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## Abstract

21

22 Birnessite minerals help control the fate of Zn in surface environments and readily  
23 fractionate Zn isotopes through adsorption reactions, yet little is known about the role  
24 played by various reactive sites in stable isotopic fractionation. Here we present the Zn  
25 isotope fractionation data caused by adsorption on birnessite under different reaction times,  
26 pH values, and Zn concentrations. We observe that isotopic equilibrium of Zn is attained  
27 after ~120 h of reaction time at pH 6. At pH 3–5 and Zn concentrations of 0.05–0.3 mM,  
28 the isotopic fractionation ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$ ) is around  $-0.46 \pm 0.04\text{‰}$ , and gradually  
29 increases to  $-0.09 \pm 0.05\text{‰}$  at pH 6–8 and Zn concentrations of 0.2 mM. The change in Zn  
30 isotopic compositions as a function of pH and Zn concentration is well described using the  
31 surface complexation model, where two binding sites are involved: external edge sites and  
32 interlayer vacancies. According to this model, two different isotopic fractionation factors  
33 of Zn are calculated:  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = -0.46 \pm 0.04\text{‰}$  for adsorption on vacancy sites  
34 and  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = 0.52 \pm 0.04\text{‰}$  for binding to edge sites. Extended X-ray absorption  
35 fine structure spectroscopy (EXAFS) demonstrates that Zn forms triple-corner-sharing  
36 (TCS) octahedral complex on birnessite vacancies at pH 3 and Zn concentrations of  
37 0.05–0.2 mM, where Zn is coordinated on one side to three oxygen atoms of the Mn  
38 vacancy ( $\sim 2.03 \text{ \AA}$ ) and to three water molecules on the other side ( $\sim 2.15 \text{ \AA}$ ), suggesting the  
39 formation of distorted Zn–O octahedra (average bond length:  $\sim 2.09 \text{ \AA}$ ). At pH 6 and 8,  
40 double-corner-sharing (DCS) complexes on layer edges formed in addition to the TCS

41 octahedral complex on vacancies. Density functional theory (DFT) optimisations suggest  
42 that DCS Zn complex exist in tetrahedral coordination. Based on EXAFS spectroscopy,  
43 DFT optimisations and surface complexation modeling, the distinct isotopic fractionation  
44 of Zn is related to the differences in Zn local structure at different reactive sites of birnessite.  
45 Our results provide a molecular-scale understanding of Zn isotopic fractionation in natural  
46 birnessite-containing settings, as well as new insights into predicting the links between  
47 adsorption and fractionation of other similar metals.

48

49 *Keywords: Birnessite; adsorption; isotope fractionation; SCM; EXAFS; DFT*

## 50 **1. Introduction**

51 Birnessite is a nanosized variety of poorly crystalline phyllosulfate that occurs  
52 widely in freshwater (Manceau et al., 2007), soils (Hochella et al., 2005), and marine  
53 sediments (Peacock and Sherman, 2007a). The crystal structure of birnessite is composed  
54 of layered edge-sharing  $\text{MnO}_6$  octahedra, and the distance between layers is approximately  
55 7.2 Å (Manceau et al., 2002). Due to the presence of Mn vacancies and lower valence  
56 cations substitution for  $\text{Mn}^{4+}$  in the  $\text{MnO}_6$  layers (Lanson et al., 2002; Toner et al., 2006;  
57 Peacock, 2009; Kwon et al., 2009), the surface of birnessite is negatively charged, which  
58 leads to the adsorption of multivalent cations. For example,  $\text{Zn}^{2+}$ ,  $\text{Ni}^{2+}$ , or  $\text{Cu}^{2+}$  adsorbs as  
59 triple-corner-sharing (TCS) surface complex (Fig. 1a-b) (Manceau et al., 2002; Peacock,  
60 2009; Kwon et al., 2009), or  $\text{Cu}^{2+}$  and  $\text{Ni}^{2+}$  incorporate into Mn vacancy sites (INCs) as  
61 part of the birnessite layer (Fig. 1c) (Peacock and Sherman, 2007b). In addition, at  
62 birnessite particle edges, the undercoordinated oxygen atoms can adsorb cations by  
63 forming double-corner-sharing (DCS) or double-edge-sharing (DES) complexes (e.g.,  $\text{Ni}^{2+}$ ,  
64  $\text{Co}^{2+}$ , or  $\text{Pb}^{2+}$ ) (Fig. 1d-f) (Villalobos et al., 2005; Bargar et al., 2009; Kwon et al., 2010;  
65 Simanova et al., 2015; Wang et al., 2018).

66 As a result of its common occurrence and high surface reactivity, birnessite plays a  
67 fundamental role in trace metal cycling, exerting a strong control on the fate and mobility  
68 of trace metals through sorption reactions (Post, 1999; Sherman and Peacock, 2010). In the  
69 ocean, for example, trace elements (e.g., Zn, Cu, and Ni) are enriched in poorly crystalline

70 birnessite in ferromanganese crusts by  $10^6$  times over their concentrations in seawater (e.g.,  
71 [Koschinsky and Hein, 2003](#)). Birnessite has also been reported to influence the cycling of  
72 trace metals in terrestrial environments ([Manceau et al., 2007](#)) and in areas influenced by  
73 anthropogenic activities from mining and smelting activities ([Juillot et al., 2011](#); [Spinks  
74 and Uvarova, 2019](#)).

75 The environmental biogeochemical cycling of metals can be investigated by applying  
76 of metal stable isotope systems, since the metal isotopic signatures can be used to identify  
77 the specific sources and geochemical processes (e.g., [Chen et al., 2008](#); [Wiederhold, 2015](#)).  
78 Metal stable isotopes, for example, have been shown to have the potential to distinguish  
79 between anthropogenic and natural sources of metals and thus enable fingerprinting of  
80 environmental pollution ([Aranda et al., 2012](#); [Bi et al., 2017](#)). The isotopic signatures from  
81 different sources can be obscured, however, by isotopic fractionation during mineral-  
82 solution interface reactions (e.g., [Gueguen et al., 2018](#); [Wang et al., 2022](#)), which may  
83 hamper the application of metal isotopes as effective tracers.

84 Similarly, metal stable isotopes have also been used to quantify the input and output  
85 of metal marine budgets ([Gall et al., 2013](#); [Little et al., 2014b](#)). Marine sediment adsorption  
86 controls the output of trace metals from the ocean; however, this process can drive  
87 significant isotopic fractionation but is poorly constrained ([Vance et al., 2016](#); [Liu et al.,  
88 2019](#); [Sorensen et al., 2020](#)). In general, investigating the links between isotopic  
89 fractionation and adsorption can improve the understanding of fractionation and inform the

90 most accurate use of metal isotopes in environmental investigations.

91 Zinc (Zn) and its isotopes are ideal candidates to examine the links between adsorption  
92 and fractionation because we already know some information about Zn adsorption and  
93 fractionation with Mn oxides (e.g., [Pokrovsky et al., 2005](#); [Bryan et al., 2015](#)). Previous  
94 work on the isotopic fractionation and the coordination environment of Zn during  
95 adsorption to different mineral surfaces can inform experimental design and interpretation  
96 (e.g., [Pokrovsky et al., 2005](#)). [Pokrovsky et al. \(2005\)](#) firstly reported the isotopic  
97 fractionation values of Zn caused by adsorption on various metal oxides, suggesting that  
98 adsorption of Zn onto surfaces of hematite, pyrolusite, corundum and gibbsite enriches  
99 heavier Zn isotopes, while birnessite and goethite prefer to adsorb lighter Zn isotopes. Later,  
100 [Bryan et al. \(2015\)](#) found that Zn isotopes show complicated behaviour during adsorption  
101 to birnessite, which may be affected by both Zn surface coverage and solution ionic  
102 strength. At low ionic strength, insignificant Zn isotope fractionation ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} =$   
103  $0.05 \pm 0.08\text{‰}$ ) is observed. In high ionic strength solution, heavy Zn are enriched on the  
104 birnessite surface, but the isotopic fractionation is strongly dependent on surface coverage,  
105 with  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = 0.16\text{‰}$  at high surface coverage, which is increased to 2.74‰ at  
106 low surface coverage. Additionally Zn is a micronutrient required by organisms and is also  
107 a potential contaminant in many scenarios (e.g., [Moore et al., 2013](#); [Sandstead, 2014](#)). For  
108 example, Zn plays a key role as a cofactor in carbonic anhydrase involved in carbon  
109 fixation, so their concentrations in seawater are critical for regulating climate (Morel et al.,

110 1994). In the ocean the concentration and isotopic compositions of Zn is controlled by  
111 adsorption to ferromanganese (oxyhydr)oxides in ferromanganese crusts and nodules and  
112 in ferromanganese particulates in oxic sediments (Little et al., 2014)

113 Knowledge of Zn surface coordination chemistry is key to understand the complex  
114 behaviour of Zn isotopes during interfacial adsorption reactions. Classic stable isotope  
115 exchange theory (Bigeleisen and Mayer, 1947) suggests that equilibrium isotopic  
116 fractionation is controlled by the bond strength between the metal of interest and its  
117 neighboring atoms (Schauble, 2004). As an advanced technique based on synchrotron  
118 radiation accelerator, EXAFS spectroscopy can determine the bond distance and  
119 coordination number of metals adsorbed on mineral surfaces. This molecular-level  
120 information can be related to isotopic behaviour to elucidate the relation between interfacial  
121 reactions and isotopic fractionation (e.g., Juillot et al., 2008). At present, Zn isotopic  
122 fractionation mechanisms have been examined for several different mineral systems by  
123 EXAFS spectroscopy (e.g., Nelson et al., 2017; Wang et al., 2022). In the Zn- $\gamma$ -Al<sub>2</sub>O<sub>3</sub>  
124 system, adsorption of heavy Zn ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = 0.47 \pm 0.03\%$ ) is explained by the  
125 presence of tetrahedral Zn with a typically shorter Zn-O bond length (1.96 Å) relative to  
126 aqueous Zn (2.06 Å) (Gou et al., 2018). Similarly, in iron oxide systems, the enrichment of  
127 heavier Zn isotopes on solid phases is related to a shorter Zn-O bond length of adsorbed  
128 Zn relative to the aqueous phase (Juillot et al., 2008). Recently, Wang et al. (2022) observed  
129 insignificant isotopic fractionation owing to the formation of outer-sphere octahedral Zn

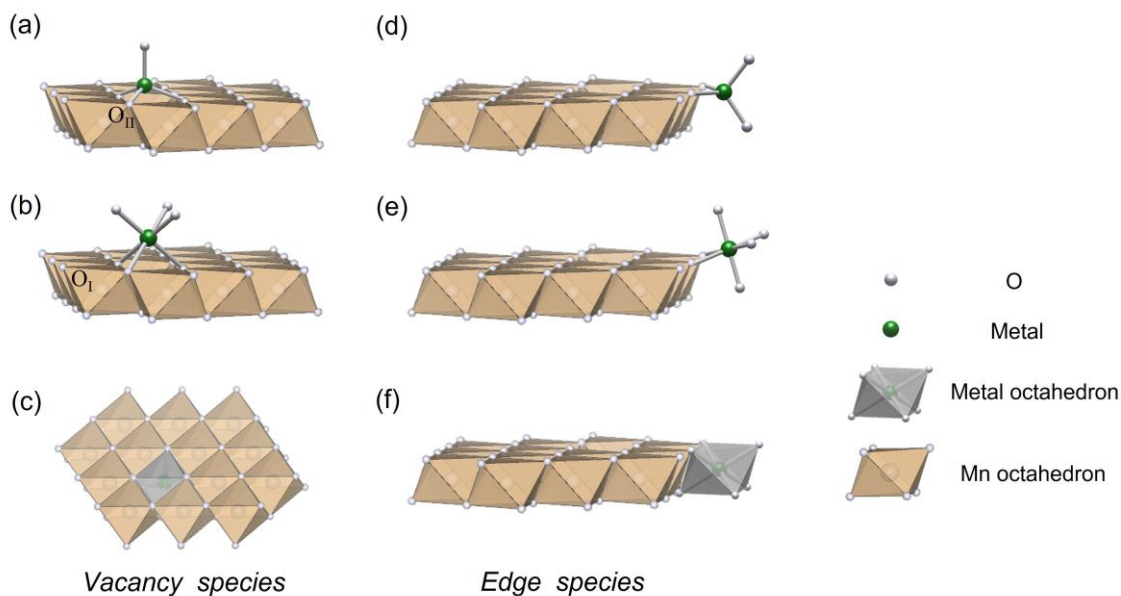


130 complexes on todorokite, in which there is no obvious change in the coordination  
131 environment relative to aqueous Zn. Despite different coordination numbers, the formation  
132 of octahedral and tetrahedral Zn complexes results in undistinguishable isotopic  
133 fractionation factors in silica systems (Nelson et al., 2017). This lack of isotopic  
134 fractionation is proposed to be caused by the small energy difference between octahedral  
135 and tetrahedral Zn species on amorphous silica and quartz surfaces (Nelson et al., 2017).  
136 For Zn adsorption on birnessite, Zn isotopes are shown to have complicated behaviour  
137 influenced by surface coverage and solution ionic strength, and different studies produce  
138 disparate results (Pokrovsky et al., 2005; Bryan et al., 2015). To date, EXAFS has not been  
139 applied to explain this complicated isotopic behaviour.

140 In this study, we investigate the adsorption of Zn onto birnessite and examine the  
141 associated isotopic fractionation behaviour under different contact times, pH values, and  
142 Zn concentrations. We follow a synergetic EXAFS spectroscopy, DFT geometry  
143 optimisation, and surface complexation modeling (SCM) approach to elucidate the  
144 fractionation mechanisms of Zn isotopes. The molecular-level structural information  
145 revealed by EXAFS spectroscopy and DFT optimisation sheds light on the nature of surface  
146 complex of Zn on birnessite and allows interpretation of the equilibrium isotopic  
147 fractionation. Additionally, we use the adsorption complexes provided by EXAFS and DFT  
148 to constrain the SCM modeling. SCM allows us to quantitatively constrain the site-specific  
149 fractionation factors of Zn isotopes caused by adsorption on birnessite. These results offer

150 a mechanistic understanding of the site-specific fractionation behaviour of Zn and  
151 potentially new insights into the links between adsorption and fractionation of other similar  
152 metals.

153



154

155 **Fig. 1** Schematic representations of possible surface complexes of metals on birnessite: (a)  
156 metal-TCS complex in tetrahedral coordination and (b) metal-TCS complex in octahedral  
157 coordination associated with vacancy sites; (c) incorporation of a metal cation into a  
158 birnessite layer vacancy; (d) metal-DCS complex in tetrahedral coordination and (e) metal-  
159 DCS complex in octahedral coordination associated with edge sites; (f) metal-DES  
160 complex in octahedral coordination associated with edge sites. O<sub>I</sub> and O<sub>II</sub> depict singly  
161 coordinated and doubly coordinated oxygen atoms, respectively. Yellow octahedra: Mn;  
162 white spheres: O; gray octahedra and green spheres: metals.

163

## 164 2. Materials and Methods

### 165 2.1. Synthesis of birnessite

166 Acid birnessite was synthesised following the procedure described by McKenzie  
167 (1971). First, 500 mL of 0.4 mol/L KMnO<sub>4</sub> solution was heated to boil in an oil bath. Then,

168 a 35 mL mixed solution of 12 mol/L HCl and 15 mL doubly deionized water (18.2 MΩ,  
169 Milli-Q, Millipore) was added slowly to the KMnO<sub>4</sub> solution at a speed of 0.7 L/min. After  
170 30 min, the crystal suspension was cooled at ambient temperature and then aged for ~12 h  
171 at ~60 °C (Feng et al., 2007). The resulting precipitates were separated by centrifugation  
172 and rinsed with deionized water until Cl<sup>-</sup> cannot be detected by 0.1 mol/L acidic AgNO<sub>3</sub>  
173 solution and then freeze-dried. The background Zn concentration in the synthetic birnessite  
174 was determined to be very low (i.e., 4.2 ppm). This low Zn concentration in the birnessite  
175 has a negligible impact in our adsorption experiments and isotopic measurements.

176

## 177 **2.2. Characterisation of birnessite**

178 A Bruker D8 diffractometer (Mo Kα radiation) was employed to collect the powder  
179 X-ray diffraction (XRD) patterns of the birnessite with an operation voltage of 40 kV/30  
180 mA. Samples were analyzed over a 5°-80° 2θ angular range with a step size of 0.02° and 1  
181 s counting time per step.

182 Transmission electron microscopy (TEM) images of the synthetic birnessite were  
183 collected on an FEI Tecnai (F20) microscope at an accelerating voltage of ~200 kV. For  
184 imaging, finely ground sample powders were suspended and dispersed in ethanol using a  
185 sonication bath. Holey carbon grids were used to load samples. The specific surface area  
186 of the birnessite was measured by the N<sub>2</sub> Brunauer-Emmet-Teller (BET) method to be 30.7  
187 m<sup>2</sup> g<sup>-1</sup>. Using an oxalic acid/permanganate back-titration method (Feng et al., 2007), the

188 average oxidation state (AOS) of Mn in our synthetic birnessite was measured to be 3.92.

189

### 190 **2.3. Zinc adsorption experiments**

191 Three different types of adsorption experiments were conducted: 1) kinetic adsorption  
192 experiments at pH ~ 6 and Zn concentration of ~0.2 mM for a duration of 0.5 – 432 h; 2)  
193 adsorption edge experiments at Zn concentration of ~0.2 mM in different ionic strength  
194 solutions (I = 0.01 and 0.1 M NaNO<sub>3</sub>) at pH 3–9 for 120 h; and 3) adsorption isotherm  
195 experiments (I = 0.1 M NaNO<sub>3</sub>) at pH ~3 and pH ~7 with Zn concentrations of 0.05–0.3  
196 mM for 120 h.

