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1	Exposure to Bioaerosols at Open Dumpsites: A Case study of bioaerosols exposure
2	from activities at Olusosun Open Dumpsite, Lagos Nigeria.
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6 7	Abstract Activities associated with the open dumping of municipal solid waste has the potential to have
8	greater impacts on the environment and public health compared to other forms of waste-to-land
9	treatment of such wastes. However, there is a lack of quantitative data on the exposure to
10	bioaerosols from open dumpsites, hence impeding the development of effective interventions
11	that would reduce the risk of respiratory symptoms among scavengers and waste workers at
12	such dumpsites. This study investigated exposure to bioaerosols at Olusosun open dumpsite,
13	Lagos Nigeria using three methodologies; (1) Conducting a cross-sectional survey on the
14	respiratory health of the population on the dumpsite, (2) Measuring bioaerosol concentrations
15	in the ambient air by measuring four bioaerosols indicator groups (total bacteria, gram-negative
16	bacteria, Aspergillus fumigatus and total fungi) using a Anderson six stage impactor sampler,
17	(3) Measuring activity related exposures to bioaerosols using an SKC button personal sampler.
18	After a cross sectional health survey of 149 participants (waste workers, scavengers,
19	middlemen, food vendors and business owners), smokers reported higher symptoms of chronic
20	cough (21%) and chronic phlegm (15%) compared to non-smokers (chronic cough 15%,
21	chronic phlegm 13%). Years of work > 5 years showed no statistically significant association
22	with chronic phlegm (OR 1.2, 95% CI 0.4-3.4; p>0.05) or asthma (OR 1.8, 95% CI 0.6-5.2;
23	p>0.05). At the 95 <sup>th</sup> percentile, the concentration of gram-negative bacteria was the highest
24	(2188 CFU/m <sup>3</sup> ), then total bacteria (2189 CFU/m <sup>3</sup> ), total fungi (843 CFU/m <sup>3</sup> ) and Aspergillus
25	fumigatus (441 CFU/m <sup>3</sup> ) after ambient air sampling. A comparison of the data showed that the

26 activity-based sampling undertaken using body worn personal sampler showed higher bioaerosols concentrations  $(10^4 - 10^6 \text{ CFU/m}^3)$ , i.e. 2-3 logs higher than those recorded from 27 28 static ambient air sampling. Bioaerosol exposure was highest during scavenging activities compared to waste sorting and site supervision . Particle size distributions showed that 41%, 29 30 46%, 76% and 63% of total bacteria, gram-negative bacteria, Aspergillus fumigatus and total 31 fungi respectively were of respirable sizes and would therefore be capable of penetrating deep 32 into the respiratory system, posing a greater human health risk. This study has shown that 33 exposure to bioaerosols can be associated with activities undertaken at open dumpsites and may contribute to the high prevalence of the chronic respiratory symptoms, among the workers 34 35 in such environments.

36

### 1. Introduction

37 The rate of solid waste generation in developing countries has grown at a steady rate in recent 38 years, especially due to urbanization and the accompanying urban population growth. Sub-39 Saharan Africa alone is estimated to generate 62 million tonnes per year of municipal solid 40 waste (MSW) with a corresponding annual rate of change of urban population of 2.27 percent 41 per year (UN-HABITAT 2009; Hoornweg and Bhada-Tata 2012b). Despite recording the 42 highest rate of urbanization in the world, most developing countries unfortunately do not have 43 the infrastructure to properly manage and treat their municipal solid waste. This has resulted in 44 waste management authorities resorting to open dumping as a cheap method for managing their 45 municipal solid waste (Srivastava et al., 2015). Hence, it is unsurprising that 17 of the world's 46 50 biggest dumpsites are located in the continent of Africa, of which 6 are in Nigeria (UNEP 47 2015). This form of waste treatment involves the uncontrolled dumping of waste MSW on open 48 land areas forming large waste hills with time. Microbial activities are central to the treatment 49 of biodegradable MSW and agitation activities, such as tipping, spreading and waste sorting 50 have been shown to aerosolise these microbial agents, including the pathogenic ones (Stagg et 51 al., 2010). Similarly, the analysis of the composition of MSW in sub-Saharan Africa have 52 shown to have high biodegradable (40-60%) and high moisture content (Ogwueleka 2009; 53 Hoornweg and Bhada-Tata 2012a). Furthermore, these characteristics in combination with the 54 hot and humid weather conditions in the region favours the growth of microorganisms 55 including pathogens that are eventually aerosolized during agitation activities at the dumpsite 56 (Odewabi et al., 2013; Epstein 2015). The individuals working on dumpsites such as scavengers, waste workers and vendors are often exposed to elevated levels of these pathogenic 57 58 microorganisms majorly by inhalation as most have been noted to work in these conditions 59 without personal protective equipment (Odewabi et al., 2013). Workers exposure to organic 60 dust laden with bioaerosols have been reported from composting and landfills facilities typically associated by high emission of organic dust during agitation activities, a condition similar to open dumpsites (Ray *et al.*, 2005; Schlosser *et al.*, 2012). In addition to exposure by inhalation, other exposure routes such as ingestion and the skin contact presents additional exposure risk to the workers. These workers have been reported to be at risk of contracting gastrointestinal diseases, blood infections such as HIV and hepatitis, physical symptoms like impetigo and musculoskeletal symptoms (Thirarattanasunthon *et al.*, 2012; Odewabi *et al.*, 2013).

68 Hitherto, respiratory health risk from exposure to bioaerosols at open dumpsite has received 69 limited attention from researchers compared to that associated with landfills and the 70 composting processes (Douwes et al., 2003; Heldal et al., 2015; Van Kampen et al., 2016). 71 Although activities such as tipping and spreading of waste may be common to landfills and 72 dumpsites, the latter has greater environmental and public health impact as there are little or no 73 controls over the safe handling, treatment and exposure limits to bioaerosols and particulate 74 matter (UNEP 2005; Hoornweg and Bhada-Tata 2012b). Moreover, there are currently very 75 limited literature on the exposure of dumpsite workers to bioaerosol emissions and the impacts on their respiratory health in a single study. As such, this research aims to investigate the 76 77 degree of exposure to bioaerosols associated with working at open dumpsites and the potential 78 impact this might have on the respiratory health of the workers at Olusosun open dumpsite.

### 79 **2. Materials and Methods**

#### 80 **2.1 Site description and waste characterization**

81 The open dumpsite used in this study was situated in Ojota, Lagos State Nigeria. Lagos State 82 has a population of 21 million with an annual growth of rate of 3.2% (LBS 2013). Of the three 83 principal dumpsites serving the urban population of Lagos i.e Ewu Elepe, Solous and Olusosun 84 dumpsites (LAWMA 2016), Olusosun is the biggest at 42.7 hectares (105.5 Acres), and is located between 6°35'40.9"N and 3°22'38.4"E (Oyeku and Eludoyin 2010). Olusosun is 85 86 estimated to receive approximately 2,100,000 tonnes of MSW, construction and E-waste per 87 year and has received approximately 17,150,000 to 24,500,000 tonnes of waste already since 88 it became active in 1992 (D-Waste 2014). Olusosun dumpsite was chosen because it receives 89 about 40% of the total municipal solid waste deposits in Lagos and the dumpsite is surrounded 90 by sensitive receptors located less than 200 m from the boundary of the dumpsite (Olorunfemi 91 2011).

#### 92 **2.2 Control site**

93 In order to be able to determine the impact of the open dumpsite on the concentration of 94 bioaerosols it was necessary to identify a 'control' site at which the concentration of 95 bioaerosols would not be impacted by activities at the open dumpsite. The two sampling points 96 were located ~ 10.79 km away from Olusosun open dumpsite within 6°29'58.9"N 3°22'47.5"E 97 and 6°29'45.3"N 3°22'45.0"E. Moreover, both locations and shared similar economic and 98 demographic characteristics with Ojota, where Olusosun dumpsite is located. The median age 99 of the sampled population at the control and Olusosun matched a national demographic 100 characteristic: 70% of adult population in Nigeria are less than 30 years in age (Reed and Mberu 2014) and are both part of the planned Lagos Central Business District. 101

#### 102 **2.3 Respiratory Health Survey**

103 A cross-sectional health survey was conducted across the Olusosun dumpsite between the  $12^{\text{th}}$ 104  $-20^{\text{th}}$  January 2017. Sample size was calculated using the metrics as recommended by WHO 105 (1991). Previous data on the prevalence of respiratory symptoms in landfill workers in India of 106 65%, was used to calculate the sample size (Ray *et al.*, 2004; 2005). With the confidence level 107 at 95%, bound on error at 5%, and statistical power at 80%, the sample size was 145 ±4 to 108 account for attrition.

