



UNIVERSITY OF LEEDS

This is a repository copy of *A Bayesian belief network framework to predict SOC dynamics of alternative management scenarios*.

White Rose Research Online URL for this paper:
<http://eprints.whiterose.ac.uk/127167/>

Version: Accepted Version

Article:

Dal Ferro, N, Quinn, CH orcid.org/0000-0002-2085-0446 and Morari, F (2018) A Bayesian belief network framework to predict SOC dynamics of alternative management scenarios. *Soil and Tillage Research*, 179. pp. 114-124. ISSN 0167-1987

<https://doi.org/10.1016/j.still.2018.01.002>

(c) 2018, Elsevier Ltd. This manuscript version is made available under the CC BY-NC-ND 4.0 license <https://creativecommons.org/licenses/by-nc-nd/4.0/>

Reuse

This article is distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs (CC BY-NC-ND) licence. This licence only allows you to download this work and share it with others as long as you credit the authors, but you can't change the article in any way or use it commercially. More information and the full terms of the licence here: <https://creativecommons.org/licenses/>

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

1 **A Bayesian Belief Network framework to predict SOC dynamics of alternative management**
2 **scenarios**

3 N. Dal Ferro^{a,*}, C. Quinn^b, F. Morari^a

4

5 ^a Department of Agronomy Food Natural resources Animals Environment, University of Padova,
6 Legnaro (PD), Italy

7 ^bSustainability Research Institute, University of Leeds, Leeds, United Kingdom

8 * Corresponding author: Tel. +39 (0)49 827 2818. Email address: nicola.dalferro@unipd.it (N. Dal
9 Ferro)

10

11

12

13

14

15

16

17

18

19

20

21 **Abstract**

22 Understanding the key drivers that affect a decline of soil organic carbon (SOC) stock in
23 agricultural areas is of major concern since leading to a decline in service provision from soils and
24 potentially carbon release into the atmosphere. Despite an increasing attention is given to SOC
25 depletion and degradation processes, SOC dynamics are far from being completely understood
26 because they occur in the long term and are the result of a complex interaction between
27 management and pedo-climatic factors. In order to improve our understanding of SOC reduction
28 phenomena in the mineral soils of Veneto region, this study aimed to adopt an innovative
29 probabilistic Bayesian Belief Network (BBN) framework to model SOC dynamics and identify
30 management scenarios that maximise its accumulation and minimise GHG emissions.

31 Results showed that the constructed BBN framework was able to describe SOC dynamics of the
32 Veneto region, predicting probabilities of general accumulation (11.0%) and depletion (55.0%),
33 similar to those already measured in field studies (15.3% and 50%, respectively). A general
34 enhancement in the SOC content was observed where a minimum soil disturbance was adopted.
35 This outcome suggested that management strategies of conversion from croplands to grasslands, no
36 tillage and conservation agriculture are the most promising management strategies to reverse
37 existing SOC reduction dynamics. Moreover, measures implying SOC stocks were also those
38 providing major benefits in terms of GHGs reduction emissions. Finally, climate change scenarios
39 slightly affected management practice. Advancements in our BBN framework might include more
40 detailed classes at higher resolution as well as any socio-cultural or economic aspect that should
41 improve the evaluation of prediction scenarios.

42

43 **Keywords**

44 Soil organic carbon; Agricultural management; Land use; Decision support system

45 **1. Introduction**

46 Soils are critical for the provision of economic goods and ecosystem services, including the
47 accumulation of atmospheric carbon (Lal, 2010). However, there is growing concern among
48 scientists and policy makers that soil organic carbon (SOC) is declining (Bouma, 2014; Stockmann
49 et al., 2015), particularly in agricultural areas, leading to a decline in service provision from soils
50 and potentially carbon release into the atmosphere (Koch et al., 2013; Smith, 2012). Monitoring
51 changes in SOC content can help identify degrading soils in order to target them for management
52 interventions that arrest declines and promote SOC accumulation.

53 Despite the attention that has been given to SOC (EC, 2012, Minelli et al., 2017), agricultural and
54 environmental impacts as a result of SOC changes in Europe still have large uncertainties associated
55 with them. These are dependent on several factors; economic (e.g., difficulty quantifying values of
56 ecosystem services), ecological (e.g., uncertainty about climate change scenarios) or socio-cultural
57 (e.g., willingness to adopt new technologies) (Burton and Schwarz, 2013; Smith et al., 2007a;
58 Yigini and Panagos, 2016). At the local scale, long-term field studies have shown different SOC
59 accumulation or depletion dynamics (Saby et al., 2008), mainly dependent on inherent pedologic
60 and climatic conditions, land use intensity, and cropping systems management (Berti et al., 2016;
61 Heikkinen et al., 2013; Maillard and Angers, 2014; Reijneveld et al., 2009). Predictions of SOC
62 dynamics under different management strategies and/or climate scenarios have been extensively
63 investigated using biogeochemical models (e.g., Borrelli et al., 2016; Lugato et al., 2014; Xu et al.,
64 2011) at the large scale (from regional to trans-national). However, these models are limited if
65 quantitative information is missing or uncertain.

66 Indeed, several SOM models rely on functional criteria related to microbial function (e.g. decay rate
67 of C pools) with the aim of representing the effect of biochemical and physical factors on SOC
68 turnover and C fluxes. However, as underlined by Dungait et al. (2012), the relative contribution of
69 biochemical and physical controls on the decay are rarely tested empirically, instead, the weakness

70 of a model's theoretical background is compensated for by calibration procedures. It follows that
71 too often models are over-calibrated in order to operate effectively in the soil systems where they
72 are validated. However, they are less consistent when applied to unusual soils or a different climate,
73 at "the edge of, or beyond, their validation" range (Dungait et al., 2012, p. 1790).

74 For these reasons, environmental processes and management have been increasingly modelled
75 following probabilistic approaches, where the uncertainty and variability of results is included in
76 modelling (Uusitalo, 2007). Bayesian belief networks (BBNs) are probabilistic models that
77 accommodate data uncertainty and variability and have increasingly been applied in ecological
78 modelling since they are able to integrate both qualitative and quantitative variables in a unique
79 model platform (Landuyt et al., 2013). By linking the different variables in a graphical interface,
80 BBN users define cause-and-effect relationships that provide both diagnosis and prognosis under
81 specific variable conditions, aiding the decision-making processing.

82 A first attempt to use BBNs to evaluate soil degradation was carried out by Hough et al. (2010) by
83 modelling peat erosion in Scotland using a combination of a national soil properties inventory and
84 local empirical observations. The authors identified climate variables the main factors associated
85 with peat erosion, while a secondary role was associated with land management practices, in
86 particular vegetation cover. Qualitative and quantitative information were merged also to evaluate
87 the risk of soil compaction (Troldborg et al., 2013), although a lack of data for model validation (at
88 field scale or from laboratory tests) partly weakened improvements in understanding factors (e.g.,
89 inherent soil characteristics, land management) and priorities to combat soil degradation.

90 In the Veneto region, north-eastern Italy, one of the most important impacts of intensive agriculture
91 on arable soils is the decline of SOC content, estimated at average rates of $1.1 \text{ Mg ha}^{-1} \text{ y}^{-1}$ (Morari
92 et al., 2006) as a result of continuous tillage, low organic inputs and over-simplification of cropping
93 systems (i.e. monocultures). In this context policy makers, as well as land managers and scientists,

94 need decision support tools to enable them to weigh up the benefits and drawbacks of different
95 agricultural systems and to explore best agri-environmental management strategies.

96 According to previous European experiences on modelling soil properties with a probabilistic
97 approach, it is expected that BBNs can provide new insights in soil management strategies. With
98 the general purpose of evaluating the feasibility of simulating the C biogeochemical cycle using
99 BBN models, this work aims: i) to quantify SOC accumulation and reduction in croplands and
100 grasslands across the Veneto region, north-eastern Italy, after independent model validation; ii) to
101 identify the main factors influencing SOC stock change dynamics; iii) to evaluate alternative
102 management scenarios that maximise SOC accumulation and simultaneously minimise GHG
103 emissions.

104

105 **2. Material and methods**

106 2.1 Study area

107 The Veneto region (NUTS-2, total area of 18,400 km²) is located in north-eastern Italy, where 55%
108 of the region is occupied by the Venetian plain, which is a complex system of urban, industrial, and
109 intensive agricultural areas characterised by high population density. According to the last
110 agricultural census (ISTAT, 2010), croplands and grasslands are mainly concentrated on the plain
111 (78%), comprising mainly cereals (maize, wheat), soybean, and fodder crops (ca. 70% of total
112 agricultural cultivations). Croplands and grasslands are generally irrigated where the shallow water
113 table, mainly located in the low-lying area around the Venice lagoon, does not contribute to soil
114 moisture in the root zone. A spatial visualisation of the Veneto region based on Corine Land Cover
115 inventory (2012) is reported in Figure 1.

116 Most of the soils of the regional low plain (<15 m a.s.l.) are Calcisols and Cambisols characterised
117 by sandy and silty-clay deposits with medium natural fertility deriving from low SOC content

118 (usually in the range of 10-20 g kg⁻¹) and low cation exchange capacity. Luvisols and Cambisols
119 (calcareous and skeletal loam, clay-loam soils) characterise mainly the high Venetian plain and hilly
120 areas in the north (15-300 m a.s.l.), while Leptosols and Cambisols are alternated in the mountains,
121 from sloping areas to valleys, respectively (WRB, 2014).