197 Prior to Zn addition, ~0.065 g L<sup>-1</sup> birnessite suspension was equilibrated at a fixed pH  
198 between 3 and 9 for ~24 h in I = 0.1 M NaNO<sub>3</sub> solutions (except for adsorption edge  
199 experiments in I = 0.1 M and 0.01 M NaNO<sub>3</sub> solutions). After this hydration step, certain  
200 amounts of ~35 mM Zn(NO<sub>3</sub>)<sub>2</sub> stock solution were added to the birnessite suspension,  
201 resulting in the desired Zn concentrations of 0.05–0.3 mM. These initial Zn concentrations  
202 were selected because the concentrations of Zn adsorbed on birnessite ( $0.037 < \text{Zn/Mn}$   
203 molar ratio  $< 0.327$ ) not only covered previously investigated Zn range, but also  
204 complemented the higher Zn part, which is essential to elucidate the link between Zn  
205 adsorption and fractionation on birnessite and further apply to natural systems with  
206 different Zn concentration. The Zn stock solution was added dropwise over a period of  
207 several minutes, while the suspensions were stirred vigorously to ensure no local

208 oversaturation of any Zn solid phases (e.g., Zn hydroxides and hydrozincite) present in the  
209 solution. Thermodynamic calculations from Visual MINTEQ.3.1 suggest that the  
210 predominant Zn species was free  $\text{Zn}(\text{H}_2\text{O})_6^{2+}$  across the pH range 3–8 in the adsorption  
211 experiments (Fig. S1).

212 After the reaction (detailed in Table 1), the suspension was centrifuged to separate the  
213 solid and solution phases. The solution was then filtered using  $\sim 0.22 \mu\text{m}$  polyethersulfone  
214 membrane filters (Sartorius, Germany). The Zn concentration was analysed using an  
215 inductively coupled plasma-optical emission spectrometer (ICP–OES) (iCAP 6000,  
216 Thermo). Selected fresh paste solids were collected for EXAFS analysis. Procedural blank  
217 experiments were included in each set of adsorption experiments (kinetic, adsorption edge,  
218 and isotherm experiments) to monitor the potential Zn contamination. All experiments used  
219 reagent grade chemicals, doubly deionized water, and purified acids (via double sub-  
220 boiling distillation). Teflon labware was used to minimize potential Zn contamination for  
221 both adsorption and isotopic experiments.

222

#### 223 **2.4. Zinc isotope measurements**

224 After batch adsorption experiments, the supernatants obtained were dried and then  
225 dissolved in  $\sim 11 \text{ M}$  HCl prior to Zn isotope analysis. The birnessite samples (after  
226 adsorption and centrifugation) were completely digested in  $\sim 11 \text{ M}$  HCl to dissolve the  
227 birnessite and release adsorbed Zn. Then the solution was evaporated to dryness and

228 redissolved in approximately 2 M HCl, and the solution was loaded in Teflon columns  
229 (height: 10 cm, diameter: 0.4 cm) containing AG MP-1 (100–200 mesh) anion-exchange  
230 resin (Bio–Rad, USA). Zn was purified following the same protocol by [Wang et al. \(2022\)](#).  
231 For each experimental run, the procedural blanks of Zn (< 20 ng) were negligible, because  
232 ~25 µg Zn was used for isotopic determination of each sample. After purification, the  
233 sample was evaporated to dryness and redissolved in ~15 M HNO<sub>3</sub> to eliminate organic  
234 resin residues. The solution was then evaporated to dryness again before the sample was  
235 dissolved in ~0.05 M HNO<sub>3</sub> for Zn isotope analysis.

236 Zn isotopic compositions were measured using a Neptune Plus multicollector  
237 inductively coupled plasma mass spectrometer (MC–ICP–MS) (Thermo Scientific). Both  
238 empirical external normalization (EEN) (i.e., doping with ERM-AE647 Cu) and the sample  
239 standard bracketing (SSB) methods were used for mass bias correction ([Marechal et al.,  
240 1999](#)). Standard and sample solutions were diluted to ~0.2 ppm Cu and ~0.5 ppm Zn for  
241 measurements. The Zn isotope analysis for each sample was performed for three blocks of  
242 60 cycles. The introduction system was washed for 40 s before each isotopic ratio  
243 measurement using a 0.05 M HNO<sub>3</sub> solution to bring the Zn signal to the baseline value.  
244 The take-up time was ~180 s for each measurement. The isobaric interference of <sup>64</sup>Ni on  
245 <sup>64</sup>Zn was evaluated by collecting <sup>62</sup>Ni, but the impact was found to be negligible. The  
246 measurements were run using wet plasma at low mass resolution modes.

247 The measured Zn isotope composition was provided in delta notation for <sup>66</sup>Zn/<sup>64</sup>Zn

248 relative to our in-house standard (HPS) (Wang et al., 2022) as follows:

$$249 \quad \delta^{66}\text{Zn} = \left( \frac{({}^{66}\text{Zn}/{}^{64}\text{Zn})_{\text{sample}}}{({}^{66}\text{Zn}/{}^{64}\text{Zn})_{\text{HPS}}} - 1 \right) \times 1000 \text{ (‰)} \quad (1)$$

250 We used the IRMM 3702 and JMC Lyon international standards to calibrate our in-house  
251 standard which gave  $\delta^{66}\text{Zn}_{\text{HPS}} = -0.36 \pm 0.04\text{‰}$  and  $-0.07 \pm 0.04\text{‰}$ , respectively (Wang et al.,  
252 2022). Over a 4-year period, the MC-ICP-MS achieved a long-term external  
253 reproducibility better than 0.03‰ (2SD) for  $\delta^{66}\text{Zn}$  based on a record of repeated analyses  
254 of geological rock standards and synthetic solutions. Several geological rock standards,  
255 such as Nod-P-1 and BCR, were analysed to assess the accuracy of the analytical procedure,  
256 which is comparable to previously reported values within error (Wang et al., 2022).

257 Zn isotopic fractionation ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$ ) between the adsorbed and aqueous phases  
258 and the mass balance offset (O) were calculated using the following two equations:

$$259 \quad \Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = \delta^{66}\text{Zn}_{\text{adsorbed}} - \delta^{66}\text{Zn}_{\text{aqueous}} \quad (2)$$

$$260 \quad \text{O} = \delta^{66}\text{Zn}_{\text{stock}} - (\delta^{66}\text{Zn}_{\text{adsorbed}} \times f + \delta^{66}\text{Zn}_{\text{aqueous}}(1 - f)) \quad (3)$$

261 where O and f are the mass balance offset and the percent of Zn adsorption, respectively.

262 The mass balance values are considered acceptable, if they are within the 0.06‰  
263 uncertainty (2SD) of the initial Zn stock ( $\delta^{66}\text{Zn}_{\text{stock}} = 0.28 \pm 0.06 \text{‰}$ ).

264

## 265 **2.5. EXAFS data collection and analysis**

266 The Zn K-edge EXAFS data were collected at Beijing Synchrotron Radiation Facility  
267 (BSRF) and Shanghai Synchrotron Radiation Facility (SSRF) on 1W1B beamline and

268 14W1B beamline, respectively. For both beamlines, the beamline monochromator consists  
269 of two parallel Si(111) crystals. Higher-order harmonics were rejected by detuning the two  
270 parallel Si(111) crystals by 30%. Experimental spectra were recorded by scanning energy  
271 from about 9570 eV to 10200 eV. Energy steps were of 10 eV for the pre-edge range, 1 eV  
272 for the absorption edge, 0.05 Å for the post-edge range and counting time for each point  
273 varied from 2 s to 5 s depending on the signal sensitivity. Calibration of the energy was  
274 performed against a Zn foil (~9659 eV). Solid samples were prepared as fresh pastes from  
275 adsorption experiments, and were mounted into plastic holders in both sides sealed using  
276 Kapton tape to prevent from sample dehydration. Samples with high Zn loading (e.g., >2%  
277 by weight) were analysed in transmission mode and otherwise in fluorescence mode. Three  
278 to five scans were collected for each sample to achieve a good signal/noise ratio. In addition  
279 to the adsorption samples, a series of Zn reference samples were also examined including  
280 Zn hydroxide, hydrozincite, chalcophanite, and tetrahedral Zn on birnessite ([Table 2](#)).

281 The IFEFFIT 1.2.11 program was used to process the EXAFS data. The  $\chi(k)$  function  
282 was extracted and weighted by  $k^3$  to enhance the oscillation in higher  $k$  range. In Fourier  
283 transform,  $k$  range of 3-11 Å<sup>-1</sup> and the Hanning window function was used. For all samples,  
284 shell-by-shell fitting was performed in R-space in the range of 1.0-3.6 Å<sup>-1</sup>. The Zn-O, Zn-  
285 Zn and Zn-Mn paths were calculated using a chalcophanite structural model. Fitting of the  
286 spectrum of Zn(NO<sub>3</sub>)<sub>2</sub> solution returned an amplitude reduction factor ( $S_0^2$ ) of ~0.97 ([Wang  
287 et al., 2022](#)) and this value was then applied to other samples. In each sample, the single



288 energy shift  $\Delta E_0$  was set to be equivalent for all shells.

289

## 290 **2.6. DFT geometry optimization**

291 Model Zn surface complexes were constructed based on an edge-sharing Mn  
292 octahedral nanodisk with a diameter of approximately 1.5 nm (i.e.,  $\text{Mn}_{19}\text{O}_{54}\text{H}_{30}$ ). For TCS  
293 species, a Mn(IV) vacancy was created by removing one Mn ion, and two Zn ions were  
294 placed above and below the vacancy site. Each Zn was coordinated with one or three  $\text{H}_2\text{O}$   
295 molecules to represent a tetrahedral or octahedral coordination, respectively. DCS Zn was  
296 initially coordinated to two singly coordinated O atoms at the lateral edge of the vacancy-  
297 free nanodisk, and four or two  $\text{H}_2\text{O}$  molecules were added to the Zn cation to represent an  
298 octahedral or tetrahedral complex. For the DES Zn complex,  $\text{Zn}\cdot 3\text{H}_2\text{O}$  was coordinated  
299 either to one triply coordinated O or one doubly coordinated O of the vacancy-free  
300 nanodisk as well as two singly coordinated O at the lateral edge. Each Zn complex model  
301 placed in a large periodic cell of  $40 \times 40 \times 30 \text{ \AA}$  was geometry-optimized using CASTEP  
302 code with the spin-polarized generalized gradient approximation using Perdew-Burke-  
303 Ernzerhof functional and ultrasoft pseudopotentials ([Vanderbilt, 1990](#); [Perdew et al., 1996](#)).  
304 The expansion of the plane wave basis set was truncated at 500 eV, and the  $k$ -point grid  
305 was  $1 \times 1 \times 1$  for the first Brillouin zone. See previous studies ([Kwon and Sposito, 2015](#);  
306 [Simanova et al., 2015](#)) for the convergence of calculated structural parameters in terms of  
307 the supercell size and the kinetic energy cutoff. Geometry optimization was performed with

308 all atomic positions relaxed using the Broyden, Fletcher, Goldfarb, Shanno (BFGS)  
309 procedure (Pfrommer et al., 1997) until the total energy change was less than  $5 \times 10^{-6}$   
310 eV/atom, and the force and atom displacement along any Cartesian component were less  
311 than 0.01 eV/Å and 0.0005 Å, respectively. The magnetic ordering among Mn ions was set  
312 to be ferromagnetic for simplicity. For some complex models, we used the dispersion  
313 correction method (DFT–D) of Grimme (2006).

314

## 315 2.7. Surface complexation modeling

316 Layered birnessite has both internal surfaces and external surfaces (e.g., Peacock and  
317 Sherman, 2007b). For the internal surfaces of birnessite, doubly coordinated oxygens ( $O_{II}$ )  
318 were associated with vacancy sites (Fig. 1a). For birnessite external surfaces, there were  
319 both singly coordinated and doubly coordinated oxygen atoms (Fig. 1b). Both doubly  
320 coordinated ( $\equiv Mn_2O^{-2/3}$ ) and singly coordinated oxygens ( $\equiv MnOH^{-1/3}$ ) showed  
321 incomplete oxygen coordination, and the resulting negative oxygen charges yielded a  
322 natural tendency toward cation binding (Peacock and Sherman, 2007b; Li et al., 2020). The  
323 triply coordinated oxygens ( $\equiv Mn_3O^0$ ) on the internal surfaces were considered inert  
324 because of the completely satisfied oxygen valence.

325 A two-site model was used to describe the birnessite surface by including both doubly  
326 coordinated interlayer sites,  $\equiv Mn_2O^{-2/3}$ , and singly coordinated edge sites  $\equiv MnOH^{-1/3}$  in  
327 this model (Zhao et al., 2018). A constant capacity model (CCM) was chosen to describe

328 the electrostatic effects of the internal surfaces because the interlayer space between two  
329 opposing  $\text{MnO}_6$  surfaces is too small for a detailed electrical double layer model (Appelo  
330 and Postma, 1999). The CCM model was based on the following two assumptions: 1)  
331 mineral surfaces only hold one adsorption electrostatic plane, and 2) all surface species  
332 were exclusively inner-sphere complexes. The capacitance for the internal sites in the CCM  
333 model was optimised based on the adsorption data.

334 For the external surfaces, charge-distribution multisite ion complexation  
335 (CD-MUSIC) combined with the triple-layer model (TLM) was used (Komárek et al.,  
336 2015; Zhao et al., 2018; Li et al., 2020). Two adsorption electrostatic planes (0-, 1-plane)  
337 and one diffusion plane (2-plane) are included in the TLM model. The capacitances of the  
338 inner and outer layers in the TLM model were set equal ( $C_1 = C_2$ ) based on previous  
339 experimental observations (Sverjensky, 2005). It was supposed that charges of Zn surface  
340 complexes distributed between the 1-plane and the 0-plane.

341 ECOSAT 4.9 and the associated fitting program FIT (Keizer and van Riemsdijk, 2009)  
342 were used to model the adsorption reactions on the basis of the speciation data obtained  
343 from EXAFS and DFT (which showed that the two most likely Zn species associated with  
344 birnessite vacancy and edge sites were TCS and DCS configurations, respectively). Zn  
345 affinity constants were calculated from the fitting of Zn adsorption data. The model  
346 parameters were finally optimized to yield the best linear correlation between the modeled  
347 and measured Zn concentrations (Fig. S2,  $R^2 = 0.90$  and  $p < 0.00001$ ).

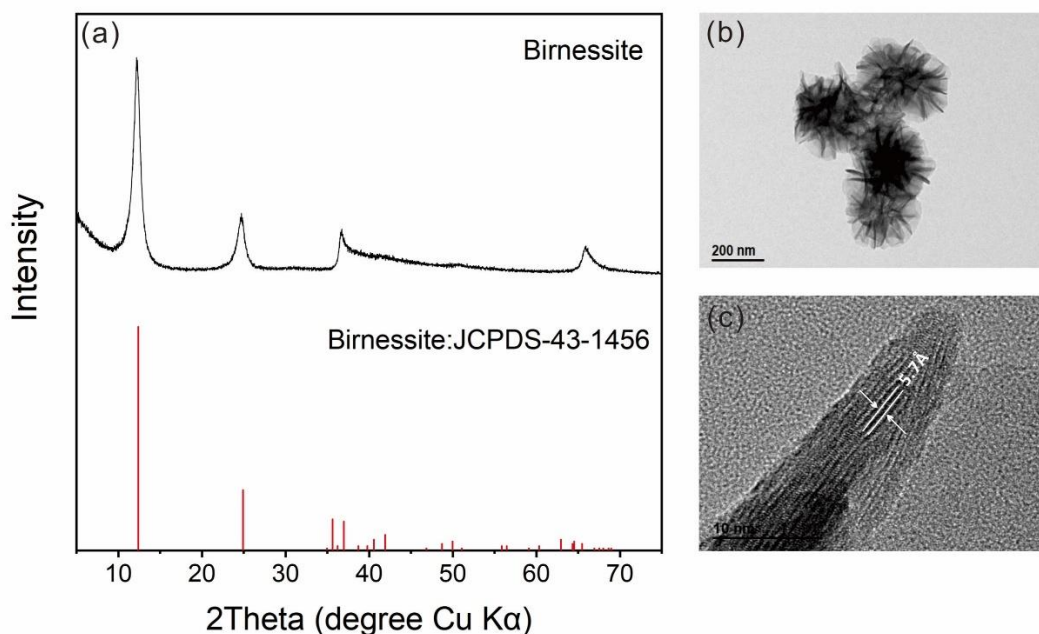
348

349

### 3. Results

#### 3.1 Characterization of synthetic birnessite

351 Synthetic birnessite (Fig. 2a) shows four broad XRD peaks at  $\sim 7.2$  Å (001),  $\sim 3.6$  Å  
352 (002),  $\sim 2.4$  Å (100), and  $\sim 1.4$  Å (110), which match well with those of typical birnessite  
353 (JCPDS 23–1239). The TEM image (Fig. 2b) shows that the particles of synthetic birnessite  
354 consist of clusters or spheroidal aggregates with a size of approximately 200 nm, in  
355 accordance with the morphology of balls of needles shown in previous studies (e.g., Atkins  
356 et al., 2014). The measured interlayer thickness of the birnessite is  $\sim 5.7$  Å (Fig. 2c), which  
357 is less than the 7.2 Å  $d$ -spacing determined via XRD, indicative of the loss of interlayer  
358 water that can occur under the high vacuum conditions required for TEM measurement  
359 (Atkins et al., 2014). The characteristic diffraction peaks together with spheroidal  
360 morphologies confirm that birnessite was successfully synthesized.



361

362

363 **Fig. 2.** XRD patterns (a) and TEM images (b, c) of synthetic birnessite. PDF JCPDS-43-  
 364 1456 for birnessite is shown for visual comparison.

365

### 366 **3.2 Batch adsorption experiments**

367 The result of the kinetic experiment on Zn adsorption to birnessite is presented in Fig.

368 [3a](#). Within the first 24 h, a relatively fast reaction occurs, accounting for ~35% of the total

369 adsorbed Zn. Following the fast adsorption stage, the reaction continues for ~120 h (~42%

370 of the total Zn). At longer reaction times, only a small amount of additional adsorption

371 occurs. After 432 h, ~49% of the initially added Zn is adsorbed. 120 h of reaction time is

372 selected for subsequent experiments.

