



UNIVERSITY OF LEEDS

This is a repository copy of *Geospatial visualisation of food contaminant distributions: Polychlorinated naphthalenes (PCNs), potentially toxic elements (PTEs) and aflatoxins.*

White Rose Research Online URL for this paper:
<http://eprints.whiterose.ac.uk/146257/>

Version: Accepted Version

Article:

Zhihua, L, Gong, Y, Holmes, M orcid.org/0000-0002-6819-1048 et al. (4 more authors) (2019) Geospatial visualisation of food contaminant distributions: Polychlorinated naphthalenes (PCNs), potentially toxic elements (PTEs) and aflatoxins. *Chemosphere*, 230. pp. 559-566. ISSN 0045-6535

<https://doi.org/10.1016/j.chemosphere.2019.05.080>

© 2019 Published by Elsevier Ltd. This manuscript version is made available under the CC-BY-NC-ND 4.0 license <http://creativecommons.org/licenses/by-nc-nd/4.0/>.

Reuse

This article is distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs (CC BY-NC-ND) licence. This licence only allows you to download this work and share it with others as long as you credit the authors, but you can't change the article in any way or use it commercially. More information and the full terms of the licence here: <https://creativecommons.org/licenses/>

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

1 Geospatial Visualisation of Food Contaminant Distributions: 2 Polychlorinated Naphthalenes (PCNs), Potentially Toxic Elements 3 (PTEs) and Aflatoxins

4 ¹Li Zhihua, ²Yunyun Gong, ²Mel Holmes*, ²Xiaoxi Pan, ¹Yiwei Xu, ¹Xiaobo Zou, ³Alwyn R.
5 Fernandes

6 ¹ *School of Food and Biological Engineering, Jiangsu University, Zhenjiang 212013, China.*

7 ² *School of Food Science and Nutrition, University of Leeds, Leeds LS2 9JT, UK.*

8 ³ *School of Environmental Sciences, University of East Anglia, Norwich NR4 7TJ, UK*

9 * Corresponding author: Mel Holmes, prcmjh@leeds.ac.uk

10 **Abstract:** Large volume of multidimensional data can be summarised, both in terms of
11 tabulated statistics, and as graphic geospatial visualisations. The latter approach allows rapid
12 interpretation and communication of complex information to stake-holders such as regulators,
13 risk assessors and policy makers. In the main study on polychlorinated naphthalene (PCN),
14 individual samples representing different edible fish species were analysed from around the
15 UK. PCNs were observed in all samples with nearly all of the twelve measured congeners being
16 detected. Summed congener concentrations ranged from 0.7 ng/kg ww (turbot) to 265 ng/kg
17 ww (sprats). The highest contamination levels were recorded for sprats and mackerel with mean
18 summed concentrations of 67 ng/kg ww and 68 ng/kg ww respectively. Two ancillary studies,
19 on potentially toxic elements (PTEs) in crabs from China and aflatoxin in children's blood
20 from Tanzania, demonstrate the wide applicability of this approach. The PTE contents in crab
21 showed strong dependence on the tested tissues and elements, and crabs from Tai and
22 Yangcheng Lakes showed obviously higher PTE levels than the other lakes. Geospatial
23 distribution of the aflatoxin biomarker AF-alb in children's serum from 3 locations showed

24 how individual anthropometric or socio-economic data reveals the relationship between family
25 size, socio-economic score and magnitude of serum aflatoxin levels. In addition to facilitating
26 the flow of interpreted data to stakeholders, these techniques can direct the formulation of risk
27 mitigation activities and help with the identification of data gaps. When combined with
28 hierarchical cluster analyses, correlations within the data can also be predicted.

29 **Keywords:** PCN-TEQ, fish, spatial analysis, visualization, risk mitigation, hierarchical
30 clustering

31 **1. Introduction**

32 The visualisation of the spatial distribution of contaminant concentrations in foods
33 sampled across large areas allows the rapid and efficient communication of information for
34 regulators, risk assessors and managers. In particular, this mode of presentation of geospatial
35 intelligence is especially relevant when multidimensional data is involved, where the sheer
36 volume of generated data makes interpretation time-consuming and difficult. When combined
37 with the outputs from the geospatial analysis of contaminant distributions in regions or
38 countries, visualisation provides a powerful tool for the study of risk identification and
39 establishment. It can also provide rapid and effective representation of affected or at-risk areas
40 which aids the development of mitigation activities. Additionally, the technique can be
41 complemented with hierarchical cluster analyses to allow the prediction of correlations within
42 data sets.

43 Some web-based mapping resources such as Google Maps (Google Maps 2016) provide
44 a powerful platform for the visualisation of geographical data. These platforms can be
45 developed to efficiently represent and explore the complex inter-relationships between food
46 contaminant occurrences, e.g. in different edible fish species or in different crop producing
47 regions, as demonstrated in this paper. Using an interactive webpage, specific aspects derived
48 from occurrence data can be represented in location and magnitude by different coloured circles

49 or radiation patterns corresponding to their GPS position and to represent different species for
50 example. The extent of contamination or the magnitude of the occurrence can be superimposed
51 on this data by varying the radii of circles or the offset distances for radiation patterns.

52 Polychlorinated naphthalenes (PCNs) are legacy industrial chemicals widely used as
53 dielectrics and flame retardants in the 20th century and are now recognised as persistent organic
54 pollutants (POPs), as evidenced by their listing in Annex A and C of the Stockholm Convention.
55 Environmentally, they occur ubiquitously with reports of contamination of biota, sediments
56 and air, both in temperate as well as Polar Regions (Biddleman et al., 2010; Braune and Muir
57 2017). Biochemically, they demonstrate properties of persistence and high bioaccumulation
58 potential, coupled with a similarity in structural configuration to polychlorinated
59 dibenzodioxins and furans (PCDD/Fs). This latter property is related to human health concerns
60 as many PCN congeners have been reported to contribute to dioxin-like toxicity (Villeneuve et
61 al., 2000; Falandysz et al., 2014; Fernandes et al., 2017) and elicit different effects such as
62 mortality, embryotoxicity, hepatotoxicity, dermal lesions, teratogenicity and carcinogenicity
63 (Behnisch et al., 2003; Falandysz et al., 2014). A recent review of their occurrence in human
64 tissue and foods (Fernandes et al., 2017) show higher levels in fish relative to other foods.

