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**Article:**

Alzraiee, A.H., Bailey, R. and Bau, D. [orcid.org/0000-0002-0730-5478](https://orcid.org/0000-0002-0730-5478) (2017) Assimilation of Historical Head Data to Estimate Spatial Distributions of Stream Bed and Aquifer Hydraulic Conductivity Fields. *Hydrological Processes*, 31 (7). pp. 1527-1538. ISSN 0885-6087

<https://doi.org/10.1002/hyp.11123>

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1 **Assimilation of Historical Head Data to Estimate Spatial Distributions of Stream Bed and**  
2 **Aquifer Hydraulic Conductivity Fields**

3

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18 September 2016

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20 Accepted version.

21 Published in: **Hydrological Processes, Volume 31, Issue 7, 30 March 2017, pp. 1527–1538**

1 **Abstract**

2 Management of water resources in alluvial aquifers relies mainly on understanding interactions  
3 between hydraulically connected streams and aquifers. Numerical models that simulate this  
4 interaction often are used as decision support tools in water resource management. However, the  
5 accuracy of numerical predictions relies heavily on the unknown system parameters (i.e. stream  
6 bed conductivity and aquifer hydraulic conductivity) which are spatially heterogeneous and  
7 difficult to measure directly. This paper employs an Ensemble Smoother to invert groundwater  
8 level measurements to jointly estimate spatially-varying streambed and alluvial aquifer hydraulic  
9 conductivity along a 35.6 km segment of the South Platte River in northeastern Colorado. The  
10 accuracy of the inversion procedure is evaluated using a synthetic experiment and historical  
11 groundwater level measurements, with the latter constituting the novelty of this study in the  
12 inversion and validation of high resolution fields of streambed and aquifer conductivities.  
13 Results show that the estimated streambed conductivity field and aquifer conductivity field  
14 produce an acceptable agreement between observed and simulated groundwater levels and  
15 stream flow rates. The estimated parameter fields are also used to simulate the spatially varying  
16 flow exchange between the alluvial aquifer and the stream, which exhibit high spatial variability  
17 along the river reach with a maximum average monthly aquifer gain of about  $2.3 \text{ m}^3/\text{day}$  and a  
18 maximum average monthly aquifer loss of  $2.8 \text{ m}^3/\text{day}$ , per unit area of streambed ( $\text{m}^2$ ). These  
19 results demonstrate that data assimilation inversion provides a reliable and computationally  
20 affordable tool to estimate the spatial variability of streambed and aquifer conductivities at high  
21 resolution in real-world systems.

1 **1. Introduction**

2 Exchange of water between groundwater systems and surface water systems can have a  
3 significant impact on biogeochemical nutrient cycling in the hyporheic zone (Frei et al., 2009;  
4 Kurtz et al., 2012), riparian zone ecology (Cey et al., 1999) and processes (e.g. vegetation  
5 growth, nutrient flux), environmental flows and associated habitat quality, mass flux of solutes  
6 between aquifer and streams (Hussein and Schwartz, 2003; Kalbus et al., 2007), and the general  
7 water balance of the stream-aquifer system (Frei et al., 2009; Kurtz et al., 2012). For the latter,  
8 water management practices can be dependent on groundwater-surface water exchange, for  
9 example significant groundwater recharge in losing reaches of a stream or stream depletion due  
10 to nearby alluvial groundwater pumping (Glover and Balmer, 1954; Jenkins, 1968; Sophocleous  
11 et al., 1995; Chen and Shu, 2002; Miller et al., 2007).

12 Fluxes between groundwater and surface water, either through groundwater discharge to  
13 streams or stream water seepage into aquifers, are governed by the position of stream stage with  
14 respect to the water table, the geometry and position of the stream channel within the alluvial  
15 plain, and the hydraulic properties of the aquifer and the streambed (Woessner, 2000; Cardenas  
16 et al., 2004). Of these, hydraulic conductivity ( $K_s$ ) of streambed sediments along the aquifer-  
17 stream interface often is the principal control, with exchange fluxes often being highly spatially  
18 variable (sometimes on the order of meters to centimeters) due to strong spatial heterogeneity of  
19 streambed  $K_s$  (Fleckenstein et al., 2006; Frei et al., 2009; Kalbus et al., 2009; Rosenberry and  
20 Pitlick, 2009; Vogt et al., 2010). Streambed  $K_s$  can range over orders of magnitude (Calver,  
21 2001) over relatively short (0.2 km to 10 km) reaches of a stream (Genereux et al. 2008; Hatch et  
22 al., 2010). Heterogeneity of aquifer properties also can have a strong impact on stream-aquifer  
23 exchange (Kalbus et al., 2009). In general, assuming complete or partial spatial uniformity in

1 streambed  $K_s$  can yield erroneous estimates of groundwater discharge and stream flow depletion  
2 (Kurtz et al., 2013; Lackey et al., 2015), with important implications for water management in  
3 coupled stream-aquifer systems. As such, a key objective in investigating groundwater-surface  
4 interactions is an accurate estimation of spatially-varying streambed  $K_s$  along a river reach.

5 Numerous methods have been employed to estimate spatially-variable streambed  $K_s$ , with the  
6 overall goal of providing reliable estimates of exchange flux in space and time. These methods  
7 include permeameter tests and seepage meters (Avery, 1994; Duff et al., 2000; Paulsen et al.,  
8 2001); electrical resistivity surveys of streambed sediment (Nyquist et al., 2008); streambed  
9 temperature mapping, vertical temperature profiling and heat transport modeling (Silliman and  
10 Booth, 1993; Silliman et al., 1995; Fryar et al., 2000; Becker et al., 2004; Keery et al., 2007; Vogt  
11 et al., 2010; Kurtz et al., 2014); water balance approaches (Krause et al., 2007); and the use of  
12 numerical groundwater models (Morway et al., 2013) or coupled surface-subsurface hydrologic  
13 models (Frei et al., 2009). For numerical models, streambed  $K_s$  is varied spatially to provide  
14 matches between observed and simulated hydraulic head data and stream stage data. As  
15 identified in recent studies, there is a need to assess streambed  $K_s$  at larger scales (i.e. longer  
16 reaches of streams) (Frei et al., 2009) while still targeting sufficient spatial resolution  
17 (Fleckenstein et al., 2010).

