# Hydrological Processes

# 4 Use of Spatially Distributed TOPMODEL to Assess the Effectiveness of

# 5 Diverse Natural Flood Management Techniques in a UK Catchment

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- 7 Qiuyu Zhu<sup>1</sup>, Megan Klaar<sup>1</sup>, Thomas Willis<sup>1, 2</sup>, and Joseph Holden<sup>1</sup>
- 8 <sup>1</sup> water@leeds, School of Geography, University of Leeds, Leeds, UK
- <sup>9</sup> <sup>2</sup> School of Geography and the Environment, University of Oxford, Oxford OX1 3QY, UK
- 10
- 11 Corresponding Author: Qiuyu Zhu, water@leeds, School of Geography, University of
- 12 Leeds, Leeds, LS2 9TJ, UK.
- 13 Email: gyqzh@leeds.ac.uk
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# 15 Abstract

16 While natural flood management (NFM) is becoming more widely used, there remains a lack of empirical evidence regarding its effectiveness. The primary uncertainties arise from two 17 key aspects: first, the determination of NFM effectiveness is constrained by the relatively 18 small catchment scales studied to date; second, the combination of multiple NFM 19 interventions within a catchment may lead to flood peak synchronisation. In this study, both 20 instream and terrestrial NFM interventions were modelled using a spatially distributed 21 hydrological model, Spatially Distributed TOPMODEL (SD-TOPMODEL). To demonstrate 22 how the scale and type of interventions interact to influence flood peaks, we integrated 23 various NFM interventions and land cover changes, including woodland planting, soil 24 25 aeration, floodplain restoration, and hedgerow planting. In comparison to previous versions of SD-TOPMODEL, we improved simulation efficiency to enable grid-based modelling of up 26 27 to a 200-year return period flood event for an 81.4 km<sup>2</sup> catchment with 5 m resolution. Following extensive parameter calibration and validation, the model demonstrated stability 28 and provided a reliable fit for flood peaks, achieving a Nash-Sutcliffe Efficiency coefficient 29 of up to 0.93 between modelled and observed discharge. The results highlighted the 30 31 effectiveness of NFM interventions in reducing flood peaks at the scale studied, particularly 32 during single-peaked storm events and under dry pre-event catchment conditions. Moreover,

- the combined use of multiple interventions was more effective and resilient than single
- 34 interventions, with flood peak reductions ranging from 4.2% to 16.0% in the study
- 35 catchment.
- 36
- 37 Keywords: nature-based solutions, flooding, flood peak reduction, peak delay, hydrological
- 38 modelling, sensitivity tests, parameterize, event characteristic
- 39

## 40 1. INTRODUCTION

Natural flood management (NFM) is a flood mitigation strategy which aims to work with 41 natural processes to restore and enhance catchment hydrological functions which limit flood 42 risk and impact (Dadson et al. 2017; Cooper et al. 2021). In particular, NFM seeks to reduce 43 and delay flood peaks by optimising the natural water retention function of the catchment and 44 to mitigate the potential hazards of flood peaks (Lane 2017; Kay et al. 2019; Black et al. 45 46 2021; Ellis et al. 2021; Kumar et al. 2021; Lashford et al. 2022). For example, altering the physical properties of soil that influence water movement and storage, such as porosity and 47 48 permeability, can enhance subsoil water storage capacity and encourage infiltration to delay flood peak time and/or reduce peak discharge during a storm event. Soil properties might be 49 altered directly through aeration and other soil management interventions or indirectly by 50 implementing afforestation, reducing grazing intensity or delivering other ecological 51 restoration practices (Grayson et al. 2010; Wahren et al. 2012; Palmer and Smith 2013; 52 Marshall et al. 2014; Dixon et al. 2016; Gao et al. 2016; Alaoui et al. 2018; Gunnell et al. 53 2019; Wilkinson et al. 2019; Bond et al. 2022; Monger et al., 2024). 54

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Evidence has shown that interventions such as those described above can potentially reduce 56 57 and slow overland flow by locally increasing soil saturated hydraulic conductivity, the depth of soil water table and surface roughness. For example, results of experimental studies at the 58 59 hillslope scale have shown that replacing grazed grassland with broadleaf woodland on hillslopes significantly increases saturated hydraulic conductivity and provides the soil with 60 61 increased capacity to store rainfall by reducing soil compaction and bulk density and increasing depths of soil water table (Marshall et al. 2009; Archer et al. 2013; Murphy et al. 62 63 2020), thus achieving a reduction in overland flows (Marshall et al. 2014; Bond et al. 2022). 64 Modelling studies by Gao et al. (2016) and Goudarzi et al. (2021) highlighted that upland 65 peat restoration, including both revegetation and gully blocking interventions, are effective in increasing static and kinematic storage of rainfall in the implementation area to reduce and 66 delay flood peaks, and such evidence has been supported in a field experimental study 67 (Shuttleworth et al. 2019). Critically, the reduction of overland flow velocities by increasing 68 surface roughness can yield reduced discharge peaks (Holden et al. 2008; Roni et al. 2015; 69 Bond et al. 2020; Bond et al. 2022). Soil aeration, which is the process of mechanically 70 71 piercing the soil to enhance porosity, has also been shown to be effective in enhancing

72 infiltration to reduce overland flow by increasing topsoil saturated hydraulic conductivity

73 (Franklin *et al.*, 2007; Alaoui *et al.* 2018; Wallace and Chappell, 2019).

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While a few NFM studies have been based on field experimental data these have been at the 75 plot or small catchment scale (Kay et al. 2019; Kumar et al. 2021; Zhu et al. 2024). The 76 majority of studies of NFM effectiveness have been conducted using modelling and have 77 78 focused on a specific NFM intervention. Most information from modelling studies is limited 79 by the lack of understanding of the upscaling impacts from the local scales of intervention 80 implementation to catchment-scale, and the modelling approaches vary in accuracy and complexity (Kumar et al. 2021; Hill et al. 2023). From previous modelling studies of 81 multiple NFM interventions, conflicting conclusions have been obtained for different 82 modelling approaches, resolution, and catchment scale. For example, in a study of the 98 km<sup>2</sup> 83 Lymington River catchment in southern England, the simplified spatially distributed 84 OVERFLOW model (20m resolution) showed up to a 20% peak reduction by increasing 85 mature floodplain forest to 20-35% of the area, with greater effects from additional 86 reforestation and sub-catchment desynchronisation (Dixon et al. 2016). Metcalfe et al. (2018) 87 used the semi-distributed Dynamic TOPMODEL with NFM interventions which enhanced 88 hillslope storage lumped in several hydrological response units (HRUs) for a 223 km<sup>2</sup> 89 catchment, reducing peak flow by a median of 5.8% and a maximum of 17.3% during one 90 91 storm. The same model was coupled with the floodplain hydrodynamic model, JFlow (2 m resolution) to assess land management and peatland restoration interventions, including a 92 93 runoff attenuation feature (RAF) in a 15 km<sup>2</sup> catchment, showing a 4%  $\pm$  2% flow reduction (Hankin et al. 2019). Dynamic TOPMODEL coupled with HEC-RAS 2D (5 m resolution) 94 95 showed up to 25% surface flow reduction from a combination of afforestation and in-channel 96 barriers in an 18 km<sup>2</sup> catchment (Ferguson and Fenner 2020). The Generalized Multistep Dynamic TOPMODEL (2 m resolution) was employed to model peatland restoration 97 scenarios in a 25 km<sup>2</sup> catchment, showing a high likelihood of > 5% peak discharge reduction 98 (Goudarzi et al., 2024). Although these modelling results all demonstrate effective flood 99 mitigation by NFM, the differences in implementation of interventions in each catchment, 100 such as intervention types, locations and area, and the differences in the complexity, spatial 101 resolution and modelling scale of each model make it difficult to compare NFM benefits 102 under the same criteria. 103

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105 Using modelling to understand the impacts of types, location and size of interventions, and the combination of interventions, is important for NFM research (Bond et al. 2022; Hill et al. 106 107 2023; Kingsbury-Smith et al. 2023; Peskett et al. 2023; Monger et al. 2024). Modelling could 108 be used to support investment decisions, especially via the assessment of NFM from an integrated catchment perspective with single and mixed types of interventions. The number 109 of UK NFM studies that have investigated multiple interventions and achieved valid flood 110 mitigation results (10 articles) is less than those that have only investigated a single 111 intervention (24 articles) (Zhu et al. 2024). Implementation of single or a combination of 112 interventions can be effective at larger catchment scales than the evidenced 20 km<sup>2</sup> limit 113 proposed by Dadson et al. (2017) and increases the resilience of NFM by combining it with 114 other forms of flood management interventions (e.g. leaky dams and runoff attenuation 115 features) (Black et al. 2021). All NFM interventions have the potential to influence the 116 synchronisation of flood peaks across tributaries in the catchment (Thomas and Nisbet, 2007; 117 Pattison et al., 2014). During large storms, multiple interventions yielded less peak reduction 118 and no peak timing impacts compared to a single intervention due to the increased likelihood 119 of peak synchronisation (Dadson et al. 2017; Kingsbury-Smith et al. 2023; Metcalfe et al. 120 2018). However, empirical evidence addressing positive or negative effects of the type of 121 122 combined interventions, location, and area on NFM effectiveness at a large catchment scale remains very scarce. 123

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Ideally, a modelling study for NFM effectiveness requires a model with sufficient spatial 125 126 resolution to simulate all land and soil management interventions yet which is simplified enough to efficiently simulate storms with different characteristics at larger catchment scales. 127 128 The Spatially Distributed TOPMODEL (SD-TOPMODEL) used in this study meets these 129 requirements (Gao et al., 2015). SD-TOPMODEL has been used to demonstrate flood peak reductions varied across catchments for different types NFM interventions. Peatland 130 revegetation scenarios tested (20 m resolution) in an 84 km<sup>2</sup> upland catchment with peatland 131 headwaters showed a 4 – 15% reduction in flood peaks (Gao et al. 2017). Hillslope grassland 132 management scenarios achieved up to 42% reduction in overland flow peaks in a 133 predominantly grassland-covered upland catchment (21 km<sup>2</sup>), where afforestation 134 intervention was most effective (Bond et al. 2022). Woodland planting scenarios 135 demonstrated up to 15.3% reduction in flood peaks in a 2.62 km<sup>2</sup> steep upland catchment 136 predominantly covered by unimproved grassland and semi-natural woodland (Monger et al. 137

2024). Kingsbury-Smith et al. (2023) used the model in a 38 km<sup>2</sup> predominantly rural upland 138 catchment for several single intervention scenarios, such as woodland and hedgerow planting, 139 riparian buffer strips, and soil aeration, and a combination of all these interventions. SD-140 TOPMODEL has been validated to effectively simulate rainfall-runoff processes in steep 141 upland catchments with land and soil management measures (Gao et al. 2018; Bond et al. 142 2022), and the resolution has been improved from 20 m to 5 m to enable representation of a 143 144 wider variety of NFM interventions (Bond et al. 2022; Kingsbury-Smith et al. 2023). However, further testing of different combinations of NFM interventions, storm 145 146 characteristics and catchment antecedent conditions by using SD-TOPMODEL at larger catchment scales is still required, as SD-TOPMODEL has not been applied to catchment 147 scales  $> 50 \text{ km}^2$  at a fine resolution of 5 m. 148