109 Participants were classed into the five major occupational groups: Scavengers (those who 110 ravage the waste pile with tongs for recyclable resources); Waste workers (staff of the 111 Government agency in charge of the safe disposing of waste and maintenance at the dumpsite); 112 Business Owners (those with kiosks rendering other side services to members of the local 113 population); Middle Men (those who buy the recovered recyclable materials from the 114 scavengers); Food vendors (people who sell food to the other members of the local population). 115 Participants were selected at random and were required to have worked at the dumpsite for at 116 least one year to be eligible. A written or verbal informed consent was sought from all 117 participant before administering the questionnaires.

### 118 **2.3.1 Operational definitions**

119 These definitions were adapted from the operation guidelines of the European Community 120 Respiratory Health Survey (ECRHS) II, and the American Thoracic Society Division of Lung 121 Disease questionnaire (ATS-DLD-78A) (Ferris 1978; ECRHS 2002; Birring 2011; Shaikh et 122 al., 2012; Gibson et al., 2016). Chronic cough was defined as a consistent cough as much as 123 4 to 6 times a day, 4 or more days a week or a cough first thing morning or both for at least 3 124 months. Chronic Phlegm was defined as expectoration of sputum for as much as twice a day 125 for at least 4 days a week or expectoration of sputum first thing in the morning or both for at 126 least 3 months. Chronic wheezing was defined as wheezing in a dusty environment that would have triggered 2 or more episodes of shortness of breath. Asthma was classified as at least two reported asthma attacks or shortness of breath with wheezing in the past two months with normal breathing between episodes of shortness of breath or a diagnosed asthmatic by a physician.

### 131 **2.4 Air sampling and sample analysis**

Two different types of air sampling were undertaken; ambient static air sampling at points
around the dumpsite and at the control site using a six stage Andersen sampler and activitybased personal sampling using a body worn SKC button sampler.

### 135 **2.4.1 Ambient air sampling**

136 In Figure 1a, the sampling points and their distances away from the active part of the dumpsite 137 are shown on a map. The sampling locations were chosen based on the prevalent wind direction as shown in the wind rose and were located as close as possible to the point at which activity 138 139 was undertaken (Environment Agency 2017). The wind rose shows the average wind direction 140 in the area over the period of 30 years (WMO 2017). The active area was considered as the point source as activities such as tipping of waste, spreading, compacting and scavenging took 141 place here most frequently. Air samples were collected one day a week, for 13 weeks from the 142 5<sup>th</sup> April to 28<sup>th</sup> August 2017. Due to the limited amount of equipment and manpower available 143 144 to the researcher it was not possible to undertake concurrent sampling at all locations.

During sampling, air was pumped through the Anderson 6 stage sampler at the rate of 28.3 L min<sup>-1</sup> with a sampling time of 2.45 mins to avoid overloading of the petri dishes containing the selective media for the bioaerosols (Reinthaler *et al.*, 1997). Samples were stored in sterile sealed bags and were taken to the laboratory for further analysis. It is important to note that at the time of sampling, site activities at Olusosun were at their peak. The microorganism, agar type, growth inhibitants, incubation temperature and duration for the different bioaerosols are 151 shown in the Table S1. After incubation, the number of colony forming units on each plate 152 were counted, summed and subjected to positive-hole correction according to Macher (1989) and based on the volume of air sampled, were expressed as cfu/m<sup>3</sup>. This method, including 153 154 bioaerosol quantification and enumeration, followed the protocol outlined in the Technical Guidance Note (M9) for monitoring of bioaerosols at regulated facilities by Environment 155 156 Agency (2017) (England and Wales). Hence, the study was designed not to identify all 157 mesophilic bacteria and fungi at the dumpsite but to align with emission standards from UK 158 facilities, since Nigeria does not have any of such.

# 159 2.4.2 Activity-based air sampling

160 Activity-based air sampling was carried out in order to measure bioaerosol exposure from 161 specific activities engaged in by members of the population at the dumpsite. Three volunteers 162 were recruited who were involved in scavenging, waste sorting and site supervision on the 163 sampling days. They were required to spend most of their active time within 70 m from the 164 active area of the dumpsite as exposure from these activities at this distance was assumed to 165 be the highest compared to the other three sampling locations. The personal air sampler (Button 166 Aerosol Sampler, SKC Inc., PA USA) was mounted vertically on the volunteers as close to the 167 breathing zone as was practically possible. Sampling involved drawing sampled air at a flow 168 rate of 4 L/min through a stainless cassette containing a 0.8 µm pore size gelatin filter while 169 they went about their normal activities (Figure S1). The sampler captured air around the 170 breathing zone for 30 mins for the three activities measured. After each sampling episode, the filters were removed from the sampler, dissolved in a sterile 30 ml vial of extraction fluid (0.01% 171 172 Tween 80 in distilled water), and stored at 4 °C for transport back to the laboratory within 24 173 h. Serial dilution were performed on all the raw samples with sterile buffered fluid (0.01% Tween 80 in distilled water) down to  $10^{-4}$  using standard aseptic techniques. Subsequently, 10 174 175 µl of the samples were plated out onto the appropriate selective media and incubated as shown

in Table S1. The bioaerosol density  $(c/m^3)$  in the raw sample from the activity-based sampling was calculated using the following equation (Herigstad *et al.*, 2001):

178 
$$Density (CFU m^{-3}) = \frac{TC}{VP} \times DF \times BV$$

179 Where *TC*: total colony count, *VP*: Plated volume, *DF*: Dilution factor, *BV*: Volume of the raw sample.

### 180 **2.5 Statistical analysis**

a) *Qualitative health Survey*: Epi info<sup>™</sup> 7 developed by the Centre for Disease Control and
Prevention (CDC), USA., was used for preliminary data processing which were later
transferred to Microsoft Excel for further processing. The relationship between predictor
variables such as chronic cough, chronic phlegm, wheezing and asthma was analysed using a
logistic regression statistic model. Covariates such as age, sex, smoking status, 'years at study
area', 'use of nose mask' were adjusted for.

187 b) Air sampling: The test for normality was assessed by Kolmogorov-Smirnov test and 188 nonparametric statistical tests were applied to the data set that violated the rule. The differences 189 between the bioaerosol concentrations for the entrance, active area, dormant area and the 190 boundary were assessed using the one-way ANOVA/Wetch ANOVA of normality. The 191 homogeneity of variance was assessed by Levene's test of homogeneity of variance. If violated 192 by any of the variables, a Krustal-Wallis test was conducted to determine the difference in 193 means across the sampling points at the dumpsite. Generally, the statistical analysis was carried 194 out using IBM SPSS Statistics 22 for Windows (Version 22.0. Armonk, NY: IBM Corp., USA), 195 Microsoft Excel and graphs generated using Origin (2015b, Origin Lab Corp., Northampton, 196 MA, USA).

#### 197 **3. Results and Discussion**

#### 198 **3.1 Ambient air sampling**

199 The result of the one-way ANOVA conducted to determine if the difference in concentration 200 of total bacteria and gram- negative bacteria across the four sampling locations were significant 201 showed no statistically significant difference in their mean concentration, irrespective of the sampling location (total bacteria: F = 1.144, p > 0.05, gram-negative bacteria: F = 2.463, p >202 203 0.05). Furthermore, running a Kruskal-Wallis one-way ANOVA on total fungi and Aspergillus 204 *fumigatus*, total fungi indicated no statistically significant differences in mean concentration across the all sampling locations ( $\chi^2$  (3) = 5.799, p = 0.122). Aspergillus fumigatus however, 205 206 showed a significant difference between the means of the concentration at the active area and every other sampled location ( $\chi^2$  (3) =14.725, p = 0.002), suggesting a steep drop in the 207 208 concentration of Aspergillus fumigatus as it travels beyond 50 m from the active area.

