

1 **Effective integration of drone technology for mapping and managing palm species in the**  
2 **Peruvian Amazon**

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23

24 **Abstract**

25 Remote sensing data could increase the value of tropical forest resources by helping to map  
26 economically important species. However, current tools lack precision over large areas, and  
27 remain inaccessible to stakeholders. Here, we work with the Protected Areas Authority of Peru  
28 to develop and implement precise, landscape-scale, species-level methods to assess the  
29 distribution and abundance of economically important arborescent Amazonian palms using  
30 field data, visible-spectrum drone imagery and deep learning. We compare the costs and time  
31 needed to inventory and develop sustainable fruit harvesting plans in two communities using  
32 traditional plot-based and our drone-based methods. Our approach detects individual palms of  
33 three species, even when densely clustered (average overall score, 74%) with high accuracy  
34 and completeness for *Mauritia flexuosa* (precision; 99% and recall; 81%). Compared to plot-  
35 based methods, our drone-based approach reduces costs per hectare of an inventory of *Mauritia*  
36 *flexuosa* for a management plan by 99% (USD 5 ha<sup>-1</sup> versus USD 411 ha<sup>-1</sup>), and reduces total  
37 operational costs and personnel time to develop a management plan by 23% and 36%,  
38 respectively. These findings demonstrate how tailoring technology to the scale and precision  
39 required for management, and involvement of stakeholders at all stages, can help expand  
40 sustainable management in the tropics.

41 **Keywords:** Palm tree detection, UAV, extensive areas, tropics, Deep learning, CNN, instance  
42 segmentation, management, Peruvian Amazon, crown, Arecaceae, Non Timber Forest  
43 Products, Aguaje, Buriti

44



## 46 1. Introduction

47

48 High-resolution UAV data promises to provide cost-effective solutions to a range of  
49 conservation challenges in the tropics<sup>1</sup>. For example, these platforms have been used to  
50 enable community-led wildlife monitoring in Borneo<sup>2</sup> and delimit priority areas for  
51 conservation and restoration in tropical dry forests in Peru<sup>3</sup>. However, despite their  
52 potential, much of the use of UAVs retains a focus on the technology, rather than leading  
53 to operational conservation success<sup>4</sup>. This failure is an example of the research-  
54 implementation gap<sup>5</sup> which is linked, in broad terms, to insufficient focus on how to link  
55 researchers and stakeholders<sup>6-8</sup>.

56

57 This issue is particularly notable in the use of UAVs to map and monitor tree species  
58 populations in moist forests<sup>9-12</sup>. Sustainable use of forest products derived from tropical  
59 trees is crucial for addressing the interlinked challenges of biodiversity conservation,  
60 supporting livelihoods of local communities and climate change mitigation, and could  
61 greatly benefit from the use of cost-effective means of mapping species populations at the  
62 scale of entire landscapes. However, there are no cases of the operational use of species-  
63 level monitoring by UAVs by stakeholders to support this goal. In contrast, current  
64 approaches with high resolution RS data focus on cases where the phenology or colour of  
65 the species are highly distinctive<sup>13,14</sup> or where the species only occurs at low densities<sup>10</sup>  
66 neither of which are focused on management needs. The challenges are two-fold. First, we  
67 need to overcome the technical challenge of the issues that stakeholders face where they  
68 require these data, and second, we need to ensure that these ‘conservation tools’ are  
69 accessible to, and adopted by, stakeholders<sup>15</sup>.

70

71 We address these twin challenges in the context of sustainable Click or tap here to enter  
72 text.harvesting<sup>16,17</sup> of the fruit of arborescent palms in Amazonia - *Mauritia flexuosa*,  
73 *Oenocarpus bataua*, and *Euterpe precatoria*<sup>18-20</sup> (Fig. 1). These species are vital for  
74 supporting local communities, providing food and habitat for wildlife<sup>18,21-23</sup> and  
75 maintaining key ecosystem services<sup>23-26</sup> including in landscapes with exceptional levels of  
76 carbon storage - *M. flexuosa* dominated palm swamps store 5.4 Gigatonnes of carbon,  
77 mostly belowground as peat<sup>27,28</sup>. These species are well-suited to sustainable management  
78 as they are among the most abundant tree species in Amazonia (so-called “hyperdominant”  
79 species)<sup>29</sup> and have a high economic value<sup>30</sup>: the gross potential value of *M. flexuosa* fruit  
80 harvesting in northern Peru was estimated at USD 41 ± 20.1 million annually<sup>16</sup>, whilst the  
81 global *E. precatoria* market was valued at USD 796.9 million in 2022 and is expected to  
82 grow at an annual growth rate of 11.3% until 2032<sup>31</sup>; the market for oil from *O. bataua*  
83 fruits is also expected to grow by 4% annually until 2031<sup>32,33</sup>. However, these species face  
84 anthropogenic threats that diminish both their abundance and regeneration potential<sup>17</sup>. To  
85 address the increasing demand for these resources, management plans that implement non-  
86 destructive methods of fruit harvesting, such as climbing, must be developed and  
87 implemented<sup>25,34</sup>.

88  
89 Numerous initiatives have been established to promote sustainable fruit harvesting from  
90 economically and ecologically important arborescent palms. However, a key challenge for  
91 developing effective management plans for these resources is accurately mapping their  
92 abundance and distribution. Traditional plot-based fieldwork methods are inefficient,  
93 particularly given the vast extent and often waterlogged conditions of these ecosystems<sup>35-</sup>  
94 <sup>37</sup>. High spatial resolution imagery is an attractive potential solution<sup>38</sup>, yet, previous studies  
95 that mapped tropical peatlands have a spatial resolution of approximately 30 m which

106 provides insufficient detail to measure the abundance of palms accurately<sup>35,36,39,40</sup>.  
107 Commercial satellite imagery with sub-50 cm resolution exists, but it is limited by cost and  
108 cloud cover, similar to the use of crewed airborne imagery<sup>38,41</sup>. In contrast, uncrewed aerial  
109 vehicles (UAVs) provide a cost-effective, safe option for obtaining very high spatial  
110 resolution imagery (approximately 10 cm) at sufficient spatial scale for management  
111 purposes (100-1000 ha)<sup>38</sup>. When combined with deep learning techniques, UAVs allow the  
112 use of automated procedures for individual tree species detection<sup>9,42,43</sup>, as well as palm  
113 species detection and quantification<sup>10,12</sup>. However, an operational method for landscape-  
114 scale mapping and quantifying the abundance of palm species in dense tropical forest,  
115 where the crowns of the same species often overlap, has not yet been implemented. While  
116 such methods hold great potential to expand the use of management plans in these  
117 ecosystems, the challenge extends beyond technological proof-of-concept. For these  
118 ‘conservation tools’ to be effective at landscape scales, they must be robust, cost-effective,  
119 easy to implement and tailored to the needs of user organizations<sup>15</sup>. Stakeholder  
120 involvement is crucial at every stage of development and, the costs - including capital  
expenditure, implementation and training - must be comparable or lower than other  
approaches<sup>15</sup>.

