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1 **Strengthening the evidence base for temperature-mediated** 2 **phenological asynchrony and its impacts**

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31 **Climate warming has caused the seasonal timing of many components of ecological**
32 **food chains to advance. In the context of trophic interactions the match-mismatch**
33 **hypothesis (MMH) postulates that differential shifts can lead to phenological**
34 **asynchrony with negative impacts for consumers. However, at present there has been**
35 **no consistent analysis of the links between temperature change, phenological**
36 **asynchrony, and individual-to-population level impacts across taxa, trophic levels and**
37 **biomes at a global scale. Here, we propose five criteria that all need to be met to**
38 **demonstrate that temperature-mediated trophic asynchrony poses a growing risk to**
39 **consumers. We conduct a literature review of 109 papers studying 129 taxa, and find**
40 **that all five criteria are assessed for only two taxa, with the majority of taxa only**
41 **having one or two criteria assessed. Crucially, nearly every study was conducted in**
42 **Europe or North America, and most studies were on terrestrial secondary consumers.**
43 **We thus lack a robust evidence base from which to draw general conclusions about**
44 **the risk that climate-mediated trophic asynchrony may pose to populations**
45 **worldwide.**

46
47 The shifting seasonal timing of key life history events, such as the budburst of trees,
48 emergence of insects or the migration and breeding times of vertebrates, is one of the three
49 universal ecological responses to climate warming^{1,2} alongside range shifts and reductions in
50 organismal body size^{3,4}. Such shifts in phenology have provided some of the earliest and
51 strongest evidence that rising temperatures have left a discernible imprint on the planet's
52 ecosystems⁵⁻⁸. For many consumer species, phenological events are timed to coincide with
53 peak abundance of a predictable food resource. However, the strength and direction of the
54 phenological response to temperature frequently differs among species occupying different
55 trophic levels, leading to asynchrony between resource and consumer (box 1). The
56 consequences of such asynchrony were first studied in the early 1900s in the context of
57 trophic interactions between fish larvae and their zooplankton resource. This generated the

58 classic match-mismatch hypothesis⁹. Fish larvae were found to spawn at a relatively fixed
59 date, but zooplankton phenology was more variable across years, causing annual variation
60 in asynchrony between consumer and resource. The degree of asynchrony (referred to as
61 mismatch by Cushing) was proposed to account for annual fluctuations in fish recruitment to
62 the population⁹. In recent years, the hypothesis that changing temperatures might increase
63 the frequency of costly trophic asynchrony between consumers and their resources has
64 been increasingly discussed¹⁰⁻¹³. The impact of asynchronous phenological interactions on
65 the fates of consumer species was identified as a key uncertainty in the fifth assessment
66 report of the IPCC¹⁴.

67 Phenological asynchrony and mismatch are often used interchangeably in the
68 ecological literature, but the meaning of the term “mismatch” is more ambiguous, as it is in
69 some cases used to imply only dissimilar responses of adjacent trophic levels¹³, and in other
70 cases implying negative impacts on the consumer¹⁵. In this paper we refer to “trophic
71 asynchrony” when the consumer demand does not coincide with the phenology of the
72 resource, and to the match-mismatch hypothesis (MMH) when asynchrony has negative
73 impacts on fitness or populations (box 1). We note that the MMH is normally conceptualised
74 from a unidirectional, bottom-up perspective (i.e., asynchrony leading to detrimental effects
75 on consumers), rather than potential top-down effects upon prey and resources¹⁶.

76 Asynchrony has been detected in many study systems^{7,17,18}, but to demonstrate
77 negative consequences of asynchrony on the consumer (i.e., the MMH), several conditions
78 need to be met. For trophic asynchrony to be identified as detrimental, the consumer must
79 depend on a short, seasonally-pulsed or ephemeral resource¹⁹⁻²⁵, and it should be
80 established whether asynchrony might be an adaptive baseline state²⁶⁻²⁸. Moreover, there
81 should be negative effects of asynchrony on consumer fitness²⁹⁻³⁴. Ultimately, asynchrony
82 becomes of conservation concern when it affects mean demographic parameters and leads
83 to population declines^{13,35-38}. Although components of the MMH and consequences for
84 population trends can be identified, these are based on very few and specific study systems.

85 We therefore lack a general overview of how often trophic asynchrony leads to population
86 declines.

Box 1: Glossary of terms widely used in the study of trophic asynchrony

Phenology: the study of cyclically recurring biological events, such as the seasonal timing of tree leafing, insect hatching, or animal migration and reproduction. In this work, we also use it to refer to the events themselves, as has become the norm in the literature.

