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The enduring world forest carbon sink

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1 **Carbon dioxide uptake by terrestrial ecosystems is critical for moderating climate change¹.**
2 **To provide a ground-based long-term assessment of the contribution of forests to**
3 **terrestrial uptake, we synthesized *in situ* forest data from boreal, temperate, and tropical**
4 **biomes spanning three decades. We found that the carbon sink in global forests was steady**
5 **at 3.6 ± 0.4 Pg year⁻¹ in the 1990s and 2000s, and 3.5 ± 0.4 Pg year⁻¹ in the 2010s. Despite**
6 **global stability, our analysis reveals major biome-level changes. Sinks have increased in**
7 **temperate ($+30 \pm 5\%$) and tropical regrowth ($+29 \pm 8\%$) forests due to increases in forest**
8 **area, but decreased in boreal ($-36 \pm 6\%$) and tropical intact forests ($-31 \pm 7\%$) due to**
9 **intensified disturbances and losses in intact forest area, respectively. Mass-balance studies**
10 **suggest Earth's land sink has increased², implying an increase in the non-forest land**
11 **carbon sink. The global forest sink is equivalent to almost half of fossil fuel emissions ($7.8 \pm$**
12 **0.4 Pg C year⁻¹, 1990-2019). However, two-thirds of the sink's benefit has been negated by**
13 **tropical deforestation (2.2 ± 0.5 Pg C year⁻¹, 1990-2019). While the global forest carbon sink**
14 **has endured undiminished for three decades despite regional variations, it may weaken**
15 **because of aging forests, continuing deforestation, and further intensification of**
16 **disturbance regimes¹. To protect the sink, land management policies must limit**
17 **deforestation, promote forest restoration and improve timber-harvesting practices^{1,3}.**

18
19
20 Atmospheric carbon dioxide (CO₂) concentration surpassed 420 ppm in 2023⁴, and climate
21 change is approaching potential tipping points that portend significant future impacts¹ without
22 urgent actions^{5,6}. While humanity has converged on the goal of achieving net zero greenhouse
23 gas emissions by 2050⁷, one of the most challenging elements is the need for large-scale

24 “negative emissions” of up to 6 Pg C year⁻¹ to compensate for the inability to eliminate all
25 emissions from fossil fuels⁸. The land sector has capacity to sequester and store additional
26 carbon because historically it has lost 180 Pg of stored C due to land use changes, and this
27 former reservoir can be restored to some extent^{5,9,10}. As forests are the dominant component of
28 the land carbon sink¹¹, we need to know how much atmospheric carbon the world’s forests have
29 been sequestering, where it is stored, and whether recent trends are consistent with the desired
30 strengthening of Earth’s land sink.

31 Recent advances in remote sensing, modelling and computation can map and model
32 Earth’s land sinks at high temporal and spatial resolution yet have difficulty in generating long-
33 term baselines and may diverge substantially in some regions and timeframes¹². In contrast, the
34 extensive ground-based and historical information from forest inventories and ecological studies
35 permits analysis of forest dynamics (growth, harvest, mortality) by region or country, all
36 ultimately based on tree-by-tree measurement of size, species and biomass. Whether regional¹³
37 or global¹¹, these data provide a unique perspective on Earth’s forests and how they are changing,
38 and are highly complementary to top-down or model-driven approaches. The length, quantity
39 and consistency of such records now permit a three-decade perspective on Earth’s global and
40 regional forest carbon balance and fluxes to span the entire period of land-use change, shifting
41 forest dynamics and accelerating climate change since the IPCC’s First Assessment Report in
42 1990¹⁴.

43 We analyzed multiple decades of ground-based measurements by the global forest
44 community (Table S1a, b, c), combined with forest area estimates based on remote sensing in
45 national forest inventories and other types of land surveys, to estimate the recent magnitude,
46 trend, impact factors, and locations of the global forest carbon sink. We constructed a global

47 record of forest inventory measurements from 1990 through 2019, supplemented with high-
48 quality data from long-term ecosystem monitoring sites. Our estimates of the forest land carbon
49 sink are largely independent of other approaches including atmospheric CO₂ observations and
50 inverse models¹⁵, dynamic global vegetation models (DGVMs)¹⁶, and mass-balance
51 assessments². The uncertainty of our estimated global forest carbon sink is ~0.4 Pg C year⁻¹,
52 compared with other estimated terrestrial sinks² having uncertainties ranging from ~0.5-1.8 Pg C
53 year⁻¹. We call for investment in specific research and monitoring priorities for reducing
54 uncertainties in forest carbon assessments.

55

56 **Global forest areas, C stocks, and sinks**

57 The world's forest area declined by 5% from 1990 to 2020, from 4,022 Mha to 3,812
58 Mha (-210 Mha) (Extended Data Table 1). This net decline in forest lands is driven by losses in
59 the tropics (-273 Mha, -13%). In contrast, temperate forest area increased (+52 Mha, +7%) while
60 the boreal forest area was stable (+12 Mha, +1%). Within the tropics, 467 Mha (26%) of intact
61 forest was lost but the area of regrowth forests expanded (+194 Mha, +56%).

62 The C stock in the world's forests in 2020 was 870 ±61 Pg C (Extended Data Table 2). In
63 boreal, temperate, and tropical regrowth forests stocks increased by 74 Pg C over three decades.
64 Meanwhile, deforestation reduced intact tropical forest carbon stocks by 149 Pg, while remaining
65 intact tropical forests sequestered 32 Pg C to make up some of the losses (Extended Data Fig.1).
66 Most of the 2020 global forest C stock is in live biomass (43%) and soils (45%), with smaller
67 proportions in dead wood (8%) and litter (4%). The fraction of total C in living biomass
68 increases towards the equator, while the proportion in soils shows the opposite pattern: boreal
69 forests stored 20% of their C in living biomass and 64% in soils; temperate forests 38% in living

70 biomass and 54% in soils; and tropical forests 57% in living biomass and 32% in soils. Total C
71 stocks were highest in the tropics, lowest in temperate and intermediate in boreal forests.

72 The C density (Mg C ha^{-1}) increased from 1990 to 2020 in each biome (Extended Data
73 Fig.2c). This suggests that global forests overall continued to gain C nearly everywhere,
74 consistent with rising CO_2 concentrations increasing photosynthetic rates globally^{17,18}. Other
75 factors, such as warmer temperatures and increased N deposition may also enhance forest C
76 densities regionally (Table S2). Nevertheless, the average global forest C density barely changed.
77 This apparent paradox is due to the loss of high density intact tropical forests and their partial
78 replacement by much lower C density regrowth forests, resulting in the average global forest C
79 density staying near constant despite density increases within each category (Table S3).

80 The C sink in Earth's forests was estimated at 3.59 ± 0.34 , 3.57 ± 0.36 , and 3.53 ± 0.41 Pg
81 C year^{-1} for the 1990s, 2000s to 2010s (Table 1), statistically stable over the decades
82 (Supplementary Information, Fig. S1). Stable global totals mask large biome-scale changes: an
83 increased sink in temperate (+30%) and tropical regrowth forests (+29%) but a decline in boreal
84 (-36%) and tropical intact (-31%) forest sinks. Further, the C sink in global established forests—
85 excluding tropical regrowth forests—declined by 19% from 2.32 ± 0.21 to 1.89 ± 0.24 Pg C year^{-1}
86 over 30 years (Table 1). After accounting for C emissions from tropical deforestation, the *net* C
87 sink (Extended Data Box 1) in Earth's forests was still positive but showed no statistically
88 significant trend (Fig. S1), being 0.93 ± 0.63 , 1.66 ± 0.56 and 1.39 ± 0.69 Pg C year^{-1} in the 1990s,
89 2000s and 2010s (Table 1).

90

91 **Forest C sinks by regions, biomes, and pools**

92 ***Boreal forests***

93 The boreal C sink declined from 508 ± 63 Tg C year⁻¹ to 324 ± 41 Tg C year⁻¹ from the
94 1990s to 2010s (Extended Data Table 3) and was strongly affected by Asian Russian forests that
95 account for 57% of the boreal area (Extended Data Table 1). The C sink in Asian Russian forests
96 declined by 42% over three decades, with the greatest reduction occurring in the late 2010s,
97 primarily due to increased severity of wildfires, insect outbreaks, and increased logging both
98 legal and illegal¹⁹ (Fig. 1). Notably, living biomass contributed a large C sink in the 1990s (145
99 Tg C year⁻¹) but switched to a source in the 2010s (-20 Tg C year⁻¹); meanwhile the deadwood
100 sink increased²⁰ by 44%. Alaska Interior managed forests were a small C sink in the 1990s,
101 which was reduced by 76% in 2010s likely due to soil warming and increasing wildfires²¹.
102 Canadian managed forests were about C neutral in the 1990s and small sources in the 2000s and
103 2010s (Extended Data Table 4). The much greater source from living biomass in 2000s (-55 Tg
104 C year⁻¹) was caused by increased outbreaks of insects and wildfires²². In the 2010s, living
105 biomass, deadwood and litter pools all became C sources while the soil sink was reduced by
106 35%, reflecting increased impacts of disturbances, warming, and droughts²².

107 Unlike Canadian managed forests which have become drier, European Russia and
108 European boreal forests have become wetter over the last half century²³. Boreal forests of
109 European Russia had a relatively stable multi-decadal C sink with a slight increase in the 2000s
110 when agricultural lands that were abandoned in the 1990s returned to forest²⁴, followed by a
111 slight decrease in 2010s likely due to increased harvesting and disturbances. However, our
112 estimates show that the soil C sink decreased by 31% in the 2010s compared to the 2000s,
113 possibly related to impacts of soil warming²⁵. European boreal forests showed an increasing C
114 sink over time resulting from improved management and growth enhancements due to CO₂
115 fertilization and lengthening growing seasons²⁶. The latest forest inventory updates from

116 Finland²⁷ and Sweden²⁸ indicate a recent sink downturn responding to a combination of drought,
117 changes in stand age structure and roundwood imports, and intensive harvests (Supplementary
118 Information).

119

120 *Temperate forests*

121 The C sink in temperate forests was 526 ± 37 Tg C year⁻¹ in the 1990s, increasing to 685
122 ± 50 Tg C year⁻¹ in the 2010s (Extended Data Table 3). The major driver was the increase in
123 China's forest area under national-scale afforestation and reforestation programs during the late
124 1980s and early 1990s, as those new forests reached their high productive stages in the 2000s
125 and 2010s, increasing the sink by 86 Tg C year⁻¹ each decade²⁹.

126 The C sink in U.S. forests decreased by 10% in the 2000s compared to the 1990s and
127 remained at that level in the 2010s (Fig. 1). In the 2000s, U.S. forests experienced increased
128 natural disturbances and summer droughts³⁰. Although the U.S. forest C sink did not recover
129 fully in the 2010s, the rate of decline was reduced. The C sink in European temperate forests
130 declined by 12% from 2000s to 2010s (Extended Data Table 3) probably because large forest
131 areas planted in the 1950s approached C saturation as they matured³¹. More recently, central
132 European forests suffered increasing bark beetle damage triggered by several years of droughts³²,
133 which could lead to forests becoming C sources at the national level, although droughts alone did
134 not seem to induce decreased growth³³.

135 In Japan, the C sink in living biomass decreased significantly in the 2010s, related to
136 aging of forests planted in the 1960s³⁴ (Extended Data Table 3). Australian forests were C
137 sources in the 1990s and 2000s and became merely neutral in the 2010s (Extended Data Table
138 3). This C source was due to extensive deforestation for agriculture, which declined in the recent

139 decade because of legislative restrictions on clearing. Carbon was also lost from harvesting of
140 native, high C density forests that were replaced by younger lower C density regrowth forests.
141 Intensified droughts and wildfires in the 2000s and 2010s also contributed to increased net
142 annual emissions.

143

144 *Tropical intact forests*

145 The C sink in tropical intact forests declined from 1284 ± 202 Tg C year⁻¹ in the 1990s to
146 881 ± 235 Tg C year⁻¹ in the 2010s (Extended Data Table 3), caused mainly by deforestation that
147 reduced the remaining intact forest area by 26%. The greatest losses proportionally occurred in
148 Southeast Asia, with 53% loss of intact forests (101 Mha) in the past 30 years, largely because of
149 expansion of oil palm plantations³⁵. The greatest losses by area were in South America (187
150 Mha, 22%) and Africa (175 Mha, 29%) (Extended Data Table 1). The C contained in deforested
151 lands (149 Pg C) had different fates: about 45% was rapidly emitted to the atmosphere, 17% lost
152 to processing harvested timber and for use of short-lived wood products such as paper, 2% was
153 stored in long-lived wood products such as construction materials, and the remaining 36%
154 continued to be stored on the land in the new land-use types, such as ranchland soils (Extended
155 Data Fig. 1).

156 The tropical intact forest C sinks declined in Southeast Asia, Africa, and South America
157 by 25%, 17% and 42% respectively (Extended Data Table 3). South America experienced the
158 largest reduction because it lost most intact forest area, and because Amazon droughts
159 contributed to increased tree mortality and slowing of tree growth rates^{36,37,38}. Consequently, the
160 2010s sink in South American intact forests was less than two-thirds that in the 1990s (Fig. 1).
161 The smallest decline in the forest C sink was in Africa, reflecting similar proportional losses of

162 forest area but less impact of drought and warming on forest processes³⁷. The decreased C sink
163 in the Southeast Asia forest was mainly driven by forestland losses.

164

165 ***Tropical regrowth forests***

166 The C sink in tropical regrowth forests increased from 1273 ± 260 Tg C year⁻¹ in the
167 1990s to 1640 ± 333 Tg C year⁻¹ in the 2010s. Despite occupying just 20% of the area of intact
168 forest in the 1990s, these forests had a similar C sink (Extended Data Table 3) because their C
169 sequestration rates are about five times higher, reflecting the early successional biomass
170 accumulation phase of tropical forests. The regrowth C sink increased greatly in the 2000s and
171 2010s with expanded areas (Extended Data Table 1). Overall, the increasing tropical regrowth
172 forest C sink balanced the declining sink in intact forests across 1990 to 2020, resulting in a near-
173 constant tropical forest C sink of $\sim 2.5 \pm 0.4$ Pg C year⁻¹ for three decades (Table 1). Although C
174 sinks in both tropical intact and regrowth forests are large, high emissions due to deforestation
175 and degradation counteracted nearly all of these remarkable sinks, making tropical forest lands
176 almost carbon neutral (Extended Data Fig.3) with a small net sink/source between -0.1 and 0.6
177 Pg C year⁻¹, fluctuating with deforestation intensities in different decades (Table 1).