122

123 2.2 Bayesian Belief Network (BBN) model construction

124 A BBN model was built with the aim of combining the climate, biogeochemical and management
125 drivers that influence SOC stock change in the 0-30 cm layer, according to the conceptual
126 framework proposed in Morari et al. (2015). Drivers leading to changes in the SOC cycle were
127 identified from either natural- or human-induced processes (e.g., net primary production, soil
128 structure degradation), whose cause-and-effect relationships were identified after an iterative
129 process that aimed to put theory into a regional context. Only agroecosystems including croplands
130 and grasslands across the Veneto region were considered in this study. The target node was SOC
131 stock change (Fig. 2), which considered climate, soil and management as the main group-factors
132 comprising a total of 22 nodes and 30 links. According to Marcot et al. (2006), the number of nodes
133 and their states was kept as low as possible in order to favour their tractability and understanding,
134 while contemporarily describing SOC processes and SOC-related phenomena. In this context, some
135 intermediate nodes were required to summarise nodes into major themes (e.g., endogen and
136 hexogen carbon, soil fertility). Parentless input nodes represented the main geographic information
137 associated with cropping systems and pedo-climatic parameters. The BBN model was built using
138 Genie Academic 2.1 software (BayesFusion LLC, University of Pittsburgh, PA, USA).

139

140 2.3 BBN model parameterisation

141 Conditional probability tables (CPTs) were incorporated into the BBN model (each node was
142 associated with a CPT) through available data, expert knowledge and existing models gathered from
143 the literature and previous work conducted in the area, while parentless nodes had unconditional
144 probability tables composed of prior knowledge on the frequencies of each state.

145 Parentless pedo-climatic nodes were populated using empirical evidence: in particular soil data
146 from the Veneto Region 1:250,000 soil map (Regione Veneto, 2005), which is linked to an
147 alphanumeric database with physicochemical characteristics (pH, texture, depth, intrinsic SOC
148 content etc.). The database is regularly revised by the Veneto Region Environmental Protection
149 Agency (ARPA Veneto), which provided an upgraded version of the database whose SOC data (0-
150 3- cm soil layer) referred to the year 2010 (http://www.arpa.veneto.it/arpavinforma/indicatori-ambientali/indicatori_ambientali/geosfera/qualita-dei-suoli/contenuto-di-carbonio-organico-nello-strato-superficiale-di-suolo/view). The database did not include soil porosity information, which
153 was estimated from bulk and particle density (Jury and Horton, 2004). Despite bulk density was
154 present in the database and represent a key parameter to determine SOC stocks, here it is was not
155 included among the basic parentless nodes. Firstly, because bulk density is correlated with soil
156 texture properties and may represent a redundant information that is not needed in the BBN (Marcot
157 et al., 2006). Secondly, because the aim of the work was to quantify the SOC stock change (rather
158 than its absolute value), whose dynamic is not correlated with bulk density which was assumed a
159 steady property.

160 The climatic database of Veneto used was that already adopted by Dal Ferro et al. (2016) in a study
161 conducted in the same area and based on 35 meteorological stations evenly spread over the region,
162 which provided 20 years of climatic data (1993-2013). Rainfall and reference evapotranspiration
163 (ET_0), calculated using Penman-Monteith equation (Allen et al., 1998) by linking vegetation,
164 temperature and time of year, were included as parentless nodes. Despite temperature is usually

165 associated with crop biomass, in our BBN framework it was not explicitly used because implicitly
166 included in the ET_0 node.

167 Parentless crops and fertiliser information were provided by the Veneto Region agricultural
168 administration (Dal Ferro et al., 2016; Regione Veneto, 2012) at the municipal level. The database
169 was used to describe cropland and grassland probability distributions across the region as well as
170 type (organic or mineral) and quantity ($\text{kg ha}^{-1} \text{y}^{-1}$) of nutrient input. Irrigation was also included in
171 the BBN model by considering the regional partition between irrigated and non-irrigated areas
172 according to the ISTAT database (ISTAT, 2010).

173 Node-associated conditional probabilities were built using to a composite approach, in some cases
174 using data derived by local field trials and modelling experiments while in others expert knowledge
175 and literature review. In particular, data on soil tillage and cover crop practices were extracted from
176 information on their spatial distribution across the Veneto region gathered through regional surveys
177 carried out by the Rural Development Programme (Regione Veneto, 2013). Probability distributions
178 of SOC turnover rate and crop biomass were derived from the modelling study of Dal Ferro et al.
179 (2016) that was conducted in the Veneto region. Following Landuyt et al. (2016) these CPTs were
180 determined based on the spatial relationship with associated parameters, such as soil fertility, ET_0 ,
181 water supply, etc. (Table 1). In this context, soil moisture was not included to affect SOC dynamics
182 because it is strictly related to soil texture. Similarly, soil nitrogen was also correlated with texture
183 parameters and therefore not sensitive to change SOC. Nevertheless, experimental and modelling
184 results showed that the fertiliser type, that in turn affected hexogen carbon, was the main factor to
185 change soil carbon-nitrogen dynamics. According to Marcot et al., (2006), pedo-climatic and childe
186 nodes were categorised by probabilistic state values (e.g., high, medium, low), defined through the
187 conversion of continuous variables. The number of categories was kept the lowest as possible,
188 although able to represent influences.

189

190 2.4 BBN scenarios

191 2.4.1 Land use and management

192 Land use and management scenarios, selected among others since the most promising and readily
193 applicable in Europe to maintain SOC in agricultural soils (Morari et al., 2015; Powlson et al.,
194 2011), have been hypothesised as the conversion from current agronomic conditions (hereafter
195 called “standard scenario”) to those adopting different strategies:

- 196 a. Croplands to 50% and alternatively 100% grassland: areas currently under arable
197 production were converted to permanent grassland where grazing, hay making or mixed
198 practices are generally applied;
- 199 b. Arable lands to 50% and alternatively 100% under no tillage practices: conventional
200 practices, which usually include several tillage operations after crop harvest (mouldboard
201 ploughing) and throughout the crop season (disk harrowing before sowing, hoeing, etc.),
202 were converted to no tillage management;
- 203 c. Croplands to 50% and alternatively 100% of continuous soil cover with cover crops: this
204 scenario simulated that cover crops followed the main crop in order to maintain continuous
205 soil cover throughout the year. Cover crops were completely incorporated (i.e., used as
206 green manure) into the soil;
- 207 d. Monoculture croplands to 50% and alternatively 100% under crop rotation: a succession of
208 different crops including legumes in arable lands replaced intensive monoculture practices
209 (mainly maize);
- 210 e. Croplands to 50% and alternatively 100% under conservation agriculture: following the
211 regional guidelines that were proposed in the RDP 2007-2013 (Regione Veneto, 2013), this
212 scenario was set up to predict the effects of conservation agriculture by including
213 simultaneously crop rotation, cover crops and no tillage management practices;

214 f. Organic (farmyard manure) to 50% and alternatively 100% of total fertiliser input: an
215 increase in the use of soil amendments (farmyard manure) was modelled as a substitute to
216 mineral fertiliser.

217 2.4.2 Climate change scenarios

218 Projections of changes in climate, as provided by the Intergovernmental Panel on Climate Change
219 (IPCC, 2007; IPCC, 2013), were combined with land use and management data in order to evaluate
220 the effectiveness of potentially adopted strategies (see paragraph 2.4.1) to mitigate climate change.
221 For this purpose, the quantification of greenhouse gas fluxes was included in the BBN model in
222 terms of net carbon dioxide (CO₂), methane (CH₄) and nitrous oxide (N₂O) changes in agricultural
223 fields. In particular, CO₂ was directly correlated with SOC dynamics, while CH₄ was associated
224 with the degree of hexogen C input and rainfall, and N₂O was linked to fertilisers type and dose as
225 well as climate conditions (i.e., temperature) (Smith et al., 2014; Smith et al., 2007b). Finally,
226 GHGs emissions were converted into CO₂ equivalent (CO₂-eq) terms to enable an evaluation of
227 integrated global warming potential (GWP) for CO₂ (GWP = 1), CH₄ (GWP = 28) and N₂O (GWP
228 = 265) over a time horizon of 100 years (Smith et al., 2007b). Equivalent CO₂ emissions were
229 modelled as utility values (Fig. 3), which refer to the combination of different management
230 strategies with climate change emission scenarios as described in Nakicenovic et al. (2000). In
231 particular, scenarios labelled as B1 (“Sustainable world”, corresponding to atmospheric CO₂
232 concentration of 538 ppm), A1B (“Rich world”, corresponding to CO₂ concentration of 674 ppm)
233 and A2 (“Separated world”, corresponding to CO₂ concentration of 754 ppm) were selected for
234 comparison in this study. Some simplifications have been done: i) climate change effects were
235 considered only in terms of rainfall and air temperature variations, neglecting the potential effects
236 of CO₂ increase on other factors such as biomass yield; ii) only climate data without any further
237 prediction on socio-cultural and economic change was considered; iii) CO₂-eq quantified only
238 emissions from the biogeochemical cycles of different crop systems, thus excluding management

239 aspects (e.g., machinery use) that directly contribute to changes in GHGs emissions; iv) despite the
240 major contribution of rice paddy fields to GHGs emissions, they were not considered in the current
241 analysis (ca. 0.9% of regional agricultural fields); v) potential adaptations of farm management
242 systems (e.g. selection of new crop species and varieties, application of efficient irrigation methods)
243 to climate change scenarios were not considered; vi) IPCC Special Report on Emission Scenarios
244 (Nakicenovic et al., 2000), instead of the most recent IPCC Representative Concentration Pathways
245 (IPCC, 2013), was used for consistency and comparison with previous studies (Lugato et al., 2015).

246 The stochastic weather generator LARS-WG (Semenov and Barrow, 2002) was used to produce a
247 daily time series of climatic variables. Weather parameters were calibrated by using probability
248 distributions of locally observed daily weather variables. Semi-empirical distributions of observed
249 data were successively found, while Fourier series were used to describe precipitation amount, solar
250 radiation, minimum and maximum temperatures. Finally, LARS-WG generated climate change
251 weather data from multi-model ensemble of 15 climate models (Semenov et al., 2013) that were
252 used in the IPCC 4th Assessment Report. In this context, the weather database for the Veneto region
253 was used to describe alternative climate scenarios and evaluate their impact on CO₂-eq emissions.

