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1 **Particulate number emissions during cold-start with diesel and biofuels: A special focus on**
2 **particle size distribution**

3
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28 **Abstract**

29 The share of biofuels in the transportation sector is increasing. Previous studies revealed that
30 the use of biofuels decreases the size of particles (which is linked to an increase in particulate
31 toxicity). Current emission regulations do not consider small particles (sub-23 nm); however,
32 there is a focus in future emissions regulations on small particles. These and the fact that
33 within cold-start emissions are higher than during the warmed-up operation highlight the
34 importance of a research that studies particulate matter emissions during cold-start. This
35 research investigates the influence of biofuel on PN and PM concentration, size distribution,
36 median diameter and cumulative share at different size ranges (including sub-23 nm and
37 nucleation mode) during cold-start and warm-up operations using diesel and 10, 15 and 20%
38 mixture (coconut biofuel blended with diesel). During cold-start, between 19 to 29% of total
39 PN and less than 0.8% of total PM were related to the nucleation mode (sub-50 nm). Out of
40 that, the share of sub-23 nm was up to 9% for PN while less than 0.02% for PM. By using
41 biofuel, PN increased between 27 to 57% at cold-start; while, the increase was between 4 to
42 19% during hot-operation. The median diameter also decreased at cold-start and the
43 nucleation mode particles (including sub-23 nm particles) significantly increased. This is an
44 important observation because using biofuel can have a more adverse impact within cold-
45 start period which is inevitable in most vehicles' daily driving schedules.

46

47 **Keywords:** Particle size distribution; PN; sub-23 nm; biofuel; cold-start.

48

49 **Word count:** 7338

Abbreviations

NOx	Nitrogen oxides
CO	Carbon Monoxide
HC	Hydrocarbon
PM	Particle mass
PN	Particle number
BSFC	Brake Specific Fuel Consumption
CMD	Count median diameter
ECU	Engine control unit
EU	European Union
VOCs	Volatile organic compounds

50

51

52 Highlights

- 53 1. This study investigated engine-out particle emissions and size distribution
- 54 2. Cold-start, hot-start, and two intermediate warm-up phases were studied
- 55 3. Nucleation mode particles increased as the engine warmed up
- 56 4. During cold-start, using biofuel decreased the size of particles
- 57 5. Sub-23 nm fraction contributed up to 0.02% and 9% to cumulative PM and PN

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61

62 **1. Introduction**

63 The transportation sector has been undergoing a significant transformation with the
64 utilisation of different strategies and technologies to reduce harmful emissions [1]. One of
65 these strategies is related to fuel. The adverse effect of using fossil fuels in the transportation
66 sector has created significant interest in renewable alternatives, such as biofuels. There have
67 been incentives in place to increase the share of such renewable alternatives. For example,
68 EC Directive 2003/30 was issued to increase the share of biofuel to 5.75% by 2010 and
69 Directive 2009/28/EC targeted a share increase to 10% by 2020. One of the reasons for
70 increasing the share of renewable fuel in the transportation sector can be the advantages it
71 has on engine performance and exhaust emissions [2-4].

72 In terms of engine performance parameters, it has been reported that using biofuels (such as
73 jatropha biodiesel) can improve the thermal efficiency of diesel engines when compared to
74 diesel fuel; however, it depends on the type of biofuel [5-7]. Friction parameters were also
75 reported to be lower with biodiesel owing to their better lubricity [8, 9]. However, fuel
76 consumption parameters such as brake specific fuel consumption (BSFC) were reported to be
77 higher and the engine power was reported to be lower with biofuel derived blends [10, 11].
78 The higher viscosity, higher density and lower calorific values of biofuels were reported to be
79 primary reasons. In terms of emissions, lower hydrocarbon (HC) and carbon monoxide (CO)
80 emissions were reported to be advantages of using biofuels [12-14]. However, it has been
81 frequently reported that nitrogen oxides (NOx) emissions increase with biofuels (some
82 articles claimed otherwise) [15-17]. Maybe the most highlighted advantage of using biofuel is
83 the decrease in particulate matter emissions; however, some articles reported a different
84 observation [18, 19]. It has been frequently reported that particulate matter emissions
85 decrease significantly due to the oxygen content of the biofuel [18, 20, 21].

86 Particulate matter emissions can adversely affect our environment and are identified as a
87 global risk factor as these emissions have been reported to be associated with
88 cardiorespiratory health problems [22-24]. It was shown that prolonged exposure to
89 particulate matter emissions is associated with an increase in free-radicals which adversely
90 impact health [25-27]. Particulate matter emissions can be evaluated from two inter-
91 correlated perspectives. The first aspect is the particulate mass (PM) which reports the mass
92 of particles. However, this measurement might not be able to provide sufficient information
93 when it comes to the health hazards from particles, as those very small particles which are
94 harmful have a lower contribution to PM [28]. However, the second perspective, particulate
95 number (PN), is more informative and has gained a lot of attention, therefore it became
96 mandated in the recent emissions regulations [29]. For passenger cars, reporting PN
97 emissions became mandated from the Euro 5 emissions regulation [29], while there was no
98 regulation on PN in Euro 1-4 [29]. The PN, which is the count of individual particles can include
99 small particles even those with nearly zero weight and contribution to PM.

100 Using biofuel has been reported to have different effects on PN emissions, some reported
101 higher PN with biofuels and some reported lower [18, 30]. However, most of the reports in
102 the literature showed that using biofuel decreases the size of particles [31, 32]. This is
103 important as it has been reported that a decrease in particle size is associated with an increase
104 in toxicity [33]. However, in the most recent emission regulations such as Euro 5 and 6 and
105 WLTP (worldwide harmonised light vehicles test procedure), there is a guideline for PN
106 measurement—PMP (particle measuring method)—which does not consider sub-23 nm
107 particles [34-36]. PMP does specify the count efficiency of 50% (D_{50}) at 23 nm and 90% (D_{90})
108 at 41 nm, and it has been reported in the literature that decreasing D_{50} from 23 nm to smaller
109 sizes such as 10 nm can significantly increase the PN emissions from vehicles [37, 38]. For

110 example, Leach et al. [38] reported that decreasing D_{50} and D_{90} to 10 nm and 23 nm leads to
111 36% higher PN. This and the increasing share of biofuel in the transportation sector highlight
112 the importance of studying PN emissions in more detail such as looking into the size of
113 particles.

114 PN emissions from an engine depend on different factors such as operating conditions [39].
115 For a high portion of vehicles, the cold-start operation is a norm which occurs mostly in the
116 morning and afternoon when people start their vehicle after some hours of engine-off and
117 drive between home and work. It has been reported that a significant number of trips
118 between home and work start and finish during the cold-start period [40]. For example, a
119 study of more than one thousand trips showed that more than 30% of the trips started and
120 finished within the cold-start period [41]. Regulation (EU Directive 2012/46/EU) considers the
121 cold-start period from engine start—after 12 hours soak (or 6 hours forced cooled)—either
122 for the first 5 minutes or during the time that the temperature of engine coolant increases to
123 70°C. Within this period, the engine temperature is sub-optimal affecting the combustion
124 process [42]. Consequently, engine emissions and performance are different in comparison
125 with hot-operation [43-45]. For example, fuel consumption and friction power were reported
126 to be higher when the engine is cold [46]. This was reported to be because of the high viscosity
127 of the engine oil at lower temperatures, which consequently leads to higher friction, therefore
128 more fuel needs to be burnt to maintain power [40, 46]. Fuel evaporation and atomisation
129 are also impacted by the lower temperature of the engine and fuel during cold-start which
130 also impacts emission and performance parameters [18].

131 It has been reported that during cold-start, emissions are higher than when the engine is fully
132 warmed up [47, 48]. For example, a study used a custom driving cycle and showed that during

133 the cold-start period of the cycle, PN emissions with biofuel increased significantly compared
134 to the hot-operation period [47]. Another study showed that around one-third of the emitted
135 PM emissions were related to the first 12% of the total distance (Phase 1) of the LA92 Unified
136 Driving Cycle [49]. It also reported that compared to Phase 3 of that cycle, in which the engine
137 was fully warmed up, during Phase 1, which was cold-start, the PM emissions were 7.5 times
138 higher. Some studies in the literature investigated other emissions such as CO, CO₂ and NO_x
139 [40, 50, 51]. However, there are a few studies that investigated PN emissions and size
140 distribution during cold-start in detail, when compared to other emissions [52]. Also, most of
141 the cold-start experimental studies in the literature used driving cycles such as NEDC (New
142 European Driving Cycle) or WLTC (worldwide harmonised light vehicles test cycle) which has
143 abrupt load/speed changes. The results of such studies were shown with the averaged value.
144 However, it is essential to study the influence of transient engine temperature while the
145 engine is warming up on exhaust emissions such as PN. It will be seen in this current study
146 that cold-start and engine warm-up have different stages and therefore different impacts on
147 exhaust emissions.

148 This research aims to study the influence of fuel and transient engine temperature at different
149 stages of engine warm-up (including cold-start) on particulate matter emissions from
150 different aspects including PN, PN and PM size distribution and median diameter, the share
151 of particles at different sizes, nucleation mode particles and sub-23 nm particles. The
152 increasing share of biofuels in the market, with the inherent decrease in particle size and the
153 associated increase in toxicity, increased emissions during cold-start coupled to future
154 emissions regulations on small particles, and the fact that the current emission regulations do
155 not consider sub-23 nm particles, all highlight the importance of this study. By using and
156 comparing different biofuel blending ratios (10, 15 and 20%), this study can be helpful to

157 engine researchers when it comes to nucleation mode and sub-23 nm particles the upcoming
158 emissions regulations.

159 **2. Methodology**

160 The engine used in this study was a Cummins ISBe220 diesel engine (designed and
161 manufactured by an American company) which is an in-line 6-cylinder, turbocharged,
162 common-rail engine, as shown in Table 1. The maximum torque and power with this engine
163 are 820 Nm (at 1500 rpm) and 162 kW (at 2500 rpm) and the engine was coupled to an in-
164 house hydraulic dynamometer (electronically-controlled). The experiments were done at
165 QUT Biofuel Engine Research Facility (BERF) in Brisbane, Australia.

166

167 Table 1 Engine specifications

Model	Cummins ISBe220 31
Aspiration	Turbocharged
Fuel injection	Common rail
Cylinders	6 in-line
Capacity	5.9 L
Bore × stroke	102 × 120 (mm)
Maximum torque	820 @ 1500 (Nm @ rpm)
Maximum power	162 @ 2500 (kW @ rpm)
Compression ratio	17.3:1

168

169 Figure 1 shows a schematic diagram of the experiment setup. Engine and dynamometer data
170 were collected with Dynolog software. The in-cylinder data such as crank angle and injection
171 signals were collected using a Kistler (6053CC60) transducer, Kistler type 2614 sensor and
172 DT9832 A-to-D convertor, which all were connected to an in-house National Instruments

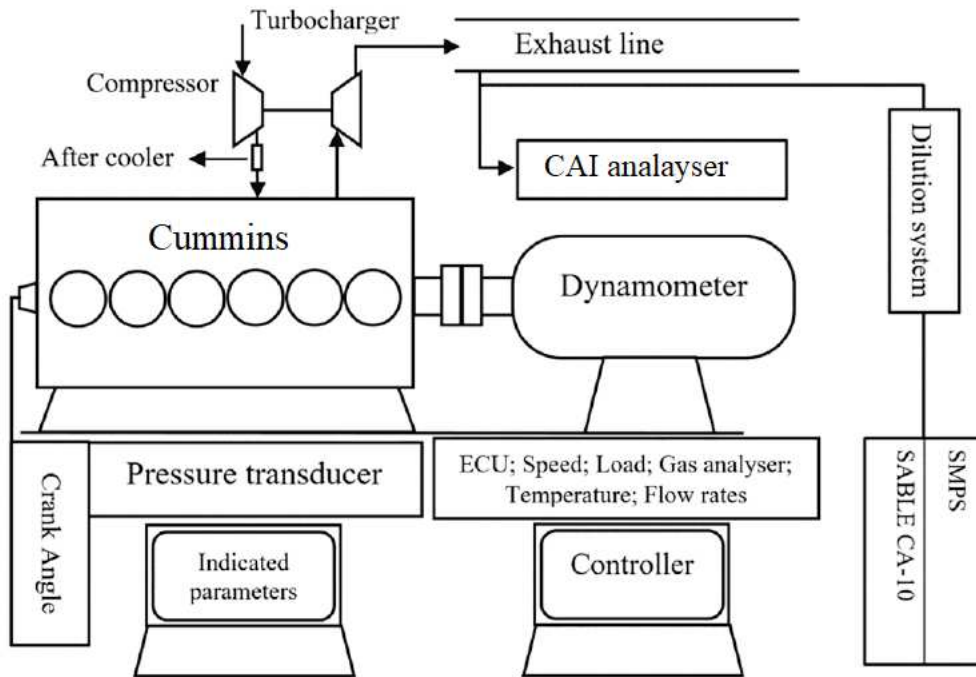
173 LabView program [53, 54]. The accuracy of the measuring instruments are shown in Appendix
174 (Table A1).