18 As an alternative to these methods, numerical hydrologic modeling coupled with data  
19 assimilation methods can be used to estimate spatially-varying streambed  $K_s$  along the stream-  
20 aquifer interface. Data assimilation methods such as the Kalman Filter and variants such as the  
21 Ensemble Kalman Filter (EnKF) (Evensen, 1994; Burgers et al., 1998) and the Ensemble  
22 Smoother (ES) (van Leeuwen and Evensen, 1996) have been used in numerous hydrologic  
23 studies to estimate aquifer hydraulic conductivity and transmissivity (Hantush and Marino, 1997;

1 Chen and Zhang, 2006; Hendricks Franssen and Kinzelbach, 2008; Alzraiee et al., 2014), first-  
2 order reaction rates of solutes (Bailey and Baù, 2011; Bailey et al., 2013), and aquifer  
3 dispersivity (Wagner, 1992; Lui et al., 2008). In these methods, system-response variables (e.g.  
4 groundwater hydraulic head, groundwater solute concentration) and system parameters (e.g.  
5 streambed conductance and hydraulic conductivity field) are jointly updated by assimilating  
6 measurement data from the true state.

7 Several recent studies (Hendricks Franssen et al., 2011; Kurtz et al., 2012; Kurtz et al., 2013),  
8 all applied to the Limmat Aquifer system near Zurich, Switzerland, used the EnKF to jointly  
9 update aquifer hydraulic conductivity,  $K_a$ , and streambed  $K_s$ . Using a variably saturated  
10 groundwater flow model with stream-aquifer interactions, Hendricks Franssen et al. (2011) and  
11 Kurtz et al. (2012) estimated stream leakage coefficients in five zones by assimilating hydraulic  
12 head data, with the latter study estimating temporal-varying stream bed  $K_s$ . Kurtz et al. (2013)  
13 estimated stream bed  $K_s$  in a synthetic system in settings of varying degrees of heterogeneity,  
14 ranging from two  $K_s$  zones to a fully heterogeneous system wherein each stream node received a  
15 different value of  $K_s$ .

16 The overall objective of this study is to jointly estimate the spatial variability of streambed  
17 conductivity,  $K_s$ , and aquifer conductivity,  $K_a$ , at relatively high resolutions (304.8 m) within a  
18 regional-scale river-aquifer system using historical data. Specifically, the Ensemble Smoother  
19 (ES) (van Leeuwen and Evensen, 1996) is used to estimate spatially-varying fields of aquifer  $K_a$   
20 and streambed  $K_s$  within a 35.6 km reach of the South Platte River in northeastern Colorado via  
21 assimilation of time series of hydraulic head data from nearby observation wells. Following a  
22 demonstrative example using synthetic head data, historical measurements are used to estimate  
23 the parameter fields. The performance of the parameter inversion is evaluated using historical

1 data from observation wells not used in parameter estimation, and the posterior uncertainty in the  
2 predicted stream-aquifer flux exchanges are quantified.

3 To our knowledge, this is the first study to use an Ensemble Smoother to assimilate historical  
4 hydraulic head data to estimate and corroborate strongly heterogeneous streambed  $K_s$  and aquifer  
5  $K_a$ . The methodology presented herein can be transferred to stream-aquifer systems in other  
6 alluvial river valleys.

## 7 **2. Site Description**

8 The South Platte Basin (Figure 1) covers approximately 21% of the State of Colorado (about  
9 57,000 km<sup>2</sup>), within which the South Platte River Basin alluvial groundwater system constitutes  
10 19% (about 10,400 km<sup>2</sup>) (Colorado Geological Survey 2003). As of 2008, the irrigated farmland  
11 was 335,000 ha, to support a population of 3.5 million (CDM Smith, 2013).

12 The surface hydrological system consists of the main stem of the South Platte River and its  
13 tributaries (Figure 1B). The alluvial deposits in the South Platte Basin consist mainly of sand  
14 and gravels. The alluvial aquifer is believed to be hydraulically connected to the surface water  
15 system throughout much of the basin (CDM Smith, 2013). The saturated thickness of the  
16 alluvial aquifer generally increases along the downstream direction (west to east), with saturated  
17 aquifer thicknesses ranging between 6 and 90 m. The aquifer hydraulic conductivity  $K_a$  ranges  
18 between approximately 30 and 600 m/day, depending on the degree of sorting and the amount  
19 of fine grain material present (CDM Smith, 2013). Agricultural irrigation is the dominant water  
20 use in the South Platte River Basin (CDM Smith, 2013).

21 In a joint effort of the Colorado Water Conversation Board and the Colorado Division of  
22 Water Resources (DWR) and as a part of the South Platte Decision Support System (SPDSS), a  
23 large-scale regional groundwater model based on MODFLOW (Harbaugh et al. 2000) was

1 developed and calibrated for a large portion of the alluvial aquifer by CDM Smith (2013). The  
2 modeled area (Figure 1A) is about 63% of the alluvial aquifer (6,400 km<sup>2</sup>), and the simulation  
3 time period is between 1950 and 2006. More information about the model is provided in  
4 Section 4.

5 In this paper and for the purpose of high-resolution parameter estimation, we focus on a  
6 smaller portion of the alluvial aquifer as shown in Figure 1B. The simulated area extends over 30  
7 km in the east-west direction along the South Platte River between the towns of Snyder and  
8 Atwood. The length of the river stem in the study area is about 35.6 km.

### 9 **3. Methodology**

#### 10 **3.1 Formulation of the Inverse Problem**

11 The rate of flux exchange between streams network and aquifer depends largely on the  
12 difference between stream stage and local water table elevation. The numerical simulation of this  
13 interaction is based on coupling the groundwater continuity equation with the stream water  
14 continuity and momentum equations. This coupling is achieved in MODFLOW's Streamflow-  
15 Routing (SFR) package (Prudic et al. 2004) by calculating the stream depth at the midpoint of  
16 each reach and assuming uniform flow between streams and aquifer over a given section of the  
17 stream and the corresponding volume of aquifer. Streamflow routing in SFR is modeled using  
18 the continuity equation and by assuming that streamflow is steady in discrete time periods, and  
19 uniform within each numerical cellblock.

