

Typology of the ecological impacts of biological invasions

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51 **Abstract**

52 Biological invasions alter ecosystems by disrupting ecological processes that can degrade
53 biodiversity, human health, and cause massive economic burdens. Existing frameworks
54 to classify the ecological impacts either miss many types of impact or conflate
55 mechanisms (causes) with the impacts themselves (consequences). We propose a
56 comprehensive typology of 19 types of ecological impact across six levels of ecological
57 organisation. This allows more accurate diagnosis of the cause of impact and can help
58 triage management options to tackle each impact-mechanism combination. We integrated
59 the typology with broad ecological concepts such as energy, mass, and information flow
60 and storage. By highlighting cascading effects across multiple levels, this typology
61 provides a clearer framework for documenting, and communicating invasion impacts,
62 thereby improving management and research.

63

64 **The need for a comprehensive impact typology**

65 Biological invasions can occur when a species is introduced into an area where it is not
66 native [1]. Once the alien (or non-native) species is established and spreading in the new
67 environment, they are classified as ‘invasive’, often with many documented **impacts**
68 (see Glossary) on biodiversity and society [2,3]. **Invasive species** are recognized as one
69 of the major causes of native population declines and species loss, as well as habitat
70 degradation and erosion of ecosystem functioning and services [3]. Due to the variety of
71 these impacts, past efforts have been made to classify them, serving as the basis for
72 impact documentation by researchers, prioritisation by practitioners and international
73 institutions like the International Union for Conservation of Nature (IUCN), and global
74 assessments on biological invasions [3–5]. Despite these advancements, current impact

75 classifications are limited in scope and precision regarding the typology of impacts,
76 reducing their overall applicability (see Supplementary material Table S1).

77 Pioneering endeavours such as the *Generic Impact Scoring System* (GISS) [6] and the
78 *Environmental Impact Classification for Alien Taxa* (EICAT) [4,7] aim to assess the
79 impacts of biological invasions systematically. These frameworks provide valuable tools
80 to classify invasive species based on impact magnitude. The GISS categorizes impacts
81 based on six ecological **mechanisms** and on six socio-economic sectors, while EICAT
82 focuses on impacts on native biodiversity through 12 mechanisms. These frameworks
83 have been applied to many taxa globally, and the EICAT has been adopted as a global
84 standard by the IUCN [4]. Despite this wide usage, the latter only considers documented
85 impacts of invasive species on native species — impacts on ecosystem processes and
86 abiotic changes alone are not captured (e.g., [8]). Furthermore, it is not unusual for studies
87 to refer to both mechanisms of impacts (e.g., predation by the invasive species) and the
88 resulting types of impact (e.g., native prey population decline) under the broad label of
89 *impacts*. However, these are structurally different: *mechanisms* represent the cause while
90 types of impact reflect the consequences. This conflation of cause and consequence
91 creates an inconsistent typology that can hinder clear assessment and communication.
92 Existing databases such as the Global Invasive Species Database (GISD) and CABI's
93 Invasive Species Compendium are valuable for cataloguing invasion-related data, but
94 their species-specific approaches can lead to inconsistencies in the categorization of
95 ecological impacts. CABI's Invasive Species Compendium, for example, provides a
96 range of ecological, economic, and social impact outcomes based on varied sources,
97 which makes cross-taxa comparisons difficult. However some progress has recently been
98 made with standardization of impact studies on GISD, which is the current home for
99 systematically collated EICAT assessments. While these original frameworks, databases,
100 and others (see [9,10]), have been instrumental in advancing our understanding of the
101 severity of invasion impacts, there is a need for a comprehensive and standardised
102 typology that also clearly separates ecological impacts from causal mechanisms.

103 Based on the growing empirical evidence for the diverse impacts of biological
104 invasions, we have developed an exhaustive typology of ecological impacts, scaled across
105 levels of biological organisation from individuals to ecosystem functions. We then
106 discuss how mechanisms acting across different levels of this hierarchy link the 19 types
107 of impacts and clarify the distinction between causes and consequences. Such a typology

108 brings research, management, and stakeholder communication closer to a more precise
109 and unified understanding of the effect of biological invasions.