254 2.5 BBN model validation

255 BBNs have been extensively used to evaluate ecosystem services and environmental management
256 without any model validation, or simply based on stakeholder evaluation (Landuyt et al., 2013).
257 However, assessing the ability of the model to represent target variables is a key step to providing
258 reliable scenarios (Death et al., 2015), particularly in the case of SOC stock change, which is rather
259 difficult to quantify without real-world data. Moreover, due to the low reactivity of SOC to
260 management changes and high spatial variability, SOC dynamics should be evaluated in the
261 medium/long term after stabilised management conditions, so as to reduce uncertainties in detecting
262 changes in SOM stocks (Kuikman et al., 2012). In this context, the model was validated by
263 comparing the BBN predictions on SOC stock change to a total of 212 unique values that were

264 obtained from different case studies (Fig. 1). Field data (187 sampling points), collected in large
265 plots (7.8×6 m) from a long-term experiment (established in 1962 and still ongoing) (Berti et al.,
266 2016) were representative of different cropping systems (e.g. monoculture, crop rotation, grassland)
267 and fertiliser inputs (e.g. mineral, organic, mixed) that are traditionally adopted across the Veneto
268 region (Regione Veneto, 2012). The experiment is located at the experimental farm of the
269 University of Padova ($45^\circ 20' N 11^\circ 18' E$, 6 m a.s.l.), characterised by a loamy Fluvi-Calcaric
270 Cambisol. Agricultural practices that have only recently been introduced in the study area (i.e., no
271 tillage, use of cover crops) were monitored in three different farms (69 sampling points) over a 3-
272 year time span (Piccoli et al., 2016). The farms are located in three different areas of the Veneto
273 region from east (Caorle municipality, $45^\circ 38' N 12^\circ 57' E$, -2 m a.s.l.; silty-clay to sandy-loam,
274 Gleyc Fluvisols or Endogleyc Flucic Cambisols) to centre (Mogliano Veneto municipality, $45^\circ 35'$
275 $N 12^\circ 18' E$, 6 m a.s.l.; silty-loam, Endogleyc Cambisols) and south-west (Ceregnano municipality,
276 $45^\circ 3' N 11^\circ 53' E$, 2 m a.s.l.; silty-loam, Endogleyc Cambisols) and well represented the pedo-
277 climatic variability of the Venetian plain.

278

279 **3. Results**

280 3.1 Model validation and sensitivity analysis

281 In general, results showed that the BBN framework was reasonably accurate in modelling the SOC
282 dynamics in the 0-30 cm profile (Fig. 4) since it was able to predict probabilities of general
283 accumulation (11.0% vs. 15.3%) and depletion (55.0% vs. 50%) as already measured in the field.
284 Small variations ($-0.1 \text{ Mg ha}^{-1} \text{ y}^{-1} < \text{SOC change} < 0.1 \text{ Mg ha}^{-1} \text{ y}^{-1}$) were also well described (34.0%
285 vs. 34.7%). Nevertheless, by analysing SOC dynamics in detail, an overestimation was observed
286 (18.0% vs 7.1%) of the “medium decrease” state value ($-0.5 \text{ Mg ha}^{-1} \text{ y}^{-1} < \text{SOC change} < 1.0 \text{ Mg}$

287 $\text{ha}^{-1} \text{y}^{-1}$), while extreme increases ($> 1 \text{ Mg ha}^{-1} \text{y}^{-1}$) or decreases ($< 1 \text{ Mg ha}^{-1} \text{y}^{-1}$) were negligible in
288 both the real and modelled state.

289 Under standard land use and management conditions, the BBN model predicted that a moderate
290 reduction in the SOC stock (here estimated in the range of $0.1 - 0.5 \text{ Mg C ha}^{-1} \text{y}^{-1}$) prevailed across
291 the Veneto region, with a probability of 34% (Fig. 2), similar to the 33% estimated for the
292 equilibrium in SOC dynamics (between -0.1 and $0.1 \text{ Mg C ha}^{-1} \text{y}^{-1}$). Further probabilities
293 emphasised land degradation conditions (total 50%), while contrasting dynamics leading to SOC
294 accumulation had a probability of only 17%, although in some cases they were estimated as greater
295 than $1.0 \text{ Mg C ha}^{-1} \text{y}^{-1}$.

296 SOC stock change dynamics were the result of a complex interaction between management and
297 pedo-climatic conditions. The influence of every node was calculated in Genie Academic 2.1
298 through a one-way sensitivity analysis, which estimated the spread of posterior probabilities of the
299 specified target node (here SOC stock change) according to Castillo et al. (1997). In this context,
300 field management practices, in particular the “Cropping system” and “Tillage operations”, were the
301 nodes that most strongly influenced SOC stock change (Table 2). A secondary role was provided
302 by: i) the intrinsic SOC content (Table 2), which depended on the peculiar pedo-climatic condition
303 of the region and was mainly classified as medium low ($10-20 \text{ g kg}^{-1}$); ii) the SOC turnover
304 coefficient, here generally implying SOC degradation conditions (89%) and associated with both
305 pedo-climatic (soil texture, soil porosity, temperature) and management factors (soil disturbance by
306 tillage). In contrast, the sensitivity analysis diagnosed negligible effects for soil-water factors
307 (rainfall, irrigation) as well as nutrient quantity-related parameters (available N input, fertiliser
308 dose), while their quality (e.g. organic amendments instead of mineral fertilisers) could partially
309 modify SOC accumulation or depletion.

310

311 3.2 Soil management scenarios

312 A change in land use and management from standard conditions to soil-improving scenarios
313 showed contrasting effects between different strategies. A general enhancement in the SOC content
314 was observed when adopting practices of minimum soil disturbance as a consequence of conversion
315 from croplands to grasslands, no tillage and conservation agriculture. Moreover, the modelled
316 scenarios showed their ability to reverse the overall SOC dynamics trend, since all predicted a
317 major accumulation that mainly offset the SOC reduction. In this context, croplands to grasslands,
318 no tillage and conservation agriculture measures were able to increase the SOC content in the 0-30
319 soil layer, whether adopted on 50% (+29%, on average) or 100% (+57.7%, on average) of current
320 arable land, with negligible differences between measures (Fig. 5). The estimated increase in SOC
321 mainly involved medium (0.5 to 1.0 Mg ha⁻¹ y⁻¹) and strong (>1.0 Mg ha⁻¹ y⁻¹) improvements,
322 overall reaching up to 60% of SOC stock change probability vs. 7% under the standard scenario.

323 By contrast, crop management strategies involving continuous soil cover and crop rotation showed
324 only minor changes in the SOC dynamics of arable lands, highlighting the slight contribution of
325 related nodes (e.g., organic carbon input from residues) as reported in the sensitivity analysis (Table
326 2). In particular, maintaining continuous soil cover through using cover crops, on both 50% and
327 100% of arable land, slightly reduced the probability of a SOC low decrease (-1%) towards
328 equilibrium (no change, +1%), while crop rotation – instead of monoculture – led to some increase
329 in medium SOC (+1%) in place of its general equilibrium (-1%).

330 Intermediate changes were observed when simulating a management change in fertiliser use,
331 especially when farmyard manure was entirely (100%) adopted. Although SOC accumulation
332 increased its overall probability by only 1% with respect to the standard scenario, the highest
333 increase was observed for the most performing categories (i.e., high increase, +2%; medium
334 increase, +1%) in place of minor changes for the others (i.e., no change, low increase). By contrast,

335 this scenario highlighted weak capabilities to reverse overall SOC accumulation/reduction dynamics
336 (Fig. 5).

337

338 3.3 GHGs emission scenarios

339 Impacts that might be generated by current and modelled management scenarios were evaluated in
340 terms of CO₂ equivalents (CO₂-eq) and predicted in the context of climate change emissions
341 scenarios (Table 3). In the standard scenario, state values of CO₂-eq balance from cropland and
342 grassland showed net emissions, quantified at 1613.9 kg ha⁻¹ y⁻¹, with major contributions of CO₂
343 and N₂O. In this context, estimated CO₂ fluxes from agricultural fields had 52% low emission
344 probability (0-1000 kg C-CO₂ ha⁻¹ y⁻¹), followed by 8% high (> 1000 kg C-CO₂ ha⁻¹ y⁻¹), while
345 those associated with N₂O were estimated 71% medium (1-3 kg N-N₂O ha⁻¹ y⁻¹), 27% low (0-1 kg
346 N-N₂O ha⁻¹ y⁻¹) and finally 2% high (> 3 kg N-N₂O ha⁻¹ y⁻¹). Methane emissions were always low
347 (0-10 kg ha⁻¹ y⁻¹). Modelled land use and management scenarios provided, in some cases, strong
348 improvements in terms of GHGs emissions (e.g., minimum soil disturbance), while in others the
349 difference with the standard scenario was negligible (e.g., continuous soil cover, conversion to
350 organic input). In particular adopting no tillage, conversion from cropland to grassland and
351 conservation agriculture (100% of the area) favoured net CO₂-eq adsorption dynamics (984 kg CO₂-
352 eq ha⁻¹ y⁻¹, on average), while 50% of their adoption involved lower equivalent CO₂ emissions (321
353 kg CO₂-eq ha⁻¹ y⁻¹, on average) with respect to the standard scenario. Modelled land use and
354 management strategies under climate change scenarios generally involved worsening conditions in
355 terms of CO₂-eq emissions with respect to the current climatic conditions although always lower
356 than 70 kg CO₂-eq ha⁻¹ y⁻¹ (Table 3). In particular, the higher temperatures affected an increase of
357 N-N₂O emissions (the “High” class increased up to 5%, on average), offsetting a lowering of CO₂
358 emissions (ca. 1%) as a result of major endogen carbon inputs. By contrast, the BBN framework

359 was seldom able to identify changes between rich (A1B), separate (A2) and sustainable (B1) world
360 scenarios since differences were always $\leq 1.0 \text{ kg CO}_2\text{-eq ha}^{-1} \text{ y}^{-1}$.