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Although no significant correlation between rainfall intensity and NFM effectiveness was 150 identified in both our review and previous modelling studies, the response of different NFM 151 intervention types to varying rainfall intensities remains variable (Gao et al. 2018; Ferreira et 152 al. 2020; Kingsbury-Smith et al. 2023; Zhu et al. 2024). Notably, there is a research gap 153 regarding the potential impacts of different rainfall event characteristics, such as single-154 155 versus multi-peaked events, in addition to rainfall intensity (Hankin et al. 2020). Furthermore, catchment antecedent conditions may influence NFM effectiveness (Wallace 156 157 and Chappell, 2019; Bond et al. 2020). The influence of these factors on NFM effectiveness have been investigated in this modelling study. 158

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To investigate and validate whether SD-TOPMODEL can accurately represent and model 160 161 soil infiltration rates, soil storage capacities and surface roughness in the catchment as a result of NFM implementation, three main NFM interventions in the catchment were selected 162 163 for this study. They are afforestation (woodland planting), soil aeration and hedgerow planting. In this study, SD-TOPMODEL is used to simulate these NFM interventions 164 impacting soil hydrological functions and surface roughness during high-flow events at a 165 catchment scale. The aim is to calibrate and validate the model with parameter choices 166 supported by evidence from previous empirical studies and improve model performance to 167 gain a full understanding of NFM effectiveness. The land cover and NFM scenarios were 168 applied to compare different combinations of NFM interventions and how they interacted 169 with seven observed storms with different characteristics at a catchment scale. 170

# 172 **2. DATA AND METHODS**

# 173 **2.1 Study Site**

174 The Upper Aire catchment is a 370.8 km<sup>2</sup> upland catchment in northern England (Figure 1).

- 175 Following extensive flooding in 2015, which exceeded a 200-year return period at the
- 176 Armley gauging station (located in the city of Leeds, Figure 1b), multiple NFM interventions
- 177 (woodland planting, hedgerow planting and soil aeration) have been implemented to slow
- 178 runoff response to rainfall and increase surface and subsurface water storage in off-channel
- areas across the catchment (Figure 1c) (Leeds City Council, 2024; Upper Aire Project, 2024).
- 180 The Upper Aire Catchment to Gargrave ( $\sim$ 81.4 km<sup>2</sup>) was chosen as the study site, as it
- 181 encompasses a significant proportion of these interventions which also have well-documented
- records about their locations and nature (Yorkshire Wildlife Trust, 2022). The catchment also
- 183 benefits from reliable river gauging data at Gargrave. The Gargrave station is located on the
- 184 main channel of the River Aire and its catchment area is dominated by a rural hilly landscape
- 185 with a maximum elevation difference of 467 m.



Figure 1 Study site: (a) location of study catchment in the UK; (b) Aire and Calder
catchment; (c) study area upstream of Gargrave gauging station including land cover data
from the UK 2015 landcover map, and locations of NFM interventions.

The catchment is dominated by improved grassland (49.7%), followed by calcareous 191 grassland (18.7%), acid grassland (12.1%) and heather grassland (10.4%), bog (3.7%), 192 broadleaf woodland (1.9%) and coniferous woodland (1.0%) (CEH, 2015). Soils are mainly 193 fine or coarse loamy, slowly permeable loamy and clayey, very shallow loamy, or well 194 drained silty soils over limestone (NATMAP, 2016), resulting in high spatial variability in 195 soil depth. The underlying geology is dominated by carboniferous limestone, along with 196 sandstone, mudstone, and shale (NATMAP, 2016). The area experiences a mean of 220 days 197 of rain, with 1510 mm of mean precipitation annually (Upper Aire Project, 2024). 198 199

# 200 2.2 Spatially distributed rainfall-runoff model: SD-TOPMODEL

- 201 To investigate the influence of land cover changes and NFM interventions on flood response
- at the catchment scale, SD-TOPMODEL was used (Gao et al. 2015). SD-TOPMODEL is a

203 spatially distributed version of the original lumped or semi-distributed TOPMODEL (Beven and Kirkby 1979). Gao et al. (2015) developed the model by downscaling the original 204 TOPMODEL equations from the catchment scale to grid cell equations. It has the advantage 205 of allowing each grid cell to be saturated at different times based on the local wetness by 206 207 using precipitation, slope, and soil water depth in each cell. The overland flow module uses the multiple-direction flow theory of Quinn et al. (2006) with a dynamic velocity parameter 208 209 related to surface roughness to conduct overland flow directions and rates in each grid cell. This facilitates the representation of hydrological variability across the land surface and 210 211 shallow subsurface conditions by adjusting parameters within each grid cell. This highresolution capability enables the inclusion of spatially specific NFM interventions. The model 212 is also well suited to simulate extreme rainfall events in catchments with steep topography 213 and shallow soils (Gao et al. 2015; Gao et al. 2016; Bond et al. 2022; Kingsbury-Smith et al. 214 2023; Monger et al. 2024), thus, is ideal for use in the Upper Aire catchment for the NFM 215 216 effectiveness study.

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218 SD-TOPMODEL can deliver three catchment outlet outputs at each timestep: overland flow, shallow subsurface flow, and the total of overland and subsurface flow outputs. Three key 219 220 parameters are employed in SD-TOPMODEL to represent the catchment physical properties: overland flow velocity, Kv equals 1/n where n is the surface roughness, soil hydraulic 221 222 conductivity Ks, and soil active water storage depth m (Gao et al., 2015). To increase the efficiency of SD-TOPMODEL for simulating extreme rainfall events, the model is written in 223 224 C++ language and was batch run on the ARC High-Performance Computer (HPC) platform 225 at the University of Leeds.

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# 227 2.2.1 Data Sources

An Ordnance Survey 5 m digital terrain model (DTM) was used for the Upper Aire
catchment (Ordnance Survey, 2022). The 2015 England Land Cover Map (CEH, 2015) and
the National Soil Map (NATMAP) were provided as vector datasets at the same resolution (5
m) to represent land cover, vegetation, and soil types. The resolution used was the highest
possible as determined by data availability and limitations of model run time (maximum 48
hrs of HPC platform runtime).

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235 For rainfall data, 15-min measured precipitation (mm) from the Malham Tarn station (Figure 1) from January 2012 to January 2022was used (Environment Agency, 2022). 15-minute 236 observed river flow (m<sup>3</sup>/s) was obtained from the Gargrave and Armley gauge over the same 237 period (Environment Agency, 2022). Within the ten-year dataset, storm events were selected 238 using the POT method in Extreme Value Analysis (EVA) (Leadbetter, 1991), identifying 239 high-flow events, and including multiple occurrences within a year. The Python package 240 'pyextremes' (https://georgebv.github.io/pyextremes/) was utilized as a selection tool for this 241 analysis, while return periods were calculated at the Armley station. 15 discrete flood events 242 were initially identified, each exceeding the discharge threshold (40 m<sup>3</sup>/s) and having a time 243 interval of more than seven days since the preceding rainfall event. We selected seven of 244 these flood events that occurred in different months with varying rainfall intensities, 245 durations, and return periods for observations at the Armley gauging station (which was used 246 due to its urban fluvial flood risk location), which covered very common, common, 247 uncommon, and rare flood events in the catchment (Table 1). Every event was initialised with 248 a base flow derived from discharge data, which served as the overland flow input into the 249 model. A 5-hour warmup runtime was required for each grid cell to reach water balance 250 within the catchment. To represent base flows under different antecedent conditions, the 251 252 warm-up period incorporated either 0 mm per timestep (for dry conditions) or 0.2 mm per timestep (for wet conditions). Dry and wet antecedent conditions were defined from soil 253 254 moisture reports (COSMOS-UK, 2020) and the Hydrological Summary for the UK (CEH, 2012-2020). This helped to categorise and understand the differences in flood mitigation 255 256 effectiveness of different NFM intervention types for flood events with different antecedent 257 characteristics.

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# 259 2.2.2 Model 0 and calibration

260 To establish a Model 0 (baseline model to start calibration) with optimal and spatially uniform parameter settings for the catchment, model calibration was conducted for each of 261 the seven observed storm events. Each model run had a 5-hour spin up before the rainfall 262 event and finished after discharge observations returned to base flow levels, with a timestep 263 of 15min and a cell size of 5 m \* 5 m. To enhance efficiency by reducing the number of 264 calibration runs and computing time, parameter spaces were selected based on prior 265 experience with SD-TOPMODEL testing and calibration, as outlined in Table 1a (Gao et al., 266 2015; Gao et al., 2016; Kingsbury-Smith et al., 2023). For calibration, 150 simulations were 267

268 conducted for each event, totalling 1050 simulations across seven events, using varying parameters with the intervals specified in Table 1. In previous studies which have applied 269 SD-TOPMODEL, the Nash-Sutcliffe efficiency coefficient (NSE) was used as a criterion for 270 evaluating and selecting the best performing model (Bond et al. 2022; Gao et al. 2017; 271 Kingsbury-Smith et al. 2023; Monger et al. 2024). The best simulation and parameter setting 272 for Model 0 were determined for each event based on the highest NSE value and minimal 273 differences in flood peak discharge and timing compared to observations (Table 1b). 274 Following this, despite some discrepancies in the best-fit parameters for Events 2, 6 and 7, 275 276 the parameters can be constrained to the following ranges: m (0.006-0.01); Ks (100-200); Kv (25-30). A generic parameter setting that can be used to represent the entire catchment was 277 derived through 100 Monte Carlo tests run for each of the seven events using the narrowed 278 range. The model was considered credible when NSE values exceeded 0.5 (Moriasi et al. 279 2007). However, NSE has been shown to be potentially insensitive to low and peak flows in 280 assessing model errors (Althoff and Rodrigues, 2021), thus, the absolute peak error and peak 281 error percentage were also considered. This generic parameter setting serves as Model 0 282 which does not account for the spatial distribution of land cover or variations between event 283 years resulting from NFM implementation and land cover changes. Model 0, with uniform 284 285 parameters, achieved NSE values exceeding 0.6 for all events, with peak discharge errors limited to no more than 32% of observed values. The best-fit model during baseline 286 287 calibration attained an NSE of 0.92, aligning with the performance reported in previous studies (Bond et al. 2022; Gao et al. 2017; Kingsbury-Smith et al. 2023; Monger et al. 2024). 288 289 Once the parameter setting was determined, all subsequent model runs used this set of 290 baseline parameter values.