Trophic level: the position that an organism occupies in the food chain. Primary consumers are herbivores (e.g., winter moth, caribou), and secondary consumers are omnivores or carnivores (e.g., great tit, herring)

Phenological sensitivity/response: the interannual variation in phenology that relates to interannual variation in a biotic or abiotic cue, for example an advance in breeding in response to temperature.

Phenological/trophic asynchrony: when the seasonal peak in consumer demand for a resource does not coincide with the seasonal peak in availability of that resource.

The match-mismatch hypothesis (MMH): poses that trophic asynchrony has negative consequences for consumer fitness or population size. This is also sometimes referred to as trophic mistiming.

87

88

89 **Five criteria for demonstrating risks of temperature-mediated asynchrony**

90 Here, based on ideas that have been widely discussed in the literature and which we outline
91 above, we propose five criteria that must all be met for temperature-mediated phenological
92 asynchrony to be both present, and causing population declines (Table 1): (1) the consumer
93 is highly reliant on a seasonally ephemeral resource; (2) the degree of trophic asynchrony
94 between consumer and resource phenology is increasing over the years (evidenced by time
95 series); (3) increasing trophic asynchrony is due to differing temperature responses of
96 consumer and resource; (4) trophic asynchrony impacts negatively on consumer fitness, and
97 (5) asynchrony impacts negatively on population growth³⁷. In Table 1 we identify some of the
98 methods that can be used to test each of these criteria. In the next section we summarise
99 the existing biological evidence for these criteria, with a particular focus on general insights
100 that have emerged from multi-species studies and formal meta-analyses on questions that

101 are pertinent to the study of the MMH. The five criteria can be seen as a best-practice
102 framework, but we realize that each study system poses unique challenges for studying
103 these criteria. We do not intend for their application to oversimplify the complex study of
104 phenology, nor do we claim that they cover everything that phenological studies need to
105 focus on.

106

107 *Evidence for phenological asynchrony*

108 Large-scale comparative analyses of phenological responses and formal meta-analyses
109 provide ample evidence that on average spring timings are advancing at mid-high latitudes,
110 and that species vary in their response to temperature^{8,17,39}. In two large multi-species
111 analyses based on phenological data from the UK, the phenology of secondary consumers
112 advanced less than primary producers and consumers over the years (criterion 2)¹⁷ and
113 secondary consumers have a lower phenological sensitivity to temperature (criterion 3)⁷. For
114 marine taxa, the magnitude of phenological advance varied among trophic groups, with
115 phytoplankton, zooplankton and bony fish all more responsive than seabirds^{40,41}. However,
116 while large-scale multi-species and multi-population studies provide valuable insights into
117 general trends and patterns of inter- and intra-specific variation in phenological responses,
118 they do not estimate *in situ* responses for specific trophic interactions, nor do they reveal
119 anything about fitness or population consequences. Of those studies that focus on trophic
120 interactions known to be important to the consumer for a short period (criterion 1) most focus
121 on a single interaction. Such studies have reported increasing asynchrony over the years
122 (criterion 2), for example in great tits, *Parus major*, and winter moth, *Operophtera*
123 *brumata*^{5,42}. However, a recent analysis of the phenological time series underpinning 27
124 species interactions (including but not limited to trophic interactions) found that whilst the
125 degree of asynchrony has changed over the years, the number of cases where asynchrony
126 had increased was roughly balanced by the number of cases where asynchrony had
127 decreased¹⁸. The same study also found that whilst phenology was responding to

128 temperature in the ecological systems considered, it was not possible to attribute temporal
129 trends in asynchrony to long-term increases in temperature (criterion 3).

130

131 *Potential consequences of trophic asynchrony*

132 The most prominent evidence for the MMH comes from intensively-studied wild systems,
133 such as that of the reliance of great tits on winter moth caterpillars to feed their young. For
134 these birds, asynchrony between the timing of peak nestling demand and peak caterpillar
135 biomass has negative consequences for individual fitness and annual mean fitness^{30,31}.
136 Meta-analyses of selection estimates in the wild report consistent selection pressures for
137 earlier phenology^{43,44}, but directional selection has not become stronger over time⁴⁴.
138 However, meta-analyses addressing selection on phenology have considered only absolute
139 timing, rather than the timing of a consumer relative to its resource¹⁰, so it is unclear from
140 these studies whether selection on consumers is being driven by asynchrony with resources.