110

111 ***Identifying and disentangling impact types***

112 A major barrier to standardising impact assessments is the complexity and the
113 interconnected nature of impacts across the different levels of biological organisation and
114 associated ecosystem processes. Different impacts can occur simultaneously across
115 multiple ecological scales, from individuals to ecosystems, and act on both biota and the
116 non-living (abiotic) environment. For example, the loss of a local population of native
117 species can trigger the loss of associated ecosystem functions [11]. Additionally, the
118 effects of biological invasions are realized through various mechanisms (causes) that are
119 often mixed with the impact types themselves (consequences). To address these
120 challenges, distinct but complementary aspects of invasion impacts need to be assessed
121 and organized.

122

123 ***Separating cause from consequence***

124 To identify and measure the impacts of invasive species accurately, one must distinguish
125 the mechanisms driving these impacts from the resulting impacts themselves. A species
126 can disrupt native ecosystems with various *mechanisms* leading to *impacts*, such as direct
127 predation leading to population collapse[12], competition leading to primary production
128 reduction and resource depletion for native species [13], and disease transmission leading
129 to negative effects on health, growth, or reproduction of individuals[14]. We define these
130 disruptive interactions as ‘mechanisms’ *sensu* [7], and their consequences as ‘impacts’.
131 For example, the brown tree snake (*Boiga irregularis*) in Guam [15] caused the extinction
132 of local fauna through direct predation. In that case, species loss is the *impact*, and
133 predation is the *mechanism*. However, many studies use these two concepts
134 interchangeably by listing for example ‘predation’ by an invasive species as an ‘impact’,
135 which conflates the two phenomena. If the impact could instead be measured
136 systematically as the *consequence* (impact definition *sensu* [16]) of the predation in this
137 case (e.g., altered behaviour of individuals, abundance declines, extinction, etc.), it would
138 clarify the much-needed distinction between these two concepts [17]. Predation by
139 invasive species such as European red foxes (*Vulpes vulpes*) and feral cats (*Felis catus*)
140 in Australia exemplifies how the same mechanism can produce various impacts, from
141 abundance declines to range retractions, and even extinctions of native species [18]).

142 Beyond predation, other mechanisms such as competition also cause impacts. For
143 example, non-native fish compete with native species, reducing **alpha** and **beta diversity**,
144 altering food-web structure, and thereby decreasing ecosystem functionality[19]. In
145 plants, competing mechanisms such as allelopathy can cause impacts that cascade from
146 the population to the ecosystem level, potentially driving long-term changes in
147 community structure and ecosystem processes [20]. These two cases are good examples
148 of different mechanisms (predation and competition) drive distinct ecological impacts,
149 each one with cascading consequences. Recognizing these differences is essential
150 because each impact-mechanism might require a distinct mitigation and management
151 response.

152

153 *Categorising all existing impact types*

154 To establish a unified standard for classifying ecological impacts of invasions, we need a
155 typology that is both comprehensive and straightforward. This typology should consist of
156 well-defined impact types, each fitting into a few distinct and easily understandable
157 categories. For widespread adoption, the scheme needs to be compatible with most
158 published studies and reach a consensus among experts of biological invasions. Currently,
159 there is no synthesis fulfilling all these criteria; the EICAT is arguably the closest, but it
160 is limited to impacts on native biodiversity and excludes *de facto* impacts on abiotic factor
161 and at the ecosystem level.

162 We first reviewed the literature on existing ecological impact typologies, (see
163 Supplementary material Table S1). These studies exhibit varying levels of organisation,
164 from extensive lists of impacts and broad ecological categorisations (e.g., [21–24]), to
165 detailed impacts focused specifically on plants (e.g., [10,25,26]). Some studies address
166 other taxonomic groups, collectively providing a comprehensive but scattered overview
167 of the diverse impacts of invasive species.