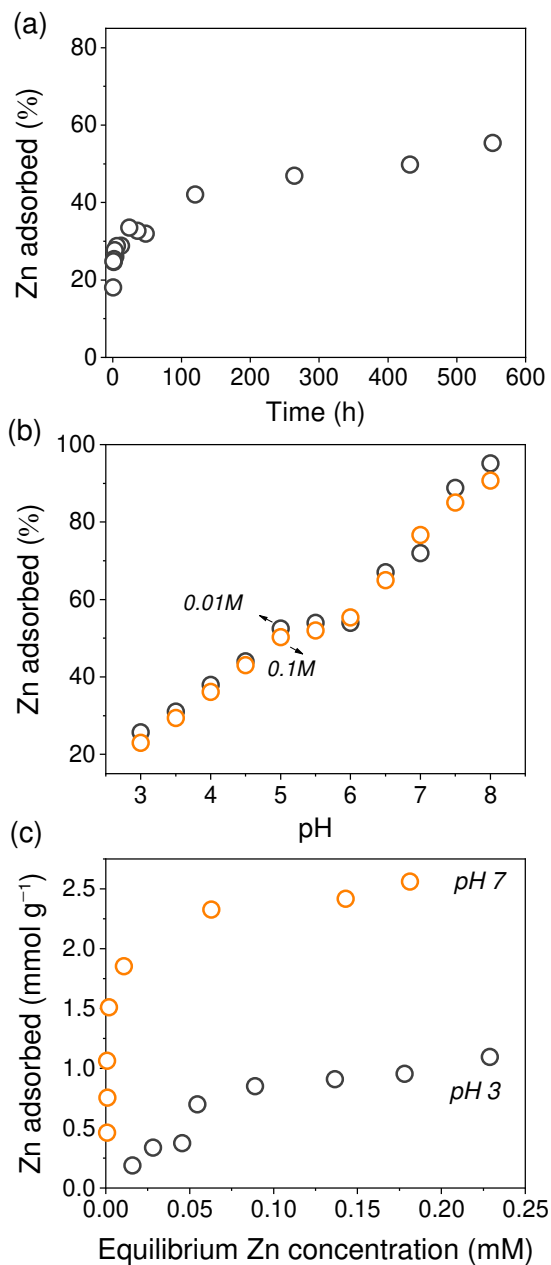
373 The effects of pH and ionic strength on Zn adsorption on birnessite are shown in Fig.

374 [3b](#). Typically, an increase in the Zn adsorption percentage from ~36% to ~96% occurs with

375 increasing pH from 3–9. Zinc adsorption did not significantly vary at ionic strength of  
376 either 0.01 or 0.1 M, suggesting the formation of an inner-sphere surface complex since  
377 metal adsorption via this mechanism can occur regardless of background electrolyte  
378 competition (Roberts et al., 2003).

379 Isotherm experiments of Zn on birnessite at pH 3 and 7 at ionic strength of 0.1 M  
380 are shown in Fig. 3c. At pH 3, Zn adsorption slowly increases with Zn solution  
381 concentration and reaches a plateau ( $\sim 1.0 \text{ mmol g}^{-1}$ ) at high Zn concentrations. In contrast,  
382 at pH 7, Zn adsorption shows a sharp increase with increasing Zn concentration in solution,  
383 resulting in a large Zn adsorption that is more than 2.5 times higher than that at pH 3.  
384 Mechanisms for the rapid increase of Zn adsorption at pH 7 is further discussed using the  
385 surface complexation modeling described in the following section.

386



387  
 388 **Fig. 3.** (a) Kinetics of Zn adsorption onto birnessite at pH 6; (b) Zn adsorption as a function  
 389 of pH and ionic strength (0.01/0.1 M NaNO<sub>3</sub>); (c) Adsorption isotherm of Zn at pH 3 and  
 390 of pH and ionic strength (0.01/0.1 M NaNO<sub>3</sub>); (c) Adsorption isotherm of Zn at pH 3 and  
 391 experiments. The Zn concentration was set to 0.2 mM except for the isotherm  
 392 was 25 ± 1 °C.

393

394

### 395 **3.3 Isotopic fractionation of Zn caused by adsorption**

396 [Fig. 4a](#) shows the changes in Zn isotopic compositions ( $\delta^{66}\text{Zn}$ ) of the aqueous and  
397 adsorbed Zn as a function of time ([Table 1](#)). Compared to the  $\delta^{66}\text{Zn}$  value of the initial Zn  
398 solution (i.e.,  $+0.28 \pm 0.06\text{‰}$ , 2 SD,  $n = 6$ ), lighter Zn isotopes are preferentially adsorbed  
399 onto the birnessite surface, resulting in heavier isotopes enriched in the aqueous phase.  
400 During the first  $\sim 120$  h of reaction, the  $\delta^{66}\text{Zn}$  value of the adsorbed phase increases rapidly  
401 from  $-0.61\text{‰}$  to  $0.08\text{‰}$ . Simultaneously, the  $\delta^{66}\text{Zn}$  value of aqueous Zn decreases from  
402  $0.62\text{‰}$  to  $0.44\text{‰}$ . As the adsorption experiment continues from  $\sim 120$  h to  $\sim 432$  h, the  $\delta^{66}\text{Zn}$   
403 values of the adsorbed ( $0.14 \pm 0.08\text{‰}$ ) and aqueous phases ( $0.44 \pm 0.05\text{‰}$ ) remain almost  
404 constant. The corresponding isotopic fractionation value ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$ ) changes from  
405  $-1.23\text{‰}$  to  $-0.36\text{‰}$  during the first  $\sim 120$  h and then remains nearly unchanged as the  
406 experiment proceeds ([Fig. 4b](#) and [Table 2](#)). This result indicates that fractionation  
407 equilibrium is achieved after approximately 120 h of adsorption.

408 The pH-dependent changes in the isotopic compositions of Zn are shown in [Fig. 5](#)  
409 and [Table 1](#). The  $\delta^{66}\text{Zn}$  value of the adsorbed phase increases gradually from  $-0.06\text{‰}$  to  
410  $0.28\text{‰}$  over a pH range of 3 to 9. The  $\delta^{66}\text{Zn}$  value of the aqueous phase remains nearly  
411 unchanged at approximately  $0.45\text{‰}$  at pH 3 to 6 before it decreases to  $\sim 0.35\text{‰}$  and then  
412 remains almost unchanged when the pH value of the starting solution increases from 7 to  
413 9. The corresponding  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  remains almost the same, with an average value



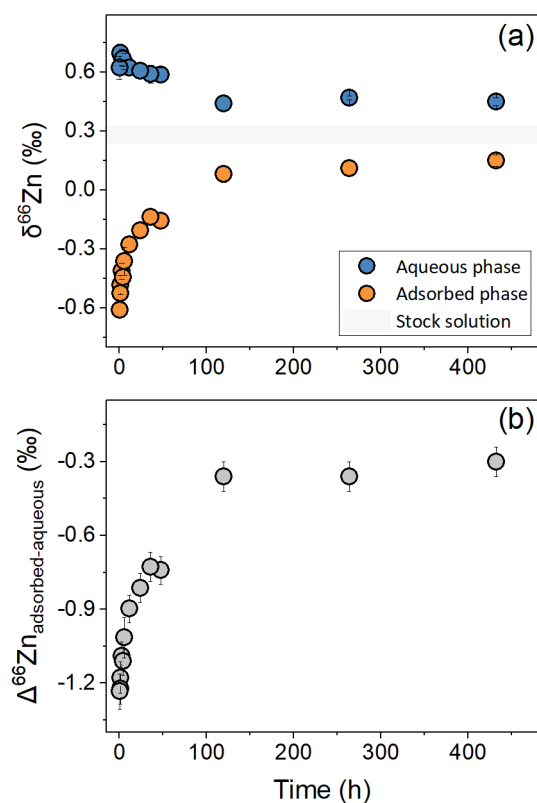
414 of  $-0.44 \pm 0.05\text{‰}$  in the pH 3 to 5 range, while the  $\Delta^{66/64}\text{Zn}_{\text{adsorbed-aqueous}}$  value increases  
415 from  $-0.25 \pm 0.06\text{‰}$  to  $-0.09 \pm 0.06\text{‰}$  in the pH 6 to 9 range (Fig. 5b).

416 The adsorption isotherm experiment was conducted to investigate the Zn isotopic  
417 fractionation behavior during adsorption at pH 3 (Fig. 5 and Table 1). Continuous  
418 enrichment of light Zn isotopes on the birnessite surface relative to aqueous solution is  
419 observed (Fig. 5c). As the initial Zn concentration increases, the  $\delta^{66}\text{Zn}$  value of the  
420 adsorbed phase changes from  $0.05\text{‰}$  at  $[\text{Zn}]_{\text{initial}} = 0.05 \text{ mM}$  to  $-0.13\text{‰}$  at  $[\text{Zn}]_{\text{initial}} = 0.3$   
421  $\text{mM}$ . Aqueous compositions decrease correspondingly from  $0.46\text{‰}$  at  $[\text{Zn}]_{\text{initial}} = 0.05 \text{ mM}$   
422 to  $0.37\text{‰}$  at  $[\text{Zn}]_{\text{initial}} = 0.3 \text{ mM}$ . The corresponding  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  remains nearly  
423 constant at  $-0.45 \pm 0.06\text{‰}$  (Fig. 5d). The mass balance offset is within  $\pm 0.05\text{‰}$  of the  
424 isotopic composition of the starting solution (Table 1), suggesting the robustness of the  
425 isotopic data.

426

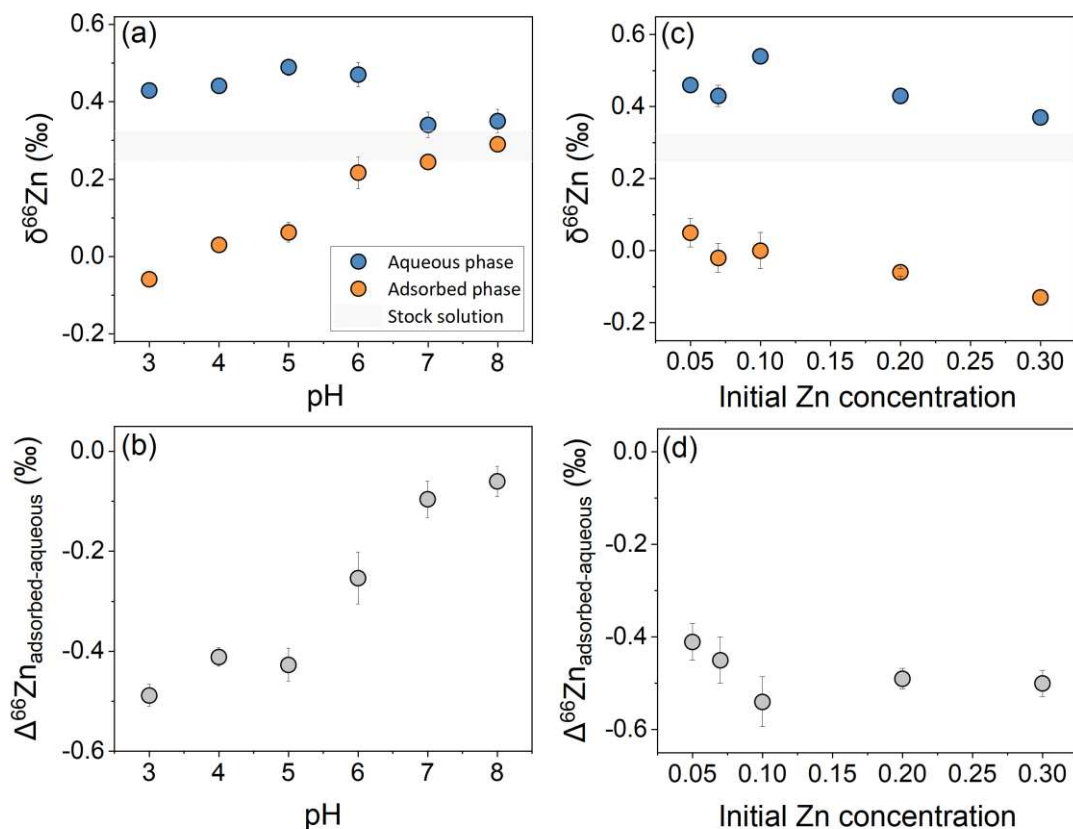
427

428



429

430 **Fig. 4.** Changes in the isotopic composition of Zn (a) and the corresponding isotopic  
 431 fractionation ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$ ) (b) with respect to different reaction time. Experiments  
 432 were performed at a total Zn concentration of 0.2 mM at pH 6 at room temperature.



433

434

435 **Fig. 5** Isotopic compositions and isotopic fractionation of Zn for adsorption edge  
 436 experiments at a total Zn concentration of 0.2 mM (a, b) and for adsorption isotherm  
 437 experiments at pH 3 (c, d).

438

439

### 440 3.4 Zinc bonding structure as determined by EXAFS

441 The  $k^3$ -weighted EXAFS spectra of the Zn-birnessite samples and four Zn reference  
 442 compounds (i.e., chalcophanite ( $\text{ZnMn}_3\text{O}_7 \cdot 3\text{H}_2\text{O}$ , a reference for octahedral Zn ( $^{\text{VI}}\text{Zn}$ ) on  
 443 birnessite), tetrahedral Zn ( $^{\text{IV}}\text{Zn}$ ) on birnessite, hydrozincite ( $\text{Zn}_3(\text{OH})_6(\text{CO}_3)_2$ ), and Zn  
 444 hydroxide ( $\text{Zn}(\text{OH})_2$ ) are shown in Fig. 6a. Chalcophanite consists of layered edge-sharing

445  $\text{Mn}^{4+}$ -O octahedra, where one in seven octahedra are vacancies (Post and Appleman, 1988).  
446 In chalcophanite, Zn octahedrally coordinated to three O atoms of Mn vacancies and three  
447 O atoms of interlayer waters (Fig. 1b). In the reference compound  $^{\text{IV}}\text{Zn}$  on birnessite, Zn is  
448 coordinated to three O atoms of the Mn vacancy and one water molecule (Fig. 1a)  
449 (Manceau et al., 2002). For Zn hydroxide and hydrozincite, Zn exists in tetrahedral  
450 coordination and in both tetrahedral and octahedral coordination to O atoms in polymerised  
451 forms, respectively (Ghose, 1964; Christensen, 1969).

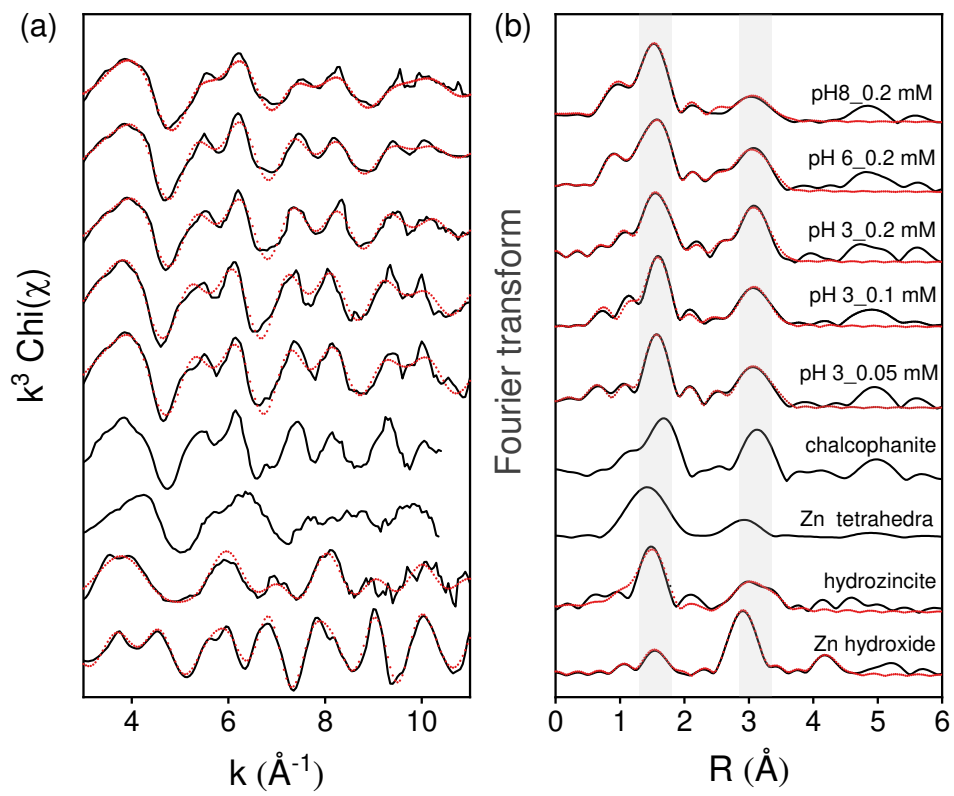
452 The EXAFS oscillations of three samples prepared for Zn concentrations of 0.05–0.2  
453 mM at pH 3 are similar to those of chalcophanite (Fig. 6a). Correspondingly, Fig. 6b shows  
454 the Fourier transformed spectra (phase shift uncorrected) of the three samples, where two  
455 main peaks are observed. These two peaks can be assigned to O and Mn backscattering  
456 based on the structure of chalcophanite. Shell-by-shell fitting results are shown in Table 2,  
457 showing first-shell Zn–O bond distances of 2.03–2.04 Å and 2.15–2.18 Å. These distances  
458 are in good agreement with those expected for the distorted octahedra of Zn on birnessite,  
459 where Zn coordinates one side to three O atoms of the Mn vacancy (2.01–2.02 Å) and on  
460 the other side to three H<sub>2</sub>O molecules (2.15–2.16 Å) (Manceau et al., 2002). The second  
461 shell fitting returns a Zn–Mn distance of 3.50–3.52 Å and coordination numbers of 6.3–7.7,  
462 which are consistent with the coordination of Zn as the TCS complex (Fig. 1b) above Mn  
463 vacancies in birnessite sheets (e.g., Drits et al., 1997; Toner et al., 2006). This configuration  
464 involves coordination of Zn to three O atoms at the Mn vacancy, which are shared by 6 Mn

465 atoms surrounding the vacancy, as occurs in chalcophanite. Previous work also found TCS  
466 tetrahedral Zn on vacancies, indicated by relatively shorter Zn–O ( $\sim 1.97$  Å) and Zn–Mn  
467 ( $\sim 3.35$  Å) interatomic distances (Marcus et al., 2004; Hinkle et al., 2017), which was not  
468 observed in our samples.