292 Table 1 Calibration parameters and results: a) parameters spaces used for narrow ranges

for calibration; b) calibration results of Model 0 for seven observed storm events (the date is

294	when	the	storm	started)

	Parameter	Lower li	mit	Upper limit	In	itervals	Model 0		
	Kv (-)	20		40		5	30		
	<i>m (m)</i>	0.002		0.018		0.004	0.009	)	
	Ks (m/h)	100		600		50	100		
295									
		Event 1 (12 Dec 2015)	Event 2 (3 Jan 2012)	Event 3 (25 Dec 2015)	Event 4 (8 Feb 2020)	Event 5 (10 Oct 2019)	Event 6 (16 Mar 2019)	Event 7 (15 Feb 2020)	
	Return Period in Armley Station (~years)	50	10	> 200	100	1	20	20	
	Storm duration (h)	16 (single- peaked)	60 (multi- peaked)	34 (multi- peaked)	50 (single- peaked)	25 (multi- peaked)	18 (single- peaked)	23 (multi- peaked)	
	Total rainfall (mm)	22.4	64.0	89.6	70.2	37.0	44.6	49.2	
	Rainfall intensity (mm/h)	1.400	1.067	2.635	1.404	1.480	2.478	2.319	
	Maximum rainfall intensity (mm/h)	6.4	4.2	5.6	7.6	5.6	4.2	5.2	
	Simulation duration (h)	45 (Wet)	90.5 (Dry)	) 65 (Wet)	55 (Dry)	41.25 (Dry)	35 (Dry)	37.5 (Wet)	
	Kv (-)	25	35	25	25	30	25	40	
	<i>m</i> ( <i>m</i> )	0.006	0.006	0.006	0.01	0.01	0.014	0.01	
	Ks (m/h)	200	400	200	200	100	200	200	
	NSE of best fit	0.93	0.80	0.87	0.84	0.83	0.96	0.86	

# 297 2.2.3 Data for determining parameter values

The model was applied for different NFM interventions. This study attempts to calibrate threeparameters of SD-TOPMODEL to larger catchment scales and more complex combinations

300 of existing NFM interventions. There are two formats to input parameters in SD-

- 301 TOPMODEL used in this study: (1) a parameter file which displays baseline values of each
- parameter conducted by Model 0 calibration (see section 2.2.2); (2) a spatially distributed

map for all three parameters based on land cover and NFM interventions to apply upon the

- 304 baseline values.
- 305

306 To represent the spatially distributed parameters, each parameter was individually spatially

307 distributed on the base map to assign the parameter to each grid cell at the same resolution as

the elevation data and the entire model. During the model run, the values within each grid cell

309 of each parameter distribution map were calculated as multipliers on the baseline values (as

the improved grassland land cover, due to its dominance in the catchment) (Bond et al., 2022;

Kingsbury-Smith et al., 2023; Monger et al., 2024). The calibration ranges for these values

were determined with reference to previous empirical studies (Table 2). The upper and lower

313 limits specified by the measured ratios allow the model to be more realistic, but also

314 constrained by the physical meaning of them.

*Table 2 Parameter multipliers used for land cover types and existing NFM interventions in the catchment and the ratio of them compared with* 

*improved grassland with references (from previous field measurements and experimental results), and the proportional area of land cover types* 

*in 2015 and 2020, as well as the proportional area of implemented NFM interventions.* 

Class name	Modelled	parameter ra	nges	Ratio of land cove	er type compared with improv	ed grassland from	Percentage of	Percentage of	
	(multiplie	rs)		literature		area (2015,	area (2020, 25m		
	<i>m (m)</i>	Ks (m/h)	Kv	М	Ks	Kv	25m raster)	raster)**	
Broadleaf	0.0135 -	200 -	12 - 22.5	1.5 - 2 (Archer et	1.8 (Murphy et al. 2020)	1.3 (Monger et al.,	2.09%	1.87%	
woodland	0.018	1000	(0.4 -	al., 2013)	2.4 (Marshall et al. 2009)	2022b)			
	(1.5 – 2)	(2 – 10)	0.75*)		12 (Gonzalez-Sosa et al.	2.4 (Manning's n			
					2010)	values, Chow,			
					11 - 20 (Monger et al.	1959)			
					2022a)				
Coniferous	0.0135 -	150 - 500	<b>150 - 500 12 - 22.5 1.2 - 5.6</b> (Arche		1.2 – 5.6 (Archer <i>et al.</i>	-	0.71%	1.04%	
woodland	0.018	(1.5 – 5)	(0.4 -		2013)				
	(1.5 – 2)		0.75*)		2.1 (Gonzalez-Sosa et al.				
					2010)				
					3.9 (Kingsbury-Smith et				
					al. 2023)				
					≤38 (Chandler <i>et al.</i> 2018)				

Arable and	0.009 (1)	50 (0.5)	22.5 (0.75)	The same as	0.5 (Holden et al. 2019)	The same as	0.34%	0.11%
horticulture				Baseline		Heather		
				(NATMAP)				
Improved	0.009 (1)	100 (1)	30 (1)	Baseline			58.78%	48.71%
grassland								
(baseline)								
Calcareous	0.009 (1)	200 - 500	15 - 24	The same as	4.6 (Gonzalez-Sosa et al.	1.4 (Monger et al.,	12.07%	18.66%
grassland		(2 – 5)	(0.5 - 0.8)	baseline	2010)	2022b)		
				_ (NATMAP)	≤4.9 (Kingsbury-Smith <i>et</i>	1.2 – 2.8 (Bond et		
Acid grassland	0.009 (1)	200 - 500	15 - 24		al. 2023)	al., 2020)	17.54%	12.11%
		(2 – 5)	(0.5 - 0.8)		≤10 (Monger <i>et al.</i> 2022b)	1.4 - 4		
					Significantly greater	(Manning's n		
					(Bond et al. 2021)	values, Chow,		
						1959)		
Heather	0.009 (1)	100 (1)	22.5 (0.75)	The same as	The same as baseline	0.75 (Manning's n	0.32%	0.00%
Heather	0.009 (1)	100 (1)	22.5 (0.75)	baseline		values, Chow,	7.73%	10.41%
grassland				(NATMAP)		1959)		
Bog	0.009 (1)	350 (3.5)	15 (0.5)	The same as	3.3 – 4.2 (Holden <i>et al</i> .	1.9 – 2.3 (Holden	3.53%	3.66%
				baseline	2007)	et al., 2008)		
				(NATMAP)				

Inland rock	0.0045	100 (1)	22.5 (0.75)	0.5 (NATMAP)	(NATMAP) The same as baseline		0.74%	0.66%
	(0.5)				(Medici et al. 2019)	values, Chow,		
						1959)		
Freshwater	0.0009	500 (5)	0.3 (0.01)	-	-	-	0.96%	0.89%
	(0.01)							
Urban	0.0009	1 (0.01)	150 (5)	-	-	Set as the highest	0.03%	0.07%
Suburban	(0.01)					value	0.71%	0.86%
NFM	Paramet	er multipliers		Ratio of NFM inte	rventions compared with im	proved grassland with	references	Percentage of
interventions	m (m)	Ks (m/h)	Kv	M	Ks	Kv		area
Hedgerow	0.009	200 - 1000	15 (0.5)	The same as basel	ine 2 (Kingsbury-Sr	nith <i>et al.</i> $1.6 - 2.3$	(Manning's n	1.30%
	(1)	$(2 - 10^*)$			2023)	values, Cl	how, 1959)	
					2-6 (Holden <i>et</i>	al. 2019)		
					22.5 – 27.7 (Wa	llace <i>et al</i> .		
					2021)			
Woodland (new	1.5 - 2	150 - 250	12 - 22.5	The same as wood	lland 1.2 (45-yr-old w	oodland, The same	as woodland	0.61%
planted)		(1.5 - 2.5)	(0.4 - 0.75)	land cover	(Archer et al. 20	(13)) land cove	r	
					2.3 (18-month-o	ld		
					saplings, (Mawd	lsley <i>et al</i> .		
					2017))			

			_		2.4 (7-yr-old broadleaf		
					woodland, (Marshall et al.		
					2009))		
Soil aeration	0.009 -	100 - 500	30(1)	1-1.5 (Willis and Klaar,	2.5 – 3 (Chehaibi <i>et al.</i>	The same as baseline	6.40%
	0.0135	(1 – 5)		2021; Kingsbury-Smith et	2010)		
	(1 - 15)			al 2023)	1 - 15 (Wallace and		
	(1 - 1.5)			u., 2023)	1 15 (Wanace and		

319 \*Bold numbers are the ranges for parameter sensitivity tests.

320 \*\*The classification of subgroups in woodlands and grasslands in CEH Land Cover Map partially changed from 2015 to 2020. Confusion and

321 misclassification within grassland classes will also occur (CEH, 2020)

### 322 2.2.4 Calibration for land cover, NFM interventions and model validation

This section describes the process of incorporating land cover and NFM interventions into the 323 hydrological model through the spatial distribution of parameter multipliers, which were 324 calibrated using an observed event (Event 1) based on the Model 0. Land cover (based on 325 CEH land cover map 2015) included broadleaf woodland, coniferous woodland, calcareous 326 grassland, and acid grassland, and NFM scenarios consisted of three implemented 327 interventions in the research catchment: hedgerow planting, woodland planting, and soil 328 aeration. Spatially distributed multipliers of each parameter (Table 2) were applied to Model 329 330 0 (uniform parameters) to represent these scenarios.

331

The multipliers summarised in Table 2 showed high variability in woodland and grassland 332 land cover types. The range of parameter choices were informed by the results of the previous 333 empirical studies while also considering the baseline values for the study catchment and the 334 limit of parameter settings in the model. Event 1 from 2015 was used for parameter testing, as 335 the shape of hydrograph for this event is characterized by concentrated rainfall and a single 336 flood peak and is particularly relevant for parameter testing since the model's land cover map 337 also corresponds to 2015, making it the most representative event. Sensitivity tests using 338 339 Event 1 were conducted within the valid parameter test ranges detailed in Table 2. For land cover types, the aim was to identify the most accurate set of parameter multipliers for each 340 341 type, while for NFM interventions, it was to identify the most sensitive parameter set.

