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Thiol-rich and ion-imprinted alginate hydrogel as a highly adsorptive and recyclable filtration membrane for rapid and selective Sr(II) removal

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Abstract

Radioactive metal ion such as strontium ion, $^{90}\text{Sr}^{2+}$, has posed severe threats to environments and humans since the wide application of nuclear power plants around the world, while a rapid remediation of Sr^{2+} contaminated water still remains challenging. The current study developed an economical biomaterial-based hydrogel adsorbent with excellent Sr^{2+} adsorption performance achieved by ion-imprinting and abundant thiol groups, which was adaptable as an adsorptive filtration membrane for efficient and rapid purification of Sr^{2+} polluted water.

The hydrogel was synthesized via a three-step route based on sodium alginate (SA). First, SA was emulsified and converted via Sr^{2+} complexation to hydrogel (SA-Sr); secondly, a thiol-rich carboxyethyl grafted pentaerythritol tetrakis (thioglycolic acid) ester (PA) synthesized by click chemistry was used to covalently crosslink the hydrogel (SA-PA-Sr) with abundant thiol groups simultaneously introduced. Lastly, a Sr^{2+} -imprinted adsorbent (SA-PA-H) was obtained via acid elution of the SA-PA-Sr gel.

The SA-PA-H was demonstrated to exhibit a superior Sr^{2+} adsorption capacity (~ 151.7 mg/g), a rapid adsorption kinetics following pseudo-second order with a rate constant of 0.669 g mg^{-1} min^{-1} , decent selectivity for Sr^{2+} , a value $\sim 1.14 \times 10^3$ mL g^{-1} when adsorbing 10 ppm Sr^{2+} from

30 concentrated (100 ppm) solutions of competitive ions (Na^+ or Mg^{2+}). The good performance was
31 maintained over a wide range of pH (4-10) and temperature (25-40 °C), and the adsorption
32 mechanism was attributed to the prevalent Sr^{2+} bindings to thiol groups and Sr^{2+} -imprinted cavities.
33 Moreover, high elasticity with a storage shear modulus ~ 10 MPa at low strains whilst rapid and
34 full self-recovery after being repeatedly damaged by large strains of the SA-PA-H were
35 demonstrated by rheology. This allowed the SA-PA-H to be adapted as a membrane for vacuum
36 filtration, giving a high removal efficiency ($> 99.2\%$) of Sr^{2+} under a high liquid flux ($\sim 40 \text{ L m}^{-2}$
37 h^{-1}). In addition, the adsorbent can be regenerated by acid washing and after four consecutive
38 adsorption-desorption cycles, the drop in removal efficiency was minor (53.51% to 36.88% for
39 100 ppm Sr^{2+}). This investigation demonstrated a novel hydrogel adsorbent advantageous in cost,
40 performance, processability, and sustainability, being applicable for rapid and complete
41 decontamination of nuclear wastewater via adsorptive membrane filtration.

42

43 **Keywords** Ion-imprinting; alginate; strontium; adsorption; thiol; membrane; hydrogel

44

45 **1 Introduction**

46 ^{90}Sr (Strontium (^{90}Sr)) is one of the major products of nuclear fission, possessing a long half-life
47 (28.79 years) with high energy β -decay. Moreover, its cation, i.e. $^{90}\text{Sr}^{2+}$, is highly soluble in water.
48 Hence, if released, $^{90}\text{Sr}^{2+}$ poses severe threats to the environment and human bodies.[1] Due to its
49 similar chemical properties to calcium, Sr^{2+} is prone to enrich in human bodies, especially in
50 bones. The good affinity of $^{90}\text{Sr}^{2+}$ to bones prevents it from being excreted by metabolism like
51 other poisons. It was demonstrated that people suffered from obvious risks of osteosarcoma, soft
52 tissue cancer and leukemia when exposed to radioactive ^{90}Sr . [2] With the wide application of
53 nuclear power around the world, how to treat polluted water containing radioactive ions including
54 $^{90}\text{Sr}^{2+}$ has become an urgent problem. On the other hand, Sr and its derivatives are important metal
55 materials and have been widely used in ceramics, coatings, electronics and pharmaceuticals.[3-7]
56 Development of efficient approaches to separating and recovering Sr^{2+} from polluted water not
57 only protects humans and the environment but also generates significant economic benefits.

58 Researchers have developed several methods to separate metal ions from polluted water, such

59 as chemical precipitation,[8] extraction[9], floatation[10, 11], adsorption[11-14] and ion
60 exchange[15, 16]. The methods have different removal efficiency and application scenes
61 according to their mechanism. Among them, adsorption is considered to be one of the best ways to
62 recover Sr^{2+} with high efficiency and various adsorptive materials are developed for the
63 enrichment and separation of Sr^{2+} . [17-21] For example, graphene oxide which has a large specific
64 surface area and oxygen functional groups is a good candidate for ion adsorption materials. Abu-
65 Nada reported graphene oxide as a Sr^{2+} adsorbent in the aqueous system, with a maximum
66 strontium adsorption capacity of 131.4 mg/g achieved. [22] Nevertheless, the traditional adsorbents
67 for Sr^{2+} , such as polyantimonic acid, [23] sodium titanate [24] and zeolite [25-27], and nanocarbon
68 materials [28], etc., are often prepared in form of powder with a tiny particle size (mostly in
69 nanoscale or submicron scale), hence may cause secondary pollution when applied in water since
70 the separation of the adsorbents are laborious. It is also difficult for the powdery adsorbents to be
71 shaped to fit purification devices. Even though the adsorbents can be loaded to a matrix to
72 improve their recovery abilities from water, [29, 30] the adsorption efficiency including maximum
73 adsorption and adsorption rate generally decreases due to the loss of active adsorption sites. The
74 relatively high cost of the adsorbent synthesis also limits its massive application. For this reason,
75 the development of highly efficient, environment-friendly, and economical adsorbents for Sr^{2+} has
76 attracted growing attention in the water treatment field.

77 Organic polymers usually have good stability and are easy to be functionalized by
78 modification of monomer or polymer products. The shape of polymer adsorbent can also be
79 tailored by various preparation techniques. The advantages have made polymers ideal candidates
80 for adsorbents. For instance, Huh et al. synthesized a microporous covalent triazine polymer and
81 loaded magnetite Fe_3O_4 nanoparticles on it. [31] The nanocomposite served as an efficient
82 adsorbent for the removal of strontium ion (Sr^{2+}) from seawater with a maximum adsorption
83 capacity of 128 mg/g. Wang et al. prepared polyaniline-functionalized porous chitosan grafted by
84 polyacrylamide bearing sulfonic acid group and used it as Sr^{2+} adsorbent. [32] The maximum
85 adsorption capacity was determined to be 88.95 mg/g. However, there are at least two main
86 drawbacks hampering the wide application of polymer adsorbents. One lies in that the introduction
87 of functional groups into polymer adsorbent for adsorption often involves extra tedious steps

88 during the synthesis of the polymers. Another is that a majority of polymers are known to be hard
89 to degrade due to their carbon skeleton, that is, the commonly argued issue of the non-
90 sustainability of plastics. To overcome the problem, bio-based polymer materials such as sodium
91 alginate (SA) have been adopted as the matrix or the core adsorptive material to prepare
92 biodegradable ion adsorbents. SA is a bio-derived polysaccharide, possessing the advantages of
93 abundant sources, low cost, nontoxicity, and biodegradability, etc.[33, 34] SA is an excellent
94 adsorbent for various metal ions because of abundant carboxylic side groups.[35-39] However,
95 there is minor ion selectivity of SA towards specific ions like Sr^{2+} , which calls for further
96 modification of SA to meet the requirement for water treatment.

97 In recent years, ion-imprinting was raised as a prominent technique to construct ion-
98 responsive materials with selectivity to certain ions.[40] In typical ion-imprinting, the ion was first
99 introduced as templates to complex with ligand groups in polymers. After the formation of the ion
100 complex, crosslinking reactions were introduced via other groups on polymers to fabricate stable
101 networks and fix the relative position of functional groups. By removing metal ions, cavities with
102 precise shapes, sizes, and charge distribution conformed to the ion are readily created in the
103 network. The cavities served as identification sites for the ion and thus endowed the polymer
104 matrix with ion selectivity, which allowed the polymer to serve as an efficient adsorbent for the
105 specific ion^[41-43]. As an example of strontium recovery, Deng et al. reported a Sr^{2+} -imprinted
106 alginate microsphere adsorbent,[44] which can successfully separate Sr^{2+} in oilfield water from
107 other ions with a maximum adsorption capacity of 177.37 mg/g, but its microsphere form and un-
108 crosslinking structure still make the adsorbent difficult to be recovered and unstable to be re-used.

109 Incorporating functional group with distinct affinity to the ion into the adsorbent is another
110 effective strategy to enhance ion selectivity. Based on hard and soft Lewis acids and bases (HSAB)
111 theory,[45] soft Lewis bases with large radii and highly deformable electron clouds are prone to
112 combine with soft Lewis acids with similar properties. As a third-period element, sulfur was a soft
113 Lewis base with a large radius and a thick electron cloud. Therefore, sulfur will preferentially
114 combine with soft Sr^{2+} instead of other hard ions in a mixture of alkali and alkaline-earth metal
115 ions including Na^+ , Mg^{2+} , Ca^{2+} , and Sr^{2+} . Taking advantage of the feature, metal sulfides were
116 demonstrated to be efficient adsorbents for Sr^{2+} and other soft metal ions such as Ba^{2+} , Co^{2+} and

117 Ni^{2+} . [46-48] Zhang et al. synthesized Na/Zn/Sn/S (NaZTS) quaternary metal sulfide nanosheets
118 for efficient adsorption of radioactive strontium ions. [49] NaZTS exhibited ultrafast kinetics with
119 an equilibrium time of 5 min, a broad active pH range (a removal rate of >98.4%), and a very low
120 Sr^{2+} desorption (< 0.04%) when employed as Sr^{2+} adsorbent in aqueous solution. The outstanding
121 ability of NaZTS to capture Sr^{2+} ions was attributed to ion exchange and strong Sr-S bonding
122 following the HSAB theory. The introduction of sulfur functional groups such as sulfonic
123 acids [50], sulfonates [51] and thiols [44] to the adsorbent is a promising way to prepare Sr^{2+} -
124 selective adsorbents. Nevertheless, the traditional sulfur-containing inorganic compounds have
125 fixed shapes and inherent brittleness, which make them difficult to be shaped and to accommodate
126 different purification devices.

127 In this study, both the HSAB strategy and the ion-imprinting strategy were employed to
128 develop a SA-based hydrogel adsorbent rich in thiol groups and owning Sr^{2+} -imprinted cavities to
129 enhance the adsorption performance and selectivity. The Sr^{2+} -imprinted cavities were introduced
130 by first using Sr^{2+} to complex the carboxyl groups of SA chains, whereby SA chains were
131 networked and thus a SA hydrogel (SA-Sr) was produced, and then ion-imprinted cavities can be
132 obtained if the Sr^{2+} were eliminated. However, the elimination of the Sr^{2+} crosslinks would
133 completely destroy the hydrogel structure. To avoid this, as well as to introduce abundant thiol
134 groups, a thiol-functionalized tetra-arm carboxylic crosslinker PA was synthesized and used to
135 covalently crosslink the SA-Sr hydrogel through *in situ* esterification, before the Sr^{2+} being
136 removed by acid elution to obtain a robust Sr^{2+} -imprinted hydrogel (SA-PA-H). The fabrication
137 and structure characterization of the SA-PA-H hydrogel was fully validated via a series of
138 techniques, and its adsorption performance of Sr^{2+} was evaluated in terms of isothermal adsorption,
139 effect of competitive ions, time, pH and temperature, etc. Moreover, the robustness of the SA-PA-
140 H hydrogel was assessed by rheology, and its application as adsorptive membranes for device
141 filtration, together with its regeneration potentials were also assessed.

