóñez, J. J. Orlando, *et al.*, Estimating the climate significance of halogen-driven ozone loss in the tropical marine troposphere, *Atmos. Chem. Phys.*, 2012, **12**, 3939–3949.
- 22 F. Iglesias-Suarez, A. Badia, R. P. Fernandez, C. A. Cuevas, D. E. Kinnison, S. Tilmes, *et al.*, Natural halogens buffer tropospheric ozone in a changing climate, *Nat. Clim. Change*, 2020, **10**(2), 147–154.
- 23 T. Sherwen, J. A. Schmidt, M. J. Evans, L. J. Carpenter, K. Großmann, S. D. Eastham, *et al.*, Global impacts of tropospheric halogens (Cl, Br, I) on oxidants and composition in GEOS-Chem, *Atmos. Chem. Phys.*, 2016, **16**(18), 12239–12271. Available from: <https://acp.copernicus.org/articles/16/12239/2016/>.
- 24 R. Hossaini, M. P. Chipperfield, S. A. Montzka, A. Rap, S. Dhomse and W. Feng, Efficiency of short-lived halogens at influencing climate through depletion of stratospheric ozone, *Nat. Geosci.*, 2015, **8**(3), 186–190, DOI: [10.1038/ngeo2363](https://doi.org/10.1038/ngeo2363).
- 25 Q. Li, R. P. Fernandez, R. Hossaini, F. Iglesias-Suarez, C. A. Cuevas, E. C. Apel, *et al.*, Reactive halogens increase the global methane lifetime and radiative forcing in the 21st century, *Nat. Commun.*, 2022, **13**(1), 2768.



- 26 W. R. Simpson, S. S. Brown, A. Saiz-Lopez, J. A. Thornton and G. R. Von, Tropospheric Halogen Chemistry: Sources, Cycling, and Impacts, *Chem. Rev.*, 2015, **115**(10), 4035–4062.
- 27 J. A. Barrera, D. E. Kinnison, R. P. Fernandez, J. F. Lamarque, C. A. Cuevas, S. Tilmes, *et al.*, Comparing the Effect of Anthropogenically Amplified Halogen Natural Emissions on Tropospheric Ozone Chemistry Between Pre-Industrial and Present-Day, *J. Geophys. Res. Atmos.*, 2023, **128**(14), e2022JD038283.
- 28 T. Sherwen, M. J. J. Evans, L. J. J. Carpenter, S. J. J. Andrews, R. T. T. Lidster, B. Dix, *et al.*, Iodine's impact on tropospheric oxidants: a global model study in GEOS-Chem, *Atmos. Chem. Phys.*, 2016, **16**(2), 1161–1186.
- 29 W. J. Bloss, M. Camredon, J. D. Lee, D. E. Heard, J. M. C. Plane, A. Saiz-Lopez, *et al.*, Coupling of HOx, NOx and halogen chemistry in the Antarctic boundary layer, *Atmos. Chem. Phys.*, 2010, **10**, 10187–10209.
- 30 M. M. J. W. van Herpen, Q. Li, A. Saiz-Lopez, J. B. Liisberg, T. Röckmann, C. A. Cuevas, *et al.*, Photocatalytic chlorine atom production on mineral dust-sea spray aerosols over the North Atlantic, *Proc. Natl. Acad. Sci. U. S. A.*, 2023, **120**(31), e2303974120, DOI: [10.1073/pnas.2303974120](https://doi.org/10.1073/pnas.2303974120).
- 31 K. A. Read, A. S. Mahajan, L. J. Carpenter, M. J. Evans, B. V. E. Faria, D. E. Heard, *et al.*, Extensive halogen-mediated ozone destruction over the tropical Atlantic Ocean, *Nature*, 2008, **453**(7199), 1232–1235.
- 32 A. S. Mahajan, J. M. C. Plane, H. Oetjen, L. M. Mendes, R. W. Saunders, A. Saiz-Lopez, *et al.*, Measurement and modelling of tropospheric reactive halogen species over the tropical Atlantic Ocean, *Atmos. Chem. Phys.*, 2010, **10**, 4611–4624.
- 33 G. Sarwar, H. Simon, J. Xing and R. Mathur, Importance of tropospheric ClNO₂ chemistry across the Northern Hemisphere, *Geophys. Res. Lett.*, 2014, **41**(11), 4050–4058.