114 Here, we therefore not only aimed to automate the detection and quantification of three  
115 economically important palm tree species - *Mauritia flexuosa*, *Oenocarpus bataua* and  
116 *Euterpe precatoria* - using a combination of field data, red-green-blue (RGB) uncrewed  
117 aerial vehicle (UAV) imagery, and deep convolutional neural networks (CNNs) - but also  
118 to demonstrate how it provides a cost- and time-effective solution for the Peruvian  
119 government’s Protected Areas authority (SERNANP) to manage these forest resources. To  
120 achieve this, we collected RGB UAV images and GPS location points from multiple sites

121 where *M. flexuosa*, *E. precatória*, or *O. bataua* occurred in the region of Loreto in northern  
122 Peru (Supplementary Fig. 1). We developed semantic segmentation maps to classify UAV  
123 mosaics pixels as one of the three palm species or as background, and then trained a model  
124 to partition the semantic segmentation maps into individual palm crowns. We tested the  
125 models using UAV mosaics spanning 70-230 hectares from the Madre de Dios region in  
126 southern Peru (Fig. 2) to assess the distribution and abundance of the palm species.  
127 SERNANP then applied this technology to complete inventories as part of developing two  
128 community-led management plans for sustainable palm fruit harvesting. Finally, we  
129 compared the costs of inventories and developing management plans using traditional plot-  
130 based versus our drone-based approach.

131 Our work is applicable to other tropical regions, as it offers a model trained across a range  
132 of forest conditions for bridging the gap between technological development and practical  
133 conservation. By demonstrating how UAV-based tools can be effectively implemented, we  
134 provide a pathway for supporting forest management and conservation outcomes globally.

135

## 136 **2. Results and discussion**

### 137 a. Landscape-scale palm species mapping

138 Our approach showed a high level accuracy for detecting the crowns of *Mauritia flexuosa*  
139 (accuracy of positive predictions: precision 99%; completeness of positive predictions:  
140 recall 69% and average overall performance: F1 score of 81%) but lower accuracy for  
141 *Euterpe precatória* (89% Precision at 50% recall and F1 of 64%) and *Oenocarpus bataua*  
142 (85% Precision at 52% recall and F1 of 65%) as they were not as abundant in the training  
143 data as *M. flexuosa* (Table 1; Supplementary Table 1). For instance, model 1, which had  
144 only 18 training palms for *E. precatória*, was unable to detect this species. In contrast,

145 model 6, which had 266 training records for this species, achieved a higher level of  
146 accuracy.

147

148 The success of this semantic segmentation method is particularly noteworthy given that  
149 the UAV mosaics used in Madre de Dios were captured by UAV cameras that were not  
150 used for model training, and that the floristic composition varies between regions<sup>40,44</sup>.  
151 This high level of robustness and generalization can be attributed to the use of a diverse  
152 set of training samples and the inclusion of data augmentation techniques. These  
153 techniques, which modify existing training images, introduce variations that simulate  
154 varying flight conditions, such as changes in flying height, illumination, wind presence,  
155 humidity, and different camera settings. By artificially expanding datasets through  
156 image augmentation, the likelihood of encountering similar cases in future data is  
157 increased. Our study therefore supports work showing that combining a diverse dataset  
158 with data augmentation is a highly effective technique for enhancing dataset quality and  
159 improving model performance<sup>45,46</sup>.

160

#### 161 b. Landscape-scale palm quantification

162 Overall, our model accurately quantifies the abundance of arborescent palm species  
163 even amidst densely clustered and large populations of palms (Fig. 3, Supplementary  
164 Fig. 2). The approach works particularly well for *M. flexuosa* (Fig. 3) but could be  
165 improved for *O. bataua* and *E. precatorea* by including more training data, especially  
166 from forest types that were not well represented in this research (e.g. *terra firme* forests).  
167 In general, the performance of the model is highest in areas where more training data  
168 was available and where palm crowns were fully visible: difficulties arise when palms  
169 are stacked on top of each other, which results in some crown centres not being visible,



170 and hence the palm crowns are not split and the number of individuals is underestimated  
171 compared to field data (Fig. 4).

172  
173 The developed method allows us to detect the centre of arborescent palms, delineate  
174 their crowns based on the distance to the centre and the learned shape of the palm, and  
175 count the number of individuals in a given area. Our approach shows a high level of  
176 generalization across lowland Amazonian regions, but it would still be valuable to  
177 evaluate the performance of our model in other Amazonian forests where these  
178 arborescent palm species also occur along with varying tree species composition, such  
179 as in pre-montane forests or other regions of Amazonia.

180  
181 In large UAV mosaics, there are some areas with artifacts that can result in  
182 misclassifications. This is evident in the case of *O. bataua*, where false positives are  
183 prone to occur when certain artifacts resemble the long leaves of this species. Although  
184 the model is able to reduce some misclassifications if the misclassified areas are smaller  
185 than the average size of the palm crowns, this issue may lead to an overestimation of  
186 the number of individuals. Additionally, some palm individuals remain undetected due  
187 to crown shape distortion, which occurs particularly when artifacts appear along the  
188 borders of mosaics or during the blending of large mosaics. This issue can be mitigated  
189 by adhering to best practices during UAV flights particularly avoiding flights during  
190 windy conditions<sup>47</sup> and during pre-processing. Clipping the edge of the mosaics can also  
191 reduce the relief displacement often associated with insufficient overlap between  
192 images<sup>48</sup>. Working with larger UAVs such as Vertical Take Off and Landing UAVs  
193 (VTOLs), could also increase the coverage extent and improve image blending, as they

194 capture more images in a single mission<sup>49</sup>. Their use and associated cost analysis remain  
195 areas for future research.

196 The high-resolution location data provided by the UAV mosaics enables us to visualize  
197 the spatial distribution and ecological associations of the palm species at a fine scale.  
198 These data therefore provide a foundation for exploring processes, such as  
199 environmental filtering, dispersal limitation, gene flux and/or conspecific interactions  
200 that may determine the distributions of tropical tree species<sup>50,51</sup>. For example, *M.*  
201 *flexuosa*, in our study area, tends to form large clusters in waterlogged areas, closer to  
202 water bodies, and *O. bataua* tends to cluster in swampy patches within terra firme  
203 forests<sup>25,29,52,53</sup>, suggesting that environmental filtering may be important for these  
204 species, whereas *E. precatória* shows a scattered distribution and forms smaller  
205 groups<sup>54</sup>, which may reflect an important role for dispersal limitation (Fig. 5).

206  
207 It is important to note that the predictions of our model are solely based on the top  
208 canopy, as the UAV mosaic only captures the upper layer of the forest. Therefore, the  
209 model detects sub-canopy and understorey palms to a much lesser extent. However, in  
210 natural forests, taller individuals of *M. flexuosa* - being in the top canopy and receiving  
211 higher light incidence- bear more and larger fruits suitable for commercialization<sup>55</sup>,  
212 with similar trends for *E. precatória*<sup>56</sup> and *O. bataua*<sup>57</sup>. Hence, fruit production is  
213 concentrated in mature canopy palms, making this underestimation negligible when  
214 using this approach to map this resource to support the development of management  
215 plans.

216

217 c. Bridging the research-implementation gap

218

219 Our approach to bridging the research-implementation gap with mapping the  
220 distribution of these palms, mirrors the framework of Reed et al. (2014)<sup>8</sup> and builds on  
221 insights from the conservation planning<sup>6</sup> and conservation technology literature<sup>15</sup>.