361

362 **4. Discussion**

363 The comparison of experimental results of SOC stock change with those from the developed
364 Bayesian Belief Network suggests that the model performed well when evaluated with independent
365 data, suggesting that the BBN was able to accurately describe the effects of different scenarios.
366 Although BBNs work effectively with retrieval of partial data (Aguilera et al., 2011) it has also
367 been recently reported in other studies (Death et al., 2015; Marcot, 2012) that steps leading to their
368 accurate application should include independent validation to avoid bias in results as a consequence
369 of expert, albeit subjective, knowledge.

370 As also observed in our study, in general the BBN simulation matched the general trend of SOC
371 accumulation and depletion dynamics, whereas some specific classes (“medium decrease”) were
372 overestimated. This is likely due to some binding balance between requirements, on the one hand of
373 detailed information, and on the other of simplification in the definition of state values and number
374 of nodes. Predictions of SOC stock change across the Veneto region by the BBN model highlighted
375 general soil degradation conditions, whose SOC reduction was quantified with high probability in
376 the “Low increase” category ($0.1\text{-}0.5 \text{ Mg C ha}^{-1} \text{ y}^{-1}$). These results were similar to those reported in
377 a study that was conducted in the same area using the DAYCENT biogeochemical model (Dal
378 Ferro et al., 2016), showing average losses of $257 \text{ kg C ha}^{-1} \text{ y}^{-1}$ (0-20 cm layer), although with
379 negative peaks lower than $-4.0 \text{ Mg C ha}^{-1} \text{ y}^{-1}$ that were conversely not found here. Very few
380 experimental results have assessed SOC stock changes on a large scale. Extensive field surveys on
381 SOC content over the period 1979-2008 were combined with a geostatistical approach by Fantappiè
382 et al. (2010) in an attempt to map Italian soil C dynamics. The authors, although with great

383 uncertainties, reported SOC stock variations of between $-1.5 \text{ Mg ha}^{-1} \text{ y}^{-1}$ and $+1.5 \text{ Mg ha}^{-1} \text{ y}^{-1}$ (0-50
384 cm) for most soils in Veneto, emphasising that a dynamic SOC input-output equilibrium was far
385 from being reached. In particular, they observed that land use type (e.g. cropland or grassland) was
386 the most important factor leading to SOC variation, while a secondary role was associated with
387 changes in land use intensity (e.g. crop system change). Similarly, the one-way sensitivity analysis
388 (Table 2) showed that the type of cropping system per se and tillage operations, which are the
389 factors that mainly characterise land use type (e.g. cropland instead of grassland), were primarily
390 involved in SOC stock change dynamics, as also observed in long-term studies that have been
391 conducted in north-eastern Italy (Morari et al., 2006). Improvements for SOC content were
392 specifically modelled with the BBN through decreasing soil disturbance with zero-tillage (both in
393 cropland and with the conversion to grassland) and maintaining a continuous soil cover (cover crops
394 and grassland), although with contrasting results. Interestingly, only the omission of tillage
395 operations was able to reverse the C dynamics trend from a general SOC reduction to major
396 accumulation, although some SOC equilibrium/reduction phenomena were still likely. Maintaining
397 continuous soil cover through cover crops had only a minor effect, even when its application was
398 extended to 100% of arable lands. Mazzoncini et al. (2011) have reported contrasting results on the
399 effects of cover crops on a loam soil in central Italy, where SOC increases were mainly observed in
400 the soil surface layer (0-10 cm). However, these effects were observed some 15 years after the
401 establishment of cover crops and the adoption of high nitrogen supply legume cover crops, which
402 are seldom adopted in the Veneto region. In addition, a recent meta-analysis on SOC sequestration
403 via cultivation of cover crops (Poeplau and Don, 2015) reported a mean annual accumulation rate of
404 $0.32 \pm 0.08 \text{ Mg ha}^{-1} \text{ y}^{-1}$ (0-22 cm soil layer) in a time span of 54 years, in contrast to our findings.
405 However, their study was conducted at the global scale including a wide variety of pedo-climatic
406 conditions.

407 Findings on the different effects of no tillage and cover crops were combined with those from crop
408 rotations in the conservation agriculture scenario, which showed comparable results to those
409 reported for no tillage practices. As a consequence, general SOC improving conditions were partly
410 mitigated by “No change” and “Low decrease” conditions. This was recently observed by Piccoli et
411 al. (2016), although they also suggested that SOC stock changes should be evaluated over a deeper
412 profile (50 cm) and longer periods of time to better evaluate the contribution of conservation
413 practices to SOC accumulation or distribution, although the wide spatial variability could
414 compensate the short-term period. Nevertheless, bias in our estimations cannot be completely
415 excluded as our BBN model validation (Fig. 3) showed, in particular, some overestimation of SOC
416 reduction rates. Moreover, the mismatch between SOC dynamics, derived from agricultural
417 experimental studies, and their representativeness whether adopted at the large-scale is still debated,
418 highlighting management and biological uncertainties on their real effectiveness (Smith et al.,
419 2005). Finally, it must be noted that differences in soil sampling and quantification of SOC content
420 may increase the uncertainty on SOC dynamics from field regional scale because of its nonlinear
421 accumulation/decomposition rate (Six and Jastrow, 2002).

422 Measures for increasing soil carbon inputs with high refractory coefficients have been suggested to
423 reduce SOC turnover and contribute to SOC stock. Recent findings (Berti et al., 2016; Kätterer et
424 al., 2011) have confirmed that farmyard manure, among different hexogen C inputs, had the greatest
425 potential in stabilising SOC content, since it shows the highest humification coefficient. In this
426 context, a massive conversion of mineral nutrients input to organic amendments (farmyard manure)
427 was hypothesised. Although the 100% application of farmyard manure instead of mineral fertiliser
428 is not realistic, it was useful to investigate here to provide evidence on its effectiveness, since it is
429 considered one of the best practices to increase SOC in mineral soils (Lal, 2004). Some benefits
430 were observed in terms of SOC increases, especially at high rates ($> 1.0 \text{ Mg ha}^{-1} \text{ y}^{-1}$), likely
431 influenced by sharp initial accumulations in arable soils of the low-lying plain that hardly receive

432 organic amendments. Nevertheless, according to early studies on SOC stock scenarios (Smith et al.,
433 1997), soils amended with organic manure has low C accumulation potential when compared to
434 other management options (Fig. 5). In addition, care should be taken to consider the overall
435 efficiency of the agricultural system when adopting organic inputs that might imply significant
436 releases of nitrogen (N), especially in the low-lying Venetian plain that often has loose soils and a
437 shallow water table, which makes it vulnerable to N leaching (Morari et al., 2012).

438 Climate variability, evaluated with the BBN in terms of climate change scenarios (temperature,
439 rainfall and crop evapotranspiration), provided information on utility values of adopting different
440 management strategies in terms of CO₂-eq emissions. The input-output CO₂-eq budget changed
441 from current climatic conditions to those foreseen by the IPCC (Nakicenovic et al., 2000), on
442 average by increasing the overall GHGs emissions as a result of increasing N₂O emissions, which
443 counterbalanced reduced CO₂ emissions (from increased SOC stock) due to its greater global
444 warming potential. However, the adoption of SOC-improving strategies (zero tillage, cropland to
445 grassland, conservation agriculture) was still able to contribute actively to reducing GHGs
446 emissions (Table 3). By contrast, marginal differences due to climate variability were observed
447 since changing scenarios resulted in similar trends on GHGs emissions, as also reported in previous
448 studies conducted at the European level (Lugato et al., 2014). Nevertheless, long-term validation is
449 still required, especially for conservation agriculture practices, to evaluate possible changes on SOC
450 and GHGs dynamics from short to long run.

451 These outcomes demonstrate that variability of management strategies across the Veneto region are
452 likely to affect the SOC stock change more than climate variability, at least at the regional level
453 (Table 2), thus emphasising the major contribution of CO₂, which is strictly related to SOC stock
454 change (Fig. 3), to CO₂-eq emissions with respect to N₂O (Montzka et al., 2011). On the other hand,
455 these results might have been affected by the sensitivity of the BBN model to slight variations in
456 temperature and rainfall. Nevertheless, improvements in the BBN model (e.g., definition of more

457 detailed classes, including experimental data at higher resolution) could overcome the low
458 sensitivity to climate variability that was found, by providing more accurate outcomes as a result of
459 slight variations in BBN parameters. Finally, at this stage the BBN framework did not take into
460 account any socio-cultural or economic aspects that might affect economical support to farmers for
461 soil-improving systems, the level of farmer expertise or technological developments leading to
462 increased applicability and acceptance of sustainable land management practices. Nevertheless, it
463 was largely achieved that BBNs can be used in an adaptive modelling framework that is often
464 missing from traditional modelling approaches (Landuyt et al., 2013). Further work will be targeted
465 to updating our framework to achieve socio-cultural and economic objectives.

466

467 **5. Conclusions**

468 The constructed BBN model well described the main management and climatic aspects related to
469 SOC dynamics in croplands and grasslands across Veneto, showing its ability to act from farm
470 (validation) to regional scale (consistent results with previous studies). By reflecting the variability
471 of SOC dynamics in real world conditions and by including quali-quantitative information
472 following a probabilistic approach, the BBN has proven to be a valuable decision support tool to
473 distinguish the effect of different management practices. Strategies to reduce SOC depletion and
474 soil degradation include minimum soil disturbance through no tillage and conversion from arable
475 lands to grasslands. Covers crops, the use of organic amendments and crop rotation had only slight
476 effects on SOC accumulation. In this context, the model was suitable to fill the gap between
477 localised experimental studies and their extension to territorial application since including
478 uncertainties that are usually not included in biogeochemical models. Finally, measures implying
479 greater SOC stock were also those providing major benefits in terms of GHGs emissions. Further
480 improvements should include socio-cultural and economic aspects, especially in the evaluation of
481 prediction scenarios.

