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1 **Green Deep Eutectic Solvents (DESs) for Sustainable Metal Recovery from Thermally**
2 **Treated PCBs: A Greener Alternative to Conventional Methods**

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10 **KEYWORDS:** Printed Circuit Boards (PCBs); Deep Eutectic Solvent (DES); Metal
11 extraction; Kinetics; DFT calculations.

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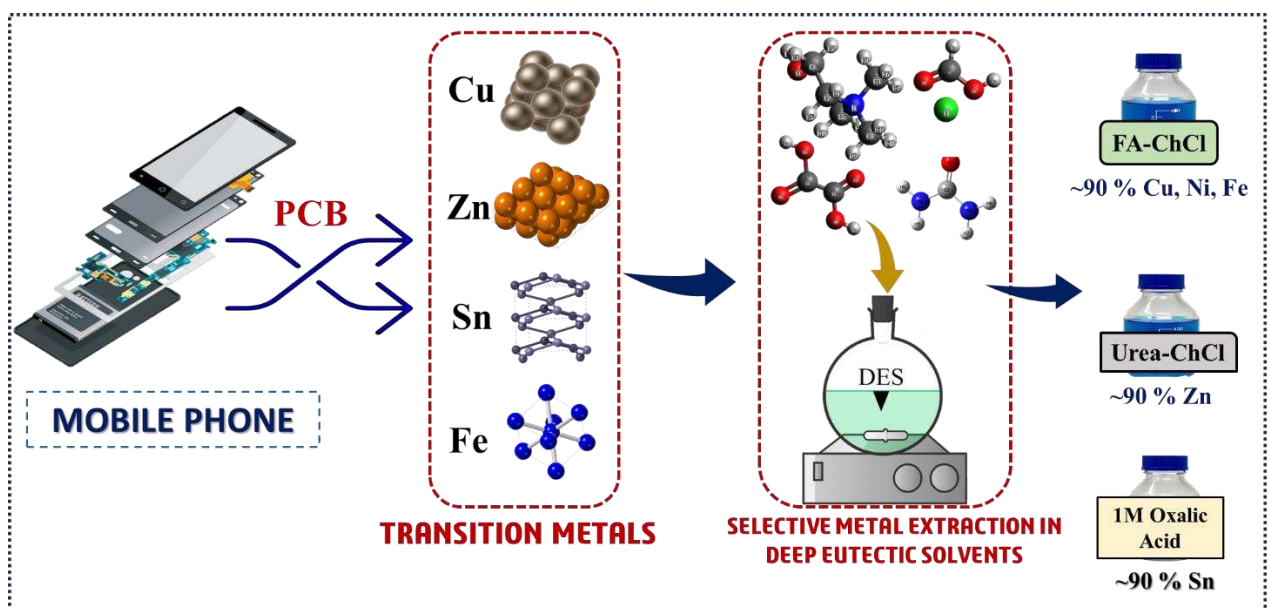
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24 **ABSTRACT**

25 Waste PCBs the core of e-waste is rich in copper, tin, zinc, iron, and nickel. Leaching base
26 metals from PCB used to be done in toxic, corrosive acidic/alkali mediums. In this work, an
27 environmentally friendly method for leaching metals from thermally treated PCBs (TPCBs)
28 of mobile phones was proposed using choline chloride based deep eutectic solvents (DES).
29 DES selectivity and solubility of metals from metal oxides were the main screening criteria.
30 FA-ChCl had the maximum solubility of Cu, Fe, and Ni, while Urea-ChCl had high Zn
31 selectivity and solubility. Oxalic acid has high selectivity for Sn. FA-ChCl extracted Cu and
32 Fe best at 16 h, 100 °C, and 1/30 g/mL. Urea-ChCl extracted Zn ($90.4 \pm 2.9\%$) from TPCBs
33 at 100 °C, 21 h, 1/20 g/mL, and 400 rpm. Oxalic acid (1M) removed $92.3 \pm 2.1\%$ Sn from
34 TPCBs in 1 h at 80 °C and 1/20 g/mL. The shrinking core model-based kinetic investigation
35 of FA-ChCl for Cu extraction showed a diffusion-controlled process. The proposed method is
36 greener than mineral acids utilized for metal extraction.

37 **GRAPHICAL ABSTRACT**

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42 **1. Introduction**

43 Electronic waste (e-waste) is a vastly underutilized resource that offers a plethora of
44 economic opportunities. E-waste contains a wealth of precious metals (platinum, gold,
45 silver), base metals (aluminum, iron, copper), heavy (zinc, mercury, chromium, lead), and
46 rare earth metals (tantalum, platinum groups) ^[1]. In 2019, the estimated value of treasured
47 metals in e-waste such as gold, copper, iron, etc., inclusive of plastics obtained from e-waste,
48 is assessed to be equivalent to USD 57 billion ^[2]. According to the UN World Economic
49 Forum in 2019, the gold present in a tonne of e-waste is 100 times more valuable than 1
50 tonne of gold ores ^[2]. Likewise, a metric ton of printed circuit board (PCB) contains 30 to 40
51 times higher copper than one metric ton of natural copper ore ^[3]. The use of e-waste as a rich
52 secondary source of metals can significantly bridge the huge gap between demand and supply
53 of several scarce metals, e.g., tin, silver, gold, platinum, lithium, etc. The Global E-waste
54 Monitor Report 2020 revealed that ca. 53.5 million metric tons (MMT) of e-waste was
55 generated globally in 2019 ^[2]. The report also predicts that an astounding 74 MMT of e-waste
56 will be generated by the year 2030, nearly double the amount generated 16 years ago ^[2].
57 Nevertheless, the recycling rate of e-waste is quite low and mostly concentrated in
58 unregulated informal sectors, often involving open dumping or burning ^[4]. Metal extraction
59 from e-waste presents opportunities in the context of a circular economy driven by the need
60 to recycle valuable metals from discarded electronics and reduce environmental impacts from
61 traditional virgin mining. Printed circuit boards (PCBs) are one of the vital components of
62 any electronic gadget, constituting 3 % of the total collected electrical and electronic waste
63 globally ^[5]. Recycling PCBs is necessary for economic and environmental reasons, as they
64 are rich in valuable metals such as copper, iron, gold, silver, and hazardous metals such as
65 lead, chromium, etc. PCBs are often referred to as an “Urban Mineral Source” due to their
66 metal concentration being 10 times richer than that of ores ^[6].

67 The development of a metal recycling process from PCBs is necessary for the electronics
68 industry to thrive sustainably. Conventional methods, such as electrostatic separation [7], froth
69 floatation [8], and vibration gas-solid fluidized bed [9], have been used to separate distinct non-
70 metallic and metallic fractions. Furthermore, to recover metals from e-waste, pyrometallurgy,
71 hydrometallurgy, and biohydrometallurgy have been employed [3]. Pyrometallurgy operates
72 at very high temperatures and produces hazardous gases like dioxins due to the presence of
73 brominated flame retardants, (hydro)chlorofluorocarbons, and mercury in PCBs [10].
74 Hydrometallurgy, a flexible alternative approach for recovering metals, uses alkaline [11],
75 cyanide [12], thiosulphate [13], mineral acid [14], etc., as leaching agents to extract metals such
76 as copper, zinc and gold. However, poor selectivity, toxicity [15], and severe environmental
77 threats have led to an exploration of greener alternatives to conventional metal extraction
78 techniques [16].

79 The replacement of toxic reagents with green solvents, such as ionic liquids and
80 supercritical fluids, has been proposed by several researchers [17]. Xiu et al. used supercritical
81 water with HCl leaching for 100% Cu recovery and iodine-iodine leaching to recover silver,
82 gold, and palladium from PCBs [18]. Golazary et al. combined the supercritical water process
83 as pre-treatment followed by supercritical CO₂ to recover 97 % copper from PCBs at 60 °C
84 and 200 bar [17a]. However, high pressure to attain the supercritical state is a concern in these
85 technologies. Ionic liquids are promising and environmentally sound alternatives to organic
86 solvents. Huang et al. used [bmim]HSO₄ ionic liquid with hydrogen peroxide to extract
87 copper from PCBs [19]. Barrueto et al. observed that acidic ionic liquid ([Bmim]HSO₄ and
88 [Hmim]HSO₄) extracted 86.2 % and 76.6 % copper, respectively, while basic ionic liquid
89 ([Bmim]Cl) extracted 47.3 % silver and 19.1 % gold [20]. Nevertheless, the exploration of
90 safer, cheaper, and tunable solvents for sustainable chemicals has received much attention in
91 recent years.

92 Deep eutectic solvents, composed of naturally occurring compounds, have high extraction
93 efficiency for metals and low toxicity with a cutting edge over ionic liquids in terms of easy
94 preparation method, less corrosive, and relatively easy biodegradation. The advantage of
95 DESs lies in the difference in the coordination ability of components and metals, providing a
96 low-cost green strategy with the potential to selectively extract metals on a large scale from
97 complex ores or waste material. Malonic acid-choline chloride (MA-ChCl), Oxalic acid-
98 choline chloride (OA-ChCl), and Ethylene glycol-choline chloride (EG-ChCl) DESs were
99 checked by Abbott et al. for transition metal oxide dissolution [24]. Selective dissolution of
100 metals was observed in choline chloride (ChCl) DES. Since then, the development of DES
101 synthesis has taken precedence in the sector as a means of effective metal extraction from
102 complicated compounds under benign conditions [25]. Choline chloride-based DESs were
103 reported to have highly efficient metal leaching from Li-ion batteries [26-27], NdFeB magnets
104 [26], ores [27], minerals [28], and electric arc furnace dust [29].
105 Recoveries of metals from spent lithium-ion batteries are extensively studied [30]. Li et al.
106 found DESs to be a greener alternative to inorganic acids with metal leaching efficiency (>99
107 %) but with slower leaching kinetics [30]. Wang et al. concluded that the high reducing
108 power of Urea-ChCl DES makes it the best DESs for cathode recycling of Li-ion batteries
109 [31]. Tran et al. observed that ethylene glycol-ChCl DESs recover 99.3 % of cobalt and
110 lithium from lithium cobalt oxide, and metal leaching efficiency increases with time and
111 temperature [30]. The use of green solvents for metal extraction from e-waste is a promising
112 area of research with the potential to reduce the environmental impacts of traditional mining
113 processes. Lactic acid-ChCl [32], malonic acid-ChCl [32], and ethylene glycol-ChCl [33] have
114 been used to extract metals from WPCBs, covering limited aspects. However, no detailed
115 study of formic acid-choline chloride and urea-choline chloride in metal extraction from
116 WPCBs has been done in the past to the best of our knowledge. Considering this research gap

117 and the Choline chloride DES advantage of selective metal extraction, the present study is
118 proposed.