65 A number of trace elements, particularly heavy metals, have long been recognised as
66 potentially toxic elements (PTEs). Nutritionally, some trace elements, such as selenium, zinc,
67 copper, etc. are essential to health but may be toxic at high levels of dietary intake.
68 Environmental sources are the main contributors to contamination of food which is the major
69 source of the overall exposure of consumers to PTEs, although other routes may also be
70 significant (e.g. oral exposure via drinking water, occupational exposure by inhalation). Certain
71 food groups naturally accumulate some elements (e.g. fish and shellfish are known to
72 accumulate arsenic and mercury and cereals can accumulate cadmium) and consequently
73 contain high concentrations of these elements compared to other foods. PTEs may enter marine

74 and aquatic environments and bio-accumulate in species at any point during growth and
75 harvesting. Chinese mitten crab (*Eriocheir sinensis*) is widely consumed in China because of
76 its unique aroma and umami taste, but has also been reported (Xu et al., 2016; Shou et al.,
77 2012) to be at high risk of PTE accumulation with the tissues of crabs exhibiting diverse PTE
78 (e.g. lead and cadmium) bioaccumulation due to feeding behaviour and environmental factors.

79 Aflatoxins are contaminants that are produced by the moulds, *Aspergillus flavus* and
80 *Aspergillus parasiticus*. Aflatoxin B1 is recognised as a genotoxic carcinogen and is the most
81 potent of the fourteen aflatoxins that are natural products of these fungi. Within crops such as
82 maize and groundnuts, the growth of these moulds is promoted by poor storage conditions
83 combined with ambient humidity and temperature. Although hepatocellular carcinogenesis is
84 generally recognised as the most lethal biological effect, the stunting of growth in childhood
85 (affecting 165 m. children world-wide) that occurs in some countries in the African continent,
86 may show an association with dietary aflatoxin occurrence (Gong et al., 2002; 2004). High
87 levels of childhood aflatoxin exposure were observed during a study on three regions of
88 Tanzania in 2010 (Shirima et al., 2013; 2015)

89 This paper focusses on the occurrence and geospatial mapping of these contaminants to
90 provide effective visualisation of their distribution within geographical regions. The main focus
91 of this work is on the distribution of PCN contamination in fish, but the applicability of this
92 approach is also demonstrated for PTEs and aflatoxins, in different geographical locations. For
93 PCNs, occurrence is mapped for different edible marine fish species from locations in the North
94 Atlantic Ocean. PTE contents of three types of tissues (muscle tissue, hepatopancreas and
95 sexual organs) of Chinese mitten crabs were determined, and mapped for samples taken from
96 5 lakes in Jiangsu province, China. For Aflatoxin B1, visualisation of human exposure as
97 measured in blood AF-alb biomarkers in 3 rural locations in Tanzania was superimposed with
98 related anthropometric data to examine correlations between these variables and exposure.

99 2. Experimental

100 2.1 Polychlorinated Naphthalene study

101 In the main study on PCNs, 75 samples of Sea Bass, Herring, Mackerel, Mullet and
102 Sardine were identified, as part of a larger study investigating other contaminants such as
103 brominated flame retardants, PCDD/Fs, etc. Samples were collected in major fishing areas,
104 mainly but not exclusively, from waters around the UK, Ireland and Northern France. In
105 general, selected fish sizes approximated to those that are commercially sold for consumption.
106 The following PCN congeners were included in this study - PCN-52/60, 53, 66/67, 68, 69,
107 71/72, 73, 74, & 75.

108 The analytical methodology for the determination of PCN concentrations in fish species
109 has been described elsewhere (Fernandes et al., 2010). In brief, sample aliquots were fortified
110 with ¹³C-labelled PCN analogues of target compounds and extracted using mixed organic
111 solvents. PCNs were chromatographically fractionated from potential interferants such as
112 PCBs, on an activated carbon column and further purified using adsorption chromatography
113 on alumina. Analytical measurement was carried out using high resolution gas chromatography
114 coupled to high resolution mass spectrometry (HRGC-HRMS). Additional control was
115 provided by the inclusion of procedural blanks and a reference material.

116 The quality control criteria used for evaluating data, and method performance parameters
117 have been reported before (Fernandes et al., 2010). There are no available reference materials
118 (RMs) specific to PCNs, but the same reference material that is used for PCDD/F and PCB
119 analysis (cod liver oil), was analysed during the course of this work with results showing good
120 consistency and agreement with established values. The reporting limits (LOQ quoted as “<”)
121 for all analytes incorporate the relevant procedural blank and were derived using the current
122 guidance on LOQ estimation (European Commission, 2017).

123 **2.2. Potentially Toxic Elements in Crabs**

124 The sampling sites included 5 major lakes in Yangtze River Delta, China, i.e., Yangcheng
125 Lake, Tai Lake, Changdang Lake, Gucheng Lake and Hongze Lake. 8 crabs were collected
126 from each lake, rinsed with tap water, and dissected quickly to obtain the sexual gland,
127 hepatopancreas and muscle. The sample tissues were digested using a microwave-assisted
128 method (López et al., 2003), then Cadmium (Cd), Lead (Pb), Mercury (Hg) and Copper (Cu)
129 in the digested solutions were measured by inductively coupled plasma mass spectrometry
130 (F.J.Sanchez L., 2003) (ICP-MS, XSeries II, Thermo Fisher Scientific, USA). Blanks and QC
131 samples included and the LOQs ranged from 3-5 µg/kg.

132 **2.3 Aflatoxins in Serum**

133 The aflatoxin data is derived from a study (Shirima et al., 2015) that investigated 3 regions
134 in Tanzania (Iringa, Tabora and Kilimanjaro). Dietary intake data and anthropometric indices-
135 body weight, recumbent length, etc. were recorded. Validated methodology was used to
136 measure aflatoxin–albumin adducts (AF-alb), as biomarkers in blood plasma taken from 166
137 children. A description of this methodology and quality parameters has already been given
138 before (Chapot., 1991) but in brief, the tests included extraction of albumin, digestion of
139 protein, purification, and ELISA quantification of AF-alb adducts. Each batch of plasma was
140 analysed with three positive controls and one negative control. Samples were measured in
141 ELISA in quadruplicate on at least two occasions on separate days. The detection limit was 3
142 pg of aflatoxin-lysine equivalents per milligram albumin.