175 To better evaluate the pure effect of fuel properties and also the engine temperature, this
176 study evaluated the engine-out emissions instead of tailpipe emissions (which are sampled
177 before any after-treatment system). This way the emissions do not depend on the after-
178 treatment systems performance/type. Therefore, the fundamental study will not be limited
179 and results can give more information about the real engine dependent emission [55]. In
180 order to measure particulate matter emissions (such as particle number, mass, size
181 distribution), this study used a TSI scanning mobility particle sizer (SMPS) consisting of a TSI
182 3071A classifier (which preselects the particles in different sizes) and a TSI 3782 CPC (which
183 grows the particles to make them detectable for the optic). SMPS is designed and
184 manufactured by TSI which is an American company. SMPS has a size resolution capability of
185 128 channels per decade, which results in 192 channels in total. The exhaust gas (which had
186 a temperature of ~ 350 °C) was directed to the SMPS, but after being diluted (~ 20 times) with
187 ambient air (~ 23 °C) which was passed through a HEPA filter in a constant volume dilution
188 system (CVS). The CVS setup was followed by the European legislation (Commission
189 Regulation (EU) 2017/2400). To calculate the dilution ratio, a CAI-600 and a SABLE CA-10 CO₂
190 analysers (designed and manufactured by American companies) were used before and after
191 the dilution system measuring the CO₂ emissions.

192 The exhaust particles are made of solid particles, volatiles and liquid droplets. Usually, the
193 PMP method uses a volatile particle remover (VPR) system including three stages of hot
194 dilution (PND₁), heated evaporative tube (ET) and cold dilution (PND₂) to minimise the effect
195 of volatiles and liquid droplets ensuring that the particle counter measures solid particles

196 (ECE/TRANS/WP.29/GRPE/2016/3 amended by GRPE-72-09-Rev.2) [56]. Regarding the
197 difficulty of sub-23 nm measurement with current instruments in the market [36], it is worth
198 mentioning that SMPS is capable to measure particles above 10 nm including solid particles
199 and volatiles.

200 This study reports PN and PM, and also PN and PM size distribution. Regarding the PN and
201 PM size distribution, SMPS measures the number of concentration in a given channel (dN)
202 and divides it by the geometric width of the size of channel, and reports the normalised
203 number concentration ($dN/d\log D_p$), where D_p is the geometric midpoint of the particle size
204 channel. The conversion of PN to PM was done through the Aerosol Instrument Manager
205 Software for SMPS Spectrometer using the formula $dM = dN \cdot (\pi/6)D_p^3 \rho$, where ρ is density,
206 and reporting the normalised mass concentration using the formula $dM/d\log D_p = dN/d\log D_p$
207 $\cdot (\pi/6) D_p^3 \rho$ [57]. In general, based on the size distribution, the measured ultrafine particles
208 can be classified into two modes; nucleation mode and accumulation mode [28]. The
209 nucleation mode particles are defined as particles with diameters less than 50 nm, and the
210 size of accumulation mode particles is between 50-500 nanometres, and these definitions are
211 used in this study to better interpret the data [28].



212

213

Figure 1 Test setup schematic diagram

214

215 Given that this study intended to investigate particulate matter emissions based on the
 216 current and future emissions regulation requirements, the selection of the fuels was
 217 conducted in a way to cover past, current and future biofuel blending ratio in the market.
 218 Therefore, it used diesel, and then made three diesel/biofuel blends of 10, 15 and 20% (by
 219 volume) denoted as D90C10, D85C15 and D80C20 using coconut oil biofuel (this study is a
 220 part of a project which investigates the potential of using different biofuels in diesel engines
 221 in Marshal Islands which have a high resource of coconut). In the tested fuel names, D stands
 222 of diesel and the digits after that show the volume of diesel in the blend. And C stands for
 223 coconut biofuel and the two digits after C shows the blending ratio. These fuel blends will be
 224 evaluated against diesel (D100).

225 Table 2 shows the tested fuel properties and Table 3 shows the chemical composition of fuels
 226 analysed by a GC/MS instrument (Trace 1310 Gas chromatograph, model ISQ, single

227 quadrupole MS). Diesel (D100) contained aromatic compounds (benzene and its derivatives,
 228 naphthalene, xylene, phthalan, mesitylene) and aliphatic compounds (mainly alkanes with 7-
 229 13 carbons, low concentrations of limonene). D90C10, D85C15, and D80C20 also contained
 230 cycloalkanes, cyclohexane and cyclooctane. With D100, the aromatic content was higher
 231 compared to other fuels. Compared to neat diesel, fuels with a lower blending ratio do not
 232 change the fuel properties significantly. For example, the density of the tested fuels, D100,
 233 D90C10, D85C15 and D80C20 were 0.84, 0.843, 0.845 and 0.846 g/cc. Also, the lower heating
 234 value of the tested fuels, D100, D90C10, D85C15 and D80C20 were 41.77, 41.31, 41.09 and
 235 40.86 MJ/kg, respectively. However, even small changes in fuel properties can affect engine
 236 performance and emissions. For example, the higher density and lower calorific value of
 237 biofuel blends can negatively impact engine power and also fuel consumption parameters
 238 such as BSFC [8]. The fuel oxygen content of biofuel is another property that distinguishes
 239 biofuel from diesel (which has no oxygen content). It has been frequently reported that fuel
 240 oxygen content is the primary driver for decreased PM emissions with biofuels [18, 19, 58].

241

242 Table 2 Fuel properties

	D100	C100	D90C10	D85C15	D80C20
Density at 15 °C (g/cc)	0.84	0.87	0.84	0.84	0.85
Kinematic viscosity at 40 °C (mm ² /s)	2.64	4.82	2.86	2.97	3.08
Cetane number	53.30	58.60	53.83	54.10	54.36
Lower heating value (MJ/kg)	41.77	37.20	41.31	41.08	40.86
Higher heating value (MJ/kg)	44.79	39.90	44.30	44.06	43.81

243

244

245 Table 3 Fuel analysis using GC/MS (Model ISQ, single quadrupole MS, Trace 1310 Gas
 246 chromatograph).

	Area %				
	aromatic	aliphatic	cyclic hydrocarbons	hydrazide	oxygenated hydrocarbons
D90C10	0.0879-0.169	0.0866-0.367	0.119-1.12	1.09-2.62	0.192-0.305
D80C20	0.0433-0.0968	0.0398-0.741	0.0394-0.222		0.0416-0.076
D100	1.43-5.66	1-12.24			

247

248 The fuels used in this study were tested under constant load (25%) and speed (1500 rpm).

249 The rationale for using a constant load and speed was to facilitate the fundamental

250 investigation into the influence of fuel type & engine temperature under cold-start and during

251 engine warm-up. Most of the studies on cold-start in the literature used a drive cycle such as

252 NEDC or WLTC and compared the first part of the cycle with the rest and made the conclusion

253 about cold-start contribution. However, these driving cycles consist of frequent load and

254 speed changes which add more variables to the analysis, complicating and limiting the

255 investigation about the pure influence of fuel type and engine temperature on emissions

256 within cold-start period. Therefore, this experimental investigation used a constant speed and

257 engine load to limit the number of influential factors that can potentially aid a better

258 judgment on the effects of engine temperature and fuel properties.

259 Experiments were done every morning with at least 12-hours of engine-off period at the

260 ambient temperature on consecutive days in an engine laboratory. Before starting each test,

261 coolant temperature and lubricating oil temperature were checked to be the same as ambient

262 temperature, as per the regulation (EU Directive 2012/46/EU). Given that the engine room

263 had an air-conditioning system, the ambient temperature during the test stayed constant

264 (23 ± 5 °C). For each test, the engine was started and ran under a quarter load (25%) at the

265 speed of 1500 rpm for more than 30 min to stabilise. The statistical analysis of the test
266 repeatability is shown in Appendix.

267 In diesel engines, the formation of particulate matter emissions depends on a number of
268 variables such as engine operating condition and fuel properties [18]. During cold-start
269 operation, the engine temperature and fuel properties are significantly influential factors.
270 Figure 2 shows how the engine coolant temperature increases during cold-start operation
271 with the tested engine for one of the tests. According to the regulation, (EU Directive
272 2012/46/EU), the cold-start period is defined from when the engine starts after 12 hours soak
273 (or 6 hours forced cooled soak) at the ambient condition either for the first 5 minutes or
274 during the time that the temperature of engine coolant increases to 70°C. However, after this
275 period, the coolant temperature still has an increasing trend indicating the sub-optimal
276 engine temperature (Figure 2). Also, it can be seen that there is a lag between the engine
277 lubricating oil temperature and coolant temperature. Compared to the coolant temperature,
278 the engine lubricating oil temperature remains sub-optimal for a longer period). The sub-
279 optimal temperature period outside the formal cold-start boundaries can affect the engine
280 performance and exhaust emissions [59, 60]. Figure 2 also shows that for the tested engine,
281 even within the formal cold-start period, when the engine temperature reaches to 65°C, the
282 start of injection is changed by the injection strategy commanded by the ECU of the tested
283 engine, and this injection strategy change affects the exhaust emissions and engine
284 performance parameters. It is worth mentioning that while the start of injection changed, the
285 injection period (therefore, the injection mass) remained constant, as the injection was
286 controlled by the ECU and, the injection period was independent of the start of injection. Also,
287 there was no modification to the engine ECU/calibration for different fuel blends. Therefore
288 the start of injection timing stayed the same for all the tests.

289 As discussed, it can be seen that from the engine start until the stable operation, there are
290 some periods with different characteristics and variables. Therefore, in order to better
291 analyse the influence of fuel properties and engine temperature, this research divides the
292 engine warm-up period into four phases to minimise the number of variables in each phase.

293 • Phase 1: Formal engine cold-start, constant start of injection, coolant and oil
294 temperatures are less than 65°C.