119 The present study uses green deep eutectic solvents and organic acid for the selective
120 recovery of Cu, Fe, Ni, Zn and Sn from PCBs. Screening of green solvents was done in the
121 first stage by dissolving various metal oxides [25b, 34]. Further, based on the solubility of
122 metals in respective DESs, each selective solvent was chosen for optimizing parameters in
123 metal extraction. FA-ChCl showed the highest solubility of Cu, Fe, and Ni, while Urea-ChCl
124 exhibited selective solubility of Zn. For Sn extraction, 1M oxalic acid demonstrated the
125 highest efficiency. TPCBs of mobile phones with high metal concentrations have been used
126 as feed materials for metal leaching experiments in the screened solvents. The influence of
127 several leaching parameters such as time, temperature, and S/L ratio, in base metal leaching
128 in DESs has been discussed in detail. Since copper has the highest contribution to the metallic
129 composition of PCBs, the shrinking core model leaching kinetics of the highly extracted
130 copper using FA-ChCl DES, and the activation energy of copper were assessed. A plausible
131 mechanism was proposed based on literature, DFT and Pourbaix diagram.

132 **2. Materials and Methods**

133 **2.1. Chemicals and Reagents**

134 Waste PCBs of miscellaneous brands used in this study were provided by Exigo Recycling
135 Pvt. Ltd, Haryana, India. The PCBs were crushed and ground to a size below 1.0 mm for the
136 experiment. For the digestion of solid samples, nitric acid (86 %, HNO₃) and hydrochloric
137 acid (37 %, HCl) from Fisher Scientific were used to prepare the aqua regia solution.

138 For DES synthesis, choline chloride (> 98 %) was purchased from Alfa Aesar, formic acid
139 (99 %), urea (> 98 %), ethylene glycol (> 95 %), malonic acid (99 %), and oxalic acid (99.5
140 %) were obtained from Fisher Scientific. Metal oxides, including copper oxide (> 99 %,
141 CuO), iron oxide (> 96 %, Fe₂O₃), zinc oxide (> 97 %, ZnO), and nickel oxide (≥ 99 %, NiO)
142 were purchased from Sigma Aldrich. For dilution, double-distilled water with a conductivity

143 of less than 1 $\mu\text{S}/\text{cm}$ was employed. All the reagents used in this experiment are of analytical
144 grade and were used as received.

145 **2.2. Experimental**

146 Waste PCB samples were comminuted and screened to obtain a powder with a particle size of
147 less than 1 mm. Particles smaller than 1 mm in size were separated using an ASTM E11 # 18
148 mesh screen (American Society for Testing and Materials). The organic content was removed
149 from the PCBs using thermal pre-treatment, which involved pyrolyzing them in an inert
150 environment. Pyrolysis was conducted in a fixed bed reactor at 500 °C for 30 min in a
151 nitrogen atmosphere, as these conditions are sufficient for the cracking of the plastic part of
152 PCBs, as discussed by Marco et al. [35]. The extraction efficiencies of Cu, Zn, and Sn from
153 TPCBs were calculated using the following formula (1).

$$154 \text{ Extraction efficiency } (E \%) = \frac{C_1 \times V_1}{W_{m\text{TPCB}}} \times 100 \quad (1)$$

155 Where C_1 is the metal content in leached solution (g/L); V_1 is the leach volume (L), and
156 $W_{m\text{TPCB}}$ is the weight of the metal in TPCBs (g).

157 **2.2.1. Synthesis of Deep Eutectic Solvents**

158 The DESs were prepared via the heating method. Hydrogen bond acceptor (HBA), ChCl , was
159 combined with hydrogen bond donors (HBD), i.e., FA, urea, ethylene glycol, malonic acid,
160 and oxalic acid in a (2:1) molar ratio in sealed 50 mL glass vials while being continuously
161 stirred at 200 rpm at 50°C to create a eutectic mixture. However, mole ratios of individual
162 components in OA- ChCl and MA- ChCl DES were 1:1 [25a]. The solvents were then stored at
163 room temperature in a desiccator to avoid moisture absorption.

164 **2.2.2. Metal leaching**

165 All leaching/ extraction experiments were carried out in 50 mL Borosil glass on a hot
166 magnetic stirrer (Cole-Parmer Stuart) with an oil bath. Primarily, for screening DESs, 0.1 g of

167 respective metal oxides: CuO, Fe₂O₃, ZnO, and NiO were dissolved in 10 mL EG-ChCl,
168 Urea-ChCl, FA-ChCl, OA-ChCl, and MA-ChCl for 24 h at 80 °C. The temperature of 80 °C
169 is selected as Abbott et al. [34] observed high solubility of zinc in urea-ChCl at a temperature
170 greater than 60 °C. Respective DES were selected for individual metal extraction. 1 M oxalic
171 acid was also tested to check the solubility of copper, iron, zinc, and nickel from oxides.

172 A parameter selectivity (S_m) was defined to evaluate the effectiveness of different leaching
173 solutions. It is calculated by dividing the concentration of the metal of interest (m) by the sum
174 of concentrations of all leached metals in the solution by equation (2).

$$175 \quad S_m = C_m / \sum_i C_i \times 100 \% \quad (2)$$

176 Here, C_m is the concentration of metal of interest in leaching solution and C_i is the
177 concentration of any metals in a leached solution.

178 A series of leaching optimization experiments were conducted to find the maximum copper,
179 zinc, and tin leaching efficiency in their respective highest soluble solvents. Filter paper with
180 a 0.22 µm pore size was used to separate the solid residue from the leach liquor. The residue
181 was washed with ultrapure water and dried for 24 h at 100 °C. For metal extraction from e-
182 waste, 1 g of TPCB sample was taken in a given volume of FA-ChCl, Urea-ChCl, and 1M
183 oxalic acid solution. The effect of time (1-32 h), temperature (20-100 °C), and S/L ratio
184 (1/10-1/40 mg/L) were evaluated for Cu, Fe, Zn, Ni, and Sn in FA-ChCl. In Urea-ChCl, the
185 temperature was varied from (20-140 °C), stirring speed (100-800 rpm), time (1- 24 h), and
186 S/L ratio (1/10-1/50 g/mL). The extraction of tin in oxalic acid time was evaluated at process
187 conditions with time (1-5 h); temperature (50-80 °C) and oxalic acid concentration 0-3 g/L.
188 High temperatures (>100 °C) were avoided considering the economical and energy point of
189 view [36]. Glass bottles were closed to prevent the loss of vapors.

190

191 **2.2.3. Individual recovery of copper**

192 The metal-leached solution obtained under optimized conditions was utilized for copper
193 recovery from FA-ChCl through the cementation process. The metallic zinc powder was
194 introduced into the leached solution to facilitate the recovery of metallic copper. The process
195 began by transferring 200 mL metal leached solution into a beaker followed by the gradual
196 addition of Zn powder. The appropriate stoichiometric amount of Zn was used to ensure
197 effective copper cementation. Once the cementation was completed, the precipitated copper
198 was recovered, dried, and subjected to analysis. The concentration of metal ions in the
199 solution is determined by ICP-MS and cementation efficiency (θ) was calculated by equation
200 (3):

$$201 \quad \theta = \left(1 - \frac{C}{C_0}\right) \times 100 \% \quad (3)$$

202 Here, θ is metal precipitation efficiency (%), C_0 is initial metal concentration in optimized
203 leaching solution, mg/L; and C is metal concentration in a solution after cementation, mg/L.

204 **2.3. Instrumentation and analysis**

205 A Nicolet iS50 FTIR spectrophotometer with an ATR crystal detector was used to record
206 infrared spectra. Samples were scanned 32 times with a resolution of 4 cm^{-1} in the range of
207 $400\text{-}4000 \text{ cm}^{-1}$. Background reference spectra were obtained before sample measurement to
208 avoid interferences. All samples were measured under atmospheric conditions.

209 The distribution of metal phases was investigated using X-ray diffraction (XRD) by Rigaku
210 Mini Flex 600 instrument from Japan. Cu-K radiations were utilized, with graphite serving as
211 a monochromator. The XRD measurements were conducted at 40 kV and 15 mA, employing
212 an X'Celerator detector. The scanning process involved a ramp rate of $5^\circ/\text{min}$, spanning an
213 angle range of $5\text{-}80^\circ$.

214 Netzsch STA449 Thermo Gravimetric Analyzer (TGA) was used to assess the thermal
215 degradation of DES. 10–20 mg of DES was placed in a nitrogen environment throughout a
216 temperature range of 25–800 °C.

217 The density of solvents was measured by Mettler Tolardo DE45 Delta Range Density meter.
218 The viscosity of DESs was measured by Anton Paar MCR302 rotational and oscillatory
219 rheometer equipped with cone and plate geometry (CP 50, diameter 50 mm, and angle 1°) at
220 303.15, 323.15 and 343.15 K temperatures at constant shear rate of 1000 s⁻¹.

221 Metal extraction in DESs was determined by Elan DRCe (Perkin Elmer) ICP-MS, diluting
222 the samples 1000 times with doubly distilled water. The solubility experiments and
223 measurements were done in triplicate.

224 **2.4. Computational Method**

225 A theoretical study of DES was done by optimizing ChCl, FA, urea, and oxalic acid.
226 Geometrical optimization of structures was done using Gaussian with B3LYP 631 g+ (d,p)
227 basis set [37]. The dispersion effect was avoided by adding the keyword “empirical
228 dispersion=D3” in the command section. Hydrogen bond distances and angles of DESs have
229 been calculated and quantified for H-bonds. The spatial distribution of electrostatic
230 interactions in the solvent system was determined by Electrostatic potential surface (ESP)
231 calculation. The interaction energy of copper with hydrogen bond donors of DES has been
232 assessed.