168 Building on this previous research, we compiled all existing impact types, regrouping
169 similar ones under broader categories to create a comprehensive, simple, and mutually
170 exclusive list. After extensive discussion and deliberation, we developed a proposed list
171 of impact types, which we then presented to 60 leading experts in the field. Using a **Delphi**
172 **process** [27], we did two rounds of voting and incorporated suggestions for improvement
173 and refinement [28]. Once we achieved a consensus, we identified the biological levels
174 of organisation at which these impacts can occur (but to which they are not limited). Our
175 assessment revealed that the impacts are expressed through 19 distinct types across six

176 levels of ecological organisation: (i) individual/organism, (ii) population, (iii) species, (iv)
 177 **assemblage**, (v) ecosystem, and (vi) abiotic environment. Each of these 19 impact types
 178 operates primarily at one of these six levels, although they can cascade to affect other
 179 levels and even other impact types (Figure 1, Table 1). Note that the typology is meant to
 180 identify and categorize the different types of impacts. However, a given category of
 181 impact can occur at different scales – for example, assemblage-level structure change can
 182 occur in a local community or at the scale of an entire region. The typology can be applied
 183 regardless of the spatial or temporal scale, and either works for single studies or data
 184 aggregation. Naturally, the spatial scale or degree of aggregation should be taken into
 185 consideration by users when using the typology, especially if it is meant for comparative
 186 purposes.

187 Besides the ecological levels of organization, we also categorized each type of impact
 188 into one of the four main components of systems ecology: energy, mass, information
 189 flow, and information storage (Table 1). For example, invasive species can disrupt energy
 190 flow by altering primary production or trophic dynamics, or affect mass by modifying
 191 nutrient cycles and habitat structure. Similarly, shifts in information flow such as
 192 behavioural changes or species interactions, and in information storage such as the loss
 193 of genetic diversity, highlight how these impacts span different dimensions of ecosystem
 194 functioning. Framing invasion impacts within these ecological components enhances
 195 comparability across studies and aligns invasion biology with broader ecosystem theory,
 196 making it easier to integrate invasion impacts into ecosystem models, conservation
 197 planning, and environmental impact assessments.

198
 199 **Table 1:** Types of impacts of invasive species, with their respective terms, definitions,
 200 ecological concepts, and associated variables to measure them, with examples. The
 201 impact types are also separated into the six ecological levels. Despite some impacts being
 202 identified in only one ecological level, they might affect others.
 203

Impact type	Definition	Ecological concept	Typically measured variable	Examples of impact description	References
Individual					
fitness and/or reproduction (1)	Change in individual reproductive capacity and overall individual fitness in native species that can influence population dynamics. Fitness or reproductive success is a combination	mass, information flow	reproductive success, survival rates	<i>Miconia calvescens</i> reduces fertility of understorey trees in Tahiti rainforests	[29]

	of survival, mating success, and fertility.				
health and/or growth (2)	Change (e.g., inhibition, increasing) of growth and adverse impacts on the physical condition of individual organisms.	energy, mass	health indices (e.g., disease prevalence, physiological stress), growth rates	Impact of <i>Carpobrotus edulis</i> on native plants. This experimental study shows the impact at different stages of plant growth. The presence of invasive insects carrying non-native fungal pathogens can reduce growth and vigour of forest trees	[30] [31]
behavioural (3)	Shifts in the actions, activities, and responses exhibited by individual organisms or populations.	mass, information flow	behavioural observations, activity patterns, habitat use	Native squirrel (<i>Sciurus vulgaris</i>) activity reduced following infection by non-native parasites (<i>Strongyloides robustus</i>). Native topminnows (<i>Skiffia bilineata</i>) reduced their foraging time when in company of invasive fish <i>Capreolus capreolus</i> decreased feeding and increased vigilance when near introduced fallow deer (<i>Dama dama</i>)	[32] [33] [34]
Population					
population size (4)	Reductions or increases in the number of individuals within populations of native species.	mass	population abundance, population growth rates, recruitment rates	Crayfish (<i>Aphanomyces astaci</i>) plague can cause large mortality events in crayfish in invaded streams and lakes. Reduction of population size of ground-nesting birds by the American mink (<i>Neogale vison</i>).	[35] [36]
genetic diversity (5)	Reduction in genetic variation and diversity within populations and species resulting from hybridization, introgression, and genetic assimilation processes. This occurs when genetic diversity within a population decreases due to factors such as genetic drift or reduced gene flow, leading to decreased adaptability and resilience.	information storage	genetic diversity indices, gene flow rates, genetic differentiation	Invasion of the invasive European barbel (<i>Barbus barbus</i>) in central Italy causes genetic introgression, threatening native barbels <i>B. plebejus</i> and <i>B. tyberinus</i> . Widespread introgression between native Oreochromines and Nile Tilapia (<i>Oreochromis niloticus</i>) in the Middle Zambezi Basin has caused almost the complete loss of <i>Oreochromis mortimeri</i> in Lake Kariba.	[37] [38]