142

143 **2 Experimental**

144 2.1 Chemicals

145 Pentaerythritol tetrakis(mercapto acetate) (90%, PET4A), strontium chloride hexahydrate

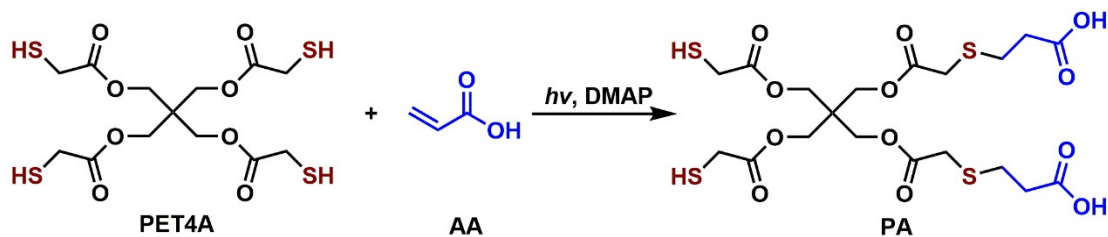
146 (99.5%, $\text{SrCl}_2 \cdot 6\text{H}_2\text{O}$), dioctyl sodium sulfosuccinate sodium (96%, AOT), benzoin dimethyl ether
147 (99%, DMPA) and sodium alginate (90%, SA) were purchased from Macklin. Acrylic acid (AA, \geq
148 99%), concentrated hydrochloric acid (36.0-38.0%), sulfuric acid (95.0-98.0%) and
149 dichloromethane were purchased from Sinopharm Chemical Reagent Co Ltd. All the reagents
150 were used as received.

151

152 2.2 Synthesis

153 2.2.1 Synthesis of thiol group-functionalized carboxylic modifier PA

154 PET4A (6.0076 g, 13.89 mmol), and acrylic acid (2.25 g, 31.2 mmol) were dissolved in 40
155 mL dichloromethane followed by an addition of DMPA (0.3965 g, 1.547 mmol) as the
156 photoinitiator. The solution contained in a round flask equipped with a water condenser was stirred
157 and exposed to continuous UV light (250-420 nm) for 2 h. The reaction (Scheme 1) generated a
158 mixture of PET4As functionalized with 1, 2, 3, or 4 carboxylic groups. The product was named PA
159 and used in the following steps without further refinement.



161 Scheme 1. Synthesis of PA from photo-initiated click reaction between PET4A and acrylic acid. A
162 typical product with two thiol groups participating in the reaction is illustrated.

163

164 2.2.2 Preparation of Sr^{2+} complexed sodium alginate (SA-Sr) hydrogel

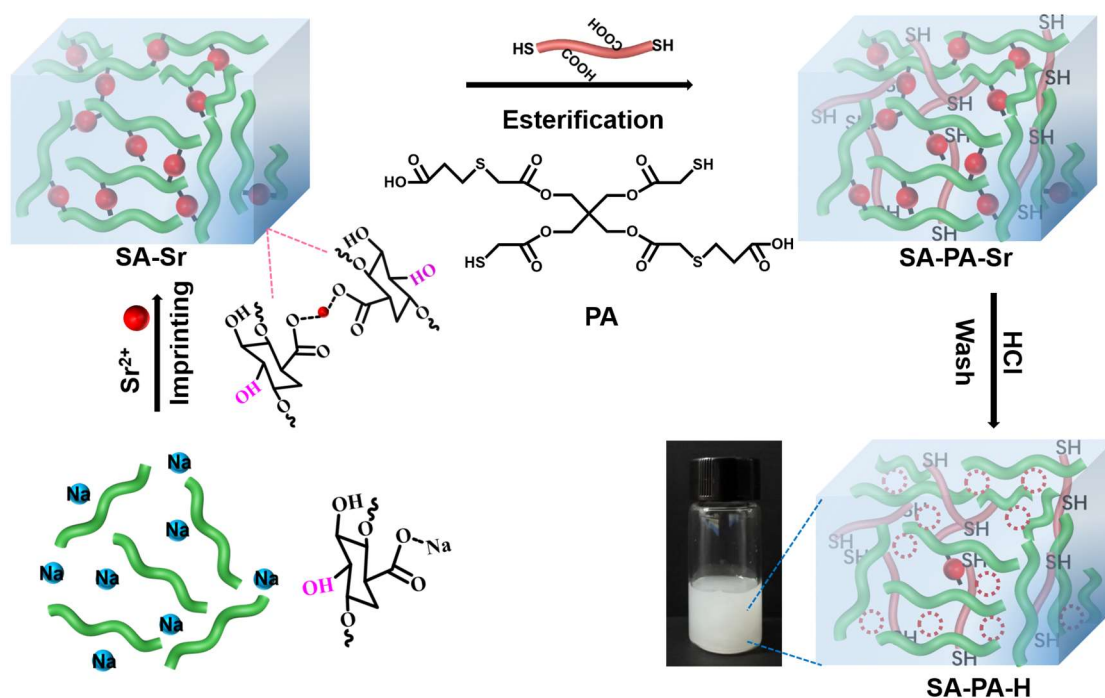
165 SA (1 g) was dissolved in 100 mL ultrapure water under stirring for 3 h to prepare a
166 homogenous and viscous SA solution in a flask for further use. AOT surfactant (1.25 g) was
167 dissolved in 50 mL dichloromethane followed by the addition of 25 mL of the as-prepared SA
168 aqueous solution (1 wt%). The mixture was ultrasonicated using a Sonic Dismembrator (BILON-
169 650Y) to prepare a W/O emulsion. Under stirring, 25 mL Sr^{2+} (0.5 wt% in water) was added
170 dropwise and solid hydrogels precipitated from the emulsion. The mixture was stirred for 24 h to
171 ensure full chelation between SA and Sr^{2+} , which generated an SA-Sr hydrogel. Then the emulsion

172 was centrifuged at 1000 rpm for 10 min to collect the hydrogel from the liquid and washed with
173 ultrapure water repeatedly to remove excess Sr^{2+} .

174

175 2.2.3 Fabrication of PA crosslinked and Sr^{2+} -imprinted composite hydrogel (SA-PA-H)

176 About 10.0 g of the as-prepared swollen SA-Sr hydrogel (solid content 0.25 g) was re-
177 dispersed in 20 mL ultrapure water followed by the addition of 20 mL dichloromethane solution of
178 4.0 g PA. 1.6 mL H_2SO_4 (0.5 wt%) was added dropwise to catalyze the esterification between the
179 hydroxyl groups in SA and the carboxyl groups in PA (Figure 1). After stirring for 24 h, the
180 chemically crosslinked hydrogel, named SA-PA-Sr, was fabricated and washed with ultrapure
181 water three times and separated by centrifugation at 1000 rpm for 10 min. Thereafter, the Sr^{2+}
182 previously chelated in the hydrogel was removed by dispersing SA-PA-Sr hydrogel in 50 mL
183 hydrochloric acid (0.5 mol/L). The dispersion was stirred for 24 h and then washed with ultrapure
184 water 3 times. The composite hydrogel product named SA-PA-H was obtained with a yield of
185 0.197 g (solid content) and 37.2%, assuming that two monosaccharide units chelating with one
186 Sr^{2+} was crosslinked by one PA molecules and Sr^{2+} was completely removed by acid wash. The
187 obtained SA-PA-H hydrogels were stored in swollen state in deionized water before use. The
188 synthesis procedure is illustrated in Figure 1.



189

190 Figure 1. Synthesis route of ion-imprinted thiol-functionalized SA-PA-H composite adsorbent
191 from SA matrix, PA crosslinker and Sr. The synthesis includes three steps: (1) The chelation
192 between Sr^{2+} and carboxylate groups in SA to prepare SA-Sr hydrogels. (2) Acid-catalyzed
193 esterification between hydroxyl groups in SA-Sr and carboxylic groups in PA to generate thiol-
194 functionalized cross-linked hydrogels SA-PA-Sr. (3) Preparation of ion-imprinted SA-PA-H
195 adsorbents by washing the SA-PA-Sr hydrogels with hydrochloride acid solutions to remove
196 chelated Sr^{2+} . The inset image displays the appearance of SA-PA-H hydrogel suspended in pure
197 water.

198

199 2.3 Materials characterization

200 2.3.1 Nuclear Magnetic Resonance (NMR). NMR spectra were acquired on a Bruker AV 400M
201 spectrometer. CDCl_3 was used as the solvent with tetramethylsilane as the internal reference.

202 2.3.2 Fourier Transform Infrared (FT-IR). The infrared absorption spectra of samples were
203 collected on a Nicolet IS50 Fourier transform infrared spectrometer (Thermo Scientific). KBr
204 tablet methods were applied to samples. Every sample was scanned 32 times with a resolution of 4
205 cm^{-1} in the range of 400 cm^{-1} to 4000 cm^{-1} and the average infrared signal was adopted.

206 2.3.3 X-ray Photoelectron Spectroscopy (XPS). The sample chemical bonds were analyzed using
207 an ESCANLAB Xi⁺ XPS system with a monochromatic Al $K\alpha$ X-ray source. An electron/ion gun
208 was used to compensate for charge build-up on the sample during measurement. The XPS peaks
209 were fitted using the CasaXPS software, and the binding energy was corrected with a reference to
210 C 1s at 284.5 eV.

211 2.3.4 Field emission scanning electron microscopy (FE-SEM) and energy dispersive X-ray
212 spectroscopy (EDX) mapping. The morphologies of the SA-Sr, SA-PA-Sr, and SA-PA-H samples
213 were observed with a Hitachi Regulus 8100 cold FE-SEM coupling with EDX. The samples were
214 sprayed with gold before the observation. The accelerating voltage was 5.0 kV and the current was
215 10 μA . The Octane Elect Plus X-ray spectrometer (EMAX, USA) in conjunction with SEM was
216 used to analyze EDX elements mapping of the samples with a current of 10 μA and an
217 accelerating voltage of 10.0 kV.

218 2.3.5 Inductively Coupled Plasma-Optical Emission Spectroscopy (ICP-OES). The concentration
219 of Sr^{2+} was measured by ICP-OES (Singapore, PerkinElmer) at 407.771 nm as a specific emission

220 wavelength of Sr²⁺. Solution samples subjected to the test were the initial Sr²⁺ solution and
221 supernatants collected after adsorption experiments. The concentration of Sr²⁺ was diluted to 0-1
222 ppm and the solution was filtered with a 0.45 µm filter before the test. The quantification was
223 done by converting the intensities of the samples according to the standard curve established.

224

225 2.4 Sr²⁺ Adsorption

226 2.4.1 Batch adsorption

227 All experiments were conducted using polypropylene vials to avoid Si contamination from
228 the glassware and any potential Sr²⁺ adsorption onto the glassware, as previously reported.¹⁰ The
229 total volume of liquid used for the adsorption was 5 mL and the solid content of hydrogel
230 adsorbents in liquid was fixed at 1 g/L throughout all the tests. SrCl₂ (⁸⁷Sr) was dissolved in
231 deionized water to simulate radioactive wastewater polluted by ⁹⁰Sr²⁺. Solutions with Sr²⁺
232 concentrations varying from 10 ppm to 500 ppm were prepared by diluting a 1000 ppm Sr²⁺ stock
233 solution. In polypropylene centrifuge tubes, suspension of hydrogel adsorbents in Sr²⁺ solution
234 was thoroughly mixed using an orbital shaker at 200 rpm for 24 h to promote the diffusion of Sr²⁺
235 into the hydrogel. Then sample tubes were centrifuged at 11,000 rpm for 15 min, and the
236 supernatant was decanted and passed through a 0.45 µm syringe filter. The concentration of Sr²⁺ in
237 the supernatant and the initial solution was determined by ICP-OES. The concentration was
238 calibrated using SrCl₂ solutions of known concentrations (0, 0.1, 0.3, 0.5, 0.8 and 1 ppm) with a
239 correlation coefficient larger than 0.999 (Figure S1).