233 **3. Results and Discussion**

234 **3.1. Characterization of PCB and TPCB**

235 As a feedstock, PCBs are rich in critical metals such as Cu, Ag, Au and Al. Table 1
236 summarizes the elemental composition of the PCB used in the current study. Among these
237 metals, Cu is the most prevalent metal in the PCB constituting 29.6 wt%, (see Table 1). After

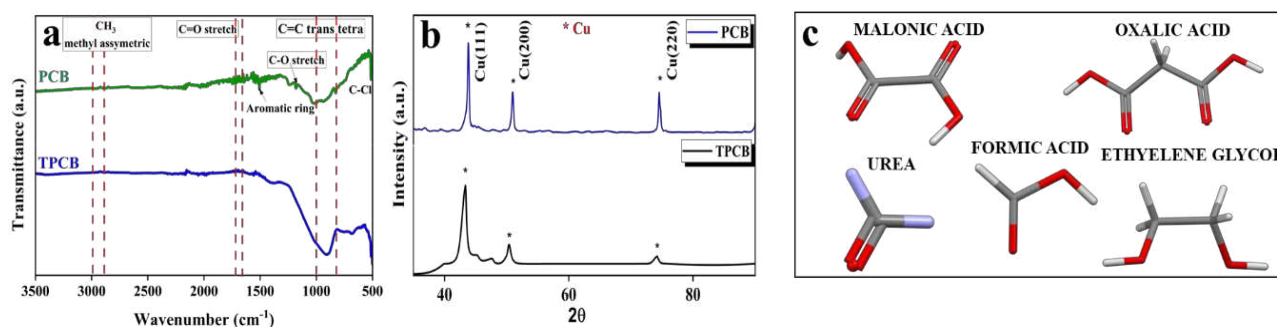
238 subjecting the PCB to the thermal treatment, the Cu content increased to 58.8 wt% due to the
 239 removal of the non-metallic fraction. The initial Fe and Zn contents in the PCB were 9.6 wt%
 240 and 0.2 wt%, respectively, and were enriched to 14.4 wt% and 0.5 wt% in the TPCBs.
 241 Additionally, the contents of Al, Ni and Sn ranged from 0.4-2 wt%, in the PCBs, and
 242 enriched to 0.6-8 wt% in the TPCBs [38]. The concentrations of transition metals in the PCBs
 243 are comparable to those previously reported [38].

244 Figure 1 compares the FTIR spectra and XRD patterns of the PCB and TPCB. The FTIR
 245 spectra (Fig. 1a) of the PCB show the presence of CH₃ methyl aromatics at 2967 cm⁻¹, C=O
 246 at 1738 cm⁻¹, benzene ring at 1610 cm⁻¹ and 1506 cm⁻¹. C-Cl and C-Br functional groups in
 247 the range of 500-750 cm⁻¹. In the case of TPCB, several characteristic peaks of PCB (C=O,
 248 halogens, C=C trans) disappeared, signifying the loss of the organic functional group after
 249 pyrolysis at 500 °C.

250 **Table 1.** Elemental composition of PCB and TPCB (w/w%).

	Cu	Fe	Al	Zn	Ni	Sn
PCB	29.63 ± 1.25	9.66 ± 0.20	0.43 ± 0.02	0.15 ± 0.01	1.26 ± 0.12	1.90 ± 0.14
TPCB	58.76 ± 2.01	14.45 ± 1.20	0.62 ± 0.03	0.52 ± 0.02	2.53 ± 0.13	7.88 ± 0.17

251
 252 Figure 1b compares XRD patterns for PCB and TPCB showing peaks at 44°, 51° and 74°,
 253 which correspond to metallic copper with Miller Indices (111), (002) and (022), respectively
 254 (JCPDS 00-004-0836). As such, the pyrolysis process does not affect the metal phase or its
 255 oxidation state. Figure 1c depicts the models of HBD involved in this study.



256

257 **Figure 1.** FTIR of PCB and TPCB (a); XRD of PCB and TPCB with clear peaks (*) at 44°,
 258 51° and 74°. (b); models of HBDs used in the DES (c).

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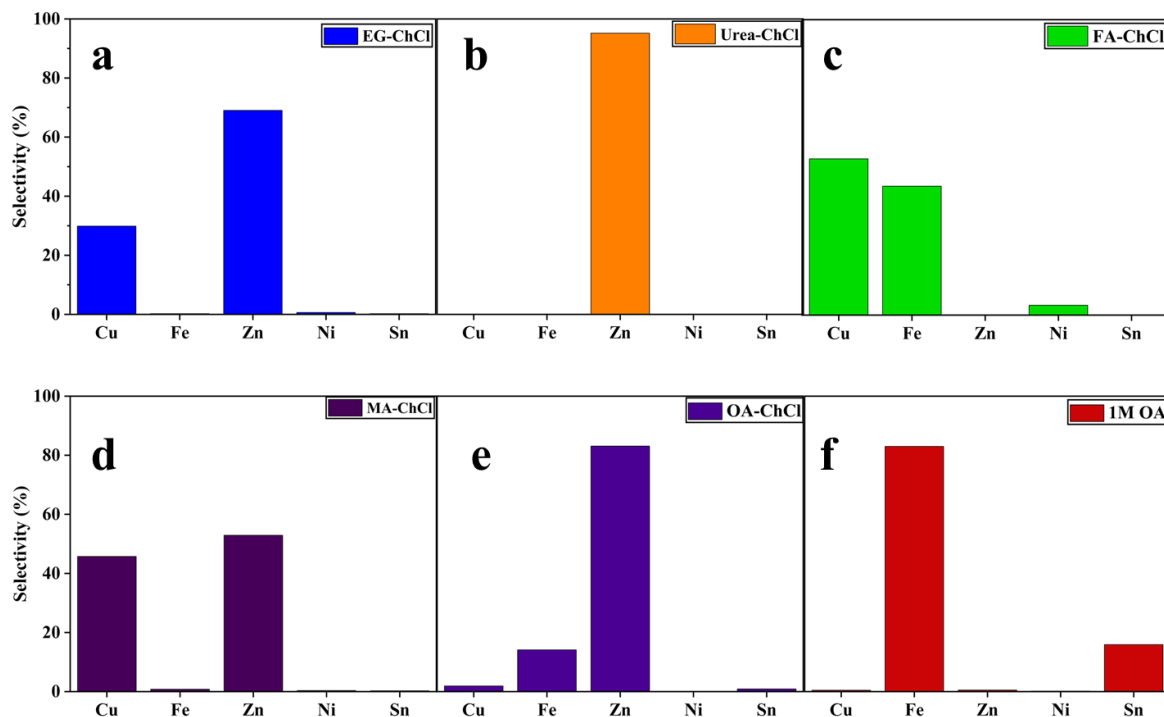
260 3.2. Screening of DES and strategies for metal extraction:

261 To determine the DES for maximum extraction of Cu, Ni, Fe, Zn and Sn, the respective metal
 262 oxides were tested for dissolution in a range of DES: EG-ChCl, Urea-ChCl, FA-ChCl, OA-
 263 ChCl, MA-ChCl, and 1M oxalic acid, as shown in Table S1.

264 The solubility of Cu in FA-ChCl was $1.60 \pm 0.06 (\times 10^5)$ ppm. However, this DES had the
 265 lowest solubility of Zn at $3.12 \pm 0.12 (\times 10^2)$ ppm, while OA-ChCl and MA-ChCl had much
 266 higher solubilities for Zn at $9.65 \pm 0.09 (\times 10^4)$ ppm and $1.62 \pm 0.05 (\times 10^4)$ ppm, respectively.
 267 For Cu, those DESs showed an intermediate solubility, which affects its selectivity towards
 268 zinc. The solubility of Zn in Urea-ChCl was in the range of 10^5 ppm, but for Cu, the
 269 solubility was very low at 10 ppm. Based on these findings, FA-ChCl was selected to extract
 270 Cu from TPCB, and Urea-ChCl for Zn extraction.

271 As for Sn, SnO₂ solubility in dicarboxylic acid DES was higher as compared to other types of
 272 DES. OA-ChCl DES also leached Cu and Zn in the solution due to the presence of Cl⁻.
 273 Oxalic acid is a good reducing agent, consequently having a high selectivity for Sn over Zn
 274 and with a high leaching efficiency (> 95 %) [39]. This high selectivity results from the
 275 interaction of 2 carbon atoms of oxalic acid, preferentially interacting with stannous salt (Sn
 276 salt) to form a Sn-dicarboxylate complex [39-40]. Furthermore, TPCB is taken as the feed

277 material to identify trends in selectivity and explore the leaching behavior of solvents under
278 the same experimental conditions as the metal oxides at 80 °C, as shown in Figure 2 and
279 elaborated in Table S1.



280

281 **Figure 2.** Selectivity of leached metals into different DESs and oxalic acid

282 3.3. Characterization of DES

283 The physicochemical properties of DES are obligatory for its industrial applications. The
284 synthesized DES, including FA-ChCl, EG-ChCl, MA-ChCl, and OA-ChCl, were
285 characterized to estimate their physical properties, including the density, viscosity, and
286 conductivity of DES. These properties have significant implications for understanding fluid
287 behavior, heat transfer, mass transport, and designing chemical processes [21]. Density and
288 viscosity were evaluated at different temperatures to comprehend the behavior of the liquid
289 under various thermal conditions.