222

223 First, the key stakeholder, the Peruvian Protected Areas Authority (SERNANP), was  
224 involved from the proposal stage (i.e. during project ‘design’<sup>8</sup>) and the research  
225 question that we address - mapping palm species in dense stands - is a key question for  
226 SERNANP (i.e. the research ‘represents’ stakeholder needs<sup>8</sup>). For example, in the  
227 region of Loreto, only 1.29% of harvested *M. flexuosa* fruits come from approved  
228 management plans (Regional Government of Loreto, 2019), highlighting the need for  
229 more effective resource inventory techniques to improve resource management. To  
230 date, SERNANP has granted 28 permissions to harvest *M. flexuosa* in the Loreto region  
231<sup>58</sup> and is in the process of issuing these permissions in the Madre de Dios region<sup>59</sup>, with  
232 our technology being used in two of these initial cases.

233

234 Second, our research has engaged stakeholders over a long period<sup>8</sup> with a strong focus  
235 on capacity building and training<sup>6</sup>. Over the past decade, multiple research projects  
236 have brought SERNANP and our research team together, exploring the distribution and  
237 carbon stores of these palm swamps<sup>27,36</sup>, the economic potential of palm fruit harvesting  
238<sup>16</sup> and the potential to identify crowns of different palm species<sup>60</sup>. Our current  
239 collaboration has involved significant engagement activities through in-person and  
240 online workshops, as well as *ad hoc* meetings. We began with an initial session to  
241 harmonize ideas and identify stakeholder needs (online, April 10, 2019, with 18  
242 participants). This was followed by drone flight training (May 24, 2019, with 7

243 participants), training on image preprocessing, including mosaicking (January 28, 2020,  
244 with 36 participants), and a session for using the model and providing feedback on its  
245 performance, primarily through visual assessments (August 2, 2020, with 4  
246 participants). Third, the technology we developed is designed to be user-friendly<sup>15</sup> based  
247 on open-source software (Palacios, Tagle et al. in prep), making it accessible and easy  
248 for stakeholders to use.

249

250 Finally, our approach is cost effective compared to existing methods for resource  
251 inventory. Traditionally, SERNANP has used plot-based methods for the resource  
252 assessments<sup>58</sup>. To compare the costs of the plot- and drone-based approaches, we used  
253 data from SERNANP to analyse the expenses associated with implementing traditional  
254 plot-based (over 10 ha) and UAV-based methods (over 200 ha) for generating these  
255 inventory data. Our UAV approach is significantly more cost-effective for mapping and  
256 quantifying the abundance of *M. flexuosa* stems, and for producing the information  
257 needed to develop management plans for this resource. Our UAV-based method reduces  
258 the costs per hectare of a resource inventory of *M. flexuosa* by 99% compared to plot-  
259 based methods (USD 5 ha<sup>-1</sup> versus USD 411 ha<sup>-1</sup>) and reduced the total operational costs  
260 of developing a management plan by 23% (Table 2). This reduction in operational costs  
261 is linked to reduced reliance on external services (Supplementary Table 2) arising from  
262 investment in capacity building. Park rangers now handle tasks that were previously the  
263 responsibility of external consultants, such as drone field surveys, data processing, and  
264 writing the resource inventory report. Consultants now focus on writing the  
265 management plan.

266

267 Plot-based methods have much lower capital costs (Table 2), but even when considering the  
268 higher initial capital costs associated with the UAV use, such as acquiring a robust workstation,  
269 the UAV itself, software licenses, and team training for drone operation and image processing,  
270 the UAV approach demonstrates a cost advantage once the number of management plans  
271 surpasses four (Fig. 6). This cost advantage arises due to its substantially lower marginal costs  
272 per additional plan (Fig. 6) and is likely to be achieved as the equipment typically lasts 3-5  
273 years and the trained personnel are often permanent staff who remain long-term.

274

275 The UAV approach also offers more than an order of magnitude more spatial coverage and this  
276 greater area not only amplifies the economic benefits of employing drones but also enables  
277 cost-effective surveying of locations that would otherwise be excluded. This advantage  
278 empowers local communities to expand their harvesting areas without requiring extensive  
279 search efforts. Additionally, it reduces the time of personnel involved in these tasks by one-  
280 third (Supplementary Table 2).

281

282 Our method therefore provides a practical, cost-and-time-effective and reliable technique for  
283 generating essential information such as the location of palm crowns and their areas across  
284 landscapes of 100-250 hectares. This method can support the effective development of  
285 management plans and has the potential to improve the spatial detail and timeliness of forest  
286 monitoring, benefiting stakeholders involved in the sustainable management of palm resources.  
287 Local communities can use it to locate their resources more efficiently, while NGOs and private  
288 companies can use it to validate the responsible use of resources. Governmental oversight  
289 agencies, such as SERNANP, can use it to estimate the amount of fruit harvested from a given  
290 protected area and investigate cases of unsustainable use. By enabling better informed decision-  
291 making and management practices, our method has the potential to contribute significantly to

292 the sustainable management of palm resources and to the protection of the intact forest  
293 landscapes where they occur.

294

295 Regarding operational matters, our method saves significant time and effort compared to the  
296 time-consuming, labour-intensive, and subjective task of visually interpreting UAV mosaics,  
297 especially when the identification of these species requires specialized training <sup>12</sup>. It can also  
298 reduce the time for labelling training data by using semi-automatic crown delineations, in  
299 contrast to the manual delineation typically used for this type of work.

300

301 Additionally, as the model has been trained to identify palms under various lighting conditions,  
302 no image editing for lighting conditions are required for the UAV mosaic. As a result,  
303 SERNANP tested our method presented here and ultimately quantified palm abundance in two  
304 communities within the Tambopata National Reserve. These inventories were then used to  
305 support the first management plans for palm fruit harvesting in this National Reserve<sup>59</sup>. Due to  
306 cost efficiency, there is potential to adopt this method for larger conservation effort in Peru.  
307 Currently, SERNANP is in the process to integrate our methodology as a standardized national  
308 protocol. To facilitate this upscaling, project pilots will be conducted in all Peruvian regions  
309 where palms are present. These pilots will gather feedback from various protected areas to  
310 ensure the methodology's effectiveness in diverse landscapes.

311

312 Our methodology can also be applied to other regions and species with distinctive crowns given  
313 its robustness, which comes from extensive data collection across a range of forest landscape  
314 and imaging conditions. We also use image augmentation techniques to increase data variability  
315 and robustness of the model, and the fact that our model and code is openly shared. For example,  
316 our approach should be explored for mapping the distribution of *Euterpe* in the dense stands on

317 the floodplains of eastern Amazonia, or for species that occur at high densities in other tropical  
318 peatlands, such as *Pandanus* spp. in Asia/Oceania or *Raphia* spp. in the Congo basin. More  
319 broadly, our approach demonstrates how the gap between research and implementation can be  
320 bridged, and these principles are applicable wherever technology is being designed to address  
321 conservation challenges.

322

### 323 3. Methods

#### 324 a. Study area

325 We developed our models based on UAV and ground reference data from 55 sites across the  
326 region of Loreto in northern Peruvian Amazonia. There are a wide variety of forest types in this  
327 region including upland forest with clay-rich and white sand soils, seasonally flooded forests  
328 and extensive palm swamps<sup>61</sup>. Surveys were carried out in collaboration with local communities  
329 and the National Service of Protected Areas – SERNANP, in areas that our partners indicated  
330 had the presence of either *Mauritia flexuosa*, *Euterpe precatoria* or *Oenocarpus bataua*. The  
331 sites focussed on seasonally flooded forests and palm swamps but also included some sites that  
332 covered planted palms in local communities, which were incorporated to enhance the  
333 generalisation of the model. Our overall approach aimed to encompass areas varying in palm  
334 density and floristic composition. Some of the sites are within protected natural areas; other  
335 sites are forests managed by local communities (Supplementary Fig. 1). Twenty sites include  
336 plots from the Amazon Forest Inventory Network (RAINFOR) which we used to supply part  
337 of the palm GPS location data; these plot data are managed using the ForestPlots.net online  
338 database<sup>62,63</sup>.