The dumpsites entrance recorded a median concentration of  $1.312 \times 10^3$  cfu/m<sup>3</sup> (231-1.849 \times 10^3 209  $cfu/m^3$ ) and 1.144 ×10<sup>3</sup>  $cfu/m^3$  (786-2.043×10<sup>3</sup>  $cfu/m^3$ ) for total bacteria and gram-negative 210 211 bacteria respectively, a 1-2 log higher concentration compared to Aspergillus fumigatus and 212 total fungi (Table 1). The reason for this disparity is unclear, however it is possible that the use 213 open-back waste trucks (1 in 4 waste trucks on dumpsite is open-back) and the extended 214 waiting time at the entrance before tipping (mean: 2 hours), would likely contribute to 215 bioaerosols concentration at that sampling location, as bioaerosols can also be emitted during 216 the transport of waste to tipping point (Schlosser et al., 2016). The weather conditions were 217 favourable for most days, except for visit day 9, which started with a light shower but ended 218 just before sampling begun at the entrance location. For the dormant area, the form of activities 219 observed to take place here were mostly sorting and loading of recovered recycled materials 220 into trucks; moreover, scavenging was also observed to sparsely take place as well. Bioaerosols sampling result shows the concentration of bacteria were >  $10^3$  cfu/m<sup>3</sup> except for total bacteria 221

on visit day 6 (884 cfu/m<sup>3</sup>) and gram-negative on visit day 9 (939 cfu/m<sup>3</sup>). Total fungi on the other hand had concentration  $< 10^3$  cfu/m<sup>3</sup> for the entire sampling visit (See Figure 2 M-P). The boundary was the closest sampling location to the nearest sensitive receptor. Bioaerosols concentration up to  $1.63 \times 10^3$  cfu/m<sup>3</sup> and  $2.26 \times 10^3$  cfu/m<sup>3</sup> for total bacteria and gram-negative bacteria respectively were measured at this location even though *aspergillus fumigatus* and total fungi were  $< 10^2$  cfu/m<sup>3</sup> and  $10^3$  cfu/m<sup>3</sup> respectively.

228 Described in Table S2 are the meteorological conditions monitored during the sampling 229 operation. The mean air temperature at the site varied from a low of 27°C up to a high of 230 35.7 °C over the entire period. The relative humidity ranged between 64.3 - 88% with average 231 wind speed of 1.2 (minimum) and 4.2 (maximum) during the same period. Assessing the 232 relationship between meteorological factors and bioaerosol concentrations across the dumpsite, 233 the coefficient of Person's product moment correlation indicated that there was a statistically 234 significant positive correlation between the concentration of gram-negative bacteria taken at 235 the dormant area and atmospheric temperature (r = 0.649 n = 13, p < 0.05). Wind speed on the 236 other hand showed a positive and negative correlation with the concentration of Aspergillus *fumigatus* at the boundary (r = 0.571, n = 13, p < 0.05) and dormant area (r = -0.57, n = 13, p < 0.05) 237 238 0.05) respectively. Relative humidity showed no statistically significant correlation with the 239 concentration of bioaerosols with respect to the sampling locations. However, upon conducting 240 a multiple regression to further explain the overall change in the concentration of bioaerosols 241 across the dumpsite in response to the independent variables (meteorological factors), the result indicated that, of the three independent variables tested, only relative humidity added 242 statistically significantly to the prediction (p < 0.01) and that total bacteria concentration [F(3, 243  $(48) = 4.140, p < 0.05, adj. R^2 = 0.156$  was the most affected by the independent variable. 244

In general, the results of the bioaerosol sampling shows a progressive decline in concentration
with distance, from the source point, downwind (see Figure 2). However, the one-way ANOVA

247 indicated no statistically significant difference in mean concentrations for the sampling points 248 across the dumpsite for total bacteria (p>0.05) and gram-negative bacteria (p>0.05). This is 249 surprising given that some of the sampling points were located a considerable distance from 250 the active area (see Figure 1a). The result suggests a 'fairly-uniform' concentration for total 251 bacteria and gram-negative bacteria across the dumpsite, which may be due to either 252 meteorological factors or further bioaerosol emission from other sampling locations or both 253 (Ogden et al., 1969; Giner et al., 1999). Moreover, although the prevailing wind direction was 254 identified prior to sampling, during sampling it may be that the wind direction was variable 255 and there may have been swirling locally. This could mean that some sampling locations may 256 have been influenced by emission from other places and that would not have happened if the 257 prevailing wind direction was maintained throughout the sampling period. Gram-negative 258 bacteria particularly have been observed to be in high prevalence at open dumpsites located in 259 these regions, hence their higher concentrations in this study is not a surprise (Odeyemi 2012). 260 The steep decline in the concentration of Aspergillus fumigatus observed between the sampling location at the active area (50 m from active point; Median: 283 CFU/m<sup>3</sup>), the dormant area 261 (531 m; median: 43 CFU/m<sup>3</sup>) and the boundary (788 m; 51 CFU/m<sup>3</sup>) showed an 80-81% 262 263 reduction in concentration downwind. This behaviour of Aspergillus fumigatus was similar to 264 what was observed by Schlosser et al., (2016) where the decline was up to 77% between the 265 tipping point at the landfill and the property boundary.

The mean bioaerosol concentration at the control site was generally 1-log lower than the concentrations observed at Olusosun dumpsite (Table 3), supporting the fact that the Olusosun dumpsite was a major source of bioaerosols within the study area. In general, the bacteria concentration at the Olusosun dumpsite was typically 1-log higher than the fungi concentration. This is not surprising as Xu and Yao (2013) also reported that there are usually higher viable bacteria concentrations when compared to viable fungi concentrations when using an Anderson 272 impactor for air sampling. When the concentration of bioaerosols at all the sampling points 273 were added together, the 95<sup>th</sup> percentile of concentration of total bacteria (2189 CFU/m<sup>3</sup>), gram negative bacteria (2352 CFU/m<sup>3</sup>), Aspergillus fumigatus (300 CFU/m<sup>3</sup>) and total fungi (824 274 275  $CFU/m^3$ ) were 2-3 log lower than reported in literature where similar activities such as tipping, 276 spreading and compaction takes place (Kalwasińska et al., 2014; Sangkham et al., 2014). The 277 main explanation for such differences probably relates to the sampling locations on-site and 278 the sampling method employed. In this study using the Anderson six stage impactor, the closest 279 sampling location to the active area was 50 m downwind; compared to other studies where 280 sampling was either carried out at the working area or up 20 m away (Breza-Boruta 2012; 281 Schlosser et al., 2016).

282 Assessing bioaerosols emissions at Olusosun dumpsite against the expected limits set in the 283 UK by the Environment Agency (2017), the concentration limits of total bacteria and gram-284 negative bacteria were exceeded on most visits to the sampling site during ambient air sampling 285 (Figure 2, Table 2). Although the Aspergillus fumigatus and total fungi concentrations never 286 exceeded the expected limits, the set limits were based on an 8-hour time weighted average 287 estimation of exposure (HSE 2014). In this study however, the median daily exposure duration 288 for participants at Olusosun dumpsite was 11 hours, indicating longer exposures, hence greater 289 chances of inhaling more bioaerosols that could potentially cause related respiratory health 290 problems to the population.

291 **3.2 Activity–based air sampling** 

The result of the activity-based air sampling shows that the scavengers and workers alike were exposed to concentrations 2 -3log higher than the mean bioaerosol concentration measured during the static ambient air sampling (Table 3). The activity-based sampling using the personal body worn sampler allowed a picture to be obtained of the exposure of workers to bioaerosols as a result of engaging in multiple activities typical of those at a dumpsite. Activity such as scavenging typically involves material collection and sorting with material collection
being most prevalent during tipping and spreading of the waste while sorting tends to take
place further away from the active area (Afon 2012).