20 Depending on the elevation of stream stage with respect to the elevation of the water table in  
21 the local aquifer, a stream can be either gaining or losing. The stream is gaining when the water  
22 table is above the stream stage elevation; in this case, the exchange flow rate is computed as:



1 
$$Q_{sa} = \frac{K_s w L}{m} (h_s - h_a) \quad (1)$$

2 where  $Q_{sa}$  is the water exchange flow rate between a given section of the stream and the local  
 3 aquifer [ $L^3 T^{-1}$ ],  $K_s$  is the hydraulic conductivity of the streambed sediments [ $L T^{-1}$ ],  $w$  is the  
 4 stream width [ $L$ ],  $L$  is the stream length in the finite difference cell [ $L$ ],  $m$  is the thickness of  
 5 stream bed deposits [ $L$ ],  $h_s$  is the head in the stream [ $L$ ], and  $h_a$  is the head in the aquifer beneath  
 6 the streambed [ $L$ ]. When the water table is below the streambed elevation, MODFLOW-SFR  
 7 package assumes that water exchange flow rate is independent of  $h_a$ . In this case, the stream-  
 8 aquifer flow is calculated by assuming head difference equal to the streambed thickness.

9 Assuming a constant streambed thickness in Equation (1),  $K_s$  and  $h_a$  (controlled by the  $K_a$   
 10 field) are the principle controlling factors of water exchange rate, and are spatially heterogeneous  
 11 fields that cannot be uniquely determined from a finite number of field samples and associated  
 12 parameter measurements. Alternatively, inverse modeling allows incorporating relatively low  
 13 cost measurements of water table elevation and stream flow rate to predict these parameters.

14 To simplify the illustration of inverse modeling for this problem, consider the following  
 15 generic model that relates an observable vector  $\mathbf{d}$  to a vector of high-dimensional input  
 16 parameters  $\mathbf{m}$ ,

17 
$$\mathbf{d} = \mathbf{G}(\mathbf{m}) \quad (2)$$

18 where  $\mathbf{d}$  is  $n_d \times 1$  vector that encompasses predicted states (e.g. hydraulic heads  $h_x^t$ ) at a set of  
 19 observable spatial locations  $x$  and at a set of times  $t$ ;  $\mathbf{m}$  is a vector with dimension  $n_m \times 1$  that  
 20 encompasses system parameters that controls observable states, and  $\mathbf{G}$  is a generic flow model  
 21 that maps input parameters to observable states.

1 In this study, we assume that uncertainty in stream-aquifer interaction is mainly attributed to  
 2 the unknown streambed  $K_s$  and aquifer  $K_a$  fields. Other factors affecting the interactions, such as  
 3 groundwater stresses and boundary conditions, are determined from field measurements and the  
 4 calibrated regional model as discussed in Section 4.1. Thus, the vector of parameters to be  
 5 determined can be written as  $\mathbf{m} = [K_s, K_a]^T$ .

6 Inverse modeling of high-dimensional parameters is usually affected by the problem of non-  
 7 uniqueness (Beven 2001), which occurs when a small number of observations are used to  
 8 estimate a larger number of system parameters. For this situation, an infinite number of solutions  
 9 to the inverse problem are possible. More realistically, all possible parameter solutions fitting to  
 10 a probability distribution function (PDF) conditional to a set of observations may be described  
 11 using Bayes' law:

$$12 \quad P(\mathbf{m} | \mathbf{d}_o) = \frac{P(\mathbf{d}_o | \mathbf{m})P(\mathbf{m})}{P(\mathbf{d}_o)} \quad (3)$$

13 where  $P(\mathbf{m} | \mathbf{d}_o)$  is the posterior probability of model parameters  $\mathbf{m}$  given a vector of  
 14 observations  $\mathbf{d}_o$ ,  $P(\mathbf{d}_o | \mathbf{m})$  is the likelihood probability distribution,  $P(\mathbf{m})$  is the prior model  
 15 parameter distribution, and  $P(\mathbf{d}_o)$  is a normalization term.

16 One of the few analytical solutions that can be obtained from Bayes' law occurs when the  
 17 forward model  $\mathbf{G}$  is linear and the PDF of system parameters in Equation (3) is multivariate  
 18 Gaussian,  $P(\mathbf{m}) \sim N(\bar{\mathbf{m}}, \mathbf{C}_m)$ , where  $\bar{\mathbf{m}}$  is the prior mean of parameters vector and  $\mathbf{C}_m$  is the  
 19 parameter prior covariance matrix. In this case, the posterior distribution also follows a Gaussian  
 20 distribution, i.e.  $P(\mathbf{m} | \mathbf{d}_o) \sim N(\hat{\mathbf{m}}, \hat{\mathbf{C}}_m)$ , where the posterior mean vector and posterior covariance  
 21 matrix are computed as follows:

$$\hat{\mathbf{m}} = \bar{\mathbf{m}} + \mathbf{C}_{md} \mathbf{C}_{dd}^{-1} (\mathbf{d}_o - \bar{\mathbf{m}}) \quad (4)$$

$$\hat{\mathbf{C}}_m = \mathbf{C}_m - \mathbf{C}_{md} \mathbf{C}_{dd}^{-1} \mathbf{C}_{md} \quad (5)$$

where  $\hat{\mathbf{m}}$  is the mean of the posterior Gaussian PDF of system parameters,  $\mathbf{C}_{md}$  is a  $n_m \times n_d$  matrix that describes the cross-covariance between system parameters and observable states,  $\mathbf{C}_{dd}$  is the auto-covariance matrix of the observable states and has a dimension of  $n_d \times n_d$ , and  $\hat{\mathbf{C}}_m \in n_m \times n_m$  is the posterior covariance matrix of system parameters.