342

343 Sensitivity tests were conducted using fixed intervals where parameters were selected and paired within the multiplier test range for five sets of tests (Table 2): (1) in land cover tests, 344 345 Ks values were tested at intervals of 2, which paired with Kv values by intervals of 0.1 and 0.15 for woodlands, with *m* tested at values of 1 and 1.5; (2) in grassland scenarios, Ks was 346 tested by intervals of 1 and Kv at intervals of 0.1; (3) in NFM scenarios, Ks values were 347 tested at intervals of 2 to represent hedgerow planting; (4) for woodland planting, Ks was 348 tested at intervals of 0.5 and Kv at intervals of 0.1 and 0.15, while m was tested at values 349 between 1 and 1.5; (5) soil aeration scenarios involved testing m at intervals of 0.25 and Ks at 350 351 intervals of 1. The land cover tests identified parameter multipliers that optimised model accuracy for each grassland and woodland type by assessing the correlation with NSE values. 352 After calibrating the multipliers for the 2015 land cover map to achieve the best NSE values 353 for Event 1 (closest to 1), all calibrated values were integrated into the land cover model. For 354

355 the NFM intervention tests, calibration prioritised maximum flood peak reduction with minimal NSE decrease to ensure intervention effectiveness and maintain the rainfall-runoff 356 relationship established in Model 0, thereby preserving model reliability. Optimal multipliers 357 were applied to represent single and combined NFM scenarios, calibrated to achieve 358 maximum flood mitigation without compromising model accuracy. Final parameter settings 359 are shown in Table 3. Pearson correlation analysis, conducted in SPSS after confirming the 360 normal distribution of the dataset, was used to assess the relationship between NSE and 361 parameter values in land cover tests, and between peak reductions and parameter values in 362 363 NFM intervention tests. Correlation strength was evaluated using Cohen's guidelines (r > 0.5= notable effect) (Cohen, 1988). Correlation analysis was employed to assess the model's 364 sensitivity to the parameters and to explore the pattern of SD-TOPMODEL's response to 365 parameter multiplication for adding spatially distributed land cover and NFM interventions to 366 the spatially uniform Model 0. This analysis identified parameter sensitivity patterns, reduced 367 model uncertainty, and prevented multiple parameter choices from yielding similar model 368 performance. 369

370

Land cover class	Parame	ter multiplier	°S	
	m	Kv	Ks	
Broadleaf woodland	1.5	0.75	8	
Coniferous woodland	1.5	0.75	4	
Improved grassland (baseline)	1	1	1	
Calcareous grassland	1	0.8	3	
Acid grassland	1	0.8	3	
NFM interventions				
Hedgerow	1	0.5	10	
Woodland (new planted)	1.5	0.6	2.5	
Soil aeration	1.5	1	4	

# 371 *Table 3 Final parameter multiplier settings after sensitivity testing*

373 For model validation, discharge results after land cover was applied to the spatially uniform Model 0 were compared with observed flow rates from the Gargrave gauging station for the 374 seven events used for calibration and three additional events (two with wet conditions and 375 one with a dry condition) from the 15 discrete flood events identified above. Validation 376 377 involved calculating NSE values (Table 4) and assessing the fit of flood peak discharge and arrival times. Following validation, NFM scenarios were applied to the land cover model, and 378 379 flood peak reduction effectiveness was assessed by comparing results with the land cover model. A radar plot quantified the area and flood mitigation effects of each scenario using 380 quantile comparisons. Results were grouped by event characteristics (single/multi-peaked and 381 dry/wet catchment conditions) and scenario characteristics (single/multiple interventions) to 382 evaluate their impacts on flood mitigation effectiveness. 383

Table 4 NSE values and peak fits between model outputs and observations for the seven storm events used in model calibration and three
additional events used for model validation (SD = standard deviation)

										Validation 1	Validation 2	Validation 3
Scenarios	Event 1	Event 2	Event 3	Event 4	Event 5	Event 6	Event 7	Mean	SD	(31 Dec	(02 Jan	(14 Oct
										2012, wet)	2015, dry)	2017, wet)
NSE (Model 0)	0.92	0.86	0.90	0.63	0.73	0.60	0.78	0.78	0.13	0.64	0.96	0.95
NSE (Land	0.93	0.77	0.82	0.68	0.66	0.59	0.70	0.74	0.12	0.61	0.97	0.91
cover model)												
Percentage of												
peak	9%	-4 %	12%	24%	5%	32%	10 %	13%	12%	-14%	-5%	-7%
differences* of	<i>y</i> / 0	170	12/0	21/0	570	5270	10 / 0	1070	12/0	11/0	270	,,,,
Model 0 (%)												
Peak time												
difference of	0.25	-0.25	1	-0.25	-0.75	-2.25	-0.5	-0.39	1.00	-1	-0.5	-0.25
Model 0 (h)												
Percentage of												
peak												
differences* of	-2%	-5%	-8%	15%	5%	26 %	2%	5%	12%	-21%	-8%	-10%
land cover												
model (%)												
Peak time	0.25	0	1.25	0	-3	-3	-0.5	-0.71	1.65	-1	-0.5	-0.25
difference of	-		-		-	-					-	-

land cover												
model (h)												
Overland flow												
peak arrival												
time different	0.25	0.25	1.25	0	-0.5	-2	-0.25	-0.14	0.99	-0.5	-0.25	-0.25
of land cover												
model (h)												

387 \*Peak discharge here is the sum of subsurface and overland flow. Percentages are the differences between the modelling peaks compared to the observed peaks

as a percentage of observed peaks.

**390 3. RESULTS** 

391 **3.1 Sensitivity tests** 

392 3.1.1 Correlation analysis for model calibration – land cover and NFM interventions

393 The sensitivity of different land cover types to the three parameters' multipliers varied, and a

significant correlation existed between *Ks*, *Kv* and model performance. As shown in Figure 3

(a), surface roughness Kv was significantly positively correlated with NSE values (r = 0.918,

p < 0.001) as the NSE values increased with the increasing *Kv* values, but not with soil

hydraulic conductivity Ks (r = 0.303, p = 0.195). NSE values increased until Ks multiplier

reached 6 after which they were no longer affected by *Ks*. This indicates that for woodland

land cover types, including broadleaf woodland and coniferous woodland, the model fit

400 improved with a marginal increase in surface roughness. Thus, the best fit of the model (the

401 maximum NSE values) for woodland land cover types was achieved when *Kv* decreased by

402 0.75 times and *Ks* increased by 8 times (as highlighted in Figure 2a). Multiplier values of 1.5

403 or 2 for soil active water storage depth m did not affect the correlation patterns between Ks,

404 *Kv* and model fit. Choosing the multiplier of 1.5 reduced the impact of model performance

405 caused by parameter changes, while a multiplier of 2 increased the standard deviation of NSE406 results in general.



Figure 2 The model parameter multipliers: surface roughness Kv and soil hydraulic
conductivity Ks, plotted with NSE difference from Model 0: (a) woodland, (b) grassland. Note
the differences in axes limits between figures.

412 As highlighted in Figure 2 (b) for grassland land cover types (calcareous grassland and acid grassland), Ks (r = -0.650, p < 0.01) was more strongly correlated to NSE values than Kv to 413 414 NSE (r = 0.625, p = 0.01). The correlation between Ks and NSE tended to arc and had a 415 relative peak when Ks multiplier was 3. The relationship between Kv and NSE values were consistent with the above results of woodland tests, i.e., increases in surface roughness led to 416 better model fit. Thus, model fit of grassland land cover types were significantly correlated 417 with both Kv and Ks. The best model fit was achieved when Ks increased to 3 times that of 418 Model 0 and Kv increased marginally from Model 0. Similar to the woodland results, m had 419 no significant correlation to the model fit of grassland scenarios. Overall, the parameter 420 multipliers for woodland and grassland scenarios were chosen using the highest NSE values 421 422 achieved in the sensitivity tests.

423

For the NFM interventions, the same sensitivity tests were used to find the most effective
combination of parameter multipliers to represent three NFM interventions. All NFM
scenarios were applied upon the land cover, and their test results were compared with the
Model 0 results. As shown in Figure 3, none of these parameter tests caused a significant

428 decrease in the NSE values from Model 0, indicating that the test did not affect model

429 performance. Thus, correlation analysis for NFM scenarios was focused on comparing the

430 effects on flood peaks.

431



m

1.00

1.25

Figure 3 The model parameters: surface roughness Kv and soil hydraulic conductivity Ks
plotted with NSE difference and peak discharge difference from Model 0: (a) woodland
planting, (b) soil aeration. Note the differences in axes limits between figures.

435

Among the three NFM interventions, there was no significant pattern among Ks, NSE values, 436 peak discharge differences and peak arrival time differences for hedgerow planting. The 437 greatest flood peak reduction occurred when Ks reached 10 times the baseline value, 438 suggesting this as the most effective multiplier for hedgerows. For woodland planting, test 439 results were first compared between m multiplier of 1.5 and 2. While the multiplier of 2 440 reduced mean flood peak discharge by  $0.58 \text{ m}^3/\text{s}$  and delayed mean peak arrival time by 0.02441 h, it increased variability as the standard deviation increased. To maintain consistency with 442 the land cover tests described above, the multiplier was chosen as 1.5 to represent newly 443 planted woodlands. Kv in woodland planting was strongly correlated with peak discharge 444 reduction (r = 0.900, p < 0.001), which indicates that the increase in surface roughness 445 446 significantly reduced peak discharge, while Ks changes had no significant impacts on peak reductions (r = -0.119, p = 0.712). When the multiplier of Ks was taken as 2.5, using a Kv447 multiplier of 0.5, 0.6 and 0.7 all increased the NSE values, with the multiplier of 0.6 resulting 448

- in the greatest reduction in peak discharge. For woodland planting, changes in *Kv* and *Ks* had
- 450 no significant impact on peak arrival delay. For soil aeration, all parameter pairs increased the
- 451 NSE value. The most effective flood reduction test results were chosen (Figure 3b). An
- 452 increase in m significantly reduced peak discharge (r = -0.619, p = 0.014), and an increase in
- 453 *Ks* slightly reduced peaks (r = -0.646, p < 0.01) (Figure 3b). Finally, a parameter pair that
- 454 maximized flood reduction and maximized the NSE value was chosen for soil aeration. Final
- 455 parameter multipliers for land cover and NFM interventions are shown in Table 3.
- 456

# 457 3.1.2 Sensitivity of peak changes – land cover and NFM interventions

For peak discharge reduction and arrival time delay in land cover tests, *Kv* was significantly 458 correlated with peak discharge reduction for woodland (r = 0.959, p < 0.001) and grassland 459 land cover types (r = 0.750, p < 0.001), but not with peak arrival time. Ks and m showed no 460 correlation with peak changes. This indicates that surface roughness increases via woodland 461 and grassland land cover significantly reduced peak discharge. For Event 1, woodland land 462 cover attenuated mean peak discharge by 1.35% (SD = 0.84%) and delayed arrival by 0.3 h 463 (SD = 0.1 h). Grassland land cover reduced mean peak discharge by 3.6% (SD = 1.6%) but 464 had no peak delay (0 h; SD = 0.17 h). Therefore, woodland was more effective at delaying 465 466 peaks, while grassland had a greater effect on peak discharge reduction.

- 467
- Parameter test results for all three NFM interventions (section 3.1.1) revealed sensitivity in
  reducing peak discharge compared to the Model 0 results for Event 1. Woodland planting and
  soil aeration decreased flood peaks by a mean value of 0.731 m<sup>3</sup>/s (1.15%) and 0.883 m<sup>3</sup>/s
  (1.40%) respectively, showing greater sensitivity than hedgerows. However, none of the three
  NFM interventions had a significant impact on flood peak delay.
- 473

# 474 **3.2 Model performance: robustness and accuracy**

To calibrate and validate the model, NSE values and peak fit results were evaluated for the baseline and land cover models (Table 4). While NSE values for some events in the land cover model were lower than in Model 0 with spatially uniform parameters, this did not indicate that land cover reduced model accuracy. Since the study focuses on flood peaks, the accuracy of peak fits was deemed more important. Hydrographs for Event 1 (Figure 4a) showed that the land cover model increased low flows at the start of the event and reduced peak flows, providing a better fit than Model 0.