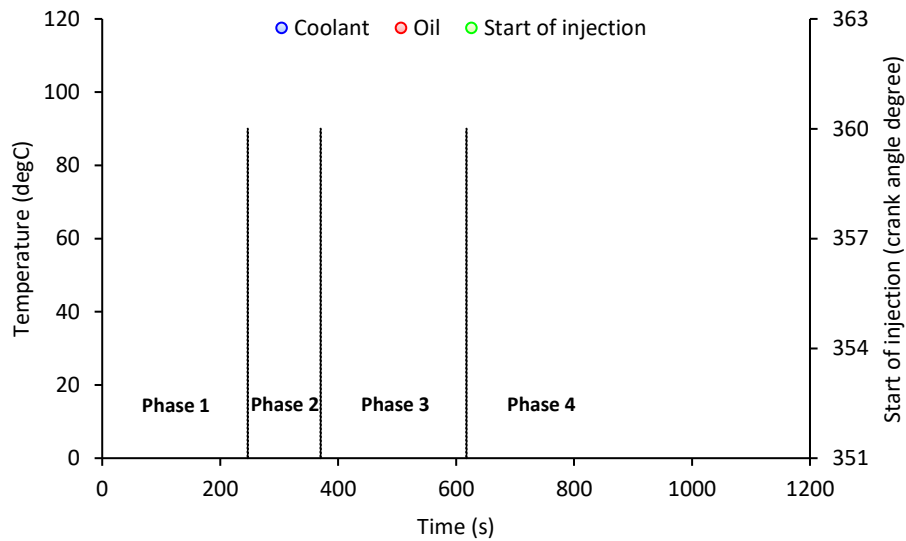
295 • Phase 2: Start of injection is increasing, coolant and oil temperatures above 65°C and
296 still increasing.

297 • Phase 3: Constant start of injection, optimal coolant temperature, sub-optimal oil
298 temperatures.

299 • Phase 4: Engine hot-operation, start of injection is constant, coolant and oil
300 temperatures are optimal.

301 This study used an SMPS analyser with a 2 min sampling time. The first two samples were
302 measured within Phase 1, the third sample was measured in Phase 2, the fourth and fifth
303 samples fell into Phase 3, and the last two samples were measured within Phase 4.

304



305

306 Figure 2 Engine coolant temperature, engine oil temperature and start of injection during the test
307 with diesel

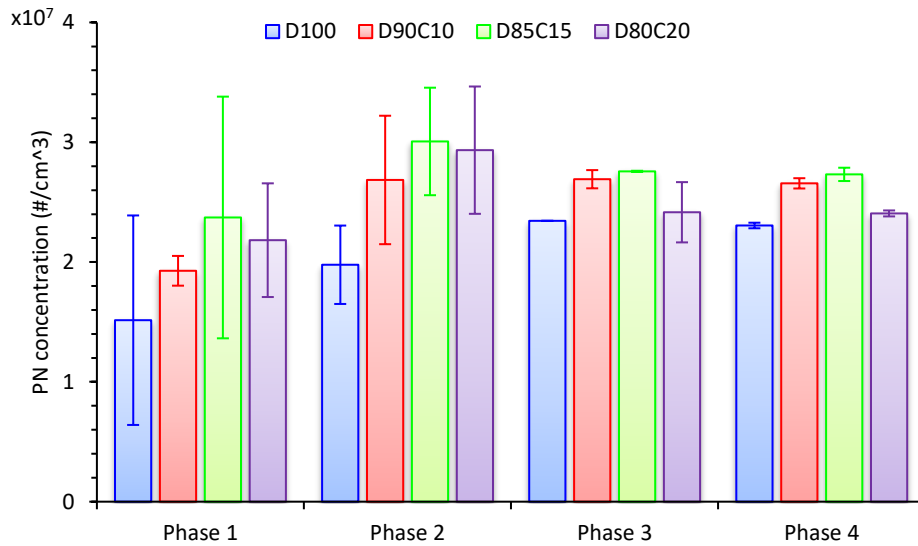
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309 3. Result and discussion

310 This section analyses the PN concentration and evaluates the particle number and mass size
311 distribution. It also investigates the median diameter and the share of PN and PM at different
312 sizes including accumulation mode, nucleation mode and sub-23 nm particles.

313 Figure 3 presents the PN concentration at the four different phases. It shows that for each
314 fuel the PN concentration in Phase 1 was lower than the other phases. Given that the engine
315 load and speed were constant within all phases, the engine temperature can be identified as
316 an influential factor, as shown in the literature [61]. Looking at each phase, it can be seen that
317 the use of biofuel increased PN emissions in all of the phases. This expected result shows that
318 fuel properties are influential, as reported before in the literature [62, 63].

319



320

321

Figure 3 PN concentration at different phases with all the tested fuels

322

323 When it comes to the potential health impact of particles, PN can also be analysed in more

324 detail by looking into the size distribution of particles [28, 33]. Reported hypotheses

325 mentioned that when compared to bigger particles, small particles penetrate in lungs deeper

326 and their relatively larger surface area increases the reaction with lung cells; a decrease in the

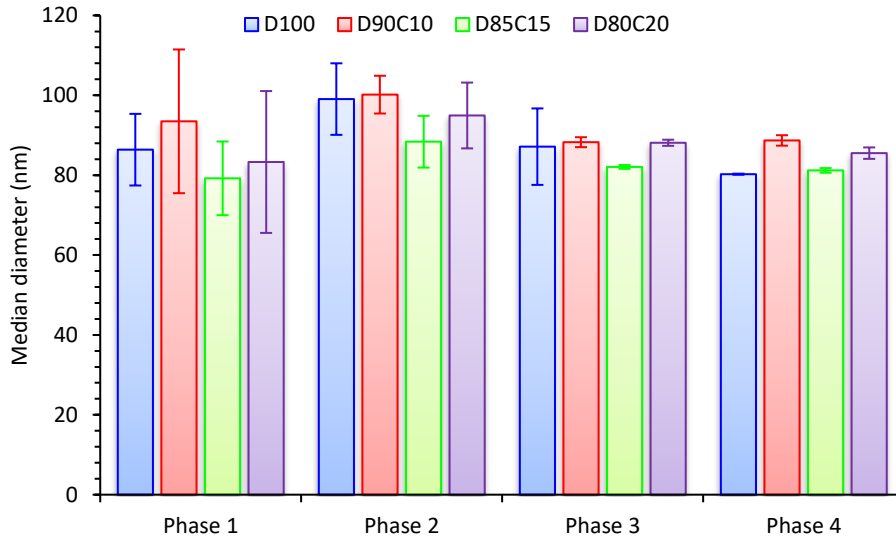
327 size of particles is also associated with an increase in toxicity [33]. Figure 4 shows the median

328 diameter of the particles with all the tested fuels at different phases. It can be seen that the

329 size of the particles changes as the engine warms up. From Phase 2 to Phase 4, it can be seen

330 that as the engine warms up, the size of particles slightly decreases [64].

331



332

333

Figure 4 PN median diameter at different phases with all the tested fuels

334

335 As discussed, cold-start operation and fuel properties are two factors that can significantly
 336 affect the PN concentration and particle size, shown in the literature as well [47]. Therefore,
 337 the following analysis is divided into two parts. It first analyses the data from the engine
 338 temperature and cold-start perspective and then discusses the influence of fuel properties.

339 3.1 Influence of engine temperature

340 Figure 5 shows the variation of PN concentration at the different phases as compared to Phase
 341 1 (cold-start). It can be seen that the PN concentration was higher at other phases when
 342 compared to Phase 1. As a constant engine load/speed was maintained through all of the
 343 phases, the engine temperature can be considered as the driving factor for any change [52].
 344 In Figure 5, the exhaust temperature is used as an indicator of the engine temperature at the
 345 different phases. It can be seen that the exhaust temperature increased from Phase 1 to
 346 Phase 4, however, the change between Phases 3 and 4 was not as significant as between
 347 Phases 1 and 2. The reason for using the exhaust temperature as an indicator instead of

348 coolant temperature, was that when the coolant temperature reaches to an optimum level,
349 the engine is still warming up.

350 Phase 1, which had the lowest PN, falls into the cold-start period. As per the regulation, EU
351 Directive 2012/46/EU, the cold-start period can last 5 min from the engine start or until the
352 engine coolant temperature reaches 70°C. During Phase 1, the engine coolant temperature
353 was less than 70°C. Also, during this phase, the start of injection did not change as the engine
354 coolant temperature was less than 65°C, which is the threshold for the injection strategy of
355 the tested Cummins engine.

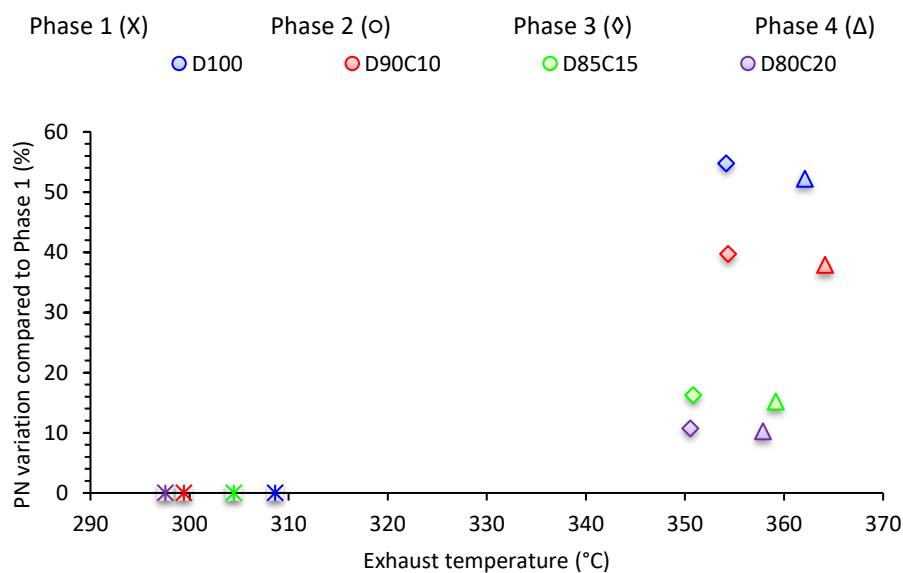
356 During Phase 2, the PN concentration was significantly higher than Phase 1. This increase was
357 between 27 to 39%. For example, with neat diesel, D100, the PN concentration increased by
358 31%, with D80C20 the PN concentration increased by 39%. The significant change between
359 Phases 1 and 2 is a combination of the increase in engine temperature and the injection
360 strategy change, with the latter being assumed as the dominant driver. During this unsteady
361 phase, coolant temperature increased to 65°C and the start of injection changed significantly
362 as by the ECU. This significant increase in the start of injection can be seen in Figure 2. Phase
363 2 cannot be considered as cold-start because the engine temperature within this phase
364 increased to higher than 70°C. In this phase, both the engine coolant and oil temperatures
365 were sub-optimal and still increasing.

366 Phase 3 had higher PN compared to Phase 1. The increase was between 11 to 55% with
367 different fuels. For example, with D80C20, the PN concentration increased by 11%, with D100,
368 PN increased by 55%. Within this phase, the engine coolant temperature was optimal.
369 However, this phase cannot be considered as hot-operation (stable operation) given that
370 within this phase the engine oil temperature was sub-optimal and still increasing. Comparing

371 this to Phase 2, within this phase, the start of injection stayed constant and did not change
372 until the end of the test.

373 Phase 4 can be considered as hot-operation given that within this phase, the engine
374 load/speed and the start of injection stayed constant, and also the engine oil and coolant
375 temperatures were optimal. Phase 4 shows that the PN concentration during this hot-
376 operation phase was higher than cold-start (Phase 1). The range of increase with different
377 fuels was from 10 to 52%. For example, with D100, the PN concentration was 52% higher than
378 Phase 1, with D80C20, the PN concentration was 11% higher. It can also be seen that, for all
379 of the fuels, the PN concentration in Phase 3 is slightly higher than Phase 4 (~2%). The
380 difference between Phase 3 and Phase 4 shows the effect of sub-optimal oil temperature
381 when the coolant temperature is optimal.