141 Trophic asynchrony becomes a matter of conservation concern if it impacts
142 negatively on population size (criterion 5)³⁷. Two long-term studies of great tits found no
143 evidence of an effect of asynchrony on population size^{31,45}, whereas a study of several
144 populations of pied flycatchers *Ficedula hypoleuca* in the Netherlands reported stronger
145 population declines where caterpillar phenology was earlier³⁶. The only multi-species studies
146 on this theme that we are aware of are for birds. One of these studies found that migrant
147 passerines breeding in European forests had declined more than resident and marsh
148 inhabiting species, which the authors attributed to the MMH⁴⁶. The other study found that,
149 across 21 UK bird species, population declines were more pronounced for species that had
150 advanced their breeding phenology least and in species whose annual productivity was most
151 reduced by asynchrony with general insect and plant phenology.⁴⁷ However, support for the
152 MMH was weak and not reflected by declines in breeding success of those species.

153

154 Table 1. Criteria of evidence that climate change induced trophic asynchrony is increasing and
 155 deleterious for the consumer (the MMH), with a consideration of the data and methods that can be
 156 used.

Criterion	Evidence required	Data and Methods
1. An ephemeral resource contributes a large proportion of the consumer's diet	A large proportion of the diet is typically composed of a species or food type that shows a pulsed seasonal distribution	A variety of methods for quantifying diet composition exist, including direct observation of feeding, gut content dissection, faecal/regurgitate dissection, metabarcoding and stable isotope analysis. Requires that relevant aspects (e.g., biomass, abundance) of the favoured resource are measured over time within at least one season and analysis reveals a pulsed intra-year relationship.
2. Asynchrony between consumer and resource phenology is increasing over time	Analysis of time-series of consumer and resource phenology, with a test of whether trends in timing differ and whether this leads to an increase or decrease in asynchrony	Requires a time series that covers a period of temperature change. A large number of phenological time series exist, as recorded by researchers, citizens, herbaria, etc. Statistical analysis of increasing asynchrony is easily achieved by including an interaction between year and species. Inference of whether asynchrony is increasing or decreasing requires inspection of predictions based on estimated elevations and slopes of the modelled relationships for each species ¹⁸ .
3. Variation in asynchrony is driven by interannual variation in temperature	Identification of the time period(s) over which consumer and resource is sensitive to temperature. Evidence that differential temperature sensitivity is the driver	A variety of methods exist for identifying the time period over which phenology of each species responds to temperature ⁴⁸⁻⁵¹ . Confidence in attribution can be increased by experiments ²⁶ or by including year as a term in the model ⁴¹ , thereby de-trending the phenology data ⁵² . Estimating temporal trends in temperature variables is also worthwhile, as differing trends may generate asynchrony ^{53,54} .
4. Asynchrony impacts negatively on consumer fitness	A suitable measure of consumer fitness decreases with increasing asynchrony	Can be assessed within years (relative fitness) or among years (mean fitness) or both ^{30,31} . Depending on how asynchrony varies across individuals or years, the relationship between fitness and asynchrony may be a linear decline or a humped relationship. If the former, care may need to be taken to establish causation ⁵⁵ . Ideally, models should take into account both asynchrony with peak resource and phenological distribution of the resource ^{56,57} . Studies of impacts on relative fitness are informative regarding selection and opportunities for adaptation, whereas studies on mean fitness may be informative regarding demographic rates ¹³

5. Asynchrony impacts negatively on consumer population size, density, or growth

Negative effects of asynchrony on fitness (4) that have a negative effect on population size/growth, as assessed over multiple years

Requires long-term data on asynchrony and population size or density. The impact of asynchrony on demographic rates can be incorporated into a population model⁵⁸ or the causal pathways between asynchrony and population growth can be assessed in a structural equation model⁵⁹. It is important to rule out a causal effect of other variables (e.g., land-use, resource availability, sea ice, range shifts) that could cause populations to change over time²⁴. Such confounding effects can partially be accounted for by including year as a term to detrend the analysis^{41,52}. An alternative approach involves modelling a population's ability to persist on the basis of demographic and quantitative genetic parameters^{45,60}.

157

158

159 **Literature survey**

160 We conducted a broad survey (n=109) of published work on the match-mismatch hypothesis
161 (MMH) across terrestrial, marine and freshwater systems with the aim of: (i) examining the
162 evidence for temperature-mediated trophic asynchrony and its impacts and (ii) identifying
163 gaps in the evidence base, thereby allowing us to (iii) make recommendations for priority
164 areas for future work. To this end, we extracted data from published, peer-reviewed original
165 research in which a trophic interaction was studied in relation to any of the five criteria we
166 proposed as vital to the MMH (Table 1). Only original studies, where the specific interaction
167 between consumer and resource could clearly be identified were included (see
168 Supplementary Information for methodological details, and Supplementary Table 1 for the
169 papers identified as relevant by the data extractors).