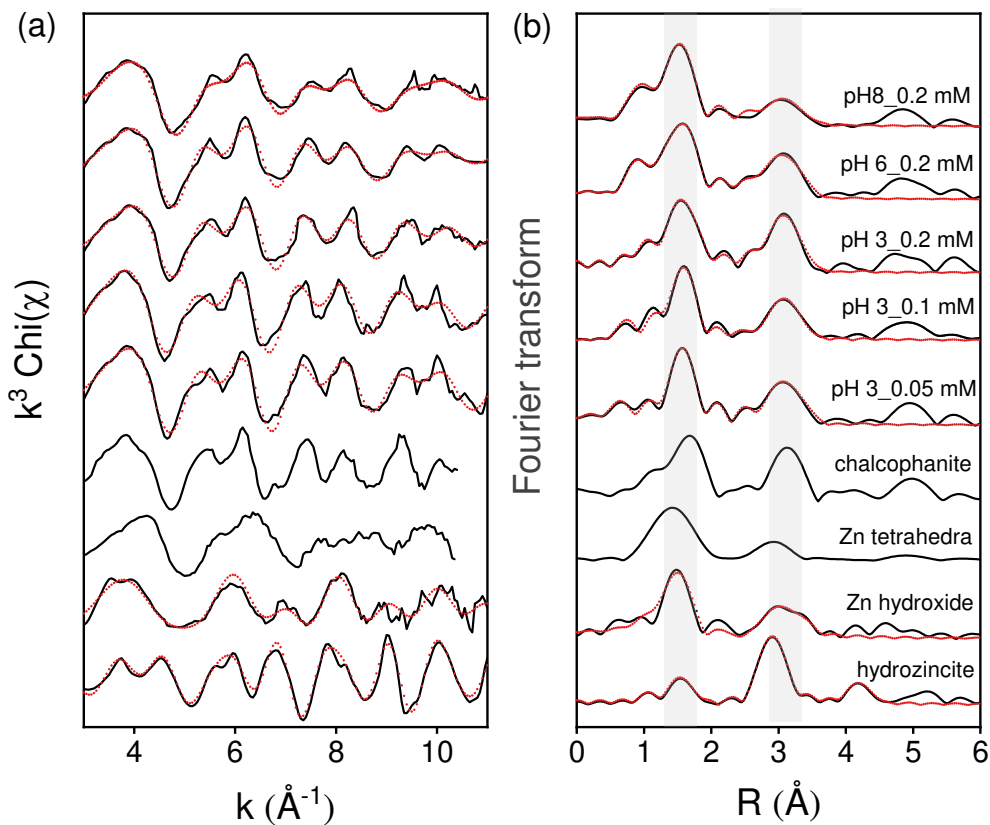
469 In contrast, the samples prepared for Zn concentration of 0.2 mM at pH 6 and pH 8  
470 exhibit slightly different features, with a “beat” pattern weakened at higher  $k$  between 7  
471 and  $8.5$  Å<sup>-1</sup>. This suggests that either the adsorption site or the bonding geometry of Zn  
472 complexes varies with Zn surface coverage (Fig. 6a). The beat patterns of the spectra  
473 prepared at pH 6 and 8 are obviously different from those of Zn hydroxide (Zn(OH)<sub>2</sub>) and  
474 hydrozincite (Zn<sub>5</sub>(CO<sub>3</sub>)<sub>2</sub>(OH)<sub>6</sub>) (Fig. 6a), indicating that polymerizing Zn as either Zn  
475 hydroxide or hydrozincite is highly unlikely. Thus, only Zn surface complexes on birnessite  
476 are considered in our systems. In the R-space, the first Zn–O shell shifts to lower values,  
477 indicating that the average Zn–O bond length ( $R_{\text{Zn-O}}$ ) decreases. Significantly lower  
478 amplitudes of the second shell for the Zn–Mn pair are observed. This low amplitude  
479 indicates a lower proportion of Mn near-neighbors per Zn adsorbed on birnessite. The fit  
480 of the first shell is employed using two sets of Zn–O shells, one corresponding to octahedral  
481 Zn (<sup>VI</sup>Zn–O) and the other to tetrahedral Zn (<sup>IV</sup>Zn–O). Such a two-shell fitting approach has  
482 been successfully used in previous studies (Lefkowitz and Elzinga, 2015). The fit yields a  
483 bond length for <sup>IV</sup>Zn–O of 2.01–2.04 Å and for <sup>VI</sup>Zn–O of 2.17–2.18 Å (Table 2). These  
484 two-shell fit results are consistent with those of Manceau et al. (2002), who reported radial

485 distances for  $^{IV}\text{Zn-O}$  in the range of 1.97-2.02 Å and for  $^{VI}\text{Zn-O}$  in the range of 2.15–2.20  
486 Å. Despite clear first-shell evidence for two Zn-adsorbed species, the second shell could  
487 be well fitted using a single Zn–Mn pair, and the fit was not improved by the addition of a  
488 second Zn–Mn pair. This fit yields an interatomic distance for Zn–Mn of 3.49 Å and CN  
489 of 3.5. This second shell Zn-Mn distance of 3.49 Å is consistent with the coordination of  
490 Zn as triple-corner-sharing complexes above Mn vacancies ([Manceau et al., 2002](#); [Hinkle  
491 et al., 2017](#)). The fitted CN of 3.5 is much smaller, however, than the CN values of 6 in  
492 triple-corner-sharing complexes above Mn vacancies. Previous studies have shown that the  
493 formation of bidentate corner-sharing and edge-sharing (i.e., DCS and DES) complexes of  
494 metals (e.g., Pb, Ni and Zn) at birnessite edge sites results in CN values of ~2 ([Villalobos  
495 et al., 2005](#); [Simanova et al., 2015](#)). DES Zn complexes are unlikely to form because Zn is  
496 observed to exist in octahedral coordination (average  $R_{\text{Zn-O}}$ : ~2.15 Å) based on our DFT  
497 calculation ([Fig. 7](#); [Table 3](#)), and the formation of DES octahedral Zn complexes  
498 contradicts the relatively short Zn-O length. We thus propose that the Zn–Mn interatomic  
499 distance, 3.49 Å, is a weighted average between TCS octahedral and DCS tetrahedral Zn  
500 complexes. The relatively large Debye-Waller factor  $\sigma^2$  of the second shell Zn-Mn  
501 correlation (0.015 Å<sup>2</sup>, [Table 2](#)) also indicates that more than one Zn–Mn pair is present in  
502 the second shell.

503



504



505  
 506 **Fig. 6** (a)  $k^3$ -weighted EXAFS spectra of the Zn-birnessite samples and model compounds;  
 507 (b) corresponding Fourier transforms in R space. Black solid lines and red dotted lines  
 508 depict experimental and fitting data, respectively. Chalcophanite and Zn tetrahedra are two  
 509 reference standards representing octahedral and tetrahedral Zn environment adsorbed on  
 510 birnessite vacancies (Manceau et al., 2002).

511

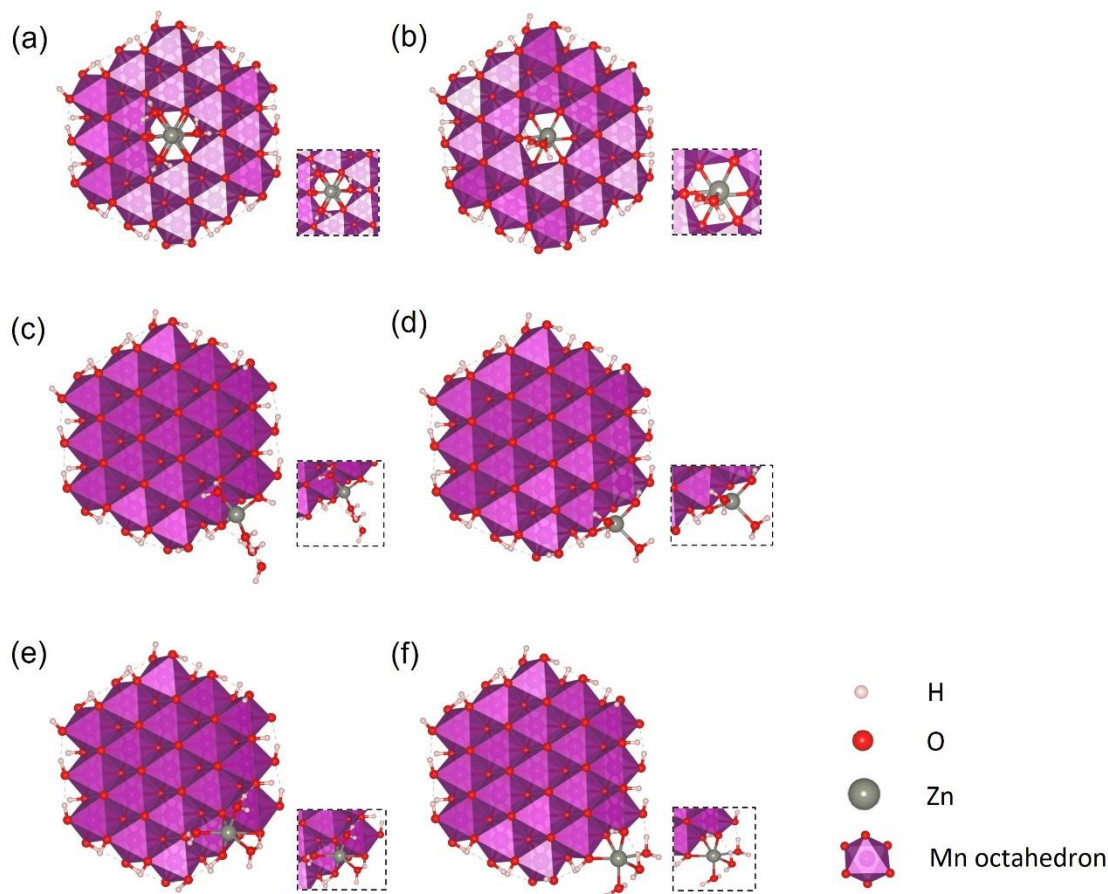
### 512 3.5 Geometry optimization of Zn surface species

513 Geometry-optimized structures of Zn surface species are presented in Fig. 7 and Table  
 514 3. At the vacancy site, Zn formed octahedral or tetrahedral coordination depending on  
 515 hydrating H<sub>2</sub>O molecules (Fig. 7a and b). Octahedral Zn TCS shows larger Zn-O and Zn-  
 516 Mn distances than tetrahedral Zn TCS (Table 3), which is consistent with previous work



517 (Post and Appleman, 1988; Manceau et al., 2002) and current EXAFS results (Table 2). At  
518 the lateral edge sites, however, we are not able to obtain six-fold coordinated DCS  
519 complexes. During the geometry optimization of the initial six-fold coordinated Zn DCS  
520 model (i.e., Zn·4H<sub>2</sub>O in Table 3), two H<sub>2</sub>O molecules leave the initially coordinating Zn  
521 cation and move to surface OH to form H-bonds with H<sub>2</sub>O and surface OH (Fig. 7c),  
522 resulting in four-fold coordinated Zn DCS, just as in the geometry optimization of the  
523 Zn·2H<sub>2</sub>O DCS model (Fig. 7d). The structural parameters of the Zn·4H<sub>2</sub>O DCS model are  
524 similar to those of the Zn·2H<sub>2</sub>O DCS model. In a recent DFT-based molecular dynamics  
525 study of hexaaquo-zinc complexes, the GGA-PBE method that we use in this study favors  
526 tetraaquo- and pentaquo-zinc complexes, and a van der Waals correction method such as  
527 DFT–D (McNellis et al., 2009) is required to obtain a hexaaquo-zinc complex, as observed  
528 in experiments (Ducher et al., 2017). We further geometrically optimized the initial sixfold  
529 coordinated Zn·4H<sub>2</sub>O DCS model using a DFT–D method; however, the DFT–D method  
530 also predicts four-fold coordinated Zn DCS, similar to GGA-PBE (Table 3). In general, the  
531 fully occupied 3*d* state of Zn<sup>2+</sup> does not allow highly asymmetric coordination with  
532 octahedral ligands and instead favors symmetric coordination, unlike transition metal  
533 cations, whose 3*d* states are partially occupied (Jensen, 2003). Therefore, at the lateral edge  
534 sites, octahedral Zn-DCS coordinated with two surface O and four water O atoms may be  
535 electronically unstable to form. On the other hand, DES complexes exhibit six-fold  
536 coordination regardless of whether Zn is coordinated with double-coordinated O or triply

537 coordinated O, and the  $d(\text{Zn-O})$  is similar to the  $d(\text{Zn-O})$  of octahedral Zn TCS.



538

539 **Fig. 7.** DFT geometry-optimized (a) Zn·3H<sub>2</sub>O and (b) Zn·1H<sub>2</sub>O TCS complex models, (c)

540 Zn·4H<sub>2</sub>O and (d) Zn·2H<sub>2</sub>O DCS complex models, and (e) Zn·3H<sub>2</sub>O (with doubly

541 coordinated oxygen) and (f) Zn·3H<sub>2</sub>O (with triply coordinated oxygen) DES complex

542 models formed at a birnessite nanodisk. The insets are the magnification of the interested

543 area. See [Table 3](#) for the structural parameters of each model.

544

### 545 3.6 Surface complexation modeling

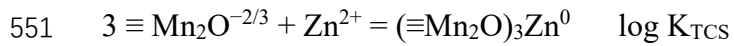
546 Based on the results from EXAFS spectroscopy and DFT geometry optimization,

547 metals can bind to both the interlayer vacancies and external edges of birnessite, and the

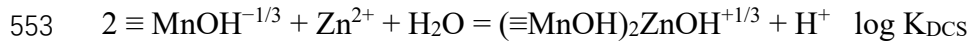
548 surface speciation of Zn on birnessite is TCS complex on vacancies and DCS complex on

549 edges. The surface adsorption reactions of Zn are thus described as follows:

550 Zn complex at internal surfaces:

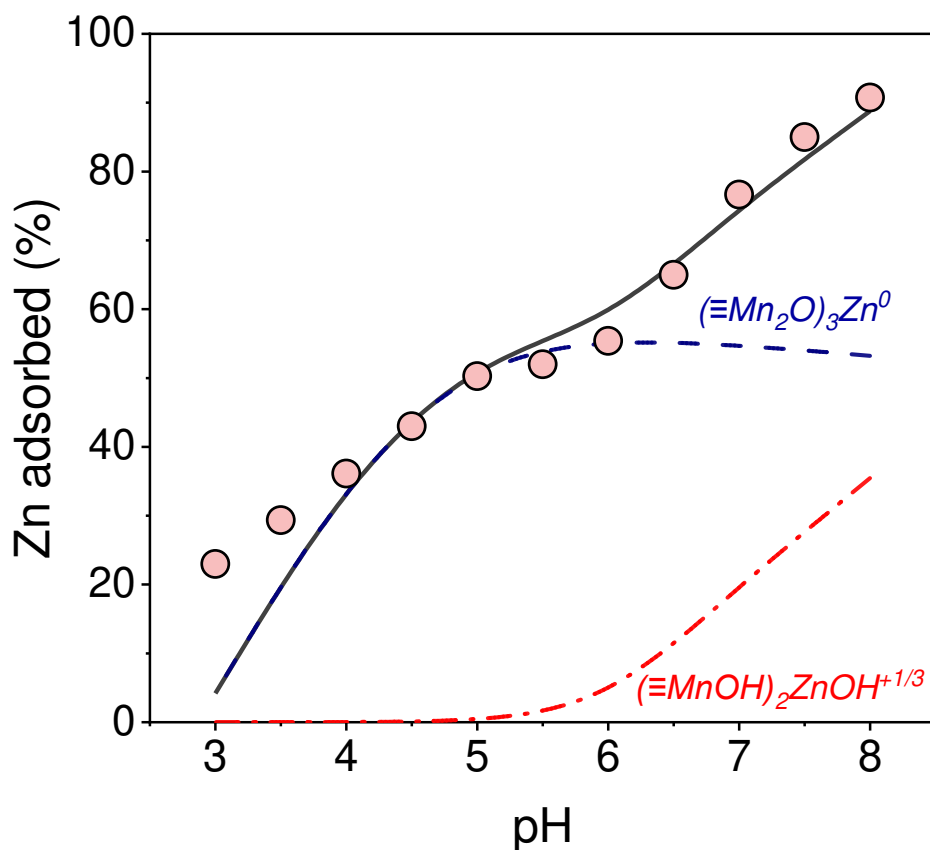


552 Zn complex at external edges:



554 The best-fit model parameters are summarized in [Table 4](#). The fitting results using two  
555 different species,  $(\equiv\text{Mn}_2\text{O})_3\text{Zn}^0$  and  $(\equiv\text{MnOH})_2\text{ZnOH}^{+1/3}$ , for the adsorption edge  
556 experiments in 0.1 M  $\text{NaNO}_3$  solution are shown in [Fig. 8](#). The results show that  
557  $(\equiv\text{Mn}_2\text{O})_3\text{Zn}^0$  is the dominant surface species, particularly at low pH values of 3 to 5, which  
558 is consistent with previous work showing that vacancy sites play a dominant role in binding  
559 metal ions due to their highly negative charge (e.g., [Toner et al., 2006](#); [Kwon et al., 2009](#);  
560 [Wang et al., 2018](#)). The proportion of  $(\equiv\text{MnOH})_2\text{ZnOH}^{+1/3}$  increases from 4.4% at pH 6 to  
561 36.2% at pH 8, suggesting that edge sites become important under alkaline conditions  
562 ([Wang et al., 2018](#); [Li et al., 2020](#)). In general, for Zn on birnessite, the model describes a  
563 two-step adsorption trend with increasing pH.