300 Table 2 shows the concentration of total bacteria, gram-negative bacteria and Aspergillus 301 fumigatus during site supervision, waste sorting and scavenging activities compared to the 302 control and ambient air sampling. Although Schlosser et al., (2016) alluded to the fact that the 303 use of different equipment for bioaerosol sampling will yield different result, but the use of a 304 combination of different sampling methods in this study produces a better validation of the 305 results (Sandelowski 2000). Unlike several other studies where bioaerosol exposure 306 measurements at landfills and dumpsites were restricted to static measurements from impaction 307 and/or sedimentation, this study combined personal sampling and static sampling, 308 consequently providing a robust dataset for the study. Furthermore, using bioaerosol data from 309 static sampling only is problematic since the participants are usually not stationed at the same 310 spots throughout the day's activities (Małecka-Adamowicz et al., 2007; Schlosser et al., 2016; 311 Fraczek et al., 2017).

312 In Figure S2, exposure to bioaerosols from the various activities are compared to ambient air 313 sampling and the result indicated higher exposure to total bacteria and gram-negative bacteria from scavenging up to  $10^6$  CFU/m<sup>3</sup>. Exposure to Aspergillus fumigatus on other hand was 314 found to be high when undertaking site-supervision activities  $(3 \times 10^5 \text{ CFU/m}^3)$  with lower 315 exposure from sorting  $(9 \times 10^4 \text{ CFU/m}^3)$  compared to scavenging (Table 3). This unusual spike 316 317 in Aspergillus fumigatus during site monitoring may not be unrelated to the property of the 318 waste (waste rich in fungi) tipped at the dumpsite at the time of sampling. Bioaerosols exposure 319 from scavenging and sorting activities accounted for the highest exposure levels at the dumpsite, 320 levels comparable to or higher than other forms waste-to-land treatment of municipal solid

waste such as landfilling and open windrow composing reported in some literature (Lis *et al.*,
2004; Odeyemi 2012; Wéry 2014b; Breza-Boruta 2016).

### 323 **3.3 Particle size distribution**

324 To further evaluate the potential degree of exposure to bioaerosols at Olusosun dumpsite, the 325 size distribution of bioaerosols were analysed and related to typical tidal volumes inhaled by 326 humans. During sampling using the six stage Andersen sampler, bioaerosols are deposited out 327 onto six separate agar plates located in the six stages of the sampler. The six stages of the 328 sampler are designed to mimic the human respiratory system with collection of progressively 329 smaller particles from stages 1 (top, upper respiratory tract) to stage 6 (bottom, lower 330 respiratory tract and alveoli) (Jensen et al., 1992; Tisch Environmental Inc. 2018). Particles of 331 inhalable fraction with aerodynamic diameter >4.7  $\mu$ m are deposited in stages 1 and 2, 332 representing deposition in the nasal area. Stages 3 and 4 (thoracic fraction with aerodynamic 333 diameter 2.1 - 4.7 µm) represents bronchial deposition, while deposition in the alveoli is 334 represented in stages 5 and 6 (respirable fraction with aerodynamic diameter  $< 2.1 \ \mu m$ ) 335 (Andersen 1958; HSE 2003, 2014).

336 Figure 3(a)-(d) shows the percentage size distribution of bioaerosols as deposited on the 337 different stages of the Anderson 6 stage sampler measured at the four sampling locations across the dumpsite. Taking into account all the bioaerosol collected from the 13 visit days, ~41% of 338 the total bacteria (ambient air: 1.9-3×10<sup>2</sup> CFU/m<sup>3</sup>; Scavenging: 1.2×10<sup>6</sup> CFU/m<sup>3</sup>; Site 339 supervision :  $6.0 \times 10^5$  CFU/m<sup>3</sup>; Sorting:  $4.8 \times 10^5$  CFU/m<sup>3</sup>) were sized at < 2.1 µm aerodynamic 340 341 diameter. Gram-negative bacteria on the other hand were slightly higher with  $\sim 46\%$  of the 342 gram-negative bacteria at Olusosun dumpsite being  $< 2.1 \,\mu$ m aerodynamic diameter. The data 343 for total fungi which also includes Aspergillus fumigatus indicates that ~76% of all the fungi 344 collected at Olusosun dumpsite were in the  $< 2.1 \,\mu m$  size range.

345 The aerodynamic diameter of bioaerosols is key to their dispersion, potential risk from 346 inhalation and disposition in either the tracheal, bronchial or the alveolar regions of the lungs 347 Bragoszewska et al., (2017) and Ferguson et al., (2017) had observed that bioaerosol exist 348 either as single cells or agglomerates depending on the season and the meteorological 349 conditions at the time of sampling. Their existence in an aggregated form, they argued, are 350 likely responsible for their settlement in the upper stages (aerodynamic diameter >  $3.3 \mu m$ ) of 351 the Anderson sampler either as aggregates of cells, cells associated with water droplets or 352 particulate matter. Moreover, high relative humidity can further increase both the size and 353 weight of the particle from absorption of ambient moisture, a phenomenon that favours higher 354 deposition at upper stages of the Anderson sampler (Total Bacteria: ~59%; Gram-negative 355 bacteria: ~54%) or dry deposition, giving a low reading for bioaerosol in the ambient air as 356 observed in this study (Liu et al., 2015; Smets et al., 2016). This is perhaps not surprising as 357 the multiple regression carried out in this study showed that, of the three metrological factors 358 tested, only relative humidity showed a statistically significant association with the concentration of bioaerosols (p < 0.01), especially total bacteria (F(3, 48) = 4.140, p < 0.05, adj. 359  $R^2 = 0.156$ ). The results shown in Figure 3 suggests a higher percentage of fungi (Aspergillus 360 361 fumigatus ~76%; and total fungi ~63%) were within respirable fraction compared to bacteria 362 (Total bacteria ~41%; gram-negative bacteria ~43%). Moreover, as observed by Deacon et al., 363 (2009), the recovery of Aspergillus fumigatus is usually highest within instruments pore size 364 range of 3.3 -1.1 µm as they are typically 2-3.5 µm in diameter. Particles of inhalable fraction 365 deposited in the nasal area and upper respiratory tract can usually be removed by the action of 366 the nasal and tracheobronchial escalators, which is a combined mucociliary function of 367 trapping deposited bioaerosol particles in mucus and removal by the action of cilia (Mason and 368 Nelson 2005). However, this does not happen when particles of respirable fraction get down in 369 to the pulmonary region where there is much more potential for health problems if deposited down there, especially those who are immunocompromised (Yoshida and Whitsett 2004;Thomas 2013).

372 The tidal volume, which is the expected volume of air displaced during normal breathing, is 373 approximately 7 mL/kg or 500 ml for a healthy young adult (Quanjer et al., 1993; Ricard 2003). 374 With a respiratory rate of 12 breaths per minute (Cretikos et al., 2008; Erden et al., 2015), the 375 scavengers, waste sorters and waste workers working 11-hours (median) at Olusosun dumpsite, will inhale approximately 3.96 m<sup>3</sup> of air containing approximately  $4.8 \times 10^5$  CFU/m<sup>3</sup> total 376 bacteria, 1.4  $\times 10^6$  CFU/m<sup>3</sup> of gram-negative bacteria and 5.2 $\times 10^4$  CFU/m<sup>3</sup> of Aspergillus 377 fumigatus during scavenging; 9.6×10<sup>4</sup> CFU/m<sup>3</sup> of total bacteria, 8×10<sup>5</sup> CFU/m<sup>3</sup> of gram-378 379 negative bacteria and 6.93×10<sup>4</sup> CFU/m<sup>3</sup> of Aspergillus fumigatus during waste sorting; 2.4  $\times 10^5$  CFU/m<sup>3</sup> of total bacteria, 9.7 $\times 10^4$  CFU/m<sup>3</sup> gram-negative bacteria and 2.3  $\times 10^4$  CFU/m<sup>3</sup> 380 381 of Aspergillus fumigatus when carrying out site supervision, Table S3.