240 The amount of Sr²⁺ adsorbed by the solids, q (mg/g), was determined using the following
241 equation:

$$242 \quad q = \frac{(C_o - C_e)V}{m} \quad (1)$$

243 where C_o and C_e are the initial and equilibrium Sr²⁺ concentrations, respectively; V (L) is the
244 volume of the suspension, and m (g) is the amount of the adsorbent (mass of solid fraction of the
245 hydrogel).

246 Adsorption kinetics was measured at room temperature using a fixed initial Sr²⁺ concentration
247 ($C_o = 100$ ppm) with a solid-to-liquid ratio of 1 g/L under neutral conditions. The concentration of
248 Sr²⁺ in the solution was measured at adsorption time varying from 10 min to 24 h. For

249 investigation of pH dependence, hydrogel adsorbents were tested across a pH range of 2 to 12 with
250 $C_o = 100$ ppm. The pH was adjusted by 1.0 M HCl or 1.0 M NaOH. All other experimental
251 parameters followed the same procedure described above.

252 The selectivity of Sr^{2+} in the presence of competitive ions was investigated by suspending 1
253 g/L hydrogel adsorbents in solutions containing Sr^{2+} (10 ppm) with varying Na^+ concentrations
254 (100, 200, 400 and 800 ppm). After the samples were shaken for 24 h, the supernatant was
255 separated and its Sr^{2+} concentration was determined by ICP-OES.

256

257 2.4.2 Desorption of Sr^{2+} and reuse of hydrogel

258 The regeneration of SA-PA-H hydrogel was investigated. For this, the hydrogel adsorbents
259 were first subjected to Sr^{2+} adsorption experiment in Sr^{2+} solution (100 ppm) with a solid content
260 of 1 g/L, orbital shaking at 200 rpm for 12 h to reach saturated adsorption. Then the saturated
261 adsorbents were fully washed with ultrapure water and submerged in 100 mL hydrochloric acid
262 (0.5 mol/L) with a solid-to-liquid ratio of 4 g/L for 6 h under stirring to remove the adsorbed Sr^{2+} .
263 After washing and centrifuging, the regenerated hydrogel adsorbent was collected and added to a
264 Sr^{2+} solution (100 ppm) again to repeat the adsorption experiment under identical conditions (i.e.
265 solid content 1g/L, orbital shaking at 200 rpm for 12 h). After the sorption experiments and
266 centrifugation, the concentration of Sr^{2+} in the filtered supernatant was measured by ICP-OES to
267 determine the adsorption capability of the regenerated adsorbent. The same procedure was
268 repeated four times to investigate the stability and reusability of the regenerated SA-PA-H
269 hydrogel as an adsorbent.

270

271 2.4.3 Filtration test

272 Filtration membranes composed of SA-PA-H hydrogel were prepared by depositing the
273 hydrogel at varying amounts (0.02-0.1 g) on a Nylon 6 filtration membrane with a pore size of 0.2
274 μm held by the sand core in a funnel with an inner diameter of 30 mm. Then the tunnel holding the
275 SA-PA-H membrane was attached to a suction filter bottle connected to a vacuum pump. Then 100
276 mL Sr^{2+} (0.78 ppm) solution was added to the tunnel and the pump was turned on to make a
277 vacuum. By adjusting the pressure in the suction bottle, the Sr^{2+} solution was filtrated through the

278 membrane driven by a pressure difference across the membrane. After the filtration was completed,
279 the filtration time was recorded to calculate the flux. The effluents were taken from the suction
280 bottle and Sr^{2+} concentrations were determined by ICP-OES.

281

282 2.5 Rheology test

283 Rheological properties of the fabricated hydrogel were measured by an HR-20 stress-
284 controlled rheometer (TA measurement, USA) with a disposable 25 mm diameter parallel plate.
285 The geometry gap was set to 1220 μm . A strain sweep test was first carried out at an oscillation
286 frequency (ω) of 1 rad/s, and then frequency sweep tests were performed within a linear
287 viscoelastic regime with a strain amplitude (γ) of 0.5%. To investigate the relaxation/recovery
288 behavior of the hydrogel in response to an applied shear strain, the hydrogel was subjected to an
289 oscillation shear with following strain amplitudes and durations (in parentheses) executed in
290 sequence: 0.1% (300 s), 1000% (300 s), 0.1% (600 s), 1000% (300 s), 0.1% (600 s), 1000% (300
291 s), 0.1% (1200 s). The temperature was maintained at 25 °C throughout all the measurements.

292

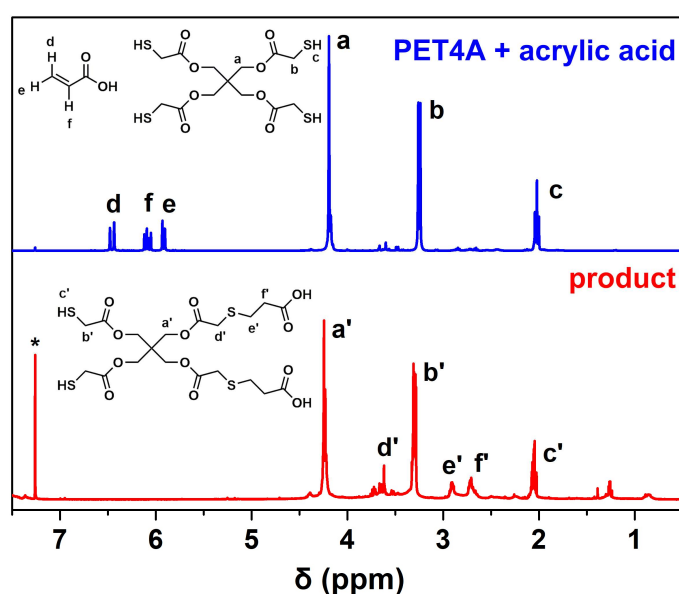
293 **3 Result and Discussion**

294 3.1 Synthesis and characterization of thiol-functionalized Sr^{2+} imprinted SA hydrogels

295 3.1.1 Synthesis of thiol group-functionalized crosslinker PA

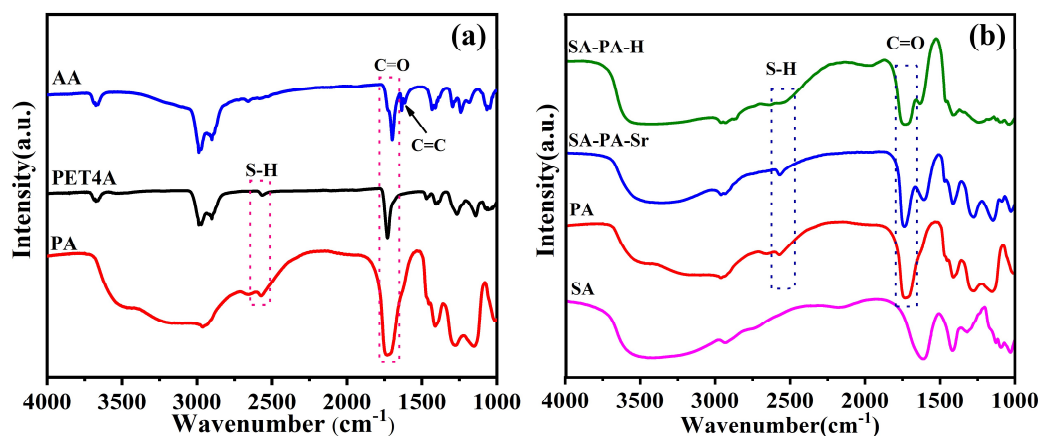
296 To develop an efficient adsorbent for Sr^{2+} removal with a high sorption capacity and
297 selectivity, strategies of Sr^{2+} imprinting and introduction of as many thiol groups as possible to the
298 adsorbent were considered in the present study based on sodium alginate (SA) hydrogel. To
299 introduce thiol groups into a SA hydrogel, PET4A owing 4 thiol groups in one molecule was
300 adopted as a crosslinker of SA via modification with acrylic acid (AA) to gain carboxylic groups
301 for esterification with SA (Scheme 1). The modification was realized by a one-step thiol-ene click
302 reaction between thiol groups of PET4A and about 2.2 equivalents of AA, with the product named
303 PA (Scheme 1). ^1H NMR spectra of the products and the substrate mixtures dissolved in CDCl_3
304 were compared in Figure 2. In the spectrum of substrates, signals with chemical shifts at 5.95-6.65
305 ppm were ascribed to alkenyl protons of acrylic acids (H^{d} , H^{f} and H^{e}), which disappeared in the
306 spectrum of PA products indicating a full conversion of acrylic acid. In the spectrum of PA product,

307 peaks ascribed to methylene protons (H^e and H^f) near thioester and carboxylic groups at 2.71
308 ppm and 2.93 ppm respectively are observed. At the same time, the peak at 2.06 ppm ascribed to
309 thiol groups remained but slightly weakened after the reaction. Using the integral of H^a ($H^{a'}$) as an
310 internal standard, the conversion of the thiol group in PET4A was determined to be 34%. The
311 results showed that an average of 1.8 thiol groups remained in every modifier/crosslinker PA
312 molecule in the mixture product. The 1H NMR demonstrated the successful synthesis of thiol- and
313 carboxylic-functionalized PA, being an ideal crosslinker of SA hydrogel whilst introducing a large
314 amount of thiol groups to improve the Sr^{2+} sorption capacity of the hydrogel.



315
316 Figure 2. 1H NMR spectra of PET4A mixed with 2.2 equiv. acrylic acid (up) and product PA
317 dissolved in $CDCl_3$ (bottom, *: CH_3Cl).

318
319 The FTIR spectrum of PA was compared with the spectra of PET4A and acrylic acid as
320 shown in Figure 3a. In contrast to the evident peak around 1650 cm^{-1} in the AA spectrum, the
321 characteristic adsorption of $C=C$ disappeared in the IR spectrum of PA, which further validated the
322 full conversion of acrylic acid. The dispersive adsorption band ascribed to O-H vibration around
323 3200 cm^{-1} increased due to the prevalent carboxylic group and adsorbed moisture in PA. The
324 adsorption signal at 2574 cm^{-1} revealed the presence of the thiol group which was also observed in
325 the spectrum of PET4A. The FTIR spectra confirmed the synthesis of PA containing essential thiol
326 groups to provide potential adsorption sites for Sr^{2+} when used to fabricate SA hydrogel adsorbent.



327

328 Figure 3. (a) FTIR spectra of AA, PET4A and PA. The characteristic S-H, C=O and C=C vibration
329 adsorption peaks are highlighted by the boxes and an arrow. (b) FTIR spectra of SA, PA, SA-PA-
330 Sr and SA-PA-H. The characteristic S-H and C=O vibration adsorption peaks are highlighted by
331 the boxes.