564



565  
 566 **Fig. 8** Experimental (pink points) and calculated (black line) Zn adsorbed on birnessite as  
 567 a function of pH in 0.1 M NaNO<sub>3</sub> solution. The model was fit based on the parameters in  
 568 [Table 3](#). The calculated Zn resulted from various proportions of Zn on the birnessite  
 569 internal surface (blue dashed line) and Zn on the birnessite external surface (red dotted  
 570 line).

571

572

573

## 4. Discussion

574

### 4.1 Constraining the equilibrium fractionation factors of Zn

575

Zn isotopes show different fractionation behaviour according to pH values for the

576

adsorption edge experiments ([Fig. 5a](#)). Light Zn isotopes are observed to adsorb on the

577 birnessite surface, with  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  ranging from  $-0.54 \pm 0.05\text{‰}$  to  $-0.06 \pm 0.04\text{‰}$ .  
578 An equilibrium fractionation process can better describe the evolution of Zn isotopic  
579 compositions, while a Rayleigh process is rejected (detailed in Fig. S3). In this process,  
580 two distinct fractionation factors are involved: one for low pH values when Zn adsorption  
581 is low and predominantly onto vacancy sites and the other for higher pH values when Zn  
582 adsorption is higher and onto edge sites. For different pH values, the distinct isotopic  
583 signatures of adsorbed Zn therefore result from different amounts of Zn adsorbed and  
584 different proportions of vacancy and edge Zn species. The Zn isotopic fractionation factor  
585 can be determined for each Zn species based on the changes in  $\delta^{66}\text{Zn}_{\text{adsorbed}}$  with pH.

586 In the experiments in the pH range 3 to 5, an average  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  of  $-0.46 \pm$   
587  $0.06\text{‰}$  is observed. According to our EXAFS, at low pH values, only vacancy sites  
588 contribute to Zn binding (Fig. 6). The isotopic fractionation factor for Zn adsorption as  
589 TCS octahedral surface complexes on vacancy sites is thus the value that we measured  
590 ( $\Delta^{66}\text{Zn}_{\text{TCS-aqueous}} = -0.46 \pm 0.03\text{‰}$ ; Table 1). In adsorption experiments in the pH range of 6  
591 to 9, isotopic fractionation magnitudes significantly decrease from  $-0.43 \pm 0.03\text{‰}$  to  $-0.09$   
592  $\pm 0.05\text{‰}$ . As indicated by both the EXAFS results, at higher pH values (e.g., pH 8), binding  
593 to edge sites appears and becomes increasingly important (Fig. 6). Vacancy site and edge  
594 site adsorption both contribute to the measured Zn isotopic fractionation at high pH values.  
595 A  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  of  $-0.06 \pm 0.05\text{‰}$  is observed at pH 8 (Fig. 5), and the model predicts  
596 39.9% Zn binding on edge sites and 59.9% Zn binding on vacancy sites (Fig. 8). Thus we

597 calculated the isotopic fractionation value for Zn adsorption as DCS tetrahedral surface  
 598 complexes on edge sites based on a two end-member mixing equation:

$$599 \quad \Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = \Delta^{66}\text{Zn}_{\text{TCS-aqueous}} * p_{\text{TCS}} + \Delta^{66}\text{Zn}_{\text{DCS-aqueous}} * p_{\text{DCS}} \quad (4)$$

600 where  $p_{\text{TCS}}$  and  $p_{\text{DCS}}$  are the respective proportions of vacancy and edge sites involved, and  
 601  $\Delta^{66}\text{Zn}_{\text{DCS-aqueous}}$  is the estimated isotopic fractionation value for Zn adsorption onto edge  
 602 sites. The Zn isotopic fractionation factor associated with edge sites is thus  $\Delta^{66}\text{Zn}_{\text{DCS-aqueous}}$   
 603  $= 0.52 \pm 0.03\%$ . We chose the sample obtained at pH 8 to calculate  $\alpha_{\text{DCS-aqueous}}$  because the  
 604 EXAFS spectrum (Fig. 6) confirms that Zn adsorbed on both the vacancy and edge sites of  
 605 birnessite at pH 8, and the surface complexation modeling (Fig. 8) fits the experimental  
 606 data at pH 8 very well.

607 The theoretical  $\delta^{66}\text{Zn}_{\text{aqueous}}$  is calculated to verify the correctness of the two  
 608 fractionation factors when both binding sites are involved during adsorption:

$$609 \quad \delta^{66}\text{Zn}_{\text{aqueous}} = \delta^{66}\text{Zn}_{\text{stock}} - 1000f \left( \frac{p_{\text{TCS}}(\alpha_{\text{TCS-aqueous}} - 1)}{1-f+(f*\alpha_{\text{TCS-aqueous}})} + \frac{p_{\text{DCS}}(\alpha_{\text{DCS-aqueous}} - 1)}{1-f+(f*\alpha_{\text{DCS-aqueous}})} \right) \quad (5)$$

610 where  $\alpha_{\text{TCS-aqueous}}$  and  $\alpha_{\text{DCS-aqueous}}$  are the isotopic fractionation coefficients associated with  
 611 vacancy and edge sites, respectively. The corresponding  $\delta^{66}\text{Zn}_{\text{adsorbed}}$  on the birnessite  
 612 surface is calculated from equation (6) as:

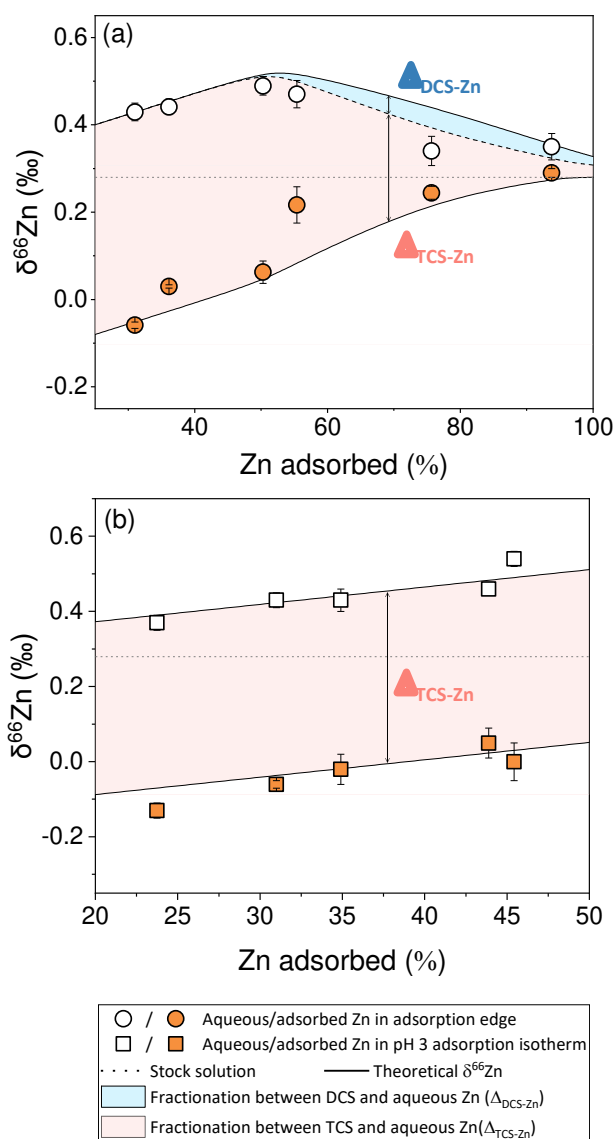
$$613 \quad \delta^{66}\text{Zn}_{\text{adsorbed}} = \delta^{66}\text{Zn}_{\text{aqueous}} + \Delta^{66}\text{Zn}_{\text{TCS-aqueous}} * p_{\text{TCS}} + \Delta^{66}\text{Zn}_{\text{DCS-aqueous}} * p_{\text{DCS}}$$

614 (6)

615 Fig. 9a shows the theoretical calculations of  $\delta^{66}\text{Zn}_{\text{adsorbed}}$  and  $\delta^{66}\text{Zn}_{\text{aqueous}}$  when both binding  
 616 sites contribute to Zn adsorption. The dashed line separates the fraction of Zn adsorbed as

617 TCS octahedral surface complexes on vacancy sites (pink area) from those adsorbed as  
618 DCS tetrahedral surface complexes on edge sites (blue area). The theoretical  $\delta^{66}\text{Zn}_{\text{adsorbed}}$   
619 and  $\delta^{66}\text{Zn}_{\text{aqueous}}$  are in good agreement with the measured Zn isotope compositions in the  
620 adsorption edge experiments. The agreement between the theoretical calculations based on  
621 the isotopic fractionation coefficients and the experimental results (Fig. 9) indicates the  
622 robustness of the two Zn isotopic fractionation factors that we propose for Zn adsorption  
623 on birnessite vacancies and edge sites. Furthermore, this good agreement between the  
624 theoretical  $\delta^{66}\text{Zn}$  based on our surface complexation model and the measured  $\delta^{66}\text{Zn}$   
625 supports the evidence of control by vacancy sites at low pH and the contribution from both  
626 the vacancy and edge sites at high pH.

627 In adsorption isotherm experiments at pH 3, the experimental data are better fitted by  
628 the equilibrium fractionation model than by the Rayleigh model (Fig. 9b and Fig. S3). This  
629 equilibrium fractionation process is consistent with previous studies on Zn fractionation  
630 caused by adsorption on various minerals (e.g., Wasylenki et al., 2014; Wang et al., 2022).  
631 Regression yields an  $\alpha_{\text{adsorbed-aqueous}}$  factor of 0.99953, which is consistent with the result  
632 ( $\alpha_{\text{TCS-aqueous}} = 0.99954$ ) for Zn adsorption on vacancy sites. The consistency between the  
633  $\alpha_{\text{adsorbed-aqueous}}$  factor for adsorption isotherm experiments at pH 3 and  $\alpha_{\text{TCS-aqueous}}$  for Zn  
634 adsorption on vacancy sites is in accordance with our EXAFS results, which shows that the  
635 vacancy sites contribute exclusively to Zn adsorption at low pH.



636

637 **Fig. 9.** Aqueous and adsorbed Zn isotopic compositions with respect to the fraction of Zn  
 638 (a) for 0.2 mM adsorption edge experiments and (b) for pH 3 adsorption isotherm  
 639 experiments. The dotted line depicts the Zn isotopic composition of the stock solution.  
 640 Error bars are 2 SD from 3 replicate analyses. Solid lines show the theoretical evolution of  
 641  $\delta^{66}\text{Zn}_{\text{adsorbed}}$  and  $\delta^{66}\text{Zn}_{\text{aqueous}}$  based on the two-site model. The dashed line separates the  
 642 fraction of  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  explained by Zn adsorption as TCS surface complexes on  
 643 vacancy sites (pink area) or DCS surface complexes on edge sites (blue area).

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645



## 646 4.2 Site-specific Zn structure controls isotopic fractionation

647 The different isotopic fractionation magnitudes of Zn at low and high pH (Fig. 10)  
648 can be interpreted by the differences in adsorption complex structures of Zn on various  
649 birnessite reactive sites because such differences in the bonding environments lead to  
650 changes in Zn-O bond strength (Schauble, 2004). According to stable isotope theory, heavy  
651 isotopes prefer to exist in stronger bonding environments (e.g., lower coordination numbers)  
652 (Bigeleisen and Mayer, 1947; Schauble, 2004).

653 The TCS octahedral Zn complexes at pH 3 on birnessite vacancy sites concentrate  
654 light Zn relative to the solution ( $\Delta^{66}\text{Zn}_{\text{TCS-aqueous}} \sim -0.46 \pm 0.04\text{‰}$ ) (Fig. 10). The octahedral  
655 complexes of Zn on birnessite indicate that the coordination number of aqueous Zn does  
656 not change during adsorption; however, significant isotopic fractionation occurs. Our  
657 EXAFS results show that the Zn octahedra on birnessite are distorted, with each Zn  
658 coordinate on one side to three oxygen atoms of the Mn vacancy (bond length: 2.03–2.04  
659 Å) and on the other side to three H<sub>2</sub>O molecules (bond length: 2.15–2.18 Å) (Table 2).  
660 Metal-oxygen polyhedral distortion is proposed as a factor driving isotopic fractionation  
661 (Nakada et al., 2017; Wasylenki et al., 2014; Yan et al., 2021). For example, the distorted  
662 structures of Ce-O coordination are speculated to be the main reason for the enrichment of  
663 light isotope on  $\delta\text{-MnO}_2$  (Nakada et al., 2017). Similarly, light Cd isotopes enriched in  
664 birnessite with the formation of a distorted Cd-O octahedron is observed (Wasylenki et al.,  
665 2014; Yan et al., 2021). In our Zn-birnessite system, the formation of Zn-O distorted

666 octahedra on vacancy sites might lead to the enrichment of light Zn. In accordance with  
667 theoretical predictions, because an average Zn-O bond length of  $\sim 2.09$  Å is obtained, this  
668 longer bond length relative to aqueous Zn ( $\sim 2.06$  Å) should lead to a weaker Zn-O bond  
669 and thus preferential enrichment of light Zn isotopes on birnessite.

670 In this study, DCS complexes of Zn on birnessite edge sites occur at higher pH values  
671 (6 to 9), and Zn adsorption on edge sites induces isotopic fractionation of  $\Delta^{66}\text{Zn}_{\text{DCS-aqueous}}$   
672  $\sim 0.52 \pm 0.04\%$  (Fig. 10). The fractionation factor of  $\Delta^{66}\text{Zn}_{\text{DCS-aqueous}}$  of  $0.52 \pm 0.04\%$   
673 seems to be comparable to those observed for Zn adsorption onto several different minerals  
674 (e.g., Juillot et al., 2008; Guinoiseau et al., 2016). Juillot et al. (2008) found that tetrahedral  
675 Zn forms on ferrihydrite, yielding a fractionation factor of  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  of  $0.52 \pm$   
676  $0.04\%$ . Similarly, a recent study found that tetrahedral Zn forms on Al oxide, leading to a  
677  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$  fractionation of  $0.49 \pm 0.06\%$  (Gou et al., 2018). In these two systems,  
678 lower coordination numbers (CN: 4) with shorter Zn-O bond lengths ( $R_{\text{Zn-O}}$ :  $\sim 1.96$  Å)  
679 relative to aqueous Zn (CN: 6,  $R_{\text{Zn-O}}$ :  $\sim 2.06$  Å) were observed, which should induce a  
680 stronger Zn-O bond for Zn adsorption, thus concentrating heavy Zn isotopes. It is  
681 noteworthy that the adsorption of Zn to amorphous  $\text{SiO}_2$  surfaces results in a much greater  
682 Zn isotope fractionation ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} \sim 0.8\text{--}1.2\%$ ) (Nelson et al., 2017). In this case,  
683 for amorphous  $\text{SiO}_2$ , EXAFS data show a tetrahedral complex but with a particularly short  
684 Zn-O bond length ( $\sim 1.93$  Å). In contrast, although a similar Zn isotope fractionation  
685 ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} \sim 0.49 \pm 0.06\%$ ) during adsorption is reported for kaolinite (Guinoiseau

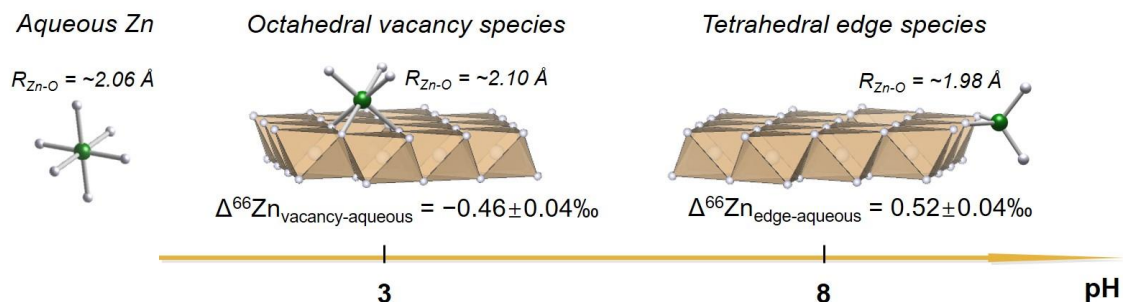
686 [et al., 2016](#)), Zn is observed to exist in octahedral coordination at the kaolinite surface, with  
687 a Zn–O distance (2.07 Å) similar to that of aqueous Zn (2.06 Å) ([Nachtegaal and Sparks,](#)  
688 [2004](#)). In the kaolinite system, the CN and Zn-O bond lengths appear to remain unchanged  
689 during Zn adsorption. The significant isotopic fractionation without a change in  
690 coordinated structure (i.e., CN and  $R_{\text{Zn-O}}$ ) is likely due, however, to a mismatch between  
691 the isotopic fractionation and EXAFS spectra because EXAFS measurements are  
692 performed at a much higher Zn concentration, and only one sample is analyzed ([Nachtegaal](#)  
693 [and Sparks, 2004](#); [Guinoiseau et al., 2016](#)). Unfortunately, additional EXAFS studies are  
694 not available to further constrain the fractionation mechanism of Zn isotope fractionation  
695 caused by adsorption on kaolinite. In our Zn-birnessite system, although we do not  
696 experimentally obtain the end-member structure of DCS complexes of Zn on birnessite,  
697 DCS Zn complexes always form along with TCS Zn complexes and the shorter EXAFS-  
698 derived Zn-O bond distances of the mixed DCS and TCS octahedral Zn (as opposed to  
699 exclusively octahedral Zn), and DFT calculations of Zn-DCS both confirm that DCS Zn  
700 exists in tetrahedral coordination ([Fig. 6, Table 3](#)). This formation of DCS tetrahedral Zn  
701 with shorter Zn-O length (1.97-1.98 Å) relative to aqueous Zn (~2.06 Å) seems to be  
702 responsible for the observed isotopic fractionation during Zn adsorption to the edge sites  
703 of birnessite.