- 34 Q. Li, A. Badia, T. Wang, G. Sarwar, X. Fu, L. Zhang, *et al.*, Potential Effect of Halogens on Atmospheric Oxidation and Air Quality in China, *J. Geophys. Res. Atmos.*, 2020, **125**(9), e2019JD032058.
- 35 D. Stone, T. Sherwen, M. J. Evans, S. Vaughan, T. Ingham, L. K. Whalley, *et al.*, Impacts of bromine and iodine chemistry on tropospheric OH and HO₂: comparing observations with box and global model perspectives, *Atmos. Chem. Phys.*, 2018, **18**(5), 3541–3561. Available from: <https://acp.copernicus.org/articles/18/3541/2018/>.
- 36 Q. Li, R. Borge, G. Sarwar, D. de la Paz, B. Gantt, J. Domingo, *et al.*, Impact of halogen chemistry on summertime air quality in coastal and continental Europe: application of the CMAQ model and implications for regulation, *Atmos. Chem. Phys.*, 2019, **19**(24), 15321–15337. Available from: <https://acp.copernicus.org/articles/19/15321/2019/>.
- 37 C. A. Cuevas, N. Maffezzoli, J. P. Corella, A. Spolaor, P. Vallelonga, H. A. Kjær, *et al.*, Rapid increase in atmospheric iodine levels in the North Atlantic since the mid-20th century, *Nat. Commun.*, 2018, **9**(1), 1452.
- 38 M. Legrand, J. R. McConnell, S. Preunkert, M. Arienzo, N. Chellman, K. Gleason, *et al.*, Alpine ice evidence of a three-fold increase in atmospheric iodine deposition since 1950 in Europe due to increasing oceanic emissions, *Proc. Natl. Acad. Sci. U. S. A.*, 2018, **115**(48), 12136–12141.
- 39 X. Zhao, X. Hou and W. Zhou, Atmospheric Iodine (127I and 129I) Record in Spruce Tree Rings in the Northeast Qinghai-Tibet Plateau, *Environ. Sci. Technol.*, 2019, **53**(15), 8706–8714.
- 40 L. J. Carpenter, S. M. MacDonald, M. D. Shaw, R. Kumar, R. W. Saunders, R. Parthipan, *et al.*, Atmospheric iodine levels influenced by sea surface emissions of inorganic iodine, *Nat. Geosci.*, 2013, **6**(2), 108–111.
- 41 S. M. MacDonald, J. C. Gómez Martín, R. Chance, S. Warriner, A. Saiz-Lopez, L. J. Carpenter, *et al.*, A laboratory characterisation of inorganic iodine emissions from the sea surface: dependence on oceanic variables and parameterisation for global modelling, *Atmos. Chem. Phys.*, 2014, **14**(11), 5841–5852.
- 42 C. Prados-Roman, C. a. Cuevas, R. P. Fernandez, D. E. Kinnison, J. F. Lamarque and A. Saiz-Lopez, A negative feedback between anthropogenic ozone pollution and enhanced ocean emissions of iodine, *Atmos. Chem. Phys.*, 2015, **15**(4), 2215–2224.
- 43 J. P. Corella, N. Maffezzoli, A. Spolaor, P. Vallelonga, C. A. Cuevas, F. Scoto, *et al.*, Climate changes modulated the history of Arctic iodine during the Last Glacial Cycle, *Nat. Commun.*, 2022, **13**(1), 6–14.
- 44 P. Vallelonga, N. Maffezzoli, A. Saiz-Lopez, F. Scoto, H. A. Kjær and A. Spolaor, Sea-ice reconstructions from bromine and iodine in ice cores, *Quat. Sci. Rev.*, 2021, **269**, 107133.
- 45 S. Zhai, J. R. McConnell, N. Chellman, M. Legrand, T. Opel, H. Meyer, *et al.*, Anthropogenic Influence on Tropospheric Reactive Bromine Since the Pre-industrial: Implications for Arctic Ice-Core Bromine Trends, *Geophys. Res. Lett.*, 2024, **51**(5), e2023GL107733.