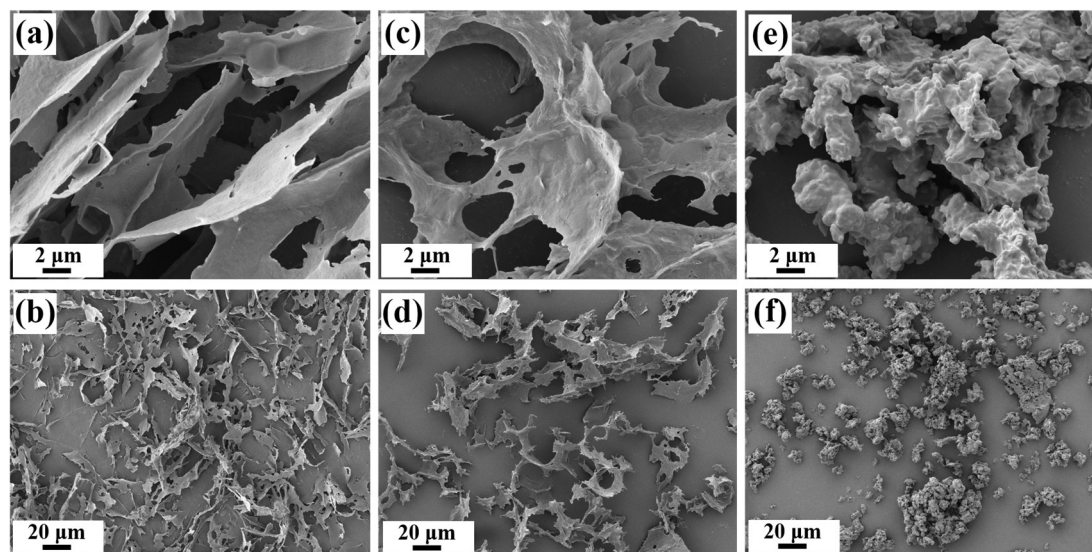
332

333 3.1.2 Fabrication of the thiol-rich and Sr²⁺ imprinted hydrogel SA-PA-H

334 A three-step route was carried out to fabricate a thiol-rich and Sr²⁺ imprinted SA-based
335 hydrogel adsorbent as illustrated in Figure 1. The synthesis procedure has been described in
336 section 2.2.2 and 2.2.3 in detail. The relatively low yield (37.2%) indicates incomplete reaction
337 between SA-Sr and PA, that is, the SA-PA-H was far from being fully crosslinked. The unreacted
338 hydroxyls and pendant chains in the hydrogels assure good hydrophilicity and enough free volume
339 of the adsorbents for the diffusion of water molecules and Sr²⁺ ions. FTIR was applied to examine
340 the structures of the prepared SA-PA-Sr and SA-PA-H hydrogels (Figure 3b). In the spectrum of
341 SA-PA-Sr, there were two carbonyl peaks at 1733 cm⁻¹ and 1631 cm⁻¹. The peak at 1631 cm⁻¹ was
342 ascribed to the vibration of C=O bonds in the carboxyl group, which is confirmed by the spectrum
343 of SA containing only carboxyl groups rather than ester groups. Meanwhile, the C=O vibration
344 peak at 1733 cm⁻¹ corresponded to ester carbonyl groups. The characteristic carbonyl absorption
345 signal was absent in the spectrum of SA but observed in the spectrum of PA which was abundant
346 with the ester group. The reduction of carboxyl absorption and the co-existence of the ester group
347 and the carboxyl group suggested partial esterification of SA with PA. The unreacted carboxyl
348 group was arguably due to the chelation of carboxyl groups with Sr²⁺, which prevented them from
349 esterification. The esterification between carboxylic groups and hydroxyl groups was also

350 supported by a slightly decreased broad O-H absorption band in the range of 3000–3500 cm^{-1} in
351 SA-PA-Sr as compared with the spectrum of SA. On the other hand, S-H vibration signals were
352 clearly observed in SA-PA-Sr indicating the existence of thiol groups in SA-PA-Sr. The FTIR
353 spectrum demonstrated that SA-PA-Sr has the expected chemical structure. The partial
354 esterification allowed SA-PA-Sr to be protonated to SA-PA-H as the intact ion-imprinted
355 crosslinked hydrogel-bearing carboxyl group and thiol group as dual adsorption sites. In contrast,
356 SA-Sr without covalent crosslinking would lose integrity in hydrochloric acid due to the break of
357 Sr^{2+} chelation.

358 Morphologies of SA-Sr, SA-PA-Sr and SA-PA-H were investigated by SEM (Figure 4). As
359 Figure 4a&b displayed, the SA-Sr hydrogel was a well-defined porous network composed of
360 layered structures with smooth surfaces. The micron-sized pores were likely to be the result of
361 water loss from swelled hydrogel during the preparation of SEM samples. SA-PA-Sr was also a
362 three-dimensional (3D) porous network composed of platelets (Figure 4c&d). However, the
363 surface of the platelets is rougher and more porous than SA-Sr. The difference in the morphology
364 could be explained by the modification of hydrogel by esterification crosslinking of SA and PA,
365 which led to slightly contracted and wrinkled surfaces. After elution with acid, the hydrogel (SA-
366 PA-H) has a morphology consisting of collapsed agglomerates (Figure 4 e&f). This is because in
367 the SA-Sr and SA-PA-Sr hydrogels, Sr^{2+} acted as non-covalent crosslinkers, by complexing with
368 carboxyl groups from different SA chains, to construct the 3D network structures of the hydrogels,
369 which were however eliminated during the acid elution. The exchange of Sr^{2+} by proton largely
370 reduced the number of the non-covalent crosslinkers, triggering partial agglomerates of the
371 framework platelets. In spite of this, the SA-PA-H still maintained the form of hydrogel containing
372 a large content of water (i.e. 94.4%) as demonstrated above. The morphology alternation confirms
373 the efficient removal of the Sr^{2+} via acid elution, forming ion-imprinted cavities in the SA-PA-H
374 as adsorption sites.

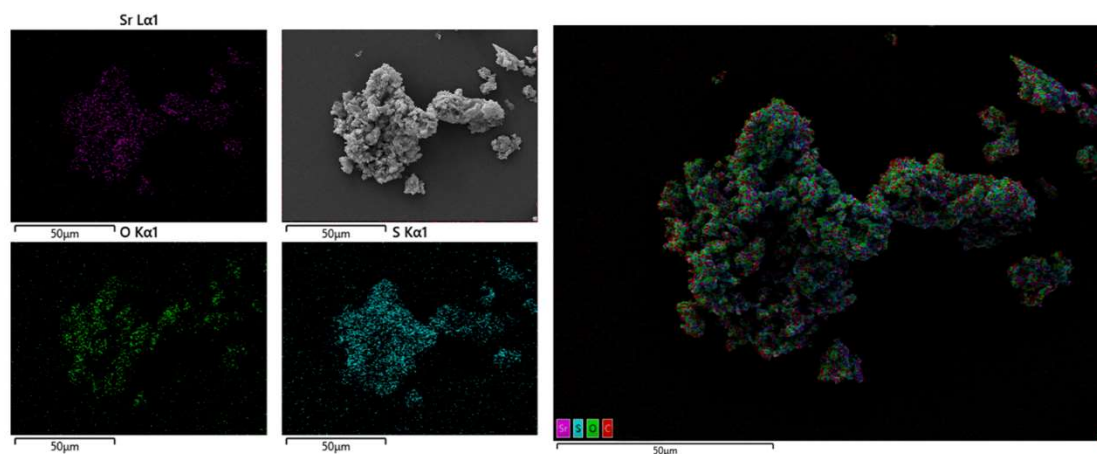


375

376 Figure. 4. SEM images of SA-Sr (a-b), SA-PA-Sr (c-d) and SA-PA-H (e-f) samples coated by gold
377 and observed at different magnifications.

378

379 The distribution of Sr, O and S elements in SA-Sr (Figure S2), SA-PA-Sr (Figure S3) and
380 SA-PA-H (Figure 5) hydrogels were further studied by SEM-EDX to confirm the successful
381 synthesis of SA-PA-H composite hydrogels with Sr-imprinted moieties. In the EDX elemental
382 map of SA-Sr (Figure S2), the intensity of S elements was barely stronger than the background
383 noise. The observed trace amount of sulfur likely stemmed from sulfonic groups in the emulsifier
384 AOT used during the preparation of SA-Sr hydrogel. After modification by PA, SA-PA-Sr and SA-
385 PA-H hydrogel showed strong signals of S elements in its EDX elemental map (Figure S3&5),
386 which indicated successful incorporation of PA along with abundant thiol groups into the
387 hydrogels. For the Sr element, both EDX elemental maps of SA-Sr and SA-PA-Sr exhibited strong
388 Sr signals (Figure S2&S3), implying the prevalence of Sr^{2+} adsorbed in the hydrogels. Compared
389 with SA-Sr and SA-PA-Sr, the intensity of Sr element diminished obviously in the EDX elemental
390 map of SA-PA-H (Figure 5). The results revealed that the majority of Sr^{2+} chelated in the hydrogel
391 was removed by washing it with hydrochloric acid. The intensities of O elements remained almost
392 unchanged through the preparation, suggesting backbones of the hydrogels remained stable. The
393 EDX elemental mapping validated the successful fabrication of the thiol-rich and ion-imprinted
394 SA-PA-H hydrogel adsorbent by the three-step protocol.

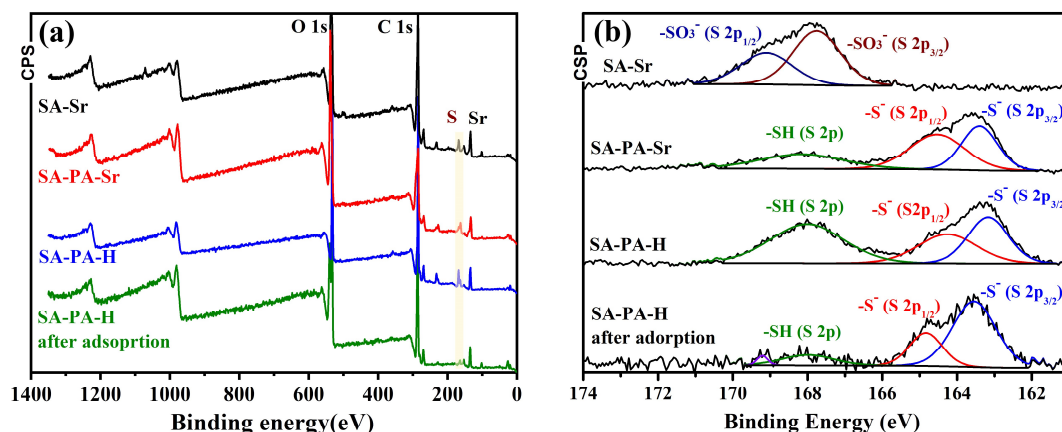


395

396 Figure. 5. EDX elemental maps along with the SEM image of SA-PA-H showing the distribution
397 of Sr, O and S in the sample by purple, green and cyan, respectively.

398

399 To further reveal the composition and interaction between components (e.g. SA, PA and Sr^{2+})
400 in the composite hydrogels, XPS spectra (Figure 6) along with high-resolution S2p spectra (Figure
401 6b) of the hydrogels were investigated. Due to spin-orbital coupling, deconvolution of every
402 merged S2p peak presented closely coupled peaks ascribed to S2p_{1/2} and S2p_{3/2} photoelectron lines
403 (Figure 6b). In the spectrum of SA-Sr hydrogel, the peaks at 169.1 and 167.7 eV can be assigned
404 to the sulfonate ($-\text{SO}_4^-$) group introduced by trace AOT during preparation. The existence of
405 sulfonate peaks explained the trace sulfur signal in the EDX mapping of SA-Sr (Figure S2). After
406 PA was added to SA-Sr followed by esterification, three deconvoluted S2p peaks with different
407 binding energies at 168.2, 164.6 and 163.4 eV appeared in the XPS spectrum of SA-PA-Sr (Figure
408 6b). The peaks at 164.6 and 163.4 eV originated from thiol anion S ($-\text{S}^-$) chelating with Sr^{2+} , and
409 the peak at 168.2 eV was attributed to protonated thiol group S ($-\text{SH}$) without Sr^{2+} bonded, which
410 overlapped with sulfonate peaks and was too weak to be deconvoluted to S2p_{1/2} and S3p_{3/2} peaks.
411 The three characteristic peaks were also observed in the XPS spectrum of SA-PA-H (Figure 6b).
412 However, the ratio between the integral area of the $-\text{S}^-$ peaks over the area of the $-\text{SH}$ peak
413 decreased in SA-PA-H as compared to SA-PA-Sr, which indicated the dissociation of Sr^{2+} from
414 sulfur and an increased amount of protonated sulfhydryl group. The results demonstrated that PA
415 successfully modified the SA hydrogel and introduces sulfur sites for Sr^{2+} binding. By washing
416 with acid, the Sr^{2+} was partially removed from SA-PA-Sr and the sulfhydryl group was recovered
417 in SA-PA-H to serve as potential Sr^{2+} adsorption moieties.