178

179 ***Necromass and Harvested Wood Products***

180 We include estimates of C stock and sink in different components of forest necromass
181 (non-living organic matter in standing and lying deadwood, litter, and soils) to enable reporting
182 of complete forest ecosystem carbon budgets even though estimation of these pools has greater
183 uncertainty. Necromass accounts for an average of 58% of total forest C stocks (514 ± 52 Pg C),
184 with proportions smallest in tropical forests (45%, 226 ± 42 Pg C), intermediate in temperate

185 forests (64%, 80 ± 9 Pg C), and greatest in boreal forests (80%, 207 ± 10 Pg C) (Extended Data
186 Table 2). The fraction of the C sink in necromass was 30% (781 ± 154 Tg C year⁻¹) of that in
187 living biomass globally, but varied greatly among biomes, averaging 184% (266 ± 48 Tg C year⁻¹)
188 in boreal forests but just 26% and 20% in temperate (109 ± 16 Tg C year⁻¹) and tropical forests
189 (406 ± 105 Tg C year⁻¹) (Extended Data Table 3).

190 Harvested wood products (HWP) is defined as a C sink, related to the amount of timber
191 harvested and the portion that remains in use or in solid waste disposal sites. Globally, only
192 ~10% of C in harvested timber is counted as HWP³⁹ because about half of the wood is used for
193 fuel and much of the rest lost during processing into wood products, followed by losses when
194 products are discarded and decompose³. The average half-life of pulp and paper products is just
195 two years while for sawnwood products it is 35 years³⁹. The annual HWP increased by 10% over
196 three decades to 0.21 Pg C year⁻¹ in the 2010s, implying more wood harvested from forests. On
197 average, HWP contributes only 6% of the global C sink (7%, 13% and 4% in boreal, temperate,
198 and tropical forests respectively) (Extended Data Table 3), although this estimate does not fully
199 account for the effects of illegal logging on wood harvesting fluxes.

200

201 **Status of the global forest carbon sink**

202 Our estimates show a large, long-term persistent sink of 3.56 ± 0.37 Pg C year⁻¹ in global
203 forests since at least 1990 with a statistically insignificant change based on Monte-Carlo
204 simulations and Cohen's *d* (Supplementary Information, Fig. S1). While stable overall, the
205 contribution to this carbon sink by different forest biomes has fluctuated greatly over time.
206 Within the tropics there has been a shift from equal contributions of intact and regrowth tropical
207 forests in the 1990s, to 65% of the sink being in regrowth forests in the 2010s as the intact sink

208 declined and the regrowth sink increased (Table 1). Boreal and temperate forests contributed
209 similar C sinks in the 1990s, but by the 2010s the boreal sink had decreased to less than half the
210 temperate sink (Table 1).

211 Carbon stock densities (Mg C ha^{-1}) in all forest biomes in all climate zones steadily
212 increased (Extended Data Fig. 2c), showing that forest ecosystems across the planet continuously
213 sequestered C, implying a universal growth factor, or several factors, enhancing forest sinks at
214 continental scales. A suite of multidisciplinary evidence suggests that the global C sink
215 persistence and C density increases were in part due to the CO_2 fertilization effect contributing to
216 substantially increased photosynthesis^{17,18}, in addition to longer growing seasons in temperate
217 and boreal regions²⁶. These may have outweighed negative effects on forest C from global
218 heating, changing rainfall patterns, and changes in the frequency and severity of natural
219 disturbances in remaining forests^{1,5}.

220

221 **Regional vulnerability of C sink and future prospects**

222 The C sink in Earth's forests is vulnerable to deforestation, degradation, and disturbances
223 triggered or intensified by climate change. In intact tropical forests, the foremost threats remain
224 ongoing deforestation and degradation, the primary causes of the declining C sink (Extended
225 Data Fig.1). More intense and frequent droughts have also killed millions of trees, contributing
226 to a weaker C sink in Amazonia^{37,40}. Given that the combined sink in intact and regrowth forest
227 is stable, the sign of the net sink for tropical forests as a whole is largely determined by the rate
228 of deforestation emissions. Only reducing deforestation and degradation will keep stored carbon
229 out of the atmosphere and by protecting tropical forests we also protect their biodiversity and
230 sink capacity in the future.

231 Boreal forests have experienced major impacts from climate change, including greater
232 increases in temperature and variability than other regions⁴¹. Climate change has disrupted C
233 dynamics in vegetation and soils, and exacerbated disturbances by wildfires, insect outbreaks,
234 and droughts. The high C stock and sink in boreal forest necromass are threatened by increased
235 decomposition rates and wildfires following dry conditions⁴². These impacts made Canadian
236 forests a C source²², while Asian Russian forests lost 42% of their sink strength over three
237 decades, particularly in the late-2010s²⁵. Future threats for boreal forest C dynamics also include
238 northward shifting of bioclimatic zones that directly causes thawing of permafrost, triggering
239 megafires such as occurred in 2020-2022, increased risk of large-scale pest outbreaks, and
240 increased rates of legal and illegal logging, which all result in release of methane and CO₂
241 (Supplementary Information).

242 Temperate forests include Earth's most intensively managed forest ecosystems. The
243 increased C sink resulted mainly from past tree planting in China²⁹. Temperate forests that
244 recovered on abandoned agricultural lands or heavily harvested forests in early-to-mid last
245 century are now approaching the age at which growth rates begin declining, though growth
246 trajectories and successional dynamics differ within the temperate forest biome^{31,34,43}. Climate
247 change has caused increases in frequency and intensity of natural disturbances, triggering
248 intensified bark beetle outbreaks following drought in some European forests³². Additionally,
249 increasing temperate zone tree harvests over the three decades (+17%) caused loss of stocks.

250 Although asynchronous regional dynamics ensured that the aggregate C sink in Earth's
251 forests was almost constant, our analysis shows how biome- and continental-scale forest C sinks
252 were susceptible to multiple environmental changes and timber harvesting. All these factors
253 impact growth, mortality, and stocks and therefore future changes will affect the persistence and

254 strength of the global forest C sink. With several strong positive and negative drivers (Table S4),
255 each likely to develop differently among biomes and regions, the global forest C sink has an
256 uncertain future. We therefore recommend carefully monitoring its future evolution.

257

258 **Comparing estimates of land C sinks using different approaches**

259 Our estimates for forests can be placed in the context of terrestrial sinks and sources
260 estimated from the Global Carbon Budget (GCB)² (Fig. 2). Both GCB's mass-balance and the
261 mean of 17 DGVMs' results estimated that the land **gross** C sink grew⁴⁴, meaning that the
262 contribution of Earth's total forest C sink (~ 3.6 Pg C yr⁻¹) to the land **gross** sink has declined
263 relatively from 75% in the 1990s to 65% in the 2010s (Extended Data Table 4). This also implies
264 that non-forest lands have been progressively removing more carbon from the atmosphere (Fig.
265 2). Our results showing relatively stable global forest **gross** sinks contrast with most carbon
266 model estimates which show C uptake is increasing across most forest biomes⁴⁴. This means that
267 the modelled future terrestrial C uptake by forests may be overestimated.

268 By contrast, over the three decades the global forest **net** sink (1.3 Pg C year⁻¹) amounted
269 to 91% of the land **net** sink (1.4 Pg C year⁻¹) (Fig.2). The forest **net** sink we estimated thus
270 compares closely to the **net** land sink independently estimated using DGVMs, of 0.9, 1.2 and 1.5
271 Pg C year⁻¹ for the 1990s, 2000s, and 2010s respectively, and is broadly comparable with inverse
272 model estimates and other methods⁴⁴. Finally, while the magnitude of the global forest **net** sink is
273 only 17% that of fossil fuel emissions, the forest **gross** C sink was of course much greater. The
274 total three-decadal sink of 106.9 Pg C is equivalent to $\sim 46\%$ of fossil fuel emissions. Even for
275 the 2010s the global forest C gain would have amounted to 37% of contemporary fossil fuel
276 emissions had deforestation ceased (Extended Data Table 4).

277

278 Uncertainties, data gaps and future research priorities

279 Uncertainty of stock-change estimates varied by biome, with the largest uncertainties in
280 tropical (+/- 27%) and boreal (+/- 13%) biomes, and the smallest in temperate biomes (+/- 7%)
281 (Extended Data Table 3). Countries with well-established national forest inventories based on
282 statistical sampling had the lowest reported uncertainty. Thus, additional ground measurements
283 and monitoring are especially needed in tropical biomes and countries that currently lack
284 statistical sampling; in soils and dead wood globally; and in areas affected by natural
285 disturbances and logging. For future global analyses based on bottom-up approaches, we
286 recommend several research and monitoring priorities:

287 1. Increased sampling of belowground biomass, dead wood, litter, and soil C. These have
288 much greater uncertainties than aboveground biomass, although smaller impacts on the total
289 uncertainty except boreal forests. For instance, if we had increased soil sink uncertainties to
290 100% in all biomes (Table S5), globally it would only increase uncertainty in the total C sink by
291 about 1% because sinks in living biomass are the dominant components. Along with increased
292 field measurements, scaling up to the region and biome should employ detailed forest type maps
293 to represent the distinct and variable forest conditions that comprise the total forest areas.

294 2. Increased research and sampling of underrepresented tropical forests, such as
295 Southeast Asian wetland forests and African dry forests, could be combined with better forest
296 type maps to mitigate potential biases from uneven sampling. This would require broad-scale
297 support and investment in long-term on-the-ground monitoring of tropical forest biomass, growth
298 and mortality, distributed across all tropical forest types. The enhanced land monitoring would
299 complement and greatly leverage investments in space-based forest monitoring, and reduce

300 uncertainties in data about changes and climate sensitivities of Earth's most productive and
301 diverse biomes.

302 3. Better information about uncertainty of forest area estimates which mostly rely on
303 remote sensing or remote-sensing based forest inventory statistics and are often reported without
304 uncertainty information⁴⁵. Uncertainties in forest areas are caused by inconsistent remote-sensing
305 data processing methods and definitions of forests, and make up a considerable proportion of the
306 uncertainty of C sink estimates.

307

308 **Enhancing the forest C sink to help attain global C neutrality**

309 Our results suggest that the single most important action for sustaining and increasing the
310 forest C sink is to stop emissions from deforestation and degradation, along with protecting the
311 large C stocks that have accumulated over centuries especially in boreal forest soils. Recovery of
312 functions by degraded forests and lands offers additional opportunities for enhancing C sinks
313 with many co-benefits such as protecting biodiversity⁴⁶. The pathways for stopping global
314 deforestation and degradation will rely on international cooperation such as UN's REDD+
315 program. Financial, legislative and other incentives are needed particularly in tropical countries.
316 Deforestation-free supply chains and well-managed selective logging can all lower deforestation
317 rates.

318 Our study demonstrates considerable impacts of large-scale reforestation and
319 afforestation on enhancing C sinks, either through natural recovery or mandated actions. Some
320 countries, such as the U.S., have lands suitable for afforestation or improved management, but
321 historically low adoption rates (Extended Data Table 1). Tropical forest regrowth represents
322 another significant opportunity to accumulate additional C on abandoned land. Declining C sink

323 strength due to forest aging has become more common in some temperate zones^{31,34}, although
324 most older forests maintain high C stocks in the absence of human disturbances and some remain
325 productive for very long times⁴³. In the future, management intensity and its effects on forest age
326 dynamics may determine C sink trends of temperate forests.

327 Strategic planning will help to prioritize forest management approaches to minimize C
328 emissions and maximize C uptake and co-benefits. For instance, adaptive and climate smart
329 forestry practices⁵ such as reduced-impact logging⁴⁷, fuel management to increase resistance to
330 wildfires⁴⁸, optimizing tree species resilience after disturbances, and restoring old-growth
331 characteristics can be highly effective⁴⁹. Protecting C stocks is also essential. For example, our
332 data show that tropical regrowth forests have high C sequestration rates but their recovering C
333 densities take many decades to reach intact forest levels. So, replacing intact forests with
334 regrowth forests having large C sinks but much lower C stocks and diminished biodiversity is
335 highly imprudent.

336 Since long-lived HWPs store C but only represent ~10% of C in harvested timber,
337 switching from short-lived products like fuelwood or pulpwood to long-lived sawnwood
338 products could sequester additional C, provided total harvest volume does not increase and
339 reduce ecosystem C stocks. Improving wood processing technologies to reduce waste⁴⁷,
340 developing new long-lived materials, and recycling⁵⁰ may benefit a sustainable and circular
341 economy as suggested by the IPCC⁵. Our estimates indicate 107 Pg C were sequestered from the
342 atmosphere by global forests since 1990, equal to 46% of fossil fuel emissions. While 63% of
343 this uptake was negated by tropical deforestation, the remaining forests helped slow climate
344 change. The global forest sequestration rate of ~3.56 Pg C year⁻¹ (~13 Gt CO₂-eq year⁻¹) for
345 1990-2019 provides a baseline for the IPCC's ambitious assessment⁴ that Agriculture, Forestry

346 and Other Land Use sectors have a combined potential to mitigate an additional 8-14 Gt CO₂-eq
347 year⁻¹ during 2020-2050. Mitigating and adapting to the climate crisis are defining challenges for
348 humanity, and these goals cannot be achieved without both protecting the carbon stocks and
349 sinks in Earth's forests and reducing emissions from fossil fuels.