170

171 *Taxonomic and geographical bias in the data*

172 The search identified 772 papers, of which the full text was found for 760. Of these, 571
173 were not relevant (e.g., no trophic interactions were reported, or none of the five criteria were
174 extractable), and 80 had no extractable data, resulting in a total of 109 papers that had

175 relevant data on 129 consumer taxa (Extended Data Fig. 2). All but six of the 109 trophic
176 interaction studies were in Europe or North America (Fig. 1). The majority of trophic
177 interactions were terrestrial (81.5% of the data), with marine (14%) and freshwater
178 interactions (4.5%) being scarcer. For most interactions the consumer was a secondary
179 consumer (58%), with studies of primary (36.5%) or higher than secondary (5.5%)
180 consumers less common. Birds made up the majority of the consumer taxa studied (53%),
181 while 29.5% of taxa were insects, 8% were fish, 5% were mammals and 4% were
182 crustaceans.

183

184 *Testing the five criteria*

185 The most tested criterion was criterion 1 (97% of interactions, $n=125/129$) - relating to
186 dependence on a seasonally pulsed resource (Fig. 2, top panels). However, rather than
187 conducting direct tests on the seasonal distribution of resources, 74% ($n=92/125$) of these
188 included only a statement based on *a priori* knowledge of the natural history of the system
189 that the resource was both ephemeral and important to the consumer.

190 Of the study systems that were tested for dependence on a seasonally pulsed
191 resource (i.e., where the resource was ephemeral and the consumer was a specialist), 42%
192 ($n=15/36$) showed such dependence. Excluding the cases where criterion 1 was not
193 explicitly tested, criterion 2 was the most frequently tested (68% of interactions, $n=88/129$),
194 relating to whether phenological asynchrony was increasing over time (Fig. 2, top panels).
195 The remaining criteria were all tested substantially less frequently, with criterion 5
196 (population consequences) being tested least often (8% of interactions, $n=10/129$).
197 Surprisingly few studies report data for criterion 3 (Fig. 2, top panels), which relates
198 temperature to asynchrony (29% of interactions, $n=37/129$), and this was almost never
199 reported for marine and freshwater taxa. The distributions of criteria tested were broadly

200 similar across consumer levels and biomes, with the exception of primary consumers for
201 which criteria 3 and 4 appear slightly more common.

202 Our analysis could identify only two out of 129 consumer taxa for which all five of our
203 criteria have been tested at least once: both of these are forest-breeding passerine birds
204 studied in Europe - the great tit and the pied flycatcher. In a further 13 taxa, four out of five
205 criteria were assessed. In the remaining 114 taxa, three or fewer criteria were studied, with
206 the majority (58%, $n=75/129$) of consumer taxa having only two of the five criteria known
207 (Fig. 2, bottom panels). Breaking this same analysis down to the per study level, no single
208 study explicitly tested all five criteria (Extended Data Fig. 3). This is generally due to a
209 tendency for studies to focus on either phenology slopes (criteria 2 & 3), or the
210 consequences of asynchrony (criteria 4 & 5). Only a handful of studies detail temporal
211 slopes, temperature slopes, and consequences of asynchrony in one study⁶¹⁻⁶³.

212

213 *Phenology slopes over time and temperature*

214 Consumer and resource responses appear to be positively correlated across studies, with
215 consumers showing a slight tendency to advance their phenology by less than their resource
216 (Fig. 3). In 61% ($n=58/95$) of the cases, the phenology slope over time was greater for the
217 resource than for the consumer (Fig 3a,b). For the phenological response to temperature,
218 the consumer slope was greater than the resource slope in 59% ($n=13/22$) of cases (Fig
219 3c,d). The degree to which these patterns differ across biomes and trophic levels could not
220 be tested with this dataset, since the number of slope estimates is too low for non-terrestrial
221 and non-secondary consumers. Based on visual inspection, it appears that especially
222 terrestrial secondary consumers tend to be slower-advancing than their resource. However,
223 more data on underrepresented groups would be required to reach robust conclusions about
224 these patterns.