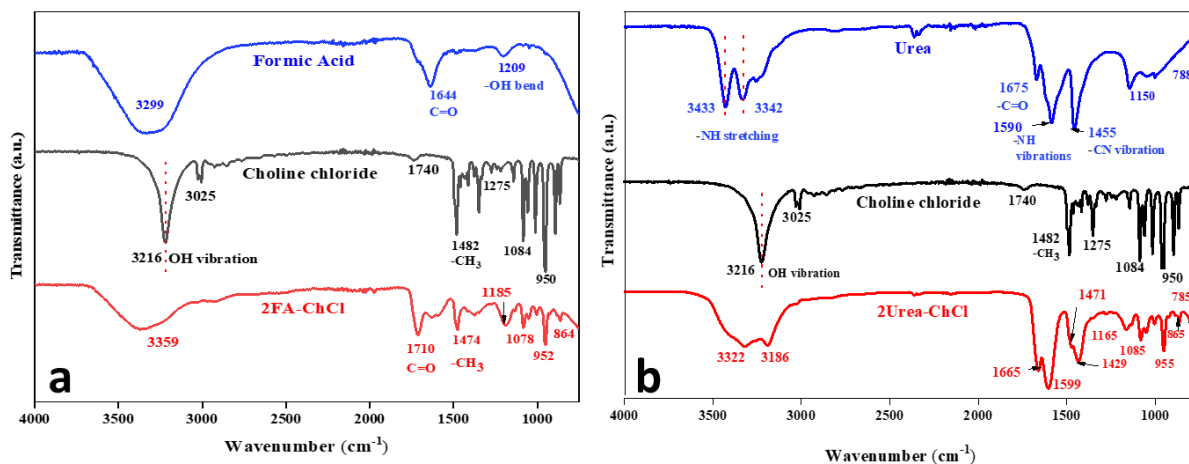
290 Choline chloride DES typically has a density of 1-1.3 g/cm³. [41]. In our last study, the
291 densities of FA-ChCl, MA-ChCl, and OA-ChCl were reported to be 1.16, 1.23 and 1.15,

292 g/cm³, respectively [41]. The density of ethylene glycol-choline chloride was 1.08 g/cm³ [42].
293 The shear viscosities of all the synthesized DES were measured at 30, 50 and 70 °C, to
294 understand the behavior of solvent viscosity with temperature according to Table 2. The
295 viscosity of FA-ChCl ranged from 0.4 to 1 mPa.s across the temperature range, which was
296 significantly lower than the viscosity of Urea-ChCl, which was several hundred mPa.s.
297 Consequently, FA-ChCl DES exhibits little mass transfer resistance compared to Urea-ChCl.
298 The high viscosity of Urea-ChCl can be attributed to factors such as the presence of a
299 prevalent hydrogen bond network, large ion size, and small void volumes. For both DES the
300 fluid viscosity decreased with increasing temperature, see Table 2 [22]. Higher temperatures
301 also enhance the leaching efficiency by enhancing the mass transfer of the dissolved species
302 [43]. In the context of extracting metal ions from electronic waste, it is often observed that a
303 reduced viscosity of the leaching agent, such as DES, is advantageous for facilitating the
304 dynamic conditions of the liquid-solid reaction and the subsequent separation process.

305 Figure 3a shows the IR spectra of FA, ChCl and 2FA-ChCl. Vibrations of FA-ChCl account
306 for OH stretching at 3359 cm⁻¹, C-H stretching at 1474 cm⁻¹, sp² C-O stretching at 1710 cm⁻¹,
307 and sp³ C-O stretching at 1078 cm⁻¹ [44]. The C-H vibration shifted from 1482 cm⁻¹ in ChCl to
308 1474 cm⁻¹ in FA-ChCl. FA-ChCl exhibits significant peaks of neat FA and ChCl peaks with
309 some shifts indicating an H-bond in between FA and ChCl.

310 Figure 3b represents the IR spectra of ChCl, urea and Urea-ChCl. The IR spectra of urea
311 show symmetric stretching C=O at 1675 cm⁻¹, which was also observed in DES with a shift
312 to 1665 cm⁻¹. The frequency of 1590 cm⁻¹ in urea represents symmetric bending NH and NH₂,
313 which was observed at 1599 cm⁻¹ in Urea-ChCl. NH stretch peak of urea at 3433 and 3342
314 cm⁻¹ has a lower shift to 3322 cm⁻¹ and 3186 cm⁻¹ in Urea-ChCl, which is attributed to the N-
315 H stretch. It should be noticed that the Urea-ChCl characteristic spectrum nearly overlaps that

316 of urea and ChCl. CH₃ at 1482 cm⁻¹, CH₂ at 1084 cm⁻¹, and CCO at 950 cm⁻¹ appeared in the
 317 DES spectrum, revealing that the structure of ChCl was retained in DES.



318
 319 **Figure 3.** FTIR of FA, ChCl and 2FA-ChCl (a); Urea, ChCl and 2Urea-ChCl (b).

320 The thermal stabilities of FA-ChCl and Urea-ChCl were evaluated using TGA. FA-ChCl
 321 degrades from 388 K, while the degradation temperature is much higher for Urea-ChCl at
 322 440 K. Therefore, the maximum operating temperature of FA-ChCl should be less than 388
 323 K, whereas, for Urea-ChCl, 440 K is the maximum operating temperature, as depicted in
 324 Figure S1.

325 **Table 2.** Viscosity (mPa.s) of the synthesized DES as a function of temperature.

DES	30 °C	50 °C	70 °C
FA-ChCl (2:1)	1.04 ± 0.001	0.54 ± 0.002	0.46 ± 0.001
Urea-ChCl (2:1)	725 ± 13.100	160 ± 2.100	99 ± 4.280

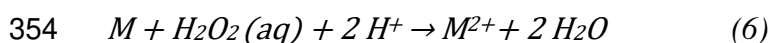
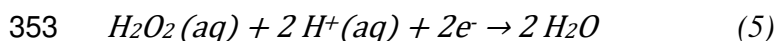
326
 327 **3.4. Parameter study on metal extraction**

328 **3.4.1. Effect of oxidizing agent**

329 DES have no oxidizing power of their own [45]. However, they could be used to dissolve
 330 metals either using oxidizing agents or as an electrolyte for the electro-oxidation of metals
 331 [45]. Without the addition of oxidizing agents, there is a slight elemental metal breakdown
 332 from e-waste into the solution, specifically in DES. A further oxidation phase with an
 333 oxidizing agent is therefore required for the dissolution of metals. Potassium permanganate

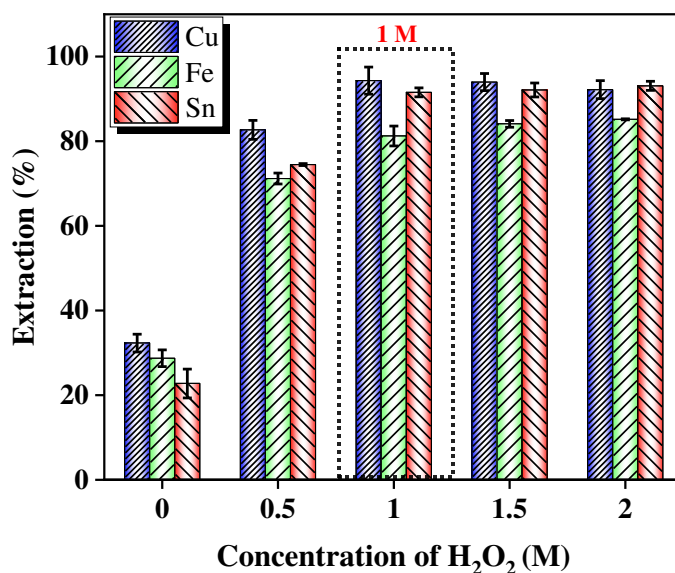
334 (KMnO₄) and hydrogen peroxide (H₂O₂) are commonly used oxidizing agents for the
335 oxidation of metals in the leaching system. In this system, both KMnO₄ and H₂O₂ were tested
336 in combination with DES to extract metals. The addition of KMnO₄ in DES showed higher
337 dissolution of copper and iron as compared to H₂O₂, as shown in Table S2. However,
338 potassium permanganate can be relatively expensive and generate significant amounts of
339 waste, making it less desirable in applications. Hence, the mild oxidizing agent hydrogen
340 peroxide was selected for further experiments.

341 Different molar concentrations of H₂O₂, ranging from 0.5 M – 2 M, have been investigated to
342 examine the impact on metal extraction while keeping other parameters constant (400 rpm,
343 24 h, 1/20 g/mL, 100 °C). After adding H₂O₂, metal extraction in FA-ChCl increases to 94.3
344 ±2 % Cu, 81.2 ±2.4 % Fe and 91.6 ±1.1 % Sn as depicted in Figure 4. There was no
345 significant change in metal extraction beyond 1M H₂O₂ in FA-ChCl DESs. Thus, 1M is
346 chosen as the optimum H₂O₂ concentration. Hydrogen peroxide has a similar effect as the
347 presence of oxygen; the extra oxygen is known as “active oxygen,” which interacts with
348 metals and changes their oxidation states. Moffett et al. discussed the mechanism of oxidation
349 or reduction reaction of H₂O₂ with Cu(II) and Fe(III) [46]. Colcleu et al. illustrated the
350 interaction of H₂O₂ with metallic copper in the presence of sulphuric acid [47]. Equation (4-6)
351 discusses the credible mechanism of metal oxidation.



355 Furthermore, the oxidized metal forms complexes either with Cl⁻ or with -COOH of DESs,
356 which enhances the efficiency of metal extraction. There have been almost negligible studies
357 on the extraction of metals from electronic waste using DESs with H₂O₂. However,

358 carboxylic acids were used as a leaching agent along with hydrogen peroxide to extract
359 metals from electronic wastes [48]. For zinc extraction in Urea-ChCl, H₂O₂ was observed to
360 have maximum extraction efficiency at 1M H₂O₂. The molar concentration of oxalic acid was
361 varied from (0.5-2 M) to check the extraction of tin according to Figure 7a. It was observed
362 that at 1 M, there is the maximum extraction of tin, and it remained the same on further
363 increasing molar concentration. However, no oxidizing agent was used to leach tin from PCB
364 using optimized 1 M oxalic acid.