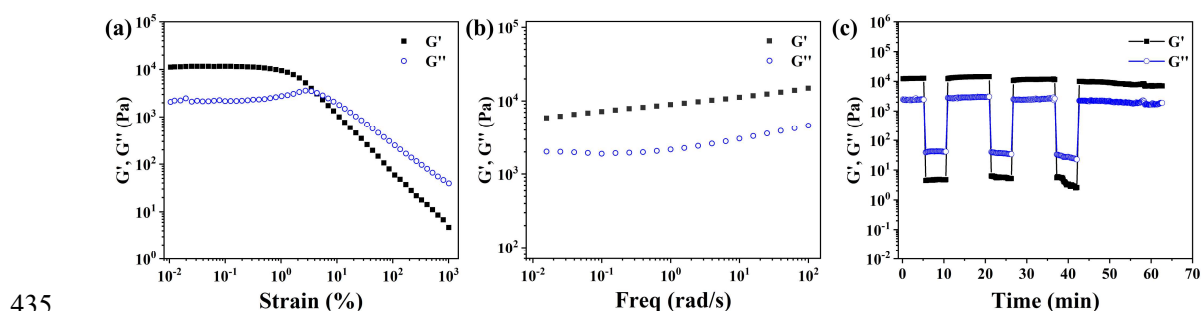


418
 419 Figure 6. (a) XPS survey spectra of SA-Sr, SA-PA-Sr, SA-PA-H and SA-PA-H after adsorption of
 420 Sr^{2+} . The signals of electrons in O 1s, C 1s and Sr are labeled close to the corresponding peaks,
 421 and the electron signal of S is highlighted by the yellow box. (b) High resolution XPS spectra of S
 422 2p peaks of SA-Sr, SA-PA-Sr, SA-PA-H and SA-PA-H after adsorption of Sr^{2+} . The S 2p peaks are
 423 deconvoluted to separated peaks labeled with corresponding sulfur-containing groups.

424

425 3.2 Hydrogel rheology

426 The viscoelasticity of the SA-PA-H hydrogel was investigated for its further application as
 427 membrane materials. The swollen hydrogel was subjected to a strain sweep test and a frequency
 428 sweep test at room temperature ($T = 25^\circ\text{C}$). The results are shown in Figure 7a&b. It is found that
 429 the SA-PA-H is elastic as storage modulus (G') > loss modulus (G'') at small strain, while yield
 430 beyond a critical strain where $\gamma = 4.4\%$ and $G' = G''$. The frequency sweep confirmed that the SA-
 431 PA-H is strongly elastic with $G' \approx 10$ kPa, and G' is larger than G'' throughout all shear
 432 frequencies. The elasticity is ascribed to the network structure of SA-PA-H crosslinked by
 433 esterification. Moreover, there would be massive hydrogen bonding formed between ester,
 434 carboxyl, hydroxyl and thiol groups, which can further strengthen the hydrogel.



435

436 Figure 7. (a) G' and G'' of SA-PA-H hydrogel in strain sweep test under oscillation shear at 1 rad/s
437 frequency. (b) Frequency-dependent G' and G'' of SA-PA-H under oscillation shear at a strain
438 amplitude of 0.5%. (c) Step-strain behavior of SA-PA-H indicating a reversible recovery in G' and
439 G'' at small strain (0.1%) after yielding at large strain (1000%).

440

441 A step-strain experiment was performed to demonstrate the self-recovery ability and the
442 processability of the SA-PA-H hydrogel (Figure 7c), which was expected to be realized by
443 reversible dissociation and reformation of non-covalent hydrogen bonds in the hydrogels. The
444 swollen hydrogel was subjected to an oscillation shear by small strains (0.1%) and large strains
445 (1000%) alternately and the result is shown in Figure 7c. Initially, the hydrogel was sheared at a
446 small strain below the yield point, where $\gamma = 0.1\%$ at $\omega = 1.0$ rad/s. The G' and G'' of the SA-PA-H
447 are about 13 kPa and 2.4 kPa, respectively, indicating that the hydrogel is in a quasi-solid state. As
448 the strain increased to 1000% beyond the yield point, G' and G'' significantly decreased to ~ 4.4 Pa
449 and ~ 40 Pa (i.e. $\tan \delta \equiv G''/G' \sim 8.8-9.3$), respectively (Figure 7c), indicating the hydrogel was in a
450 liquid state. Nevertheless, when the strain was decreased to 0.1%, the G' recovered to 13 kPa
451 immediately and $\tan \delta$ of the hydrogel decreased to 0.21 at the same time, revealing that the
452 hydrogel recovered to a quasi-solid state. In the subsequent strain-step alternations, repeated
453 yielding and recovery of the hydrogel were observed, demonstrating no significant loss in the
454 viscoelasticity nor a change in the rate of structural reformation of the hydrogel even underwent
455 repeated large strain deformations. Only a small drop of G' and G'' was observed in the last strain-
456 step test (4th cycle) due to the loss of water from the hydrogel after long-time measurement (Figure
457 7c). The behavior demonstrated that SA-PA-H hydrogel can be readily processed and reshaped
458 without damaging its mechanical properties thanks to the presence of massive hydrogen bonds
459 between polymer chains. The robustness with strong elasticity allows us to reshape the fabricated
460 hydrogel into membrane filtration devices for facile water purification, which will be discussed in
461 section 3.3.5.

462

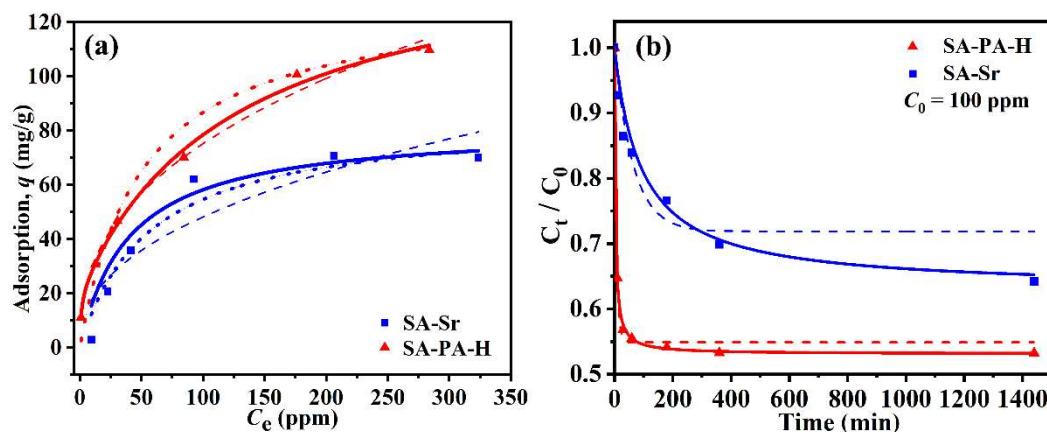
463 3.3 Strontium adsorption

464 3.3.1 Isothermal adsorption

465 To investigate the ability of the hydrogels to remove radioactive Sr^{2+} from polluted water, we
 466 examined the isothermal adsorption behavior of SA-PA-H and SA-Sr in simulated Sr^{2+} solutions.
 467 The well-known Langmuir model was applied to describe Sr^{2+} adsorption behaviors, assuming: (i)
 468 surface monolayer sorption, (ii) a finite number of binding sites, (iii) uniform sorption energies,
 469 and (iv) no transmigration of sorbates in the plane of the surface. Adsorption (q , mg of adsorbate/g
 470 of adsorbent) of Sr^{2+} by the hydrogels was plotted against equilibrium Sr^{2+} concentration (C_e , ppm)
 471 in solution as shown in Figure 8a. The adsorption data were fitted by the Langmuir isotherm:

$$472 \quad q = \frac{bq_m C_e}{1 + bC_e} \quad (2)$$

473 where b is the Langmuir constant related to the affinity coefficient of binding sites (L/mg) and q_m
 474 is the maximum sorption capacity of Sr^{2+} for the adsorbent.



475 Figure 8. (a) Sr^{2+} adsorption isotherm with adsorption (q) plotted against the equilibrium
 476 concentration (C_e) for SA-PA-H and SA-Sr. The dotted, solid and dashed lines are fittings
 477 according to single-site Langmuir, dual-site Langmuir and Freundlich isotherm models,
 478 respectively; (b) Sr^{2+} adsorption kinetics by SA-PA-H and SA-Sr at $C_0 = 100$ ppm. The dashed and
 479 solid lines are fittings according to PFORE and PSORE, respectively.
 480

481
 482 As shown in Figure 8a, the adsorption data of both SA-PA-H and SA-Sr can be fitted by
 483 single-site Langmuir isotherm with a decent correlation coefficient (dotted lines, $R^2 = 0.961$ and
 484 0.971 , respectively). Table 1 lists the Langmuir fitting parameters, q_m and b of the hydrogels. As
 485 can be seen, the SA-PA-H had an excellent adsorption capacity with a high q_m value of 130.0 mg/g.
 486 In contrast, the q_m of SA-Sr is much lower (87.9 mg/g) due to that a majority of the sorption sites
 487 (e.g. the carboxyl groups in SA) were already occupied by Sr^{2+} . The adsorption results indicated

488 the importance of the Sr²⁺-imprinted cavities obtained via acid elution of the SA-PA-Sr to remove
 489 pre-complexed Sr²⁺ and of the thiol groups introduced by PA in enhancing the Sr²⁺ adsorption
 490 capacity of the SA-PA-H hydrogel.

491

492 Table 1. Single-site, dual-site Langmuir and Freundlich isotherm models for Sr²⁺ adsorption

Sample	Langmuir model			dual-site Langmuir model					Freundlich model		
	q_m (mg/g)	b (L/mg)	R^2	q_{m1} (mg/g)	b_1 (L/mg)	q_{m2} (mg/g)	b_2 (L/mg)	R^2	K_F (mg ¹⁻ⁿ •L ⁿ /g)	n	R^2
SA-PA-H	130.0	0.02	0.961	20.5	0.98	131.2	0.007	0.991	11.92	0.400	0.989
SA-Sr	87.9	0.01	0.971	25.7	0.01	62.2	0.011	0.939	6.750	0.427	0.813

493

494 Considering that in the SA-PA-Sr hydrogel both the ion-imprinted cavities located at the
 495 carboxyl groups in SA chains and the thiol groups introduced from PA would contribute to the Sr²⁺
 496 adsorption as different sorption sites with different sorption energies, the dual-site Langmuir
 497 adsorption model was also used to fit the adsorption data:

$$498 \quad q = \frac{b_1 q_{m1} C_e}{1 + b_1 C_e} + \frac{b_2 q_{m2} C_e}{1 + b_2 C_e} \quad (3)$$

499 where the fitting parameters b_1 and b_2 , and q_{m1} and q_{m2} represent the affinity coefficients (L/mg)
 500 and the maximum sorption capacity (mg/g) for the two types of sorption sites, respectively.

501 As shown in Figure 8a (the solid lines) and Table 1, the Dual-site Langmuir model better
 502 described the adsorption behavior of the SA-PA-H with a higher R^2 (0.991) than the single-site
 503 Langmuir model (0.961). On the other hand, the adsorption data of SA-Sr was poorly fitted by the
 504 dual-site Langmuir model ($R^2 = 0.939$), suggesting that the SA-Sr was a single-site adsorbent
 505 (Figure 8a). The fitting results validated the two sorption sites for Sr²⁺ in the SA-PA-H hydrogel,
 506 that is, the thiol group and the ion-imprinted carboxyl group, with a total maximum adsorption
 507 capacity (q_m) of 151.7 mg/g. This well meets the purpose of designing the SA-PA-H using ion-
 508 imprinting and thiol-rich strategies to enhance the Sr²⁺ adsorption performance. The role of the
 509 thiol group was also confirmed by the XPS measurement of the SA-PA-H hydrogel after the
 510 adsorption experiment. The peak ascribed to the thiol group (-SH) diminished while the peaks of
 511 thiol anion (-S⁻) were amplified when compared to the signal before adsorption (Figure 6b),
 512 confirming the adsorption of Sr²⁺ to the thiol group by forming S⁻(1/2Sr²⁺) complex.