143 **2.4 Geo-spatial mapping**

144 For the geo-spatial mapping, an interactive webpage based on Google Maps was used (Li
145 et al., 2018) to integrate sample location (GPS), with contaminant concentration data. The
146 webpage primarily consisted of user selective control fields such as individual or summed

147 contaminant, species, occurrence scaling, colour legend of selected species/contaminants and
148 a help field. Each sample was presented with a circle centred at its GPS coordinate location.
149 The magnitude of occurrence was indicated by the radius of each circle (ng/kg for PCNs; µg/kg
150 for PTEs; pg/mg for AF-alb), which can be simultaneously colour coded to variables such as
151 socio-economic status. The design allows multiple species or congeners to be selected,
152 differentiated by colour coding and line type (solid, dashed, dotted) as applied to relevant circle
153 representations. The aflatoxin study used plasma AF-alb biomarker data and/or contaminant
154 levels within maize/other foods at the designated sampling region together with hazard
155 indicators such as child weight z-scores. This approach offers efficient data exploration without
156 full statistical analyses.

157 With the volumes of data generated in this study, hierarchical clustering (HC) analysis can
158 complement the above approach, a feature which would allow the user to investigate if certain
159 locations or species/individuals are susceptible to concentration levels of concern. HC is a
160 typical algorithm to analyse the similarities (or dissimilarities) of objects in variable space
161 (Smoliński et al., 2002). In order to have a better understanding of the geographical
162 distribution, or independence of occurrence of the congeners, HC was employed in the PCN
163 study e.g. by dividing the fish samples into 3 clusters with the PCN congener concentration as
164 input variables to investigate if correlations existed between the species and spatial locations.
165 HC analysis was performed using R language with the Ward method as the amalgamation rule
166 and Euclidean distance as metric (Smoliński et al., 2002).

167 **3. Results and Discussion**

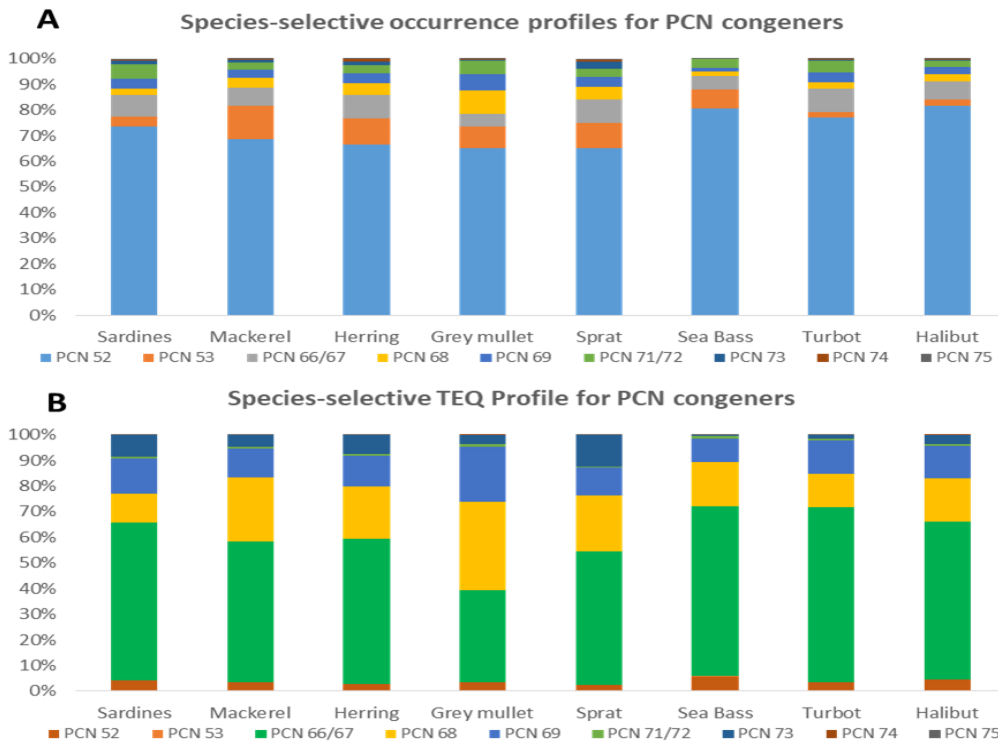
168 The raw data from the PCN study is very large and has been presented in a sponsor report
169 (Fernandes et al., 2015). Raw data from the mitten crab study is provided in the supplementary
170 information (SI). Descriptive statistics are summarised in Tables 3.1 and 3.2, by analyte type.

171 3.1 PCN

172 PCN concentrations are reported in ng/kg wet weight (ww) as per convention. Congeners
173 were selected for measurement based on toxicological significance as described in earlier
174 studies (Fernandes et al 2010; Falandysz et al., 2014). From the point of view of food safety,
175 and given the known AhR receptor active responses (Falandysz et al., 2014; Fernandes et al.,
176 2017), these concentrations have also been converted to toxic equivalents (TEQs) using a
177 similar approach to PCDD/Fs and dioxin-like PCBs (Van den berg et al., 2006). The TEQ
178 values were calculated using relative potencies (REPs) as used in other studies (Fernandes et
179 al., 2010; 2011; 2017; Jongchu et al., 2018; Lili et al., 2018).

180 PCNs were detected in all of the samples analysed with variations in patterns of
181 occurrence and levels of concentrations depending on the species and location. Higher
182 concentration levels were recorded for penta- and hexa-chlorinated PCNs (e.g. PCNs 52, 53
183 and 66/67) with lowest levels occurring mainly for the octa-chlorinated PCN-75. The extent of
184 positive detection was high, at 86% of all measurements. The patterns of occurrence are
185 dominated by PCN-52 but the occurrence of other congeners varies widely depending on the
186 species as shown in Figure 3.1A (average pattern for each species, except for halibut which
187 was a single sample), whereas the absolute concentrations appeared to be more dependent on
188 the location of the sample. Summed concentrations (sum of twelve measured congeners)
189 showed a range extending from 0.7 ng/kg ww (turbot) to 265 ng/kg ww (sprats). Highest
190 concentrations were recorded for sprats and mackerel with mean concentrations of 67 ng/kg
191 ww and 68 ng/kg ww respectively. Given the large geographical range of the sampling region,
192 these observations are indicative, although other studies (Fernandes et al., 2009) appear to show
193 similar findings. However, other parameters such as size and the age of the fish, and seasonal
194 physiological changes (e.g. spawning periods, etc.) will undoubtedly influence the findings.