- 46 S. Zhai, X. Wang, J. R. McConnell, L. Geng, J. Cole-Dai, M. Sigl, *et al.*, Anthropogenic Impacts on Tropospheric Reactive Chlorine Since the Preindustrial, *Geophys. Res. Lett.*, 2021, **48**(14), 1–12.
- 47 J. F. Lamarque, L. K. Emmons, P. G. Hess, D. E. Kinnison, S. Tilmes, F. Vitt, *et al.*, CAM-chem: description and evaluation of interactive atmospheric chemistry in the Community Earth System Model, *Geosci. Model Dev.*, 2012, **5**(2), 369–411. Available from: <https://gmd.copernicus.org/articles/5/369/2012/>.
- 48 S. Tilmes, J. F. Lamarque, L. K. Emmons, D. E. Kinnison, D. Marsh, R. R. Garcia, *et al.*, Representation of the Community Earth System Model (CESM1) CAM4-chem within the Chemistry-Climate Model Initiative (CCMI), *Geosci. Model Dev.*, 2016, **9**(5), 1853–1890. Available from: <https://gmd.copernicus.org/articles/9/1853/2016/>.
- 49 R. P. Fernandez, R. J. Salawitch, D. E. Kinnison, J. F. Lamarque and A. Saiz-Lopez, Bromine partitioning in the tropical tropopause layer: implications for stratospheric injection, *Atmos. Chem. Phys.*, 2014, **14**(24), 13391–13410. Available from: <https://acp.copernicus.org/articles/14/13391/2014/>.
- 50 C. Ordóñez, J. F. Lamarque, S. Tilmes, D. E. Kinnison, E. L. Atlas, D. R. Blake, *et al.*, Bromine and iodine



- chemistry in a global chemistry-climate model: description and evaluation of very short-lived oceanic sources, *Atmos. Chem. Phys.*, 2012, **12**(3), 1423–1447. Available from: <https://acp.copernicus.org/articles/12/1423/2012/>.
- 51 A. Saiz-Lopez, R. P. Fernandez, C. Ordóñez, D. E. Kinnison, J. C. Gómez Martín, J. F. Lamarque, *et al.*, Iodine chemistry in the troposphere and its effect on ozone, *Atmos. Chem. Phys.*, 2014, **14**(23), 13119–13143. Available from: <https://acp.copernicus.org/articles/14/13119/2014/>.
- 52 T. Claxton, R. Hossaini, C. Wilson, S. A. Montzka, M. P. Chipperfield, O. Wild, *et al.*, A Synthesis Inversion to Constrain Global Emissions of Two Very Short Lived Chlorocarbons: Dichloromethane, and Perchloroethylene, *J. Geophys. Res. Atmos.*, 2020, **125**(12), e2019JD031818, DOI: [10.1029/2019JD031818](https://doi.org/10.1029/2019JD031818).
- 53 A. B. Guenther, X. Jiang, C. L. Heald, T. Sakulyanontvittaya, T. Duhl, L. K. Emmons, *et al.*, The Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): an extended and updated framework for modeling biogenic emissions, *Geosci. Model Dev.*, 2012, **5**(6), 1471–1492. Available from: <https://gmd.copernicus.org/articles/5/1471/2012/>.
- 54 J. A. Barrera, R. P. Fernandez, F. Iglesias-Suarez, C. A. Cuevas, J. F. Lamarque and A. Saiz-Lopez, Seasonal impact of biogenic very short-lived bromocarbons on lowermost stratospheric ozone between 60N and 60S during the 21st century, *Atmos. Chem. Phys.*, 2020, **20**(13), 8083–8102. Available from: <https://acp.copernicus.org/articles/20/8083/2020/>.
- 55 WMO, *Scientific Assessment of Ozone Depletion: 2010 Global Ozone Research and Monitoring Project Report 52*, WMO, 2011.
- 56 J. Lelieveld, W. Peters, F. J. Dentener and M. C. Krol, Stability of tropospheric hydroxyl chemistry, *J. Geophys. Res. Atmos.*, 2002, **107**(D23), 1–11, DOI: [10.1029/2002JD002272](https://doi.org/10.1029/2002JD002272).