382 Several inhalation studies have investigated the effect of bioaerosols particle size on bioaerosol 383 lungs deposition and the lethal dose required to initiate an infection. Although there are several 384 animal models to this effect, no human studies exist investigating this relationship (Thomas 385 2013; Dabisch et al., 2017). The use of mouse models to demonstrate infection from 386 Aspergillus fumigatus have been widely reported in literature, and that perhaps is because mice 387 displays pathological consequences similar to humans (Kupfahl et al., 2006; Dagenais and Keller 2009). Sheppard et al., (2004) administered a dose of 2.4 ×10<sup>3</sup> CFU/mouse of 388 389 Aspergillus fumigatus conidia to immunocompromised mice, resulting in a lethal pulmonary 390 infection in most of the mice, surviving between 5 to 12 days. The exposure to Aspergillus 391 *fumigatus* in this research (from scavenging, waste sorting and site supervision) however, was up to 2.74 x 10<sup>5</sup> CFU/day, a 2-log higher than reported by Sheppard et al., (2004). However, 392 393 considering the complex nature of the structure and defence mechanism of the human 394 respiratory system, it is unlikely that a similar dose will be lethal to humans, however, a prolong

exposure and neutropenia to will increase the risk of developing invasive aspergillosis by the
exposed workers especially, the immunocompromised (Dagenais and Keller 2009; Wéry
2014a). Gram negative bacteria have also been reported to be a potent pro-inflammatory
stimulus in the human lungs due to flagellin release from their flagella (Liaudet *et al.*, 2003).

# 399 **3.4 Respiratory symptom from workers**

Summarised in Table 4 is the socio-demographic description of the participants in the cross sectional health survey. Of the 149 participants, 130(87%) and 19(13%) were males and females respectively. The result also shows that participants spent 11 hours daily (median) at the dumpsite and recorded a median "year at work" as 5. Moreover, the median age of the participants was 30 years, and the smoker and non-smoker population was 61(41%) and 84(56%) respectively.

406 As shown in Figure 4 (A), smokers reported higher symptoms of chronic cough (21%) and 407 chronic phlegm (15%) compared to non-smokers (chronic cough 15%, chronic phlegm 13%). 408 This was however different with chronic wheezing and asthma, as non-smokers recorded higher prevalence (chronic wheezing 4%, asthma 1.3) when compared to smokers (chronic 409 410 wheezing 3.5%, asthma 0.7%). Figure 4 (B) further shows reported symptoms among the non-411 smoker population in relation to how long the members have been at the study location. The 412 occurrence of chronic cough (10%) and chronic phlegm (10%) were the highest for members 413 of the population that have been at the study location in the first 5 years. Chronic phlegm was 414 also noticed to be consistently reported to varying degree by non-smokers, irrespective of the 415 years spent working at the dumpsite. Generally, the reported chronic respiratory symptoms 416 were observed to have reduced as the years of spent at the dumpsite increased.

417 The result of the logistic regression showed a weak association between the independent 418 variable 'Use of Nose Mask' and occurrence of chronic cough (p > 0.05). Another independent variable was 'Years of work' which showed no statistically significant association with chronic
cough, chronic phlegm and asthma, for participants with years of work > 5 years (Table 5).

421 The prevalence of respiratory symptoms among the sampled population is not surprising since 422 previous researches have also reported similar respiratory symptoms among waste workers and 423 dumpsite scavengers. Athanasiou et al., (2010) and Garrido et al., (2015) for instance reported 424 high prevalence of asthma and COPD symptoms among solid waste workers in municipality 425 of Keratsini, Greece and city of Hamburg, Germany, respectively. Black et al., (2019) also 426 reported that 70% of the 1278 informal waste workers surveyed in their study at the dumpsite 427 in Nuwakot district Nepal, reported one form of respiratory symptoms or other. Moreover, 428 increased alveolar macrophages, neutrophils, eosinophils and lymphocytes in the sputum 429 samples of examined female rag pickers (female scavengers) working on a waste dumpsite in 430 Kolkata, India have been reported by Ray et al., (2009). Although the study did not report the 431 possible exposure concentration, the observed biomarkers from the body of the female rag 432 pickers were evidence indicating a heightened immune system, which would suggest that they 433 have been exposed to bacteria, endotoxins and viruses through inhalation.

434 Figure 6A shows that chronic cough and chronic phlegm recording a higher prevalence among 435 smokers was not surprising, since chronic respiratory symptoms are most of the time associated 436 with smoking (Abramson et al., 2002; Frank et al., 2006; Garrido et al., 2015). The differences 437 in percentage points in the prevalence of the reported respiratory symptom between smokers 438 and non-smokers was marginal (chronic cough -6%, Chronic phlegm- 2\%, chronic wheezing-439 0.5% and asthma-0.6%) and suggests that the reported respiratory symptoms were largely of 440 occupational origin than life style. This further supports growing evidence that chronic cough 441 and chronic phlegm are risk factors for chronic obstructive pulmonary disease (COPD) and 442 that reported cases of COPD among non-smokers are on the rise in developing countries (Salvi 443 and Barnes, 2009; de Marco et al., 2007).

444 It was expected that the non-use of a nose mask would greatly increase the odds of developing 445 these outcomes; however, the model was not sensitive enough to measure this. This perhaps is 446 because 87% of the respondents admitted not to have used a nose mask while working at the 447 dumpsite, and the few that did (13%), used them only occasionally when they had access to 448 PPE. This result should not be viewed as undermining the need to provide appropriate safety 449 interventions, as such interventions have been recommended by researchers to reducing the 450 risk of acquiring work related respiratory health problems (Lavoie et al., 2006; Kuijer et al., 451 2010). It was also observed that all the respiratory health outcomes were not associated with 452 'Years of work' > 5 years (p > 0.05) (Table 5). In their study of the respiratory health of brick 453 kiln workers, Shaikh et al., (2012) observed a similar result, however only for workers with 454 'years of work' > 10 years. The median 'years of work' at Olusosun dumpsite was 5 years and 455 this represented 46% of the participants (Table 4). By implication, the majority of the 456 participants who had reported to suffer chronic respiratory symptoms, would have contracted 457 either chronic phlegm or asthma within the first 5 years of working at Olusosun open dumpsite. 458 In a different report of a 5-year follow-up study of compost workers, Bünger et al., (2007) 459 observed a doubling of the number of workers with chronic bronchitis (RR 1.41; 95% CI 1.3-460 1.5) over the study period. Van Kampen et al., (2016) also observed an increase relative risk of 461 cough (RR 1.28; 95% CI 1.2-1.5) and phlegm (RR 1.32; 95% CI 1.2-1.5) and among compost 462 workers working for at least 5 years at the composting facility irrespective how long their years 463 of work was. They however, reported COPD as having no significant association with years of 464 work. This study has demonstrated that working at dumpsites such as Olusosun may increase 465 the odds of contracting chronic respiratory symptoms. This perhaps is primarily due to the 466 limited use of PPE by the population, limited control over activities on the dumpsite and longer 467 exposure hours, rather than longer years of work in such environments. In undertaking this 468 study, there were some limitations which were largely due to limited resources and lack of 469 health records. Firstly, there was no morbidity record for the workers at the dumpsite making 470 it difficult to compare the number of those who would have left the dumpsite due to ill health 471 from exposure, with those currently working at the dumpsite within the same period. This may 472 result in the underestimation of the extent of the respiratory health problems suffered by the workers in Olusosun dumpsite. Secondly, the findings for this study were largely based on a 473 474 subjective inquiry without a complementary objective clinical measure. Spirometry and 475 haematological profiling could not be carried out to assess the lungs functions and detect 476 inhaled PM in the blood. The third limitation was that only the areas downwind of the 477 dumpsite (selection based on the predominant wind direction across the dumpsite) was selected 478 for sample group near the dumpsite. This may present a bias in the data, as there was regular 479 swirling in the local wind during the survey activity, hence a possible exclusion of an important 480 exposed group. Other cofounding factors such as vehicular emissions and industrial emissions 481 at the study locations were not considered and consequently adjusted for in this study. Lastly, 482 the probable underestimation of the actual bioaerosol exposure concentration due to 483 uncultivatable and non-viable fraction of airborne microorganisms; a weakness inherent in the 484 use of the impaction sampling technique (HSE 2003; Persoons et al., 2010).