225

226 *Fitness and demographic consequences*

227 Fitness consequences in relation to trophic asynchrony (criterion 4) are studied in 36%
228 (47/129) of the consumers (Fig. 2, top panels). Consequences of asynchrony for offspring
229 (n=44) are studied over three times as often as consequences for adults (n=14, Fig. 4),
230 though it is possible that this reflects a research bias to study fitness components that are
231 more sensitive to asynchrony. In 28% of consumer taxa (n=13/47), no negative effect of
232 asynchrony on fitness was reported (Fig. 4). The least studied consequence of trophic
233 asynchrony is its effects on population demography (criterion 5, Fig. 2, 4). In half of these
234 interactions (n=5/10), no effect of asynchrony was reported.

235

236 **Discussion**

237 Our literature survey on the ecological impacts of temperature-mediated trophic asynchrony
238 reveals that the full causal chain from temperature change, to temperature-driven shifts in
239 seasonal timing, consumer-resource synchrony, and individual-to-population level impact
240 has rarely been studied. Only two out of 129 taxa were studied for all criteria, and for the
241 majority of study systems, only one or two out of five criteria were met. The available studies
242 were strongly biased toward terrestrial secondary consumers (especially birds) in the
243 Northern Hemisphere (largely Europe and North America). Notably, the effects of climate
244 warming on trophic asynchrony in aquatic systems and in the Southern Hemisphere are
245 understudied⁶⁴, although this could be reflective of the small amount of temperate land mass
246 in the Southern hemisphere. Tropical studies are also under-represented, but this may partly
247 indicate a reduced importance of temperature as a phenological cue in tropical ecosystems⁸.
248 Crucially, demographic consequences of trophic asynchrony are the least studied of the five
249 criteria, despite this knowledge being the most important to conservation.

250

251 *Bias in the Match-Mismatch Hypothesis evidence base across biomes*

252 Terrestrial systems were by far the most represented of the three environments that we
253 considered, presumably by virtue of the comparative ease of collecting data on both
254 phenology and fitness in these systems. This ease of data collection is evident in the great
255 contribution that citizen science data collectors have made to the study of terrestrial
256 phenology^{47,65–68}, which is rare for aquatic systems. Monitoring phenology of many aquatic
257 organisms is hampered by their wide ranges and underwater habitats⁶⁹, and compounded by
258 the logistic and financial challenges encountered during offshore research. As a result,
259 relatively few multi-decadal phenological time series have been collected at sufficient
260 resolution to capture seasonal changes^{70,71}. Moreover, separate sampling programmes are
261 often needed for consumer and resource (e.g., piscivorous birds and their prey)⁷², and even
262 if resources can be quantified, many aquatic organisms are generalist feeders, further
263 adding to the difficulties in quantifying the MMH. Citizen scientists can, however, collect
264 valuable data on the terrestrial stages of aquatic organisms (e.g., dragonflies), or aquatic
265 seasonal events that can be observed from shore (e.g., amphibian spawning, floating algal
266 blooms). Furthermore, with ongoing technological innovation in data collection
267 methodologies, it may become possible to widen the aquatic evidence base for some taxa.
268 For example, radar can be used to quantify aquatic-terrestrial subsidies based on insect
269 emergence, providing detailed measures of the timing and size of resource pulses⁷³, and
270 satellite-based observation tools are providing a wide-scale perspective on phytoplankton
271 phenology changes⁷⁴. It would therefore be valuable to consider how diverse data sources,
272 and lines of evidence, can be fruitfully combined to advance our knowledge of the
273 importance of the MMH in aquatic systems.

274 Despite their overrepresentation in MMH research, even in terrestrial systems there
275 are biases and gaps in the evidence-base that extend beyond the aforementioned
276 geographic biases (Fig 1). Of the terrestrial studies, temperate forest taxa and birds in
277 particular predominate, which is likely due to the fact that seasonality increases with latitude.
278 Temperate forests experience a pronounced seasonal temperature-mediated pulse in

279 resources⁴⁶, and they present particularly suitable study systems to study individual fitness in
280 the wild (e.g., cavity nesting birds). In aquatic systems, individual marking of philopatric
281 seabirds and pinnipeds permits some components of fitness to be monitored⁷⁵, but this is
282 much harder for underwater organisms⁷⁶. Likewise, for many widely-distributed groups such
283 as fish, invertebrates, and plankton, individuals cannot be sampled repeatedly, and
284 populations can rarely be sampled to the extent that demographic implications of asynchrony
285 can be assessed. On the other hand, invertebrates are more amenable to experimental
286 study³², and numerous national surveys of population sizes exist^{77,78} that could be used to
287 infer demographic consequences of trophic asynchrony. Another key research gap in
288 aquatic systems involves the specific role of cross system consumer-resource interactions in
289 mediating trophic asynchrony. For example, some freshwater consumers feed upon
290 terrestrial resources, which represents a substantial source of nutrients⁷⁹. The delivery of at
291 least some of this material is strongly seasonal. Leaf fall, for example, is triggered by
292 photoperiod in conjunction with drought and temperature⁸⁰. Aquatic phenology research
293 would greatly benefit from increased consideration of the synchrony between freshwater
294 consumers and terrestrial resources.