483	Peak fits were assessed by comparing maximum flood peaks and arrival times to observations
484	for Model 0 and the land cover model, respectively (Table 4). The land cover model had a
485	better fit for peak discharge but not for arrival times. When comparing the total runoff and
486	overland flow time series data in the model results, there was a difference in peak arrival
487	times as illustrated in Figure 4. These biases were from 0 to 2.5 h in several events. It is likely
488	the bias may be due to model characteristics and does not represent an error in the model
489	results. Moreover, it was verified that the difference between the peak arrival times of
490	overland flow time series in the land cover model and observations were negligibly small and
491	did not affect subsequent NFM scenario results (Table 4). This is the reason why only
492	overland flow discharge data were used for the analysis of the NFM scenario results.
493	





496 Figure 4 Hydrographs comparing observations, Model 0, land cover model and NFM

497 scenarios for Event 1 (16-h storm event on 11/12/2015, 45 hr simulation duration without

498 showing the first 5 h of precondition preparation) with inset showing close-up details of flood

499 *peaks; (a) total runoff, (b) overland flow discharge.* 

500

# 501 **3.3 NFM scenarios**

After obtaining all parameter expressions as shown in Table 3 through sensitivity testing, all land cover and NFM scenarios were put into the model and run for all seven storm events. There was variability in the effectiveness of the different NFM interventions for each storm event and the impact of the different events by NFM scenarios. The details of overland flow results for NFM scenarios were compared to overland flow results for the land cover model as detailed below.

508

# 509 3.3.1 Impacts of single intervention scenarios

510 The soil aeration intervention was implemented across the largest proportion of the study catchment (6.40%), which is much higher than the area of the other two interventions: 511 woodland planting (0.61%) and hedgerow planting (1.30%) (Figure 1). The axes of the radar 512 plot (Figure 5) represent the quartiles of peak discharge reduction and area proportion. 513 Notably, Event 4 showed the highest flood peak reduction percentage compared to other 514 events (Figure 5a). The overland flow peak reduction effect of the single intervention 515 scenarios varied. Discharge reduction by soil aeration was largely proportional to the increase 516 in area of implementation. The results showed that the discharge reductions achieved by soil 517 518 aeration interventions were consistently above the 50th quantile for peak reduction in Figure 5b. Woodland planting represented the smallest area yet achieved effective flood peak 519 reduction in all events. This contrasts with the results for hedgerow planting. Although 520 hedgerow planting increased Ks ten-fold and doubled the Kv of baseline values, neither 521 resulted in a significant reduction in flood peaks, with a maximum reduction of only 1.3% 522 across all seven events (Table 5). 523



Figure 5 Radar plot of overland flow peak reductions for single and combined NFM
intervention applications among seven storm events (the axis labels are in percentage and
axes were divided into quartiles); (b) excluded Event 4 from (a).

529 Consistent with the results of the previous parameter sensitivity tests of NFM scenarios for Event 1 (Figure 3), soil aeration was more sensitive in reducing peaks than woodland 530 planting, but both had a greater effect than hedgerow planting (Figure 5). This was reflected 531 in the mean values of overland flood peak reductions calculated for seven events in Table 5. 532 While the standard deviations of peak reductions across the seven events were relatively 533 similar, hedgerows led to slightly higher peak flow variability compared to the other two 534 interventions. The high standard deviation values indicated that the characteristics of the 535 events may have contributed to the increased heterogeneity in the results of NFM scenarios. 536 To better interpret the validity of different NFM scenarios, the results were analysed by 537 grouping according to event characteristics: the shape of hydrographs (single or multi-538 peaked), the wet or dry preconditions of the catchment and their combinations (Figure 6). The 539

- 540 grouped results showed that the same NFM interventions were less effective in reducing
- 541 overland flow peaks in multi-peaked flood events compared to single-peaked events. The
- 542 peak reduction driven by NFM interventions in dry conditions was approximately twice as
- 543 effective as in wet conditions. NFM interventions can therefore have greater effectiveness
- under certain conditions, such as dry antecedence in soil during a single-peaked flood event.



Figure 6 Overland flow peak reduction (%) grouped by characteristics of flood events. (a) &
(c) single intervention, (b) & (d) multiple interventions.

548	Table 5 Percentage reduction in overland flow peak discharge across seven storm events for NFM scenarios, and proportion of catchment area
549	for each scenario (SD = standard deviation, SEM = standard error of mean)

	Overland	Overland flow peak reduction (%)											lay (h)	Area (%)
NFM Scenarios	Event 1 (wet)	Event 2 (dry)	Event 3 (wet)	Event 4 (dry)	Event 5 (dry)	Event 6 (dry)	Event 7 (wet)	Mean	SD	SEM	Mean	SD	SEM	
Soil aeration only	4.43	4.02	5.17	15.00	4.04	3.79	5.82	6.04	3.72	1.41	0.11	0.23	0.09	6.40
Woodland only	1.23	0.66	1.18	11.79	1.33	1.05	2.34	2.80	3.70	1.40	0.11	0.18	0.07	0.61
Hedgerow only	0.31	0.14	0.23	11.62	-0.09	-0.12	0.49	1.80	4.02	1.52	0.00	0.19	0.07	1.30
Hedgerow & Woodland	1.39	1.02	1.75	12.61	0.15	0.55	2.03	2.79	4.06	1.53	0.07	0.11	0.04	1.91
Hedgerow & Soil Aeration	4.75	4.39	5.04	15.87	4.81	4.02	6.20	6.44	3.90	1.47	0.04	0.16	0.06	7.70
Woodland & Soil Aeration	5.90	4.24	5.76	15.83	5.68	5.16	6.62	7.03	3.65	1.38	0.14	0.12	0.05	7.01
Hedgerow & Woodland & Soil Aeration	6.17	4.20	6.21	15.99	6.36	5.04	7.91	7.41	3.66	1.38	0.18	0.17	0.07	8.31
Mean	3.45	2.67	3.62	14.10	3.19	2.78	4.49							4.75
SD	2.24	1.80	2.29	1.86	2.48	2.06	2.61							3.08
SEM	0.85	0.68	0.86	0.70	0.94	0.78	0.99							1.16

# 551 3.3.2 Impacts of combinations of scenarios

Results from the single NFM intervention scenarios were combined into pairs or with all 552 three and tested for seven storm events. The overland flow peak reductions varied among 553 different combinations of NFM interventions (Table 5). Comparing the mean of reductions 554 shows that the flood mitigation effectiveness of the combined intervention scenarios is not 555 simply equivalent to the sum of the effects of single intervention scenarios, where different 556 combinations may have enhanced or reduced effects on the flood mitigation. For example, 557 the mean overland flow reductions for the hedgerow & woodland planting combination and 558 559 for the single woodland planting intervention were almost identical. However, for the hedgerow & soil aeration intervention combined, the mean discharge reduction increased by 560 0.4% compared to the single soil aeration intervention. These increases by combining 561 hedgerow planting with another intervention are all less than the mean discharge reduction of 562 1.8% that can result from the hedgerow planting intervention alone. The standard deviations 563 in Table 5 are all between 3.5 - 4%, which is relatively high compared to the mean indicating 564 that the impact of the interventions vary widely between events. 565

566

Although there is a clear flood mitigation effect in each combination, not all events have a 567 568 stronger peak reduction by combining NFM interventions. Figure 5 shows that the effectiveness of overland flow peak reduction varies among events. For example, the 569 570 hedgerow & woodland combination (red line) achieved an overland flow peak reduction effect comparable to the proportion of area implemented for all events except Event 5. This 571 572 differed from the results for the woodland only scenario, where the inclusion of the hedgerow intervention significantly increased peak mitigation in Event 4 but had the opposite effect in 573 574 Event 5. The soil aeration intervention was the most effective of the single interventions despite being applied to the smallest proportion of the catchment and had positive interaction 575 effects when woodland or hedgerow planting were combined with it among all seven events. 576 The combination of woodland and soil aeration had a more effective overland flood peak 577 reduction effect than the combination of hedgerow and soil aeration in five events. The 578 combination of three interventions resulted in the maximum peak reduction, except in Event 579 580 2 where the effect was slightly lower than that of hedgerow & soil aeration and woodland & soil aeration combinations and in Event 6 where the effect was slightly lower than that of the 581 woodland & soil aeration combination. Thus, even though the addition of woodland planting 582 and soil aeration interventions to the combinations were effective in peak reduction, there 583

were still differences in response to different storm events. Comparing Figure 6 (a) and (b) 584 for event groups, the increase in the median under all four groups shows that multiple 585 interventions enhanced the overland flow peak reduction effect overall. A similar finding is 586 shown in Figure 6 (c) and (d), where multiple interventions were effective at enhancing NFM 587 performance under unfavourable conditions, such as multi-peaked events and wet soil 588 conditions. Thus, the weakening of NFM effectiveness due to multi-peaked flooding and wet 589 590 conditions were less pronounced under multiple interventions compared to a single intervention. 591

592

In scenario tests, overland flow peaks were delayed in arrival by up to 0.5 h. Several 593 scenarios had overland flow peaks that were advanced by one timestep (0.25 h) in some 594 events. Hedgerow & woodland, woodland & soil aeration and all three interventions 595 scenarios resulted in no advance in overland flood peak arrival time among the events. 596 Overall, the mean delay for each scenario across the seven events ranged between 0 and 0.18 597 h. The mean delays for soil aeration and woodland planting were the same but woodland 598 planting had a smaller standard deviation suggesting less individual variation among events. 599 600 The hedgerow planting scenario had no effective peak delay compared to the other scenarios. 601 The mean delays for various combinations of scenarios generally followed the pattern of the overland flow peak reductions: adding soil aeration and woodland planting on any 602 603 interventions increased the delay slightly, while the combination of three interventions resulted in the greatest delay. 604

605

#### 606 4. DISCUSSION

This study used SD-TOPMODEL to investigate the impacts of different land cover types and 607 NFM interventions at an 81.4 km<sup>2</sup> catchment. These effects include impacts on model 608 609 accuracy and performance, flood peak reduction and arrival time delays, and impacts on the interaction of subsurface and overland flows. In general, it was found that SD-TOPMODEL 610 can efficiently and accurately simulate different types of NFM interventions at this catchment 611 scale, validated against multiple storm events, while allowing high resolution (5 m) spatial 612 distribution. The modelling results indicated that multiple interventions were not always the 613 most effective. Event characteristics and antecedent conditions played a significant role in 614 determining the level of flood mitigation. 615