482

483 **Acknowledgements**

484 The research leading to these results has received funding from the European Union Seventh
485 Framework Programme (FP7/2007-2013) under grant agreement no. 603498 (RE CARE project).
486 Special thanks to ARPAV (Environmental Protection Agency of Veneto Region, Italy) for
487 providing soil and meteorological data and the Agri-environment, Hunting and Fishing Direction of
488 Veneto Region for the support in land use and management data collection.

489

490 **References**

- 491 Aguilera, P.A., Fernández, A., Fernández, R., Rumí, R., Salmerón, A., 2011. Bayesian networks in
492 environmental modelling. *Environ. Model. Softw.* 26, 1376-1388.
- 493 Allen, R.G., Pereira, L.S., Raes, D., Smith, M., 1998. Crop evapotranspiration—Guidelines for
494 computing crop water requirements—FAO Irrigation and drainage paper 56. FAO, Rome.
- 495 Berti, A., Morari, F., Dal Ferro, N., Simonetti, G., Polese, R., 2016. Organic input quality is more
496 important than its quantity: C turnover coefficients in different cropping systems. *Eur. J. Agron.* 77,
497 138-145.
- 498 Borrelli, P., Paustian, K., Panagos, P., Jones, A., Schütt, B., Lugato, E., 2016. Effect of Good
499 Agricultural and Environmental Conditions on erosion and soil organic carbon balance: A national
500 case study. *Land Use Policy* 50, 408-421.
- 501 Bouma, J., 2014. Soil science contributions towards Sustainable Development Goals and their
502 implementation: linking soil functions with ecosystem services. *J. Plant Nutr. Soil Sci.* 177, 111-
503 120.

504 Burton, R.J.F., Schwarz, G., 2013. Result-oriented agri-environmental schemes in Europe and their
505 potential for promoting behavioural change. *Land Use Policy* 30, 628-641.

506 Castillo, E., Gutiérrez, J.M., Hadi, A.S., 1997. Sensitivity analysis in discrete Bayesian networks.
507 *IEEE Trans. Syst. Man Cybern.* 27, 412-423.

508 IPCC, 2007. *Climate Change 2007: The Physical Science Basis. Contribution of Working Group I*
509 *to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change.* Cambridge
510 University Press, Cambridge and New York.

511 IPCC, 2013. *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I*
512 *to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change.* Cambridge
513 University Press, Cambridge and New York.

514 Dal Ferro, N., Cocco, E., Lazzaro, B., Berti, A., Morari, F., 2016. Assessing the role of agri-
515 environmental measures to enhance the environment in the Veneto Region, Italy, with a model-
516 based approach. *Agric., Ecosyst. Environ.* 232, 312-325.

517 Death, R.G., Death, F., Stubbington, R., Joy, M.K., Belt, M., 2015. How good are Bayesian belief
518 networks for environmental management? A test with data from an agricultural river catchment.
519 *Freshwat. Biol.* 60, 2297-2309.

520 Dungait, J.A.J., Hopkins, D.W., Gregory, A.S., Whitmore, A.P., 2012. Soil organic matter turnover
521 is governed by accessibility not recalcitrance. *Global Change Biol.* 18, 1781-1796.

522 European Commission, 2012. Report from the Commission to the European Parliament, the
523 Council, the European Economic and Social Committee and the Committee of the Regions, The
524 implementation of the Soil Thematic Strategy and ongoing activities. (COM(2012) 46). Official
525 Journal of the European Union, Brussels.

526 Fantappiè, M., L'Abate, G., Costantini, E.A.C., 2010. Factors influencing soil organic carbon stock
527 variations in Italy during the last three decades, in: Zdruli, P., Pagliai, M., Kapur, M., Faz Cano, A.
528 (Eds.), Land degradation and desertification: assessment, mitigation and remediation. Springer,
529 London, pp. 435-465.

530 Heikkinen, J., Ketoja, E., Nuutinen, V., Regina, K., 2013. Declining trend of carbon in Finnish
531 cropland soils in 1974–2009. *Global Change Biol.* 19, 1456-1469.

532 Hough, R.L., Towers, W., Aalders, I., 2010. The risk of peat erosion from climate change: land
533 management combinations—an assessment with bayesian belief networks. *Hum. Ecol. Risk Assess.*
534 16, 962-976.

535 IPCC, 2013. Climate Change 2013: The Physical Science Basis. Contribution of Working Group I
536 to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, in: Stocker,
537 T.F., Qin, D., Plattner, G.-K. Tignor, M., Allen, S.K., Boschung, J. Nauels, A. Xia, Y., Bex, V.,
538 Midgley, P.M. (Eds.), Cambridge University Press, Cambridge, New York, pp. 1535.

539 ISTAT, 2010. Italian general agricultural census. Italian National Institute of Statistics. [http://dati-](http://dati-censimentoagricoltura.istat.it/)
540 [censimentoagricoltura.istat.it/](http://dati-censimentoagricoltura.istat.it/) (accessed 18.07. 17).

541 IUSS Working Group WRB, 2014. World reference base for soil resources 2014. International soil
542 classification system for naming soils and creating legends for soil maps. World Soil Resources
543 Reports No. 106, FAO, Rome.

544 Jury, W A, Horton, R., 2004. Soil physics, 6th edition. John Wiley & Sons Inc., Hoboken.

545 Kätterer, T., Bolinder, M.A., Andrén, O., Kirchmann, H., Menichetti, L., 2011. Roots contribute
546 more to refractory soil organic matter than above-ground crop residues, as revealed by a long-term
547 field experiment. *Agric., Ecosyst. Environ.* 141, 184-192.

548 Koch, A., McBratney, A., Adams, M., Field, D., Hill, R., Crawford, J., Minasny, B., Lal, R.,
549 Abbott, L., O'Donnell, A., 2013. Soil security: solving the global soil crisis. *Glob. Policy* 4, 434-
550 441.

551 Kuikman, P.J., Ehlert, P.A.I., Chardon, W.J., van Beek, C.L., Tóth, G., Oenema, O., 2012. Soil
552 organic matter decline, in: van Beek, C.L., Tóth, G. (Eds.), *Risk assessment methodologies of soil*
553 *threats in Europe*. Publications Office of the European Union, Luxembourg, pp. 41-50.

554 Lal, R., 2010. Managing soils and ecosystems for mitigating anthropogenic carbon emissions and
555 advancing global food security. *Bioscience* 60, 708-721.

556 Lal, R., 2004. Soil carbon sequestration to mitigate climate change. *Geoderma* 123, 1-22.

557 Landuyt, D., Broekx, S., D'hondt, R., Engelen, G., Aertsens, J., Goethals, P.L.M., 2013. A review
558 of Bayesian belief networks in ecosystem service modelling. *Environ. Model. Softw.* 46, 1-11.

559 Landuyt, D., Broekx, S., Goethals, P.L.M., 2016. Bayesian belief networks to analyse trade-offs
560 among ecosystem services at the regional scale. *Ecol. Ind.* 71, 327-335.

561 Lugato, E., Bampa, F., Panagos, P., Montanarella, L., Jones, A., 2014. Potential carbon
562 sequestration of European arable soils estimated by modelling a comprehensive set of management
563 practices. *Global Change Biol.* 20, 3557-3567.

564 Maillard, E., Angers, D.A., 2014. Animal manure application and soil organic carbon stocks: A
565 meta-analysis. *Glob. Change Biol.* 20, 666-679.

566 Marcot, B.G., 2012. Metrics for evaluating performance and uncertainty of Bayesian network
567 models. *Ecol. Model.* 230, 50-62.

568 Marcot, B.G., Steventon, J.D., Sutherland, G.D., McCann, R.K., 2006. Guidelines for developing
569 and updating Bayesian belief networks applied to ecological modeling and conservation. *Can. J.*
570 *For. Res.* 36, 3063-3074.

571 Mazzoncini, M., Sapkota, T.B., Bàrberi, P., Antichi, D., Risaliti, R., 2011. Long-term effect of
572 tillage, nitrogen fertilization and cover crops on soil organic carbon and total nitrogen content. *Soil*
573 *Till. Res.* 114, 165-174.

574 Minelli, A., Erlewein, A., Castillo, V. 2017. Land Degradation Neutrality and the UNCCD: From
575 Political Vision to Measurable Targets, in: Ginzky H., Heuser, I.L., Qin, T., Ruppel, O.C., Wegerdt,
576 P. (Eds.), *International Yearbook of Soil Law and Policy 2016*. Springer, Cham, pp. 85-104.

577 Montzka, S.A., Dlugokencky, E.J., Butler, J.H., 2011. Non-CO₂ greenhouse gases and climate
578 change. *Nature* 476, 43-50.

579 Morari, F., Lugato, E., Berti, A., Giardini, L., 2006. Long-term effects of recommended
580 management practices on soil carbon changes and sequestration in north-eastern Italy. *Soil Use*
581 *Manage.* 22, 71-81.

582 Morari, F., Lugato, E., Polese, R., Berti, A., Giardini, L., 2012. Nitrate concentrations in
583 groundwater under contrasting agricultural management practices in the low plains of Italy. *Agric.,*
584 *Ecosyst. Environ.* 147, 47-56.

585 Morari, F., Panagos, P., Bampa, F., 2015. Decline of organic matter in mineral soils, in: Stolte, J.,
586 Tesfai, M., Øygarden, L., Kværnø, S., Keizer, J., Verheijen, F., Panagos, P., Ballabio, C., Hessel, R.
587 (Eds.), *Soil threats in Europe*. Publications Office of the European Union, Luxembourg, pp. 55-68.

588 Nakicenovic, N., Alcamo, J., Davis, G., de Vries, B., Fenham, J., Gaffin, S., 2000. IPCC special
589 report on emissions scenarios. IPCC, Cambridge.