In practice, the assumptions of model linearity and parameter Gaussianity restrict the wide applications of this formulation. Additionally, it is computationally intensive to compute the parameter-state cross-covariance matrix for high-dimensional models. Evensen (1994) proposed an ensemble-based formulation of the Kalman Filter for high-dimensional problems. In this formulation, the prior PDF is approximated using an ensemble of parameter-state realizations produced through a Monte Carlo simulation by:

$$\mathbf{X}_p = \begin{bmatrix} \mathbf{m}_1 & \dots & \mathbf{m}_N \\ \mathbf{d}_1 & \dots & \mathbf{d}_N \end{bmatrix} \quad (6)$$

where  $N$  is the number of realizations in the ensemble and  $\mathbf{X}_p$  is the parameter-state forecast (prior) matrix with dimensions  $(n_m + n_d) \times N$ . Using this matrix, the prior ensemble covariance matrix can be calculated as

$$\mathbf{C}_p = \frac{(\mathbf{X}_p - \hat{\mathbf{X}}_p)(\mathbf{X}_p - \hat{\mathbf{X}}_p)^T}{N-1} \quad (7)$$

1 where  $\hat{\mathbf{X}}_p$  is a matrix with dimension  $(n_m + n_d) \times N$  where each column is the prior ensemble  
 2 mean vector. Following Equations (4) and (5), the update forecast matrix and update covariance  
 3 matrix can be written as follows:

$$4 \quad \mathbf{X}_u = \mathbf{X}_p + \mathbf{\Phi}(\mathbf{D} - \mathbf{H}\mathbf{X}_p) \quad (8)$$

$$5 \quad \mathbf{C}_u = (\mathbf{I} - \mathbf{K}\mathbf{H})\mathbf{C}_p \quad (9)$$

6 where  $\mathbf{X}_u$  is the update parameter-state matrix,  $\mathbf{D}$  is the perturbed measurements matrix with  
 7 dimension  $n_d \times N$ ,  $\mathbf{H}$  is a binary matrix ( $n_d \times n_m$ ) that is used to extract model predictions at  
 8 locations and times of observations data,  $\mathbf{C}_u$  is the update covariance matrix and  $\mathbf{I}$  is the identity  
 9 matrix.  $\mathbf{\Phi}$  is the so-called Kalman Gain matrix ( $n_m \times n_d$ ), computed as:

$$10 \quad \mathbf{\Phi} = \mathbf{C}_p \mathbf{H}^T (\mathbf{H}\mathbf{C}_p \mathbf{H}^T + \mathbf{R})^{-1} \quad (10)$$

11 In equation (10),  $\mathbf{R}$  is the covariance matrix  $n_d \times n_d$  of measurement errors computed from  
 12 uncorrelated error realizations generated from  $\sim N(0, \sigma_e)$ , where  $\sigma_e$  is the error standard  
 13 deviation that reflects confidence in measurements. When the observation vector  $\mathbf{d}_o$  incorporates  
 14 data at multiple times, i.e.  $\mathbf{d}_o = [\mathbf{d}_{t=1}, \dots, \mathbf{d}_{t=n_t}]$ , the objective of the Bayesian update is to compute  
 15 the posterior distribution  $P(\mathbf{m} | \mathbf{d}_{t=1}, \dots, \mathbf{d}_{t=n_t})$ , where  $n_t$  is the number of temporal measurements.  
 16 For this situation, it is straightforward to expand the Ensemble Kalman Filter to the Ensemble  
 17 Smoother (ES) (van Leeuwen and Evensen, 1996) that assimilates all available measurements  
 18 from any time into a single update step. To implement the ES, the forecast matrix incorporates  
 19 parameters and model responses at all observable locations and times as follows:

$$\mathbf{X}_p = \begin{bmatrix} \mathbf{m}_1 & \dots & \mathbf{m}_N \\ \mathbf{d}_1^{t=1} & & \mathbf{d}_N^{t=1} \\ \vdots & \dots & \vdots \\ \mathbf{d}_1^{t=n_t} & & \mathbf{d}_N^{t=n_t} \end{bmatrix} \quad (11)$$

2 The forecast matrix in Equation (11) is used to calculate the spatio-temporal cross-covariance  
3 matrix using Equation (7). Similar to the EnKF, Equations (8) to (10) can be used to achieve the  
4 update in ES.

## 5 **4. Application**

### 6 **4.1 Model Settings and Field Observations**

7 The model input data are extracted from the regional SPSS MODFLOW groundwater  
8 model (CDM Smith, 2013). The numerical model domain consists of a single layer and 100  
9 columns and 100 rows. Each cell is 304.8 m x 304.8 m (1000 ft). The saturated thickness in the  
10 simulated area ranges between 60 m to 88 m along the river pathway and decreases away from  
11 the river to a minimum of about 15 m along the edges of the alluvial aquifer.

12 The extracted simulated period spans from 2000 to 2006. The system stresses include: (a)  
13 spatially variable recharge, accounting for deep percolation resulting from precipitation,  
14 irrigation return flow, and seepage from ditches and canals (CDM Smith, 2013); (b) spatially  
15 variable evapotranspiration (ET) computed internally by the MODFLOW-ETS package as a  
16 function of groundwater depth and measured reference ET; and (c) about 500 pumping and  
17 injection wells with flow rates changing seasonally. The system stresses change monthly to allow  
18 for seasonal variation. Thus, the simulation period is divided into 84 transient state stress-periods  
19 and one steady-state simulation in the first month of 2000.

1 The upstream and downstream boundary conditions are chosen to be Neumann-type time  
2 variable lateral flow conditions. The monthly groundwater lateral flow rates are extracted from  
3 the regional groundwater flow model using a zonal mass balance analysis for the study area,  
4 resulting in a generally west-to-east flow regime (Figure 2). The boundary conditions on the  
5 northern and southern sides of the model are simulated to be variable flux and are obtained from  
6 the regional groundwater model as well.

## 7 **4.2 Numerical Experiments**

8 Implementing Data Assimilation techniques for system parameter estimation is performed in  
9 two stages: a forecast or simulation stage, and an update or assimilation stage.