References

1. IPCC AR6 WGII, “Climate Change 2022: Impacts, Adaptation and Vulnerability. Contribution of Working Group II to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change” (Cambridge University Press, Cambridge, UK and New York, NY, USA, 2022),
2. Friedlingstein, P. et al. Global Carbon Budget 2021. *Earth Syst. Sci. Data* **14**, 1917–2005 (2022).
3. Peng, L. et al. The carbon costs of global wood harvests. *Nature* **620**, 110–115 (2023). <https://doi.org/10.1038/s41586-023-06187-1>.
4. NOAA, <https://gml.noaa.gov/ccgg/trends/> (2023).
5. Nabuurs, G.J. et al. “Chapter 7: Agriculture, Forestry and other Land Uses”. in *the IPCC Working Group III Report to the Sixth Assessment* (Cambridge University Press) (2022).
6. McKay, D.I.A. et al. Exceeding 1.5°C global warming could trigger multiple climate tipping points. *Science* **377(1126)**, DOI: 10.1126/science.abn7950 (2022).
7. United Nations / Framework Convention on Climate Change. *Adoption of the Paris Agreement, 21st Conference of the Parties, Paris, United Nations* (2015).
8. Fuss, S. et al. Negative emissions—Part 2: Costs, potentials and side effects. *Environ. Res. Lett.* **13**, 063002 (2018).
9. IPCC WG I, “Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change” (Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2018).
10. Walker, W.S. et al. The global potential for increased storage of carbon on land. *PNAS* **119(23)**, e2111312119 (2020).
11. Pan, Y. et al. A large and persistent carbon sink in the world’s forests. *Science* **333**, 988-993 (2011).
12. Araza, A. et al. Past decade above-ground biomass change comparisons from four multi-temporal global maps. *Inter. J. Appl. Earth Observation and Geoinformation* **118**, 103274 (2023).
13. Suarez, D.R. et al. Estimating aboveground net biomass change for tropical and subtropical forests: Refinement of IPCC default rates using forest plot data. *Glob Change Biol.* **25**, 3609–3624 (2019).

14. IPCC. *Climate Change: The IPCC 1990 and 1992 Assessments*. (https://www.ipcc.ch/site/assets/uploads/2018/05/ipcc_90_92_assessments_far_full_report.pdf) (1992).
15. Schimel, D., Stephens, B.B. & Fisher, J. B. Effect of increasing CO₂ on the terrestrial carbon cycle. *PNAS* **112** (2), 436-441 (2014).
16. Sitch, S. et al. Recent trends and drivers of regional sources and sinks of carbon dioxide. *Biogeosciences* **12**, 653–679 (2015).
17. Pan, Y. et al. Contrasting responses of woody and grassland ecosystems to increased CO₂ as water supply varies. *Nat. Ecol. Evol.* **6**, 315-323 (2022).
18. Walker, A.P. et al. Integrating the evidence for a terrestrial carbon sink caused by increasing atmospheric CO₂. *New Phytologist* **229**, 2413–2445 (2021).
19. Shvetsov, E.G. Kukavskaya, E.A., Shestakova, T.A., Laflamme, J. & Rogers, B.M. Increasing fire and logging disturbances in Siberian boreal forests: a case study of the Angara region. *Environ. Res. Lett.* **16**, 115007 (2021).
20. Fan, L. et al. Siberian carbon sink reduced by forest disturbances. *Nature Geoscience* **10**, 1038/s41561-0 22-01087-x (2022).
21. Wang, J.A., Baccini, A., Farina, M., Randerson, J.T. & Friedl, M.A. Disturbance suppresses the aboveground carbon sink in North American boreal forests. *Nat. Clim. Chang.* **11**, 435–441 (2021).
22. Kurz, W.A. et al. Carbon in Canada’s boreal forest - A synthesis. *Environ. Rev.* **21** (4), (2013).
23. Dai, A. Increasing drought under global warming in observations and models. *Nature Clim Change* **3**, 52–58 (2013).
24. Gutman, G. & Radeloff, V.C. (Eds). *Land use and land cover change in Eastern Europe after the collapse of the Soviet Union in 1991* (Springer, Dordrecht, 2016).
25. Walker, X. J. et al. Increasing wildfires threaten historic carbon sink of boreal forest soils. *Nature* **572**, 520–523 (2019).
26. Kauppi, P. E. et al. Managing existing forests can mitigate climate change. *For. Ecol. Manag.* **513**, 10.1016/j.foreco.2022.120186 (2022).
27. Henttonen, H.M., Nöjd, P., Mäkinen H. Environment-induced growth changes in forests of Finland revisited - a follow-up using an extended data set from the 1960s to the 2020s. *For. Ecol. Manag.* **551**, <https://doi.org/10.1016/j.foreco.2023.121515> (2024).
28. Korosuo, A. et al. The role of forests in the EU climate policy: are we on the right track?

- Carbon Balance Manage* **18**, 15 (2023). <https://doi.org/10.1186/s13021-023-00234-0>.
29. Yang, C. et al. Updated estimation of forest biomass carbon pools in China, 1977–2018. *Biogeosciences* **19**, 2989–2999 (2022).
 30. Domke, G.C. et al. “Chapter 9: Forests”, in *Second State of the Carbon Cycle Report (SOCCR2): A Sustained Assessment Report*. N. Cavallaro, G. Shrestha, R. Birdsey, M. A. Mayes, R. G. Najjar, S. C. Reed, P. Romero-Lankao, Z. Zhu, Eds. U.S. Global Change Research Program, Washington, DC, USA, pp. 365-398, <https://doi.org/10.7930/SOCCR2.2018.Ch9> (2018).
 31. Nabuurs, G.J. et al. First signs of carbon sink saturation in European forest biomass. *Nat. Clim. Chan.* **3**, 792 (2013).
 32. Hlásny, T. et al. “Living with bark beetles: impacts, outlook and management options. From Science to Policy 8” (European Forest Institute, 2019).
 33. Salomón, R.L. et al. The 2018 European heatwave led to stem dehydration but not to consistent growth reductions in forests. *Nat Commun* **13**, 28 (2022).
 34. MAFF (Ministry of Agriculture, Forestry and Fisheries), “State of Japan’s Forests and Forest Management: 3rd Country Report of Japan to the Montreal Process, 2019” (<https://www.maff.go.jp/e/policies/forestry/attach/pdf/index-8.pdf>, 2019)
 35. Vijay, V., Pimm, S.L. Jenkins, C. N. & Smith, S. J. The impacts of oil palm on recent deforestation and biodiversity loss. *PLoS ONE* **11** (7), e0159668 (2016).
 36. Hubau, W. et al. Asynchronous carbon sink saturation in African and Amazonian tropical forests. *Nature*, **579** (7797), 80-87 (2022).
 37. Phillips O.L. et al. Drought sensitivity of the Amazon rainforest. *Science* **323**, 1344–47 (2009).
 38. Lewis, S.L. et al. . The 2010 Amazon drought. *Science* **331** (2011).
 39. IPCC, “Refinement to the 2006 IPCC guidelines for national greenhouse gas inventories” (<https://www.ipcc.ch/report/2019-refinement-to-the-2006-ipcc-guidelines-for-national-greenhouse-gas-inventories/>, 2019).
 40. Brienen, R.J.W. et al. Long-term decline of the Amazon carbon sink. *Nature* **519**, 344–348 (2015).

41. Intergovernmental Panel on Climate Change, *Climate Change 2014—Impacts, Adaptation and Vulnerability: Regional Aspects*. (Cambridge University Press. doi: 10.1017/CBO9781107415416 (2014).
42. Phillips, C.A. et al. Escalating carbon emissions from North American boreal forest wildfires and the climate mitigation potential of fire management. *Sci. Adv.* **8**, eabl7161 (2022).
43. Law, B.E. et al. Land use strategies to mitigate climate change in carbon dense temperate forests. *Proc. Nat. Acad. Sci.* **115(14)**, 3663-3668, <https://doi.org/10.1073/pnas.1720064115> (2018).
44. Friedlingstein, P. et al. Global Carbon Budget 2022. *Earth Syst. Sci. Data* **14**, 4811–4900 (2022).
45. Henson, M.C. et al. High-resolution global maps of 21st-Century forest cover change. *Science* **342**, 850-853 (2013).
46. Almut, A. et al. Restoring Degraded Lands. *Ann. Rev. Envir. Resour.* **46**, 569-599 (2021)
47. Sasaki, N. et al. Sustainable Management of Tropical Forests Can Reduce Carbon Emissions and Stabilize Timber Production. *Front. Environ. Sci.*, doi: 10.3389/fenvs.2016.00050 (2016).
48. Hurteau, M.D., North, M.P., Koch, G.W. & Hungate, B.A. Managing for disturbance stabilizes forest carbon. *PNAS* **116 (21)**, 10193–10195 (2019).
49. Thom, D. et al. The climate sensitivity of carbon, timber, and species richness covaries with forest age in boreal–temperate North America. *Global Change Biology* **25(7)**, 2446-58 (2019).
50. Birdsey, R. et al. Climate, Economic, and Environmental Impacts of Producing Wood for Bioenergy. *Environmental Research Letters* **13**, 050201 (2018).

Figure Legends

Figure 1. Carbon sinks/sources (Pg C year^{-1}) in the world's forests through the decades. Green bars represent established forests (boreal, temperate, and tropical intact forests), while brown bars represent tropical regrowth forests. Positive values (with downward bars) indicate C sinks, while negative values (with upward bars) show C sources. Detailed uncertainties of sink/source are shown in Extended Data Table 3. We grouped a few regions/countries shown in the detailed table to fewer categories in order to keep the graphic from getting too cluttered, including Europe (Europe temperate + Other Europe), Japan/Korea (Japan + Korea), South Asia (India + Other South Asia), and Mexico/Central America (Mexico + Central America) (see Extended Data Table 3).

Figure 2. Comparison and integration of inventory-based global forest carbon sink (-) and source (+) (Pg C year^{-1}) estimates with those from the Global Carbon Budget (GCB)². E_{FOS} , S_{OCEAN} , E_{GLUC} , S_{GLAND} , and S_{NLAND} (blue) were estimated by or derived from GCB. S_{GFOR} , E_{DFOR} , S_{NFOR} (black) and differences ($E_{\text{GLUC}} - E_{\text{DFOR}}$, $S_{\text{GLAND}} - S_{\text{GFOR}}$) were estimated in this study. Our values for the global forest gross sink (S_{GFOR}) and tropical deforestation gross emissions (E_{DFOR}), when compared to GCB total land estimates, provide new ground-inventory constraints with which to derive global *non-forest land* LULC gross emissions and gross sink estimates for each decade since 1990 (see Extended Data Table 4 for further details).

1 **Methods**

2 ***Forest biomes and lands***

3 Estimates of C stocks and stock changes are reported for forests partitioned into three biomes:
4 boreal, temperate, and tropical (including subtropical); and by carbon component (living
5 biomass, dead wood, litter, soil, and harvested wood products). Forests in boreal and temperate
6 biomes include both “forest land remaining forest land” and reforested or afforested lands
7 (collectively “new forests”), while tropical forests are separated into remaining forests (intact
8 forests) and regrowth forests. The area of global forests used as a basis for estimating C stocks
9 and sinks was ~4.0 billion hectares, representing 95% of the global forested lands⁵¹ (Extended
10 Data Table 1, Table S6). The 5% not covered are some remote forest areas including unmanaged
11 forests in northern Canada and Alaska Interior, and some areas of west/central Asia with sparse
12 forests, where we lacked credible ground data. We do not include non-forested peatlands or
13 wetlands, or coastal mangrove forests which commonly contain high C in soil or sediments⁵².

14

15 ***Definitions of forest carbon pools and stocks***

16 We generally followed the definitions from Table 3.1.2 in the IPCC Good Practice Guidance⁵³.
17 Definitions of five main carbon pools are detailed in the Supplement: living biomass, dead wood,
18 litter, soil organic matter, and harvested wood products.

19

20 *Carbon stock* – carbon contained in different carbon pools or in all carbon pools.

21

22 *Carbon stock change (or C flux)* – change in carbon stocks between time points, which can
23 represent carbon gain (sink) or carbon loss (source).

24

25 ***Overview of data and calculation methods***

26 Sources of data used in this study

27 This study covers three decades (1990s, 2000s and 2010s) with available data from 1990 to 2020
28 (Table S1a, S1b, S1c). Since our last study (for 1990-2007)¹⁰, country-scale greenhouse gas
29 (GHG) inventories in temperate and boreal countries/regions have expanded to include more
30 countries and have been updated. Networks of sample plots in tropical regions of Amazonia,
31 Africa, and Southeast Asia have expanded. Our data are not always consistent with what

32 individual countries report to FAO or IPCC. We use FAO data as reported in FRA 2020⁵¹ to
33 establish the total forest area by country or region. These data are a credible source for trend
34 information about forest area over decades and across geographies.

35

36 Accounting approaches to calculations for different forest regions/biomes

37 There are slightly different accounting approaches used in this paper because the available data
38 have been developed and presented in different ways in inventories, country reports, and the
39 literature. Estimates were harmonized between accounting systems by carefully defining land
40 areas and matching these with the sources of data, and by adjusting reported estimates where
41 necessary to account for known inconsistencies. Our calculation methods are summarized in
42 Table S1 and described in more detail Supplementary Information.

43

44 Either the “**stock-change**” or the “**default**” approaches were used for boreal and temperate
45 biomes, following the guidance from IPCC^{53,54}. The stock change approach was also applied for
46 several tropical countries or regions (only intact forests) including India, other South Asia
47 countries, Mexico, Central America and Caribbean. If there is no land-use change, then the
48 stock-change approach is nearly identical to estimating the land-atmosphere CO₂ flux, with the
49 exception of “lateral transfers” of C which primarily include river erosion, transport, outgassing,
50 and deposition; and harvested wood products. One exception is Canada, which reports C stock
51 changes based on the “**gain-loss**” approach. The default approach commences with a single
52 forest inventory and then adds C gains from forest growth and losses from harvest, fires and
53 decomposition without confounding estimates through C transfers between land-use categories⁵³.

54

55 We accounted for harvested wood products (HWP) but not for other lateral transport, which may
56 be responsible for a significant global C sink into coastal oceans from forests that is not reflected
57 in the stock-change method. If there is land-use change, then the stock-change accounting
58 overestimates the C uptake by forests in proportion to the area of afforestation during the period
59 of change, because existing C stocks on new forest land (primarily soil C) appear instantaneously
60 in the forest carbon inventory, transferred from the previous land use category. Conversely, the
61 stock-change approach may underestimate C uptake by forests in proportion to the area of

62 deforestation because existing soil C may be moved to a non-forest land category and appear as a
63 loss of C from forest. We corrected for this apparent loss in our accounting.