#### 485 **4. Conclusion**

486 This study has demonstrated that open waste dumpsites are a major source of bioaerosols that 487 can contribute adversely to the respiratory health of the waste workers, scavengers and others 488 that are working at the dumpsite. Bioaerosols concentrations at Olusosun open dumpsite were 489 2-3 log higher when compared to the control site. The results from the activity-based air 490 sampling suggests that exposure to bioaerosol at Olusosun open dumpsite may also depend on 491 the activities engaged in at the dumpsite. Scavenging recorded the highest level of bioaerosols 492 exposure when compare to waste sorting, site supervision and the sampled ambient air. Gram-493 negative bacteria represented the highest bioaerosol concentration in Olusosun dumpsite. The 494 particle size distribution for Aspergillus fumigatus and total fungi showed a greater portion approximately 76% and 63% respectively were sizes  $< 3.3 \mu m$ . These particle sizes are 495 496 respirable and can be deposited at the lower respiratory tracts. Observed also was the 497 prevalence of chronic respiratory symptoms such as chronic cough and chronic phlegm, with 498 earlier showing equally prevalence in both smokers and non-smokers and the later, more 499 prevalent among the non-smoker population. Further clinical investigation into the respiratory 500 health of the workers at the Olusosun dumpsite is necessary to determine the lungs function in 501 responds to exposure to the bioaerosols.

# 502 **5. Supplementary data**

503 Supplementary data associated with this article is available at <u>https://doi.org/10.5518/449</u>

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# 508 Ethics approval

509 Ethical approval was obtained from the Engineering Faculty Research Ethics Committee,510 University of Leeds, UK.

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# 513 **Referances**

- ABRAMSON, M., M. MATHESON, C. WHARTON and M. SIM. 2002. Prevalence of respiratory
   symptoms related to chronic obstructive pulmonary disease and asthma among
   middle aged and older adults. *Respirology*, 7(4), pp.325-331.
- AFON, A. O. 2012. A survey of operational characteristics, socio-economic and health effects
   of scavenging activity in Lagos, Nigeria. *Waste Management & Research*, **30**(7),
   pp.664-671.
- AGARWAL, S. 2017. Seasonal variability in size-segregated airborne bacterial particles and
   their characterization at different source-sites. *Environmental Science and Pollution Research*, 24(15), pp.13519-13527.
- ANDERSEN, A. A. 1958. New sampler for the collection, sizing, and enumeration of viable
   airborne particles. *Journal of Bacteriology*, **76**(5), p471.
- 525ATHANASIOU, M., G. MAKRYNOS and G. DOUNIAS. 2010. Respiratory health of municipal526solid waste workers. Occupational medicine, **60**(8), pp.618-623.
- BIRRING, S. S. 2011. Controversies in the evaluation and management of chronic cough.
   American journal of respiratory and critical care medicine, 183(6), pp.708-715.
- BLACK, M., J. KARKI, A. LEE, P. MAKAI, Y. BARAL, E. KRITSOTAKIS, A. BERNIER and A. F.
   HECKMANN. 2019. The health risks of informal waste workers in the Kathmandu
   Valley: a cross-sectional survey. *Public health*, **166**, pp.10-18.
- BRĄGOSZEWSKA, E., A. MAINKA and J. S. PASTUSZKA. 2017. Concentration and Size
   Distribution of Culturable Bacteria in Ambient Air during Spring and Winter in
   Gliwice: A Typical Urban Area. *Atmosphere*, 8(12), p239.
- BREZA-BORUTA, B. 2012. Bioaerosols of the municipal waste landfill site as a source of
   microbiological air pollution and health hazard. *Ecological Chemistry and Engineering. A*, **19**(8), pp.851-862.
- BREZA-BORUTA, B. 2016. The assessment of airborne bacterial and fungal contamination
   emitted by a municipal landfill site in Northern Poland. *Atmospheric Pollution Research*, 7(6), pp.1043-1052.
- 541 BÜNGER, J., B. SCHAPPLER-SCHEELE, R. HILGERS and E. HALLIER. 2007. A 5-year follow-up
  542 study on respiratory disorders and lung function in workers exposed to organic dust
  543 from composting plants. *International archives of occupational and environmental*544 *health*, **80**(4), pp.306-312.
- BURKOWSKA, A., M. SWIONTEK-BRZEZINSKA and A. KALWASIŃSKA. 2011. Impact of the
   municipal landfill site on microbiological contamination of air. *Contemporary Problems of Management and Environmental Protection*, 9, pp.71-7.
- 548 CRETIKOS, M. A., R. BELLOMO, K. HILLMAN, J. CHEN, S. FINFER and A. FLABOURIS. 2008.
   549 Respiratory rate: the neglected vital sign. *Medical Journal of Australia*, 188(11),
   550 p657.
- 551D-WASTE. 2014. Waste Atlas 2014 Report: The World's 50 Biggest Dumpsites [online].552[Accessed 29/07/2018]. Available from: <a href="http://www.d-waste.com/">http://www.d-waste.com/</a>.
- DABISCH, P., Z. XU, J. BOYDSTON, J. SOLOMON, J. BOHANNON, J. YEAGER, J. TAYLOR, R.
   REEDER, P. SAYRE and J. SEIDEL. 2017. Quantification of regional aerosol deposition
   patterns as a function of aerodynamic particle size in rhesus macaques using PET/CT
   imaging. *Inhalation toxicology*, **29**(11), pp.506-515.
- DAGENAIS, T. R. and N. P. KELLER. 2009. Pathogenesis of Aspergillus fumigatus in invasive
   aspergillosis. *Clinical microbiology reviews*, **22**(3), pp.447-465.

559 DEACON, L., L. PANKHURST, G. DREW, E. T. HAYES, S. JACKSON, P. LONGHURST, J. 560 LONGHURST, J. LIU, S. POLLARD and S. TYRREL. 2009. Particle size distribution of 561 airborne Aspergillus fumigatus spores emitted from compost using membrane 562 filtration. Atmospheric Environment, 43(35), pp.5698-5701. 563 DOUWES, J., P. THORNE, N. PEARCE and D. HEEDERIK. 2003. Bioaerosol health effects and 564 exposure assessment: progress and prospects. Annals of Occupational Hygiene, 565 **47**(3), pp.187-200. 566 ECRHS. 2002. The European Community Respiratory Health Survey II. European Respiratory 567 Journal, 20(5), pp.1071-1079. 568 ENVIRONMENT AGENCY. 2017. Technical Guidance Note M9. Environmental Monitoring of 569 Bioaerosols at regulated facilities 1ed. United Kingdom: Environment Agency. 570 EPSTEIN, E. 2015. Disposal and Management of Solid Waste: Pathogens and Diseases. 6000 571 Broken Sound Parkway NW, Suite 300: CRC Press, Taylor & Francis Group. 572 ERDEN, F., A. Z. ALKAR and A. E. CETIN. 2015. Contact-free measurement of respiratory rate 573 using infrared and vibration sensors. Infrared Physics & Technology, 73, pp.88-94. 574 FERGUSON, R. M. W., CHARLOTTE E. NEATH, ALEX J. DUMBRELL, CORINNE WHITBY and I. 575 COLBECK. 2017. Novel insights into the size distribution of bacterial bioaerosols at 576 composting sites. In: The Aerosol Society Focus Meeting 10-Bioaerosols, 8 June, 2017, 577 University of Bristol, UK. The Aeorosl Society. 578 FERRIS, B. G. 1978. Epidemiology Standardization Project-American Thoracic Society. Am 579 *Rev Respir Dis*, **118**, pp.1-120. 580 FRĄCZEK, K., J. KOZDRÓJ, R. GÓRNY, M. CYPROWSKI and M. GOŁOFIT-SZYMCZAK. 2017. 581 Fungal air contamination in distinct sites within a municipal landfill area. 582 International Journal of Environmental Science and Technology, 14(12), pp.2637-583 2648. 584 FRANK, P., J. MORRIS, M. HAZELL, M. LINEHAN and T. FRANK. 2006. Smoking, respiratory 585 symptoms and likely asthma in young people: evidence from postal questionnaire 586 surveys in the Wythenshawe Community Asthma Project (WYCAP). BMC pulmonary 587 *medicine*, **6**(1), p10. 588 GARRIDO, M. V., C. BITTNER, V. HARTH and A. M. PREISSER. 2015. Health status and health-589 related quality of life of municipal waste collection workers-a cross-sectional survey. 590 Journal of Occupational Medicine and Toxicology, **10**(1), p22. 591 GIBSON, P., G. WANG, L. MCGARVEY, A. E. VERTIGAN, K. W. ALTMAN, S. S. BIRRING, T. M. 592 ADAMS, A. F. BARKER, F. BLACKHALL and D. C. BOLSER. 2016. Treatment of 593 unexplained chronic cough: CHEST guideline and expert panel report. Chest, 149(1), 594 pp.27-44. 595 GINER, M. M., J. S. C. GARCÍA and J. G. SELLÉS. 1999. Aerobiology of Artemisia airborne 596 pollen in Murcia (SE Spain) and its relationship with weather variables: annual and 597 intradiurnal variations for three different species. Wind vectors as a tool in 598 determining pollen origin. International Journal of Biometeorology, 43(2), pp.51-63. 599 GLADDING, T. L. and C. L. GWYTHER. 2017. A study of the potential release of bioaerosols 600 from containers as a result of reduced frequency residual waste collections. Science 601 of the Total Environment, 576, pp.481-489. 602 GREEDY, D. and J. THRANE. 2012. Closed for business- A look at the closure of open dumps. 603 Waste Management World 9(6).