195



196

197 Figure 3.1 Average pattern (A), of PCN congener occurrence and (B), TEQ, based on species

198 Table 3.1. Statistical summary of PCN occurrence in various species of edible marine fish

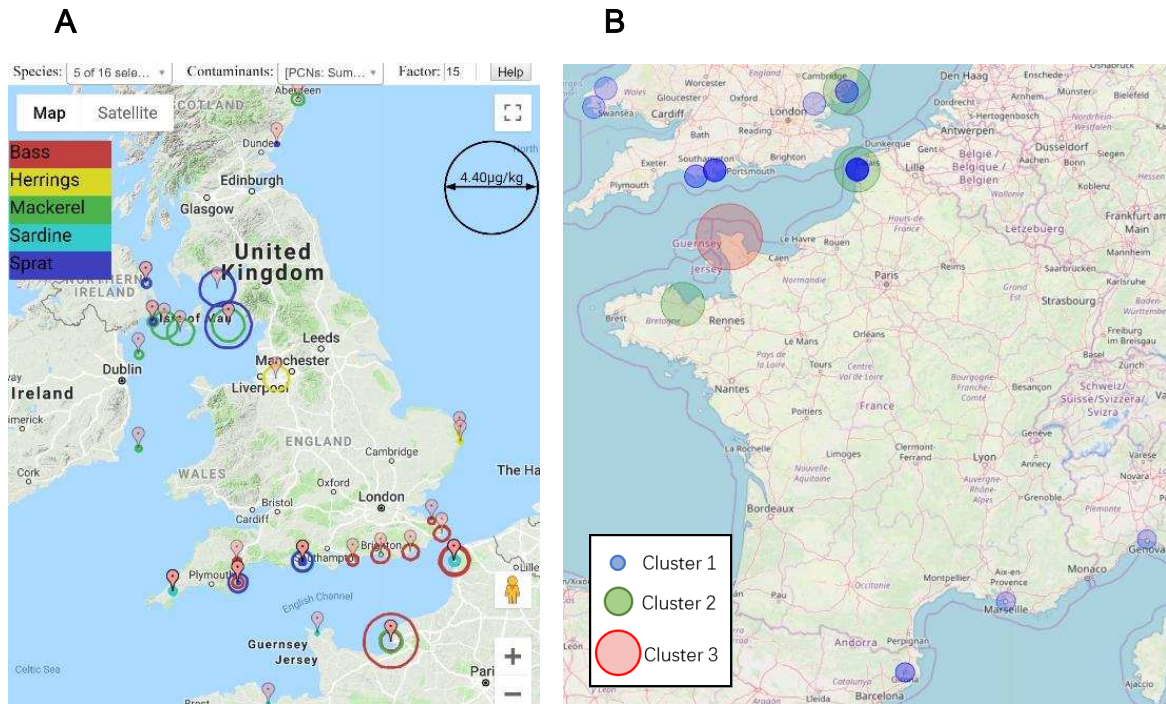
Species (N) & average fat		Sum	Sum	*PCN	Species (N) & average fat	Sum	Sum	*PCN	
		PCNs Lipid basis	PCNs whole weight	TEQ whole weight		PCNs Lipid basis	PCNs whole weight	TEQ whole weight	
		ng/kg fat	ng/kg w.w.				ng/kg fat	ng/kg w.w.	
Sardines (12)	MIN	22	5.4	0.004	Sprat (15)	MIN	138	29.4	0.014
	MEDIAN	172	16.6	0.007		MEDIAN	335	46	0.027
	MEAN	291	19.8	0.009		MEAN	680	66.5	0.044
	MAX	1215	63.1	0.031		MAX	2390	264.8	0.204
Mackerel (14)	MIN	112	10.1	0.002	Sea Bass (13)	MIN	302	14.2	0.004
	MEDIAN	451	50.5	0.024		MEDIAN	848	29.2	0.008
	MEAN	648	68	0.035		MEAN	999	29.4	0.01
	MAX	1654	243	0.17		MAX	3084	48.5	0.026
Herring (6)	MIN	141	18.3	0.009	Turbot (6)	MIN	132	0.7	<0.001
	MEDIAN	231	29.7	0.016		MEDIAN	246	3.5	0.002
	MEAN	431	38.7	0.024		MEAN	343	5.3	0.003
	MAX	1342	89.5	0.069		MAX	828	15.5	0.009
Grey mullet (9)	MIN	125	4.2	0.001	Halibut (1)		253	5.85	0.003
	MEDIAN	554	12.4	0.006					
	MEAN	1293	14.7	0.007					
	MAX	7572	33.5	0.014					

199 *PCN TEQ calculated using REP values from Fernandes et al. 2010 Env. Sci. Technol., 44, 3533–3538. Sum

200 PCNs = sum of 12 congeners

201 The variation in congener occurrence is reflected in the TEQ distribution (Figure 3.1B),
202 although here, the higher REPs of congeners PCN-66/67, PCN-68 and PCN-69 (Falandysz et
203 al., 2014) make them the largest contributors to the cumulative TEQ. The TEQ values for these
204 fish samples (Table 3.1), range from <0.001 to 0.2 ng/kg ww and correlate well with the
205 summed concentrations, with highest average TEQ concentrations for mackerel and sprats at
206 0.035 and 0.044ng/kg ww. This range of values is substantially lower than the corresponding
207 TEQ arising from PCDD/Fs and PCBs in fish, but it does contribute to the overall burden of
208 dioxin-like toxicity.

209 Preliminary geospatial mapping analysis showed that the variables that would provide the
210 most useful interpretation of the results were likely to be location, species and PCN congener
211 (including sum of congeners). The visualisation plots of the concentrations (Figure 3.2A)
212 revealed that although locations across the southern/eastern UK coasts and northern France
213 showed a majority of concentrations above the average for sum PCNs (39 ng/kg ww), the
214 highest PCN concentrations were recorded in samples of mackerel and sprats from the Irish
215 Sea. This distribution is confirmed when plots of the indicative PCN congeners are examined
216 e.g. PCN-52 as the most abundant congener measured in this study (Figure 3.3A), and PCN-
217 66/67 as the largest contributor to the PCN TEQ (Figure 3.3B). In terms of species, Figure 3.2A
218 also reveals that of the measured species, mackerel, sprats and sea-bass are the most susceptible
219 to PCN contamination. Considering the significant difference between species, analysis of
220 individual species was necessary to examine the spatial distribution of PCNs exposure. Using
221 grey mullet as an example, HC analysis was performed to determine the correlations between
222 grey mullet samples with PCN congener concentrations used as input variables. Figure 3.2B
223 clearly shows that cluster 2 and cluster 3, present higher level of PCNs and appeared along the
224 north coast of France.