64

65 For the tropics (Southeast Asia, Africa and South America), C sinks and sources (or net fluxes)
66 were estimated using a “**flow**” approach because most tropical areas lack the repeated national-
67 scale forest inventories that are the basis for the stock-change approach. This approach is similar
68 to the IPCC “tier 2” methods⁵³ that multiply region-specific estimates of C density or change in
69 C density times the associated areas represented by the region-specific estimates. For intact
70 tropical forests (not affected by land use or change), fluxes were estimated from measured C
71 stock changes on permanent sample plots, which is nearly equivalent to forest-atmosphere C
72 exchange except for river transport and deposition of C. The approach allows accounting for C
73 gains in forests, including some impacts of forest degradation affecting rates of C gains, but not
74 C losses due to deforestation because C stored in deforested areas is accounted separately in our
75 global budget (Extended Data Fig. 1).

76

77 The effects of land-use change and harvesting on C flux were estimated separately using a
78 **bookkeeping** approach⁵⁵ that keeps track of ecosystem C emissions and harvested wood
79 products from deforestation and logging, and ecosystem C uptake on regrowing forests.

80

81 Estimates of C stock changes pertain to “forest land remaining forest land” plus “afforested land”
82 for boreal and temperate forests. For tropical intact and regrowth forests of Southeast Asia,
83 Africa, and South America, and also for tropical regrowth forests of Mexico and Central
84 America/Caribbean, changes in C density times the associated areas were used. Estimates of C
85 stocks for specific years (Extended Data Table 2) pertain to the total area of forest land in the
86 given year and therefore include C stocks lost because of deforestation, which are not included in
87 Extended Data Table 3. Thus, it is not possible to consistently match the estimates between these
88 two tables, which is particularly true for tropical intact forests – the only biome that has lost
89 substantial forest area (Extended Data Table 1).

90

91 Forest area and area change

92 Area estimates (Extended Data Table 1) are from country-level forest inventories or reports
93 based on forest inventories. Forest inventories typically use remote sensing combined with
94 ground observations to estimate forest area and area changes following FAO forest definitions,
95 excluding “other wooded land”. Where forest inventory data direct from countries are lacking,
96 particularly in the tropics, FAO statistics were used to estimate total forest area for 1990, 2000,
97 2010, and 2020⁵¹. In some regions, particularly the tropics and Russia, the quality of data
98 reported to FAO is poor and the protocols may be subject to change over time. Because tropical
99 intact forests defined in this study are not the same as primary forests defined in FAO statistics
100 (see the definition in Extended Data Box 1), we used the area estimates of tropical intact forests
101 from published studies for Southeastern Asia, Africa, and South America³⁵. The difference
102 between total tropical forest from the FAO⁵¹ and the area of tropical intact forest for these
103 regions was assumed to be the area of tropical regrowth forest. We attempted to establish good
104 consistency between the change in reported areas from the years of 1990, 2000, 2010, and 2020,
105 and estimated areas of afforestation and deforestation from inventories, country reports, and
106 analyses of emissions from land-use changes.

107

108 Carbon stocks and carbon stock changes

109 Where available, C stock and density estimates are from country-level forest inventories or
110 reports based on national forest inventories (NFI). Most countries in temperate and boreal
111 biomes have established NFIs with repeated measurement of permanent sample plots. Generally,
112 sample plots are randomly located across all areas of the country, and measurements taken on
113 those plots that are located on forest land. Thus, the inventory is an unbiased sample of the
114 population of trees in the country, and the precision of estimates may be calculated. The re-
115 measurement interval is typically between 5 and 10 years. At each sample plot, individual trees
116 are selected for measurement of diameter, height, species, and condition. Re-measurement
117 determines the basic tree population dynamics: growth, mortality, and harvest. Additional
118 measurements may be taken to include understory vegetation, woody debris, litter, and soils. For
119 some temperate or boreal countries where direct access to inventory data is not available, we
120 used a biomass expansion factor (BEF) approach, which converts estimates of growing stock
121 volume to estimates of biomass or C stocks. The measured data may be used to estimate the C
122 stocks and C stock changes using a variety of country-specific methods (described in

123 Supplementary Information), but generally following guidelines provided by IPCC^{53,54}. For
124 example, the basic tree measurements of diameter and height are used to estimate tree biomass
125 and carbon using allometric models and conventional statistical methods.

126
127 For tropical intact forests of Southeast Asia, Africa, and South America, we used data from
128 repeated long-term measurements of networks of ecological research plots, upscaled to the
129 regions to estimate biomass and other C pools for the region's forest areas^{35,36,37}. For tropical
130 regrowth forests, which lack sufficient ground-based data, we followed the bookkeeping
131 approach⁵⁶ which is based on a literature review of regrowth rates and C stocks and knowledge
132 of forest areas and conditions, averaged over different ecozones (tropical wet, moist and dry
133 forests) for each region⁵⁷. These methods are described in more detail in Supplementary
134 Information.

135
136 The data from regions, countries or continents were aggregated to global biomes: boreal,
137 temperate, and tropical forests. For countries and regions that do not allow access to original
138 data, the data from the FRA regional reports⁵¹ were used to fill the data gaps (Table S1b).
139 Available data allowed C stock and area estimates to be compiled for 1990, 2000, 2010, and
140 2020, and annual changes in C stocks (sometimes referred to as “sink” if there was a C gain, and
141 as “source” if there was a C loss) to be estimated for three time periods: 1990-1999, 2000-2009,
142 and 2010-2019.

143
144 More data are available for live biomass and biomass changes than for other C pools. Some
145 forest inventories and many ecological studies also collect and report data for dead wood and
146 litter, though less consistently than for biomass; therefore, empirical models are often the source
147 of estimates for these C pools. Inventories of forest soil carbon across the landscape are scarcer
148 than inventories of biomass or other ecosystem C pools, and sampling methods include varying
149 soil depths for sampling among regions and countries. There are existing soil surveys in different
150 countries, but very rarely with periodic revisits and rarely associated with documented
151 information about aboveground forest vegetation. To evaluate forest soil C change over time is
152 particularly challenging because the formation and respiration of soil C is affected by various
153 biological, environmental and geographical factors; and land-use history; and not always

154 correlated with more easily observable vegetation traits. In almost every region, empirical
155 modeling methods were used to combine data from soil surveys and field studies for developing
156 soil C estimates.

157

158 Harvested wood products (HWP)

159 HWP is defined as a component of the C sink in this study and included in the C stock change
160 category. Where available, estimates of carbon in HWP are from country level inventory reports.
161 Otherwise, harvested roundwood data were derived from FAO annual statistics (see
162 Supplementary Information). Generally, estimates of carbon in HWP account for its temporary
163 exclusion from the atmosphere, which includes both the wood products in use and discarded
164 wood products remaining in landfills or dumps. For countries that lacked reported estimates of
165 HWP, we derived a simple conversion factor from the countries that did report: the ratio of C
166 flux in HWP (Tg C year^{-1}) to the quantity of harvested roundwood (million m^3) according to
167 FAO reports⁵¹, which is 0.095.

168

169 *Approaches to estimate uncertainty*

170 We report the Standard Error (SE) for estimates of C stocks and changes in C stocks, using the
171 95% confidence level. Values presented as “ $y \pm x$ ” should be interpreted to mean that the authors
172 are 95% certain the actual value is between $y - x$ and $y + x$. The 95% boundary was chosen to
173 communicate the high degree of certainty that the actual value was in the reported range and the
174 low likelihood (5% or less) that it was outside that range. This characterization is not, however, a
175 statistical property of the estimate, and should not be confused with statistically defined 95%
176 confidence intervals.

177

178 We report uncertainty using two approaches depending on the availability of uncertainty
179 estimates from data sources: quantitative estimates and expert opinion. Quantitative estimates are
180 based on remote sensing and sampling combined with empirical models, using either error
181 propagation methods or Monte Carlo simulation approaches to combine all C pools together, and
182 including the uncertainty of area estimates. The expert opinion approach is based on that adopted
183 by IPCC for reporting in global assessments (described in Supplemental Information).

184 Quantitative estimates are more commonly available for data derived from national forest

185 inventories or extensive sampling plot networks, whereas expert opinion is used where
186 quantitative estimates are unavailable, a method which has been used in previous large-scale
187 analyses⁵⁸— see Supplementary Information for details. In applying these approaches, we ensured
188 that estimates based on expert opinion were not overly optimistic compared with estimates from
189 similar countries or regions that reported quantitative estimates.

190

191 *Evaluating major uncertainties in different biomes and C components*

192 We reported uncertainties for the aggregated sums of individual C pools (such as litter and
193 deadwood) but not for each individual pool because this detailed information is not regularly
194 included in the publicly available estimates, even though the uncertainty of each individual C
195 pool is included in the aggregated estimates of carbon stocks and stock changes that are
196 estimated using error propagation approaches.

197

198 Uncertainty estimates for stock change in boreal forests are $\sim\pm 13\%$ and possibly more
199 considering uncertainty of soil C estimates. The largest stock change by far is in Russian boreal
200 forests, and the uncertainty is particularly significant because of the large sink estimated in this
201 region. The main reasons for the uncertainty of boreal forest estimates involve incomplete
202 sampling of large areas of Alaska, Canada, and Russia, combined with poor data on soil C and
203 wildfires particularly in the Asian part of Russia.

204

205 Uncertainty estimates for stock change in temperate forests are about $\pm 7\%$ representing the
206 lowest value among all biomes. This is mainly because most temperate countries have strong and
207 repeated forest inventory sampling programs that cover most of the forest area. The greatest
208 uncertainty in temperate forests is for changes in soil carbon, which is not monitored as easily or
209 as often as the other carbon pools.

210

211 Uncertainty estimates for stock change in tropical intact forests are about $\pm 27\%$ in the most
212 recent period, largely because the estimates are based on a relatively small number of intensively
213 monitored sites whose data are individually quite accurate but not conducive to scaling because
214 representation of the larger population of forests by the collection of sites is unknown. This
215 uncertainty is particularly notable because the largest component of the global forest C sink is in

216 tropical forests. The effects of disturbances, particularly drought, are difficult to quantify, and
217 there is relatively little data about the C pools other than live biomass.

218
219 Uncertainty estimates for stock change in tropical regrowth forests are about +/- 20%, a little
220 lower than estimates for intact forests. The area of tropical regrowth forests is not well known,
221 and there is relatively little sampling done. The error estimates, based on expert opinion,
222 probably underestimate the true uncertainty of this increasingly important component of the
223 global budget.

224
225 The uncertainties of stock-change estimates for soil C, dead wood, litter, and HWP are high in
226 boreal regions and the tropics. However, the size of the sink in these pools is relatively small
227 compared with living biomass, except boreal forests, so the contribution to overall uncertainty is
228 also small. As shown by uncertainty experiments (Table S5), while ignoring soil C sinks would
229 reduce the estimated global forest C sink by $\sim 400 \text{ Tg C year}^{-1}$, it would have minimal impact on
230 the global and biome-level temporal trends. Increasing 100% uncertainties in soil sinks, the total
231 C sinks in boreal, temperate and tropical forests increased their uncertainties by 15%, 2% and
232 $<1\%$, respectively, yet with error propagation it increased uncertainty in total global C sink by \sim
233 1%.

234
235 Additional sources of uncertainty are described in the Supplementary Information.

236 237 *Assessing our approach vs. modeling/remote-sensing approaches*

238 Remote-sensing and modeling estimates of the forest sector are subject to significant
239 uncertainties and inconsistency between different studies^{59,60,61}, compared with ground data that
240 are based on more standard definitions and protocols^{51,53,54}. Different representations and
241 complexity of regional ecological processes, and limited calibration with data for
242 parameterization are often the cause for inconsistencies in model results^{62,63}. Indeed, remote
243 sensing and modeling approaches are dependent on summarized “standard” per-hectare biomass
244 estimates derived from field studies. Ground data have improved significantly, and multiple
245 carbon pools are more often measured and monitored. Our estimates represent a credible
246 complement to the remote-sensing and model-based estimates used for the land part of the

247 Global Carbon Budget^{1,43}, with terrestrial data in GCB being based on an average of models^{62,64}.
248 It is important to use multiple methods to contrast and compare calculations in order to improve
249 overall estimates of land C sinks.

250

251

252 **Data availability**

253 Data used for synthesis and analysis in this paper are derived from more detailed measurements
254 and are fully contained in the spreadsheets with embedded formulas for access
255 (<https://doi.org/10.2737/RDS-2023-0051>). Our results can be replicated beginning with these
256 spreadsheets. The estimates used for tables and figures of the main text and Extended Data are
257 also in the data repository. The data repository also includes original measurements of a few
258 countries and the source data information for others with DOIs and websites for accessing
259 original data. Because policies for data sharing vary from country to country, some sources
260 include original measurement data from sampling with fully open access, while some do not
261 include original data but rather aggregated data. Most original data are publicly available through
262 direct access, while in a few cases where the data is not publicly available, requests from
263 regional authors are needed. Full descriptions of regional datasets and estimation approaches,
264 including links, are provided in Supplementary Information.