339

340 To test the models, we used four UAV mosaics from the region of Madre de Dios in southern  
341 Peruvian Amazonia. Here, the UAV flights were carried out over palm swamps in the

342 Tambopata National Reserve, which is situated in the Tambopata River basin near Puerto  
343 Maldonado. The UAV mosaics can be accessed at [https://doi.org/10.4121/70a8cec0-dfa7-4963-](https://doi.org/10.4121/70a8cec0-dfa7-4963-ba8a-612e738ec0cb.v1)  
344 [ba8a-612e738ec0cb.v1](https://doi.org/10.4121/70a8cec0-dfa7-4963-ba8a-612e738ec0cb.v1) SERNANP works closely together with local communities in this  
345 region to develop sustainable commercial activities, such as Brazil nut harvesting<sup>44</sup>. More  
346 recently, in response to the growing demand for palm fruits, there has been an increased focus  
347 on harvesting the fruits of *M. flexuosa* in this reserve<sup>44</sup>.

348

349 b. Ground reference data collection

350 For the training and validation data, 5,089 individuals of *M. flexuosa* (4497), *E. precatória*  
351 (282) and *O. bataua* (310) palms were identified and georeferenced using a handheld  
352 Trimble Geo7X GPS-receiver and the dual-frequency GNSS Trimble Tornado antenna,  
353 with an average error of approximately 5 m from 2017 to 2019 across all 55 sites.

354

355 c. UAV missions

356 For the training and validation data, UAV data were collected concurrently with ground  
357 data collection using small commercial multi-rotors (DJI Phantom 4 Pro and DJI Phantom  
358 4 RTK)<sup>64</sup> over 55 sites from 2017 to 2019. Some sites were surveyed every year and others  
359 only once during this period.

360 For the testing data, SERNANP conducted missions using commercial small multi-rotors  
361 (DJI Phantom 4 and DJI Mavic 2, the latter possessing slightly different camera  
362 characteristics) across three sites from 2019 to 2022, flying over the Sandoval lake twice  
363 — in 2019 and in 2021.

364

365 To ensure the generalisability of the model against variations in the spatial resolution of the  
366 UAV mosaics, the missions were conducted at various flying heights, ranging from 60 to



367 150 meters above ground level (AGL). It is important to note that the maximum flying  
368 height permitted by national legislation is 150 m AGL, which precluded capturing images  
369 from higher altitudes (up to 500 m AGL) which could otherwise have been useful<sup>60</sup>. The  
370 forward and side overlap ranged from 80 to 90% and the camera angle was mostly at the  
371 nadir position (90°)<sup>64</sup>.

372

#### 373 d. Data processing

374 The data processing involved five stages: pre-processing, training an image semantic  
375 segmentation model, training an instance segmentation model (Supplementary Fig. 3),  
376 accuracy assessment and model testing, and cost analysis. Pre-processing was conducted  
377 using various software platforms, detailed in the following subsection. The remaining stages  
378 were conducted entirely in the Python programming language<sup>65,66</sup>, with specific packages  
379 referenced as needed.

380

#### 381 **Pre-processing: Training and Validation data preparation**

382 The pre-processing consisted of 4 steps: (1) mosaicking, (2) multiresolution superpixel  
383 partitioning and labelling, (3) tiling and (4) image augmentation. The UAV images collected  
384 on the missions were mosaicked using the software Pix4DMapper. Due to the intricate structure  
385 of vegetation, different parameters were tested to obtain mosaics with as few artifacts as  
386 possible<sup>60</sup>. In some cases, the mosaics were generated from a single mission, while in other  
387 cases, images from different flights over the same site were combined. Eighty-nine UAV  
388 mosaics were obtained in total<sup>67</sup>.

389 To reduce the time spent on data labelling, mosaics were then used as input for a multiresolution  
390 superpixel partitioning that delineate the crowns<sup>60</sup>. Each crown was assigned the species label  
391 corresponding to the ground reference data (the palm tree locations recorded with the GPS). To

392 ensure the accuracy of the data, shapefile layers containing the ground reference points were  
393 overlaid on the RGB mosaic using open-source software Quantum GIS (QGIS). This process  
394 was conducted to verify whether the location points aligned with the palm tree crowns in the  
395 mosaic. In cases where reference palm trees were misaligned, they were either manually  
396 adjusted or excluded from the classification if the corresponding palm tree was not clearly  
397 identifiable in the mosaic. Subsequently, the shapefiles containing the delineated crowns with  
398 their assigned species labels were rasterized to match the same extent as the UAV mosaics using  
399 the Python programming language<sup>65</sup>. This approach saves time in training data preparation, as  
400 the conventional practice involves manual delineation, which is time-consuming and costly,  
401 especially when verifying a large number of tree crowns requires cross-checking by an  
402 experienced visual interpreter<sup>68</sup>.

403

404 The third and fourth steps were also conducted in Python. In the third step, mosaics with the  
405 labelled data were sliced into tiles of 512 x 512 pixels, ranging from 4x4m to 30x30m on the  
406 ground, depending on the mosaic's spatial resolution. This size is sufficient to capture at least  
407 two palms per tile, as shown in our crown measurements (Supplementary Table 3), following  
408 the approach for sample selection by Brodrick et al. (2019)<sup>69</sup>. The tiles were split into a training  
409 set (80%) and a validation set (20%). To test the accuracies and generalization of the model,  
410 seven combinations of tiles were used to ensure that the trained model could effectively handle  
411 diverse characteristics associated with UAV data collection or geographical locations. These  
412 characteristics encompassed factors such as illumination conditions, mosaic spatial resolution,  
413 and floristic composition (Table 1). To increase the ability of the model to generalize, some of  
414 the tiles were augmented using up to two different augmenters per batch, applied randomly<sup>70</sup>.  
415 Image augmentation artificially expands datasets, increasing the likelihood of encountering  
416 similar cases in future data, which improves dataset variability and model performance<sup>45,46</sup>. The

417 augmenters used were affine image transformations such as flipping (50%), rotating ( $\pm 20^\circ$ )  
418 and zooming in and out (0.8 to 1.2) to simulate different flying paths and altitudes and color  
419 modifications as the change of brightness ( $\pm 20\%$ ) and saturation (-20% to +10%) to resemble  
420 different illumination conditions adding blur to resemble the presence of humidity/light fog or  
421 water droplets, motion blur to simulate different wind conditions, elastic transformations to  
422 resemble artifacts in the mosaics, and JPEG compression to simulate different camera sensors.

423

### 424 **Species mapping: Image Semantic Segmentation model**

425 We used a deep convolutional neural network (CNN), selecting a semantic segmentation  
426 architecture and task formulation, rather than object detection. This decision was based on  
427 the feedback from our main stakeholders, who indicated that having delineated crowns was  
428 an important asset for them and it has been shown that having crown area information (i.e.  
429 dominance) is more effective for forest management<sup>68</sup>. We selected the DeepLab v3+  
430 architecture, which has as its backbone MobileNet-v2 and atrous spatial pyramid pooling  
431 (ASPP), allowing enlarging the field of view of filters to incorporate multiple scales context  
432 but maintaining localization accuracy<sup>71</sup>. We did not perform instance segmentation  
433 simultaneously at this point due to the high computational costs and complexity<sup>9,68</sup>.