10 In the first stage (forecast), a Monte Carlo simulation is conducted in which a number of  
11 realizations of streambed  $K_s$  and the aquifer  $K_a$  fields are generated and processed in the flow  
12 model. To generate prior realizations,  $y = \log(K_a)$  is assumed to fit to an isotropic and  
13 stationary Gaussian process (de Marsily 1986) with a prescribed covariance model  
14  $C_{yy}(d; y, y)$ . Similarly,  $z = \log(K_s)$  is assumed to fit to a one-dimensional correlated random  
15 process representing the spatial variability of streambed conductivity  $C_{zz}(d; z, z)$  along the  
16 stream pathway. The parameters  $\lambda$  and  $\sigma^2$  represent the correlation length [L] and the variance of  
17 the random processes, respectively. The stationary means of the two fields are  $\mu_y$  and  $\mu_z$ .

18 Table 1 summarizes the geostatistical properties of the two fields. In this study, a spherical  
19 covariance function is assumed for both  $C_{yy}$  and  $C_{zz}$ , yet other covariance functional forms can  
20 also be used. The number of generated realizations for both  $K_s$  and  $K_a$  fields is 500. The range of  
21 spatial variability of the specific yield is typically narrow, thus this parameter is assumed  
22 homogenous with a value of 0.2 (CDM Smith, 2013).

1 In the second stage (update), the system parameters and states are updated using Equations 8  
 2 and 9. The update ensemble can be used to quantify the uncertainty in posterior estimates.

3 Two sets of experiments are performed in this study (Table 2). The first experiment, termed  
 4 A, provides an initial test of the methodology by assimilating synthetic hydraulic head values  
 5 generated from a flow model that simulates known reference  $K_s$  and  $K_a$  fields. The spatial  
 6 locations and times of the synthetic observations are the same as those of the available historical  
 7 observations. The purpose of choosing this spatio-temporal configuration of synthetic  
 8 observations is to evaluate possible biases in inversion results introduced from the number of  
 9 observations and their spatio-temporal distribution. The objective of this experiment is to test the  
 10 ability of the ensemble smoother (ES) to estimate the true  $K_s$  and  $K_a$  fields using only hydraulic  
 11 head data.

12 To evaluate the performance of inverse parameter estimations in experiment A, the estimated  
 13 parameter fields are compared with referenced ones using two performance statistics: (1) the  
 14 mean absolute error  $L_1$ , and Pearson's correlation coefficient  $r$ , which are respectively calculated  
 15 as follows:

$$16 \quad L_1 = \frac{1}{n} \sum_{i=1}^{i=n} |\phi_{\text{true}}(i) - \phi_{\text{est}}(i)| \quad (12)$$

17

$$18 \quad r = \frac{\sum_{i=1}^n \left[ \begin{matrix} \text{true} & (i) & \text{---} \\ \text{true} & & \end{matrix} \right] \cdot \left[ \begin{matrix} \text{est} & (i) & \text{---} \\ \text{est} & & \end{matrix} \right]}{\sqrt{\sum_{i=1}^n \left[ \begin{matrix} \text{true} & (i) & \text{---} \\ \text{true} & & \end{matrix} \right]^2 \cdot \sum_{i=1}^n \left[ \begin{matrix} \text{est} & (i) & \text{---} \\ \text{est} & & \end{matrix} \right]^2}} \quad (13)$$

1 where  $\phi_{\text{true}}$  is the true (or reference) parameter vector and  $\phi_{\text{est}}$  is the estimate parameter vector. The  
2 Pearson's correlation coefficient  $r$  provides a measure for the linear correlation between the  
3 estimated and the reference parameter vectors.

4 Experiment B-1 assimilates historical groundwater level data collected from the field. Figure  
5 2 shows the locations of the 16 observations wells used in the assimilation. The temporal span of  
6 data varies from well to well within the period 2000-2006. The total number of water table  
7 measurements from the 16 wells is 4,944. Only half of the available water table elevation data  
8 are assimilated, in particular those collected during the first half of the 2000-2006 period.

9 In experiment B-2, results of experiment B-1 are validated by comparing observed states  
10 (water table elevation and streamflow) with states obtained numerically by simulating the  
11 updated  $K_s$  and  $K_a$  fields. Comparisons are carried out using Equations 12 and 13, with  $\phi$   
12 representing groundwater hydraulic head or streamflow. In this analysis, only water table  
13 elevation data collected in the second half of the 2000-2006 period are used. The comparison is  
14 also carried out with respect to streamflow rates observed at a stream gage located 6 km from the  
15 upstream end of the model domain (Figure 2). Note that, due to the relative proximity of the  
16 stream flow gauge to the upstream end, the contribution of aquifer losses or gains to the South  
17 Platte River is relatively small. Therefore, streamflow data was not used in updating parameter  
18 fields.

## 19 **5. Results and Discussion**

### 20 **5.1 Assimilation of Synthetic Hydraulic Head Data**

21 Figure 3 shows the reference streambed  $K_s$  field and the mean of the  $K_s$  fields as updated by  
22 assimilating synthetic hydraulic head data in experiment A. The two fields are very close in  
23 magnitude and patterns of spatial variation. Figure 4 shows a scatter plot of the updated mean



1 and reference  $K_s$  fields in relation to a 1-1 line. The correlation between the two fields is high  
2 with  $r = 0.98$  and  $L_1 = 0.0379$  (Equations 12 and 13), indicating that when system stresses are  
3 known, the streambed  $K_s$  field can be estimated effectively using hydraulic head data only. This  
4 notion can be understood by observing that Equation (1) relates aquifer water losses to the  
5 difference between head in the aquifer and stream stage. Since the uncertainty in stream stage is  
6 relatively smaller than the uncertainty in water table elevation (because stream-bed elevation can  
7 be directly measured and stage variability is typically small at the site), the uncertainty of flux  
8 depends largely on the adjacent aquifer head field, which is controlled principally by the spatial  
9 distribution of  $K_a$ . That is to say, the exchange flux rates between the aquifer and the river are  
10 governed mainly by the aquifer hydraulic head data. However, this situation likely is not realized  
11 in reality, since other sources of uncertainty, for example of conceptual and structural nature, can  
12 contribute to the prediction errors.