513 The adsorption data are also fitted to the Freundlich isotherm:

$$514 \quad q = K_F c_e^n \quad (4)$$

515 where K_F is the Freundlich sorption coefficient and n is the adsorption intensity [52].

516 As shown in Figure 8a (dashed lines), the Fredulich isotherm successfully captures the
517 adsorption behaviors of SA-PA-H ($R^2 = 0.989$, Table 1) while it fails to fit the adsorption data of
518 SA-Sr ($R^2 = 0.813$). The Freundlich isotherm describes multilayer adsorption and assumes
519 exponential decay in the energy distribution of heterogeneous adsorbed sites [52]. The excellent
520 fitting of the Freundlich isotherm to SA-PA-H adsorption data but not the SA-Sr confirms the
521 exsistence of more than one type of adsorption sites in SA-PA-H, that is, the carboxylic and thiol
522 groups. This is consistent with the dual-site Langmuir model, and the values of n are lower than 1,
523 indicating a nonlinear sorption on the adsorbents.

524 Moreover, to confirm the essential role of PA in the fabrication of highly efficient adsorbent,
525 a control sample of Sr^{2+} -imprinted SA hydrogel without PA component, named SA-H, was
526 prepared by eluting SA-Sr with hydrochloride acid, and its Sr^{2+} adsorption performance was
527 evaluated, with data shown in Figure S4. Since the data failed to be fitted by Langmuir models
528 (see discussion in the Supporting Information), the comparison of adsorption performance
529 between SA-H and SA-PA-H, SA-Sr was done at given C_e s. The general trend clearly showed a
530 worse adsorption capability of the SA-H than SA-PA-H. For example, at $C_e = 250$ ppm, the q of
531 SA-H was only 48 mg/g, while q of SA-PA-H was ~ 103.9 mg/g. Surprisingly, at a given C_e , the q
532 of SA-H was even smaller than that of SA-Sr whose carboxyl groups had been chelated with Sr^{2+} .
533 One possible reason was that during the acid elution, progressive esterification between carboxyl
534 groups and hydroxyl groups in SA would occur as catalyzed by acid, which can significantly
535 abolish the adsorption sites of carboxyl groups. In contrast, in the presence of PA, hydroxyl groups
536 in SA had been consumed by esterification with carboxyl groups from PA. Thus, the carboxyl
537 groups of SA were kept intact and imprinted by Sr^{2+} during the acid elution process, potentially
538 acting as selective adsorption sites for Sr^{2+} . This indicates that PA is an indispensable modifier for
539 SA to achieve the successful preparation of efficient Sr^{2+} -imprinted adsorbents.

540

541 3.3.2 Adsorption kinetics

542 Sr^{2+} adsorption kinetics of SA-PA-H and SA-Sr were studied. As shown in Figure 8b, SA-PA-
 543 H adsorbed Sr^{2+} in a very fast manner, with the adsorption equilibrium almost reached in 30 min.
 544 In contrast, Sr^{2+} adsorption of SA-Sr was relatively slow, having the adsorption equilibrium even
 545 not fully reached in the experiment time (12 h). To quantify the adsorption kinetics of the
 546 hydrogels, the data were fitted by a pseudo-first order rate equation (PFORE) and a pseudo-second
 547 order rate equation (PSORE):

$$548 \quad \log(q_e - q_t) = \log(q_e) - \frac{k_1}{2.303} t \quad (5)$$

$$549 \quad \frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (4-6)$$

550 where q_t and q_e are adsorbed Sr^{2+} quantities (mg/g) at time t and equilibrium respectively, and k_1
 551 and k_2 are a first order and a second order rate constant, respectively.

552 The fitting parameters are shown in Table 2. The PSORE fitted the adsorption kinetics of SA-
 553 PA-H very well with $R^2 > 0.99$, while the PFORE failed to describe the adsorption kinetics of both
 554 the SA-PA-H and the SA-Sr hydrogels. The possible reason for the preference to the PSORE
 555 kinetics instead of PFORE kinetics is that the adsorption of a single Sr^{2+} is a second order reaction
 556 involving physicochemical interactions between two phases, that is, a Sr^{2+} cation from solution
 557 and a binding group, i.e. carboxylic or thiol group, from solid phase. In the fitting results based on
 558 PSORE, significantly higher k_2 ($0.67 \text{ g mg}^{-1} \text{ min}^{-1}$) and q_e (46.9 mg g^{-1}) of SA-PA-H than SA-Sr
 559 indicated an excellent adsorption activity of the SA-PA-H. The enhanced adsorption kinetics of the
 560 SA-PA-H were attributed to the multiple sorption sites and the likely promotion by Sr^{2+} -imprinting
 561 and the additional adsorption sites of thiol groups.

562

563 Table 2 Fitting parameters of PFORE and PSORE for Sr^{2+} sorption by SA-PA-H and SA-Sr

Sample	PSORE			PFORE		
	k_2 ($\text{g mg}^{-1} \text{ min}^{-1}$)	q_e (mg g^{-1})	R^2	k_2 ($\text{g mg}^{-1} \text{ min}^{-1}$)	q_e (mg g^{-1})	R^2
SA-PA-H	0.67	46.9	0.99	0.15	45.1	0.86
SA-Sr	0.03	37.0	0.91	0.02	28.1	0.69

564

565

566 3.3.3 pH and temperature effect

567 Polluted water from different resources might have various pH values and the temperature of
568 the polluted water can also vary depending on the environment. Therefore, it is crucial to study the
569 stability of the composite hydrogel when used to uptake Sr^{2+} in environments of different pH and
570 temperatures to better guide its practical application in recovering Sr^{2+} from real wastewater.

571 For pH effect, adsorption experiments were carried out at $C_o = 100$ ppm in a pH range from 2
572 to 12 (Figure 9a). The adsorption behavior of SA-PA-H for Sr^{2+} are evaluated by the adsorption
573 capacity and removal efficiency (E). The filtration efficiency is evaluated by removal efficiency (E)
574 calculated from equation (7):

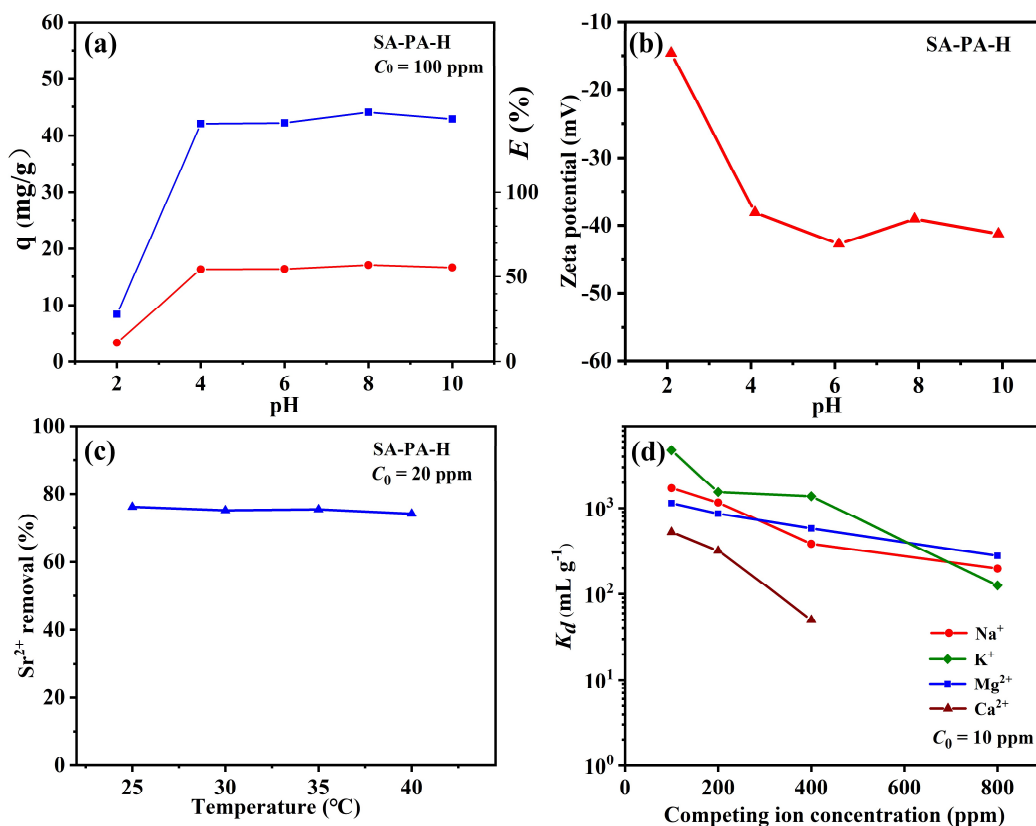
$$575 \quad E(\%) = \frac{C_o - C_e}{C_o} \times 100\% \quad (67)$$

576 where C_o and C_e are initial and equilibrium concentrations of Sr^{2+} measured by ICP-OES.

577 Over a wide range of pH (4-10), the sorption capacity of SA-PA-H towards Sr^{2+} was almost
578 unchanged with a $q \sim 41$ mg/g and a removal efficiency around 54%-57%. The SA-PA-H showed
579 superior stability in solutions with various pH. However, when the solution pH was lowered to 2,
580 the sorption capacity q dropped to 17.7 mg/g, only $\sim 40\%$ of that of the other pH values. The
581 suppressed adsorption here was owing to the high concentration of H^+ competing with Sr^{2+} to bind
582 with carboxyl groups. On the other hand, when suspended in a more basic Sr^{2+} solution with pH of
583 12, the swollen SA-PA-Hs were observed to degrade into water soluble substances with the
584 suspension becoming a clear solution. This is likely due to that the base environment catalyzed the
585 hydrolysis of ester crosslinks in SA-PA-H, resulting in the degradation of the SA-PA-H into water
586 soluble anionic SA and PA. Therefore, the adsorption of SA-PA-H at more alkaline pH cannot be
587 evaluated, while the experiment demonstrated a possible way of degradation of the adsorbents
588 after use.

589 To further elucidate the effect of H^+ on SA-PA-H hydrogel, the Zeta potential of the hydrogel
590 was measured (Figure 7b). When the pH value was greater or equal to 4, the majority of carboxyl
591 groups in SA-PA-H were deprotonated and existed as carboxy anions, which negatively charged
592 the surface of the hydrogel with a low Zeta potential around -40.0 mV. As the pH was decreased to
593 2, the carboxyl groups in the hydrogel were protonated, resulting in an increase of the Zeta
594 potential to -14.6 mV. The different states of the carboxyl groups revealed by the Zeta potential
595 measurement demonstrated the important role of carboxy anion as available sites for Sr^{2+} sorption,

596 which could be modulated by pH to realize Sr^{2+} desorption and regeneration of the SA-PA-H
 597 adsorbent.
 598



599
 600 Figure 9. The effect of pH, temperature and competing ions on SA-PA-H and its adsorption. (a)
 601 Sr^{2+} sorption capacity (q) and Sr^{2+} removal efficiency at $C_0 = 100$ ppm versus pH from 2-10; (b)
 602 pH dependence of the Zeta potential of SA-PA-H versus pH from 2-10; (c) Sr^{2+} removal efficiency
 603 at $C_0 = 20$ ppm versus temperature from 25°C to 40°C ; (d) Distribution coefficient K_d of Sr^{2+} (C_0
 604 = 10 ppm) in brine solutions containing Na^+ , K^+ , Ca^{2+} or Mg^{2+} for SA-PA-H plotted versus
 605 concentration of the competing ions ranging from 200 ppm to 800 ppm. The adsorption of Sr^{2+} in
 606 the presence of 800 ppm Ca^{2+} was undetected.