434

### 435 **Palm quantification: Instance segmentation model**

436 In tropical forests, the clustering of individual species at close proximity presents a  
437 challenge for quantifying species' abundances. Directly quantifying individuals from  
438 semantic segmentation maps is inaccurate due to masks potentially encompassing multiple  
439 crowns. Hence, a method is required to split these multi-crown segments without high  
440 computational costs or complexity. We used a simple yet powerful convolutional neural

441 network-based method for instance segmentation based on semantic segmentation masks  
442 called Deep Watershed Transform<sup>72</sup>, which learns how to identify the centre of the palm  
443 trees. This method is inspired by the classical watershed transform algorithm, where the  
444 distance to the boundary helps to discriminate crowns<sup>72</sup>. The model uses the segmented  
445 image plus the UAV mosaic as input to detect the instances and delineate “basins”, where  
446 each basin corresponds to an individual palm crown.

447

448 To ensure accuracy, the removal of small predicted pieces of crowns is conducted by a post-  
449 process that first fills small holes to keep the integrity of an instance, filling a maximum of  
450 1000 pixels per instance. Then, the instances are eroded to make the spacing between  
451 crowns clearer, using the Scikit morphology binary erosion (enlarging darker regions, thus  
452 the spaces between crowns). Next, structural erosion from SciPy is applied to maintain the  
453 crown shape, where the kernel sizes depend on the species. Later, the instances smaller than  
454 the average UAV measurements of crown size (Supplementary Table 3) are removed and  
455 holes are filled after erosion using the Scikit morphology module.

456

### 457 **Accuracy assessment and model testing**

458 To evaluate the model's transferability, we utilized full UAV orthomosaics to assess the  
459 accuracy and robustness of the seven models trained on different data subsets as described  
460 on Table 1. Among these models, six were trained using different arrangements of training  
461 and test data, while the "Final" model incorporated all the training data from the region of  
462 Loreto and was tested using the data from the region of Madre de Dios region (Fig. 2).

463

464 The training and test data arrangements were designed to cover data scenarios of increasing  
465 complexity. Model 1 involved a dataset from a single location with a similar floristic  
466 composition, images captured on similar dates within the same year, and the utilization of  
467 the same UAV. The objective was to establish a baseline for the model's performance. For  
468 training and testing this model, five UAV mosaics from the Veinte de Enero community in  
469 Loreto from October 2017 were used.

470

471 Models 2 to 4 aimed to test the model's robustness over time and involved three different  
472 combinations of datasets grouped by the year of data collection. These combinations  
473 utilized two years of data for training and one year for testing. This approach allowed us to  
474 account for variations in illumination conditions, habitat diversity, and spatial resolutions  
475 resulting from different flying heights.

476

477 Model 5 assessed the model's robustness across different locations. In this case, the model  
478 was trained on data from one hydrological basin around the Allpahuayo-Mishana National  
479 Reserve and then tested on data from another basin in the Pacaya Samiria National Reserve.

480

481 Model 6 explored the model's performance when the floristic composition of the forests  
482 differed slightly. For this scenario, we used most of the available training data from various  
483 areas around the Pastaza- Marañon (PM) Foreland Basin, encompassing different habitat  
484 types, illumination conditions, and spatial resolutions. The model was then tested with  
485 mosaics from the Nueva Jerusalen site, in the north of the Loreto Region, close to the border  
486 with Colombia<sup>73</sup>.

487

488 The general performance of the different models was evaluated with the Precision (user's  
489 accuracy), Recall (producer's accuracy) and F1 score from the Scikit-learn Package<sup>74</sup>. For  
490 the species mapping assessment, a Point-in-polygon method was used<sup>75</sup>, comparing the  
491 ground data polygon with the predicted points. Given that the prediction points are  
492 exclusively generated for the target species, in order to evaluate whether the model is  
493 predicting non-palm trees as palm trees (commission error), we manually designated other  
494 objects, not belonging to the target species, as points in the background class. This was done  
495 in areas where the presence of the three palm species was not visually identified, such as  
496 the crowns of other trees. The selection of these points mirrored a similar number to those  
497 allocated for the target species within each plot. In addition, the overall accuracy and the  
498 confusion matrices were also calculated.

499

500 The Precision— user's accuracy ( $UA$ ) —is the number of correctly classified objects  
501 (true positives,  $tp$ ) in a class divided by the total number of points that were predicted  
502 by the model:

$$UA_{class} = \frac{tp_{class}}{N_{classified}}, \quad (1)$$

503

504 The Recall— producer's accuracy ( $PA$ ) —is derived by dividing the number of  
505 correctly classified objects per class ( $tp$ ) by the total number of polygons according to  
506 the ground reference:

$$PA_{class} = \frac{tp_{class}}{N_{ground\ reference\ class}}, \quad (2)$$

507

508 The F1 score is the harmonic mean of recall and precision to provide a comprehensive  
509 assessment of a model's performance and thus expresses the balance between recall  
510 and precision:

$$F1\ score = 2 \times \frac{precision \times recall}{precision + recall}, \quad (3)$$

511 The F1 score was used to assess overall performance, instead of the overall accuracy,  
512 because *M. flexuosa* was more abundant in most plots compared to other palm species.

513

514 The overall accuracy (*OA*) is the total number of correctly classified pixels (*tp*),  
515 divided by the total number of samples ( $N_c$ ):

$$OA = \frac{\sum_{class}^i tp}{N_c}, \quad (4)$$

516 where *i* is the number of classes.

517

518 To evaluate our approach to counting individual trees, the predicted number of  
519 individual arborescent palm trees was compared to the visible number of palm trees per  
520 UAV mosaic across fifty-five sites (Fig. 3). Subsequently, we calculated the  $R^2$ , RMSE,  
521 and the Normalized RMSE to assess the relationship between the predicted and visible  
522 counts. For sites with fewer than a thousand individuals, the visible palm values in the  
523 UAV mosaics relied on the count of GPS locations; for sites with a higher number of  
524 individuals, the reference values were based on manually located crowns. This approach  
525 enabled us to assess the scalability of our method across both smaller, uniform areas  
526 and larger, more variable regions.

527

528 **Cost analysis**

529 We sourced the costs associated with developing management plans to support  
530 sustainable palm fruit harvesting of *M. flexuosa*, based on both plot- and drone-based  
531 inventories, from SERNANP. We grouped the costs as external services provided by  
532 consultants (staff for field survey, data processing and reporting), capital costs including  
533 capital equipment (small boat, drone, appropriate computer with licenced software) and  
534 capacity building (drone pilot licences for the park rangers, training on image  
535 processing), recurring costs (field consumables not provided by the consultant team),  
536 and SERNANP permanent staff costs (Supplementary Table 2).

537 Costs were quantified in Peruvian Nuevos Soles and converted to USD using the  
538 exchange rate for 2021<sup>76</sup>. The costs were based on actual expenditure corresponding to  
539 the Master Plan of the Tambopata National Reserve 2019-2023<sup>44</sup>.

540

541

542

543



544 **Data availability**

545 Source data are provided with this paper. The UAV mosaics and their details can be found at  
546 4TU ResearchData [<https://doi.org/10.4121/70a8cec0-dfa7-4963-ba8a-612e738ec0cb.v1>] <sup>67</sup>.