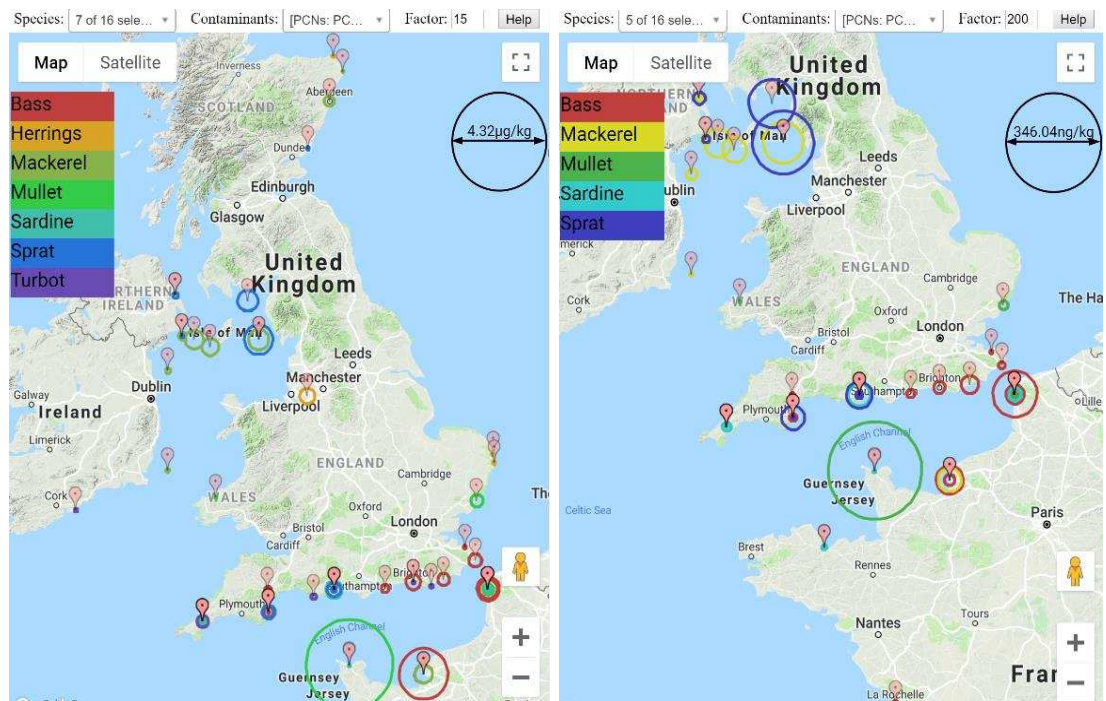
225

226

Figure 3.2. (A) Spatial distribution of sum PCN congeners for locations around the UK;

227

(B) HC analysis for PCN levels of Grey mullet



228

229

Figure 3.3. Spatial distribution of A. PCN-52 and B. PCN-66/67

230 3.2 Analysis and visualization of PTEs in Crab

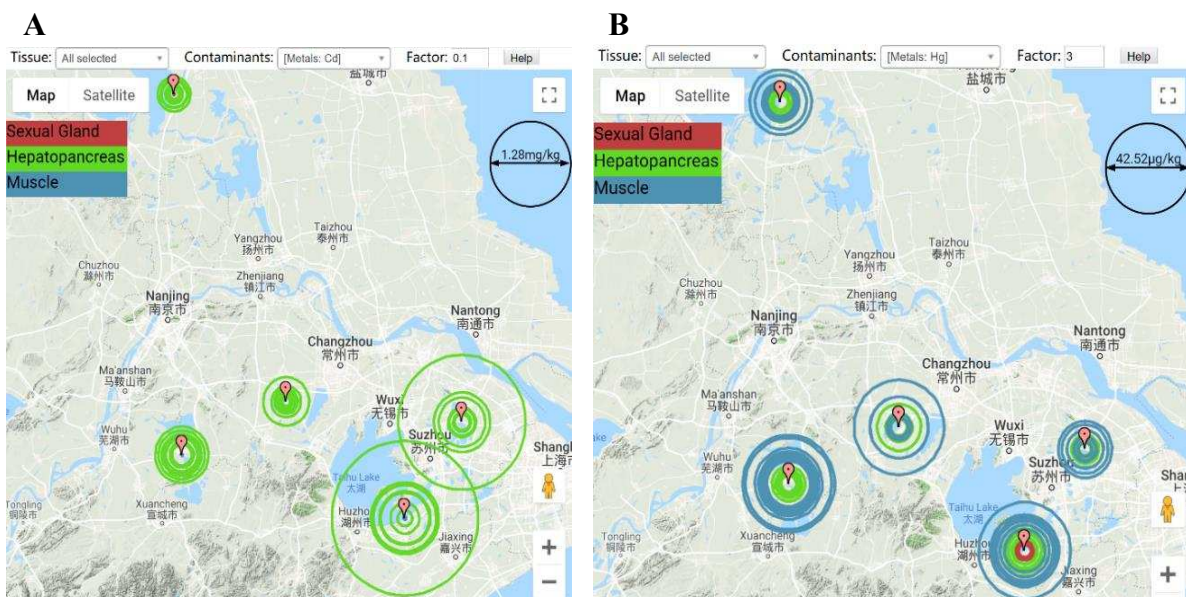
231 A statistical summary of the measured PTE concentrations in different crab tissues is
232 presented in Table 3.2. The relatively high occurrence of Cu in all samples follows a normal
233 distribution for fish and shellfish (De Mora et al., 2004) species but concentrations for some of
234 the other PTEs showed a strong dependency on the type of tissue. The geometric mean
235 concentration for Cd in the hepatopancreas was significantly higher than in the sexual gland or
236 muscle. Conversely, the mean concentration of Hg in muscle was approximately double the
237 concentration in the hepatopancreas and about 50-fold higher than in the sexual gland. There
238 was no significant difference observed for Pb in different tissue in terms of the geometric mean.
239 The maximum concentrations generally reflect the means apart from Hg in the sexual gland
240 which shows a large difference. Such characteristics of PTE occurrence in crab tissues can be
241 conveniently superimposed on the geographical locations of the samples using geospatial
242 mapping. In Figure 3.4A the distribution of Cd with location as a variable, immediately displays
243 the higher occurrence levels in the Tai and Yangcheng lakes relative to the other lakes. In this
244 plot, only Cd in the hepatopancreas is immediately apparent, mainly due to the higher
245 concentrations in this tissue (131 to 2382 $\mu\text{g}/\text{kg}$ with an average value of 579 $\mu\text{g}/\text{kg}$, compared
246 to e.g. <5.0 to 21 $\mu\text{g}/\text{kg}$ with average value of 2.4 $\mu\text{g}/\text{kg}$ in the sexual gland). Although the
247 immediate effect is to obscure the detail for the muscle and the sexual gland, this can easily be
248 overcome by using the scaling factor to expand the scale of the plot as shown in Figure 3.4B
249 which shows the distribution for Hg plotted at a much higher magnification. Here the highest
250 concentrations occur in the muscle tissue. The relatively high accumulation of Cd in
251 hepatopancreas was also observed by Ma et al., 2008 and Wang et al., 2008, who suggested
252 that this could be due to the hepatopancreas playing an important role in Cd detoxification in
253 this species (Hopkin and Nott, 1980).