607

608 Temperature effect on the adsorption performance of SA-PA-H hydrogel was further studied
 609 by testing Sr^{2+} adsorption behaviors of the SA-PA-H at temperatures from 25°C to 40°C (Figure
 610 9c) in Sr^{2+} solutions with an initial concentration $C_0 = 20$ ppm. As shown, for the measured
 611 temperatures, the removal rate of Sr^{2+} by SA-PA-H remained consistent, more than 75%, almost
 612 independent of the temperature. That is, the SA-PA-H hydrogel exhibited good temperature

613 stability within the range of 25-40 °C. Based on adsorption data at different temperatures, the
614 thermodynamics parameters like Gibbs free energy (ΔG^0), enthalpy (ΔH^0) and entropy (ΔS^0) of
615 the adsorption can be calculated from the following equations (8-10)[53]:

$$616 \quad K_a = \frac{q_e}{C_e} \quad (8)$$

$$617 \quad \ln K_a = \frac{\Delta S^0}{R} - \frac{\Delta H^0}{RT} \quad (9)$$

$$618 \quad \Delta G^0 = \Delta H^0 - T\Delta S^0 \quad (10)$$

619 where K_a is the distribution coefficient, which equals to the ratio between adsorption (q) and
620 equilibrium concentration of the solutes in the solution. R is the gas constant. By plotting $\ln K_a$
621 against $1/T$, the temperature-dependent adsorption data can be fitted by the *Van't Hoff* equation
622 (Figure S5). The ΔH^0 and ΔS^0 are calculated to be -27.1 kJ/mol and -83.2 J/(mol·K), respectively.

623 The thermodynamics parameters obtained are listed in Table S2.

624 As seen in the results (Table S2), the negative ΔH^0 indicates that the adsorption of Sr^{2+} by
625 SA-PA-H is exothermic, supported by the Sr^{2+} adsorption decline with increased temperature. On
626 the other hand, the negative ΔS^0 shows an entropy reduction during the adsorption of Sr^{2+} by SA-
627 PA-H, which can be a result of the configuration loss of carboxylic groups and thiol groups after
628 binding with divalent ion Sr^{2+} . The negative ΔG^0 confirms that the adsorption occurs
629 spontaneously under the tested conditions.

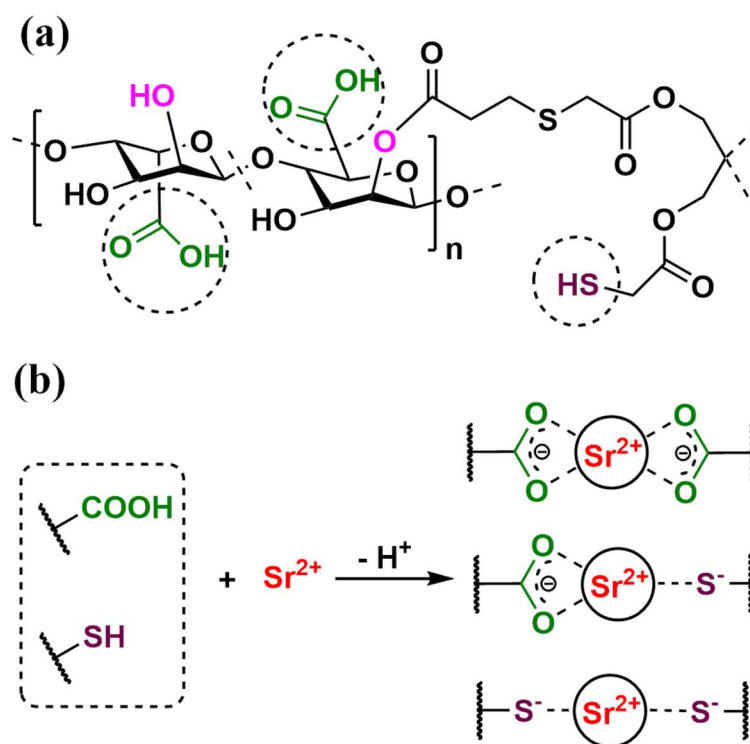
630 In summary, the adsorption of Sr^{2+} on SA-PA-H adsorbent is a spontaneous and endothermic
631 process, even being efficient in complicated environments with various temperatures and pH.

632

633 3.3.4 Adsorption Mechanism

634 In sum of the isothermal adsorption, adsorption kinetics and thermodynamics studies, an
635 adsorption mechanism is proposed. In the suspension of SA-PA-H, the SA-PA-H hydrogel is
636 swollen by the water molecules with its functional groups hydrated in the medium (Figure 10a).
637 The Sr^{2+} dissolved in the solution exchanges with protons from adsorption-active groups, *i.e.*,
638 carboxylic and thiol groups within the ion-imprinted hydrogel. The generated Lewis bases,
639 carboxylate and thiolate groups, serve as anionic ligands to fix Sr^{2+} in the ion-imprinted cavities
640 whereby a single Sr^{2+} is chelated with two adsorbate groups, which could be two carboxylate

641 ligands, two thiolate ligands or a carboxylate ligand along with a thiolate ligand (Figure 10b). The
 642 binding of Sr^{2+} by the thiols to form thiolate species can be demonstrated by the XPS spectrum of
 643 SA-PA-H after adsorption (Figure 6b), where the -SH signals significantly diminished while an
 644 enhanced S^- signal is observed compared to those in the SA-PA-H spectrum. Since the hydrogel is
 645 demonstrated to be homogeneous with carboxylic and thiol groups distributed evenly within the
 646 hydrogel by the EDX maps (Figure 5) and the hydrogel exhibits multilayer adsorption behavior
 647 according to the fitting by the Freundlich model, it is arguably that the Sr^{2+} can be adsorbed by all
 648 the active adsorption groups throughout the fully swollen porous hydrogel, where the diffusion of
 649 Sr^{2+} is less hindered.



650
 651 Figure 10. (a) The chemical structure of SA-PA-H (PA-crosslinked alginic acid) with adsorption-
 652 active carboxylic and thiol groups circled. (b) Ion exchange reaction between Sr^{2+} and protons in
 653 carboxylic and thiol groups. A single Sr^{2+} is bound by two carboxylate ligands, two thiolate
 654 ligands or a carboxylate ligand and a thiolate ligand.

655

656 3.3. 5 Competing ion effect

657 Sodium (Na^+), potassium (K^+), calcium (Ca^{2+}) and magnesium (Mg^{2+}) are the most common
 658 competing ions for Sr^{2+} when adsorbents were used in real polluted water. To figure out the

659 adsorption selectivity of SA-PA-H towards Sr^{2+} , the hydrogel was dispersed in Sr^{2+} solutions ($C_0 =$
660 10 ppm) containing Na^+ , K^+ , Ca^{2+} or Mg^{2+} with varied ion concentrations from 100 to 800 ppm.
661 Sr^{2+} selectivity was evaluated by the distribution coefficient, K_d :

$$662 \quad K_d = \frac{C_0 - C_e}{C_e} \frac{V}{m} \quad (11)$$

663 where C_0 and C_e are the initial and equilibrium concentrations of Sr^{2+} , V is the volume (mL) of the
664 solution, and m is the mass (g) of the absorbent.

665 The Sr^{2+} selectivity K_d was plotted against competing ion concentrations (Figure 9d). When
666 the initial Na^+ concentration was 100 ppm, K_d of SA-PA-H to Sr^{2+} was as high as $1.72 \times 10^3 \text{ mL g}^{-1}$.
667 Although the K_d of SA-PA-H gradually decreased as the concentration of Na^+ was increased, SA-
668 PA-H still showed a good adsorption selectivity towards Sr^{2+} with a K_d of $1.97 \times 10^2 \text{ mL g}^{-1}$ when
669 the Na^+ concentration reached 800 ppm (Figure 9d). The good selectivity to Sr^{2+} was also
670 observed in the presence of K^+ with K_d decreasing from $4.27 \times 10^3 \text{ mL/g}$ to $1.25 \times 10^2 \text{ mL/g}$ as the
671 concentration of K^+ increased from 100 to 800 ppm. Similarly, when the initial Mg^{2+} concentration
672 was 100 ppm, the K_d towards Sr^{2+} was $1.14 \times 10^3 \text{ mL g}^{-1}$, which decreased to $2.80 \times 10^2 \text{ mL g}^{-1}$ as
673 the Mg^{2+} concentration reached 800 ppm (Figure 9d). On the other hand, SA-PA-H selectively
674 adsorbs Sr^{2+} at low concentrations of Ca^{2+} (100-400 ppm) with K_d s from $5.27 \times 10^2 \text{ mL/g}$ to 50
675 mL/g . However, when the Ca^{2+} is as high as 800 ppm, the adsorption of Sr^{2+} by SA-PA-H is minor.
676 Since there is a few Sr^{2+} remaining in SA-PA-H after acid wash, the remained Sr^{2+} may exchange
677 with Ca^{2+} who is in overwhelming excess in the competing adsorption test. Thus, the adsorption of
678 Sr^{2+} in the presence of high concentrations (e.g. 800 ppm) of Ca^{2+} was suppressed, which may be
679 resolved by washing the SA-PA-H hydrogel with acids for more times. All in all, the experiments
680 demonstrated the decent selectivity of SA-PA-H to Sr^{2+} over most competing ions (Na^+ , K^+ and
681 Mg^{2+}) except Ca^{2+} at high concentrations. The selectivity was attributed to the thiol group and the
682 ion-imprinting involved in the hydrogel. According to the hard and soft acids and bases (HSAB)
683 principle[45], both strontium and sulfur are ‘soft’ atoms with thick electronic cloud be easy to
684 deform because of their large atomic numbers. Thus, the thiol group was prone to bind with Sr^{2+}
685 instead of Na^+ , K^+ and Mg^{2+} which are regarded as ‘hard’ cations. Meanwhile, ion-imprinting
686 created cavities with specific shapes and ligand distribution adapted to the geometry and charge
687 distribution of Sr^{2+} , acting as ion sieves for Sr^{2+} . Since Sr, Ca and Mg are alkaline-earth metal

688 elements with the same bivalent charge when presented as cations, the volume of ion-imprinted
689 cavities was expected to play an essential role in selectively recovering Sr^{2+} from concentrated
690 Mg^{2+} solutions or diluted Ca^{2+} solutions in light of the size exclusion effect of the cavities. The
691 two factors contributed together to the good selectivity of SA-PA-H towards Sr^{2+} .

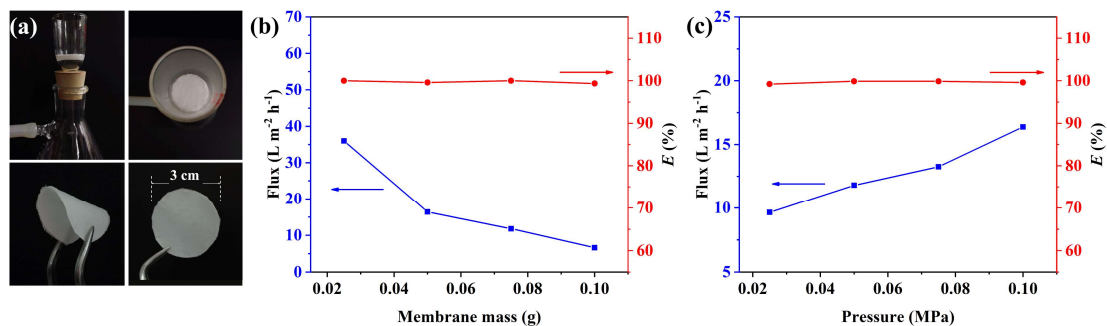
692

693 3.3.6 Membrane filtration

694 Membrane filtration is an advanced technique for desalination and water purification with several
695 advantages such as low cost, facile operation and ready portability.[54-56] The rheological study
696 in section 3.2 has demonstrated that the SA-PA-H hydrogel was recoverable after large
697 deformation, which enables SA-PA-H to accommodate certain shapes for device filtrations. A
698 simple vacuum membrane filtration device was established by depositing swollen SA-PA-H
699 hydrogels onto the sand core in a suction funnel equipped with a suction bottle (Figure 10a). The
700 adsorption efficiency of the SA-PA-H membrane was investigated under vacuum filtration
701 conditions, where 100 mL Sr^{2+} solution with a C_0 of 1 ppm passed through the membrane driven
702 by vacuum.