13 Figure 5 shows the reference  $K_a$  field and the mean of the updated  $K_a$  ensemble. A visual  
14 comparison shows that the two fields are very similar in values and spatial distribution. Figure 6  
15 shows a scatter plot of the estimated and reference field with a 1-1 line. The correlation between  
16 the two fields is high with  $r = 0.98$  and  $L_1 = 0.153$  indicating a high performance of the ES in  
17 estimating the reference  $K_a$  field. It is important to recall that the hydraulic stresses used to  
18 generate the synthetic measurements are the same as those used to generate the realizations in the  
19 forecast state-parameter matrix, i.e. the discrepancy between the prior head ensemble and the  
20 synthetic measurements comes in this case only from the unknown system parameters.

## 21 **5.2 Assimilation of Historical Groundwater Level Measurements (Experiment B-1)**

22 Figure 7a shows the ensemble mean of the updated  $K_a$  fields from assimilating half of the  
23 available head data as in experiment B-1. The spatial variability of the estimated  $\ln(K_a)$  ranges

1 between 1 and 10 (ln(m/day)), which are reasonable values for an alluvial aquifer. To evaluate  
2 the efficiency of this estimate, the posterior standard deviation is plotted in Figure 7b. The  
3 posterior standard deviation ranges between 0.5 and 1.2 (ln(m/day)). These values correspond to  
4 coefficient of variations that range between 5.5% and 60%. A close analysis of the spatial  
5 distribution of the standard deviation of the updated  $K_a$  field reveals that regions of low standard  
6 deviation coincide with the locations of observation wells (Figure 2).

7 Figure 8 shows the prior and posterior ensembles of  $K_s$ . The posterior ensemble mean of  
8 streambed  $K_s$  is also shown. One can see that the prior ensemble mean is constant with a value of  
9  $z = 0$  (equivalent to 1 m/day), while the posterior mean is spatially variable. The posterior mean  
10 of  $\ln(K_s)$  ranges between the values -0.5 and 0.5 (equivalent to 0.61-1.65 m/day), which are  
11 within the range of published conductivity values (0.01 to 85 m/day) published by Calver (2001).

12 One important observation is that the range of variability of the estimated  $K_s$  values is  
13 relatively small when compared to published values (Calver, 2001), which could have a wider  
14 range of 1-100 m/day at the same site. This can be explained by recalling that the cell size in the  
15 model is about 304.8 m and thus the resulting estimates are the effective stream conductivity on  
16 a support scale of about 304.8 m and stream width of 14 m. The sensitivity of calibration results  
17 to uncertainty in the prior standard deviation was investigated by repeating the calibration using  
18 different prior standard deviations ( $\sigma_z = 0.1, 1.0, 2.0$ ). Results indicate that applying different  
19 standard deviations does not have a large effect on the posterior ensemble. These results are not  
20 shown here.

21 It is worth recalling that these results rely on the assumption that  $K_s$  is constant with time. A  
22 number of studies have shown that this is not always the case, as flood events and streambed  
23 erosion might introduce changes in magnitudes and spatial distribution of  $K_s$  (Springer et al.,

1 1999; Cardenas et al. 2004). In this respect, the  $K_s$  estimates shown in Figure 8 represent  
2 “effective”  $K_s$  values over the period 2000-2006.

3 The statistical properties of the total stream-aquifer flux exchange along the simulated reach (  
4 about 35.6 km long) are summarized in Figure 9, which shows the prior and posterior  
5 distribution of the total flux. Flow values are calculated for each of the 500 realizations by  
6 temporally averaging (over 84 months stress periods) the total stream exchange flow between the  
7 35.6 km river and the aquifer. Both prior and posterior average total flows are negative which  
8 indicate the case of an aquifer discharging groundwater to the stream (gaining stream), which is  
9 expected according to common understanding of the South Platte River interaction with the  
10 alluvial aquifer. The prior mean total flow is about  $-0.83 \times 10^5 \text{ m}^3/\text{day}$  and the posterior mean is -  
11  $1.34 \times 10^6 \text{ m}^3/\text{day}$ . These values are equivalent to  $2.33 \text{ m}^3/\text{day}$  per unit length of the stream for the  
12 prior flow and  $4.01 \text{ m}^3/\text{day}$  per unit length of the stream for the posterior flow.

13 To gain more insight on the spatial distribution of flux exchange, Figure 10 shows the  
14 posterior ensemble and the posterior ensemble mean of the stream-aquifer flux along the length  
15 of the South Platte River. While the flow in general is from the aquifer to the stream, some  
16 segments experience flow from the stream to the aquifer. The spatial variability along the river  
17 reach has a maximum average monthly aquifer gain of about  $0.98 \times 10^4 \text{ m}^3/\text{day}/\text{per stream reach}$   
18 ( $2.3 \text{ m}^3/\text{day}/\text{m}^2$ ) and a maximum average monthly aquifer loss of  $1.2 \times 10^4 \text{ m}^3/\text{day}/\text{per stream}$   
19 reach ( $2.8 \text{ m}^3/\text{day}/\text{m}^2$ ).

### 20 **5.3 Validation of Assimilation Results (Experiment B-2)**

21 As indicated in Section 4.2, available observation data consist of groundwater hydraulic head  
22 time series at 16 observation wells and streamflow at a stream gauge located 6 km from the  
23 upstream end of the model domain. In experiment B-2, half of the hydraulic head data and the

1 streamflow data are used to validate the updated  $K_a$  and  $K_s$  fields. To do so, the mean of the  
2 update ensembles of  $K_s$  and  $K_a$  are simulated to predict the hydraulic head at the locations and  
3 times of observed heads and the streamflow at the site of the stream gage. Figure 11 shows the  
4 comparison between the simulated and observed heads. The correlation between observed and  
5 simulated head is  $r = 0.99$  and  $L_1 = 1.50$ , indicating good performance of the inversion process.