Species					
species range (6)	Shifts in the geographical distribution of species, including expansions, contractions, or shifts in habitat occupancy.	mass	geographic distribution, habitat suitability, dispersal ability	Contraction of the range of a native animal species due to competition with invasive species; replacement of <i>Sciurus vulgaris</i> by <i>S. carolinensis</i>	[39]
species loss (7)	Decline or disappearance of native species within a particular ecosystem or geographical area.	mass, information storage	species richness, community composition	Predation by <i>Boiga irregularis</i> extirpated bird species from Guam	[12]
Assemblage					
assemblage structure (8)	Alterations in the diversity and abundance of species within assemblages, which can scale from local communities to large-scale species pools.	energy, information storage	alpha, beta and gamma diversity indices	<p>Fish faunas across continental United States have become more similar because of widespread introductions of cosmopolitan species</p> <p>In Australian grasslands, dominant invasive grasses, <i>Bromus diandrus</i> and <i>Avena fatua</i>, altered community composition and reduced the cover of native species</p> <p>Litter leachate of invasive blue gum <i>Eucalyptus globulus</i> reduces more biodiversity of understorey plants compared to its native range</p>	[40] [41] [42]
successional patterns (9)	Involves alterations to the temporal sequence and trajectory of ecological succession within ecosystems.	energy, information flow	successional stage, vegetation composition, community turnover rates, disturbance regime	Invasion of many non-native plant species in old fields in Tennessee disrupts native species interactions and accelerates successional patterns by shifting native co-occurrence from structured to random, and promotes the dominance of non-native woody species that alter forest development	[43]
soundscape (10)	Changes in the acoustic environment.	information flow	acoustic diversity, sound intensity, sound frequency, temporal patterns of vocalisation, species composition, species richness, species evenness, community diversity indices	<p>Invasion of spotted knapweed (<i>Centaurea stoebe</i>) in savannahs reduced habitat quality for chipping sparrows (<i>Spizella passerina</i>), leading to fewer older song model birds and resulted in lower song diversity and greater song similarity among yearlings.</p> <p>Invasive cane toads (<i>Rhinella marina</i>) disrupt the communication systems of native frogs.</p>	[44,45] [46]

Ecosystem function/ service					
primary production (11)	Changes in the rate and magnitude of biomass production by primary producers (e.g., plants, algae) within ecosystems.	energy, mass	biomass accumulation, photosynthetic rates, primary productivity	Reduction in plant biomass production due to competition with invasive plants. Increase in algal blooms leading to enhanced primary production in aquatic ecosystems affected by invasive species.	[47] [48]
ecological function (12)	Impairment or disruption of ecosystem processes, such as nutrient cycling, pollination, or decomposition.	energy, mass, information flow, information storage	functional diversity indices (e.g., functional richness, evenness, divergence), rates of ecological processes (e.g., pollination rates, decomposition rates, nutrient cycling rates), species interactions, trophic dynamics	Extirpation of native pollinator species due to competition with invasive pollinators, resulting in reduced pollination services and decreased reproductive success for native plant species. Disruption of soil microbial communities by invasive plant species with allelopathic traits, reducing nutrient cycling rates and impairing soil fertility.	[49] [50]
food web (13)	Changes in the structure and dynamics of food chains and trophic interactions.	energy	trophic interactions, food chain length, energy flow	Disruption of native insect-plant interactions by invasive herbivores. Alteration of predator-prey dynamics in aquatic ecosystems due to introduction of invasive fish species. Invasive lake trout (<i>Salvelinus namaycush</i>) disrupt and reorganise lake trophic pathways and outcompete bull trout (<i>S. confluentus</i>) despite bull trout shifting resource consumption patterns.	[51] [52] [53]
habitat or refugia (14)	Deterioration, substitution, or disappearance of critical habitats or refuge areas for native species.	mass	habitat quality, habitat availability, habitat complexity	Degradation of nesting habitats for native bird species due to invasive vegetation encroachment loss of sheltering refugia for aquatic organisms following habitat alteration by invasive species	[54] [55]
Abiotic environment					
hydrology / water quality /	Changes related to water-related factors such as	energy, mass	water quality (e.g., pH, nutrient	Higher water use by alien plants (e.g., tamarisk,	[56] [57]