254 Besides single-factor comparison, multiple factors including tissue, location and element,

255 could be compared at the same time with a radiation pattern. As shown in Figure 3.5, each
 256 vertex of the radiation pattern presents an element, the colour presents tissue part. Figure 3.5
 257 clearly shows that PTE occurrence in the crabs strongly depended on tissue type and individual
 258 PTEs, which is consistent with Table 3.2 and Figure 3.4. Spatially, crabs from Tai Lake and
 259 Yangcheng Lake showed higher level of PTE bio-accumulation than the other 3 lakes.

260 Table 3.2. Statistical summary of the measured PTE contents in different crab tissues

Tissue parts	Statistics	PTEs ($\mu\text{g}/\text{kg}$)			
		Cd	Pb	Hg	Cu
Hepatopancreas	Min.	131	14.6	<3.0	3687
	Median	461	40.9	11.5	25123
	Mean	579.	52.3	11.4	30424
	Max.	2382	210.2	25.2	79060
Sexual Gland	Min.	<5.0	12.9	<3.0	3996
	Median	<5.0	42.4	<3.0	19469
	Mean	2.4	59.0	0.4	24278
	Max.	21.0	277.1	9.2	59659
Muscle	Min.	<5.0	17.2	<3.0	7763
	Median	<5.0	32.1	23.6	15485
	Mean	3.9	50.2	21.7	16040
	Max.	42	349	47	31425



261

262

Figure 3.4. Spatial distribution of A. Cd and B. Hg for different tissues of crab



263

264

Figure 3.5. Radiation pattern of Cd, Hg and Pb for different tissues

265 3.3 Aflatoxin

266

267

268

269

270

271

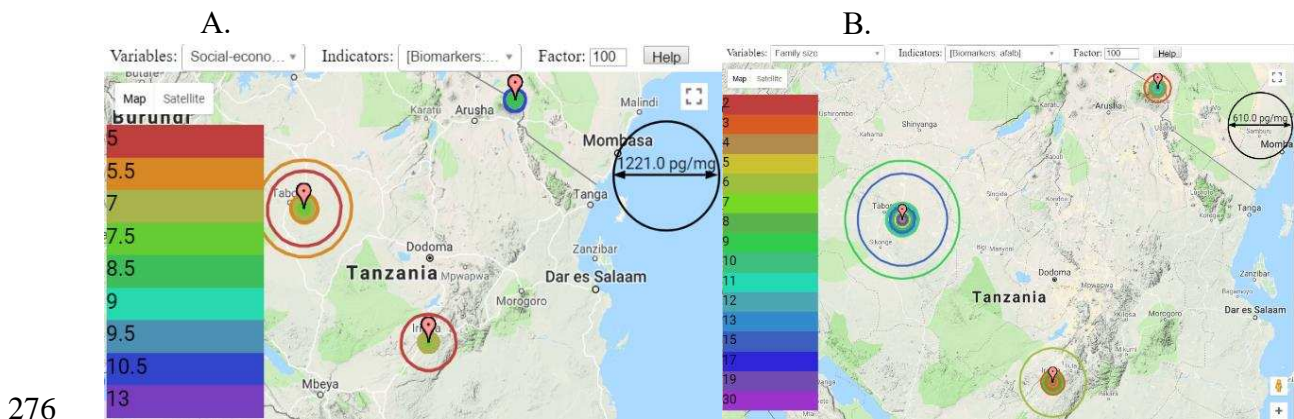
272

273

274

275

These preliminary results of aflatoxin analysis demonstrate the initial exploratory potential of the application. Geospatial analysis was applied to the distribution of the aflatoxin biomarker AF-alb in children's blood from 3 locations, to individual anthropometric or socio-economic data. The examples in Figures 3.6A and 3.6B show the magnitude of AF-alb content overlaid with socio-economic status and family size respectively and shows how potential associations for indicators may be explored. The Tabora region (located centre of Figures 3.6A & B) shows relatively stronger associations. Generally, the characteristic of poorer social economic state (lower score) show higher level of aflatoxin biomarker, which means more risk for aflatoxin exposure. Similarly, correlation to child hazard data (e.g. stunting) can be performed to evaluate biomarker to hazard indicators.



276
 277 Figure 3.6. Selected spatial distribution for the aflatoxin biomarker AF-alb in Tanzania with A.
 278 associated socio-economic status colour coding and B. family size colour-coding, respectively.