382

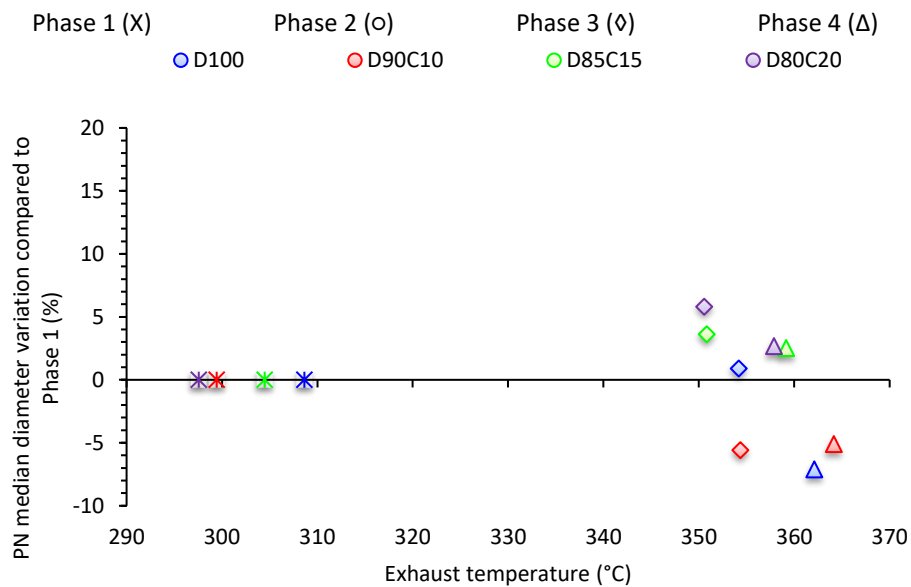


383

384 Figure 5 PN variation compared to Phase 1 vs. exhaust gas temperature at different phases with all
385 the tested fuels

386 The main contributors to PN emissions are nanoparticles [28]. Figure 6 shows that for each
 387 fuel, compared to Phase 1, the median diameter increased significantly in Phase 2 and then
 388 decreased in Phases 3 and 4. It is reported in the literature that increased exhaust gas
 389 temperature is associated with an increase in nanoparticles [65]. This can be seen in Figure 7,
 390 where for each fuel from Phase 2 to Phase 4 the increased exhaust temperature is associated
 391 with a decreased median diameter. This can also be seen in Figure 7 where moving from Phase
 392 2 to Phase 4 is associated with an increased number of nanoparticles.

393



394

395 Figure 6 Median diameter variation compared to Phase 1 vs. exhaust gas temperature at different
 396 phases with all the tested fuels

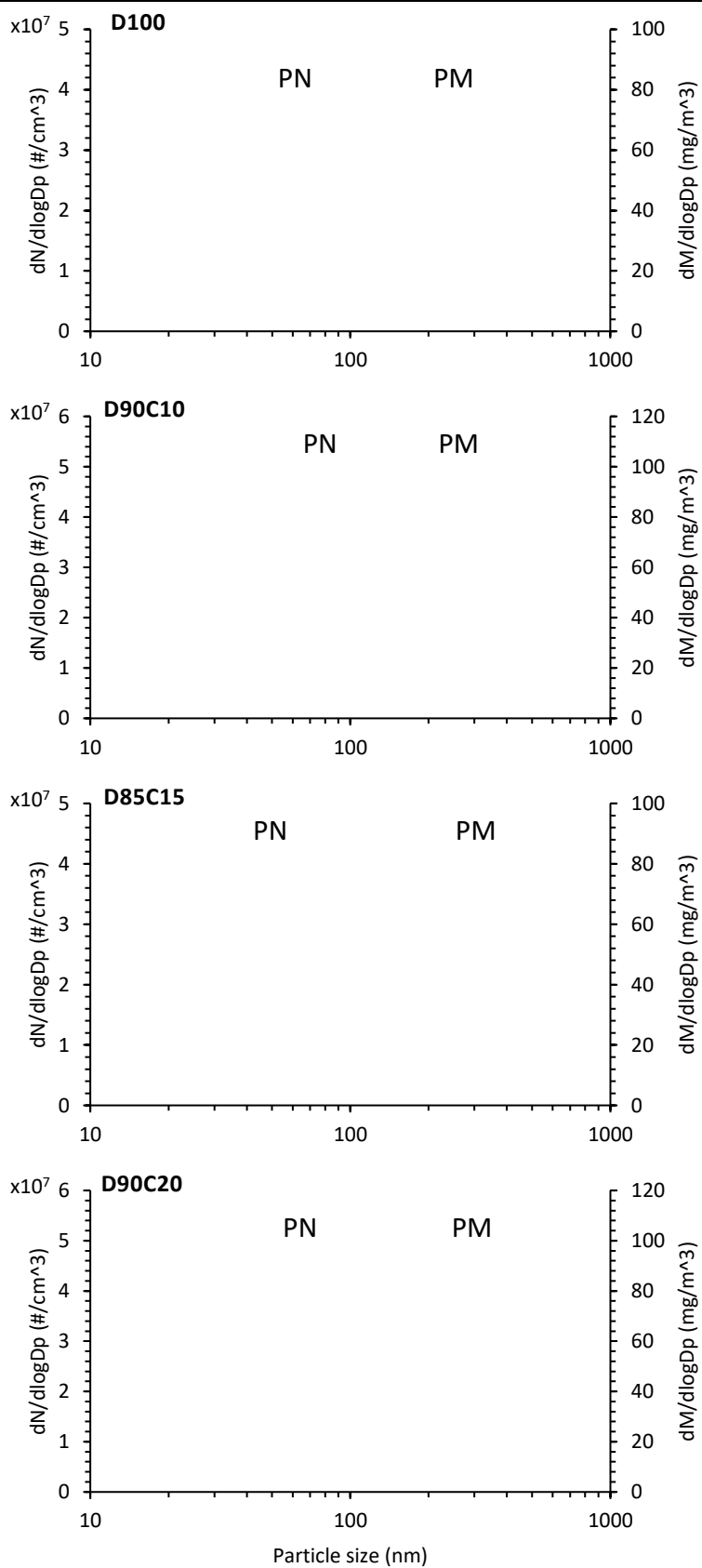
397

398 Figure 7 shows the PN and PM size distribution at the four phases of engine warm-up. In this
 399 study, particles are categorised into nucleation (5-50 nm) or accumulation modes (50-500 nm)
 400 [28]. Nucleation mode particles, which contribute negligibly to PM, but significantly to PN, are
 401 predominately volatile organic compounds (VOCs) and sulfur compounds—formed within

402 dilution and sampling—and also metal compounds and solid carbons—formed during the
403 combustion process [28]. In Figure 7, comparing the PN and PM size distribution graphs for
404 each fuel at each phase, which are shown with the same colour but two different shapes (o
405 and Δ) in each sub-figure, shows that for the size range of 5 to 50 nm in which the number of
406 particles was significantly high, the mass of particles was low. This can be better observed in
407 Figure 8 where the cumulative share of PN and PM at different size ranges is presented. As
408 can be seen, for all of the fuels at all the phases, the cumulative share of sub-50 nm particles
409 was between 0.3 to 0.9% of the total mass, while depending on the phase/fuel 11 to 29% of
410 the total PN were related to sub-50 nm. Also, depending on phase/fuel, for the sub-30 and
411 sub-23 nm particles, the PM cumulative share were up to 0.06 and 0.02%; while, their
412 cumulative PN share were 13 and 9%, respectively. Therefore, excluding sub-23 nm particles
413 in the current regulation means neglecting a significantly high number of small and toxic
414 particles, considering the number of vehicles in cities and the PN limit in the emissions
415 standards (e.g. 6×10^{11} #/km in Euro 6 and WLTP).

416

○ Phase 1 ● Phase 2 ● Phase 3 ● Phase 4



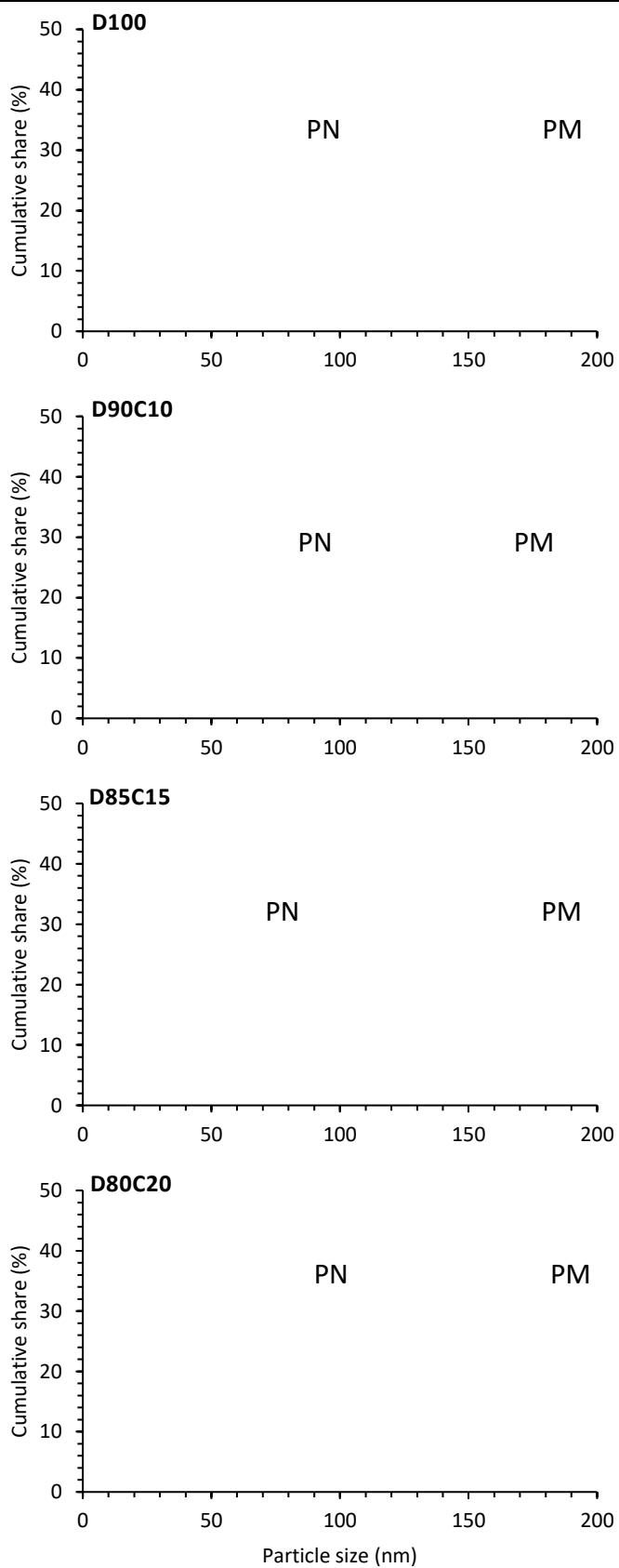
417

Figure 7 PN and PM size distribution at different phases with all the tested fuels

418 The main contributor to the accumulation mode particles are carbonaceous agglomerates
419 and adsorb materials [28]. The agglomeration of particles in the nucleation mode by the
420 condensation of volatile materials can also form some particles in the accumulation mode
421 [28]. Figure 8 shows that sub-100 nm particles contributed to 53-67% of PN while only to 8-
422 14% of PM. 92-96% of PN was related to sub-200 nm, while only 50-66% of the total mass was
423 related to these particles. It can be seen that in the accumulation mode, the share of PN
424 decreased as the share of PM increased. Comparing PN and PM size distribution graphs in
425 Figure 7 shows that the contribution of particles above 50 nm to PM was significant. This is
426 more significant for the sizes above 100 nm where the share of PN was decreasing. This can
427 be better explained by comparing Figure 9, which shows the PM median diameter, with Figure
428 4, which shows the PN median diameter. It can be seen that the PM median diameter was
429 between 173 to 213 nm, while for PN the median diameter changed between 79 to 100 nm.
430 This can clearly show that accumulation particles have a major contribution to PM. The PN
431 and PM size distribution in each phase are different and need to be further investigated.