References of Method

51. Food and Agriculture Organization (FAO), “*Global Forest Resource Assessment 2020: Main Report*” (Rome. <https://doi.org/10.4060/ca9825en>) (2020).
52. Murdiyarso, D., Kauffman, J.B. & Verchot, L. Climate change mitigation strategies should include tropical wetlands. *Carbon Management* **4(5)**, 509-517 (2013).
53. IPCC. *Good practice guidance for land use, land-use change, and forestry* (IGES, Japan), <http://www.ipcc-nggip.iges.or.jp/public/gpplulucf/gpplulucf.html>, (2003).
54. IPCC. *IPCC guidelines for national greenhouse gas inventories* (IGES, Japan, <http://www.ipcc-nggip.iges.or.jp/public/2006gl/index.html>) (2006).
55. Houghton, R.A. & Castanho, A. Annual emissions of carbon from land use, land-use change, and forestry 1850–2020, *Earth Syst. Sci. Data Discuss.* [in press], <https://doi.org/10.5194/essd-2022-351>, in review (2022).
56. Houghton, R.A. Terrestrial fluxes of carbon in GCP carbon budgets. *Global Change Biology* DOI: 10.1111/gcb.15050 (2020).
57. Cook-Patton, S. et al, Mapping potential carbon capture from global natural forest regrowth. *Nature* **585**, 545–550 (2020).
58. CCSP. *The First State of the Carbon Cycle Report (SOCCR): The North American Carbon Budget and Implications for the Global Carbon Cycle*, A. W. King, L. Dilling, G.P. Zimmerman, D.M. Fairman, R.A. Houghton *et al* Eds. (U.S. Climate Change Science Program and the Subcommittee on Global Change Research, National Oceanic and Atmospheric Administration, National Climatic Data Center, Asheville, NC, USA (2007).
59. Lang, N. et al. Global canopy height regression and uncertainty estimation from GEDI LIDAR waveforms with deep ensembles. *Remote Sensing of Environment* **268**, 112760 (2022).
60. Bastos, A. et al. Sources of uncertainty in regional and global terrestrial CO₂ exchange estimates. *Global Biogeochemical Cycles*, 34 (2020).
61. Ciais, P. et al. Definitions and methods to estimate regional land carbon fluxes for the second phase of the Regional Carbon Cycle Assessment and Processes Project (RECCAP-II). *Geoscientific Model Development*, **15(3)**, 1289-1316. (2022).
62. O’Sullivan, M.. et al. Process-oriented analysis of dominant sources of uncertainty in the land carbon sink. *Nat Commun* 13, 4781 (2022).
63. Fatichi, S., Pappas, C, Zscheischler, J. & Leuzinger, S. Modelling carbon sources and sinks in terrestrial vegetation. *New Phytologist* **221**, 652–668 (2019).

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Author contributions

Y.P. and R.A.B. were lead authors, synthesized the data and drafted the manuscript; O.L.P, R.A.H. J.F. provided critical concepts and substantial editing; Y.P, R.A.B., O.L.P, R.A.H., J.F. P.K, H.K., W.A.K., A.I., S.L.L, G.-J.N, A.S. contributed primary datasets and analyses, led regional estimates and writing of the methods; S.H., B.L, A.C., D.S., and D.M. contributed regional estimates and methodology documentation. All authors contributed in writing, discussions, or comments.

Table 1. The global forest carbon sinks and source (Pg C year⁻¹) over 3 decades from 1990 to 2019.

Carbon sink and source in biomes	1990-1999	2000-2009	2010-2019	1990-2019 (mean)	1990-2019 (total, Pg C)
Boreal forest	0.51 ± 0.06	0.49 ± 0.05	0.32 ± 0.04	0.44 ± 0.05	13.18 ± 0.29
Temperate forest	0.53 ± 0.04	0.59 ± 0.04	0.68 ± 0.05	0.60 ± 0.04	18.02 ± 0.24
Tropical intact forest	1.28 ± 0.20	1.03 ± 0.19	0.88 ± 0.24	1.07 ± 0.21	31.95 ± 1.15
Tropical regrowth forest	1.27 ± 0.26	1.46 ± 0.29	1.64 ± 0.33	1.46 ± 0.30	43.72 ± 1.62
C sink in World forests ⁴	3.59 ± 0.34	3.57 ± 0.36	3.53 ± 0.41	3.56 ± 0.37	106.88 ± 2.02
Global established forests (excluding tropical regrowth)	2.32 ± 0.21	2.11 ± 0.20	1.89 ± 0.24	2.11 ± 0.22	63.15 ± 1.21
Tropical intact forest	1.28 ± 0.20	1.03 ± 0.19	0.88 ± 0.24	1.07 ± 0.21	31.95 ± 1.15
Tropical regrowth forest	1.27 ± 0.26	1.46 ± 0.29	1.64 ± 0.33	1.46 ± 0.30	43.72 ± 1.62
All tropical forests	2.56 ± 0.33	2.49 ± 0.35	2.52 ± 0.41	2.52 ± 0.36	75.68 ± 1.99
Tropical deforestation gross emissions	-2.66 ± 0.53	-1.91 ± 0.43	-2.13 ± 0.56	-2.24 ± 0.51	-67.05 ± 2.79
Global forest net C sink	0.93 ± 0.63	1.66 ± 0.56	1.39 ± 0.69	1.33 ± 0.63	39.83 ± 3.45
Equations of global forest C fluxes:					
$F_{GLOBAL\ GROSS\ FOREST\ SINK} = F_{BOREAL} + F_{TEMPERATE} + F_{TROPICAL\ INTACT} + F_{TROPICAL\ REGROWTH}$					(Eq. 1.1)
$F_{ESTABLISHED\ FORESTS} = F_{BOREAL} + F_{TEMPERATE} + F_{TROPICAL\ INTACT}$					(Eq. 1.2)
$F_{ALL\ TROPICAL\ FORESTS} = F_{TROPICAL\ INTACT} + F_{TROPICAL\ REGROWTH}$					(Eq. 1.3)
$F_{GLOBAL\ FOREST\ NET\ SINK} = F_{GLOBAL\ FOREST\ GROSS\ SINK} + F_{TROPICAL\ DEFORESTATION\ GROSS\ EMISSION}$					(Eq. 1.4)

Notes: the definitions of forest biomes and C fluxes in the table and equations refer to Extended Data Box 1.

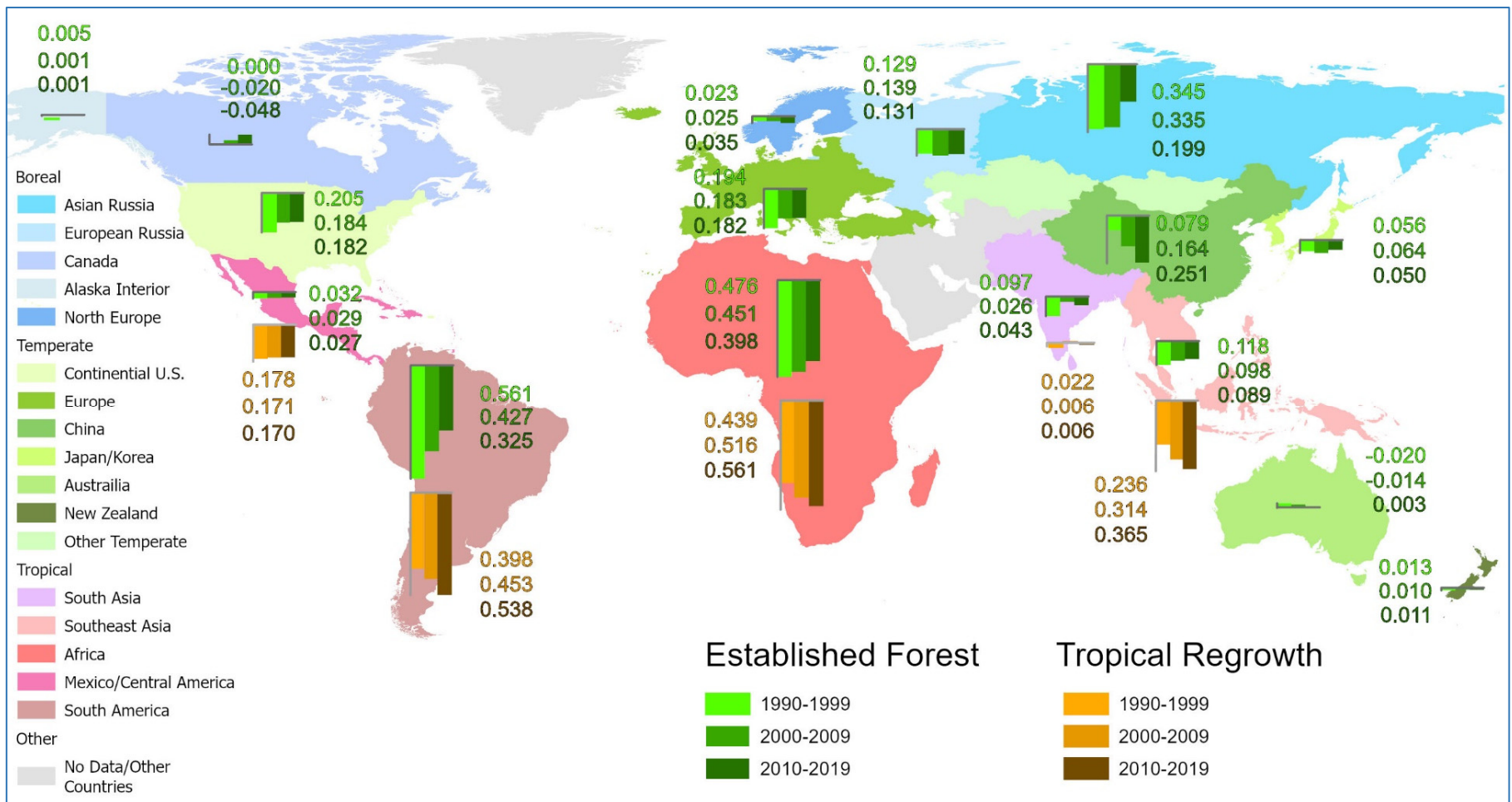
Extended Data Figure Legends

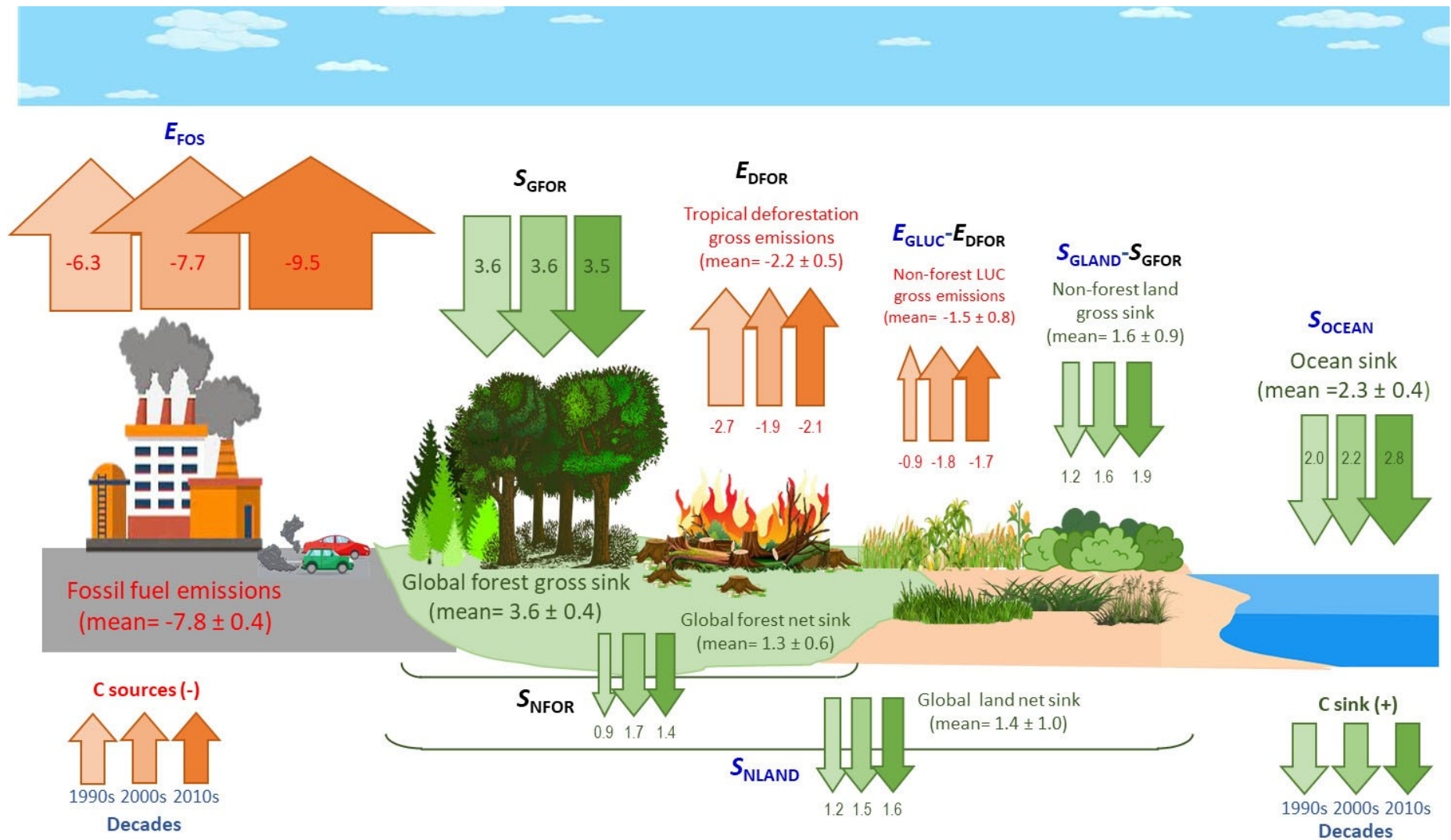
Extended Data Fig. 1. Why have tropical intact forests lost carbon stocks yet also provided a carbon sink? From 1990 to 2019, tropical intact forests that remain intact continued to sequester carbon by 32.0 Pg C (Table 1). Deforestation reduced the area of tropical intact forests by 467 Mha (containing C stocks of 149.4 Pg C). About 45% of C stocks in the deforested lands was emitted to the atmosphere shortly after the deforestation (mainly due to the slash-and-burning practice for agricultural land conversion), 36% was transferred to other land-uses such as agricultural lands (mostly as soil carbon), 17% was lost in processing harvested timber such as via wood shavings or stored in short-lived products such as fuelwood and paper, and 2% was retained in harvested wood products (HWP) such as long-lived construction materials. Because the remaining intact forests had provided a 32.0 Pg C sink, the net C stock loss from the intact forests was 117.5 Pg C.

Extended Data Fig. 2. Forest areas, carbon stocks, and carbon stock changes (fluxes) in the global forest and in forest biomes for the decades from 1990 to 2020: (a) forest areas; (b) carbon stocks; (c) carbon stock densities; (d) carbon stocks by pool; (e) carbon stock change (sinks); and (f) carbon stock change per hectare. The error bars represent standard deviations. For (a) we assumed 10% uncertainty in forest areas due to lack of documented uncertainty in remotely-sensed data; for (d) the uncertainty values of individual carbon pools were not included with most data sources, so we assumed that deadwood, litter and soil carbon pools have twice the uncertainty of the biomass pool, and estimated the uncertainty values of the individual carbon pools from the total carbon stock uncertainty. Uncertainties in the remaining charts are calculated based on data in Extended Data Table 2 and Extended Data Table 3.