- 57 A. Saiz-Lopez and R. P. Fernandez, On the formation of tropical rings of atomic halogens: Causes and implications, *Geophys. Res. Lett.*, 2016, **43**(6), 2928–2935, DOI: [10.1002/2015GL067608](https://doi.org/10.1002/2015GL067608).
- 58 C. A. Cuevas, R. P. Fernandez, D. E. Kinnison, Q. Li, J. F. Lamarque and T. Trabelsi, The influence of iodine on the Antarctic stratospheric ozone hole, *Proc. Natl. Acad. Sci. U. S. A.*, 2022, **119**(7), e2110864119.
- 59 R. P. Fernandez, D. E. Kinnison, J. F. Lamarque, S. Tilmes and A. Saiz-Lopez, Impact of biogenic very short-lived bromine on the Antarctic ozone hole during the 21st century, *Atmos. Chem. Phys.*, 2017, **17**(3), 1673–1688. Available from: <https://acp.copernicus.org/articles/17/1673/2017/>.
- 60 T. K. Koenig, S. Baidar, P. Campuzano-Jost, C. A. Cuevas, B. Dix, R. P. Fernandez, *et al.*, Quantitative detection of iodine in the stratosphere, *Proc. Natl. Acad. Sci. U. S. A.*, 2020, **117**(4), 1860–1866, DOI: [10.1073/pnas.1916828117](https://doi.org/10.1073/pnas.1916828117).
- 61 J. E. Mak, G. Kra, T. Sandomenico and P. Bergamaschi, The seasonally varying isotopic composition of the sources of carbon monoxide at Barbados, West Indies, *J. Geophys. Res. Atmos.*, 2003, **108**(D20), 4635, DOI: [10.1029/2003JD003419](https://doi.org/10.1029/2003JD003419).
- 62 C. Caram, S. Szopa, A. Cozic, S. Bekki, C. A. Cuevas and A. Saiz-Lopez, Sensitivity of tropospheric ozone to halogen chemistry in the chemistry–climate model LMDZ-INCA vNMHC, *Geosci. Model Dev.*, 2023, **16**(14), 4041–4062. Available from: <https://gmd.copernicus.org/articles/16/4041/2023/>.
- 63 Y. J. Tham, X. C. He, Q. Li, C. A. Cuevas, J. Shen, J. Kalliokoski, *et al.*, Direct field evidence of autocatalytic iodine release from atmospheric aerosol, *Proc. Natl. Acad. Sci. U. S. A.*, 2021, **118**(4), 1–8.
- 64 J. M. Roberts, H. D. Osthoff, S. S. Brown and A. R. Ravishankara, N₂O₅ oxidizes chloride to Cl₂ in acidic atmospheric aerosol, *Science*, 2008, **321**(5892), 1059.
- 65 S. Bleicher, J. C. Buxmann, R. Sander, T. P. Riedel, J. A. Thornton, U. Platt, *et al.*, The influence of nitrogen oxides on the activation of bromide and chloride in salt aerosol, *Atmos. Chem. Phys. Discuss.*, 2014, **14**(7), 10135–10166.
- 66 J. Dai, G. P. Brasseur, M. Vrekoussis, M. Kanakidou, K. Qu, Y. Zhang, *et al.*, The atmospheric oxidizing capacity in China-Part 1: Roles of different photochemical processes, *Atmos. Chem. Phys.*, 2023, **23**(22), 14127–14158.
- 67 H. Chen, P. Liu, Q. Wang, R. Huang and G. Sarwar, Impact and pathway of halogens on atmospheric oxidants in coastal city clusters in the Yangtze River Delta region in China, *Atmos. Pollut. Res.*, 2024, **15**(2), 101979.
- 68 L. K. Whalley, K. L. Furneaux, A. Goddard, J. D. Lee, A. S. Mahajan, H. Oetjen, *et al.*, The chemistry of OH and HO₂ radicals in the boundary layer over the tropical Atlantic Ocean, *Atmos. Chem. Phys.*, 2010, **10**(4), 1555–1576.
- 69 R. Sommariva, W. J. Bloss, N. Brough, N. Carslaw, M. Flynn, A. L. Haggerstone, *et al.*, OH and HO₂ chemistry during NAMBLEX: roles of oxygenates, halogen oxides and heterogeneous uptake, *Atmos. Chem. Phys.*, 2006, **6**, 1135–1153.