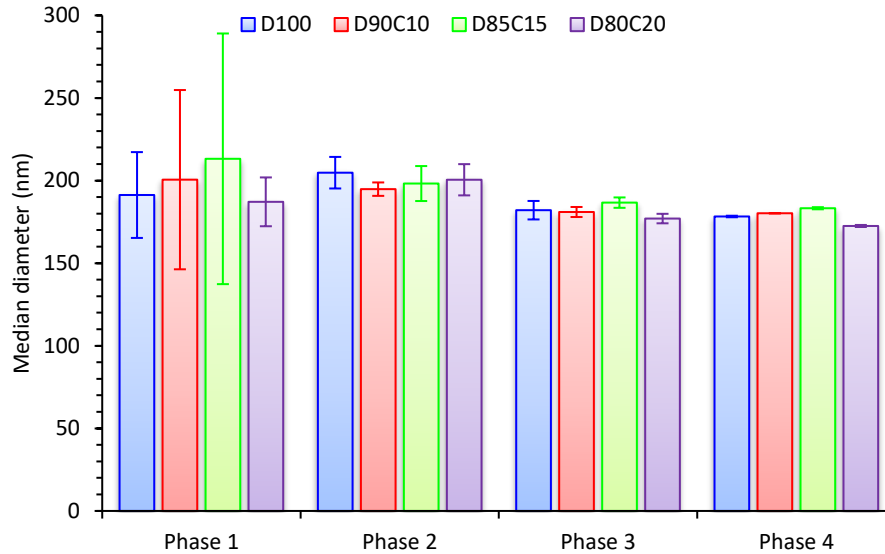
432

○ Phase 1 ● Phase 2 ● Phase 3 ● Phase 4



433

Figure 8 Cumulative PN and PM share at different phases with all of the tested fuels



434

435 Figure 9 PM median diameter at different phases with all of the tested fuels

436

437 A decrease in the size of particles can indicate a higher share of nucleation mode. The median
 438 diameter trend observed in Figure 4 reflects the trend of PN share in the nucleation mode in
 439 Figure 8. The median diameter of the particles increases from Phase 1 to Phase 2 (because of
 440 the change in injection strategy) and decreases gradually through the engine warm-up period.
 441 Therefore, an inverse trend can be seen for the share of nucleation mode particles where the
 442 cumulative PN share of sub-50 nm and sub-23 nm particles decreased from Phase 1 to Phase
 443 2 and then increases gradually.

444 Figure 4 showed that within Phase 1, which is the formal cold-start period, the PN median
 445 diameter with different fuels was between 79 to 94 nm. Figure 6 also showed how the median
 446 diameter changed through different phases as the engine exhaust temperature increased.
 447 Figure 8 shows that in Phase 1, with different fuels, between 19 to 29% of PN and less than
 448 0.8% of PM are in the nucleation mode (5-50 nm). Out of that, the share of sub-23 nm was
 449 between 1 to 9% for PN and less than 0.02% for PM. During cold-start (Phase 1), the low
 450 temperature of the exhaust pipe can decrease the exhaust gas temperature and within this

451 cooling process, the volatile materials can nucleate homogeneously into particles [66]. The
452 reason for this is that a decrease in the exhaust gas temperature increases the saturation
453 vapour pressure and saturation ratio of the volatile materials, which can lead to nucleation
454 and condensation of volatile materials [67]. It is reported that VOCs during cold-start are
455 higher than hot-start [68], therefore there is a high chance of nucleation. This is more
456 significant with biofuel blends owing to their higher VOCs compared to diesel [69].

457 Figure 7 shows that in Phase 2, the particle size distribution graph tends toward higher PN
458 and also bigger particles when compared to Phase 1. This can be better seen in Figure 6 where
459 the median diameter of particles increased from Phase 1 to Phase 2. As an example, the PN
460 median diameter increased from 86 nm to 99 nm with D100. The increase in median diameter
461 from Phase 1 to Phase 2 was between 7 to 15%. Moving toward bigger particles in Phase 2
462 can also be seen by comparing the red colour to the blue colour in Figure 7 and in Figure 8
463 where the cumulative share of PN in the nucleation mode decreased significantly in Phase 2
464 compared to Phase 1. For example, with D100, the share of nucleation mode particles
465 decreased from 20% to 12%, with D85C15 the decrease was from 30 to 20%. The share of
466 sub-23 nm particles also decreased in Phase 2. The reason for this move toward bigger
467 particles can be due to the injection strategy change within this unsteady warm-up phase in
468 which the start of injection increased significantly and led to decreased ignition delay
469 adversely affecting fuel atomisation and premixed combustion. The other variable between
470 Phases 1 and 2 was the engine temperature, however, this parameter might not be the reason
471 for the increased diameter size. The observed trend in Figure 6 showed that increasing the
472 engine temperature can lead to a decrease in the size of particles. The decreasing trend of
473 median diameter between Phase 2 to Phase 4 in Figure 6 can show that.

474 In Phase 3, particles were smaller than Phase 2, however, they are still bigger than Phase 1
475 (with most of the fuels). This can be seen in Figure 6 and Figure 7. In this phase, the median
476 diameter of the particles with different fuels was mostly between 82 to 88 nm (Figure 4). In
477 this phase, the start of injection stayed constant, however, the engine temperature—
478 presented by exhaust temperature—was still increasing. Comparing Phase 3 to Phase 2,
479 Figure 6 shows that the increasing trend of exhaust temperature is associated with a decrease
480 in median diameter. This can also be seen in Figure 8 where the share of nucleation mode
481 particles increased compared to Phase 2. For example, with D100, the sub-50 nm particle
482 share increased from 12 to 18% and from 14 to 17% with D85C15. Also, in this phase, the
483 share of sub-23 nm particles increased.

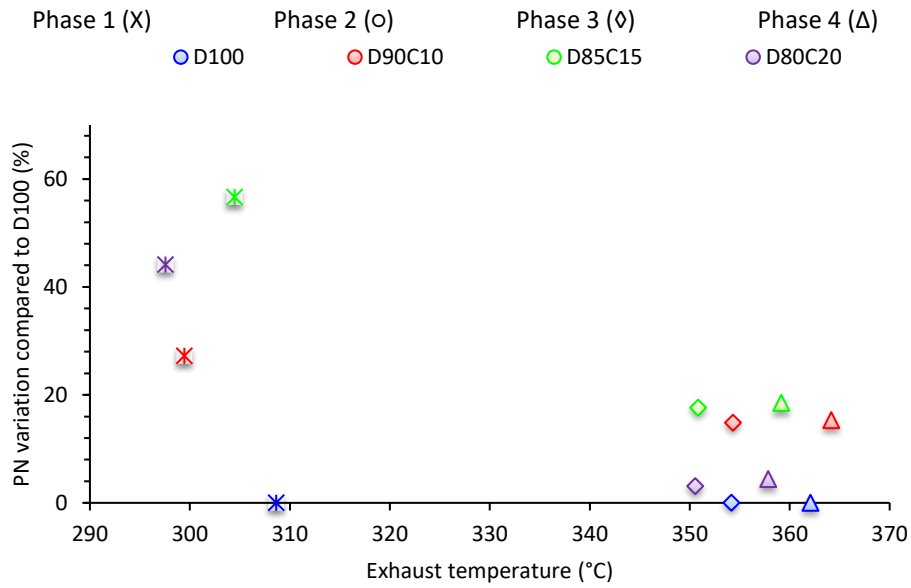
484 Comparing Phase 4 to Phase 3 and Phase 2 shows that the increased exhaust temperature is
485 associated with a decrease in particle size. This decreasing trend between the median
486 diameter and exhaust temperature can be seen in Figure 6. It can also be seen in Figure 8 that
487 the share of nucleation mode particles in Phase 4 was higher than Phase 3. This can be due
488 to the effect of sub-optimal engine oil temperature. However, the change was small with most
489 of the fuels. For example, with D80C20, the nucleation mode particles increased from 17 to
490 18%. Comparing Phase 4 to Phase 1 shows a different trend with different fuels, which means
491 that the fuel properties are likely to be the driving factor.

492 **3.2 Effect of fuel**

493 Figure 3 showed that the PN concentration increased in all of the phases when the fuel blends
494 were used instead of diesel. For example, using D90C10 increased PN by 27, 36, 15 and 15%
495 from Phase 1 to 4, respectively. The reason for this can be the higher volatile organic
496 compounds (VOCs) of biofuels when compared to diesel [69]. It was also observed in Figure 5

497 that compared to D100, these fuels showed less difference between cold-start and hot-
498 operation. Comparing Phase 3 and 4 to Phase 1, in Figure 5, shows that the PN variation
499 decreased by increasing the share of biofuel in the blend. For example, with D100, D90C10,
500 D85C15 and D80C20, PN during Phase 4 was 52, 38, 15 and 10% higher than Phase 1.
501 However, this can be misleading, and the data can be evaluated from another aspect. Figure
502 10 shows how the PN concentration changed with different fuels compared to D100. It can
503 be seen that using D90C10, D85C15 and D80C20, instead of D100, increased the PN
504 concentration in all of the phases. However, the increase was more significant during cold-
505 start. For example, in Phase 1, which is cold-start, using these fuels instead of D100 increased
506 PN between 27 to 57%; while, during hot-operation the increase range was between 4 to 19%.
507 This is an important observation because using these fuels has more adverse impacts within
508 the cold-start period, which is an inevitable part of daily driving for most vehicles. This analysis
509 shows that the evaluation of using such fuels needs to be done not only during hot-operation,
510 but also cold-start. A reason for the increased PN with the fuel blends during cold-start could
511 be the higher viscosity of biofuels, which adversely affects fuel atomisation and evaporation
512 [70]. Another reason could be the increased number of small particles. The higher oxygen
513 content of biofuels can also result in higher PN. In biofuels, there is more oxygen available in
514 the center of the diffusion flame (where usually the concentration of HC is higher). This could
515 result in smaller primary particle diameters and an increase of smaller size particles.

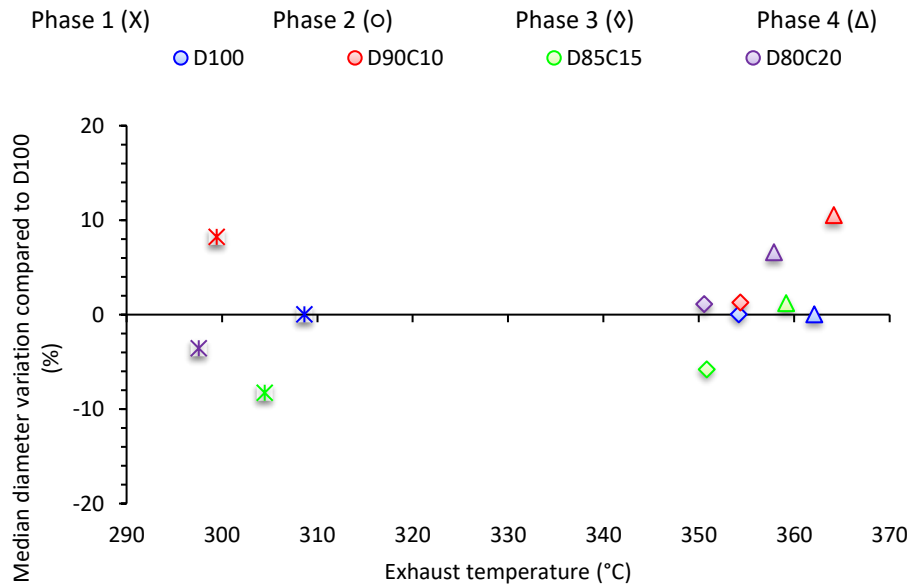
516 The adverse effect of using biofuels can be further highlighted in Figure 11, where the PN
517 median diameter with these fuels during the first two phases is smaller than D100 (in most of
518 the cases); while during Phase 4, D100 has the lowest median diameter. This means that these
519 fuels emit smaller particles during cold-start. This is important as a decrease in particle size
520 has been shown to be associated with an increase in toxicity [33].