704 At a given temperature and pressure, equilibrium isotopic fractionation factors of an  
705 element are theoretically predicted to be controlled by the relative bond strengths between

706 that element and its coordinating atoms (Bigeleisen and Mayer, 1947; Schauble, 2004),  
707 which are jointly determined by multiple factors, including bond length, coordination  
708 number, oxidation state and electronic configuration (Schauble et al., 2004; Young et al.,  
709 2009). In general, longer bond lengths correspond to weaker bond strengths and have lower  
710 vibrational frequencies, and thus elements in longer bonding environments should prefer  
711 to incorporate lighter isotopes relative to those in shorter bonding environments. A  
712 correlation between the average Zn–O bond lengths obtained from EXAFS spectroscopy  
713 and their corresponding isotopic fractionation magnitudes caused by adsorption on a  
714 variety of different mineral surfaces (Fig. 11) suggests that there is a general rule that Zn–  
715 O bond lengths impart a first-order control on Zn isotope fractionation. Specifically, the Zn  
716 with the longest Zn–O bond distance ( $2.10 \pm 0.02 \text{ \AA}$ ) is related to the enrichment of the  
717 lightest Zn isotopes ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = \sim -0.46\text{‰}$ ), while the Zn with the shortest Zn–O  
718 length ( $1.93 \text{ \AA}$ ) is related to the enrichment of the heaviest Zn ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = \sim$   
719  $0.8\text{--}1.2\text{‰}$ ). In the work here Zn–O bond length is interpreted as a function of tetrahedral  
720 vs octahedral complex, but in addition the extended bonding environment around adsorbed  
721 Zn may also influence the first shell bond length, especially for Zn adsorption onto different  
722 minerals.

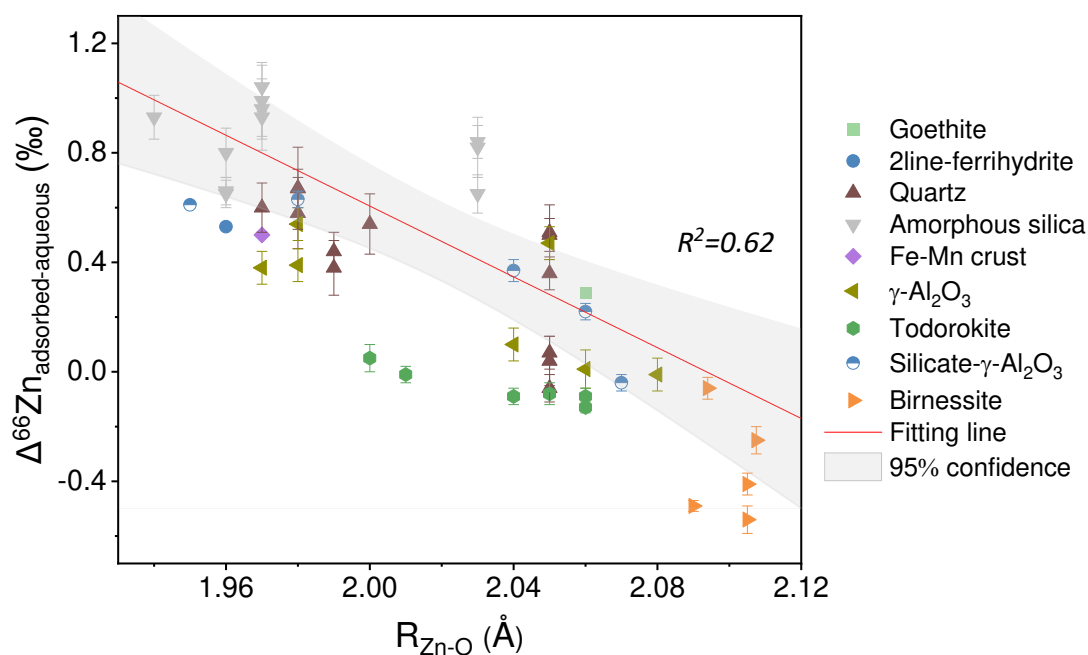
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**Fig. 10.** Schematic representation of site-specific Zn isotopic fractionation caused by adsorption on birnessite and the dominant Zn species on birnessite at different pH values.



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**Fig. 11.** Correlation of isotopic fractionation ( $\Delta^{66}Zn_{\text{adsorbed-aqueous}}$ ) caused by sorption to various minerals and the associated structural chemistry parameters (i.e.,  $R_{Zn-O}$ ). Plotted data refer to those reported previously for goethite (Juilot et al., 2008), 2-line ferrihydrite (Juilot et al., 2008), quartz (Nelson et al., 2017), amorphous silica (Nelson et al., 2017), Fe-Mn nodules (Little et al., 2014a),  $\gamma\text{-Al}_2\text{O}_3$  (Gou et al., 2018), todorokite (Wang et al., 2022), silicate- $\gamma\text{-Al}_2\text{O}_3$  (Gou et al., 2022), and birnessite (this study). The solid red fitting line was obtained by linear regression analysis. The gray area depicts the 95% confidence band. The error bars represent an analytical uncertainty of 2 SD.

#### 740 4.3 Implications for understanding Zn isotope fractionation in natural birnessite-

741 **containing settings**

742 Birnessite is a major mineral phase in freshwater (Manceau et al., 2007) and many  
743 soils (Hochella et al., 2005; Lanson et al., 2008), as well as an important component in  
744 marine sediments and ferromanganese crusts (Post, 1999; Bodei et al., 2007). It is also  
745 commonly found as a secondary product in metal ore deposits (Post, 1999; Spinks and  
746 Uvarova, 2019). Due to its unusually high adsorption capacities and scavenging  
747 capabilities, birnessite readily participates in adsorption and oxidation–reduction reactions  
748 and therefore provides one of the primary controls on trace metal cycling in Earth’s surface  
749 environments (Post, 1999; Peacock and Sherman, 2007b; Atkins et al., 2016).

750 Zinc, as an important micronutrient required by organisms (Sinoir et al., 2012) and at  
751 elevated concentrations a toxic contaminant (Sandstead, 2014), has been shown to  
752 commonly coexist with birnessite in natural environments, such as acid mine drainage sites,  
753 soils, and marine sediments (Kay et al., 2004; Mayanna et al., 2015). Zinc isotopes may  
754 fractionate during environmental interactions with birnessite (e.g., Bryan et al., 2015). In  
755 this study, we show that light Zn isotopes are preferentially adsorbed on birnessite surfaces,  
756 with a fractionation factor of  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = -0.46 \pm 0.04\text{‰}$  at acidic pH and a low  
757 Zn/Mn molar ratio (0.037–0.170) compared to a less fractionation factor ( $\Delta^{66}\text{Zn}_{\text{adsorbed-}}$   
758  $\text{aqueous} \sim -0.09 \pm 0.04\text{‰}$ ) at circumneutral pH and a high Zn/Mn molar ratio (0.170–0.327)  
759 (Fig. 12). Combining our results with those in the literature, however, we can chart and  
760 predict the evolution of Zn isotope fractionation with Zn speciation as a function of pH and

761 Zn/Mn molar ratio (Manceau et al., 2002; Grangeon et al., 2012) (Fig 12). Specifically, in  
762 hydrogenetic marine Mn crusts (birnessite as the main Mn mineral phase), pH and Zn/Mn  
763 molar ratio appear to be decoupled, in that pH is high (pH~8) but Zn/Mn molar ratio is very  
764 low (~0.004), yet Zn is reported to form TCS tetrahedral surface complexes on vacancy  
765 sites of birnessite (Little et al., 2014a) with a fractionation between crusts and seawater of  
766 ~0.5‰ (e.g., Marechal et al., 2000; Conway et al., 2013). This suggests that at low surface  
767 coverage, the TCS tetrahedral complex on vacancy sites forms independently of pH  
768 (vacancy tetrahedra  $V_{tet}$ , Fig. 12). However, as the surface coverage increases, Zn isotope  
769 fractionation should decrease because of the formation of TCS octahedral Zn at higher  
770 Zn/Mn ratios (Manceau et al., 2002) ( $V_{tet}$  and vacancy octahedra  $V_{oct}$ , Fig. 12). Then, at  
771 higher surface coverage, we still observe that at Zn/Mn ratios from ~0.037 to ~0.170, light  
772 Zn isotopes are adsorbed at the birnessite surface, with a fractionation factor  $\Delta^{66}Zn_{adsorbed-}$   
773  $aqueous$  of ~-0.46‰ where Zn forms fully TCS octahedral complexes ( $V_{oct}$ , Fig. 12). Then,  
774 as the surface coverage and pH further increase (i.e.,  $0.170 < Zn/Mn < 0.327$ , pH 6–9), we  
775 observe that Zn isotope fractionation decreases with a fractionation factor  $\Delta^{66}Zn_{adsorbed-}$   
776  $aqueous$  of ~ -0.09‰, where Zn exists in both TCS octahedral and DCS tetrahedral  
777 coordination ( $V_{oct}$  and edge tetrahedra  $E_{tet}$ , Fig. 12). It should be noted that the exact pH  
778 and Zn/Mn molar ratios correlating to the associated Zn end-member structures (i.e., TCS  
779 tetrahedral  $V_{tet}$ , TCS octahedral  $V_{oct}$ , and DCS tetrahedral Zn  $E_{tet}$ ), which are then used to  
780 predict Zn isotope fractionation, may change with respect to different structural

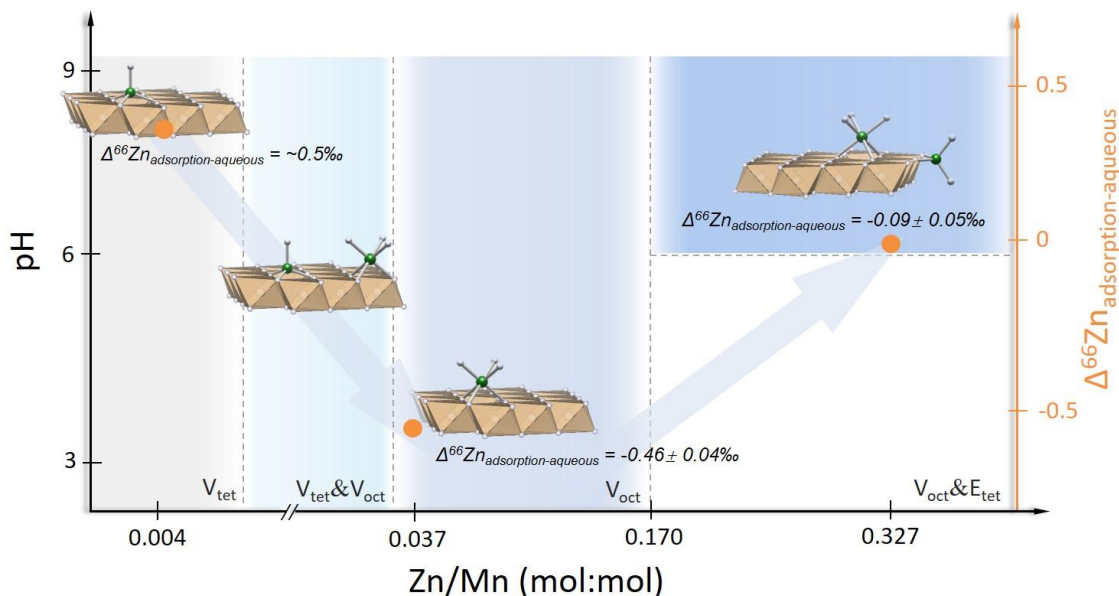
781 characteristics of birnessite (e.g., quantity of Mn vacancies and Mn AOS) (Manceau et al.,  
782 2002; Kwon et al., 2009; Yin et al., 2018), but this should not change the overall predicted  
783 evolutionary trend of Zn isotope fractionation as a function of pH and Zn/Mn ratios.

784 Simplified conditions are used in the adsorption experiments (absence of multiple  
785 ligands and a focus on one mineral phase), and Zn isotopic fractionation may be influenced  
786 by complexation with different ligands and by adsorption to other mineral surfaces (e.g.,  
787 Fe oxides) in natural environments (Fujii et al., 2014; Spinks and Uvarova, 2019). The  
788 exact fractionation magnitudes of Zn caused by natural birnessite adsorption could thus be  
789 either slightly smaller (e.g., as influenced by chloride complexation) (Fujii et al., 2014) or  
790 larger (e.g., influenced by ferrihydrite adsorption) (Juillot et al., 2008) than the values from  
791 our experiments. Therefore, in natural conditions, multiple factors, such as  
792 organic/inorganic ligands, different mineral surfaces, and solution conditions, should be  
793 taken into consideration when studying the fractionation magnitudes of Zn isotopes caused  
794 by birnessite adsorption. It is apparent, however, that Zn isotope fractionation caused by  
795 adsorption on birnessite is significant and should be carefully considered when using Zn  
796 isotopes to investigate processes in natural Mn-bearing environments.

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800 **Fig. 12** Schematic illustration of the evolution of Zn speciation on birnessite and the  
 801 predicted isotopic fractionation (orange circles) as a function of pH values and Zn/Mn  
 802 molar ratios. Dashed lines and shaded areas are drawn based on Zn isotopic data and  
 803 EXAFS spectroscopic analyses.  $V_{tet}$  and  $V_{oct}$  refer to triple-corner-sharing TCS Zn  
 804 complexes, existing in tetrahedral and octahedral coordination on vacancy sites,  
 805 respectively.  $E_{tet}$  refers to double-corner-sharing DCS tetrahedral Zn complexes on edge  
 806 sites. The arrows depict the predicted trend of Zn isotope fractionation according to Zn  
 807 speciation.

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## 5. Conclusion

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In this study, Zn isotope fractionation was quantified during adsorption onto synthetic

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birnessite. Additionally, we constrain the fractionation mechanisms by coupling EXAFS

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spectroscopy, DFT geometry optimisation, and SCM modeling. The results suggest that the

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adsorbed Zn exhibits a preferential enrichment of light isotopes, with  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$

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ranging from  $-0.46 \pm 0.04\text{‰}$  at low pH to  $-0.09 \pm 0.05\text{‰}$  at high pH. The change in Zn

815

isotopic fractionation as a function of pH is described using a two-site adsorption model

816 where two binding sites are included: interlayer vacancies and external edge sites.  
817 According to this model, two distinct isotopic fractionation factors of Zn are calculated:  
818  $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}} = -0.46 \pm 0.04\text{‰}$  for adsorption onto vacancy sites and  $\Delta^{66}\text{Zn}_{\text{adsorbed-}}$   
819  $\text{aqueous} = 0.52 \pm 0.04\text{‰}$  for binding to edge sites. The site-specific Zn isotopic fractionation  
820 can be explained by the local structures of Zn at molecular-level. Zn adsorbs to the vacancy  
821 sites of birnessite in TCS octahedral coordination at low pH (average bond length:  $\sim 2.09$   
822  $\text{\AA}$ ) with the enrichment of light Zn on birnessite. In contrast, Zn begins to form a DCS  
823 tetrahedral complex at higher pH values on edge sites at the birnessite surface with a  
824 calculated fractionation factor ( $\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$ ) of  $0.52 \pm 0.04\text{‰}$ . We assert that  
825 differences in bonding structures result in the differences in isotopic fractionation factors  
826 for binding to different reactive sites of the birnessite surface, as TCS octahedral Zn  
827 formation on vacancy sites with a shorter bond length relative to DCS tetrahedral Zn on  
828 edge sites is associated with a lower vibrational frequency and thus incorporates lighter  
829 isotopes.

830 Site-specific isotopic fractionation is strongly dependent on solution pH, suggesting  
831 the importance of isotopic fractionation of Zn caused by adsorption on birnessite in natural  
832 environments with varying pH values. Because birnessite minerals are highly reactive and  
833 control the concentrations and speciation of various trace metals, notably Zn, in both  
834 terrestrial and marine environments (Post, 1999; Manceau et al., 2002; Koschinsky and  
835 Hein, 2003), the chemical conditions of birnessite-rich environments should be carefully

836 considered in any isotopic studies of biogeochemical cycling. Moreover, the site-specific  
837 reactivity of birnessite provided by Zn isotopes in this study offers new insights into the  
838 links between adsorption and fractionation, which can be useful for predicting the  
839 fractionation behavior of other similar metals.

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854 for providing Zn reference EXAFS spectra on birnessite.

855

856 **Appendix A. Supplementary Material**

857 The supplementary material for this article includes supplementary figures and a file  
858 of Research Data. Supplementary figures contains speciation of aqueous Zn (Figure S1),  
859 plots of the correlation between modeled and measured Zn adsorbed on birnessite (Figure  
860 S2), Equilibrium or Rayleigh fractionation model (Figure S3). The file of Research Data  
861 includes the raw data of Fig. 6 and 11.

862

863 **References**

864

865 Appelo, C. A. J. and Postma, D. (1999) A consistent model for surface complexation on  
866 birnessite ( $\delta$ -MnO<sub>2</sub>) and its application to a column experiment. *Geochim.*  
867 *Cosmochim. Acta* **63**, 3039–3048.

868 Aranda S., Borrok D. M., Wanty R. B. and Balistrieri, L. S. (2012) Zinc isotope  
869 investigation of surface and pore waters in a mountain watershed impacted by acid  
870 rock drainage. *Sci. Total Environ.* **420**, 202-213.

871 Atkins A. L., Shaw S. and Peacock C. L. (2014) Nucleation and growth of todorokite from  
872 birnessite: implications for tracemetal cycling in marine sediments. *Geochim.*  
873 *Cosmochim. Acta* **144**, 109–125.

874 Atkins, A.L., Shaw, S. and Peacock, C.L. (2016). Release of Ni from birnessite during  
875 transformation of birnessite to todorokite: implications for Ni cycling in marine

876 sediments. *Geochim. Cosmochim. Acta* **189**, 158–183.

877 Balistrieri L. S., Borrok D. M., Wanty R. B. and Ridley W. I. (2008) Fractionation of Cu  
878 and Zn isotopes during adsorption onto amorphous Fe(III) oxyhydroxide:  
879 experimental mixing of acid rock drainage and ambient river water. *Geochim.*  
880 *Cosmochim. Acta* **72**, 311–328.

881 Bargar J. R., Fuller C. C., Marcus M. A., Brearley A. J., Perez Dela Rosa M., Webb S. M.  
882 and Caldwell W. A. (2009) Structural characterization of terrestrial microbial Mn  
883 oxides from Pinal Creek, AZ. *Geochim. Cosmochim. Acta* **73**, 889–910.