590 Piccoli, I., Chiarini, F., Carletti, P., Furlan, L., Lazzaro, B., Nardi, S., Berti, A., Sartori, L., Dalconi,
591 M.C., Morari, F., 2016. Disentangling the effects of conservation agriculture practices on the
592 vertical distribution of soil organic carbon. Evidence of poor carbon sequestration in North-Eastern
593 Italy. *Agric., Ecosyst. Environ.* 230, 68-78.

594 Poeplau, C., Don, A., 2015. Carbon sequestration in agricultural soils via cultivation of cover
595 crops—A meta-analysis. *Agric., Ecosyst. Environ.* 200, 33-41.

596 Powlson, D.S., Whitmore, A.P., Goulding, K.W.T., 2011. Soil carbon sequestration to mitigate
597 climate change: A critical re-examination to identify the true and the false. *Eur. J. Soil Sci.* 62, 42-
598 55.

599 Regione Veneto, 2005. Carta dei suoli del Veneto alla scala 1:250,000. ARPAV – Osservatorio
600 Regionale Suolo, Castelfranco Veneto.

601 Regione Veneto, 2012. Valutazione in itinere del programma di sviluppo rurale 2007-2013 della
602 Regione Veneto. Aggiornamento Relazione di Valutazione Intermedia. [http://www.regione.veneto.
603 it/web/agricoltura-e-foreste/valutazione-psr/](http://www.regione.veneto.it/web/agricoltura-e-foreste/valutazione-psr/) (accessed 18.07.17).

604 Regione Veneto, 2013. Programma di sviluppo rurale per il Veneto 2007-2013.
605 <https://www.regione.veneto.it/web/agricoltura-e-foreste/psr-2007-2013/> (accessed 18.07.17).

606 Reijneveld, A., van Wensem, J., Oenema, O., 2009. Soil organic carbon contents of agricultural
607 land in the Netherlands between 1984 and 2004. *Geoderma* 152, 231-238.

608 Saby, N., Bellamy, P.H., Morvan, X., Arrouays, D., Jones, R.J.A., Verheijen, F.G.A., Kibblewhite,
609 M.G., Verdoodt, A.N.N., Üveges, J.B., Freudenschuß, A., Simota, C., 2008. Will European soil-
610 monitoring networks be able to detect changes in topsoil organic carbon content? *Glob. Change*
611 *Biol.* 14, 2432-2442.

612 Semenov, M A, Barrow, E.M., 2002. LARS-WG: a stochastic weather generator for use in climate
613 impact studies. Version 3.0 user manual.

614 Semenov, M.A., Pilkington-Bennett, S., Calanca, P., 2013. Validation of ELPIS 1980-2010 baseline
615 scenarios using the observed European Climate Assessment data set. *Clim. Res.* 57, 1-9.

616 Six, J., Jastrow, J.D., 2002. Organic matter turnover, in: Lal, R. (Ed.), *Encyclopedia of Soil Science*,
617 Marcel Dekker, New York, 936-942.

618 Smith, P., Powelson, D., Glendining, M., Smith, J.O., 1997. Potential for carbon sequestration in
619 European soils: Preliminary estimates for five scenarios using results from long-term experiments.
620 *Global Change Biol.* 3, 67-79.

621 Smith, P., Andrén, O., Karlsson, T., Perälä, P., Regina, K., Rounsevell, M., Wesemael, B., 2005.
622 Carbon sequestration potential in European croplands has been overestimated. *Global Change Biol.*
623 11, 2153-2163.

624 Smith, P., Martino, D., Cai, Z., Gwary, D., Janzen, H., Kumar, P., McCarl, B., Ogle, S., O'Mara, F.,
625 Rice, C., 2007a. Policy and technological constraints to implementation of greenhouse gas
626 mitigation options in agriculture. *Agric., Ecosyst. Environ.* 118, 6-28.

627 Smith, P., Martino, D., Cai, Z., Gwary, D., Janzen, H., Kumar, P., McCarl, B., Ogle, S., O'Mara, F.,
628 Rice, C., Rice, C., 2007b. Agriculture, in: Metz, B., Davidson, O.R., Bosch, P.R., Dave, R., Meyer,
629 L.A. (Eds.), *Climate change 2007: mitigation. Contribution of Working Group III to the Fourth*
630 *Assessment Report of the Intergovernmental Panel on Climate Change.* Cambridge University
631 Press: Cambridge, UK, pp. 497-540.

632 Smith, P., 2012. Agricultural greenhouse gas mitigation potential globally, in Europe and in the
633 UK: what have we learnt in the last 20 years? *Global Change Biol.* 18, 35-43.

634 Smith, P., Clark, H., Dong, H., Elsiddig, E.A., Haberl, H., Harper, R., House, J., Jafari, M., Masera,
635 O., Mbow, C., 2014. Agriculture, forestry and other land use (AFOLU), in: Edenhofer, O., Pichs-
636 Madruga, R., Sokona, Y., Farahani, E., Kadner, S., Seyboth, K., Adler, A., Baum, I., Brunner, S.,
637 Eickemeier, P., Kriemann, B., Savolainen, J., Schlömer, S., von Stechow, C., Zwickel, T., Minx,
638 J.C. (Eds.), *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group*
639 *III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge
640 University Press, Cambridge, pp. 811-922.

641 Stockmann, U., Padarian, J., McBratney, A., Minasny, B., de Brogniez, D., Montanarella, L., Hong,
642 S.Y., Rawlins, B.G., Field, D.J., 2015. Global soil organic carbon assessment. *Glob. Food Sec.* 6, 9-
643 16.

644 Troldborg, M., Aalders, I., Towers, W., Hallett, P.D., McKenzie, B.M., Bengough, A.G., Lilly, A.,
645 Ball, B.C., Hough, R.L., Hough, R.L., 2013. Application of Bayesian Belief Networks to quantify
646 and map areas at risk to soil threats: Using soil compaction as an example. *Soil Till. Res.* 132, 56-
647 68.

648 Uusitalo, L., 2007. Advantages and challenges of Bayesian networks in environmental modelling.
649 *Ecol. Model.* 203, 312-318.

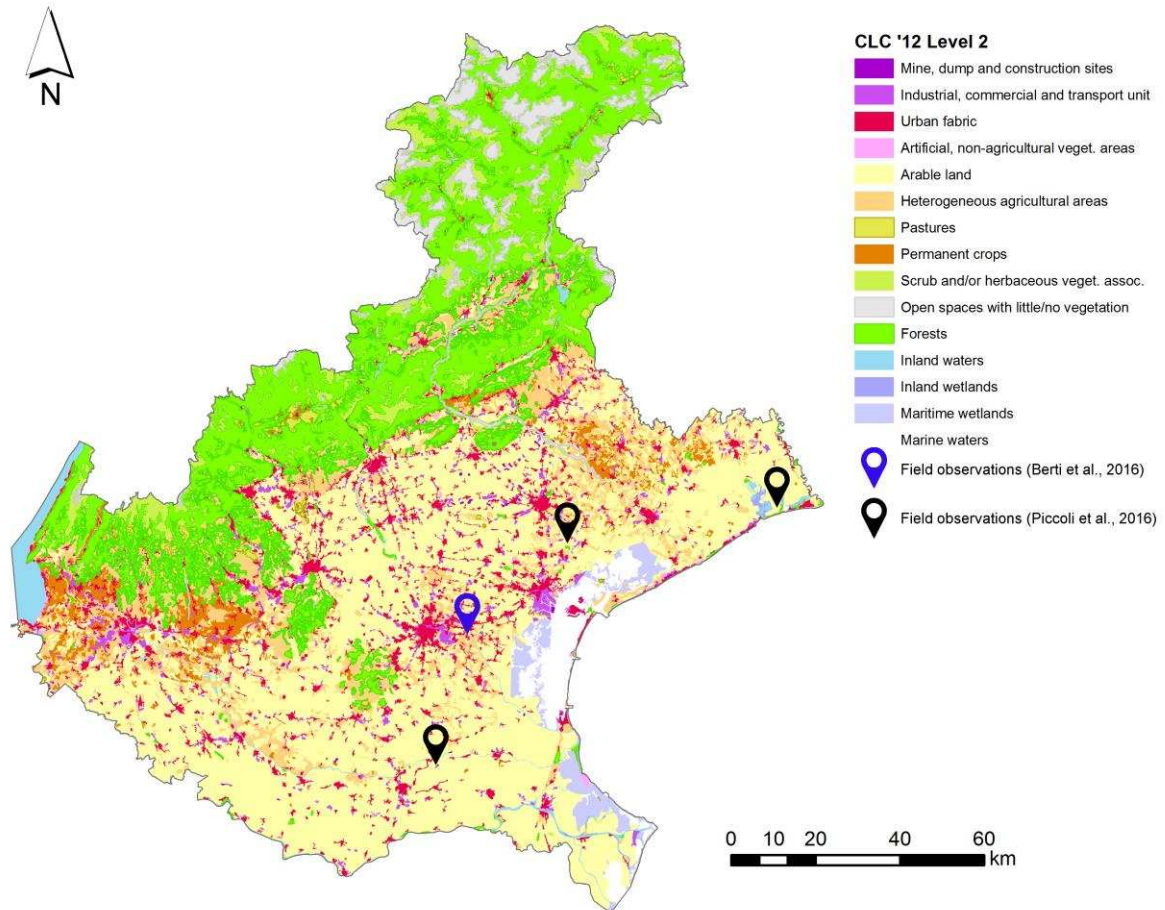
650 World Reference Base for Soil Resources, IUSS Working Group, 2014. *International Soil*
651 *Classification System for Naming Soils and Creating Legends for Soil Maps*. FAO, World Soil
652 Resources Reports No. 106. Rome.

653 Xu, X., Liu, W., Kiely, G., 2011. Modeling the change in soil organic carbon of grassland in
654 response to climate change: effects of measured versus modelled carbon pools for initializing the
655 Rothamsted Carbon model. *Agric., Ecosyst. Environ.* 140, 372-381.