Extended Data Fig.3. Carbon sinks and sources in global forests (Pg C year⁻¹) expressed as the mean annual rate across the full three-decadal period 1990 to 2019. Positive values represent carbon sinks, while negative (red) values carbon sources. Because carbon fluxes estimated in temperate and boreal forests were based on the “stock change” method, which included carbon gains and losses (from temporarily harvested forests), the C sink estimated was a net sink. Because carbon fluxes estimated in tropical forests were based on the “flux” method, C sinks estimated were gross sinks.

Tropical deforestation emissions were estimated by a book-keeping model. The tropical forest net sink, therefore, was the balance of C sinks and emissions (see Methods for concepts and details).





Box 1. Definitions of forest lands, features, and fluxes

Forest – The definition of forest varies slightly from country to country, but generally follows the FAO FRA definition (see Supplementary Information). (Note: Our forest definition does not wholly conform to the “managed– unmanaged lands” distinction that is compulsory in the reporting to UNFCCC and as used in global integrated assessment models, since we cover a large portion of unmanaged forests.

Forest land remaining forest land – forests that do not undergo land-use change during the reporting period, including forests that are harvested and regenerate back to forest.

Afforestation – land that has changed from non-forest to forest.

Deforestation – land that has changed from forest to non-forest.

Boreal and temperate forests – comprised of “forest land remaining forest land” plus new forests (afforested land), including primary forests, secondary forests that have regrown back either from harvesting historically or more recently, harvested forests that have temporarily lost tree cover, and land afforested from other non-forest land-uses.

Tropical intact forest – tropical forest areas that have not been strongly modified structurally by human activities. Tropical intact forests include primary forests, alongside slightly modified forests to a maximum modification from low-intensity selective logging, and some long-established secondary forests.

Established forest – used in this study to represent existing forests including boreal, temperate, and tropical intact forests.

Tropical regrowth forest – tropical forests regrowing on abandoned lands that have been previously deforested or logged and used for agriculture or other non-forest land-use types.

Gross C sink– Total C sequestered by forest (or land).

Net C sink – the *gross C sink* subtracting C emissions from forest deforestation and degradation (or from land-use changes).

Extended Data Table 1. Area of forests (10⁶ ha) and land-use change by biome, country, or region, and year

Biome and country /region	Total forest area, 1990 (10 ⁶ ha)	Total forest area, 2000 (10 ⁶ ha)	Total forest area, 2010 (10 ⁶ ha)	Total forest area, 2020 (10 ⁶ ha)	1990 -1999			2000 -2009			2010 -2019		
					Afforestation (10 ⁶ ha)	Deforestation (10 ⁶ ha)	Net change (10 ⁶ ha)	Afforestation (10 ⁶ ha)	Deforestation (10 ⁶ ha)	Net change (10 ⁶ ha)	Afforestation (10 ⁶ ha)	Deforestation (10 ⁶ ha)	Net change (10 ⁶ ha)
Boreal Forest¹													
Asian Russia	650.7	652.6	658.1	651.9	3.000	0.100	2.900	3.712	0.220	3.492	2.132	0.130	2.002
European Russia	170.7	173.3	177.6	181.8	3.201	0.100	3.101	3.596	0.130	3.466	3.028	0.120	2.908
Canada	226.9	226.5	226.0	225.5	0.082	0.579	-0.497	0.032	0.519	-0.487	0.009	0.544	-0.535
Alaska Interior	24.5	24.5	24.5	24.5	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000
European boreal ²	62.1	62.7	62.7	62.7	0.100	0.300	0.070	0.165	0.037	0.128	0.056	0.030	0.030
Subtotal	1134.7	1139.4	1148.8	1146.7	6.308	0.552	5.756	7.505	0.906	6.599	5.225	0.824	4.401
Temperate Forest¹													
United States ³	257.1	257.4	257.4	257.0	1.327	1.294	0.033	1.466	1.474	-0.008	1.399	1.721	-0.322
European temperate ⁴	104.5	110.7	116.2	119.0	4.510	0.745	3.765	3.641	0.918	2.723	2.542	0.923	1.619
Other Europe ⁵	40.9	42.0	43.4	45.0	0.797	0.139	0.658	0.431	0.51	0.280	0.517	0.230	0.286
China	139.7	142.8	163.5	174.1	44.349	41.227	3.122	40.295	19.587	20.708	16.372	5.778	10.594
Japan	25.2	25.1	25.1	25.0	0.488	0.577	-0.092	0.309	0.350	-0.044	0.208	0.250	-0.043
Korea	6.6	6.5	6.4	6.3	0.000	0.075	-0.075	0.000	0.089	-0.089	0.000	0.100	-0.100
Australia	134.4	132.7	131.1	134.1	0.633	0.155	0.478	3.085	6.523	-3.438	4.358	4.221	0.136
New Zealand	9.4	9.9	9.8	9.9	0.633	0.155	0.478	0.178	0.180	-0.002	0.120	0.076	0.045
Other countries ⁶	17.5	17.4	17.3	17.6	0.153	0.246	-0.093	0.503	0.657	-0.155	0.794	0.433	0.362
Subtotal	742.2	751.0	776.4	794.1	54.504	50.888	3.616	49.909	29.952	19.957	26.312	13.755	12.557
Tropical Intact Forest													
India	56.0	59.5	60.7	63.0	0.749	0.384	0.365	0.820	0.630	0.191	1.870	1.337	0.533
Other South Asia ⁷	14.0	15.1	14.6	14.3	0.284	0.076	0.208	0.301	0.053	0.249	0.110	0.095	0.015
Southeast Asia ⁸	190.6	136.9	118.4	90.1	10.670	30.240	-19.570	11.203	13.572	-2.369	2.685	10.316	-7.631
Africa	600.2	531.8	477.8	425.5	7.233	24.080	-16.847	8.317	23.144	-14.828	4.304	21.424	-17.121
Mexico	39.9	34.8	32.5	32.0	0.465	0.686	-0.221	0.281	0.425	-0.144	0.127	0.377	-0.250
Central America	10.7	9.2	7.7	6.1	0.095	1.137	-1.042	0.089	0.924	-0.835	0.196	0.603	-0.406
South America	885.2	817.2	756.3	698.5	6.813	58.405	-51.592	14.325	66.815	-52.491	5.559	31.639	-26.080
Subtotal	1796.6	1604.6	1468.1	1329.6	26.309	115.008	-88.699	35.335	105.562	-70.226	14.850	65.791	-50.941
Tropical Regrowth Forest													
India	8.0	8.1	8.8	9.2	na	na	na	na	na	na	na	na	na
Other South Asia	2.0	2.0	2.1	2.1	na	na	na	na	na	na	na	na	na
Southeast Asia	53.1	85.2	99.5	116.1	na	na	na	na	na	na	na	na	na
Africa	142.6	178.2	198.2	211.1	na	na	na	na	na	na	na	na	na
Mexico	30.7	33.6	34.4	33.7	na	na	na	na	na	na	na	na	na
Central America	23.2	23.4	23.5	24.2	na	na	na	na	na	na	na	na	na
South America	88.5	105.4	113.9	145.7	na	na	na	na	na	na	na	na	na
Subtotal	348.1	436.1	480.4	541.9									
All Tropical Forest													
India	63.9	67.6	69.5	72.2	0.749	0.384	0.365	0.820	0.630	0.191	1.870	1.337	0.533
Other South Asia	16.0	17.2	16.8	16.4	0.284	0.076	0.208	0.301	0.053	0.249	0.110	0.095	0.015
Southeast Asia	243.7	222.1	217.9	206.2	10.670	30.240	-19.570	11.203	13.572	-2.369	2.685	10.316	-7.631
Africa	742.8	710.0	676.0	636.6	7.233	24.080	-16.847	8.317	23.144	-14.828	4.304	21.424	-17.121
Mexico	70.6	68.4	66.9	65.7	0.465	0.686	-0.221	0.281	0.425	-0.144	0.127	0.377	-0.250
Central America	34.0	32.6	31.2	30.3	0.095	1.137	-1.042	0.089	0.924	-0.835	0.196	0.603	-0.406
South America	973.7	922.6	870.2	835.2	6.813	58.405	-51.592	14.325	66.815	-52.491	5.559	31.639	-26.080
Subtotal	2144.7	2040.6	1948.4	1871.5	26.309	115.008	-88.699	35.335	105.562	-70.226	14.850	65.791	-50.941
Global Total	4021.8	3931.2	3873.8	3812.0	87.196	166.975	-79.779	92.749	136.419	-43.670	46.387	80.370	-33.983

Note: ¹Includes carbon stock for the reporting year on "forest land remaining forest land" and "new forest land" (afforested land).

²Europe (boreal) includes Norway, Sweden, and Finland.

³Excluding part of Interior Alaska and Hawaii.

⁴Europe (temperate) includes European countries (EU-28) Albania, Bosnia Herzegovina, Serbia, Switzerland, except for Norway, Sweden, and Finland.

⁵Other Europe includes Ukraine, Belarus, Georgia, Armenia, Azerbaijan, and Turkey.

⁶Other countries include Mongolia and Kazakhstan.

⁷Other South Asia includes Afghanistan, Pakistan, Nepal, Bhutan, Bangladesh, Sri Lanka

⁸Southeast Asia includes Indonesia, Malaysia, Philippines, Vietnam, Cambodia, Thailand, Myanmar, Laos.

Extended Data Table 2. Forest carbon stocks (Pg C) by biome, country, or region for 1990, 2000, 2010, and 2020.