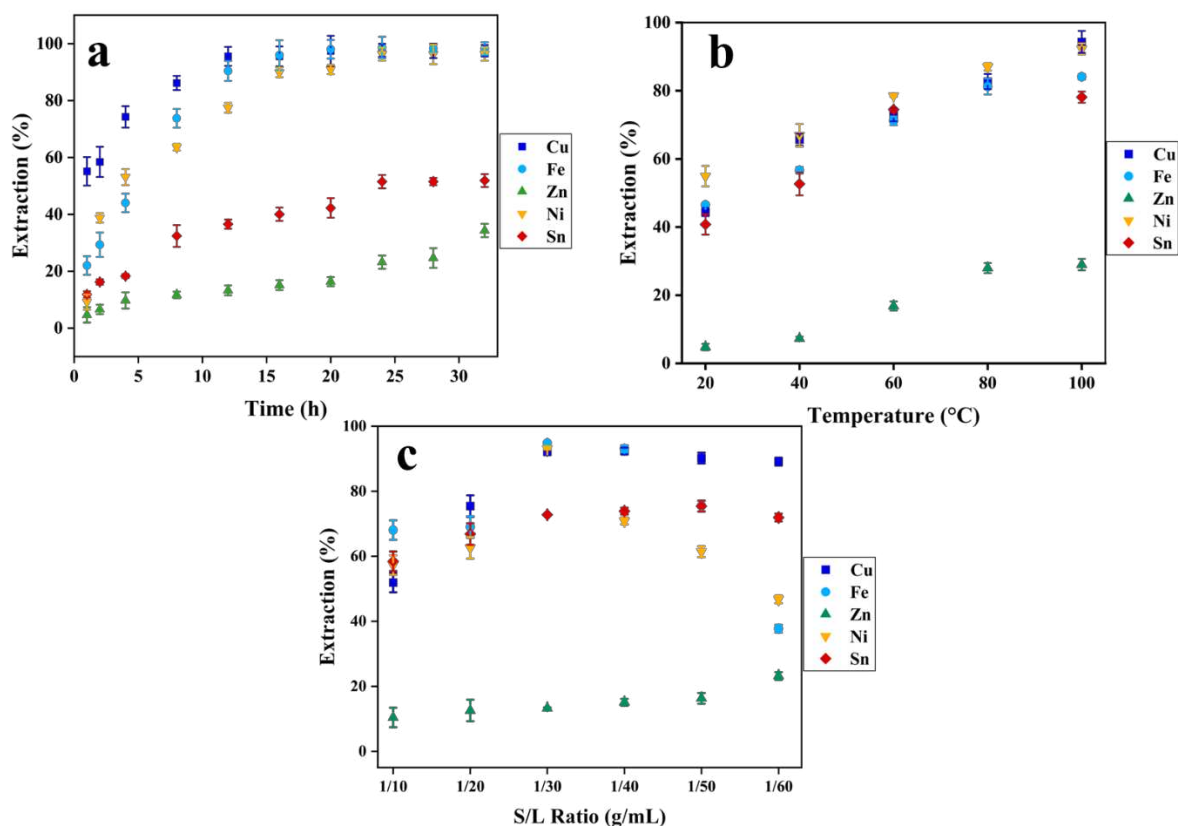


365
366 **Figure 4.** Effect of H₂O₂ concentration on metal extraction from PCB using FA-ChCl (400
367 rpm, 24 h, 1/20 g/mL and 100 °C)

368 3.4.2. Effect of temperature

369 Based on the thermal stability of solvents by TGA, the maximum operating temperatures
370 were selected. The efficiency of metal extraction by changing reaction temperature was
371 studied in the individual solvent system by ranging temperature from 20-100 °C for FA-
372 ChCl, 20-100 °C for 1M Oxalic acid, and varying 20-130 °C for Urea-ChCl. For FA-ChCl,
373 all the other parameters were kept constant, time at 21 h, solid-liquid ratio of 1/20 g/mL, and

374 stirring speed of 300 rpm. Figure 5b depicts that an increase in temperature had a remarkable
 375 effect on metal extraction efficiency from ~40-42 % of copper, nickel, and iron to greater
 376 than 90 % of metal extraction in FA-ChCl at 100 °C. An increase in reaction temperature
 377 enhances the collisions between reactants, thereby enhancing metal extraction [49]. The
 378 optimum temperature of 100 °C was selected for FA-ChCl.



379

380 **Figure 5.** Effect of Time (100 °C; 1/20 g/mL; 300 rpm) (a); Temperature (21 h; 1/20 g/mL;
 381 300 rpm) (b); S/L ratio (100 °C; 21 h; 300 rpm) (c); in metal extraction from TPCB to FA-
 382 ChCl.

383 Low-moderate reaction temperatures (20-140 °C) were selected for Urea-ChCl to evaluate
 384 temperature effect on metal extraction efficiency, keeping other parameters viz, time (21 h),
 385 solid-liquid ratio (1/20 g/mL), stirring speed (300 rpm) constant. In Urea-ChCl, zinc
 386 extraction drastically increased to 90.37 ± 1.97 % on increasing temperature to 80 °C, as

387 shown in Figure 6a. In the temperature range between 80-120 °C, selective extraction of Zn
388 was also observed. Therefore, 80 °C is preferred as the optimum temperature for Urea-ChCl.
389 1M oxalic acid was examined from 20-50 °C to evaluate the extraction of tin from TPCBs, as
390 shown in Figure 7c. It was observed that extraction of tin at 20 °C was $75.32 \pm 1.06 \%$ while
391 it raised to $91.28 \pm 2.32 \%$ at 80 °C in 1 h, 1/20 g/mL, 300 rpm.

392 **3.4.3. Effect of time**

393 The effect of metal extraction efficiency in individual DESs was studied by varying time
394 from 1-32 h at 100 °C constant temperature and 1/20 g/mL in FA-ChCl. It was observed that
395 copper, nickel, and iron extraction reached $93.5 \pm 3.47 \%$, $90 \pm 1.73 \%$, and $93.5 \pm 2.52 \%$
396 extraction at 21 h and are almost at equilibrium after 21 h. After 24 h, the extraction
397 efficiency of these three metals in FA-ChCl was above 93 %, as demonstrated in Figure 5a.
398 However, low soluble metals zinc and tin reach $34.3 \pm 2.34 \%$ and $51.8 \pm 2.27 \%$ metal
399 extraction till 32 h, respectively, in FA-ChCl.

400 The effect of time on metal extraction in Urea-ChCl was studied by varying it from 1-24 h at
401 (100 °C; 1/20 g/mL; 400 rpm). It was observed that after 16 h, zinc extraction was 90.1 ± 4.1
402 % and did not have significant changes in metal extraction, as shown in Figure 6b. Hence, 16
403 h is the optimum time for maximum Zn extraction.

404 The time effect on metal extraction from PCBs with oxalic acid was studied from 30 min to 2
405 h, as shown in figure 7b. Temperature of 100 °C, S/L ratio of 1/20 g/mL, and 300 rpm were
406 fixed. It was observed that in 1h, extraction of tin was $91.3 \pm 2.1 \%$ which varies barely on
407 increasing time till 5 h. Therefore, 1h is chosen as the optimum time for maximum tin
408 extraction.

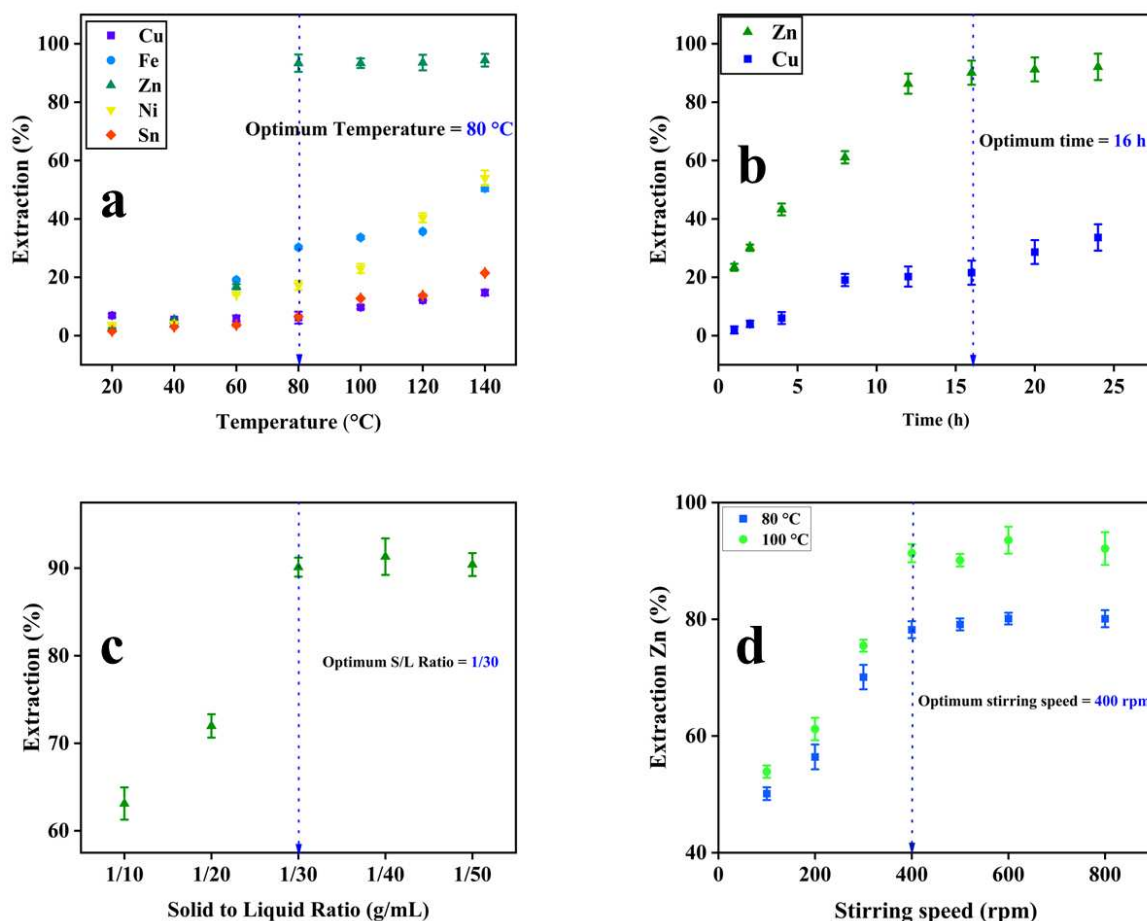
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410

411 3.4.4. *Effect of S/L ratio and stirring speed*

412 Metal extraction efficiency in DES is highly dependent on the available area per unit volume
413 of the solution. The solid-liquid ratio was examined in FA-ChCl for metal extraction, ranging
414 from 1/20 to 1/40 g/mL while keeping other process parameters constant: time (16 h),
415 temperature (80 °C) and stirring speed (300 rpm). It must be noted that along with DESs, 1M
416 H₂O₂ has been used for metal leaching experiments in DES systems. It was observed that
417 1/30 g/mL is the optimum solid-liquid ratio for 94.5 ± 2% copper, 95.1 ± 1.2% iron, and 95.6
418 ± 1.8 % nickel extraction in FA-ChCl as shown in Figure 5c. At a higher S/L ratio (1/10
419 g/mL) there is a reduction in metal extraction due to the diminution of available surface area
420 per volume of solution resulting in low interaction of PCB powder and FA-ChCl solution.
421 Furthermore, at a low S/L ratio, there is an increase in solution viscosity due to the
422 agglomeration of PCB powder particles, which is unfavorable for mass transfer and diffusion
423 resulting in low metal extraction [43]. On further decreasing the S/L ratio, there is an
424 excessive solvent that does not have a significant effect on metal extraction efficiency [50].