884 Bi X. Y., Li Z. G., Wang S. X., Zhang L., Xu R., Liu J. L., Yang H. M. and Guo M. Z.  
885 (2017) Lead isotopic compositions of selected coals, Pb/Zn ores and fuels in China  
886 and the application for source tracing. *Environ. Sci. Technol.* **51**, 13502–  
887 13508. Bigeleisen J. and Mayer M. G. (1947) Calculation of equilibrium constants for  
888 isotopic exchange reactions. *J. Chem. Phys.* **15**, 261–267.

889 Bigeleisen J. and Mayer M. G. (1947) Calculation of equilibrium constants for isotopic  
890 exchange reactions. *J. Chem. Phys.* **15**, 261–267.

891 Bodei S., Manceau A., Geoffroy N., Baronnet A. and Buatier M. (2007) Formation of  
892 todorokite from vernadite in Ni-rich hemipelagic sediments. *Geochim. Cosmochim.*  
893 *Acta* **71**, 5698–5716.

894 Bryan A. L., Dong S., Wilkes E. B. and Wasylenki L. E. (2015) Zinc isotope fractionation  
895 during adsorption onto Mn oxyhydroxide at low and high ionic strength. *Geochim.*

896 *Cosmochim. Acta* **157**, 182–197.

897 Chen, J., Gaillardet, J. and Louvat, P. (2008) Zinc isotopes in the Seine river waters, France:  
898 A probe of anthropogenic contamination. *Environ. Sci. Technol.* **42**, 6494–6501.

899 Christensen A. N. (1969) The crystal structure of c-Zn(OH)<sub>2</sub>. *Acta Chem. Scand.* **23**, 2016–  
900 2020.

901 Conway T. M., Rosenberg A. D., Adkins J. F. and John S. G. (2013) A new method for  
902 precise determination of iron, zinc and cadmium stable isotope ratios in seawater by  
903 double-spike mass spectrometry. *Anal. Chim. Acta* **793**, 44–52.

904 Dong S. and Wasylenki L. E. (2016) Zinc isotope fractionation during adsorption to calcite  
905 at high and low ionic strength. *Chem. Geol.* **447**, 70–78.

906 Drits V. A., Lanson B., Bougerol-Chaillout C., Gorshkov A. I. and Manceau A. (2002)  
907 Structure of heavy-metal sorbed birnessite: Part 2. Results from electron diffraction.  
908 *Am. Mineral.* **87**, 1646–1661.

909 Drits V. A., Silvester E., Gorshkov A. I. and Manceau A. (1997) Structure of synthetic  
910 monoclinic Na-rich birnessite and hexagonal birnessite; I, Results from X-ray  
911 diffraction and selected area electron diffraction. *Am. Mineral.* **82**, 946–961.

912 Ducher M., Pietrucci F., Balan E., Ferlat G., Paulatto L. and Blanchard M. (2017) van der  
913 Waals contribution to the relative stability of aqueous Zn<sup>2+</sup> coordination states. *J.*  
914 *Chem. Theory Comput.* **13**, 3340–3347.

915 Feng X. H., Zhai L. M., Tan W. F., Liu F. and He J. Z. (2007) Adsorption and redox

916 reactions of heavy metals on synthesized Mn oxide minerals. *Environ. Pollut.* **147**,  
917 366–373.

918 Feng X., Zhu M., Ginder-Vogel M., Ni C., Parikh S. and Sparks D. (2010) Formation of  
919 nanocrystalline todorokite from biogenic Mn oxides. *Geochim. Cosmochim. Acta* **74**,  
920 3232–3245.

921 Fujii T., Moynier F., Blichert-Toft J. and Albarede F. (2014) Density functional theory  
922 estimation of isotope fractionation of Fe, Ni, Cu, and Zn among species relevant to  
923 geochemical and biological environments. *Geochim. Cosmochim. Acta* **140**, 553–576

924 Fujii T., Moynier F., Telouk P. and Abe M. (2010) Experimental and theoretical  
925 investigation of isotope fractionation of zinc between aqua, chloro, and macrocyclic  
926 complexes. *J. Phys. Chem.* **114**, 2543–2552.

927 Gou W., Li W., Ji J. and Li W. (2018) Zinc isotope fractionation during sorption onto Al  
928 oxides: Atomic level understanding from EXAFS. *Environ. Sci. Technol.* **52**, 9087-  
929 9096.

930 Gou W., Li W., Siebecker M., Zhu M., Li L. and Sparks D. L. (2022) Coupling molecular-  
931 scale spectroscopy with stable isotope analyses to investigate the effect of Si on the  
932 mechanisms of Zn-Al LDH formation on Al oxide. *Environ. Sci. Technol.* (just  
933 accepted).

934 Gall, L., Williams, H. M., Siebert, C., Halliday, A. N., Herrington, R. J. and Hein, J. R.  
935 (2013) Nickel isotopic compositions of ferromanganese crusts and the constancy of

936 deep ocean inputs and continental weathering effects over the Cenozoic. *Earth Planet.*  
937 *Sci. Lett.* **375**, 148–155.

938 Ghose S. (1964) The crystal structure of hydrozincite,  $Zn_5(OH)_6(CO_3)_2$ . *Acta Cryst.* **17**,  
939 1051–1057.

940 Grangeon, S., Fernandez-Martinez, A., Claret, F., Marty, N., Tournassat, C., Warmont, F.  
941 and Gloter, A. (2017) In situ determination of the kinetics and mechanisms of nickel  
942 adsorption by nanocrystalline vernadite. *Chem. Geol.* **459**, 24–31.

943 Grimme, S. (2006). Semiempirical GGA-type density functional constructed with a long-  
944 range dispersion correction. *J. Comput. Chem.* **27**, 1787-1799.

945 Gueguen B., Sorensen J. V., Lalonde S. V., Pena J. T., Brandy M. and Rouxel O. (2018)  
946 Variable Ni isotope fractionation between Fe-oxyhydroxides and implications for the  
947 use of Ni isotopes as geochemical tracers. *Chem. Geol.* **481**, 38–52.

948 Guinoiseau D., Gelabert A., Moureau J., Louvat P. and Benedetti M. F. (2016) Zn isotope  
949 fractionation during sorption onto kaolinite. *Environ. Sci. Technol.* **50**, 1844–1852.

950 He J., Zhang L., Jin S., Zhu Y. and Liu F. (2008) Bacterial communities inside and  
951 surrounding soil iron-manganese nodules. *Geomicrobiol. J.* **25**, 14–24.

952 Hein J. R., Koschinsky A., Halbach P., Manheim F. T., Bau M., Kang J. K. and Lubick N.  
953 (1997) Iron and manganese oxide mineralization in the Pacific. In *Manganese*  
954 *Mineralization: Geochemistry and Mineralogy of Terrestrial and Marine Deposits*  
955 (eds. K. Nicholson, J. R. Hein, B. Buhn and S. Dasgupta). Geological Society Special



956 Publication No. 119, pp. 123–138.

957 Hein, J. R., Koschinsky, A. and Kuhn, T. (2020). Deep-ocean polymetallic nodules as a  
958 resource for critical materials. *Nature Reviews Earth & Environment*, **1**, 158-169.

959 Hinkle, M. A. G., Dye, K. G. and Catalano, J. G. (2017) Impact of Mn(II)-manganese oxide  
960 reactions on Ni and Zn speciation. *Environ. Sci. Technol.* **51**, 3187–3196.

961 Hochella, Jr., M. F., Kasama T., Putnis A., Putnis C. V. and Moore J. N. (2005)  
962 Environmentally important, poorly crystalline Fe/Mn hydrous oxides: ferrihydrite and  
963 a possibly new vernadite-like mineral from the Clark Fork River Superfund Complex.  
964 *Am. Mineral.* **90**, 718–724.

965 Jensen W. B. (2003) The place of zinc, cadmium, and mercury in the periodic table. *J.*  
966 *Chem. Educ.* **80**, 952–961.

967 Juillot F., Marechal C., Morin G., Jouvin D., Cacaly S., Telouk P., Benedetti M. F.,  
968 Ildefonse P., Sutton S., Guyot F. and Brown, G. E. (2011) Contrasting isotopic  
969 signatures between anthropogenic and geogenic Zn and evidence for postdepositional  
970 fractionation processes in smelter-impacted soils from Northern France. *Geochim.*  
971 *Cosmochim. Acta* **75**, 2295–2308.

972 Juillot F., Marechal C., Ponthieu M., Cacaly S., Morin G., Benedetti M., Hazemann J. L.,  
973 Proux O. and Guyot F. (2008) Zn isotopic fractionation caused by sorption on goethite  
974 and 2-Lines ferrihydrite. *Geochim. Cosmochim. Acta* **72**, 4886–4900.

975 Keizer, M.G. and van Riemsdijk, W.H. (2009) ECOSAT: A Computer Program for the

976 Calculation of Speciation and Transport in Soil-water Systems. Wageningen  
977 University, The Netherlands.

978 Komárek, M., Koretsky, C. M., Stephen, K. J., Alessi, D. S. and Chrastný, V. (2015)  
979 Competitive adsorption of Cd (II), Cr (VI), and Pb (II) onto nanomaghemite: a  
980 spectroscopic and modeling approach. *Environ. Sci. Technol.* **49**, 12851–12859.

981 Koschinsky A. and Hein J. R. (2003) Uptake of elements from seawater by ferromanganese  
982 crusts: solid-phase associations and seawater speciation. *Mar. Geol.* **198**, 331–351.

983 Kwon K. D., Refson K. and Sposito G. (2009) Zinc surface complexes on birnessite: a  
984 density functional theory study. *Geochim. Cosmochim. Acta* **73**, 1273–1284

985 Kwon K. D., Refson K. and Sposito G. (2010) Surface complexation of Pb(II) by hexagonal  
986 birnessite nanoparticles. *Geochim. Cosmochim. Acta* **74**, 6731–6740.

987 Kwon K. D. and Sposito G. (2015) Mechanistic understanding of metal sorption by  
988 phylломanganates through density functional theory, *Advances in the environmental*  
989 *biogeochemistry of manganese oxides. Am. Chem. Soc.* **51–64**.

990 Lanson B., Drits V. A., Gaillot A.-C., Silvester E., Plancon A. and Manceau A. (2002)  
991 Structure of heavy-metal sorbed birnessite: Part 1. Results from X-ray diffraction. *Am.*  
992 *Mineral.* **87**, 1631–1645.

993 Lanson B., Marcus M. A., Fakra S., Panfili F., Geoffroy N. and Manceau A. (2008)  
994 Formation of Zn–Ca phylломanganate nanoparticles in grass roots. *Geochim.*  
995 *Cosmochim. Acta* **72**, 2478–2490.

996 Lefkowitz, J. P. and Elzinga, E. J. (2017) Structural alteration of hexagonal birnessite by  
997 aqueous Mn(II): impacts on Ni(II) sorption. *Chem. Geol.* **466**, 524–532.

998 Li, Y., Zhao, X., Wu, J. and Gu, X. (2020). Surface complexation modeling of divalent  
999 metal cation adsorption on birnessite. *Chem. Geol.* **551**, 119774.

1000 Little S. H., Sherman D. M., Vance D. and Hein J. R. (2014a) Molecular controls on Cu  
1001 and Zn isotopic fractionation in Fe–Mn crusts. *Earth Planet. Sci. Lett.* **396**, 213–222.

1002 Little S. H., Vance D., McManus J. and Severmann S. (2016) Key role of continental  
1003 margin sediments in the oceanic mass balance of Zn and Zn isotopes. *Geology* **44**,  
1004 207–210.

1005 Little S. H., Vance D., Walker-Brown C. and Landing W. (2014b) The oceanic mass balance  
1006 of copper and zinc isotopes, investigated by analysis of their inputs, and outputs to  
1007 ferromanganese oxide sediments. *Geochim. Cosmochim. Acta* **125**, 673–693.

1008 Little, S. H., Archer, C., McManus, J., Najorka, J., Wegorzewski, A. V. and Vance, D.  
1009 (2020). Toward balancing the oceanic Ni budget. *Earth Planet. Sci. Lett.* **547**, 116461.

1010 Liu, S. A., Liu, P. P., Lv, Y., Wang, Z. Z. and Dai, J. G. (2019). Cu and Zn isotope  
1011 fractionation during oceanic alteration: implications for Oceanic Cu and Zn cycles.  
1012 *Geochim. Cosmochim. Acta* **257**, 191-205.

1013 Manceau A., Lanson B. and Drits V. A. (2002) Structure of heavy metal sorbed birnessite:  
1014 part III. The results from powder and polarized extended X-ray absorption fine  
1015 structure spectroscopy. *Geochim. Cosmochim. Acta* **66**, 2639–2663.

1016 Manceau A., Lanson M. and Geoffroy N. (2007) Natural speciation of Ni, Zn, Ba and As  
1017 in ferromanganese coatings on quartz using X-ray fluorescence, absorption and  
1018 diffraction. *Geochim. Cosmochim. Acta* **71**, 95–128.

1019 Marcus M. A., Manceau A. and Kersten M. (2004) Mn, Fe, Zn and As speciation in a fast-  
1020 growing ferromanganese marine nodule. *Geochim. Cosmochim. Acta* **68**, 3125–3136.

1021 Marechal C. N., Telouk P. and Albarede F. (1999) Precise analysis of copper and zinc  
1022 isotopic compositions by plasma-source mass spectrometry. *Chem. Geol.* **156**, 251–  
1023 273.

1024 Marechal C., Nicolas E., Douchet C. and Albarede F. (2000) Abundance of zinc isotopes  
1025 as a marine biogeochemical tracer. *Geochem. Geophys. Geosyst.* **1**, 1015.

1026 McKenzie R. M., (1971) The synthesis of birnessite, cryptomelane, and some other oxides  
1027 and hydroxides of manganese. *Mineral. Mag.* **38**, 493–503

1028 McNellis, E. R., Meyer, J. and Reuter, K. (2009). Azobenzene at coinage metal surfaces:  
1029 Role of dispersive van der Waals interactions. *Phys. Rev. B*, **80**, 205414.

1030 Moon E. M. and Peacock C. L. (2012) Adsorption of Cu(II) to ferrihydrite and ferrihydrite-  
1031 bacteria composites: Importance of the carboxyl group for Cu mobility in natural  
1032 environments. *Geochim. Cosmochim. Acta* **92**, 203–219.

1033 Moore, C.M., et al., 2013. Processes and patterns of oceanic nutrient limitation. *Nat. Geosci.*  
1034 **6**, 701–710.

1035 Nachtegaal M. and Sparks D. L. (2004) Effect of iron oxide coatings on zinc sorption

1036 mechanisms at the clay-mineral/water interface. *J. Colloid Interface Sci.* **276**, 13–23.

1037 Nakada, R., Tanaka, M., Tanimizu, M. and Takahashi, Y. (2017). Aqueous speciation is  
1038 likely to control the stable isotopic fractionation of cerium at varying pH. *Geochim.*  
1039 *Cosmochim. Acta* **218**, 273-290.

1040 Nelson, J.; Wasylenki, L.; Bargar, J. R.; Brown, G. E. and Maher, K. (2017) Effects of  
1041 surface disorder and surface coverage on isotopic fractionation during Zn(II)  
1042 adsorption onto quartz and amorphous silica surfaces. *Geochim. Cosmochim. Acta*  
1043 **215**, 354–376.

1044 Pavlov, M., Siegbahn, P. E. and Sandström, M. (1998). Hydration of beryllium, magnesium,  
1045 calcium, and zinc ions using density functional theory. *J. Phys. Chem. A* **102**, 219-228.

1046 Peacock C. L. (2009) Physiochemical controls on the crystalchemistry of Ni in birnessite:  
1047 genetic implications for ferromanganese precipitates. *Geochim. Cosmochim. Acta* **73**,  
1048 3568–3578

1049 Peacock C. L. and Moon E. M. (2012) Oxidative scavenging of thallium by birnessite:  
1050 explanation for thallium enrichment and stable isotope fractionation in marine  
1051 ferromanganese precipitates. *Geochim. Cosmochim. Acta* **84**, 297–313.

1052 Peacock C. L. and Sherman D. M. (2007a) The crystal chemistry of Ni in marine  
1053 ferromanganese crusts and nodules. *Am. Mineral.* **92**, 1087–1097.

1054 Peacock C. L. and Sherman D. M. (2007b) Sorption of Ni by birnessite: equilibrium  
1055 controls on Ni in seawater. *Chem. Geol.* **238**, 94–106.

1056 Perdew J. P., Burke K. and Ernzerhof M. (1996) Generalized gradient approximation made  
1057 simple. *Phys. Rev. Lett.* **77**, 3865–3868.

1058 Pfrommer B. G., Cote M., Louie S. G. and Cohen M. L. (1997) Relaxation of crystals with  
1059 the quasi-Newton method. *J. Comput. Phys.* **131**, 233–240.

1060 Pokrovsky O. S., Viers J. and Freydier R. (2005) Zinc stable isotope fractionation during  
1061 its adsorption on oxides and hydroxides. *J. Colloid Interface Sci.* **291**, 192–200.

1062 Post J. E. (1999) Manganese oxide minerals: crystal structures and economic and  
1063 environmental significance. *Proc. Natl. Acad. Sci.* **96**, 3447–3454.

1064 Post J. E. and Appleman D. E. (1988) Chalcophanite,  $ZnMn_3O_7 \cdot 3H_2O$ : New crystal-  
1065 structure determinations. *Am. Mineral.* **73**, 1401–1404.

1066 Post J. E., Heaney P. J. and Hanson J. (2003) Real-time synchrotron X-ray powder  
1067 diffraction studies of the structure and dehydration of todorokite. *Am. Mineral.* **88**,  
1068 142–150.

1069 Pye C. C., Corbeil C. R. and Rudolph W. W. (2006) An ab initio investigation of zinc chloro  
1070 complexes. *Phys. Chem. Chem. Phys.* **8**, 5428–5436.

1071 Ravel, B. Kelly, S.D., 2007. The difficult chore of measuring coordination by EXAFS. In:  
1072 *X-ray Absorption Fine Structure–XAFS* **13**, pp. 150–152.