soil moisture (15)	hydrology, water quality, and soil moisture.		concentration), soil moisture content, hydrological regimes	<p>mesquite, <i>Prosopis</i>) can reduce soil moisture, runoff, and baseflow.</p> <p>Macrophytes (e.g., <i>Salvinia</i>, Eurasian watermilfoil, <i>Sagittaria</i>) can increase flood risk by reducing flow velocities and water passage.</p> <p>Invasive plants (e.g., willows, poplars) and animals (e.g. beavers, coypu, carp) can alter channel form and hydraulics, changing flow patterns and flood risk.</p> <p>Dissolved oxygen declines in the Hudson River associated with invasion of zebra mussel (<i>Dreissena polymorpha</i>).</p>	
nutrient pool and fluxes (16)	Changes in the availability, cycling, and distribution of nutrients.	energy, mass	nutrient concentrations (e.g., nitrogen, phosphorus), nutrient cycling rates, soil nutrient content	Introduced hippopotamus as ecosystem engineers in Colombia, importing terrestrial organic matter and nutrients with detectable impacts on ecosystem metabolism and community structure in the early stages of invasion.	[58]
fire regime (17)	Changes in the frequency, intensity, and spatial patterns of wildfires.	energy, mass	fire occurrence, fire severity, fire spread rates	<p>Alteration of fire frequency and intensity in grassland ecosystems invaded by flammable exotic plant species.</p> <p>Changes in fire spread patterns in forested areas following introduction of invasive shrub species.</p>	[59] [60]
soil / sediment (18)	Changes in the physical, chemical, and biological properties of soil or sediment substrates	mass	soil properties (e.g., texture, pH), sediment characteristics, mineral concentrations, heavy metal bioavailability	<p>Invasive plants altering soil chemistry.</p> <p>Increase in heavy metal bioavailability by plants.</p> <p>Changes in soil physical properties and geomorphology.</p>	[61] [62] [63]
micro-climate (19)	Alterations in local or regional climatic conditions.	energy	temperature, precipitation, humidity, wind patterns, evapotranspiration rates, albedo,	Invasive plant <i>Impatiens glandulifera</i> alters temperature and soil humidity.	[64] [65,66]

			carbon dioxide concentration	Dense stands of <i>Ammophila arenaria</i> reduce temperatures and available light.	
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205 ***Cascading impacts***

206 Invasive species can directly induce one or multiple types of impacts within invaded
207 ecosystems, often with interrelated repercussions across impact types (e.g., [67,68])
208 (Figure 2), which can complicate the understanding of cumulative impacts in the absence
209 of a structured typology. Ecological impacts can, however, be positioned along a gradient
210 ranging from proximal to distal effects. At the proximal end, immediate consequences
211 stem directly from the presence and activities of invasive species, manifesting as
212 observable impacts in the short term (i.e., months to years). These initial impacts can
213 cascade through ecosystems, generating diverse and increasingly complex ecological
214 effects over time (Figure 2). For example, the introduction of a lethal pathogen can swiftly
215 reduce native populations through disease transmission causing higher mortality,
216 illustrating a direct and immediate impact. One example includes the introduction of
217 invasive amphibians that carry and spread chytrid fungus (*Batrachochytrium*
218 *dendrobatidis* and *B. salmandrivorans*) to native amphibian populations. This has
219 occurred frequently in many parts of the world, causing the extinction of native
220 populations [69–72]. However, alien or invasive parasites can have their own impacts,
221 both on local species and the invader (see Box 1). More distal impacts are subsequent
222 consequences that emerge from the cascading effects of the initial impact. The extinction
223 of a native species due to a disease can alter altered food-web dynamics. Such changes
224 can disrupt the trophic interactions and energy flow within the ecosystem, potentially
225 shifting ecosystem functions or services. For example, when native amphibian
226 populations began disappearing in Central America after the introduction of chytrid fungi,
227 the resulting loss of predation on mosquito larvae and adults caused an explosion of
228 mosquito populations; this in turn increased the incidence of pathogenic insect-borne
229 diseases such as malaria in humans living nearby [73]. Over time, these functional
230 disruptions can culminate in habitat modification because the altered processes reshape
231 the physical environment and the structure of the biotic community, disturbing the energy,
232 mass, information flow, and storage of the ecosystem.