Biome and country /region	1990							2000							2010							2020							
	Total living biomass	Dead wood	Litter	Soil	Total C stock	-ty of total C stock (±)	Carbon density (MgC ha ⁻¹)	Total living biomass	Dead wood	Litter	Soil	Total C stock	-ty of total C stock (±)	Carbon density (MgC ha ⁻¹)	Total living biomass	Dead wood	Litter	Soil	Total C stock	-ty of total C stock (±)	Carbon density (MgC ha ⁻¹)	Total living biomass	Dead wood	Litter	Soil	Total C stock	-ty of total C stock (±)	Carbon density (MgC ha ⁻¹)	
Boreal Forest¹																													
Asian Russia	24.1	8.8	7.2	104.7	144.8	6.8	222.5	25.6	9.4	7.5	105.8	148.3	6.7	227.2	26.8	10.0	7.9	106.9	151.6	6.2	230.4	26.6	10.8	8.2	107.8	153.4	6.5	235.3	
European Russia	9.1	2.5	1.5	25.4	38.5	2.2	225.5	9.8	2.8	1.6	25.7	39.9	2.1	230.2	10.5	3.0	1.7	25.9	41.1	2.0	231.4	11.0	3.4	1.8	26.0	42.2	2.2	232.1	
Canada	13.3	4.6	10.7	18.6	47.3	7.1	208.3	13.2	4.5	10.9	18.6	47.2	7.1	208.2	12.6	4.6	10.9	18.7	46.8	7.0	207.1	12.4	4.5	10.7	18.7	46.3	6.9	205.2	
Alaska Interior	0.5	0.2	1.1	5.6	7.4	2.3	300.3	0.5	0.2	1.1	5.6	7.4	2.3	302.2	0.6	0.2	1.1	5.6	7.4	2.3	302.0	0.6	0.2	1.1	5.6	7.4	2.3	302.3	
European boreal ²	2.2	0.2	0.3	11.71	14.4	2.9	232.4	2.3	0.2	0.4	11.7	14.6	2.9	232.4	2.5	0.3	0.4	11.8	14.9	3.0	238.4	2.8	0.4	0.4	12.0	15.6	3.1	248.6	
Subtotal	49.2	16.3	20.9	166.0	252.4	10.7	222.4	51.4	17.1	21.4	167.4	257.3	10.6	225.8	53.0	18.0	21.9	168.9	261.8	10.3	227.9	53.4	19.3	22.2	170.0	264.9	10.5	231.1	
Temperate Forest⁴																													
United States ³	14.0	1.9	2.7	27.2	45.8	4.9	178.1	15.4	2.2	2.7	27.2	47.5	5.1	184.5	16.7	2.4	2.8	27.2	49.1	5.3	190.9	18.2	2.7	2.8	27.2	50.9	5.5	198.1	
European temperate ⁴	5.7	0.2	0.7	5.6	12.2	2.5	116.4	6.2	0.2	0.7	5.1	12.2	2.4	110.0	7.8	0.4	0.8	6.7	15.7	3.1	135.4	8.1	0.5	0.8	6.9	16.4	2.9	137.8	
Other Europe ⁵	1.6	0.0	0.3	2.1	3.9	0.7	96.0	2.0	0.0	0.3	2.3	4.6	0.8	109.0	2.3	0.0	0.3	2.4	5.1	0.9	116.6	2.6	0.0	0.4	2.7	5.7	1.0	126.4	
China	5.4	0.4	0.5	14.6	20.9	1.7	149.6	5.9	0.4	0.5	15.0	21.8	1.7	152.7	7.1	0.5	0.5	17.3	25.4	2.0	155.4	8.0	0.5	0.5	18.6	27.6	2.1	158.5	
Japan	1.8	0.1	0.1	3.6	5.7	2.1	224.4	2.1	0.1	0.1	3.6	5.9	2.1	235.1	2.3	0.1	0.1	3.6	6.1	2.1	244.2	2.4	0.1	0.1	3.6	6.2	2.1	247.8	
Korea	0.1	0.0	0.0	0.1	0.3	0.1	39.4	0.3	0.0	0.0	0.2	0.5	0.1	74.1	0.4	0.0	0.0	0.3	0.8	0.2	124.3	0.5	0.0	0.0	0.5	1.0	0.3	166.2	
Australia	8.4	1.6	0.6	11.5	22.1	7.4	164.4	8.4	1.6	0.6	11.4	22.0	7.4	165.8	8.3	1.6	0.6	11.3	21.8	7.3	166.3	8.3	1.6	0.6	11.4	21.9	7.3	160.0	
New Zealand	1.6	0.2	0.1	0.9	2.8	0.4	300.7	1.7	0.2	0.1	0.9	3.0	0.4	299.6	1.8	0.2	0.1	0.9	3.0	0.5	306.9	1.8	0.2	0.1	0.9	3.1	0.5	310.7	
Other countries ⁶	0.7	0.1	0.3	1.5	2.7	0.6	151.7	0.7	0.1	0.3	1.5	2.7	0.6	152.3	0.7	0.1	0.3	1.5	2.7	0.6	153.9	0.7	0.1	0.3	1.6	2.7	0.7	154.5	
Subtotal	39.6	4.5	5.4	67.1	116.5	9.7	157.0	42.8	4.8	5.4	67.3	120.3	9.8	160.2	47.8	5.4	5.6	71.3	130.0	10.2	167.4	51.0	5.8	5.6	73.4	135.8	10.1	171.0	
Tropical Intact Forest																													
India	2.1	0.0	0.1	3.3	5.5	1.4	97.7	2.7	0.0	0.1	3.4	6.2	1.5	104.2	2.7	0.0	0.1	3.5	6.3	1.5	103.5	2.8	0.0	0.1	3.6	6.5	1.5	103.2	
Other South Asia ⁷	1.3	0.0	0.0	0.5	1.9	0.4	134.1	1.4	0.0	0.0	0.6	2.1	0.5	141.9	1.4	0.0	0.0	0.6	2.1	0.5	143.9	1.4	0.0	0.0	0.6	2.1	0.5	146.1	
Southeast Asia ⁸	45.8	8.5	0.7	17.3	72.3	9.4	379.5	33.7	6.2	0.5	12.5	52.9	7.2	386.7	29.9	5.5	0.5	10.8	46.6	6.6	393.8	23.3	4.2	0.4	8.3	36.2	5.4	402.0	
Africa	85.4	18.8	1.3	64.7	170.3	48.3	283.7	79.4	17.2	1.3	57.5	155.3	44.2	292.1	74.9	15.9	1.2	51.8	143.8	40.6	300.9	69.6	14.5	1.1	46.3	131.6	37.5	309.2	
Mexico	1.6	0.2	0.1	2.6	4.5	0.6	113.4	1.4	0.1	0.1	2.3	3.9	0.5	113.4	1.5	0.1	0.1	2.2	3.9	0.5	118.4	1.5	0.1	0.1	2.1	3.8	0.5	118.4	
Central America	1.7	0.3	0.0	1.0	3.1	0.4	285.2	1.5	0.3	0.0	0.8	2.7	0.3	293.0	1.3	0.2	0.0	0.7	2.3	0.3	300.1	1.1	0.2	0.0	0.6	1.9	0.2	308.9	
South America	142.6	26.9	2.6	81.1	253.2	32.2	286.0	136.2	25.4	2.5	75.1	239.1	31.1	292.6	129.4	23.9	2.4	69.6	225.4	46.4	298.0	121.9	22.4	2.3	64.6	211.2	44.7	302.3	
Subtotal	280.4	54.7	5.0	170.6	510.7	58.9	284.3	256.3	49.2	4.6	152.2	462.4	54.5	288.2	241.1	45.7	4.4	139.2	430.3	62.0	293.1	221.7	41.5	4.1	126.0	393.2	58.7	295.8	
Tropical Regrowth Forest																													
India	0.1	0.0	0.0	0.5	0.6	0.2	75.8	0.3	0.0	0.0	0.5	0.8	0.2	95.7	0.3	0.0	0.0	0.5	0.8	0.3	94.1	0.3	0.0	0.0	0.5	0.9	0.3	95.5	
Other South Asia	0.1	0.0	0.0	0.1	0.2	0.1	84.1	0.1	0.0	0.0	0.1	0.2	0.1	107.7	0.1	0.0	0.0	0.1	0.2	0.1	109.9	0.1	0.0	0.0	0.1	0.2	0.1	111.0	
Southeast Asia	3.0	0.6	0.1	2.9	6.5	2.3	123.0	5.0	0.7	0.1	3.1	8.9	2.2	104.4	7.6	0.8	0.1	3.5	12.0	3.0	120.9	10.7	0.9	0.1	4.0	15.7	3.7	135.1	
Africa	3.6	1.7	0.1	8.7	14.1	2.4	98.6	7.1	1.8	0.1	9.4	18.4	3.9	103.5	11.3	1.9	0.2	10.2	23.6	5.1	119.1	15.9	2.1	0.2	11.0	29.2	6.1	138.4	
Mexico	0.3	0.0	0.0	0.5	0.8	0.1	26.8	0.4	0.0	0.0	0.7	1.2	0.2	36.6	0.6	0.0	0.0	0.8	1.4	0.2	41.8	0.5	0.0	0.0	0.8	1.4	0.2	41.5	
Central America	0.6	0.2	0.0	1.4	2.2	0.3	95.0	1.4	0.2	0.0	1.5	3.2	0.4	135.1	2.3	0.2	0.0	1.6	4.1	0.5	175.7	3.1	0.3	0.0	1.7	5.1	0.6	212.0	
South America	2.8	2.1	0.2	4.0	9.2	1.7	103.5	6.3	2.2	0.2	4.4	13.1	3.3	124.6	10.2	2.4	0.3	4.9	17.7	4.4	155.2	14.8	2.5	0.3	5.4	23.0	5.2	158.4	
Subtotal	10.5	4.6	0.4	18.0	33.6	3.8	96.4	20.7	5.0	0.5	19.7	45.9	5.6	105.2	32.4	5.4	0.6	21.6	59.9	7.4	124.8	45.5	5.9	0.7	23.5	75.6	8.8	139.4	
All Tropical Forest																													
India	2.2	0.0	0.1	3.8	6.1	1.4	95.0	3.0	0.0	0.1	3.9	7.0	1.4	103.2	3.0	0.0	0.1	4.0	7.1	1.4	102.3	3.1	0.0	0.1	4.1	7.4	1.6	102.2	
Other South Asia	1.4	0.0	0.1	0.6	2.0	0.4	127.8	1.5	0.0	0.1	0.7	2.4	0.5	137.9	1.5	0.0	0.1	0.7	2.3	0.5	139.6	1.5	0.0	0.1	0.7	2.3	0.5	141.7	
Southeast Asia	48.8	9.2	0.8	20.2	78.9	9.7	323.6	38.7	6.9	0.6	15.6	61.8	44.3	278.4	37.5	6.3	0.6	14.3	58.7	7.3	269.2	34.0	5.2	0.5	12.2	51.9	6.5	251.7	
Africa	88.9	20.5	1.5	73.4	184.3	48.4	248.1	86.6	18.9	1.4	66.9	173.8	44.4	244.8	86.2	17.8	1.4	61.9	167.4	40.9	247.6	85.5	16.6	1.4	57.3	160.8	38.0	252.5	
Mexico	1.9	0.2	0.1	3.1	5.3	0.6	75.7	1.9	0.2	0.1	3.0	5.2	0.4	75.7	2.1	0.2	0.1	3.0	5.3	0.4	79.0	2.0	0.2	0.1	2.9	5.2	0.5	79.0	
Central America	2.3	0.5	0.0	2.4	5.3	0.5	155.2	3.0	0.5	0.0	2.4	5.9	0.5	179.6	3.6	0.5	0.0	2.3	6.4	0.6	206.2	4.2	0.5	0.1	2.3	7.0	6.1	231.6	
South America	145.4	29.0	2.9	85.1	262.3	32.2	269.4	142.4	27.6	2.8	79.5	252.3	31.2	273.4	139.6	26.3	2.7	74.5	243.1	46.6	279.3	136.8	24.9	2.6	70.0	234.2	45.0	277.5	
Subtotal	291.0	59.3	5.4 </																										

Extended Data Table 3. Estimated annual change in forest C stock (Tg C year⁻¹) by biome, country, or region for 3 decades respectively from 1990 to 2020.

Biome and country /region	1990-1999								2000-2009								2010-2019								
	Total living biomass	Dead wood	Litter	Soil	Harvested wood products	Net C stock change	Uncertainty of net stock change (±)	Stock change per area (MgC ha ⁻¹ year ⁻¹)	Total living biomass	Dead wood	Litter	Soil	Harvested wood products	Net C stock change	Uncertainty of net stock change	Stock change per area (MgC ha ⁻¹ year ⁻¹)	Total living biomass	Dead wood	Litter	Soil	Harvested wood products	Net C stock change	Uncertainty of net stock change	Stock change per area (MgC ha ⁻¹ year ⁻¹)	
Boreal Forest¹																									
Asian Russia	145.1	58.7	31.7	105.5	6.3	347.3	59.6	0.53	127.7	61.6	34.5	109.3	4.5	337.6	48.2	0.51	-19.8	84.5	35.8	95.7	5.9	202.1	22.8	0.30	
European Russia	68.0	25.9	10.8	20.1	7.1	131.9	19.4	0.75	74.9	23.3	13.3	24.9	6.2	142.6	18.3	0.79	69.2	28.1	13.8	17.3	5.7	134.1	13.1	0.73	
Canada	-6.4	-16.7	10.0	2.6	10.9	0.4	0.1	0.00	-55.1	11.1	8.8	2.7	12.2	-20.2	5.5	-0.09	-29.3	-8.2	-17.9	1.8	5.9	-47.8	12.9	-0.21	
Alaska Interior	4.7	0.1	-0.2	0.0	0.0	4.5	1.4	0.19	3.8	-0.3	-2.8	0.0	0.0	0.7	0.2	0.03	3.7	-0.2	-1.2	-1.3	0.0	1.1	0.3	0.04	
European boreal ²	14.7	1.5	2.2	-0.6	5.5	23.3	7.0	0.37	13.1	1.3	2.0	1.9	6.8	25.1	8.0	0.40	20.4	2.0	2.0	3.0	7.4	34.8	10.4	0.40	
Subtotal	226.1	69.5	54.5	127.6	29.8	507.5	63.1	0.44	164.5	97.1	55.8	138.8	29.8	485.8	52.5	0.42	44.1	106.2	32.5	116.5	24.9	324.3	41.1	0.28	
Temperate Forest¹																									
United States ³	143.4	26.7	1.7	1.1	32.3	205.1	22.2	0.80	134.5	26.9	0.5	-2.1	24.3	184.0	19.9	0.71	132.3	27.5	-0.8	-0.4	23.3	181.9	19.6	0.71	
European temperate ⁴	89.5	20.9	2.0	5.3	8.2	125.9	18.7	1.17	101.3	3.9	2.1	7.9	17.0	132.3	20.9	1.17	79.8	3.0	3.0	11.9	18.9	116.6	18.1	0.99	
Other Europe ⁵	41.3	0.2	3.0	21.1	2.1	67.8	14.8	1.64	35.5	0.2	1.5	11.3	2.1	50.5	10.5	1.18	30.6	0.1	2.3	28.9	3.3	65.3	16.4	1.48	
China	46.0	2.6	0.6	10.9	18.6	78.7	12.7	0.56	121.2	6.8	6.0	12.9	17.5	164.4	25.8	1.07	180.2	7.4	6.1	32.6	24.2	250.5	38.0	1.48	
Japan	27.1	0.0	0.0	-1.3	7.7	33.5	9.6	1.33	25.4	0.0	0.0	-0.6	7.5	32.3	8.8	1.29	12.7	0.0	0.0	-0.8	10.9	22.8	4.1	0.91	
Korea	11.6	0.0	0.9	9.7	0.4	22.5	5.7	1.71	16.4	0.0	1.3	13.7	0.6	32.0	8.0	2.50	13.1	0.0	1.0	12.0	0.6	26.8	6.8	2.15	
Australia	-15.9	-2.0	-0.1	-4.1	1.9	-20.2	6.8	-0.17	-12.5	-0.9	0.5	-3.4	1.9	-14.4	4.8	-0.12	7.4	-0.8	0.5	-5.3	1.2	3.0	1.0	0.02	
New Zealand	6.3	1.0	0.3	4.6	1.1	13.3	3.2	1.38	7.9	0.7	0.2	-0.3	1.4	9.9	2.8	1.01	7.8	0.9	-0.1	0.3	2.2	11.1	2.8	1.13	
Other countries ⁶	-0.1	0.0	0.0	-0.2	0.05	-0.4	0.1	-0.02	0.11	0.0	0.0	0.2	0.06	0.5	0.2	0.03	1.7	0.3	0.7	3.8	0.06	6.6	2.0	0.38	
Subtotal	349.2	49.3	8.3	47.0	72.4	526.2	37.4	0.70	429.8	37.6	12.1	39.7	72.4	591.6	42.2	0.77	465.5	38.5	12.8	83.0	84.7	684.7	50.0	0.87	
Tropical Intact Forest																									
India	50.2	0.1	3.4	10.1	13.8	77.6	12.9	1.34	0.4	0.0	0.8	7.1	16.8	25.2	2.9	0.42	12.5	0.1	0.4	8.5	18.5	40.0	4.2	0.65	
Other South Asia ⁷	7.8	0.4	-0.1	6.0	4.8	18.9	5.5	1.30	-3.5	-0.2	-0.1	0.1	4.3	0.6	1.4	0.04	-1.2	0.0	0.0	0.0	4.4	3.3	1.2	0.21	
Southeast Asia	82.2	10.4	1.7	4.0	19.6	118.0	45.0	0.72	70.9	8.9	1.5	2.7	13.8	97.8	36.5	0.77	61.1	7.8	1.3	3.8	14.9	88.8	34.1	0.85	
Africa	379.2	48.2	8.0	13.9	27.2	476.4	138.7	0.84	355.3	45.1	7.5	10.7	32.9	451.4	94.3	0.89	298.1	37.9	6.3	16.7	38.6	397.5	133.3	0.88	
Mexico	17.9	2.3	0.4	0.9	1.1	22.5	6.2	0.60	16.1	2.1	0.3	0.8	2.3	21.7	6.0	0.64	15.5	2.0	0.3	0.7	2.3	20.9	5.7	0.65	
Central America	6.7	0.6	0.1	0.2	2.3	9.9	2.7	0.99	3.7	0.5	0.1	0.2	2.6	7.0	1.9	0.83	2.3	0.3	0.0	0.2	2.7	5.6	1.5	0.82	
South America	451.1	57.3	9.5	20.9	21.7	560.5	138.4	0.66	341.4	43.5	7.2	16.5	18.6	427.3	160.0	0.54	241.7	30.7	5.1	26.0	20.8	324.3	190.5	0.45	
Subtotal	995.2	119.3	23.0	56.0	90.4	1283.8	201.6	0.75	784.3	99.9	17.3	38.0	91.5	1030.9	189.4	0.67	630.0	78.7	13.5	56.2	102.2	880.6	235.1	0.63	
Tropical Regrowth Forest																									
India	16.4	0.0	0.5	-0.1	0.0	16.8	4.9	2.09	1.3	0.0	0.2	3.8	0.0	5.3	1.9	0.63	3.4	0.0	0.1	1.6	0.0	5.1	1.3	0.57	
Other South Asia ⁷	3.9	0.1	0.0	1.3	0.0	5.2	2.1	2.59	0.4	0.0	0.0	0.5	0.0	1.0	0.3	0.46	0.1	0.0	0.0	0.8	0.0	0.9	0.4	0.45	
Southeast Asia	198.1	7.1	1.4	29.1	0.0	235.6	105.6	3.41	263.5	9.4	1.9	38.8	0.0	313.6	125.4	3.39	306.1	10.9	2.2	45.3	0.0	364.5	151.4	3.38	
Africa	356.1	13.2	2.6	67.4	0.0	439.3	168.8	2.74	418.0	15.5	3.1	79.1	0.0	515.6	153.6	2.74	454.7	16.8	3.4	86.0	0.0	560.9	194.5	2.74	
Mexico	67.0	4.9	0.1	10.2	0.0	82.2	22.6	2.56	60.4	4.5	0.1	9.2	0.0	74.2	20.4	2.18	57.9	4.3	0.1	8.8	0.0	71.1	19.6	2.09	
Central America	82.6	2.9	0.6	9.8	0.0	95.9	26.4	4.11	83.6	2.9	0.6	9.9	0.0	96.9	26.7	4.13	85.4	3.0	0.6	10.0	0.0	99.0	27.2	4.15	
South America	343.4	12.0	2.4	40.7	0.0	398.4	163.3	4.11	390.2	13.6	2.7	46.1	0.0	452.6	212.4	4.13	464.4	16.1	3.2	54.5	0.0	538.3	220.9	4.15	
Subtotal	1067.4	40.1	7.5	158.3	0.0	1273.4	259.9	3.25	1217.4	45.9	8.6	187.3	0.0	1459.1	292.5	3.18	1372.1	51.2	9.5	207.0	0.0	1639.8	332.7	3.21	
All Tropical Forest																									
India	66.6	0.1	3.9	10.0	13.8	94.4	13.8	1.44	1.7	0.1	1.0	10.9	16.8	30.5	3.4	0.44	15.8	0.1	0.5	10.2	18.5	45.1	4.0	0.64	
Other South Asia ⁷	11.7	0.5	-0.1	7.3	4.8	24.2	5.8	1.45	-3.0	-0.2	-0.1	0.5	4.3	1.5	1.4	0.09	-1.2	0.0	0.0	0.8	4.4	4.0	1.2	0.24	
Southeast Asia	280.3	17.5	3.1	33.1	19.6	353.6	114.8	1.52	334.4	18.3	3.4	41.5	13.8	411.4	130.6	1.87	367.3	18.7	3.5	49.0	14.9	453.4	155.2	2.14	
Africa	735.3	61.3	10.6	81.2	27.2	915.7	218.5	1.26	773.2	60.6	10.6	89.7	32.9	967.0	180.3	1.40	752.8	54.7	9.6	102.7	38.6	958.4	235.8	1.46	
Mexico	84.9	7.2	0.5	11.1	1.1	104.7	23.4	1.51	76.5	6.5	0.4	10.0	2.3	95.8	21.2	1.42	73.4	6.3	0.4	9.6	2.3	92.0	14.6	1.39	
Central America	89.3	3.4	0.7	10.0	2.3	105.7	26.5	3.18	87.2	3.4	0.7	10.0	2.6	103.9	26.7	3.26	87.7	3.3	0.6	10.3	2.7	104.6	27.3	3.40	
South America	794.5	69.2	11.9	61.6	21.7	958.9	214.0	1.01	731.6	57.1	9.9	62.6	18.6	879.9	265.9	0.98	706.1	46.8	8.3	80.7	20.8	862.8	291.7	1.01	
Subtotal	2062.6	159.4	30.5	214.3	90.4	2557.2	328.9	1.22	2001.7	145.8	25.9	225.2	91.5	2490.0	348.5	1.25	2002.1	129.9	23.0	263.2	102.2	2520.5	407.4	1.32	
Global Total	2637.9	278.3	93.3	389.0	192.6	3591.0	337.0	0.90	2596.0	280.4	93.7	403.7	193.7	3567.4	354.9	0.91	2511.7	274.7	68.3	462.8	211.9	3529.8	412.5	0.92	