425 In Urea-ChCl, zinc extraction reaches 90.1 ± 1.1 % at a 1/30 g/mL ratio, with conditions of
426 100 °C, 21 h and 400 rpm stirring speed. Further, increasing the volume of solution only
427 yields a marginal difference in extraction efficiency. Thus, a 1/30 solid-liquid ratio is
428 considered the optimum condition for achieving maximum zinc extraction in Urea-ChCl, as
429 shown in figure 6c. In the case of oxalic acid, after a 1/20 g/mL ratio, there was no significant
430 difference in tin extraction as depicted in figure 7d; Wang et al. conducted selective
431 extraction of tin from zinc leaching residue using oxalic acid-sulfuric acid mixture and
432 concluded that a lower S/L ratio is beneficial for efficient mass diffusion from solid to liquid
433 [51].



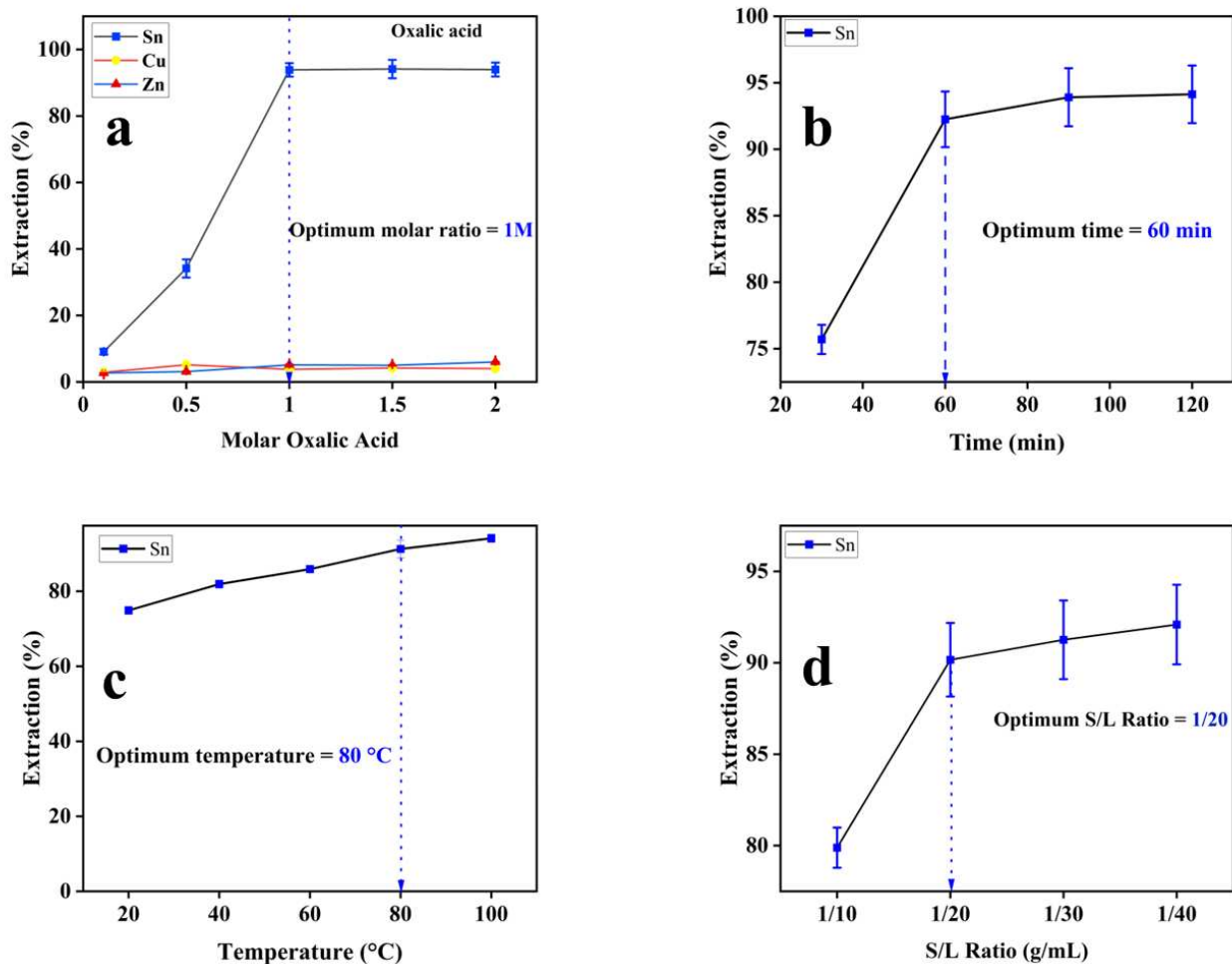
434

435 **Figure 6.** Effect of Temperature (21 h; 1/20 g/mL; 400 rpm) (a); Time (100 °C; 1/20 g/mL;
 436 400 rpm) (b); S/L Ratio (100 °C; 21 h; 400 rpm) (c); and stirring speed (d) in metal extraction
 437 from TPCB in Urea-ChCl.

438 A high S/L ratio is required to increase process yield; however, it reduces metal extraction
 439 efficiency. Hence, there should be a trade-off between metal extraction efficiency and process
 440 yield. Thus, a 1/20 g/mL and a 1/30 g/mL ratio are the optimum solid-liquid ratios for
 441 maximum tin extraction in oxalic acid and base metals extraction in FA-ChCl, respectively.

442 In order to reduce the impact of external mass transfer, the effect of stirring speed was
 443 visualized by changing the stirring speed in extremely viscous Urea-ChCl DESs from 100 to
 444 800 rpm, as shown in figure 6d. It was observed that 400 rpm is the optimum stirring speed
 445 for maximum zinc recovery. An increase in stirring speed leads to efficient mass transfer
 446 without agglomeration, thereby decreasing mass transfer resistance and increasing metal

447 extraction. However, 300 rpm is selected for FA-ChCl and oxalic acid as in literature, 300
 448 rpm was found to be the optimum stirring speed [52].



449
 450 **Figure 7.** Effect of Molar concentration of oxalic acid (100 °C; 1 h; 1/30 g/mL) (a);
 451 Reaction time (100 °C, 1/20 g/mL) (b); Reaction temperature (1 h; 1/20 g/mL) (c); S/L Ratio
 452 (100 °C; 1 h) (d). stirring speed is constant= 300 rpm in oxalic acid.

453 3.5. Kinetic study

454 The dissolution of metals from PCBs to organic acid-choline chloride DES is a solid-liquid
 455 reaction. Liquid goes through the solid with an effective diffusivity while solid was supposed
 456 to homogeneously diffuse all over the material. In the present work, the reaction for metal
 457 extraction begins at the particle's outer surface, which causes the reaction zone to contract as

458 the reaction time increases. Due to this, the reaction zone moves inward, and the extracted
 459 metals stay in the aqueous solution. There are two major assumptions of the shrinking core
 460 model. The first assumption is that the reacting solid particle has a spherical shape. This
 461 simplifies the mathematical description of the reaction because it allows for a uniform
 462 reaction rate at the particle's surface. Particles can have more complex shapes, but the
 463 spherical symmetry assumption is often used as an approximation^[53]. The second
 464 assumption is that the rate of the chemical reaction at the surface of the solid particle is
 465 uniform^[53]. As copper is a bulk metal, copper dissolution kinetics in FA-ChCl DESs are
 466 discussed in detail in this study. Metal leaching has three vital steps, which involve (1)
 467 Diffusion of reactants: molecules/ions (H⁺, FA, Urea, Cl⁻, etc.) into the solid surface of PCBs
 468 (Cu, Fe, Zn, Ni, Sn) and adsorbed (2) Interfacial surface chemical reaction by complexation
 469 reaction. (3) Diffusion of products: dissociation of the metal complex (Zn-Urea, Cu-Choline
 470 chloride, etc.) and diffusion in DES solution. As copper content in TPCBs is maximum,
 471 hence kinetics of copper extraction in FA-ChCl is discussed in detail. The process of metal
 472 extraction in organic acids involves the principle of complexation, where the organic acid
 473 forms a complex with the metal ions, which makes them more soluble in the aqueous
 474 solution. It can be implied from the abovementioned discussion that extraction of metals
 475 depends on mass transfer and chemical reaction on the surface.

476 **Table 3.** Kinetic parameters of interfacial chemical reaction controlled and diffusion-
 477 controlled models at different reaction temperatures.

Temperature (°C)	$1-(1-\chi)^{1/3}$		$1 - 2/3\chi - (1 - \chi)^{2/3}$	
	$k(\text{h}^{-1})$	R^2	$k(\text{h}^{-1})$	R^2
20	0.01159	0.870	0.0029	0.958
40	0.01209	0.916	0.00404	0.975
60	0.01917	0.913	0.00848	0.988
80	0.01840	0.951	0.00851	0.972
100	0.02481	0.922	0.01169	0.977

478
 479 Organic acids form complexes with the metal ions, and the resulting complexes are soluble in
 480 water, which facilitates their separation from the solid material and leaves an inert solid

481 residue without metals. The leaching process was analyzed using a kinetics model based on
 482 the shrinking core concept. This model was employed due to the observation that the solid
 483 surface underwent shrinkage during the reaction, resulting in the inward movement of the
 484 reaction zone and the manifestation of a shrinking core phenomenon. Within the leaching
 485 system, the step that exhibits the lowest rate is referred to as the rate-determining phase. The
 486 extraction process is diffusion controlled or chemical reaction controlled is denoted by
 487 equations (7-8)

$$488 \quad 1 - \frac{2}{3}\chi - (1 - \chi)^{2/3} = kt \quad (\text{diffusion controlled}) \quad (7)$$