547

548 **Code availability**

549 The code for training the model and making predictions is available at Code Ocean

550 [<https://doi.org/10.24433/CO.0764353.v1>] <sup>66</sup>

551 For the model training workflow, open the Jupyter notebook "1.PalmsCNN\_Tutorial." To

552 work only with the predictions, use "2.PalmsCNN\_Tutorial\_Prediction."

553

554 The Google Collab notebook and single python scripts can be found at GitHub:

555 <https://github.com/iiap-gob-pe/PalmsCNN/tree/main>

556

557

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#### 782 **Author contributions**

783 X.T.C., F.C.D., D.D.C.T. and T.R.B. conceived the study. X.T.C., D.M., H.B. and M.H.  
784 designed the methodology. X.T.C., A.D.D., L.F., S.D.L.P., E.F.G., P.P.P., F.C.D and G.F.L.  
785 collected and processed the ground data; X.T.C., A.D.D., E.F.G, R.C.V., S.P., S.D.L.P. and  
786 L.F. processed the UAV data; X.T.C., D.M., H.B., N.E.T., G.M., E.F.G., T.R.B., and M.H.  
787 conducted the formal analysis. X.T.C., M.H., E.N.H.C., T.R.B, R.C.V., L.F., S.P. and  
788 S.D.L.P. designed the graphics, X.T.C. wrote the manuscript with critical inputs from  
789 E.N.H.C., E.F.G., G.C., H.B., D.M., F.C.D., N.E.T., G.F.L., G.M., L.F. and T.R.B. All  
790 authors have read and agreed to the published version of the manuscript.

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792 **Competing Interests**

793 The authors declare no competing interests.

794 **Supplementary Information**

795 Contains Resumen en Español, Supplementary Tables 1-3, and Supplementary Figures. 1-3.

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Table 1. Average classification accuracies across three species of arborescent

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palms for assessing the robustness of the seven different approaches for model

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training and testing. Source data are provided as a Source Data file.

Model	Training				Testing			Differences over*	Precision	Recall	F1-score
	Images from	Year	No. of UAV mosaics	No. of tiles used	Mosaic from	Year	No. of UAV mosaics				
1	Veinte de Enero site	2017	4	764	Veinte de Enero site	2017	1	ic	0,63	0,60	0,61
2	All sites	2017 + 2018	26	10505	Nueva York & 2 de Mayo de Muyuy	2019	2	ic, fc, sr	0,59	0,60	0,59
3	All sites	2018 + 2019	33	10420	Veinte de Enero & Parinari	2017	2	ic, fc, sr	0,86	0,46	0,45
4	All sites	2017 + 2019	34	11651	Jenaro Herrera & Iquitos	2018	2	ic, fc, sr	0,77	0,54	0,56
5	Around the National Reserve Allpahuayo Mishana	All years	13	5157	Around the National Reserve Pacaya - Samiria	2019	2	ic, sp, fl, gl	0,86	0,59	0,67
6	Within the Pastaza Maraño n (PM) Foreland Basin	All years	79	27992	Nueva Jerusalen site	2019	2	ic, sp, fl, gl, td	0,72	0,77	0,72
Final	All the training sites from the Loreto region.	All years	81	29902	All the sites in Madre de Dios region	2019-2022	4	ic, sp, fl, gl, td	0,88	0,67	0,74

\* ic: illumination conditions, fc: floristic composition, sr: spatial resolution, gl: geographical location, td:









amount of training data

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Table 2. Cost comparison between of expenditure by SERNANP on traditional plot-based fieldwork and drone (UAV) surveys for developing management plans for sustainable management of *M. flexuosa*. Total costs are the sum of operational expenditure and capital costs; resource inventory costs are one component of the operational expenditure. Source data are provided as a Source Data file.

	<b>Area Covered (ha)</b>	<b>Total costs (USD)</b>	<b>Total time (person-hrs)</b>	<b>Operational expenditures (USD)</b>	<b>Capital costs (USD)</b>	<b>Resource inventory costs (USD)</b>	<b>Inventory costs per area (USD/ha)</b>
<b>Plot-based method</b>	10 ha	\$29,863	1,136 h	\$14,521	\$15,342	\$4,110	\$411/ha
<b>UAV survey</b>	230 ha	\$39,890	724 h	\$11,205	\$28,685	\$1,0956	\$5/ha
Change in costs based on UAV utilization (%)		34%	-36%	-23%	87%	-73%	-99%

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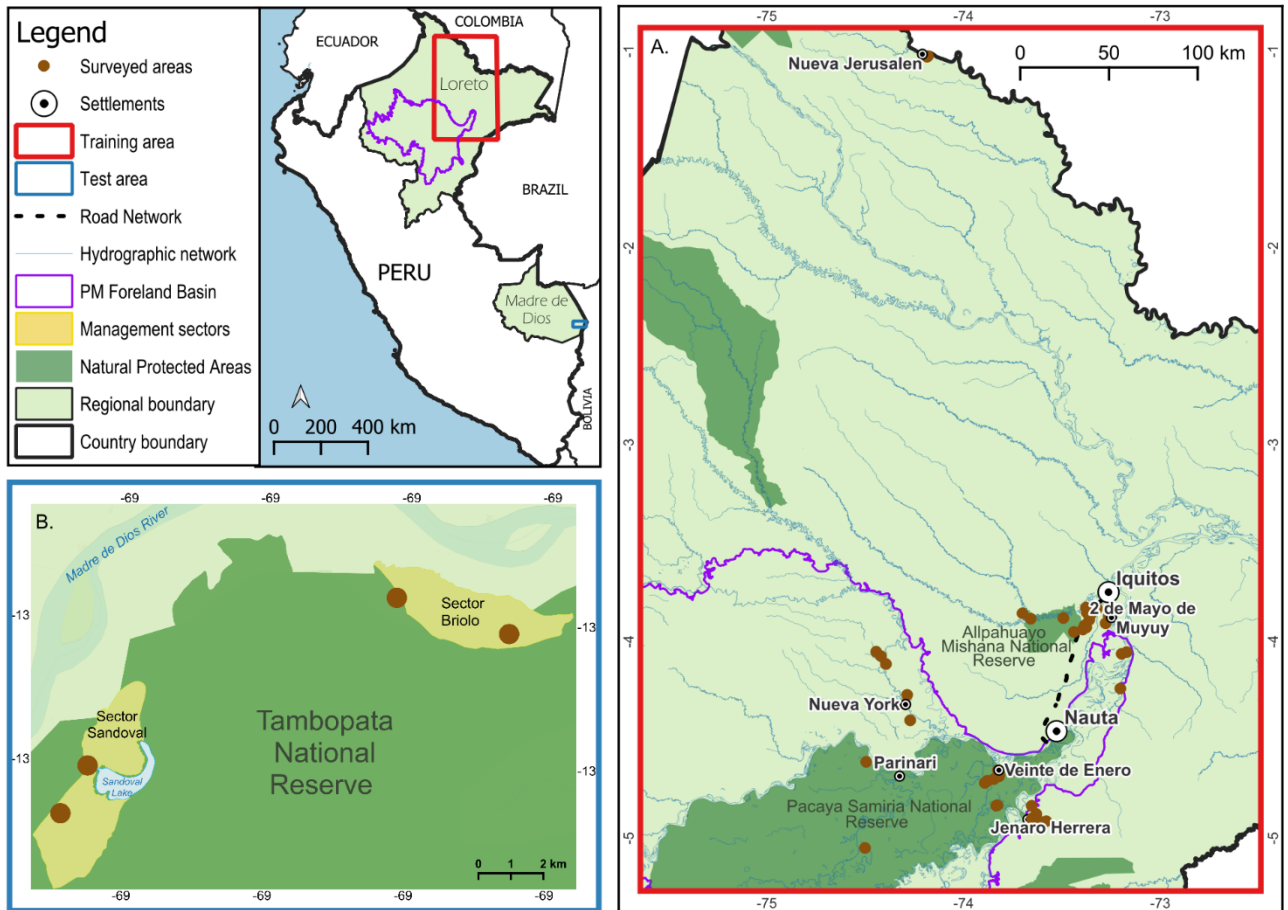
Species local names	Fruit	Ground view	Drone view
<i>Mauritia flexuosa</i> aguaje/buriti			
<i>Oenocarpus bataua</i> ungurahui/bataua			
<i>Euterpe precatoria</i> huasai/acai berry			