Note: ¹Includes carbon stock for the reporting year on "forest land remaining forest land" and "new forest land" (afforested land).

²Europe (boreal) includes Norway, Sweden, and Finland.

³Excluding Interior Alaska and Hawaii.

⁴Europe (temperate) includes European countries (EU-28), Albania, Bosnia Herzegovina, Serbia, Switzerland, except for Norway, Sweden, and Finland.

⁵Other Europe includes Ukraine, Belarus, Georgia, Armenia, Azerbaijan, and Turkey.

⁶Other countries include Mongolia and Kazakhstan.

⁷Other South Asia includes Afghanistan, Pakistan, Nepal, Bhutan, Bangladesh, Sri Lanka.

⁸Southeast Asia includes Indonesia, Malaysia, Philippines, Vietnam, Cambodia, Thailand, Myanmar, Laos.

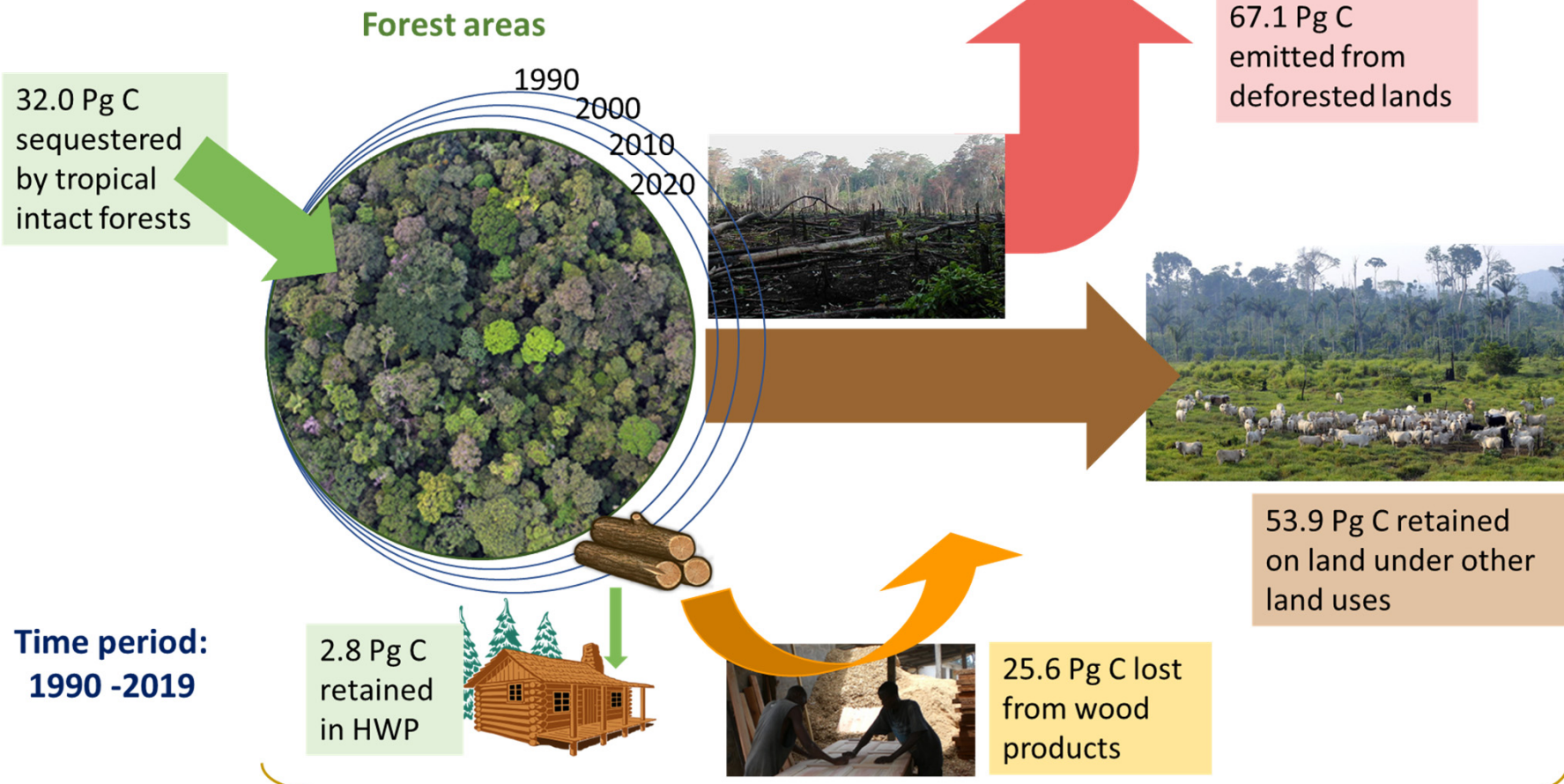
Extended Data Table 4. Alternative accounting of the Global Carbon Budget (Pg C year⁻¹)

Global C fluxes and budget	1990-1999	2000-2009	2010-2019 ¹	1990-2019 (Mean)	1990-2019 (Total, PgC)
Sources (C emissions):					
Fossil fuel emission (E_{FOS}) ²	-6.3 ± 0.3	-7.7 ± 0.4	-9.5 ± 0.5	-7.8 ± 0.4	-235 ± 2.2
Land-use change (LUC) gross emission (E_{GLUC}) ³	-3.6 ± 0.6	-3.7 ± 0.6	-3.8 ± 0.6	-3.7 ± 0.6	-111 ± 3.3
Total gross C emission ($E_{FOS} + E_{GLUC}$) ⁴	-9.9 ± 0.7	-11.4 ± 0.7	-13.3 ± 0.8	-11.5 ± 0.7	-346 ± 4.0
Sinks (C sequestration):					
Atmosphere (G_{ATM}) ⁵	3.1 ± 0.02	4.0 ± 0.02	5.1 ± 0.02	4.1 ± 0.02	122 ± 0.11
Ocean (S_{OCEAN}) ⁶	2.0 ± 0.4	2.2 ± 0.4	2.8 ± 0.4	2.3 ± 0.4	70 ± 2.2
Global land gross C sink (S_{GLAND}) ⁷	4.8 ± 0.8	5.2 ± 0.8	5.4 ± 0.9	5.1 ± 0.8	154 ± 4.6
Land C fluxes:					
Global C sink in established forests ⁸	2.3 ± 0.2	2.1 ± 0.2	1.9 ± 0.2	2.1 ± 0.2	63.2 ± 1.2
Global gross C sink in all Earth's forests ⁹	3.6 ± 0.3	3.6 ± 0.4	3.5 ± 0.4	3.6 ± 0.4	106.9 ± 2.0
Global non-forest land gross C sink ($S_{GLAND} - S_{GFOR}$) ¹⁰	1.2 ± 0.9	1.6 ± 0.9	1.9 ± 1.0	1.6 ± 0.9	47.1 ± 5.1
Tropical deforestation gross emission (E_{DFOR}) ¹¹	-2.7 ± 0.5	-1.9 ± 0.4	-2.1 ± 0.6	-2.2 ± 0.5	-67.1 ± 2.8
Global non-forest LUC gross emission ($E_{GLUC} - E_{DFOR}$) ¹²	-0.9 ± 0.8	-1.8 ± 0.7	-1.7 ± 0.8	-1.5 ± 0.8	-43.9 ± 4.2
Global land net sink (S_{NLAND}) ¹³	1.2 ± 1.0	1.5 ± 1.0	1.6 ± 1.1	1.4 ± 1.0	43 ± 5.7
Global forest net sink (S_{NFOR}) ¹⁴	0.9 ± 0.6	1.7 ± 0.6	1.4 ± 0.7	1.3 ± 0.6	39.8 ± 3.5

Notes and definitions of C fluxes in the table and the global carbon budget, red and (-) values are C sources, while black and (+) values are C sinks:

1. Estimates are derived from the Global Carbon Budget (GCB) Table 6 of Friedlingstein et al.², in which the last decade is presented as 2011-2020, while in this study 2010-2019.
2. **Fossil fuel emissions** (E_{FOS}) are derived from Table 6 of Friedlingstein et al.².
3. **Land-use change (LUC) gross emissions** (E_{GLUC}) are derived from Table 5 of Friedlingstein et al.², which are all LUC gross emissions including tropical deforestation gross emissions.
4. **Total gross C emissions** are the result of $E_{FOS} + E_{GLUC}$
5. **Atmosphere C growth** (G_{ATM}) is derived from Table 6 of Friedlingstein et al.², which is the increase of atmospheric carbon (in the CO₂ form).
6. **Ocean C sequestration** (S_{OCEAN}) was derived from Table 6 of Friedlingstein et al.², which is the carbon absorbed by oceans.
7. **Global land gross C sink** (S_{GLAND}) is the result of **Total gross C emissions** minus the C growth in the Atmosphere (G_{ATM}) and carbon sequestration by Ocean (S_{OCEAN}), so often viewed as the residual sink.
8. **Global C sink in established forests** include boreal, temperate and tropical intact forests (excluding tropical regrowth forest, which means excluding LUC).
9. **Global gross C sink in all Earth's forests** (S_{GFOR}) is the estimate from this study (Table 1).
10. **Global non-forest land gross C sink** is the result of $S_{GLAND} - S_{GFOR}$
11. **Tropical deforestation gross emission** (E_{DFOR}) is the estimate from this study (Table 1).
12. **Global non-forest gross emission** is the result of $E_{GLUC} - E_{DFOR}$.
13. **Global land net sink** (S_{NLAND}) is the balance between S_{GLAND} and E_{GLUC} .
14. **Global forest net sink** (S_{NFOR}) is the balance between S_{GFOR} and E_{DFOR} and the estimate from this study (Table 1).

Gross loss of C stocks from tropical intact forests: 149.4 Pg C



Net loss of C stocks from tropical intact forests: 117.5 Pg C