6 In a similar manner, the simulated and observed stream flow at the stream gage is shown in  
7 Figure 12. Figure 12a compares monthly simulated and observed stream flow rates. A general  
8 agreement between the two time-series is observed, with  $r = 0.685$  (Figure 12b) and  $L_1 = 4.2 \times$   
9  $10^5$ . Streamflow estimation could be significantly improved if the stream gage was located  
10 further downstream within the study area. Since the stream gage is located only 6 km from the  
11 upstream end, the updated  $K_s$  and  $K_a$  fields do not have a strong influence on surface water –  
12 groundwater exchange rates.

13 To explore the impact of not calibrating the spatial variability of streambed  $K_s$ , the ES is used  
14 to recalibrate the aquifer conductivity field ( $K_a$ ) assuming spatially constant streambed  $K_s$  equal  
15 to the posterior average streambed conductivity ( $\overline{K_s}$ ) estimated in experiment B1. In this  
16 calibration experiment, the forecast is achieved by simulating an ensemble of spatially variable  
17  $K_a$  realizations, whereas the streambed  $K_s$  is assumed to be spatially constant and deterministic ( $\ln(\overline{K_s}) \approx -0.008$ ). The recalibrated conductivity field  $K_a'$  is compared to the conductivity field  
18  $K_a$  estimated in experiment B1 (Figure 7a). The spatial variability of the difference between the  
19 two fields ( $\ln(K_a) - \ln(K_a')$ ) is shown in Figure 13a, whereas Figure 13b compares between the  
20 observed hydraulic heads and the simulated heads using  $K_a'$  field. Erroneously disregarding the  
21 calibration of the  $K_s$  field produces a suboptimal estimation of  $K_a$  field that compensates for this  
22

1 error and still minimizes the calibration residual error. As a result, the correlation between the  
2 simulated heads and observed heads decreased from  $r = 0.99$  for the case wherein the spatial  
3 variability of  $K_s$  is calibrated to  $r = 0.97$  where  $K_s$  calibration is disregarded. It is worth noting  
4 that disregarding the calibration of  $K_s$  can be seen as adopting a different parametrization scheme  
5 for the unknown parameters that still minimizes the calibration residual error (The equifinality of  
6 inverse problem (Beven, 2001)).

## 7 **6. Summary and Conclusions**

8 This study implements data assimilation of groundwater level measurements using the  
9 Ensemble Smoother to estimate the spatial heterogeneity of both spatially-varying streambed and  
10 hydraulic conductivity along a 35.6 km reach of the South Platte River in northeastern Colorado.  
11 The two fields were parameterized using cellblocks with sizes of 304.8 m square. Two numerical  
12 experiments were conducted to explore the performance of data assimilation: (1) assimilating  
13 synthetic data and (2) assimilating historical groundwater levels from 16 observation wells. In  
14 the synthetic experiment, assimilated groundwater head measurements were obtained from  
15 known streambed and aquifer hydraulic conductivity fields, with measurements having the same  
16 spatial locations and temporal frequencies as the historical data. In assimilating the historical  
17 head data, half of the available groundwater level measurements are used in the assimilation,  
18 while the other half and streamflow measurements are used to evaluate the accuracy of the  
19 estimated fields.

20 Results show that the Ensemble Smoother reproduces the synthetic streambed and aquifer  
21 hydraulic conductivity fields with very good agreement to the reference fields. In assimilation of  
22 historical data, results show that simulated groundwater levels and stream flow rates using the  
23 estimated streambed and aquifer hydraulic conductivity fields are in reasonably good agreement

1 with observed data. The posterior ensemble means of estimated  $K_s$  and  $K_a$  fields were used to  
2 estimate the spatial variability of stream-aquifer flux exchange, which show high degree of  
3 spatial variability. While applying data assimilation to estimate the parameters of groundwater  
4 systems is still limited in practice, this work shows that the approach can provide a reliable and  
5 computationally affordable inversion tool and the methods described in this paper can be applied  
6 to other stream-aquifer systems.

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1 **Table 1. Groundwater Model Setting and Properties of the Geostatistical Model for Hydraulic Parameters.**

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<b>Properties of Finite Difference Grid</b>	
Horizontal Domain Dimensions [x,y]	[30 km,30 km]
Saturated Thickness [z]	15 m - 88 m
Rows, Columns, and Layers	[100,100,1]
Number of Active Cells	3461
<b>Simulation Times</b>	
Simulated Period	2000 to 2006
Steady State stress period	One month (1/2000)
Transient State Stress Period	84 Months (7 years)
Time step	1 day
<b>Boundary Conditions</b>	
Upper Boundary Condition (Fig. 2)	Variable flux for each stress period.
lower Boundary Condition (Fig. 2)	Variable flux for each stress period.
Northern Side Boundary Condition (Fig. 2)	Variable flux for each stress period.
Southern Side Boundary Condition (Fig. 2)	Variable flux for each stress period.
Initial Boundary Condition (Fig. 2)	Interpolated groundwater table measurements for January 2000
<b>Geostatistical Properties of Aquifer Parameters</b>	
Log( $K_a$ ) - 2D isotropic field [ $\mu_y, \sigma_y^2, \lambda_y$ ]	[5,1.5,4000m]
Log( $K_s$ ) - 1D field [ $\mu_z, \sigma_z^2, \lambda_z$ ]	[0.1,0.1,2000m]
Sy	Constant Value (0.20)