295

296 *Further challenges in studying the MMH*

297 We recognise that studying these five criteria and improving the evidence base regarding the
298 risks posed by the MMH will not be straightforward and we have already discussed how
299 aquatic environments present particular challenges, but other complexities remain. While
300 criteria 2 and 3 are perhaps the easiest to satisfy, even here challenges exist in attributing a
301 change in phenology/asynchrony to temperature, as sensitivity estimates can be obscured
302 by non-climate drivers or compensatory mechanisms^{62,81}. For example, changing nutrient
303 availability and light conditions can influence the seasonal timing of phytoplankton
304 blooms^{82,83}, but would not be expected to affect consumer organisms in the same way. In
305 this article we have simplified the interaction between resource and consumer to a single

306 metric, the asynchrony between the peak demand of consumer and availability of the
307 resource. However, as the MMH predicts that consumer fitness relates to resource
308 availability during a particular window⁹, consumers might in addition to asynchrony be
309 sensitive to the height and width of the resource^{11,37}, either of which could be sensitive to
310 temperature and exacerbate or ameliorate effects on fitness. Although the potential for
311 resource abundance to influence fitness is widely acknowledged, it is unusual for studies on
312 the MMH in relation to fitness (criterion 4) or population size (criterion 5) to include its
313 effect^{56,57}. We realize that especially criterion 4 may be hard to satisfy for study systems
314 where individuals cannot be studied, which should not discourage people from working on
315 such systems. Whilst fulfilling all the other criteria would allow one to infer whether the
316 mismatch is causing population declines, there are clear advantages of studying individuals
317 within populations. Apart from the fact that such data helps demonstrate causative effects of
318 asynchrony on fitness, it can tell us whether seasonal timing will be under directional
319 selection.

320 Where the resource is in fact a guild (caterpillars, phytoplankton) rather than a
321 species, temperature-mediated shifts in the aggregate phenology may arise from a variety of
322 processes, from similar plastic responses of different species, to changes in the relative
323 abundance of early- and late-blooming constituent species, even when these species
324 independently might show no or weak phenological shifts⁸⁴. While the effect of asynchrony
325 on the consumer may not be sensitive to these two scenarios, if we want to project
326 phenological changes into the future we need to understand the processes that underpin
327 community phenological responses. An obvious solution to this problem is to improve the
328 species level resolution of sampling⁸⁵, but this can be costly and impractical in the short
329 term, and might require new sampling approaches such as eDNA⁸⁶.

330

331 *Is trophic asynchrony of conservation concern?*

332 This review reveals a lack of robust evidence for the MMH, and even the two best studied
333 taxa in terrestrial systems present a mixed message. In great tits, matching with the
334 caterpillar peak has fitness impacts at both the individual and population level^{30,31}, but trophic
335 asynchrony currently poses no threat to their population persistence^{31,45}. Pied flycatchers
336 also perform worse when poorly matched with the caterpillar peak^{22,87}, but, in contrast to
337 great tits, declines in asynchronous flycatcher populations have been recorded in the
338 Netherlands³⁶. Nevertheless, those pied flycatcher populations have been increasing again
339 since 2002⁸⁸. Interestingly, pied flycatchers breed about two weeks later than tits⁸⁹, the
340 average nest is rarely matched with the caterpillar peak⁹⁰, and a long-term study in the
341 Netherlands found no correlation between annual mean asynchrony with the caterpillar peak
342 and the strength of the seasonal decline in the number of recruits⁹¹. Moreover, pied
343 flycatchers are more generalist than tits in the nestling diet⁹², so it remains uncertain to what
344 extent these flycatcher populations will be negatively affected by trophic asynchrony
345 compared to specialists.

346 It would nevertheless be premature to conclude from this that a relative shortage of
347 evidence for demographic consequences of trophic asynchrony constitutes evidence of
348 absence of an effect. Even in species for which negative population consequences are not
349 yet apparent, such as great tits, it is possible that continued increases in temperature will be
350 problematic. Application of an integral projection model to a UK population of great tits
351 suggested that under a high emission scenario, more rapid responses of the prey species
352 (the winter moth caterpillar) coupled with limits to plasticity in great tit hatch date being
353 reached, lead to an acceleration in directional selection. An increase in evolution of hatch
354 date timing was to an insufficient degree to prevent negative consequences of trophic
355 asynchrony, and the population in that scenario is projected to have an increased risk of
356 extinction⁹³. Such demographic approaches should be greatly expanded upon, and provide a
357 unique way to understand which life stages will likely matter from the perspective of
358 pathways leading to shifts in population growth rate and density⁵⁸.