521

522 Figure 10 PN variation compared to D100 vs. exhaust gas temperature at different phases with all of
 523 the tested fuels

524



525

526 Figure 11 Median diameter variation compared to D100 vs. exhaust gas temperature at different
 527 phases

528

529 The increase in smaller particles is shown in Figure 12 which shows the cumulative share of
 530 nucleation mode particles at different phases. Figure 12 shows that during cold-start the

531 share of nucleation mode particles increased significantly when D100 was replaced by the
532 fuel blends. For example, the share of sub-50 nm particles with D100, D90C10, D85C15 and
533 D80C20 were 20, 20, 29 and 23% in Phase 1, and 12, 12, 20 and 14% in Phase 2, respectively.
534 A similar trend was also observed for sub-23 nm particles. For example, the share of these
535 particles in Phase 1 increased from 1 to 9% when D100 was replaced with D85C15.

536 It was mentioned that nucleation mode particles (sub-50 nm) are mainly volatile organic
537 compounds (VOCs), sulfur compounds, metal compounds and solid carbons [28]. It is known
538 that biofuels have higher VOCs compared to diesel [69], therefore the higher VOCs of the
539 tested fuels compared to D100 could be the reason for the increased share of nucleation
540 mode particles and therefore the total PN. A greater amount of VOCs with biofuel was also
541 reported by Hedayat et al. [71], which used the same type of engine as this study, under a
542 stationary cycle.

543

○ D100 ○ D90C10 ○ D85C15 ○ D80C20

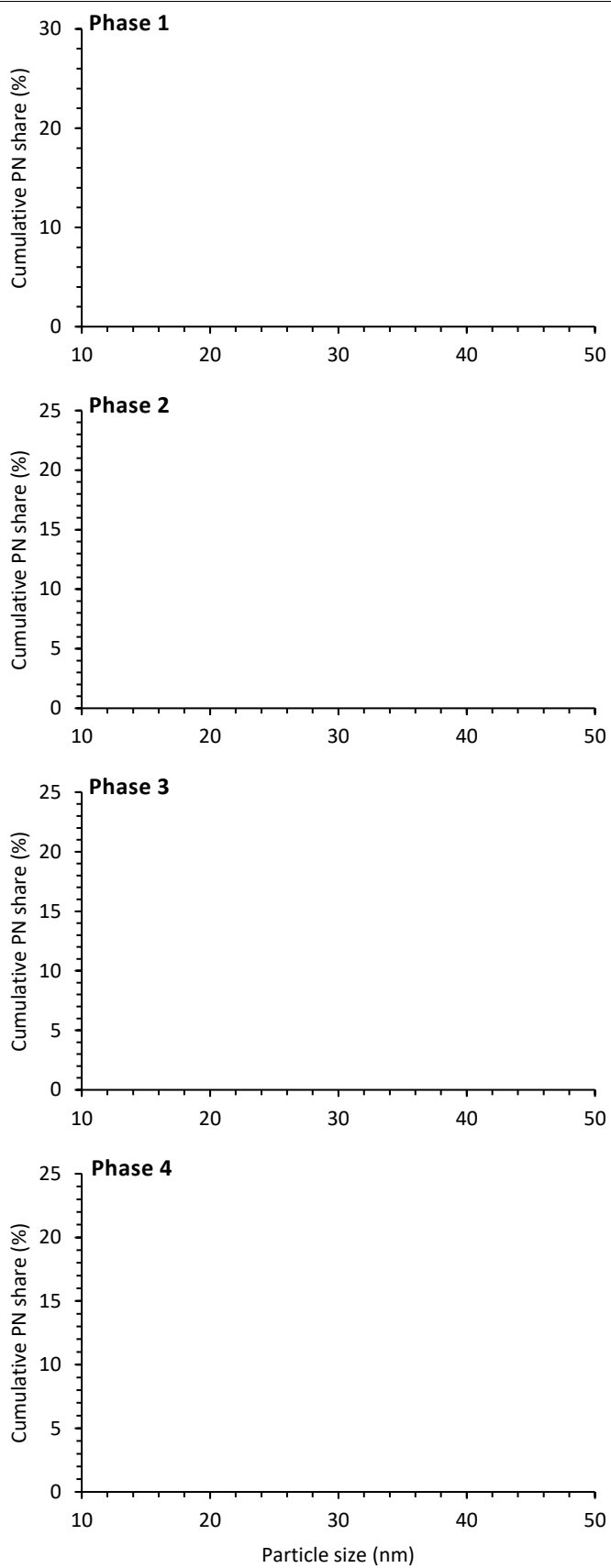


Figure 12 Cumulative PN share at different phases with all of the tested fuels

545 **3.3 Practical implications of this study**

546 Due to the high price and adverse health effects of fossil fuels, there is a focus to increase the
547 use of biofuels and measures have been in place. However, vehicle emissions are still an issue
548 and governments and authorities are trying to limit the amount of emission by tightening the
549 emissions levels from vehicles in the emission regulations. This research is related to the
550 future emissions regulations (Euro 7) and introduces important knowledge about regulated
551 and currently-unregulated particulate number emissions during cold-start and engine warm-
552 up, which is an emerging area in the literature. This study shows that when it comes to very
553 small particles (which are more toxic), the current market trend (increasing the share of
554 biofuel) can have a more adverse impact (because of more small particles) within the cold-
555 start period—which is inevitable in most vehicles’ daily driving schedule. But, even the latest
556 emission regulations, such as WLTP, only consider particles with a size above 23 nm. Most
557 cold-start operations occur in residential areas, and this study shows that a significant portion
558 of emitted particles are smaller than 23 nm during cold-start, and also the current fuel market
559 leads to an even higher number of these small particles. And the current regulations do not
560 include a practical enough measure to limit such emissions. Including sub-23 nm particles in
561 regulations (through PMP) increases the number of particles measured during the
562 homologation tests, and this might lead many vehicles to fail to comply with the PN limit in
563 emissions certification tests. However, passing or failing the certification test does not change
564 the fact that these days many certified vehicles on the street are emitting a huge number of
565 small particles affecting people’s health and still the emission regulations are unable to stop
566 them. This study emphasizes the importance of including smaller (sub-23 nm) particles in
567 future regulations, or even in the new amendments of the current regulations.

568 **4. Conclusions**

569 Using a diesel engine fueled by diesel and biofuels, this research evaluated PN, PN and PM
570 median diameter, PN and PM size distribution, the share of particles at different sizes
571 including nucleation mode and sub-23 nm particles. This research divided the engine warm-
572 up period into four phases of cold-start (Phase 1) to hot-operation (Phase 4) and two
573 intermediate unsteady warm-up phases which could be counted neither as cold-start (defined
574 in the regulations) nor as hot-operation.

575 Results showed that for all of the fuels, compared to Phase 1, PN increased by 27-39% in
576 Phase 2, 11-55% in Phase 3 and 10-52% in Phase 4. In the cold-start phase, between 19 to
577 29% of the total PN and less than 0.8% of the total PM were related to particles in the
578 nucleation mode (sub-50 nm). Out of that, the share of sub-23 nm was between 1 to 9% for
579 PN and less than 0.02% for PM. By using biofuel blends instead of diesel, PN increased
580 between 27 to 57% during cold-start; while, during hot-operation, the increase in PN ranged
581 between 4 to 19%. Also, the PN median diameter decreased during the first two phases and
582 nucleation mode particles increased significantly. For all of the fuels at the different phases,
583 the PM median diameter was between 173 to 213 nm, while for PN the median diameter was
584 between 79 to 100 nm. The cumulative share of sub-50 nm particles was 0.3-0.9% of the total
585 mass, while 11-29% of the total PN. For the sub-30 and sub-23 nm particles, the PM
586 cumulative share was up to 0.06 and 0.02%; while, their PN cumulative share was 13 and 9%,
587 respectively. The contribution of particles above 50 nm to PM was significantly high. This was
588 more significant for the sizes above 100 nm where the share of PN was decreasing. Sub-100
589 nm particles contributed to 53-67% of PN while only to 8-14% of PM. 92-96% of PN was
590 related to sub-200 nm, while only 50-66% of the total mass was related to this size range. In
591 the accumulation mode, the share of PN decreased as the share of PM increased.

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596

597 6. References

- 598 1. Joshi, A., *Review of Vehicle Engine Efficiency and Emissions*. SAE International Journal
599 of Engines, 2020. 10.4271/2020-01-0352.
- 600 2. Verma, T.N., P. Nashine, P.K. Chaurasiya, U. Rajak, A. Afzal, S. Kumar, D.V. Singh and
601 A.K. Azad, *The effect of ethanol-methanol-diesel-microalgae blends on performance,*
602 *combustion and emissions of a direct injection diesel engine*. Sustainable Energy
603 Technologies and Assessments, 2020. **42**: p. 100851.
604 <https://doi.org/10.1016/j.seta.2020.100851>.
- 605 3. Sathyamurthy, R., D. Balaji, S. Gorjian, S.J. Muthiya, R. Bharathwaaj, S.
606 Vasanthaseelan and F.A. Essa, *Performance, combustion and emission characteristics*
607 *of a DI-CI diesel engine fueled with corn oil methyl ester biodiesel blends*. Sustainable
608 Energy Technologies and Assessments, 2021. **43**: p. 100981.
609 <https://doi.org/10.1016/j.seta.2020.100981>.
- 610 4. Abrar, I. and A.N. Bhaskarwar, *Performance and emission characteristics of constant*
611 *speed diesel engine fueled by surfactant-free microemulsions*. Sustainable Energy
612 Technologies and Assessments, 2021. **47**: p. 101414.
613 <https://doi.org/10.1016/j.seta.2021.101414>.
- 614 5. Elkelawy, M., A.E. Kabeel, E.A. El Shenawy, H. Panchal, A. Elbanna, H.A.-E. Bastawissi
615 and K.K. Sadasivuni, *Experimental investigation on the influences of acetone organic*
616 *compound additives into the diesel/biodiesel mixture in CI engine*. Sustainable Energy
617 Technologies and Assessments, 2020. **37**: p. 100614.
618 <https://doi.org/10.1016/j.seta.2019.100614>.
- 619 6. Nabi, M.N., A. Zare, F.M. Hossain, T.A. Bodisco, Z.D. Ristovski and R.J. Brown, *A*
620 *parametric study on engine performance and emissions with neat diesel and diesel-*
621 *butanol blends in the 13-Mode European Stationary Cycle*. Energy Conversion and
622 Management, 2017. **148**: p. 251-259.
623 <https://doi.org/10.1016/j.enconman.2017.06.001>.
- 624 7. Verma, T.N., P. Shrivastava, U. Rajak, G. Dwivedi, S. Jain, A. Zare, A.K. Shukla and P.
625 Verma, *A comprehensive review of the influence of physicochemical properties of*
626 *biodiesel on combustion characteristics, engine performance and emissions*. Journal
627 of Traffic and Transportation Engineering (English Edition), 2021. **8**(4): p. 510-533.
628 <https://doi.org/10.1016/j.jtte.2021.04.006>.
- 629 8. Zare, A., M.N. Nabi, T.A. Bodisco, F.M. Hossain, M.M. Rahman, Z.D. Ristovski and R.J.
630 Brown, *The effect of triacetin as a fuel additive to waste cooking biodiesel on engine*