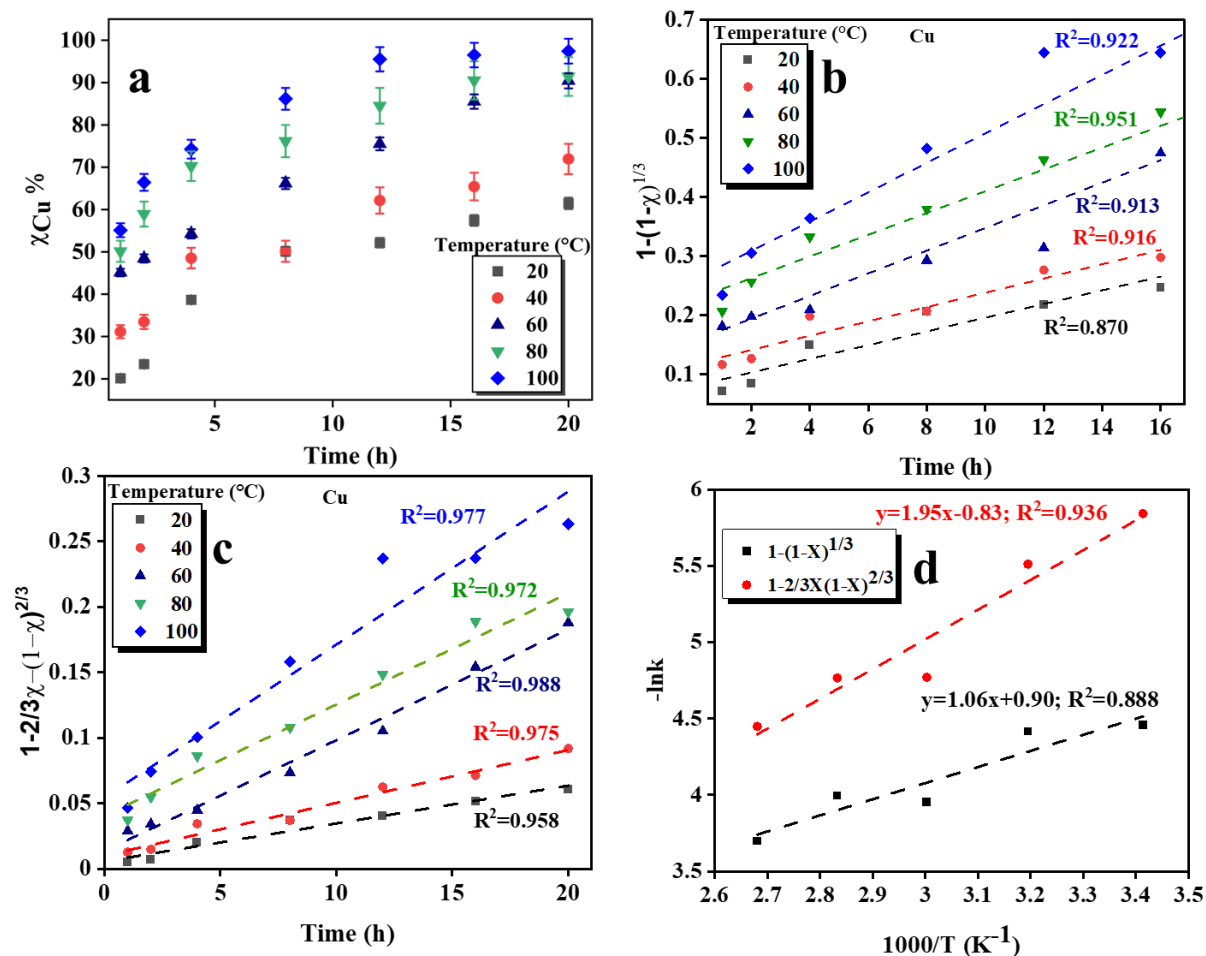
$$489 \quad 1 - (1 - \chi)^{1/3} = kt \quad (\text{interfacial chemical reaction controlled}) \quad (8)$$

490 Here, χ is the leaching efficiency of metals (Cu), %; k is the rate constant, h^{-1} ; t is time in h.
 491 Figure 8 (a-d) represents the experimental data fitting in the equations (7-8). It is evident
 492 from Figure 8 (b and c) that the extraction of copper in FA-ChCl fits well with equation 7 and
 493 is a diffusion-controlled phenomenon. The rate constants are determined in the temperature
 494 range of 20-100 °C using linear regression. The k value and R^2 are represented in Table 3,
 495 through which it can be concluded that R^2 of diffusion-controlled model $1 - \frac{2}{3}\chi -$
 496 $(1 - \chi)^{2/3}$ is near to 1. Hence, in the extraction of Cu from PCBs to FA-ChCl, diffusion is a
 497 rate-determining step. The activation energy (E_a) for Cu extraction was regulated by a
 498 linearised form of the Arrhenius equation (9) and portrayed in Figure 8d.

$$499 \quad \ln k = \ln A - \frac{E_a}{RT} \quad (9)$$

500

501



502

503 **Figure 8.** Plots of extraction of Cu with respect to time and temperature (a); $[1-(1-\chi)^{1/3}]$ vs.
 504 time at different temperatures and time (b); $[1-2/3\chi-(1-\chi)^{2/3}]$ s time at different temperatures
 505 and time (c); Arrhenius plot for extraction of Cu in both models (d).

506 Here, k is the reaction rate constant (h^{-1}), A is the frequency factor, E_a is apparent activation
 507 energy (kJ/mol), R is gas constant 8.3145 (J/K-mol) , and T is reaction temperature in K . The
 508 E_a value of Cu extraction in FA-ChCl comes to 16.21 kJ/mol in the diffusion-controlled
 509 model. It has been thoroughly reported in the literature that E_a less than 25 kJ/mol suggests a
 510 diffusion-controlled region, while greater than 30 kJ/mol suggests a chemical reaction-
 511 controlled process [54]. The shrinking core model was used by Huang et al. to fit the copper
 512 leaching data (in ionic liquids), the reaction was diffusion-controlled and the apparent
 513 activation energy calculated was 25.36 kJ/mol [19]. Topcu et al. evaluated activation energy
 514 for the dissolution of copper, 8.86 kJ/mol and zinc, 14.48 kJ/mol from copper converter slag

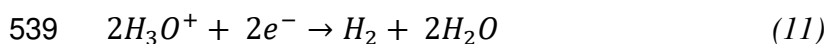
515 to urea-ChCl DESs [54a]. The activation energy of the dissolution of copper in conventional
516 acids is in the range of 11-55 kJ/mol quite higher than the activation energy of copper
517 extraction in DES [55].

518 3.6. Mechanism of metal leaching

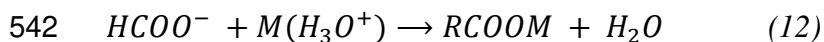
519 DESs have an interaction of hydrogen bond donors and acceptors and form hydrogen bonds,
520 hence conferring high dissolution properties. Abbott et al. discussed the chemistry of metal
521 complexes with carboxylic acid-choline chloride DESs [34]. There was a correlation between
522 the solubility of metal oxide in malonic acid-choline chloride and HCl. It was concluded that
523 the proton acts as an oxygen acceptor, forming metal complexes. Detailed analysis of metal
524 complexes revealed the formation of MCl_x^- in most choline chloride DESs [34]. Figure S3
525 reveals the map of the electrostatic potential calculation of choline chloride, which reveals
526 that chloride ions have the most negative electrostatic potential values at an iso value of 0.02.
527 The electrostatic potential is an important descriptor to understand the reactive behaviour of
528 molecules such as to determine nucleophilic and electrophilic sites in molecules. In choline
529 chloride surface analysis chloride has the most negative potential, hence more prone to
530 electrophilic attack [56]. Hartley et al. investigated the metal solubility in diols-based DESs
531 where metal speciation has proven to be a controlling factor with the dominance of anion of
532 the solvent [57]. Organic acids supply protons and ligands to dissolve metals by acidification
533 and complexation reactions. Organic acid dissociates to protons (H^+) and promotes
534 dissolution. Interactions of metal with organic acids have been discussed in the literature [58].
535 Nevertheless, there is still a lack of literature and understanding of metal dissolution in
536 organic acid based DESs.



538 Proton reduction leads to hydrogen and oxidizes metal.



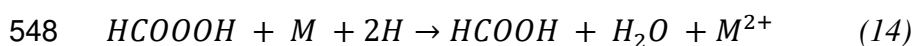
540 Complexation reaction occurs alike formate in formic acid and acetate in acetic acid to form
541 stable complex as denoted by equation (12)



543 Another reason for an increase in metal extraction with the addition of H_2O_2 is the formation
544 of peroxy carboxylic acid; in the case of formic acid, performic acid is formed as shown in
545 (13) [59]:



547 Performic acid has high oxidizing power; hence it oxidizes metal (14).



549 Urea-ChCl DES form metal complex anion $[\text{M}_x\text{ClO}(\text{urea})^-]$ from its metallic oxides or
550 sulphate forms. Urea act as a ligand and forms metal complexes anions.



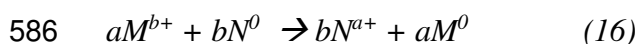
552 The energy difference between the Highest Occupied Molecular Orbitals (HOMO) and the
553 Lowest Unoccupied Molecular Orbital (LUMO) of a complex is a crucial factor that helps in
554 understanding the electron properties and reactivity. High and low energy difference has a
555 significant implication for the complex's behaviour. The energy difference of HOMO LUMO
556 of metal complexes is illustrated in Table S5. Formic acid interaction with copper has the
557 lowest energy gap, which reveals its higher reactivity with the complex as compared to other
558 hydrogen bond donors. However, the low energy gap of oxalic acid interaction with copper
559 reveals its tendency to form copper oxalate. Figure S4 shows the interaction of the Cu₄
560 cluster with different electron-deficient sites of components of DES.

561 Eh-pH diagram, commonly known as the Pourbaix diagram, aids in understanding the
562 behavior of individual metals in aqueous solutions. Materials project an open-source
563 platform, was utilized to plot Eh-pH diagram of Fe, Cu, Ni, Zn and Sn as shown in Figure S2
564 (a-e) [60]. Fe^{2+} , Cu^{2+} , Zn^{2+} , Ni^{2+} and Sn^{2+} are the major metal species present in solution.