359

360 **Research Priorities**

361 Based on our five criteria and our review of the literature we identify six priorities for future
362 work to properly test the match-mismatch hypothesis and its impacts.

363 1. **From cause to effect - focusing on population consequences:** There is an urgent
364 need for studies that consider the full causal chain, from climate driver to seasonal
365 timing, synchrony, and individual-to-population level impact. In particular, we need
366 many more tests of the impact of asynchrony on population change (criterion 5),
367 across taxa and habitat types. This most important criterion from the perspective of
368 conservation and policy^{13,37} and yet has received the least attention. Furthermore,
369 given that the population impacts of trophic asynchrony at one location may be
370 buffered by matching at another location⁹⁰, we strongly advocate expanding the
371 spatial scale of current research to include multi-population studies. This will allow
372 the consequences of phenological shifts to be interpreted in the context of other
373 universal climate warming responses such as range shifts.

374 2. **Balancing the evidence - data collection and synthesis for aquatic systems:**
375 Despite the marine origin of the MMH, current monitoring and research has so far led
376 to a limited understanding of the MMH in marine and freshwater systems, compared
377 to terrestrial habitats. It is imperative for funders to continue to support time series,
378 since with each passing year the statistical power of these to reveal patterns
379 improves. We further recommend for underused historic records, including museum
380 collections and naturalist observations, to be coupled with new work on these
381 systems to create well documented long time series within a matter of years.
382 However, we must also ask how additional monitoring approaches (e.g., eDNA,
383 radar) might be usefully combined with “traditional” monitoring approaches, to
384 expand the species representation, monitoring of individual states and fitness

385 consequences, and spatial coverage of aquatic ecosystem studies, and support a
386 broader understanding of changes in phenological asynchrony (criterion 2) and the
387 role of temperature as a driver (criterion 3) in these systems.

388 **3. Environmental drivers of phenology - beyond temperature:** Here, we have
389 addressed phenological asynchrony in relation to temperature (criterion 3), the best-
390 studied driver. However, the environmental drivers of phenology vary geographically.
391 For instance, at lower latitudes seasonally pulsed precipitation is a more important
392 driver of phenology⁸, and at higher latitudes the timing of snow melt is a key
393 mechanism⁹⁴⁻⁹⁹. In order to gain a global perspective on the risks posed by climate-
394 mediated phenological asynchrony poses there is an urgent need to apply our
395 framework to alternative environmental drivers of phenology.

396 **4. Assessing the risks - global predictions and species traits:** We need more
397 studies on trophic asynchrony and its drivers at different latitudes and many more to
398 be conducted outside of Europe and North America (see Fig. 1). As data on the MMH
399 accumulate, a fruitful approach would be to conduct comparative analyses to identify
400 the taxonomic groups, trophic levels, environments and regions where fitness or
401 population impacts of phenological asynchrony (criterion 4 and 5) are most likely.
402 Based on first principles we may expect temperature-mediated asynchrony to be
403 more frequent and deleterious when the consumers are endotherms rather than
404 ectotherms⁸, income rather than capital breeders^{24,37}, and at higher latitude regions
405 experiencing the most seasonal climates and the most rapid climate change⁸.
406 However, empirical validation of these predictions is lacking.

407 **5. Observing interactions - enhancing the role of citizen science:** Mass
408 participation citizen science has resulted in millions of phenological records that
409 underpin many of the studies quantifying phenological shifts^{7,17,65,100} and can even be
410 used to project weather records into the past¹⁰¹. A strength of these schemes is their
411 spatial as well as temporal coverage. In some instances it is possible to identify the
412 phenology of consumer species and their resources from existing datasets¹⁰², but this

413 requires the assumption that co-occurring species are actually interacting. While
414 using data amassed over larger spatial scales (e.g., via citizen science or remote
415 sensing) is attractive as a means to examining geographic variation in temporal
416 trends in asynchrony (criterion 2) and temperature sensitivity (criterion 3) or fitness
417 consequences (criteria 4&5), care is required in matching data at a resolution that is
418 pertinent to the trophic interaction^{103,104}. Moreover, we are not aware of any study
419 combining citizen science-derived datasets to study the impacts of asynchrony of
420 specific trophic interactions on population change (criterion 5). Therefore an
421 opportunity exists for development or extensions of citizen science schemes to
422 collect data on the phenology of species interactions across trophic levels and on the
423 fitness and/or population sizes of the consumer.