- 631 *performance and exhaust emissions*. Fuel, 2016. **182**: p. 640-649.
632 <https://doi.org/10.1016/j.fuel.2016.06.039>.
- 633 9. Zare, A., R.J. Brown and T. Bodisco, *Ethanol Fumigation and Engine Performance in a*
634 *Diesel Engine, in Alcohol as an Alternative Fuel for Internal Combustion Engines*.
635 2021, Springer. p. 191-212.
- 636 10. Nabi, M.N., A. Zare, F.M. Hossain, M.M. Rahman, T.A. Bodisco, Z.D. Ristovski and R.J.
637 Brown, *Influence of fuel-borne oxygen on European Stationary Cycle: Diesel engine*
638 *performance and emissions with a special emphasis on particulate and NO emissions*.
639 Energy Conversion and Management, 2016. **127**: p. 187-198.
640 <https://doi.org/10.1016/j.enconman.2016.09.010>.
- 641 11. Zare, A., T.A. Bodisco, M.N. Nabi, F.M. Hossain, Z.D. Ristovski and R.J. Brown, *Engine*
642 *Performance during Transient and Steady-State Operation with Oxygenated Fuels*.
643 Energy & Fuels, 2017. **31**(7): p. 7510-7522. 10.1021/acs.energyfuels.7b00429.
- 644 12. Nabi, M.N., A. Zare, F.M. Hossain, Z.D. Ristovski and R.J. Brown, *Reductions in diesel*
645 *emissions including PM and PN emissions with diesel-biodiesel blends*. Journal of
646 Cleaner Production, 2017. **166**: p. 860-868.
647 <https://doi.org/10.1016/j.jclepro.2017.08.096>.
- 648 13. Bencheikh, K., A.E. Atabani, S. Shobana, M.N. Mohammed, G. Uğuz, O. Arpa, G.
649 Kumar, A. Ayanoglu and A. Bokhari, *Fuels properties, characterizations and engine*
650 *and emission performance analyses of ternary waste cooking oil biodiesel–diesel–*
651 *propanol blends*. Sustainable Energy Technologies and Assessments, 2019. **35**: p.
652 321-334. <https://doi.org/10.1016/j.seta.2019.08.007>.
- 653 14. Attia, A.M.A., M. Nour, A.I. El-Seesy and S.A. Nada, *The effect of castor oil methyl*
654 *ester blending ratio on the environmental and the combustion characteristics of*
655 *diesel engine under standard testing conditions*. Sustainable Energy Technologies
656 and Assessments, 2020. **42**: p. 100843. <https://doi.org/10.1016/j.seta.2020.100843>.
- 657 15. Lapuerta, M., O. Armas and J. Rodríguez-Fernández, *Effect of biodiesel fuels on diesel*
658 *engine emissions*. Progress in Energy and Combustion Science, 2008. **34**(2): p. 198-
659 223. <https://doi.org/10.1016/j.pecs.2007.07.001>.
- 660 16. de Moraes, A.M., S. de Moraes Hanriot, A. de Oliveira, M.A.M. Justino, O.S. Valente
661 and J.R. Sodr , *An assessment of fuel consumption and emissions from a diesel power*
662 *generator converted to operate with ethanol*. Sustainable Energy Technologies and
663 Assessments, 2019. **35**: p. 291-297. <https://doi.org/10.1016/j.seta.2019.08.005>.
- 664 17. Singh, G., A.P. Singh and A.K. Agarwal, *Experimental investigations of combustion,*
665 *performance and emission characterization of biodiesel fuelled HCCI engine using*
666 *external mixture formation technique*. Sustainable Energy Technologies and
667 Assessments, 2014. **6**: p. 116-128. <https://doi.org/10.1016/j.seta.2014.01.002>.
- 668 18. Giakoumis, E.G., C.D. Rakopoulos, A.M. Dimaratos and D.C. Rakopoulos, *Exhaust*
669 *emissions of diesel engines operating under transient conditions with biodiesel fuel*
670 *blends*. Progress in Energy and Combustion Science, 2012. **38**(5): p. 691-715.
671 <https://doi.org/10.1016/j.pecs.2012.05.002>.
- 672 19. Datta, A. and B.K. Mandal, *A comprehensive review of biodiesel as an alternative fuel*
673 *for compression ignition engine*. Renewable and Sustainable Energy Reviews, 2016.
674 **57**: p. 799-821. <https://doi.org/10.1016/j.rser.2015.12.170>.
- 675 20. Nabi, N., A. Zare, M. Hossain, M.M. Rahman, D. Stuart, Z. Ristovski and R. Brown.
676 *Formulation of new oxygenated fuels and their influence on engine performance and*
677 *exhaust emissions*. The Combustion Institute Australia and New Zealand Section.

- 678 21. Verma, P., S. Stevanovic, A. Zare, G. Dwivedi, T. Chu Van, M. Davidson, T. Rainey, R.J.
679 Brown and Z.D. Ristovski, *An overview of the influence of biodiesel, alcohols, and*
680 *various oxygenated additives on the particulate matter emissions from diesel*
681 *engines*. *Energies*, 2019. **12**(10): p. 1987
- 682 22. World Health Organization, *Review of evidence on health aspects of air pollution–*
683 *REVIHAAP Project*. 2013
- 684 23. Vaughan, A., S. Stevanovic, A.P. Banks, A. Zare, M.M. Rahman, R.V. Bowman, K.M.
685 Fong, Z.D. Ristovski and I.A. Yang, *The cytotoxic, inflammatory and oxidative*
686 *potential of coconut oil-substituted diesel emissions on bronchial epithelial cells at an*
687 *air-liquid interface*. *Environmental Science Pollution Research*, 2019. **26**(27): p.
688 27783-27791
- 689 24. Vaughan, A., S. Stevanovic, M. Rahman, A. Zare, B. Miljevic, Z. Ristovski, R. Bowman,
690 K. Fong and I. Yang, *N-acetyl cysteine (NAC) intervention attenuates the effects of*
691 *diesel and biodiesel emission exposure on human bronchial epithelial cells, 16HBE, at*
692 *air-liquid interface*. *European Respiratory Journal*, 2016
- 693 25. Stevanovic, S., A. Vaughan, F. Hedayat, F. Salimi, M.M. Rahman, A. Zare, R.A. Brown,
694 R.J. Brown, H. Wang and Z. Zhang, *Oxidative potential of gas phase combustion*
695 *emissions-An underestimated and potentially harmful component of air pollution*
696 *from combustion processes*. *Atmospheric environment*, 2017. **158**: p. 227-235
- 697 26. Vaughan, A., S. Stevanovic, A. Zare, M. Rahman, B. Miljevic, Z. Ristovski, R. Bowman,
698 K. Fong and I. Yang, *Coconut oil substitution in diesel fuel alters human bronchial*
699 *epithelial cell responses to diesel emission exposure at the air-liquid interface in vitro:*
700 *to O22*. *Respirology*, 2016. **21**
- 701 27. Vaughan, A., S. Stevanovic, L.E. Morrison, A.M. Pourkhesalian, M.M. Rahman, A.
702 Zare, B. Miljevic, F. Goh, V. Relan and R. Bowman, *Removal of organic content from*
703 *diesel exhaust particles alters cellular responses of primary human bronchial*
704 *epithelial cells cultured at an air-liquid interface*. *Journal of Environmental Analytical*
705 *Toxicology*, 2015. **5**(5): p. 100316-1
- 706 28. Kittelson, D.B., *Engines and nanoparticles: a review*. *Journal of aerosol science*, 1998.
707 **29**(5-6): p. 575–588
- 708 29. Delphi Technologies, *Worldwide emissions standards-Passenger cars and light duty*
709 *vehicles*. Delphi, 2019. <https://www.delphi.com/innovations>.
- 710 30. Chen, L., S. Ding, H. Liu, Y. Lu, Y. Li and A.P. Roskilly, *Comparative study of*
711 *combustion and emissions of kerosene (RP-3), kerosene-pentanol blends and diesel in*
712 *a compression ignition engine*. *Applied Energy*, 2017. **203**: p. 91-100
- 713 31. Zare, A., T.A. Bodisco, M.N. Nabi, F.M. Hossain, M.M. Rahman, Z.D. Ristovski and R.J.
714 Brown, *The influence of oxygenated fuels on transient and steady-state engine*
715 *emissions*. *Energy*, 2017. **121**: p. 841-853.
716 <https://doi.org/10.1016/j.energy.2017.01.058>.
- 717 32. Chen, L., X. Hu, J. Wang and Y. Yu, *Impacts of alternative fuels on morphological and*
718 *nanostructural characteristics of soot emissions from an aviation piston engine*.
719 *Environmental science & technology*, 2019. **53**(8): p. 4667-4674
- 720 33. Krahl, J., J. Büniger, O. Schröder, A. Munack and G. Knothe, *Exhaust emissions and*
721 *health effects of particulate matter from agricultural tractors operating on rapeseed*
722 *oil methyl ester*. *Journal of the American Oil Chemists' Society*, 2002. **79**(7): p. 717-
723 724

- 724 34. Giechaskiel, B., T. Lähde, S. Gandi, S. Keller, P. Kreutziger and A. Mamakos,
725 *Assessment of 10-nm Particle Number (PN) Portable Emissions Measurement*
726 *Systems (PEMS) for Future Regulations*. International Journal of Environmental
727 Research and Public Health, 2020. **17**(11). 10.3390/ijerph17113878.
- 728 35. Giechaskiel, B., *Effect of Sampling Conditions on the Sub-23 nm Nonvolatile Particle*
729 *Emissions Measurements of a Moped*. Applied Sciences, 2019. **9**(15).
730 10.3390/app9153112.
- 731 36. Giechaskiel, B., T. Lähde, A.D. Melas, V. Valverde and M. Clairotte, *Uncertainty of*
732 *laboratory and portable solid particle number systems for regulatory measurements*
733 *of vehicle emissions*. Environmental Research, 2021. **197**: p. 111068.
734 <https://doi.org/10.1016/j.envres.2021.111068>.
- 735 37. Andersson, J., A. Mamakos, Z.C. Samaras, Z. Toumasatos, A. Kontses, L.D.
736 Ntziachristos, A. Bergmann, S. Hausberger, L. Landl, M. Bainschab, J. Keskinen and C.
737 Haisch, *Measuring Automotive Exhaust Particles Down to 10 nm*. SAE International
738 Journal of Advances and Current Practices in Mobility, 2020. **3**(1): p. 539-550.
739 <https://doi.org/10.4271/2020-01-2209>.
- 740 38. Leach, F., A. Lewis, S. Akehurst, J. Turner and D. Richardson, *Sub-23 nm particulate*
741 *emissions from a highly boosted GDI engine*. SAE Technical Paper, 2019. 0148-7191.
- 742 39. Lodi, F., A. Zare, P. Arora, S. Stevanovic, P. Verma, M. Jafari, Z. Ristovski, R. Brown
743 and T. Bodisco, *Characteristics of Particle Number and Particle Mass Emissions of a*
744 *Diesel Engine during Cold-, Warm-, and Hot-Start Operation*. SAE Technical Paper,
745 2021. <https://doi.org/10.4271/2021-01-5061>.
- 746 40. Reiter, M.S. and K.M. Kockelman, *The problem of cold starts: A closer look at mobile*
747 *source emissions levels*. Transportation Research Part D: Transport and Environment,
748 2016. **43**: p. 123-132
- 749 41. André, M., *In actual use car testing: 70,000 kilometers and 10,000 trips by 55 French*
750 *cars under real conditions*. SAE transactions, 1991: p. 65-72
- 751 42. Zare, A., S. Stevanovic, M. Jafari, P. Verma, M. Babaie, L. Yang, M.M. Rahman, Z.D.
752 Ristovski, R.J. Brown and T.A. Bodisco, *Analysis of cold-start NO₂ and NO_x emissions,*
753 *and the NO₂/NO_x ratio in a diesel engine powered with different diesel-biodiesel*
754 *blends*. Environmental Pollution, 2021. **290**: p. 118052.
755 <https://doi.org/10.1016/j.envpol.2021.118052>.
- 756 43. Shi, Z., C.-f. Lee, H. Wu, Y. Wu, L. Zhang and F. Liu, *Optical diagnostics of low-*
757 *temperature ignition and combustion characteristics of diesel/kerosene blends under*
758 *cold-start conditions*. Applied Energy, 2019. **251**: p. 113307.
759 <https://doi.org/10.1016/j.apenergy.2019.113307>.
- 760 44. Lodi, F., A. Zare, P. Arora, S. Stevanovic, M. Jafari, Z. Ristovski, R.J. Brown and T.
761 Bodisco, *Combustion Analysis of a Diesel Engine during Warm up at Different Coolant*
762 *and Lubricating Oil Temperatures*. Energies, 2020. **13**(15): p. 3931
- 763 45. Verma, P., M. Jafari, A. Zare, E. Pickering, Y. Guo, C.G. Osuagwu, S. Stevanovic, R.
764 Brown and Z. Ristovski, *Soot particle morphology and nanostructure with oxygenated*
765 *fuels: A comparative study into cold-start and hot-start operation*. Environmental
766 Pollution, 2021. **275**: p. 116592. <https://doi.org/10.1016/j.envpol.2021.116592>.
- 767 46. Zare, A., T.A. Bodisco, M.N. Nabi, F.M. Hossain, Z.D. Ristovski and R.J. Brown, *A*
768 *comparative investigation into cold-start and hot-start operation of diesel engine*
769 *performance with oxygenated fuels during transient and steady-state operation*.
770 Fuel, 2018. **228**: p. 390-404. <https://doi.org/10.1016/j.fuel.2018.05.004>.