703 After the filtration, a trace amount of Sr^{2+} was detected in the effluent in all the conditions
704 (Figure S6), and the ion was removed completely after filtration by SA-PA-H membrane in all the
705 conditions with extremely high removal efficiency ($> 99.2\%$) (Figure 10b&c), which revealed the
706 excellent adsorption efficiency of SA-PA-H in fast water purification. Varying the mass of SA-PA-
707 H fabricating the membrane did not affect the removal efficiency of the membrane, and a
708 complete removal ($> 99.99\%$) of Sr^{2+} was achieved with only 0.025 g SA-PA-H (Figure 10b). An
709 increased mass of SA-PA-H to 0.1 g decreased the water flux from $36.0 \text{ L m}^{-2} \text{ h}^{-1}$ to $6.6 \text{ L m}^{-2} \text{ h}^{-1}$
710 because of augmented resistance due to the thicker membrane. When the trans-membrane pressure
711 was raised from 0.025 MPa to 0.1 MPa (Figure 10c), the flux increased accordingly from $9.6 \text{ L m}^{-2} \text{ h}^{-1}$
712 to $16.4 \text{ L m}^{-2} \text{ h}^{-1}$ while the removal efficiency remained unimpaired ($> 99.2\%$). The flux can
713 be readily regulated to produce pure water at desired rates without risks of unexpected percolation
714 of solutes. The excellent membrane filtration performance is undoubtedly ascribed to the large
715 adsorption capacity and the fast adsorption equilibrium of the SA-PA-H. Moreover, after the
716 filtration the SA-PA-H membrane can be recovered from the suction funnel entirely and after

717 drying in the air it maintained as a free-standing membrane that can be bent elastically (Figure
 718 10a). The good mechanical property in the dry state grants SA-PA-H great potential in fabricating
 719 portable and disposable membranes for transportation and large-scale application.



720
 721 Figure 11. (a) Device consisted of a filtration funnel for vacuum membrane filtration and external
 722 appearance of a dried SA-PA-H membrane fabricated in its plat and bent forms. (b) Variation of
 723 solution flux (square, left axis) and removal efficiency (circle, right axis) as a function of
 724 membrane mass ranging from 0.02 to 0.1 g in membrane filtration experiments. (c) Variation of
 725 solution flux (square, left axis) and removal efficiency (circle, right axis) as a function of trans-
 726 membrane pressure ranging from 0.02 to 0.1 MPa in membrane filtration experiments.

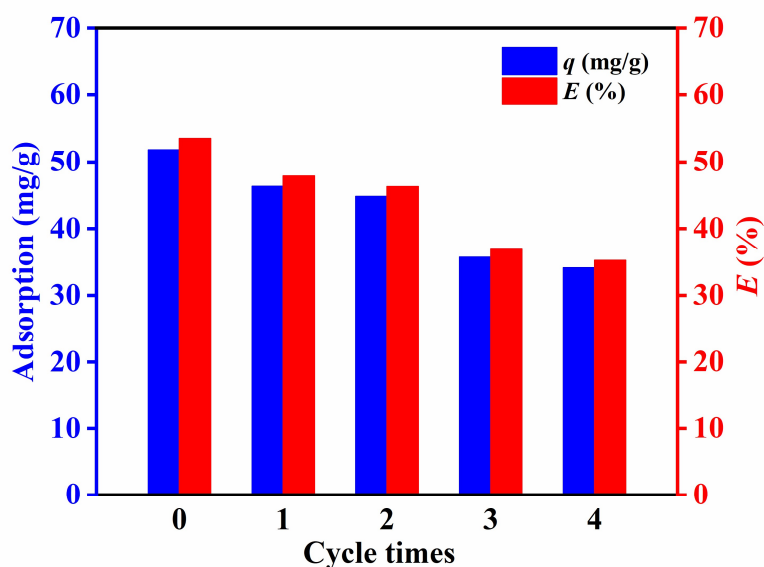
727

728 3.3.7 Recyclability

729 Adsorption and desorption of Sr²⁺ are essential to the service life of adsorbents. An
 730 economical and environment friendly adsorbent is expected to have good recyclability and
 731 stability. To assess the recyclability and stability of the SA-PA-H adsorbent, Sr²⁺ was removed
 732 from the saturated adsorbent by hydrochloric acid to recover sorption sites after one batch
 733 adsorption experiment. Then the regenerated adsorbent was subject to a new adsorption
 734 experiment with the same initial Sr²⁺ concentration ($C_0 = 100$ ppm). In every adsorption cycle,
 735 adsorption efficiency (E) was calculated by equation (7).

736 As Figure 11 shows, the E of freshly-prepared SA-PA-H was 53.51% and the adsorption
 737 capacity (q) was 51.8 mg/g. In the first, second, third and fourth desorption-adsorption cycle, E of
 738 regenerated adsorbents dropped slightly to 47.83%, 46.28%, 36.88% and 35.22%, respectively.
 739 The attenuation of adsorption efficiency was only about 10%-20% in every regeneration taking the
 740 hydrogel in the last cycle as the reference, indicating decent stability of the hydrogel. Since there
 741 was an unavoidable loss of SA-PA-H hydrogel content during the recycling involving

742 centrifugation, the actual recycle efficiency was expected to be higher than the experimental
743 values. The loss of adsorption capacity may also be attributed to the deterioration of adsorption
744 sites resulting from the acid-catalyzed esterification between hydroxyl and carboxylic groups
745 during acid wash in the regeneration step. The results showed that through washing with acids SA-
746 PA-H adsorbent could be regenerated conveniently. In the meantime, protonation of the hydrogels
747 released the adsorbed Sr^{2+} into the acid solutions, which enables an efficient separation of Sr^{2+}
748 from the contaminated water. The robustness allowed SA-PA-H to serve as a recyclable and
749 environment friendly adsorbent to recover Sr^{2+} from polluted water with little waste and reagent
750 contamination.



751
752 Figure 12. Sr^{2+} adsorption capacity and adsorption efficiency (E) of SA-PA-H hydrogel in batch
753 adsorption experiments with an initial Sr^{2+} concentration of 100 ppm as a function of regeneration
754 time. The SA-PA-H used was regenerated by washing the adsorbents with 100 mL 0.5 mol/L
755 hydrochloric acid.

756

757 3.3.8 Comparison of adsorption performance

758 Table 3 compares the as-developed SA-based adsorbent with other reported adsorbents. As
759 shown, the SA-PA-H hydrogel is among the most effective Sr^{2+} adsorbents in terms of maximum
760 adsorption capacity. Moreover, the SA-PA-H adsorbent can reach adsorption equilibrium by a fast
761 uptake kinetics and it is also sufficiently flexible to be shaped or loaded to fit different devices.
762 These features enabled its application in continuous water purification processes such as column

763 filtration and membrane filtration, where the contact time between contaminated water and the
764 adsorbent is very limited that the adsorbents with slow uptake kinetics fail to achieve good
765 purification. Although there are adsorbents having significantly high adsorption capacity and short
766 equilibrium time, they are mostly in form of powder, difficult to be applied in real water treatment
767 because of their low availability to be fitted into common purification devices and the difficulties
768 of post-treatment after adsorption (e.g. easily causing secondary pollution). Taking advantage of
769 excellent adsorption efficiency, good selectivity, good pH stability and outstanding modality, SA-
770 PA-H is a promising adsorbent candidate for real water treatment.
771

772 Table 3 Comparison of the adsorbent developed in this work with other Sr²⁺ adsorbents reported.

Adsorbent	State	q _{max} (mg/g)	Equilibrium (min)	pH range	Reference
Na Titanate	powder	49.6	60	3-10	[24]
polyantimonic acid- polyacrylonitrile (PAA-PAN)	powder	49	500	n. a.	[23]
Graphene oxide (GO)	powder	131.4	20	3-11	[22]
Covalent triazine polymer- Fe ₃ O ₄ (CTP-Fe ₃ O ₄)	powder	128	200	7-14	[31]
layered vanadosilicate (SGU- 7)	powder	109	10	3-13	[57]
Na ₂ Sn ₃ S ₇ (NaTS)	powder	80.0	1	3-13	[46]
K _{1.87} ZnSn _{1.68} S _{5.30} (KZTS)	powder	19.3	1	3-11	[47]
Na ₅ Zn _{3.5} Sn _{3.5} S ₁₃ ·6H ₂ O (NaZTS)	powder	40.40	5	3-12	[49]
Impregnated nano-zeolite (NAASMS-Z)	powder	350	30	7-10	[26]
K ₂ Sb ₄ S ₇ ·2H ₂ O (SbS-1K)	powder	61.1	40	4-11	[58]
Thioglycolic acid-modified alginate sodium (SA-TGA)	microsphere	177.37	180	4-8	[44]
metakaolin/slag-based zeolite (M/SZMs)	microsphere	54.90	30	4-8	[59]
Chitosan	Fiber and microsphere	20.0	200	3-7	[60]
Graphene oxide/Nickel-metal- organic framework composite (GO/Ni-MOF)	membrane	72	>500	n.a.	[61]
Bacterial cellulose membrane modified with ethylenediaminetetraacetic acid (BCM@APTES-EDTA)	membrane	44.86	600	n.a.	[30]
SA-PA-H	moldable hydrogel	151.7	30	4-10	This work

773

774 4 Conclusion

775 A Sr²⁺-imprinted and thiol-functionalized hydrogel based on SA (*i.e.* SA-PA-H) was successfully
776 synthesized and used to selectively adsorb Sr²⁺ from contaminated water. The synthesis of
777 hydrogel involved three steps including the formation of SA hydrogel in emulsion via
778 complexation with Sr²⁺, crosslinking of the alginate gels via esterification with thiol-
779 functionalized tetra-arm carboxylic acids (PA), and lastly activation by acid elution to form Sr²⁺-
780 imprinted cavities. The synthesis route was verified by a series of techniques including NMR,
781 FTIR, XPS and SEM, etc.

782 The incorporation of abundant thiol groups and the imprinted cavities with predetermined sizes
783 endow the prepared SA-PA-H hydrogel with an excellent adsorption ability and decent selectivity
784 towards Sr²⁺ within a wide range of pH (4-10) and temperature (25-40 °C). The adsorption process
785 followed second-order kinetics at an adsorption rate constant of 0.669 g mg⁻¹ min⁻¹. The
786 isothermal adsorption behavior can be well described by the dual-site Langmuir model with a
787 maximum adsorption capacity of 151.7 mg g⁻¹. The Sr²⁺ binding to thiol and carboxyl groups has
788 been confirmed to be the two main adsorption mechanisms. The selectivity was validated by that
789 the existence of competing ions such as Mg²⁺, K⁺ and Na⁺ hardly weakened the adsorption
790 efficiency of SA-PA-H to Sr²⁺, having a K_d greater than 1.97×10² mL/g obtained even at high
791 competing ion concentrations.

792 Rheological investigation reveals that SA-PA-H hydrogel can self-recover after being damaged by
793 large strains. The hydrogel can be easily fabricated into membranes for Sr²⁺ separation with high
794 removal efficiency (> 99.2%) within a wide range of liquid flux. The SA-PA-H adsorbent can be
795 readily regenerated by washing with HCl solutions and the regenerated adsorbents possess
796 comparable adsorption capability as the pristine hydrogel. The research developed a facile way to
797 fabricate alginate-based hydrogel adsorbent for selective Sr²⁺ separation. Thanks to its outstanding
798 adsorption capability and good mechanical properties, SA-PA-H is promising in strontium
799 recovery for contaminated water treatment.

800

801

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808

809 **Appendix A. Supplementary data**

810 Supplementary data to this article can be found online at: XXXX

811

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815 **Author Contributions**

816 #B. Zheng and J. Yin contributed equally to this work.

817 **Notes**

818 The authors declare no competing interest.

819

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