424 6. **Clarifying the concept - “asynchrony” or “mismatch”**: There exists a degree of
425 terminological inconsistency in relation to the MMH, which may confuse attempts at
426 achieving a common understanding of the potential importance of this phenomenon.
427 Many studies that claim to address “mismatch” identify the conditions that could lead
428 to greater asynchrony, but stop short of explicitly testing whether asynchrony leads to
429 any negative consequences for the consumer. Where no evidence for negative
430 repercussions is presented we encourage authors to use the term “asynchrony”,
431 rather than “mismatch”, which implies a negative consequence.

432

433 **Concluding remarks**

434 Temperature-mediated trophic asynchrony and its consequences are widely discussed in
435 global change research and have been intensively studied over the past two decades. In this
436 study we have presented five criteria that together provide a causal chain to explicitly
437 demonstrate the risk that temperature-mediated asynchrony poses to populations, which we
438 hope will strengthen future work. In an extensive review of the literature we found that no
439 single study and only two study systems have tested all five criteria, with a clear deficit of

440 studies considering the impact of asynchrony on population size, which is the most important
441 criterion from a conservation perspective³⁷. This means that at present we cannot state from
442 the literature that temperature-mediated trophic asynchrony will have a widespread negative
443 impact on consumer population size or growth. We identify six research priorities, which
444 need to be tackled to get a comprehensive understanding of the frequency and magnitude of
445 trophic asynchrony and its impacts on consumers. A more consistent approach to the study
446 of the match-mismatch hypothesis and its population consequences at the global scale will
447 allow us to better target conservation efforts and provide much needed evidence for possible
448 consequences of one of the most intriguing impacts of climate change on global biota:
449 phenological change.

450

451 **Figure legends**

452 Figure 1. Locations of studies on phenological asynchrony identified by our analysis,
453 subdivided by biome (light blue = freshwater, dark blue = marine, orange = terrestrial) and
454 consumer trophic level (triangles = primary, squares = secondary, circles = >secondary).
455 There is a clear geographical bias of studies, with a considerable overrepresentation of
456 Europe and North America.

457

458 Figure 2. Individual criteria tested across taxa (a,b), and the total number of criteria tested
459 per taxon (c,d). The most tested criteria (a,b) were 1 “ephemeral resource” and 2
460 “phenological change over the years”. The total number of criteria tested (c,d) was two out of
461 five for most taxa, and all five criteria were tested for only two out of 129 taxa (c,d). The left
462 panels (a, c) are divided by trophic level, and the right panels by biome (b, d).

463

464 Figure 3. Consumer versus resource slopes in relation to year and temperature. Symbol
465 shapes represent consumer trophic level (triangles = primary, squares = secondary, circles =
466 >secondary), colour represents biome (light blue = freshwater, dark blue = marine, orange =
467 terrestrial) and larger symbols are from longer time series (average 21 years, range 6 to 119
468 years). The solid diagonal line represents an equal rate of change by consumer and
469 resource. Where the resource slope < 0 , points above the line represent systems where
470 resource phenology is advancing by more than that of the consumer, whereas points below
471 the line represent systems where consumer phenology is advancing more rapidly than
472 resource phenology. Where resource slope > 0 , points below the line represent systems
473 where resource phenology is delaying by more than that of the consumer, whereas points
474 above the line represent systems where consumer phenology is delaying more rapidly than
475 resource phenology.

476

477 Figure 4. Number of taxa in which consequences of trophic asynchrony were studied,
478 divided into those where the effect reported was negative or neutral (statistically non-
479 significant, no positive effect of trophic asynchrony was ever reported for this taxon). Results
480 are clearly biased toward juvenile rather than adult life stages.

481

482 **Data availability**

483 All data files related to this review are available at the Open Science Framework:

484 <https://osf.io/c8xzd/>.

485

486 **Code availability**

487 All R code to generate the results in this paper can be combined with the data files, and are

488 available at the Open Science Framework: <https://osf.io/c8xzd/>.

489

490

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772 TH, OWL, NP, and BCS contributed to conceiving ideas. All authors contributed to editing
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774 JCW extracted data for the analyses. JMS, ABP, AA, CH, KK, and SJT contributed to writing

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777

778 **Competing Interests**

779 The authors declare no competing interest.