279 3.4 General Discussion with focus on advantages of geo-spatial visualisation

280 Figures 3.2-3.6 demonstrate rapid and effective geo-spatial visualisation of the large
 281 amount of data associated with these studies. Although spatially distributed data can also be
 282 summarised in tabular format, navigating listed numerical data can be time consuming and
 283 tedious when compared to the amount of information that is quickly apparent from the visual
 284 plots. Areas of higher contamination that may relate to relatively higher risk, or locations
 285 requiring attention, are immediately apparent from the plots and enable rapid qualitative
 286 assessment of potential associations within data. So for example, in the case of the PCN
 287 contamination of edible fish species, it is immediately apparent from the plots that fish from
 288 the northern reaches of the Irish Sea between Northern Ireland and South-West England, and
 289 the English Channel show a higher levels of PCNs. This also allows for possible geographical
 290 associations with sources to be made e.g. Commercial PCN sold as Seekay waxes were
 291 manufactured at Runcorn (Crookes and Howe, 1993) on the southern bank of the river Mersey
 292 which flows into the Irish Sea. In the case of the AF-alb data from Tanzania, the introduction
 293 of socio-economic data relating to the families of the children who provided blood samples, as
 294 a variable within the geospatial analysis, allowed potential associations with poorer socio-
 295 economic scores and larger family sizes to be correlated with higher AF-alb levels in the blood

296 (Figure 3.6). The plot also shows that the relatively affluent area around Kilimanjaro in the
297 north of the country was associated with smaller family sizes and lower AF-alb blood levels.

298 If sufficient data are available, more detailed analyses, e.g. HC, can be conducted or
299 alternatively, where data are lacking, this promotes more informed planning, e.g. where
300 additional research/data acquisition is required, and enables greater transparency across
301 stakeholders. These techniques and the resulting visualised data also allow interpreted
302 information to become more amenable to non-specialist stakeholders, e.g. risk assessors and
303 managers who are then empowered to direct the formulation of risk mitigation activities and
304 also help with the identification of data gaps. Another aspect of the techniques is the ability to
305 retrospectively add new types of data that may become available e.g. rates of crop failure or
306 any changes due to temporal or climatic conditions or spontaneous events.

307 **4. Conclusions**

308 The geo-spatial visualisation approach applied to the data presented here allows the rapid
309 and effective evaluation of complex and multidimensional data sets. Three diverse datasets on
310 PCNs, PTEs and aflatoxins were analysed demonstrating the versatility of the system. In the
311 main study on PCNs, correlations between location and species with PCN congeners was
312 demonstrated, showing that mackerel and sea-bass were the edible species that were most
313 susceptible to PCN contamination, and the Northern reaches of the Irish Sea and the English
314 Channel showed relatively higher levels of contamination than the other marine areas studied.
315 In two ancillary studies, data on potentially toxic elements (PTEs) in crabs from 5 lakes in
316 China and aflatoxin levels in maize from three regions of Tanzania, demonstrate the wide
317 applicability of this visualisation approach. The PTE contents in crab showed strong
318 dependence on the tested tissues and elements, and crabs from Tai Lake and Yangcheng Lake
319 showed obviously higher level of PTE accumulation than the other lakes. Geospatial
320 distribution of the aflatoxin biomarker AF-alb in children's blood from 3 locations in Kenya

321 showed how individual anthropometric or socio-economic data reveals the relationship
322 between family size, socio-economic score and magnitude of serum aflatoxin levels. In
323 addition to facilitating the flow of interpreted data to stakeholders, these techniques can help
324 direct the formulation of risk mitigation activities and also help identify gaps in data and
325 knowledge.

326 **Acknowledgements**

327 Data for the fish species was derived from work funded by the UK Food Standards
328 Agency. The study on crabs was funded by the National Natural Science Foundation of China
329 (31801631)

330 **Declarations of interest:** None

331 **References**

332 Behnisch, P., Hosoe, K., Sakai, S., 2003. Brominated dioxin-like compounds: in vitro assessment in
333 comparison to classical dioxin-like compounds and other polyaromatic compounds. *Environ. Int.* 29,
334 861-877.

335 Bidleman, T., Helm, P., Braune, B., Gabrielsen, G. 2010. Polychlorinated naphthalenes in polar
336 environments - a review. *Sci. Total Environ.* 408(15), 2919-35.

337 Braune, B., Muir, D. 2017. Declining Trends of Polychlorinated Naphthalenes in Seabird Eggs from
338 the Canadian Arctic, 1975-2014. *Environ. Sci. Technol.* 51(7), 3802-3808.

339 Chapot B., Wild CP. 1991. ELISA for quantification of aflatoxin-albumin adducts and their application
340 to human exposure assessment. *Techniques in Diagnostic Pathology, Vol 2.* San Diego, CA: Academic
341 Press, 135–155.

342 Crookes, M., Howe, D. 1993. Environmental hazard assessment: Halogenated naphthalenes; Report
343 TSD/13; Department of the Environment: London, 1993.

344 De Mora, S., Fowler, S., Wyse, E., Azemard, S. 2004. Distribution of heavy metals in marine bivalves,
345 fish and coastal sediments in the Gulf of Oman. *Marine Poll. Bull.* 49(5-6):410–24.

346 European Commission, 2017. Guidance document on the estimation of LOD and LOQ for
347 measurements in the field of contaminants in feed and food. Available at: [https://ec.europa.eu/jrc/en/publication/guidance-document-estimation-lod-and-loqmeasurements-field-](https://ec.europa.eu/jrc/en/publication/guidance-document-estimation-lod-and-loqmeasurements-field-contaminants-feed-and-food)
348 [contaminants-feed-and-food](https://ec.europa.eu/jrc/en/publication/guidance-document-estimation-lod-and-loqmeasurements-field-contaminants-feed-and-food).

350 Falandysz, J., Fernandes, A., Gregoraszczyk, E., Rose, M. 2014. The Toxicological Effects of
351 Halogenated Naphthalenes: A Review of Aryl Hydrocarbon Receptor-Mediated (Dioxinlike) Relative
352 Potency Factors. *J. Environ. Sci. Health*, 32, 239-272.

353 Fernandes, A., Smith, F., Petch N., Brereton N., Bradley E., Panton S., Carr, M., Rose, M., 2009.
354 Investigation into the levels of environmental contaminants in Scottish marine and freshwater fin fish
355 and shellfish. Report to the Food Standards Agency Scotland, Fera Report 2 FD 09/01.

356 Fernandes, A., Mortimer, D., Gem, M., Smith, F., Rose, M., Panton, S., Carr, M., 2010. Polychlorinated
357 Naphthalenes (PCNs): Congener specific analysis, occurrence in food and dietary exposure in the UK.
358 *Environ. Sci. Technol.* 44, 3533–3538.14.