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811 Figure 1. Images of the three most ecologically and economically important arborescent

812 palm species in the Peruvian Amazon

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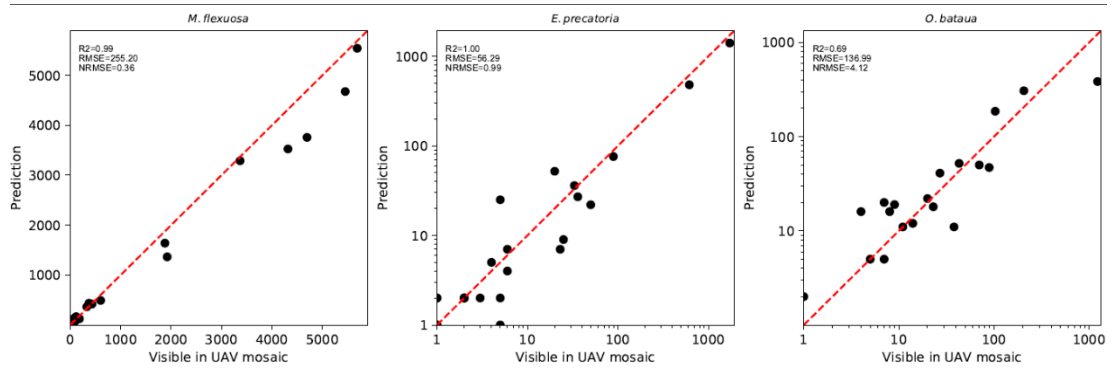
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Figure 2. Distribution of the locations surveyed for (A) training and (B) testing a convolutional neural network (CNN) model for detecting three species of arborescent palms using large-scale UAV mosaics. The brown dots correspond to the sites where the UAV surveys were conducted. The purple line corresponds to the Pastaza-Maranon (PM) Foreland Basin. Sources: Cartographic base layers belong to the National Geographic Institute of Peru – IGN (2017) and to the Ministry of Environment of Peru MINAM (2019). Source data are provided as a Source Data file.



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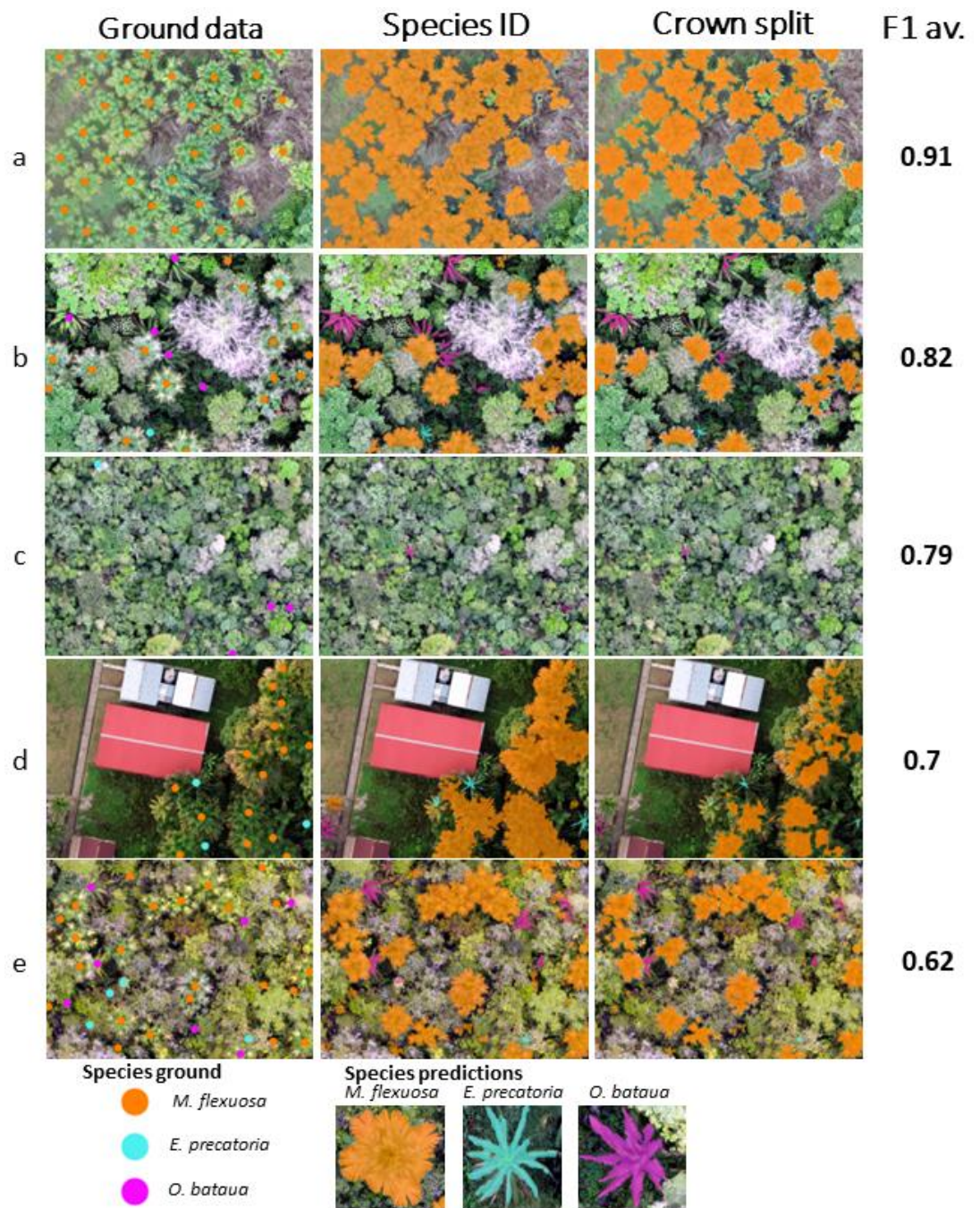
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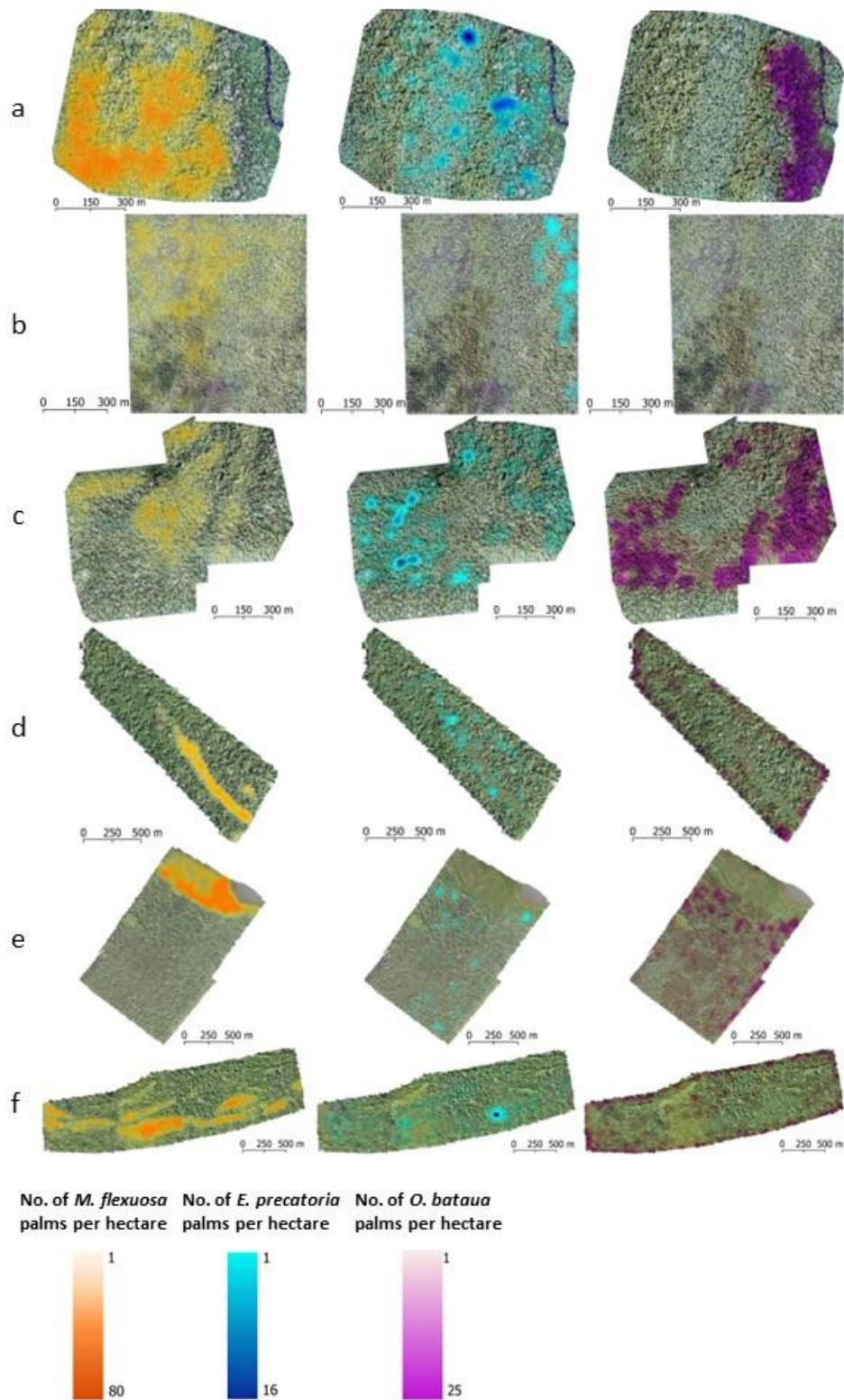
Fig. 3. Comparison of the number of three species of arborescent palm (*Mauritia flexuosa*, *Euterpe precatorea* and *Oenocarpus bataua*) visible in the UAV mosaics with model-predicted results across fifty-five sites. For sites with less than a thousand individuals, the number of palms in the UAV mosaics was based on the count of GPS locations of palms with visible crowns in the canopy at each site. For the sites with a greater number of individuals, the reference values were the total count of manually located crowns in the UAV mosaics. The red lines show the 1:1 relationship in each case. Source data are provided as a Source Data file.