233 The causal relationship between more proximal and distal impacts often spans
234 ecological scales, especially where a decline in the abundance of native species

235 populations (e.g., a flowering plant) can cascade to disrupt the population dynamics of
236 interacting native species (e.g., its pollinators), the structure of the community itself (e.g.,
237 diversity of insects), and even beyond to erode ecosystem function (e.g., pollination). For
238 instance, invasive plants strongly influence plant-pollinator network structure ([74], and
239 reviewed in [75]). The ecological scale at which impacts occur can also affect our
240 perception of overall impacts, because structural changes are more easily perceived at
241 broader ecosystem scales. For example, habitat degradation or changes in fire regime are
242 generally more noticeable than changes to individual fitness or behaviour, or genetic
243 changes in populations. Invasive species do not only degrade ecosystem function and
244 services ([76,77]), they can also have more subtle effects across all ecological scales.

245 Invasive species can also affect ecosystems beyond their immediate environment by
246 changing the flow of nutrients and species across boundaries (i.e., cross-ecosystem
247 interactions) [68]. The invasive willow tree (*Salix spp.*) in Australia altered riparian
248 vegetation structure, and increased leaf litter input and stream shading, reducing light
249 availability and suppressing algal growth. This shift redirected the aquatic food web
250 toward detritus-based energy pathways, leading to changes in fungal, algal, and
251 macroinvertebrate communities. As a result, algal production declined, while detritivore
252 macroinvertebrates became more dominant [78]. Although important, these cross-
253 ecosystem interactions are currently understudied and the impact categorisation can
254 support the identification of the range, connections, and breadth of these impacts
255 occurring at all scales (see [68,79]).

256

257 *Concluding remarks*

258 Our aim here is to provide a clear and standardised terminology for classifying impacts
259 across all ecological levels. We introduced a comprehensive typology encompassing 19
260 distinct types of impacts caused by invasive species, organised into six ecological levels.
261 We also differentiated these impact types from their underlying causes, emphasising the
262 ecological mechanisms through which invasive species affect native ecosystems, and
263 outlined a gradient of proximal and distal impacts that often cascade through these
264 systems. Recognising the full spectrum of these impacts and their interconnections is
265 necessary to develop effective conservation and management strategies.

266 The adoption of a standardized typology for ecological impacts has the potential to
267 improve data harmonization and interoperability across invasion biology databases and
268 frameworks. By transitioning to an impact-centred typology, researchers can standardize

269 how impacts like habitat degradation, disruption of nutrient cycling, or declines in
270 population size are recorded, without focusing solely on the identity of the invasive
271 species. Furthermore, adopting an impact-based framework could be instrumental in
272 assessing the effectiveness of global biodiversity monitoring and management initiatives,
273 such as the Kunming-Montreal Global Biodiversity Framework. Our typology provides
274 a standardized foundation for ecological impact indicators that can be tracked over time
275 across diverse ecosystems. Such a framework can also support decision-making by
276 streamlining data reporting and making invasion impacts more directly comparable. This
277 would ultimately support better prioritization of invasive species management by
278 enabling clearer assessments, including quantification, of ecological impacts of invasions.
279 Our typology can also complement existing frameworks such as the EICAT or the GISS
280 offering researchers, managers, stakeholders and others a tool to organise and
281 communicate the impacts of invasive species. We hope to standardise future research and
282 facilitate clearer definitions and distinctions across studies, ultimately advancing the field
283 of invasion biology (see Outstanding questions).

284

285 **Glossary**

286 **Alpha diversity** – The diversity of species within a specific habitat or ecosystem, often
287 measured as species richness. It represents local biodiversity and the complexity of an
288 ecosystem.

289 **Assemblage** – A group of species that coexist in the same geographical area, which can
290 vary in spatial scale from local to regional. It includes communities, which are generally
291 considered to be restricted to a specific ecosystem or habitat.