656 Yigini, Y., Panagos, P., 2016. Assessment of soil organic carbon stocks under future climate and
657 land cover changes in Europe. *Sci. Total Environ.* 557, 838-850.

658 **List of figures**

659 **Figure 1** - Veneto region study area according to 2-level Corine Land Cover inventory (2012).



660

661

662

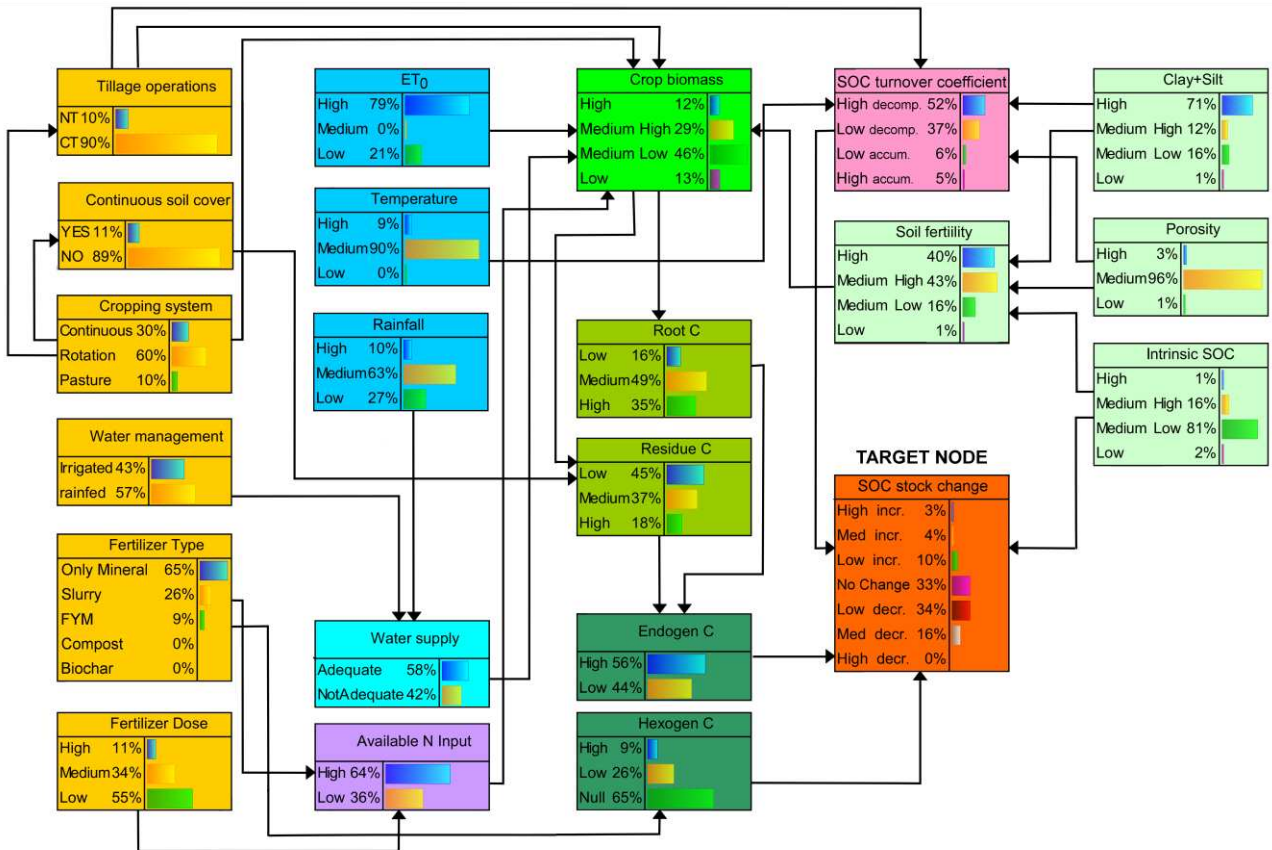
663

664

665

666

667 **Figure 2** - Bayesian belief network showing factors determining SOC stock change in the 0-30 cm
 668 soil layer. Each node represents a specific factor that, interacting with other factors, influences the
 669 SOC stock change. The arrows represent the cause-and-effect direction between nodes. Each node
 670 can have a range of values (e.g. high, medium, low), each associated to a conditional probability.



671

672

673

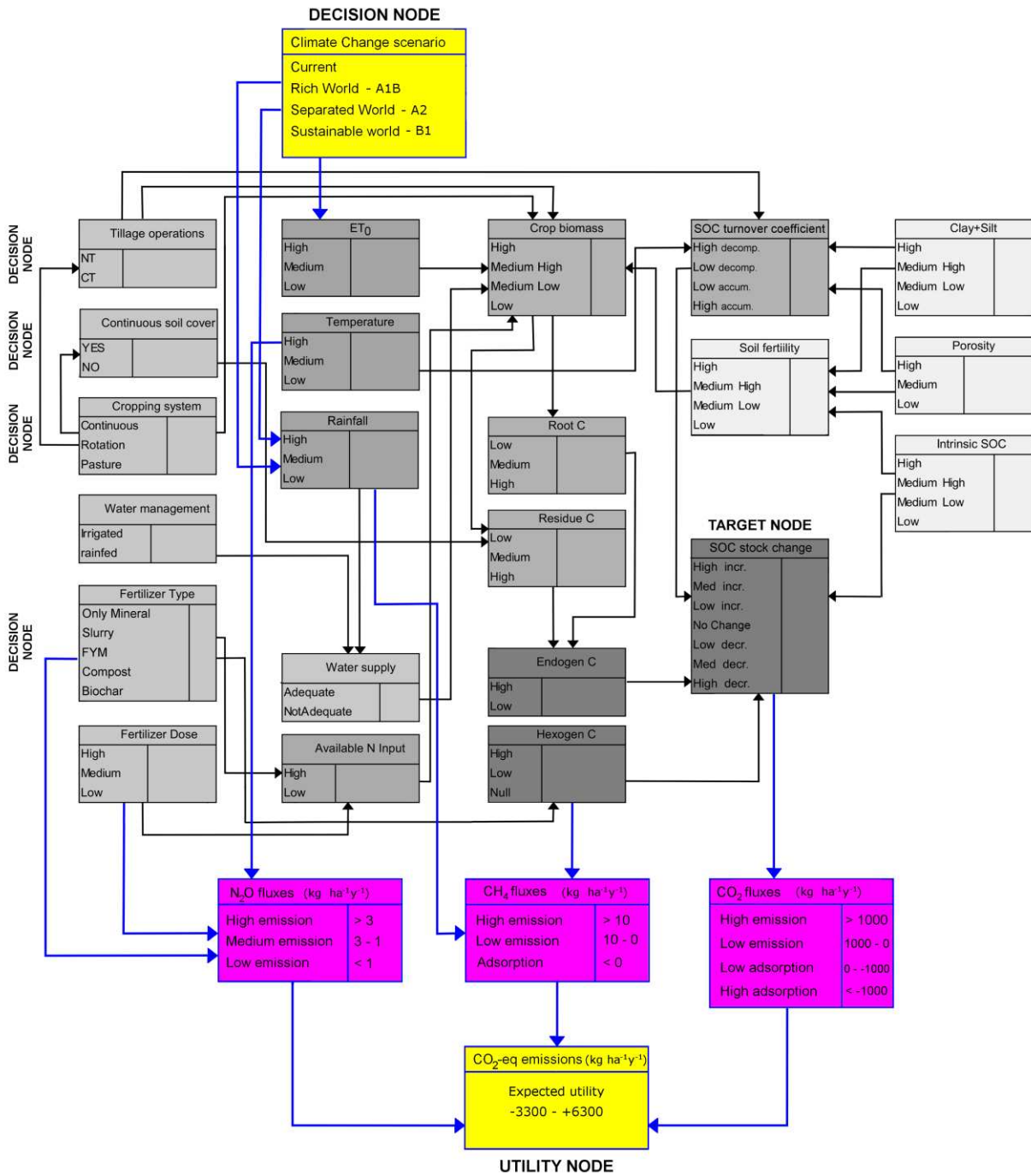
674

675

676

677

678 **Figure 3** - BBN with utility values for climate change emissions scenarios.



679

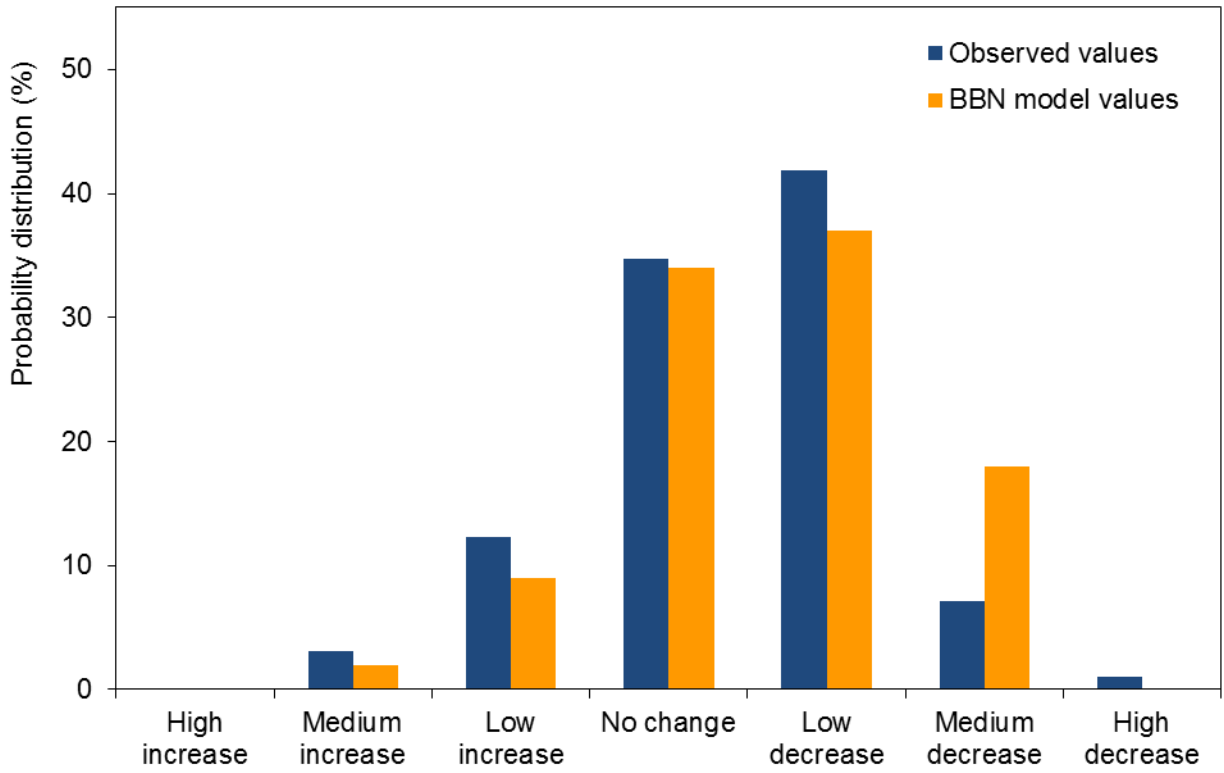
680

681

682

683

684 **Figure 4** - Comparison of SOC stock change probability distributions as a result of field surveys
685 and BBN modelling.