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835 Figure 4. Examples of the final model predictions for the location and crown delineation of  
 836 three species of palm tree in five habitat types: (a) plantation, (b) swamp forest, (c) terra firme,  
 837 (d) urban, and (e) pole forest. For each habitat, the average F1 score across species per site in  
 838 the region of Loreto is also shown. Source data are provided as a Source Data file.



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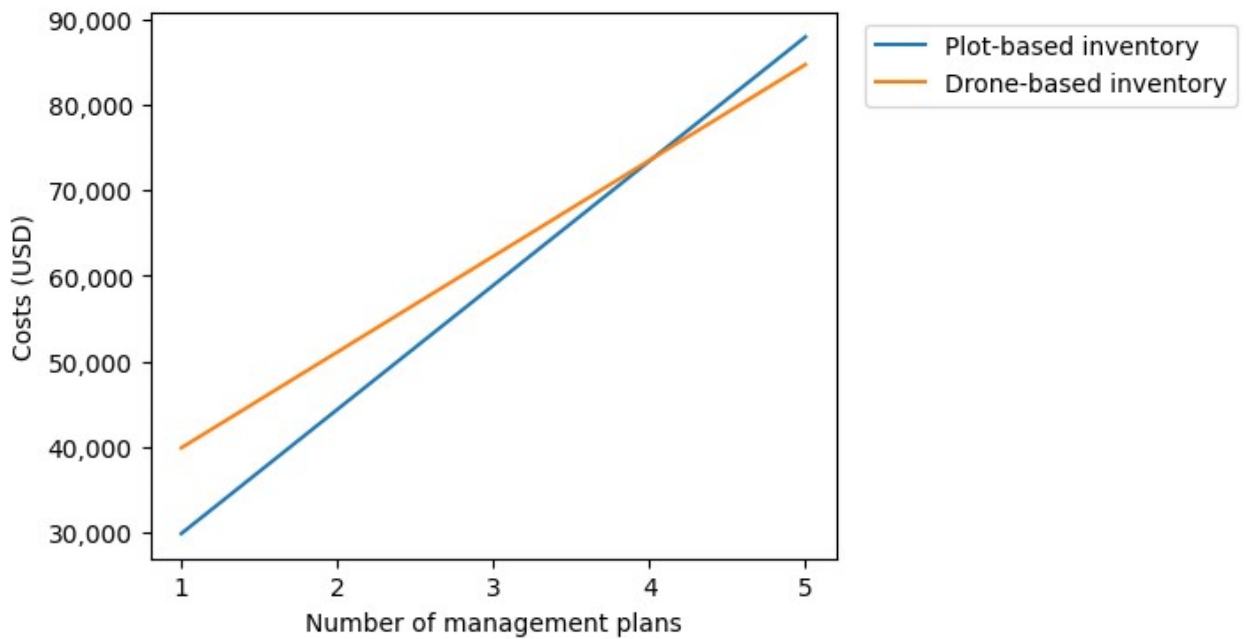
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Fig. 5. Variation in stem density of three arborescent palm species (*Mauritia flexuosa*, *Euterpe precatorea* and *Oenocarpus bataua*) across six UAV mosaics covering 70 to 230 hectares in Loreto (a-c) and Madre de Dios (d-f). (a) Parinari

843 community, palm swamp, (b) Nueva York Community, pole forest with no  
844 presence of *Oenocarpus*, (c) Piura community, palm swamp, (d) Sector Briolo  
845 – Elina, palm swamp/ *terra firme* forest, (e) Around the Sandoval lake, palm  
846 swamp/ *terra firme* forest, (f) Sector Briolo – Brigida, palm swamp/ *terra firme*  
847 forest.

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850 Fig. 6. Comparison of the total costs of producing different numbers of  
851 management plans between approaches that use traditional plot-based  
852 fieldwork or drone (UAV) surveys for mapping the abundance of arborescent  
853 palms. Source data are provided as a Source Data file.

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