### 617 **4.1 Evaluating model parameters and model performance**

In the parameter calibration and validation of the model, the evaluating criteria are dynamic, 618 including model fit metrics (e.g., NSE values) and flood peak reduction, which are adapted 619 based on the objectives of each calibration scenario. Multi-objective optimization has been 620 justified to improve compromised solutions and enhance the hydrological consistency of 621 parameter settings (Dung et al., 2011; Wöhling et al. 2013; Shafii and Tolson, 2015; Althoff 622 and Rodrigues, 2021). A multi-objective calibration approach is needed for the accuracy of 623 the simulation as indicated by the NSE values and the effectiveness against flood mitigation 624 625 in this study, which depend on the purpose of the test (Efstratiadis and Koutsoyiannis, 2010). For example, during the calibration of parameters for land cover types, results indicated that 626 increasing surface roughness was the most effective factor in reducing peak discharge, but the 627 multipliers with least amount of change within the range were chosen. This is because the 628 best model accuracy (NSE values closer to 1) was achieved when the roughness parameter 629 was increased by the least amount. The land cover model aims to determine the best fit to the 630 actual land cover distribution by comparing to observations, thus, NSE values were more 631 632 critical criteria for the land cover model. On the contrary, in the NFM scenarios, while the NSE values were used as criteria for model stability, the determination of the parameter 633 634 values depended on the best flood peak reduction that could be achieved with similar NSE values. This dual-objective approach aligns with the methods used in recent studies that 635 636 emphasize balancing model stability with functional performance (e.g., flood peak reduction) (Shafii and Tolson, 2015; Althoff and Rodrigues, 2021). Future research could build on this 637 638 strategy by exploring how different catchment characteristics influence the trade-off between stability and effectiveness, potentially leading to the development of adaptive calibration 639 640 frameworks tailored to specific hydrological contexts.

641

# 642 4.2 Impacts of land cover type and NFM interventions on flood peak reduction and timing

Results from this study suggest that NFM interventions achieve maximum effectiveness
during single-peaked storms, particularly under dry antecedence. In previous modelling
studies of NFM on flood peak reduction, it has been shown that the reductions vary between
events, and that this is mainly related to event characteristics (Gao *et al.* 2016; Gao *et al.*2018; Ferreira *et al.* 2020), particularly pre-event soil moisture (Wahren *et al.* 2012). Some
studies also considered seasonal rainfall, which could be adjusted by model parameters
(Gabriels *et al.* 2022). The seven events used in this study occurred in winter months from

October to March and were grouped based on event characteristics and catchment antecedent
conditions allowing them to be compared by their impact on overland flow peak reduction
(Figure 6).

653

Overland flow peak reductions delivered by NFM scenarios were not significantly different 654 for high or low rainfall intensity events when rainfall was concentrated. The most prominent 655 of these was Event 4, which resulted in a greater peak reduction than the other events in each 656 of the single intervention scenarios. This may be because Event 4 was a rapid flow event that 657 658 occurred under dry antecedent conditions and had a much higher flood peak discharge than the other events. Overall, the results suggest that greater overland flow peak reduction occurs 659 under dry antecedence compared to wet antecedence (Figure 6), which aligns with findings 660 from other monitoring and modelling studies on land cover and land use (Bond et al., 2020; 661 Breuer et al., 2009; Wallace and Chappell, 2019). However, the sample size of events in this 662 study is small and further testing for rainfall varieties is needed in the future. 663

664

665 The overland flow peak reduction obtained in this study for the three combined NFM interventions was in the range of 4.20% - 15.99%. We found modelling the effects of upland 666 667 interventions on downstream runoff at a larger catchment scale (81.4 km<sup>2</sup>) did not yield significantly different results to other local scale modelling studies (Bond *et al.*, 2022; 668 669 Hankin et al. 2019; Kingsbury-Smith et al. 2023; Monger et al., 2024). No significant correlation was observed between the peak reduction and rainfall intensity, consistent with 670 671 findings from the data synthesis by Zhu et al. (2024). The study with the closest results to ours also employed SD-TOPMODEL though at a lower resolution (20 m) and concluded that 672 673 upland land management scenarios covering most of the catchment (84 km<sup>2</sup>) resulted in 3.9% 674 to 15% flood reduction with various rainfall intensities (Gao et al. 2017). In contrast, our 675 study used a higher resolution and covered all realistically existing land cover types, which is closer to the reality. 676

677

The closest catchment used for a published NFM simulation to that of the Upper Aire was for Bishopdale in northern England, a study which utilised SD-TOPMODEL and yielded flood peak reductions up to 11% (Kingsbury-Smith *et al.* 2023). Their study concluded that the scenario combining all types of NFM interventions across a large percentage of the 38 km<sup>2</sup> catchment achieved the smallest peak reduction for a 100-year rainfall event and suggested 683 that such an effect may be caused by increased synchronisation between small tributaries. This is contrary to the findings of our study in our larger catchment (81.4 km<sup>2</sup>), which 684 showed that a combination of existing interventions (covering 8.31% of the catchment) 685 produced greater reductions in peak flow than any single intervention, including soil aeration 686 (6.40% of the catchment), the largest and most effective individual intervention. The larger 687 catchment scale of our study offers greater potential for a strategic distribution of multiple 688 NFM interventions leading to desynchronisation of peak flows and achieving solid peak 689 reductions with a lower percentage of intervention areas. However, the differences in NFM 690 691 modelling results between catchments and scenarios highlight the importance of modelling potential responses on a case-by-case basis for each individual catchment (Zhu et al., 2024). 692 This is essential for the effective planning and implementation of a portfolio of catchment 693 based NFM interventions both within the UK and internationally. Additionally, we tested 694 observed rainfall events ranging from 1-year to over 200-year return periods, and calibrated 695 and validated the model using observed data, enhancing the model's credibility and realism. 696 Another study conducted in Swindale in northern England used the Dynamic TOPMODEL 697 combined with a 2D hydraulic model to find flow reductions of 2-6% by NFM (Hankin et 698 699 al. 2019). This suggests that even though the study is for a combination of NFM interventions 700 and excludes differences in catchment size and rainfall intensity, the area and location of the interventions, the characteristic or precondition of the catchment, and the model used, as well 701 702 as its resolution and parameter choices, all have an impact on the simulation results.

703

704 We also found that the flood reduction from different NFM combinations is not directly proportional to the sum of the areas where interventions were implemented. The hydrological 705 706 response of NFM interventions is primarily attributed to the attenuation of the main flood 707 wave traveling through the intervention area (Dixon et al. 2016). When upscaling the study 708 area, the peak flow reduction effects from different sub-catchments did not simply accumulate at the whole catchment scale (Pattison and Lane, 2012). As a result, potential 709 synchronisation between sub-catchments means that the NFM effect on peak discharge does 710 not increase in direct proportion to the area and number of interventions implemented. These 711 712 subtle differences in peak discharge make it challenging to isolate model uncertainty. 713

# 714 **4.3 Implications for SD-TOPMODEL and NFM**

715 There was a bias in the flood peaks in the three outputs of the model (total runoff, subsurface flow, overland flow), which is dictated by model characteristics. A lag time exists between 716 the peak of total runoff and the peak of overland flow (Figure 4). The arithmetic mechanism 717 of the model is to prioritise subsurface runoff and to start generating saturation-excess 718 overland flow when the soil is saturated, and these flows are calculated separately in each 719 grid and each timestep (Gao et al. 2015), which allows the model to fully represent the 720 721 rainfall-runoff process within the topsoil. This procedure allows the model to output the amount of runoff from the subsurface and surface at each timestep and each grid. Thus, the 722 723 model produces total runoff and overland flow rising and falling limbs at different times, and 724 their delay lengths also vary between events (Figure 4).

725

The scenarios in this study significantly reduced total runoff compared to the spatially 726 uniform Model 0, while overland flow decreased, and subsurface flow increased. Similarly, 727 Monger et al. (2024) found that woodland scenarios critically affected the interactions of 728 subsurface and overland flows in SD-TOPMODEL results and reduced total runoff. 729 Increasing the woodland cover improved soil permeability, which increased subsurface flow 730 731 and reduced its conversion to overland flow (Monger et al. 2022a; Monger et al. 2022b; 732 Monger et al. 2024). Grasslands were found to have higher surface roughness and lower soil permeability than woodlands (Bond et al. 2020; Monger et al. 2022b), which may influence 733 734 overland flow more than subsurface flow in SD-TOPMODEL (Bond et al. 2022). When land cover and NFM scenarios were applied, the subsurface flow increased at the beginning of 735 736 storm events, which delayed the time before the onset saturated-excess overland flow. This is because all simulated scenarios increase soil hydraulic conductivity and surface roughness in 737 738 general. Higher infiltration rates and longer infiltration times allow more water to be stored in shallow soils. Moreover, the land cover and NFM scenarios increase the efficiency of the soil 739 740 saturation process while increasing the active area of subsurface runoff in the catchment. Therefore, overland flow peaks are reduced by these scenarios. The simulation results of SD-741 TOPMODEL efficiently and accurately demonstrate the effectiveness of NFM. 742

743

# 744 **5. CONCLUSIONS**

Our study demonstrates a successful application of SD-TOPMODEL in a catchment (81.4

- km<sup>2</sup>) at a 5 m resolution, achieving a strong fit to observed data, with NSE values reaching up
- to 0.93 and minimal peak flow errors. In the modelling of land cover types in the study

748 catchment, woodland was found to be more effective in delaying peaks, whereas grassland was more effective in reducing peak discharge. Among the existing NFM interventions, soil 749 aeration emerged as the most effective individual measure, achieving greater peak discharge 750 reduction results than woodland planting and hedgerow planting. However, the effectiveness 751 752 of NFM interventions was influenced by flood and rainfall characteristics, as well as preevent catchment conditions (wet or dry). Notably, greater flood peak reductions were 753 754 observed during single-peaked events and in dry pre-event conditions. Furthermore, multiple interventions proved more effective and resilient than single interventions in attenuating 755 756 floods at the catchment scale we examined. The results also revealed that the area and number of interventions were not decisive in flood mitigation. This finding presents an 757 opportunity to strategically plan multiple NFM interventions at the catchment scale, enabling 758 a trade-off between intervention area and cost-effectiveness. Therefore, we recommend that 759 high-resolution, spatially distributed modelling of more catchments be undertaken to 760 investigate the impact of catchment characteristics on the effectiveness of NFM. This would 761 support the optimisation of spatial planning and enhance the integration of NFM with other 762 763 flood risk management measures during the planning stage.

764

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768

# 769 DATA AVAILABILITY

- 770 Sensitivity tests and all simulations with land cover and NFM interventions/scenarios for the
- 771 model are available at <u>https://doi.org/10.5518/1628</u>.
- 772

# 773 **REFERENCES**

- Alaoui, A., Rogger, M., Peth, S. and Blöschl, G. (2018). Does soil compaction increase
- floods? A review. *Journal of Hydrology*, 557, 631-642.
- 776
- Althoff, D., & Rodrigues, L. N. (2021). Goodness-of-fit criteria for hydrological models:
- 778 Model calibration and performance assessment. *Journal of Hydrology*, 600, 126674.