292 **Beta diversity** – The variation in species composition between different habitats,
293 ecosystems, or geographical areas.

294 **Delphi process** – A structured, iterative method for expert consensus used in research
295 and decision-making. It involves multiple rounds of anonymous surveys, where experts
296 provide input, receive feedback, and refine their responses to reach a collective
297 agreement.

298 **Gamma diversity** – The diversity of the whole region or area of interest, usually
299 measured by pooling multiple samplings in the study area; estimated with similar
300 metrics to alpha diversity.

301 **Impact (consequence)**– Any measurable change in ecological, economic, or social
302 systems resulting from an invasive species (Ricciardi et al. 2013). The typology
303 concerns only to ecological impacts.

304 **Invasive species** – An alien or non-native species that is transported beyond its natural
305 biogeographic range. When it establishes and spreads (i.e., stages of the invasion
306 process), they are usually referred as invasive species. Here we consider that any
307 species can cause impacts regardless the stage of invasion and we refer to all of them as
308 ‘invasive species’ throughout the text.

309 **Mechanism** – The process through which an invasive species exerts its impact.

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311

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313

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326 **Declaration of interests**

327 The authors declare no competing interests.

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533
 534

535 **Figure legends**

536 **Figure 1:** Nineteen impact types of invasive species categorized across six ecological
 537 levels. Each impact type is numbered and represented by an icon and label, illustrating
 538 its position within the ecological hierarchy. The arrows indicate increasing levels, from
 539 individual-level impacts to broader abiotic effects, highlighting how impacts can

540 accumulate and propagate across scales. Note that all scales are connected to each other
541 and the impacts can have multiple connections between each other.

542

543 **Figure 2:** Examples of connections between more proximal (black arrows) and more
544 distal (grey arrows) impacts of biological invasions. The colours for each impact type
545 represent one of the six ecological scales provided in Figure 1, as do the numbers
546 associated with impact types. (A) The golden apple snail *Pomacea canaliculata* reduced
547 the population of aquatic plants, which led to planktonic algae dominating the food web,
548 and consequently a shift to turbid water by released nutrients [80]; (B) The introduction
549 of brown trout *Salmo trutta* caused changes in invertebrate grazing behaviour, replaced
550 the population of nonmigratory galaxiid fish, altered crayfish and large invertebrate
551 distributions, and changed algal species assemblage structure, causing higher algal
552 primary productivity, and consequently altering nutrient flux [81].

553

554 **Box**

555

Box 1: Invasive parasites: mediating native ecological influence and invasive host impact

Biological invasions often involve many organisms. Invasive plants, vertebrates, and invertebrates can carry symbionts [82], including a microbiome (mutualistic or commensal microbes) or a pathobiome (parasites), into new environments [83]. When parasites co-invade with their invasive hosts, they might impact only their invasive host or also infect native hosts, potentially becoming ‘invasive’ themselves [84]. By affecting the health of their invasive hosts, these parasites can reduce the host’s impact on the ecosystem, acting as a form of biological control on invasive populations [84,85].

Alternatively, invasive parasites can also infect native species, posing their own set of impacts. They can adversely affect native population size, health, and ecological roles [86] and in these cases, they have impacts similar to those that invasive hosts have directly on native species. However, in those cases, the literature should be (but rarely is) clear whether the invasive parasite or the invasive host is responsible for the impact on native species. For example, the grey squirrel not only outcompetes the native red squirrel through ecological competition, but the invader also carries squirrel poxvirus, which accelerates the red squirrel's decline upon infection [39,87]. The two viewpoints, that the invasive host is only the carrier of the invasive pathogen with the impact, or that the invasive host has the impact by spreading pathogens (i.e., via apparent competition), seem equally defensible, but the distinction should be clearly made. In our typology, we propose one type that corresponds to the former (typically the first effects, at the individual level), but also to the latter (changing species interactions).

Parasites can have a positive effect on the invasive host by affecting their native competitors or other enemies more, creating an invasional meltdown during which the invasive host is helped in its invasion by its parasite [88]. All three aspects make the impacts of invasive parasites more complicated cases, calling for additional clarity in reporting their impacts, or the impacts of their invasive hosts (see Supplementary material Table S2).

556