- HELDAL, K. K., L. MADSO and W. EDUARD. 2015. Airway inflammation among compost
   workers exposed to actinomycetes spores. *Annals of Agricultural and Environmental Medicine*, 22(2).
- HERIGSTAD, B., M. HAMILTON and J. HEERSINK. 2001. How to optimize the drop plate
   method for enumerating bacteria. *Journal of microbiological methods*, 44(2), pp.121 129.
- HOORNWEG, D. and P. BHADA-TATA. 2012a. What a waste: a global review of solid wastemanagement.
- HOORNWEG, D. and P. BHADA-TATA. 2012b. What a waste: a global review of solid waste
   management. Urban Development Series, Urban Development and Loca
   Government Unit, WB, Washington, DC 20433 USA: World Bank.
- HSE. 2003. Occupational and environmental exposure to bioaerosls from compost and
   potential health effects a critical review of published data [online]. [Accessed
   23/03/3018]. Available from: http://www.hse.gov.uk/research/rrpdf/rr130.pdf.
- HSE. 2014. General methods for sampling and gravimtric analysis of respirable thoracic and
   *inhalable aerosols* [online]. [Accessed]. Available from:
   <u>http://www.hse.gov.uk/pubns/mdhs/pdfs/mdhs14-4.pdf</u>.
- JENSEN, P. A., W. F. TODD, G. N. DAVIS and P. V. SCARPINO. 1992. Evaluation of eight
   bioaerosol samplers challenged with aerosols of free bacteria. *American Industrial Hygiene Association Journal*, 53(10), pp.660-667.
- KALWASIŃSKA, A., A. BURKOWSKA and M. SWIONTEK BRZEZINSKA. 2014. Exposure of
   Workers of Municipal Landfill Site to Bacterial and Fungal Aerosol. *CLEAN–Soil, Air, Water,* 42(10), pp.1337-1343.
- KUIJER, P., J. K. SLUITER and M. H. FRINGS DRESEN. 2010. Health and safety in waste
   collection: Towards evidence based worker health surveillance. *American journal of industrial medicine*, 53(10), pp.1040-1064.
- KUPFAHL, C., T. HEINEKAMP, G. GEGINAT, T. RUPPERT, A. HÄRTL, H. HOF and A. A.
  BRAKHAGE. 2006. Deletion of the gliP gene of Aspergillus fumigatus results in loss of
  gliotoxin production but has no effect on virulence of the fungus in a low dose
  mouse infection model. *Molecular microbiology*, 62(1), pp.292-302.
- LAVOIE, J., C. J. DUNKERLEY, T. KOSATSKY and A. DUFRESNE. 2006. Exposure to aerosolized
   bacteria and fungi among collectors of commercial, mixed residential, recyclable and
   compostable waste. *Science of the total environment*, **370**(1), pp.23-28.
- 637 LAWMA. 2016. Landfill Operations [online]. [Accessed 30th August 2016]. Available from:
   638 <u>http://www.lawma.gov.ng/lawma\_landfill.html</u>.
- LAWMA. 2017. Statistics of Refuse Deposited at Various Landfill Sites for 2014-2016. In: P. R.
  A. S. DEPARTMENT (Ed.).
- 641 LBS. 2013. *Household Survey 2013*. Lagos Welfare and Service Delivery Survey, Lagos
   642 Nigeria: Lagos Bureau of Statistics.
- LIAO, C.-M. and W.-C. LUO. 2005. Use of temporal/seasonal-and size-dependent bioaerosol
   data to characterize the contribution of outdoor fungi to residential exposures.
   *Science of the Total Environment*, **347**(1-3), pp.78-97.
- 646 LIAUDET, L., C. SZABO, O. V. EVGENOV, K. G. MURTHY, P. PACHER, L. VIRÁG, J. G. MABLEY, A.
  647 MARTON, F. G. SORIANO and M. Y. KIROV. 2003. Flagellin from gram-negative
  648 bacteria is a potent mediator of acute pulmonary inflammation in sepsis. *Shock*,
  649 **19**(2), pp.131-137.

- LIS, D. O., K. ULFIG, A. WLAZŁO and J. S. PASTUSZKA. 2004. Microbial air quality in offices at
   municipal landfills. *Journal of occupational and environmental hygiene*, 1(2), pp.62 68.
- LIU, Z., B. HU, L. WANG, F. WU, W. GAO and Y. WANG. 2015. Seasonal and diurnal variation
   in particulate matter (PM 10 and PM 2.5) at an urban site of Beijing: analyses from a
   9-year study. *Environmental Science and Pollution Research*, 22(1), pp.627-642.
- MACHER, J. M. 1989. Positive-hole correction of multiple-jet impactors for collecting viable
   microorganisms. *American Industrial Hygiene Association Journal*, **50**(11), pp.561 568.
- MAŁECKA-ADAMOWICZ, M., J. KACZANOWSKA and W. DONDERSKI. 2007. The Impact of a
   Landfill Site in Żółwin--Wypaleniska on the Microbiological Quality of the Air. *Polish* Journal of Environmental Studies, **16**(1).
- MASON, C. M. and S. NELSON. 2005. Pulmonary host defenses and factors predisposing to
   lung infection. *Clinics in chest medicine*, **26**(1), pp.11-17.
- MOORE, M., P. GOULD and B. S. KEARY. 2003. Global urbanization and impact on health.
   *International journal of hygiene and environmental health*, **206**(4), pp.269-278.
- ODEWABI, A. O., O. A. OGUNDAHUNSI, A. A. ODEWABI, K. S. ORITOGUN and M. EKOR. 2013.
   Adenosine deaminase activity and immunoglobulin levels as potential systemic
   biomonitors of occupational hazards and health status in municipal solid waste
   management workers. *Environmental toxicology and pharmacology*, **35**(1), pp.1-12.
- ODEYEMI, A. 2012. Antibiogram Status of Bacterial Isolates from Air Around Dumpsite of
   Ekiti State Destitute Centre at Ilokun, Ado-Ekiti, Nigeria. *Journal of Microbiology Research*, 2(2), pp.12-18.
- 673OGDEN, E. C., J. V. HAYES, RAYNOR and G. S. 1969. Diurnal patterns of pollen emission in674Ambrosia, Phleum, Zea, and Ricinus. American Journal of Botany, 56(1), pp.16-21.
- 675 OGWUELEKA, T. 2009. Municipal solid waste characteristics and management in Nigeria.
   676 Journal of Environmental Health Science & Engineering, 6(3), pp.173-180.
- OLORUNFEMI, F. 2011. Landfill development and current practices in Lagos metropolis,
   Nigeria. Journal of Geography and Regional Planning, 4(12), p656.
- 679 OYEKU, O. and A. ELUDOYIN. 2010. Heavy metal contamination of groundwater resources in
   a Nigerian urban settlement. *African Journal of Environmental Science and Technology*, 4(4).
- PERSOONS, R., S. PARAT, M. STOKLOV, A. PERDRIX and A. MAITRE. 2010. Critical working
  tasks and determinants of exposure to bioaerosols and MVOC at composting
  facilities. *International journal of hygiene and environmental health*, 213(5), pp.338347.
- 686 QUANJER, P. H., G. TAMMELING, J. COTES, O. PEDERSEN, R. PESLIN and J. YERNAULT. 1993.
   687 Lung volumes and forced ventilatory flows. 6(16), pp.5-40.
- RAY, M., G. MUKHERJEE, S. ROYCHOWDHURY and T. LAHIRI. 2004. Respiratory and general
   health impairments of ragpickers in India: a study in Delhi. *International Archives of Occupational and Environmental Health*, **77**(8), pp.595-598.
- RAY, M. R., S. ROYCHOUDHURY, G. MUKHERJEE, S. ROY and T. LAHIRI. 2005. Respiratory and
  general health impairments of workers employed in a municipal solid waste disposal
  at an open landfill site in Delhi. *International Journal of Hygiene and Environmental Health*, 208(4), pp.255-262.
- RAY, M. R., S. ROYCHOUDHURY, S. MUKHERJEE, S. SIDDIQUE, M. BANERJEE, A. AKOLKAR, B.
   SENGUPTA and T. LAHIRI. 2009. Airway Inflammation and Upregulation of. BETA. 2