780	Archer, N.A.L., Bonell, M., Coles, N., MacDonald, A.M., Auton, C.A. and Stevenson, R.
781	(2013). Soil characteristics and landcover relationships on soil hydraulic conductivity at a
782	hillslope scale: A view towards local flood management. Journal of Hydrology, 497, 208-
783	222.
784	
785	Beven, K.J. and Kirkby, M.J. (1979). A physically based, variable contributing area model of
786	basin hydrology / Un modèle à base physique de zone d'appel variable de l'hydrologie du
787	bassin versant. Hydrological Sciences Bulletin, 24(1), 43-69.
788	
789	Black, A., Peskett, L., MacDonald, A., Young, A., Spray, C., Ball, T., Thomas, H. and
790	Werritty, A. (2021). Natural flood management, lag time and catchment scale: Results from
791	an empirical nested catchment study. Journal of Flood Risk Management, 14(3), 16.
792	
793	Bond, S., Kirkby, M.J. and Holden, J. (2021). Upland grassland management influences
794	organo - mineral soil properties and their hydrological function. Ecohydrology, 14(8).
795	
796	Bond, S., Kirkby, M.J., Johnston, J., Crowle, A. and Holden, J. (2020). Seasonal vegetation
797	and management influence overland flow velocity and roughness in upland grasslands.
798	Hydrological Processes, 34(18), 3777-3791.
799	
800	Bond, S., Willis, T., Johnston, J., Crowle, A., Klaar, M.J., Kirkby, M.J. and Holden, J.
801	(2022). 'The influence of land management and seasonal changes in surface vegetation on
802	flood mitigation in two UK upland catchments. Hydrological Processes, 36(12).
803	
804	Breuer, L., Huisman, J. A., Willems, P., Bormann, H., Bronstert, A., Croke, B. F. W., Frede,
805	H. G., Gräff, T., Hubrechts, L., Jakeman, A. J., Kite, G., Lanini, J., Leavesley, G.,
806	Lettenmaier, D. P., Lindström, G., Seibert J., Sivapalan, M. and Viney, N. R. (2009).
807	Assessing the impact of land use change on hydrology by ensemble modeling (LUCHEM). I:
808	Model intercomparison with current land use. Advances in water resources, 32(2), 129-146.
809	
810	CEH (2012-2020). Monthly Hydrological Summaries [Online]. Available at:
811	https://nrfa.ceh.ac.uk/monthly-hydrological-summary-uk. [Accessed 6 July 2023]
812	

813	CEH (2015, 2020).	UKCEH Land Cover Maps	[Online]. Available at:
-----	-------------------	-----------------------	-------------------------

814 https://www.cen.ac.uk/data/ukcen-land-cover-maps. [Accessed 1 / Ma
--

- 815
- 816 Chandler, K.R., Stevens, C.J., Binley, A. and Keith, A.M. (2018). Influence of tree species
- and forest land use on soil hydraulic conductivity and implications for surface runoff
- 818 generation. *Geoderma*, 310, 120-127.
- 819
- 820 Chehaibi, S., Khelifi, M. and Abrougui, K. (2010). Effects of mechanical aeration on the
- 821 compaction and permeability of a grassy sward. In XVIIth World Congress of the
- 822 International Commission of Agricultural and Biosystems Engineering (CIGR), Québec City,

823 Canada, Canadian Society for Bioengineering (CSBE/SCGAB), June, 13-17.

- 824
- 825 Cooper, M.M.D., Patil, S.D., Nisbet, T.R., Thomas, H., Smith, A.R. and McDonald, M.A.
- 826 (2021). Role of forested land for natural flood management in the UK: A review. *Wiley*
- 827 *Interdisciplinary Reviews-Water*, 8(5), e1541.
- 828
- 829 Cohen, J. (1988). *Statistical Power Analysis for the Behavioral Sciences (2nd ed.)*.
  830 Routledge.
- 831
- 832 COSMOS-UK (2020). Monthly summary of UK soil moisture status Archive of Monthly
- 833 Summaries [Online]. Available at: <u>https://cosmos.ceh.ac.uk/archive-monthly-summaries</u>.
- 834 [Accessed 6 July 2023]
- 835
- Big Dadson, S.J., Hall, J.W., Murgatroyd, A., Acreman, M., Bates, P., Beven, K., Heathwaite, L.,
- 837 Holden, J., Holman, I.P., Lane, S.N., O'Connell, E., Penning-Rowsell, E., Reynard, N., Sear,
- B38 D., Thorne, C. and Wilby, R. (2017). A restatement of the natural science evidence
- 839 concerning catchment-based 'natural' flood management in the UK. Proceedings of the Royal
- 840 Society a-Mathematical Physical and Engineering Sciences, 473(2199), 20160706.

841

- Dixon, S.J., Sear, D.A., Odoni, N.A., Sykes, T. and Lane, S.N. (2016). The effects of river
- 843 restoration on catchment scale flood risk and flood hydrology. *Earth Surface Processes and*

844 *Landforms*, 41(7), 997-1008.

846	Dung, N. V., Merz, B., Bárdossy, A., Thang, T. D., and Apel, H. (2011). Multi-objective
847	automatic calibration of hydrodynamic models utilizing inundation maps and gauge data.
848	Hydrological Earth System Science, 15(4), 1339–1354.
849	
850	Efstratiadis, A., & Koutsoyiannis, D. (2010). One decade of multi-objective calibration
851	approaches in hydrological modelling: a review. Hydrological Sciences Journal-Journal Des
852	Sciences Hydrologiques, 55(1), 58-78.
853	
854	Ellis, N., Anderson, K. and Brazier, R. (2021). Mainstreaming natural flood management: A
855	proposed research framework derived from a critical evaluation of current knowledge.
856	Progress in Physical Geography-Earth and Environment, 45(6), 819-841.
857	
858	Environment Agency (2018). Working with Natural Processes-Evidence Directory. Bristol,
859	England: Environment Agency. Available at: https://www.gov.uk/flood-and-coastal-erosion-
860	risk-management-research-reports/working-with-natural-processes-to-reduce-flood-risk.
861	[Accessed 8 March 2023]
862	
863	Ferguson, C. and Fenner, R. (2020). The impact of Natural Flood Management on the
864	performance of surface drainage systems: A case study in the Calder Valley. Journal of
865	Hydrology, 590, 125354.
866	
867	Ferreira, C.S.S., Mourato, S., Kasanin-Grubin, M., Ferreira, A.J.D., Destouni, G. and
868	Kalantari, Z. (2020). Effectiveness of Nature-Based Solutions in Mitigating Flood Hazard in
869	a Mediterranean Peri-Urban Catchment. Water, 12(10), 2893.
870	
871	Franklin, D. H., Cabrera, M. L., West, L. T., Calvert, V. H., & Rema, J. A. (2007). Aerating
872	grasslands: Effects on runoff and phosphorus losses from applied broiler litter. Journal of
873	environmental quality, 36(1), 208-215.
874	
875	Gabriels, K., Willems, P. and Van Orshoven, J. (2022). An iterative runoff propagation
876	approach to identify priority locations for land cover change minimizing downstream river
877	flood hazard. Landscape and Urban Planning, 218, 104262.
878	

- Gao, J., Holden, J. and Kirkby, M. (2015). A distributed TOPMODEL for modelling impacts
  of land cover change on river flow in upland peatland catchments. *Hydrological Processes*,
  29(13), 2867-2879.
- 882
- Gao, J., Holden, J. and Kirkby, M. (2016). The impact of land-cover change on flood peaks in
  peatland basins. *Water Resources Research*, 52(5), 3477-3492.
- 885
- Gao, J.H., Holden, J. and Kirkby, M. (2017). Modelling impacts of agricultural practice on
- flood peaks in upland catchments: An application of the distributed TOPMODEL.
- 888 *Hydrological Processes*, *31*(23), 4206-4216.
- 889
- Gao, J.H., Kirkby, M. and Holden, J. (2018). The effect of interactions between rainfall
- patterns and land-cover change on flood peaks in upland peatlands. *Journal of Hydrology*,
  567, 546-559.
- 893

Gonzalez - Sosa, E., Braud, I., Dehotin, J., Lassabatère, L., Angulo - Jaramillo, R., Lagouy,
M., Branger, F., Jacqueminet, C., Kermadi, S. and Michel, K. (2010). Impact of land use on
the hydraulic properties of the topsoil in a small French catchment. *Hydrological Processes*,
24(17), 2382-2399.

- 898
- Goudarzi, S., Milledge, D.G., Holden, J., Evans, M.G., Allott, T.E.H., Shuttleworth, E.L.,
  Pilkington, M. and Walker, J. (2021). Blanket Peat Restoration: Numerical Study of the
  Underlying Processes Delivering Natural Flood Management Benefits. *Water Resources Research*, 57(4), 25.
- 903
- Goudarzi, S., Milledge, D., Holden, J., Evans, M., Allott, T., Johnston, A., Shuttleworth, E.,
  Kay, M., Brown, D., Rees, J., Edokpa, D., Spencer, T. (2024). Natural flood management
  through peatland restoration: Catchment-scale modeling of past and future scenarios in
  Glossop. *Water Resources Research*, 60(8), e2024WR037320.
- 908
- 909 Grayson, R., Holden, J. and Rose, R. (2010). Long-term change in storm hydrographs in
- 910 response to peatland vegetation change. *Journal of Hydrology*, *389*(3-4), 336-343.
- 911

- Gunnell, K., Mulligan, M., Francis, R.A. and Hole, D.G. (2019). Evaluating natural
- 913 infrastructure for flood management within the watersheds of selected global cities. *Science*
- 914 *of Total Environment*, 670, 411-424.
- 915
- Hankin, B., Metcalfe, P., Beven, K. and Chappell, N.A. (2019). Integration of hillslope
- 917 hydrology and 2D hydraulic modelling for natural flood management. *Hydrology Research*,
- 918 *50*(6), 1535-1548.
- 919
- 920 Hill, B., Liang, Q., Bosher, L., Chen, H., & Nicholson, A. (2023). A systematic review of
- natural flood management modelling: Approaches, limitations, and potential solutions.
- *Journal of Flood Risk Management, 16*(3), e12899.
- 923
- Holden, J., Grayson, R.P., Berdeni, D., Bird, S., Chapman, P.J., Edmondson, J.L., Firbank,
- 925 L.G., Helgason, T., Hodson, M.E., Hunt, S.F.P., Jones, D.T., Lappage, M.G., Marshall-
- Harries, E., Nelson, M., Prendergast-Miller, M., Shaw, H., Wade, R.N. and Leake, J.R.
- 927 (2019). The role of hedgerows in soil functioning within agricultural landscapes. *Agriculture*,
- 928 Ecosystems & Environment, 273, 1-12.
- 929
- 930 Holden, J., Kirkby, M.J., Lane, S.N., Milledge, D.G., Brookes, C.J., Holden, V. and
- 931 McDonald, A.T. (2008). Overland flow velocity and roughness properties in peatlands. *Water*
- 932 *Resources Research*, 44(6).
- 933
- Holden, J., Shotbolt, L., Bonn, A., Burt, T.P., Chapman, P.J., Dougill, A.J., Fraser, E.D.G.,
- 935 Hubacek, K., Irvine, B., Kirkby, M.J., Reed, M.S., Prell, C., Stagl, S., Stringer, L.C., Turner,
- A. and Worrall, F. (2007). Environmental change in moorland landscapes. *Earth-Science*
- 937 *Reviews*, 82(1-2), 75-100.
- 938
- Iacob, O., Brown, I. and Rowan, J. (2017). Natural flood management, land use and climate
- 940 change trade-offs: the case of Tarland catchment, Scotland. *Hydrological Sciences Journal*-
- 941 *Journal Des Sciences Hydrologiques*, 62(12), 1931-1948.
- 942