686

687

688

689

690

691

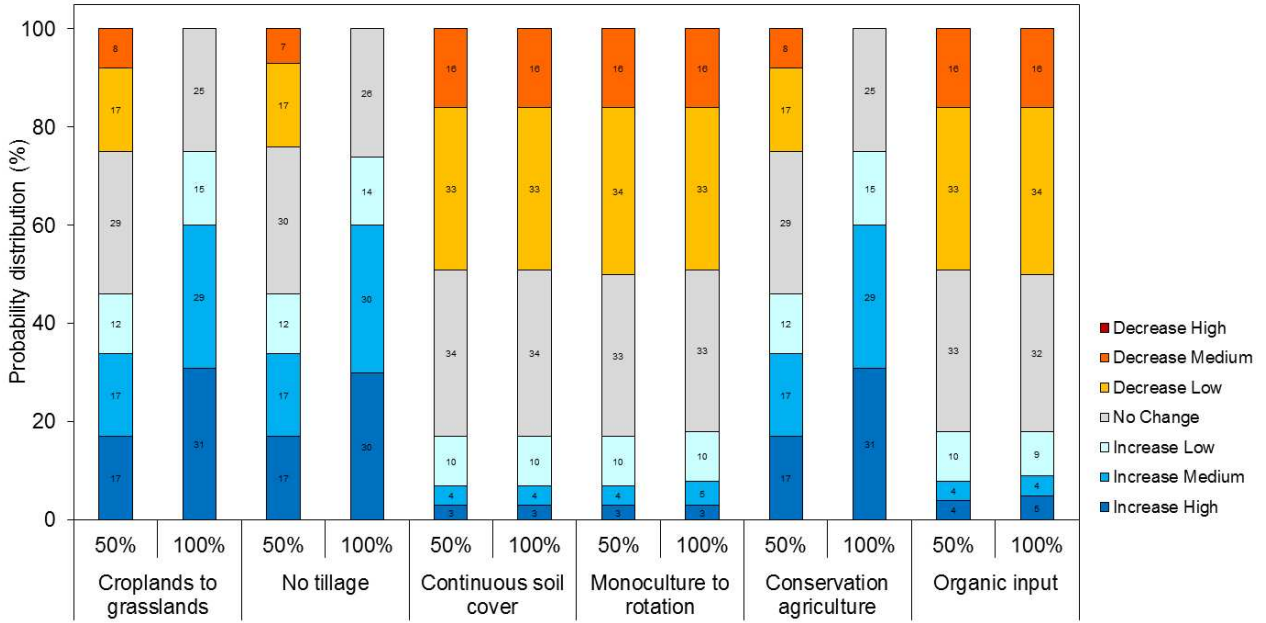
692

693

694

695

696 **Figure 5** - SOC stock change probability distribution under different land use and management
 697 scenarios.



698

699 **Table 1** Description of nodes included in the BBN their state values to evaluate SOC stock change.

	Node	State value	Value/Description	Type of information
Pedo-climatic nodes	Intrinsic SOC content (g kg ⁻¹)	High	> 40	Soil map (Regione Veneto, 2005)
		Medium high	40 – 20	
		Medium Low	20 – 10	
		Low	< 10	
	Soil porosity (m ³ m ⁻³)	High	> 0.55	Soil map (Regione Veneto, 2005)
		Medium	0.55 – 0.40	
		Low	< 0.40	
	Clay + Silt (kg kg ⁻¹)	High	> 0.6	Soil map (Regione Veneto, 2005)
		Medium high	0.6 – 0.4	
		Medium low	0.4 – 0.2	
		Low	< 0.2	
	ET ₀ (mm)	High	> 1000	derived from Penman-Monteith equation on data from the Environmental Protection Agency (ARPAV)
		Medium	1000 – 800	
		Low	< 800	
Rainfall (mm)	High	> 1200	Environmental Protection Agency (ARPAV)	
	Medium	1200 – 1000		
	Low	< 1000		
Temperature (°C)	High	> 13	Environmental Protection Agency (ARPAV)	
	Low	< 13		
Management nodes	Crop system	Grassland		Regione Veneto (2012)
		Rotation		
	Monoculture			
Fertiliser type	Mineral		Regione Veneto (2012)	
	Slurry			
	Farmyard manure			
	Biochar			
N fertiliser dose (kg ha ⁻¹ y ⁻¹)	Compost		Regione Veneto (2012)	
	High			> 340
	Medium			340 – 170

	Tillage operation	Low Tillage No tillage	< 170	Regione Veneto (2013)
	Continuous soil cover	Yes No		Regione Veneto (2013)
	Water management	Irrigated Rainfed		ISTAT, 2010
Child nodes	Available N input (kg ha ⁻¹)	High	> 200	Expert opinion
		Low	< 200	
	Crop biomass (Mg ha ⁻¹ d.m.)	High	> 30	Dal Ferro et al., 2016
		Medium high	30 – 20	
		Medium low	20 – 10	
	Endogen OC input (Mg ha ⁻¹ y ⁻¹)	Low	< 10	Expert opinion
		High	> 4.0	
	Hexogen OC input (Mg ha ⁻¹ y ⁻¹)	Low	< 4.0	Expert opinion
		High	> 4.0	
	Root carbon (Mg ha ⁻¹ y ⁻¹)	Low	0.0 – 4.0	Expert opinion
		Null	0.0	
		High	> 4.0	
	Residue carbon (Mg ha ⁻¹ y ⁻¹)	Medium	4.0 – 2.0	Expert opinion
		Low	< 2.0	
		High	> 4.0	
SOC turnover coefficient (y ⁻¹)	Medium	4.0 – 2.0	Six and Jastrow, 2002	
	Low	< 2.0		
	High decomposition	> 0.02		
Soil fertility	Low decomposition	0.0 – 0.02	Literature review; Expert opinion	
	Low accumulation	0.0 – -0.02		
	High accumulation	< -0.02		
		High		
		Medium high		
		Medium low		
		Low		

Water supply	Adequate		Literature review; Expert opinion
	Not adequate		
SOC stock change (Mg ha ⁻¹ y ⁻¹)	High increase	> 1.0	
	Medium increase	1.0 – 0.5	
SOC stock change (Mg ha ⁻¹ y ⁻¹)	Low increase	0.5 – 0.1	Dal Ferro et al., 2016
	No change	0.1 – -0.1	
	Low decrease	-0.1 – -0.5	
	Medium decrease	-0.5 – -1.0	
	High decrease	< -1.0	

700

701

702

703

704

705

706

707

708

709

710

711

712

713

714 **Table 2** One-way sensitivity analysis of posterior probabilities for the node SOC stock change.

715

Order	Node	Sensitivity node
1	Cropping system	0.374
2	Tillage operations	0.226
3	Intrinsic SOC	0.139
4	SOC turnover coefficient	0.049
5	Fertiliser type	0.027
6	Clay+Silt	0.021
7	Endogen C	0.016
8	Porosity	0.015
9	Residue C	0.010
10	Hexogen C	0.009
11	Temperature	0.006
12	Fertiliser dose	0.005
13	Soil cover	0.004
14	Root C	0.004
15	Rainfall	0.001
16	Water management	0.001
17	Water supply	0.001
18	Soil fertility	0.001
19	Crop biomass	0.001
20	ET ₀	0.000
21	Available N input	0.000

716

717

718

719

720 **Table 3** Utility values of equivalent CO₂ emissions (CO₂-eq, kg ha⁻¹ y⁻¹) under different land use and management and climate scenarios. The
 721 higher are the values, the greater are the CO₂-eq emissions.

722

Land use and management	Area investment	Climate scenarios			
		Current	Rich – A1B	Separate – A2	Sustainable – B1
Standard		1613.9	1647.2	1646.3	1647.2
Croplands to grasslands	50%	311.4	361.9	361.9	361.9
	100%	-991.0	-923.4	-922.4	-923.4
No tillage	50%	326.7	378.1	378.1	378.1
	100%	-972.9	-904.3	-904.3	-904.3
Continuous soil cover	50%	1617.7	1651.0	1651.0	1651.0
	100%	1621.5	1656.7	1656.7	1656.7
Monoculture to rotation	50%	1613.9	1647.2	1647.2	1646.3
	100%	1612.0	1645.3	1645.3	1645.3
Conservation agriculture	50%	324.8	376.2	376.2	376.2
	100%	-990.1	-923.4	-923.4	-923.4
Organic input	50%	1604.3	1643.4	1643.4	1643.4
	100%	1558.6	1588.1	1588.1	1588.1

723

724

725