359 Fernandes, A., Tlustos, C., Rose, M., Smith, F., Carr, M., Panton S. 2011. Polychlorinated Naphthalenes
360 (PCNs) in Irish Foods: Occurrence and Human Exposure. *Chemosphere*, 85, 322-328.

361 Fernandes, A., Rose, M., Smith, F., Panton, S., 2015. Geographical Investigation for chemical
362 contaminants in fish collected from UK and proximate marine waters. Report FD 15/04 to the UK Food
363 Standards Agency. London.

364 Fernandes A., Rose, M., Falandysz J. 2017. PCNs in Food and Humans. *Environ Int.* 104, 1-13

365 Fernandes, A., Mortimer, D., Holmes, M., Rose, M., Zhihua, L., Huang, X., Smith, F., Panton, S.,
366 Marshall, L. 2018. Occurrence and Spatial Distribution of Chemical Contaminants in Edible Fish
367 Species Collected from UK and Proximate Marine Waters. *Env. Int.*, 114, 219-230.

368 F.J.Sanchez, L., M.D.Gil, G., N.P.Sanchez, M., J.L.Martinez, V. 2003. Determination of heavy metals
369 in crayfish by ICP-MS with a microwave-assisted digestion treatment. *Ecotox. Environ. Safety.* 45(2),
370 223-228.

371 Gong, Y., Cardwell, K., Hounsa, A., Egal, S., Turner, P., Hall, A., et al. 2002. Dietary aflatoxin exposure
372 and impaired growth in young children from Benin and Togo: cross sectional study. *Brit. Medical J.*
373 325(7354), 20–1.

374 Gong, Y., Hounsa, A., Egal, S., Turner, P., Sutcliffe, A., Hall, A., et al. 2004. Postweaning exposure to
375 aflatoxin results in impaired child growth: a longitudinal study in Benin, West Africa. *Environ. Health*
376 *Perspect.* 112(13), 1334–8.

377 Google Maps, 2016. Available at: <https://maps.google.com>

378 Hopkin, P., Nott, J. A. 1980. Studies on the digestive cycle of the shore crab *Carcinus maenas* (L.) with
379 special reference to the B cells in the hepatopancreas. *J. Mar. Biol. Assoc. U. K .*, 60, 891 - 907.

380 Jongchu, K., Eun-Su, S., Sung-Deuk, C., Jiping, Z., Yoon-Seok, C. 2018. Polychlorinated naphthalenes
381 (PCNs) in seafood: Estimation of dietary intake in Korean population. *Sci. Tot. Environ.* 624, 40-47.

382 Kannan, K., Imagawa, T., Blankenship, A L., Giesy, J P. 1998. Isomer-Specific Analysis and Toxic
383 Evaluation of Polychlorinated Naphthalenes in Soil, Sediment, and Biota Collected near the Site of a
384 Former Chlor-Alkali Plant. *Environ. Sci. Technol.* 32 (17), 2507-2514.

385 Li, Z., Panton, S., Marshall, L., Fernandes, A., Rose, M., Smith, F., Holmes, M. 2018. *Chemosphere.*
386 195, 727-734.

387 Lili, C., Shasha, W., Lirong, G., Huiting, H., Dan, X., Lin, Q., Wenbin, L. 2018. Concentrations and
388 trophic magnification of polychlorinated naphthalenes (PCNs) in marine fish from the Bohai coastal
389 area, China. *Environ. Poll.* 234, 876-884.

390 López, F., Garcia, M., Morito, N., Vidal, J. 2003. Determination of heavy metals in crayfish by ICP-
391 MS with a microwave-assisted digestion treatment. *Ecotoxicol. Environ. Safety.* 54(2), 223-228.

392 Ma, W., Wang, L., He, Y., Yan, Y. 2008. Tissue-specific cadmium and metallothionein levels in
393 freshwater crab *Sinopotamon henanensis* during acute exposure to waterborne cadmium. *Environ.*
394 *Toxicol.* 23. 393-400.

395 Shirima, C., Kimanya, M., Kinabo, L., Routledge, M., Srey, C., Wild, C., et al. 2013. Dietary exposure
396 to aflatoxin and fumonisin among Tanzanian children as determined using biomarkers of exposure.
397 *Mol. Nutr. Food Res.* 57(10), 1874–1881.

398 Shirima, C., Kimanya, M., Srey, C., Kinabo, J., Humpf, H., Wild, C., Gong, Y. 2015. A Prospective
399 Study of Growth and Biomarkers of Exposure to Aflatoxin and Fumonisin during Early Childhood in
400 Tanzania *Env. Health Persp.* 123(2) 173-178.

401 Shou Z., Chenghong F., Weimin Q., Xiaofeng C., Junfeng N., Zhenyao S. 2012. Role of living
402 environments in the accumulation characteristics of heavy metals in fishes and crabs in the Yangtze
403 River Estuary, China. *Marine Pollution Bulletin.* 64 (6), 1163-1171.

404 Smoliński, A., Walczak, B., Einax, J. 2002. Hierarchical clustering extended with visual complements
405 of environmental data sets. *Chemometrics Intell. Lab. Sys.* 64, 45-54.

406 Van den Berg, M., Birnbaum, L., Denison, M., De Vito, M., Farland, W. et al. 2006. The 2005 World
407 Health Organization Re-evaluation of Human and Mammalian Toxic Equivalency Factors for Dioxins
408 and Dioxin-like Compounds. *Toxicol. Sci.* 93, 223–241.

409 Villeneuve, D., Khim, J., Kannan, K., Falandysz, J., Blankenship, A., Nikiforov, V., Giesy, J. 2000.
410 Relative potencies of individual polychlorinated naphthalenes to induce dioxin-like response in fish and
411 mammalian in vitro bioassays. *Arch. Environ. Contam. Toxicol.* 39, 273–281.

412 Wang, L., Yan, B., Liu, N., Li, Y., Wang, Q. 2008. Effects of cadmium on glutathione synthesis in
413 hepatopancreas of freshwater crab, *Sinopotamon yangtsekiense*. *Chemosphere*, 74(1), 51-56.

414 Xu Y., Zhang W., Shi J., Zou X., Li Z., Zhu Y. 2016. Microfabricated interdigitated Au electrode for
415 voltammetric determination of lead and cadmium in Chinese mitten crab (*Eriocheir sinensis*). *Food*
416 *Chemistry*, 201, 190-196.