- Kay, A.L., Old, G.H., Bell, V.A., Davies, H.N. and Trill, E.J. (2019). An assessment of the
  potential for natural flood management to offset climate change impacts. *Environmental Research Letters*, *14*(4), 044017.
- 946

947 Kingsbury - Smith, L., Willis, T., Smith, M., Boisgontier, H., Turner, D., Hirst, J., Kirkby,

948 M. and Klaar, M. (2023). Evaluating the effectiveness of land use management as a natural

949 flood management intervention in reducing the impact of flooding for an upland catchment.

950 *Hydrological Processes*, *37*(4).

951

952 Kumar, P., Debele, S.E., Sahani, J., Rawat, N., Marti-Cardona, B., Alfieri, S.M., Basu, B.,

- Basu, A.S., Bowyer, P., Charizopoulos, N., Gallotti, G., Jaakko, J., Leo, L.S., Loupis, M.,
- 954 Menenti, M., Mickovski, S.B., Mun, S.J., Gonzalez-Ollauri, A., Pfeiffer, J., Pilla, F., Proll, J.,
- 955 Rutzinger, M., Santo, M.A., Sannigrahi, S., Spyrou, C., Tuomenvirta, H. and Zieher, T.

956 (2021). Nature-based solutions efficiency evaluation against natural hazards: Modelling

- 957 methods, advantages and limitations. *Science of the Total Environment*, 784, 147058.
- 958
- Lane, S.N. (2017). Natural flood management. *Wiley Interdisciplinary Reviews-Water*, 4(3),
  e1211.
- 961
- Lashford, C., Lavers, T., Reaney, S., Charlesworth, S., Burgess-Gamble, L. and Dale, J.
- 963 (2022). Sustainable Catchment-Wide Flood Management: A Review of the Terminology and
- 964 Application of Sustainable Catchment Flood Management Techniques in the UK. *Water*,
  965 14(8), 1204.
- 966
- Leadbetter, M. R. (1991). On a basis for 'Peaks over Threshold' modelling. *Statistics & Probability Letters*, 12(4), 357-362.
- 969
- 970 Leeds City Council (2024). Leeds FAS2 Natural Flood Management [Online]. Available at:
- 971 https://leedscitycouncilfloodresilience.commonplace.is/en-GB/proposals/leeds-fas2-natural-
- 972 <u>flood-management/step1</u>. [Accessed 10 June 2024].
- 973
- 974 Marshall, M.R., Ballard, C.E., Frogbrook, Z.L., Solloway, I., McIntyre, N., Reynolds, B. and
- 975 Wheater, H.S. (2014). The impact of rural land management changes on soil hydraulic

976	properties and runoff processes: results from experimental plots in upland UK. Hydrological
977	Processes, 28(4), 2617-2629.
978	
979	Marshall, M.R., Francis, O.J., Frogbrook, Z.L., Jackson, B.M., McIntyre, N., Reynolds, B.,
980	Solloway, I., Wheater, H.S. and Chell, J. (2009). The impact of upland land management on
981	flooding: results from an improved pasture hillslope. Hydrological Processes, 23(3), 464-
982	475.
983	
984	Mawdsley, T., Chappell, N.A. and Swallow, E. (2017). Hydrological change on Tebay
985	Common following fencing and tree planting: A preliminary dataset. Report in support of the
986	Woodland Trust Upland Planting Research Programme: Lancaster University, Lancaster, UK.
987	
988	Medici, G., West, L.J. and Banwart, S.A. (2019). Groundwater flow velocities in a fractured
989	carbonate aquifer-type: Implications for contaminant transport. Journal of Contaminant
990	Hydrology, 222, 1-16.
991	
992	Metcalfe, P., Beven, K., Hankin, B. and Lamb, R. (2018). A new method, with application,
993	for analysis of the impacts on flood risk of widely distributed enhanced hillslope storage.
994	Hydrology and Earth System Sciences, 22(4), 2589-2605.
995	
996	Monger, F., Bond, S., Spracklen, D.V. and Kirkby, M.J. (2022a). Overland flow velocity and
997	soil properties in established semi-natural woodland and wood pasture in an upland
998	catchment. Hydrological Processes, 36(4), e14567.
999	
1000	Monger, F., Spracklen, D.V., Kirkby, M.J. and Schofield, L. (2022b). The impact of semi-
1001	natural broadleaf woodland and pasture on soil properties and flood discharge. Hydrological
1002	Processes, 36(1), e14453.
1003	
1004	Monger, F., Spracklen, D.V., Kirkby, M.J. and Willis, T. (2024). Investigating the impact of
1005	woodland placement and percentage cover on flood peaks in an upland catchment using
1006	spatially distributed TOPMODEL. Journal of Flood Risk Management, 17(2), e12977.
1007	

1008	Moriasi, D. N., Arnold, J. G., Van Liew, M. W., Bingner, R. L., Harmel, R. D., & Veith, T.
1009	L. (2007). Model evaluation guidelines for systematic quantification of accuracy in watershed
1010	simulations. Transactions of the ASABE, 50(3), 885-900.
1011	
1012	Murphy, T.R., Hanley, M.E., Ellis, J.S. and Lunt, P.H. (2020). Native woodland
1013	establishment improves soil hydrological functioning in UK upland pastoral catchments.
1014	Land Degradation & Development, 32(2), 1034-1045.
1015	
1016	Ordnance Survey (2022). Terrain 5 [ASC geospatial data, Scale 1:10000, Updated: 13
1017	November 2021], Ordnance Survey (GB), EDINA Digimap Ordnance Survey Service
1018	[Online]. Available at: https://digimap.edina.ac.uk. [Accessed 25 March 2022]
1019	
1020	NATMAP (2016). National soil map of England and Wales – NATMAP [Online]. Available:
1021	https://www.landis.org.uk/data/nmvector.cfm. [Accessed 17 May 2023]
1022	
1023	Palmer, R.C. and Smith, R.P. (2013). Soil structural degradation in SW England and its
1024	impact on surface-water runoff generation. Soil Use and Management, 29(4), 567-575.
1025	
1026	Pattison, I., Lane, S. N., Hardy, R. J., & Reaney, S. M. (2014). The role of tributary relative
1027	timing and sequencing in controlling large floods. Water Resources Research, 50(7), 5444-
1028	5458.
1029	
1030	Pattison, I., & Lane, S. N. (2012). The link between land-use management and fluvial flood
1031	risk: a chaotic conception? Progress in Physical Geography, 36(1), 72-92.
1032	
1033	Peskett, L.M., Heal, K.V., MacDonald, A.M., Black, A.R. and McDonnell, J.J. (2023). Land
1034	cover influence on catchment scale subsurface water storage investigated by multiple
1035	methods: Implications for UK Natural Flood Management. Journal of Hydrology: Regional
1036	Studies, 47, 101398.
1037	
1038	Quinn, P., Beven, K., Chevallier, P. and Planchon, O. (2006). The prediction of hillslope flow
1039	paths for distributed hydrological modelling using digital terrain models. Hydrological
1040	<i>Processes</i> , 5(1), 59-79.

1042	Roni, P., Beechie, T., Pess, G., Hanson, K. and Jonsson, B. (2015). Wood placement in river
1043	restoration: fact, fiction, and future direction. Canadian Journal of Fisheries and Aquatic
1044	Sciences, 72(3), 466-478.
1045	
1046	Shafii, M., & Tolson, B. A. (2015). Optimizing hydrological consistency by incorporating
1047	hydrological signatures into model calibration objectives. Water Resources Research, 51(5),
1048	3796-3814.
1049	
1050	Shuttleworth, E.L., Evans, M.G., Pilkington, M., Spencer, T., Walker, J., Milledge, D. and
1051	Allott, T.E.H. (2019). Restoration of blanket peat moorland delays stormflow from hillslopes
1052	and reduces peak discharge. Journal of Hydrology X, 2, 100006.
1053	
1054	Thomas, H., & Nisbet, T. R. (2007). An assessment of the impact of floodplain woodland on
1055	flood flows. Water and Environment Journal, 21(2), 114-126.
1056	
1057	Upper Aire Project (2024). The Upper Aire Project website [Online]. Available at:
1058	https://upperaire.org.uk/. [Accessed 10 June 2024]
1059	
1060	Wahren, A., Schwarzel, K. and Feger, K.H. (2012). Potentials and limitations of natural flood
1061	retention by forested land in headwater catchments: evidence from experimental and model
1062	studies. Journal of Flood Risk Management, 5(4), 321-335.
1063	
1064	Wallace, E.E. and Chappell, A.N. (2019). Blade Aeration Effects on Near - Surface
1065	Permeability and Overland Flow Likelihood on Two Stagnosol Pastures in Cumbria, UK.
1066	Journal of Environmental Quality, 48(6), 1766-1774.
1067	
1068	Wallace, E.E., McShane, G., Tych, W., Kretzschmar, A., McCann, T. and Chappell, N.A.
1069	(2021). The effect of hedgerow wild-margins on topsoil hydraulic properties, and overland-
1070	flow incidence, magnitude and water-quality. Hydrological Processes, 35(3), e14098.
1071	

- Wilkinson, M.E., Addy, S., Quinn, P.F. and Stutter, M. (2019). 'Natural flood management:
  small-scale progress and larger-scale challenges. *Scottish Geographical Journal*, *135*(1-2),
  23-32.
- 1075
- 1076 Wöhling, T., Samaniego, L., & Kumar, R. (2013). Evaluating multiple performance criteria
- 1077 to calibrate the distributed hydrological model of the upper Neckar catchment. *Environmental*
- 1078 *earth sciences, 69, 453-468.*
- 1079
- 1080 Zhu, Q., Klaar, M., Willis, T., Holden, J. (2024). A quantitative review of natural flood
- 1081 management research. *Wiley Interdisciplinary Reviews: Water*, e1765.
- 1082