- 771 47. Zare, A., M.N. Nabi, T.A. Bodisco, F.M. Hossain, M.M. Rahman, T. Chu Van, Z.D.
772 Ristovski and R.J. Brown, *Diesel engine emissions with oxygenated fuels: A*
773 *comparative study into cold-start and hot-start operation*. Journal of Cleaner
774 Production, 2017. **162**: p. 997-1008. <https://doi.org/10.1016/j.jclepro.2017.06.052>.
- 775 48. Lodi, F., A. Zare, P. Arora, S. Stevanovic, M. Jafari, Z. Ristovski, R.J. Brown and T.
776 Bodisco, *Engine Performance and Emissions Analysis in a Cold, Intermediate and Hot*
777 *Start Diesel Engine*. Applied Sciences, 2020. **10**(11): p. 3839
- 778 49. Nam, E., *Analysis of particulate matter emissions from light-duty gasoline vehicles in*
779 *Kansas City*. 2008: US Environmental Protection Agency.
- 780 50. Lee, D.-W., J. Johnson, J. Lv, K. Novak and J. Zietsman, *Comparisons between*
781 *vehicular emissions from real-world in-use testing and EPA moves estimation*. Texas
782 Transportation Institute, 2012.
- 783 51. Roberts, A., R. Brooks and P. Shipway, *Internal combustion engine cold-start*
784 *efficiency: A review of the problem, causes and potential solutions*. Energy
785 Conversion and Management, 2014. **82**: p. 327-350
- 786 52. Zare, A., T.A. Bodisco, P. Verma, M. Jafari, M. Babaie, L. Yang, M.M. Rahman, A.
787 Banks, Z.D. Ristovski and R.J. Brown, *Emissions and performance with diesel and*
788 *waste lubricating oil: A fundamental study into cold start operation with a special*
789 *focus on particle number size distribution*. Energy Conversion and Management,
790 2020. **209**: p. 112604
- 791 53. Bodisco, T. and R.J. Brown, *Inter-cycle variability of in-cylinder pressure parameters*
792 *in an ethanol fumigated common rail diesel engine*. Energy, 2013. **52**: p. 55-65
- 793 54. Bodisco, T., P. Tröndle and R.J. Brown, *Inter-cycle variability of ignition delay in an*
794 *ethanol fumigated common rail diesel engine*. Energy, 2015. **84**: p. 186-195
- 795 55. Su, S., T. Lv, Y. Lai, J. Mu, Y. Ge and B. Giechaskiel, *Particulate emissions of heavy*
796 *duty diesel engines measured from the tailpipe and the dilution tunnel*. Journal of
797 Aerosol Science, 2021. **156**: p. 105799.
798 <https://doi.org/10.1016/j.jaerosci.2021.105799>.
- 799 56. Giechaskiel, B., A.D. Melas, T. Lähde and G. Martini, *Non-Volatile Particle Number*
800 *Emission Measurements with Catalytic Strippers: A Review*. Vehicles, 2020. **2**(2).
801 10.3390/vehicles2020019.
- 802 57. TSI, *Aerosol Instrument Manager® Software for Scanning Mobility Particle Sizer™*
803 *(SMPS™) Spectrometer*. 2010, TSI. p. 71.
- 804 58. Giakoumis, E.G., C.D. Rakopoulos and D.C. Rakopoulos, *Assessment of NOx Emissions*
805 *during Transient Diesel Engine Operation with Biodiesel Blends*. Journal of Energy
806 Engineering 2014. **140**(3): p. A4014004. doi:10.1061/(ASCE)EY.1943-7897.0000136.
- 807 59. Van, T.C., A. Zare, M. Jafari, T.A. Bodisco, N. Surawski, P. Verma, K. Suara, Z.
808 Ristovski, T. Rainey and S. Stevanovic, *Effect of cold start on engine performance and*
809 *emissions from diesel engines using IMO-Compliant distillate fuels*. Environmental
810 Pollution, 2019. **255**: p. 113260
- 811 60. Mitchell, B.J., A. Zare, T.A. Bodisco, M.N. Nabi, F.M. Hossain, Z.D. Ristovski and R.J.
812 Brown, *Engine blow-by with oxygenated fuels: A comparative study into cold and hot*
813 *start operation*. Energy, 2017. **140**: p. 612-624
- 814 61. Zare, A., T.A. Bodisco, M. Jafari, P. Verma, L. Yang, M. Babaie, M.M. Rahman, A.
815 Banks, Z.D. Ristovski, R.J. Brown and S. Stevanovic, *Cold-start NOx emissions: Diesel*
816 *and waste lubricating oil as a fuel additive*. Fuel, 2021. **286**: p. 119430.
817 <https://doi.org/10.1016/j.fuel.2020.119430>.

- 818 62. Chien, S.-M., Y.-J. Huang, S.-C. Chuang and H.-H. Yang, *Effects of biodiesel blending*
819 *on particulate and polycyclic aromatic hydrocarbon emissions in*
820 *nano/ultrafine/fine/coarse ranges from diesel engine*. *Aerosol and Air Quality*
821 *Research*, 2016. **9**(1): p. 18-31
- 822 63. Jafari, M., P. Verma, T.A. Bodisco, A. Zare, N.C. Surawski, P. Borghesani, S.
823 Stevanovic, Y. Guo, J. Alroe, C. Osuagwu, A. Milic, B. Miljevic, Z.D. Ristovski and R.J.
824 Brown, *Multivariate analysis of performance and emission parameters in a diesel*
825 *engine using biodiesel and oxygenated additive*. *Energy Conversion and*
826 *Management*, 2019. **201**: p. 112183.
827 <https://doi.org/10.1016/j.enconman.2019.112183>.
- 828 64. Yusuf, A.A. and F.L. Inambao, *Effect of cold start emissions from gasoline-fueled*
829 *engines of light-duty vehicles at low and high ambient temperatures: Recent trends*.
830 *Case Studies in Thermal Engineering*, 2019. **14**: p. 100417.
831 <https://doi.org/10.1016/j.csite.2019.100417>.
- 832 65. Kittelson, D.B., W.F. Watts, J.P. Johnson, C. Thorne, C. Higham, M. Payne, S. Goodier,
833 C. Warrens, H. Preston and U. Zink, *Effect of fuel and lube oil sulfur on the*
834 *performance of a diesel exhaust gas continuously regenerating trap*. *Environmental*
835 *science & technology*, 2008. **42**(24): p. 9276-9282
- 836 66. Ning, Z., C.S. Cheung and S.X. Liu, *Experimental investigation of the effect of exhaust*
837 *gas cooling on diesel particulate*. *Journal of Aerosol Science*, 2004. **35**(3): p. 333-345
- 838 67. Kasper, M., *Sampling and measurement of nanoparticle emissions for type approval*
839 *and field control*. SAE Technical Paper, 2005. 0148-7191.
- 840 68. Tsai, J.-H., S.-Y. Chang and H.-L. Chiang, *Volatile organic compounds from the exhaust*
841 *of light-duty diesel vehicles*. *Atmospheric environment*, 2012. **61**: p. 499-506
- 842 69. Pourkhesalian, A.M., S. Stevanovic, F. Salimi, M.M. Rahman, H. Wang, P.X. Pham, S.E.
843 Bottle, A.R. Masri, R.J. Brown and Z.D. Ristovski, *Influence of fuel molecular structure*
844 *on the volatility and oxidative potential of biodiesel particulate matter*.
845 *Environmental science & technology*, 2014. **48**(21): p. 12577-12585
- 846 70. Armas, O., J.J. Hernández and M.D. Cárdenas, *Reduction of diesel smoke opacity from*
847 *vegetable oil methyl esters during transient operation*. *Fuel*, 2006. **85**(17-18): p.
848 2427-2438
- 849 71. Hedayat, F., S. Stevanovic, A. Milic, B. Miljevic, M.N. Nabi, A. Zare, S.E. Bottle, R.J.
850 Brown and Z.D. Ristovski, *Influence of oxygen content of the certain types of*
851 *biodiesels on particulate oxidative potential*. *Science of The Total Environment*, 2016.
852 **545-546**: p. 381-388. <https://doi.org/10.1016/j.scitotenv.2015.12.036>.

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857 **7. Appendix A**

858 Table A1 shows the accuracy of the measuring instruments.

859 Table A1 Instrument accuracy

Instrument	Accuracy
Kistler 6053CC60 piezoelectric transducer	≈ -20 pC/bar
Kistler type 2614	0.5 crank angle degrees
CAI-600 NDIR CO ₂ analyser	Linearity > 0.5% and repeatability > 1% of full scale
SABLE CA-10 Carbon CO ₂ gas analyser	1% of reading within the range of 0-10%
Dynamometer	±0.5%

860

861 Table A2 shows the statistical analysis of the test repeatability for cold-start and warm-start
862 tests with diesel using average and coefficient of variation (CoV) of engine speed and torque.

863 As can be seen, the difference between the tests is indicating the repeatability of the
864 experiment. In addition to these two important parameters, Table A3 shows the CO₂
865 repeatability. This experimental study used a non-dispersive infrared CAI-600 CO₂ gas
866 analyser which is a piece of high-tech quality equipment in the market used by different car
867 industries and research groups to check the CO₂ emission and confirm the test repeatability.

868 It is worth to mention that using a correlation car/engine and measuring CO₂ in the repeated
869 tests and then calculating the variation between the correlation tests is a very common and
870 trustworthy method of uncertainty measurement and test repeatability check in emissions
871 laboratories of the automotive industries.

872

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874

875 Table A2 Test repeatability: statistical analysis

		Engine speed (rpm)		Engine torque (Nm)	
		Mean	CoV (%)	Mean	CoV (%)
Warm-start	Test I	1499.49	0.14	242.02	1.00
	Test II	1498.94	0.15	238.28	1.34
	Difference	0.04%		1.5%	
Cold-start	Test I	1499.19	0.13	227.20	5.42
	Test II	1498.87	0.15	225.28	3.74
	Difference	0.02%		0.82%	

876

877 Table A3 Test repeatability: CO₂ correlation test

		Mean	CoV (%)
Warm-start	Test I	6.64	0.36
	Test II	6.47	0.51
	Difference	0.17%	
Cold-start	Test I	6.51	2.27
	Test II	6.36	0.97
	Difference	0.12%	

878