1073 Roberts D. R., Ford R. G., and Sparks D. L. (2003) Kinetics and mechanisms of Zn  
1074 complexation on metal oxides using EXAFS spectroscopy. *J. Colloid Interface Sci.*  
1075 **263**, 364–376.

1076 Sandstead, H.H., 2014. Zinc. In: *Handbook on the Toxicology of Metals Elsevier*, pp. 1369–  
1077 1386.

1078 Schauble E. A. (2004) Applying stable isotope fractionation theory to new systems. *Rev.*  
1079 *Mineral. Geochem.* **55**, 65–111.

1080 Scheinost A. C., Abend S., Pandya K. I. and Sparks D. L. (2001) Kinetic controls on Cu  
1081 and Pb sorption by ferrihydrite. *Environ. Sci. Technol.* **35**, 1090–1096.

1082 Sherman D. M. and Peacock C. L. (2010) Surface complexation of Cu on birnessite ( $\delta$ -  
1083 MnO<sub>2</sub>): controls on Cu in the deep ocean. *Geochim. Cosmochim. Acta* **74**, 6721–6730.

1084 Simanova A. A., Kwon K. D., Bone S. E., Bargar J. R., Refson K., Sposito G. and Pena J.  
1085 (2015) Probing the sorption reactivity of the edge surfaces in birnessite nanoparticles  
1086 using nickel (II). *Geochim. Cosmochim. Acta* **164**, 191–204.

1087 Sinoir, M., Butler, E.C., Bowie, A.R., Mongin, M., Nesterenko, P.N., and Hassler, C.S.  
1088 (2012) Zinc marine biogeochemistry in seawater: A review: *Marine & Freshwater*  
1089 *Research*, v. 63, p. 644–657, doi:10.1071/MF11286

1090 Skierszkan E. K., Mayer K. U., Weis D. and Beckie R. D. (2016). Molybdenum and zinc  
1091 stable isotope variation in mining waste rock drainage and waste rock at the Antamina  
1092 mine, Peru. *Sci. Total Environ.* **550**, 103-113.

1093 Sorensen, J.V., Gueguen, B., Stewart, B.D., Peña, J., Rouxel, O., Toner, B.M., 2020. Large  
1094 nickel isotope fractionation caused by surface complexation reactions with hexagonal  
1095 birnessite. *Chem. Geol.* **537**, 119481.

1096 Spinks, S. C., and Uvarova, Y. (2019). Fractionation of Zn isotopes in terrestrial  
1097 ferromanganese crusts and implications for tracing isotopically heterogeneous metal  
1098 sources. *Chem. Geol.* **529**, 119-314.

1099 Sverjensky, D. A. (2005) Prediction of surface charge on oxides in salt solutions: Revisions  
1100 for 1:1 (M+L-) electrolytes. *Geochim. Cosmochim. Acta* **69**, 225–257.

1101 Toner B., Manceau A., Webb S. M. and Sposito G. (2006) Zinc sorption to biogenic  
1102 hexagonal-birnessite particles within a hydrated bacterial biofilm. *Geochim.*  
1103 *Cosmochim. Acta* **70**, 27–43.

1104 Vance D., Little S. H., Archer C., Cameron V., Andersen M. B., Rijkenberg M. J. and Lyons  
1105 T. W. (2016) The oceanic budgets of nickel and zinc isotopes: the importance of  
1106 sulfidic environments as illustrated by the Black Sea. *Phil. Trans. R. Soc. A* **374**,  
1107 20150294.

1108 Vanderbilt D. (1990) Soft self-consistent pseudopotentials in a generalized eigenvalue  
1109 formalism. *Phys. Rev.* **B** 41, 7892–7895.

1110 Villalobos M., Bargar J. R. and Sposito G. (2005) Mechanisms of Pb(II) sorption on a  
1111 biogenic manganese oxide. *Environ. Sci. Technol.* **39**, 569–576.

1112 Wang Y., Benkaddour S., Marafatto F. F. and Pena J. (2018) Diffusion- and pH-dependent  
1113 reactivity of layer-type MnO<sub>2</sub>: Reactions at particle edges versus vacancy sites.  
1114 *Environ. Sci. Technol.* **52**, 3476–3485.

1115 Wang Z., Kwon K. D., Peacock C., Mo X., Gou W., Feng X. and Li W. (2022) Zn stable



1116 isotope fractionation during adsorption onto todorokite: A molecular perspective from  
1117 X-ray absorption spectroscopy and density functional theory. *Geochim. Cosmochim.*  
1118 *Acta* **327**, 116–136.

1119 Wasylenki L. E., Rolfe B. A., Weeks C. L., Spiro T. G. and Anbar A. D. (2008)  
1120 Experimental investigation of the effects of temperature and ionic strength on Mo  
1121 isotope fractionation during adsorption to manganese oxides. *Geochim. Cosmochim.*  
1122 *Acta* **72**, 5997–6005.

1123 Wasylenki L. E., Swihart J. W. and Romaniello S. J. (2014) Cadmium isotope fractionation  
1124 during adsorption to Mn oxyhydroxide at low and high ionic strength. *Geochim.*  
1125 *Cosmochim. Acta* **140**, 212–226.

1126 Waychunas G. A., Fuller C. C. and Davis J. A. (2002) Surface complexation and precipitate  
1127 geometry for aqueous Zn(II) sorption on ferrihydrite I: X-ray absorption extended fine  
1128 structure spectroscopy analysis. *Geochim. Cosmochim. Acta* **66**, 1119–1137.

1129 Wiederhold, J. G. (2015) Metal stable isotope signatures as tracers in environmental  
1130 geochemistry. *Environ. Sci. Technol.* **49**, 2606-2624.

1131 Yan, X., Zhu, M., Li, W., Peacock, C. L., Ma, J., Wen, H., Liu, F., (2021) Cadmium isotope  
1132 fractionation during adsorption and substitution with iron (oxyhydr) oxides. *Environ.*  
1133 *Sci. Technol.* **55**, 11601–11611.

1134 Yin H., Wang X., Qin Z., Ginder-Vogel M., Zhang S., Jiang S., Liu F., Li S., Zhang J. and  
1135 Wang Y. (2018) Coordination geometry of Zn<sup>2+</sup> on hexagonal turbostratic birnessites

1136 with different Mn average oxidation states and its stability under acid dissolution. J.  
1137 *Environ. Sci. (China)* **65**, 282–292.

1138 Young E. D., Tonui E., Manning C. E., Schauble E. and Macris C. A. (2009) Spinel–olivine  
1139 magnesium isotope thermometry in the mantle and implications for the Mg isotopic  
1140 composition of Earth. *Earth Planet. Sci. Lett.* **288**, 524–533.

1141 Zhao, W., Tan, W., Wang, M., Xiong, J., Liu, F., Weng, L. and Koopal, L.K. (2018) CD-  
1142 MUSICEDL modeling of Pb<sup>2+</sup> adsorption on birnessites: role of vacant and edge sites.  
1143 *Environ. Sci. Technol.* **52**, 10522–10531.

1144 Zhao, Y., Vance, D., Abouchami, W. and de Baar, H. J. W. (2014) Biogeochemical cycling  
1145 of zinc and its isotopes in the Southern Ocean. *Geochim. Cosmochim. Acta* **125**, 653–  
1146 672.

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1152 **Table 1.** Isotopic data of Zn in the adsorption experiments.

Experiments	Time (h)	[Zn] <sub>initial</sub> (mM)	Adsorbed Zn(%)	Zn/Mn (mol:mol)	pH	$\delta^{66}\text{Zn}_{\text{aqueous}}$	2SD	$\delta^{66}\text{Zn}_{\text{adsorbed}}$	2SD	$\Delta^{66}\text{Zn}_{\text{adsorbed-aqueous}}$	2SD	Mass balance
Birnessite adsorption kinetics	0.5	0.2	18.1	0.061	6	0.62	0.06	-0.61	0.03	-1.23	0.07	0.06
	1	0.2	24.8	0.084	6	0.70	0.05	-0.53	0.03	-1.22	0.06	0.07
	1.5	0.2	24.7	0.084	6	0.69	0.03	-0.48	0.05	-1.18	0.06	0.06
	3	0.2	27.7	0.094	6	0.68	0.04	-0.41	0.01	-1.09	0.04	0.05
	4	0.2	26.0	0.088	6	0.67	0.02	-0.44	0.03	-1.11	0.04	0.06
	6	0.2	28.8	0.097	6	0.65	0.03	-0.36	0.07	-1.02	0.08	0.04
	12	0.2	28.9	0.098	6	0.62	0.02	-0.28	0.03	-0.90	0.03	0.05
	24	0.2	33.6	0.114	6	0.61	0.01	-0.21	0.01	-0.81	0.04	0.03
	36	0.2	32.7	0.111	6	0.59	0.05	-0.14	0.04	-0.73	0.05	0.03
	48	0.2	32.0	0.108	6	0.59	0.04	-0.16	0.02	-0.74	0.05	0.02
	120	0.2	42.1	0.129	6	0.44	0.04	0.08	0.04	-0.36	0.06	0.01
	264	0.2	47.0	0.159	6	0.47	0.01	0.10	0.04	-0.36	0.08	0.01
432	0.2	49.8	0.169	6	0.42	0.02	0.14	0.03	-0.30	0.05	0.01	
Adsorption edges 0.2 mM	120	0.2	40.0	0.105	3	0.43	0.02	-0.06	0.01	-0.49	0.02	0.00
	120	0.2	36.1	0.122	4	0.44	0.02	0.03	0.00	-0.41	0.02	0.01
	120	0.2	50.3	0.170	5	0.49	0.02	0.06	0.03	-0.43	0.03	-0.01
	120	0.2	55.4	0.187	6	0.47	0.03	0.22	0.04	-0.25	0.05	0.05
	120	0.2	75.6	0.256	7	0.34	0.03	0.24	0.01	-0.10	0.04	-0.01
Adsorption isotherms pH 3	120	0.2	93.7	0.317	8	0.35	0.01	0.29	0.04	-0.06	0.04	0.00
	120	0.05	43.9	0.037	3	0.46	0.00	0.05	0.04	-0.41	0.04	0.01
	120	0.07	34.9	0.041	3	0.43	0.03	-0.02	0.04	-0.45	0.05	-0.01
	120	0.1	45.4	0.077	3	0.54	0.02	0.00	0.05	-0.54	0.05	0.01
	120	0.2	40.1	0.105	3	0.43	0.02	-0.06	0.01	-0.49	0.02	0.00
	120	0.3	23.7	0.121	3	0.37	0.02	-0.13	0.02	-0.50	0.03	-0.03

1153 *Note.*  $^{66}\text{Zn}_{\text{stock}}$  was  $0.28 \pm 0.06\text{‰}$  (2SD, n = 6). Each isotopic data was the average value calculated from three measurements. The 2 SD

1154 values were the internal precisions of replicates. Mass balance offset =  $(\delta^{66}\text{Zn}_{\text{aqueous}} (1-\text{adsorbed \%}) + \delta^{66}\text{Zn}_{\text{adsorbed}} \text{adsorbed \%}) - \delta^{66}\text{Zn}_{\text{stock}}$ .

1155 **Table 2.** Fitting results of the EXAFS spectra for Zn-birnessite samples and model compounds  
 1156 ( $S_0^2 = 0.97$ )

Samples	Shell no.	Path	CN	R (Å)	$\sigma^2(\text{Å}^2)$	R-factor	$\Delta E_0$ (eV)
Zinc hydroxide	1	Zn–O	3.7	1.98	0.006	0.0009	5.16
	2	Zn–Zn	10.4	3.23	0.009		
	3	Zn–Zn	3.8	4.62	0.004		
Hydrozincite	1	Zn–O	4.5	1.98	0.009	0.0082	–5.98
	2	Zn–Zn	1.4	3.12	0.012		
	2	Zn–Zn	2.2	3.52	0.006		
Zn tetrahedra	1	Zn–O	3.3	1.97			
	2	Zn–Mn	2.9	3.35			
	2	Zn–Mn	2.0	3.50			
Chalcophanite	1	Zn–O	3	2.07			
	1	Zn–O	3	2.14			
	2	Zn–Mn	3	3.49			
	2	Zn–Mn	3	3.50			
pH 8_0.2 mM	1	Zn–O	2.1	2.01	0.005	0.0007	7.24
	1	Zn–O	2.3	2.17	0.015		
	2	Zn–Mn	3.5	3.49	0.015		
pH 6_0.2 mM	1	Zn–O	2.9	2.04	0.008	0.0005	5.86
	1	Zn–O	2.7	2.18	0.020		
pH 3_0.2 mM	2	Zn–Mn	7.1	3.49	0.016	0.0008	5.97
	1	Zn–O	2.5	2.03	0.006		
	1	Zn–O	2.5	2.15	0.012		
pH 3_0.1 mM	2	Zn–Mn	6.3	3.50	0.011	0.0041	3.02
	1	Zn–O	3.0	2.04	0.004		
	1	Zn–O	3.0	2.17	0.010		
pH 3_0.05 mM	2	Zn–Mn	7.5	3.52	0.012	0.0011	5.72
	1	Zn–O	2.6	2.03	0.001		
	1	Zn–O	2.6	2.18	0.003		
	2	Zn–Mn	7.7	3.52	0.012		

The parameters of Zn tetrahedra are from EXAFS fits for Zn adsorbed birnessite mainly in tetrahedral coordination (sample ZnBi8) (Manceau et al., 2002). Chalcophanite values are from XRD (Post and Appleman, 1988). Bir pH 3\_0.2 means Zn local structure on birnessite at pH 3 and initial Zn concentration of 0.2 mM.

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1160 **Table 3.** DFT-calculated structure parameters of Zn surface complexes formed in the birnessite nanodisk model

	TCS		DCS			DES		1161
	Zn·3H <sub>2</sub> O	Zn·1H <sub>2</sub> O	Zn·4H <sub>2</sub> O	Zn·4H <sub>2</sub> O <sup>a</sup>	Zn·2H <sub>2</sub> O	Zn·3H <sub>2</sub> O <sup>c</sup>	Zn·3H <sub>2</sub> O <sup>d</sup>	
<i>d</i> (Zn–O <sub>1st</sub> )	2.07	1.91	1.95	1.93	1.85	2.01	2.07	1164
	2.09	1.95	1.96	1.95	1.90	2.08	2.09	1165
	2.08	1.95	1.98	1.98	2.14	2.13	2.10	1166
	2.18	2.09	2.03	2.03	2.10	2.20	2.11	1167
	2.17					2.20	2.23	1168
	2.17					2.25	2.26	1169
	<2.13>	<1.98>	<1.98>	<1.97>	<2.00>	<2.15>	<2.16>	1170
C.N.	6	4	4 <sup>b</sup>	4 <sup>b</sup>	4	6	6	1171
<i>d</i> (Zn–Mn <sub>1st</sub> )	3.50	3.29	3.27	3.32	3.25	2.96	2.97	1172
	3.51	3.31	3.36	3.36	3.16	2.98	3.13	1173
	3.50	3.33						1174
	3.50	3.32						1175
	3.51	3.32						1176
	3.50	3.31						1177
	<3.50>	<3.31>	<3.32>	<3.34>	<3.21>	<2.97>	<3.05>	1178
C.N.	6	6	2	2	2	2	2	1179

1180 <sup>a</sup> Calculation with dispersion correction (DFT–D).

1181 <sup>b</sup> Initial octahedral arrangement always ended up with tetrahedral coordination.

1182 <sup>c</sup> Complex coordinated with terminal bridging oxygen.

1183 <sup>d</sup> Complex coordinated with bridging oxygen.

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1186 **Table 4.** Model parameters for protons and Zn adsorption to birnessite.

Protons complex on external surface	$\Delta z_0$	$\Delta z_1$	$\Delta z_2$	$\log K$
$\equiv\text{MnOH}^{-1/3} + \text{H}^+ = \equiv\text{MnOH}_2^{+2/3}$	1	0	0	4.6 <sup>a</sup>
$\equiv\text{MnOH}^{-1/3} + \text{Na}^+ = \equiv\text{MnOH}^{-1/3} \dots \text{Na}^+$	0	1	0	-0.6 <sup>a</sup>
$\equiv\text{MnOH}^{-1/3} + \text{H}^+ + \text{NO}_3^- = \equiv\text{MnOH}_2^{+2/3} \dots \text{NO}_3^-$	0	-1	0	3.92 <sup>a</sup>
Protons complex on internal surface	$\Delta z_0$			$\log K$
$\equiv\text{Mn}_2\text{O}^{-2/3} + \text{H}^+ = \equiv\text{Mn}_2\text{OH}^{+1/3}$	1			1.2 <sup>a</sup>
Zn complex on external surface	$\Delta z_0$	$\Delta z_1$	$\Delta z_2$	$\log K$
$2 \equiv\text{MnOH}^{-1/3} + \text{Zn}^{2+} + \text{H}_2\text{O} = (\equiv\text{MnOH})_2\text{ZnOH}^{+1/3} + \text{H}^+$	0.51	0.49	0	-3.54
Zn complex on internal surface	$\Delta z_0$			$\log K$
$3 \equiv\text{Mn}_2\text{O}^{-2/3} + \text{Zn}^{2+} = (\equiv\text{Mn}_2\text{O})_3\text{Zn}^0$	2			-0.36
External surface: $\text{SSA}^b = 176 \text{ m}^2 \text{ g}^{-1}$ ; site density <sup>c</sup> = 14.9 sites $\text{nm}^{-2}$ ; $C_1 = C_2 = 2.5 \text{ F m}^{-2}$ .				
Interlayer surface: $\text{SSA}^b = 729 \text{ m}^2 \text{ g}^{-1}$ ; site density <sup>c</sup> = 8.34 sites $\text{nm}^{-2}$ ; $C = 2.0 \text{ F m}^{-2}$ .				
<sup>a</sup> From <a href="#">Li et al. (2020)</a> . <sup>b</sup> The SSA of external and internal surfaces are assumed based on the calculation of the crystal structure of birnessite ( <a href="#">Zhao et al., 2018</a> ). <sup>c</sup> The site densities of birnessite are calculated according to Mn AOS and the crystal structure based on <a href="#">Zhao et al. (2018)</a> .				

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