565 Standard electrode potential of (metals and H₂O₂) and pH of all DES are tabulated in Tables
566 S2 and S3, respectively. Electrode potential differences of Fe, Cu, Zn, Ni, Sn and H₂O₂ are
567 1.336 V, 1.436 V, 2.536 V, 2.026 V and 1.63 V, respectively. Considering the Eh-pH
568 diagram of all metals, Fe²⁺, Cu²⁺, Zn²⁺, Ni²⁺ and Sn²⁺ species exist in the acidic pH range
569 above -0.5 V, -0.4 V, -0.8 V, -0.2 V and -0.5 V, respectively. Metal species form metal
570 complexes as denoted by equation (12-13) resulting in metal extraction. For tin extraction in
571 oxalic acid, it was presumed to form tin oxalate. Wang et al. discussed the reduction of Sn
572 (IV) to Sn (II) by E-pH diagram [51]. Sn⁴⁺ has a stable region in the highly acidic
573 environment; however, the required pH range for the stable region of Sn²⁺ is below pH 3.
574 Compared to sulphuric acid and hydrogen peroxide, E (CO₂/H₂C₂O₄) reduction potential is -
575 0.59 V, while E(Sn⁴⁺/Sn²⁺) is 0.15; hence thermodynamically oxalic acid can reduce Sn (IV)
576 to Sn (II), thereby has higher extraction.

577 **3.7. Separation of metals from leached solution**

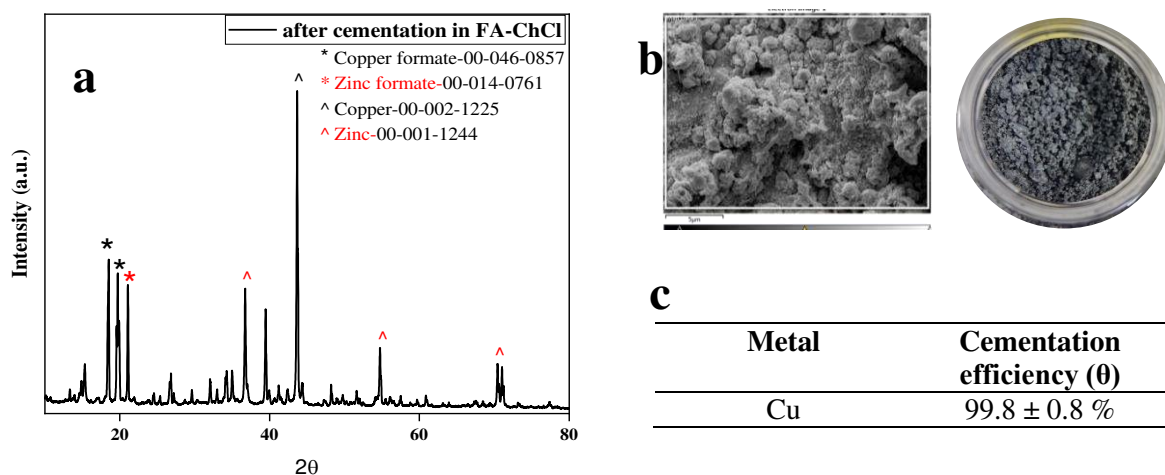
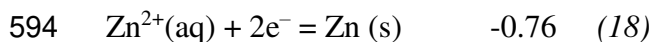
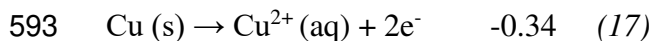
578 Cementation is an electrochemical phenomenon characterized by the displacement of a metal
579 ion from solution by more reactive metals (sacrificial). Cementation has the advantage of
580 easy operation, cost and high metal recovery and is a common technique used in metallurgy
581 and hydrometallurgy for metal recovery from leached solution. The cementation process is
582 classified as a galvanic process due to its lack of dependence on an external current source.
583 The phenomenon of metal ion reduction in a solution to its metallic form is attributed to the
584 disparity in standard electrode potential between the two metals involved. The equation of the
585 typical cementation process is:



587 Here, M^{b+} is the target metal, M⁰ is the cemented target metal in metallic form, N⁰ is a
588 sacrificial metal used for cementation while N^{a+} is a sacrificial metal used for cementation.

589 Table S3. illustrates the standard electrode potential of metals. The disparity in standard

590 reduction potentials between copper (Cu) at 0.34 V and zinc (Zn) at -0.76 V instigates the
 591 reduction of copper ions into their elemental form. The standard electrode potentials of
 592 copper and zinc are illustrated by



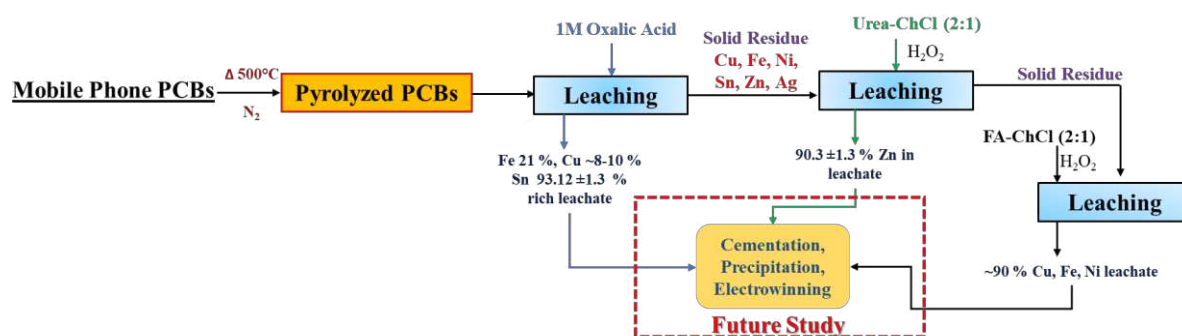
595 **Figure 9.** XRD of recovered powder after cementation from FA-ChCl leached solution (a).
 596 FESEM of recovered salt along with visual representation (b). Cementation efficiency of
 597 copper from FA-ChCl leached solution using zinc dust (c).

598 As discussed in our previous paper by Jadhao et al. (2023), the cementation of copper is
 599 thermodynamically feasible^[52]. An optimized metal leached solution of FA-ChCl was used
 600 for this process. At room temperature, metallic zinc powder was added to the leached FA-
 601 ChCl solution. The cementation efficiency of copper was observed to be $99.8 \pm 0.8 \%$.
 602 However, XRD of recovered powder reveals the presence of copper formate along metallic
 603 copper. Figure 9 depicts the XRD of recovered powder after cementation along with FESEM
 604 and cementation efficiency.

605 3.8. Proposed integrated process.

606 The cumulative multistage leaching process was employed for the selective extraction of
 607 metals from PCBs taking 5 g of TPCBs. Initially, 1M Oxalic acid was used to leach Sn from
 608 the PCB for 1 h, in 1/20 g/mL ratio at 80 °C, 300 rpm. ICP-MS analysis of leachate revealed

609 that $93.5 \pm 0.4 \%$ of Sn was leached in 1M oxalic acid solution. Alongside, iron extraction
 610 was $34.7 \pm 1.2 \%$ and $10-8 \%$ of Cu in the solution. Further, the solid residue was washed and
 611 treated with Urea-ChCl for 16 h, 1/30 g/mL ratio, $80 \text{ }^\circ\text{C}$, 400 rpm to extract $90.3 \pm 1.3 \%$ of
 612 zinc. The solid residue was washed and dried and treated with FA-ChCl DES to extract $90 \pm$
 613 0.9% Cu, $89 \pm 0.2 \%$ Fe, and $88 \pm 1.9 \%$ Ni in a leachate solution. Figure 10 depicts the
 614 proposed process flowsheet.



615

616 **Figure 10.** Flowsheet of proposed work for selective extraction of metals from PCBs

617 There are various aspects in metal extraction from DES which is still not much explored.
 618 Exploration of electrowinning integrated with DES metal extraction can be done. The
 619 integrated process environmental impact could be evaluated using LCA studies. Alongside,
 620 scaleup study to set up greener technology alternative to conventional technology ought to be
 621 explored.

622 4. Conclusion:

623 In this work, DESs (FA-ChCl, Urea-ChCl) and organic acid have been tested as potential
 624 leaching agents for the selective extraction of metals from PCBs. Leaching experiments states
 625 that FA-ChCl showed high copper, iron, and nickel solubility in the range of 10^5 ppm while
 626 negligible solubility of Zn. However, Urea-ChCl was observed to be selective in the
 627 solubility of Zn. Metal leaching experiments in FA-ChCl with H_2O_2 uncover that extraction
 628 of Cu, Fe, Ni, and Zn is greater than 90% at optimized conditions ($100 \text{ }^\circ\text{C}$, 1/30 g/mL, 16 h,
 629 300 rpm). Zn extraction from TPCBs was observed to be $90.4 \pm 2.9 \%$ at ($100 \text{ }^\circ\text{C}$, 1/20 g/mL,

630 21 h, 400 rpm). 1M oxalic acid selectively extracted 92.3 ± 2.09 % Sn from TPCBs under
631 optimal conditions (1 h, 80 °C, 1/20 g/mL). The diffusion-controlled reaction is suggested by
632 a shrinking core model-based kinetic investigation of copper extraction in FA-ChCl. The
633 apparent activation energy of copper was 16.21 kJ/ mol. Metal extraction mechanism in DES
634 is a two-step process with a major complexation reaction of metals with ligands of hydrogen
635 bond donor or acceptor anion, which is explained by Eh-pH diagram of metals. This study
636 provided a detailed investigation of metal extraction in green solvents, which paved a new
637 approach for the selective extraction of metals from PCBs. Further analysis of the mechanism
638 of metal extraction in the green solvent must be studied in detail, correlating theoretical
639 calculations with experimental results to further understand at a molecular scale. Moreover,
640 detailed research on metal speciation in DES systems is required to fully comprehend the
641 mechanism.

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680 ABBREVIATIONS
681 B3LYP: Beck's three parameter functional Lee Yang Parr
682 ChCl: Choline Chloride
683 DESs: Deep Eutectic Solvents
684 DFT: Density Functional Theory
685 EG: Ethylene Glycol
686 EG-ChCl: Ethylene Glycol-Choline chloride
687 FA: Formic Acid
688 FA-ChCl: Formic Acid-Choline chloride
689 FTIR: Fourier Transform Infrared Spectroscopy
690 HBA: Hydrogen Bond Acceptor
691 HBD: Hydrogen Bond Donor
692 ILs: Ionic Liquids
693 MA-ChCl: Malonic Acid-Choline chloride
694 OA: Oxalic Acid
695 OA-ChCl: Oxalic Acid-Choline chloride
696 PCB: Printed Circuit board
697 TPCBs: Thermally treated printed circuit boards
698 TGA: Thermo Gravimetric Analysis
699 Urea-ChCl: Urea-Choline chloride

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