697 Mac-1 Integrin Expression on Circulating Leukocytes of Female Ragpickers in India. 698 Journal of occupational health, **51**(3), pp.232-238. 699 REED, H. E. and B. U. MBERU. 2014. Capitalizing on Nigeria's demographic dividend: reaping 700 the benefits and diminishing the burdens. Etude de la population africaine = African 701 *population studies*, **27**(2), pp.319-330. 702 REINTHALER, F. F., E. MARTH, U. EIBEL, U. ENAYAT, O. FEENSTRA, H. FRIEDL, M. KÖCK, F. P. 703 PICHLER-SEMMELROCK, G. PRIDNIG and R. SCHLACHER. 1997. The assessment of 704 airborne microorganisms in large-scale composting facilities and their immediate 705 surroundings. Aerobiologia, 13(3), pp.167-175. 706 RICARD, J.-D. 2003. Are we really reducing tidal volume—And should we? : Am Thoracic Soc. 707 SANDELOWSKI, M. 2000. Focus on research methods combining qualitative and quantitative 708 sampling, data collection, and analysis techniques. Research in nursing & health, 23, 709 pp.246-255. 710 SANGKHAM, S., P. SAKUNKOO and C. ARUNLERTAREE. 2014. Concentrations of airborne 711 fungi from dumping landfill a case study with a Khon Kaen municipality site, Khon Kaen Province. Journal for Public Health Research-วารสาร วิจัย สาธารณสุข ศาสตร์ มหาวิทยาลัย ขอนแก่น, 712 713 7(2). 714 SCHLOSSER, O., A. HUYARD, D. RYBACKI and Z. DO QUANG. 2012. Protection of the vehicle 715 cab environment against bacteria, fungi and endotoxins in composting facilities. 716 Waste management, **32**(6), pp.1106-1115. 717 SCHLOSSER, O., S. ROBERT and C. DEBEAUPUIS. 2016. Aspergillus fumigatus and mesophilic 718 moulds in air in the surrounding environment downwind of non-hazardous waste 719 landfill sites. International journal of hygiene and environmental health, 219(3), 720 pp.239-251. 721 SHAIKH, S., A. A. NAFEES, V. KHETPAL, A. A. JAMALI, A. M. ARAIN and A. YOUSUF. 2012. 722 Respiratory symptoms and illnesses among brick kiln workers: a cross sectional study 723 from rural districts of Pakistan. BMC Public Health, **12**(1), p999. 724 SHEPPARD, D. C., G. RIEG, L. Y. CHIANG, S. G. FILLER, J. E. EDWARDS and A. S. IBRAHIM. 725 2004. Novel inhalational murine model of invasive pulmonary aspergillosis. 726 Antimicrobial agents and chemotherapy, **48**(5), pp.1908-1911. 727 SMETS, W., S. MORETTI, S. DENYS and S. LEBEER. 2016. Airborne bacteria in the 728 atmosphere: presence, purpose, and potential. Atmospheric Environment, 139, 729 pp.214-221. 730 SRIVASTAVA, V., S. A. ISMAIL, P. SINGH and R. P. SINGH. 2015. Urban solid waste 731 management in the developing world with emphasis on India: challenges and 732 opportunities. Reviews in Environmental Science and Bio/Technology, 14(2), pp.317-733 337. 734 STAGG, S., A. BOWRY, A. KELSEY and B. CROOK. 2010. Bioaerosol emissions from waste 735 composting and the potential for workers' exposure. Health and Safety Executive 736 Research report, 786. 737 THIRARATTANASUNTHON, P., W. SIRIWONG, M. ROBSON and M. BORJAN. 2012. Health risk 738 reduction behaviors model for scavengers exposed to solid waste in municipal dump 739 sites in Nakhon Ratchasima Province, Thailand. Risk management and healthcare 740 *policy*, **5**, p97. 741 THOMAS, R. J. 2013. Particle size and pathogenicity in the respiratory tract. Virulence, 4(8), 742 pp.847-858.

- TISCH ENVIRONMENTAL INC. 2018. *Viable Andersen Cascade Impactors* [online]. [Accessed
   28th March]. Available from: <u>www.tisch-env.com</u>.
- 745 UN-HABITAT. 2009. *Global Report on Human settelements*. Planning Sustainable Cities,
   746 (GRHS/09/FS1). Nirobi Kenya: United Nations Human Settlements Programme (UN 747 HABITAT).
- UNEP. 2005. Closing an open dumpsite and shifting from open dumping to controlled
   dumping and to sanotary landfilling In: U. N. E. PROGRAMME (Ed.) Nairobi Kenya:
   UNEP Division of Technology, Industry and Economics (DTIE).
- UNEP. 2015. DTI /1957/JA. *Global Waste Management Outlook*. GWMO, United Nations
   Environment Programme.
- VAN KAMPEN, V., F. HOFFMEYER, A. DECKERT, B. KENDZIA, S. CASJENS, H. NEUMANN, M.
  BUXTRUP, E. WILLER, C. FELTEN and R. SCHÖNEICH. 2016. Effects of bioaerosol
  exposure on respiratory health in compost workers: a 13-year follow-up study. *Occup Environ Med*, **73**, pp.829-837.
- WÉRY, N. 2014a. Bioaerosols from composting facilities—a review. *Frontiers in cellular and infection microbiology*, 4, p42.
- WÉRY, N. 2014b. Bioaerosols from composting facilities—a review. *Frontiers in cellular and infection microbiology*, 4.
- WHO. 1991. Sample size determination in health studies: a practical manual. Epidemiology
   and Statistics. Geneva World Health Organization.
- WMO. 2017. World Weather Information Service [online]. [Accessed 10th March]. Available
   from: <u>http://worldweather.wmo.int/en/city.html?cityId=258</u>.
- XU, Z. and M. YAO. 2013. Monitoring of bioaerosol inhalation risks in different environments
   using a six-stage Andersen sampler and the PCR-DGGE method. *Environmental monitoring and assessment*, **185**(5), pp.3993-4003.
- YOSHIDA, M. and J. WHITSETT. 2004. Interactions between pulmonary surfactant and
   alveolar macrophages in the pathogenesis of lung disease. *Cellular and molecular biology (Noisy-le-Grand, France)*, **50**, pp.OL639-48.
- 771
- 772