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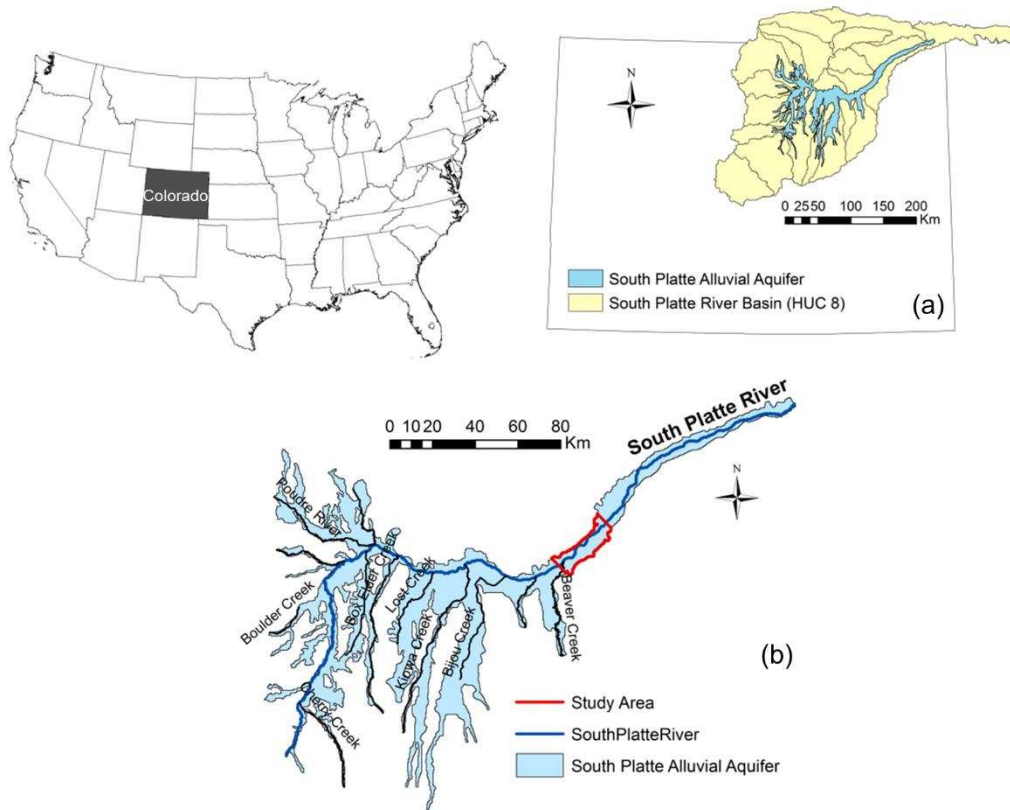
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5 **Table 2. Data Assimilation Experiments and Cross-validation of Estimated Parameters**

Experiment	Description	Data Used	Ensemble Size
A	Synthetic Data Assimilation	Synthetic Head Data	500 realizations
B	B-1 Assimilation of Actual Field Data	Actual Head Data at 16 Observation wells at 2472 different times.	500 realizations
	B-2 Cross-Validation of Estimated Parameters.	<ul style="list-style-type: none"> <li>Actual Head Data at 16 Observation wells at 2472 different times.</li> <li>Month Stream flow Data for the period 2000 to 2006.</li> </ul>	N/A

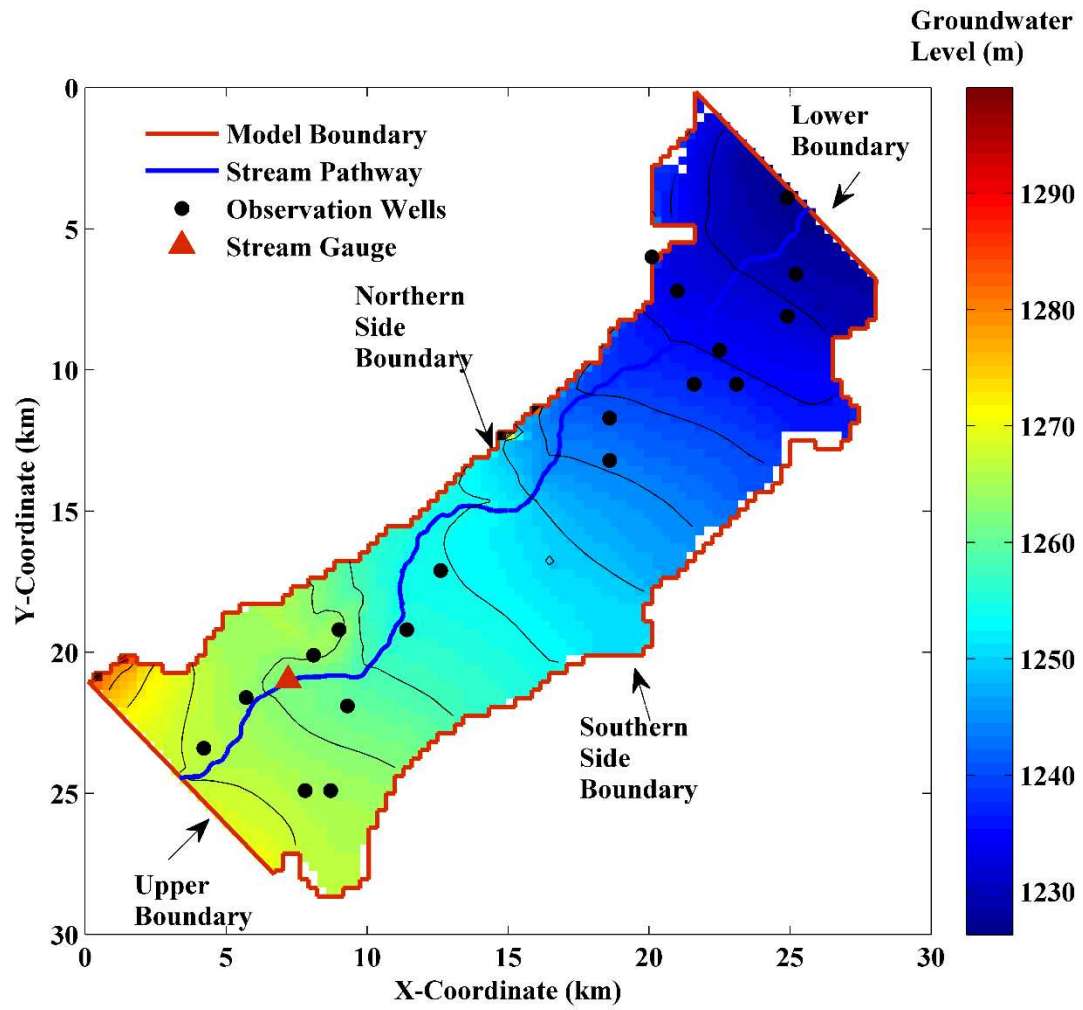
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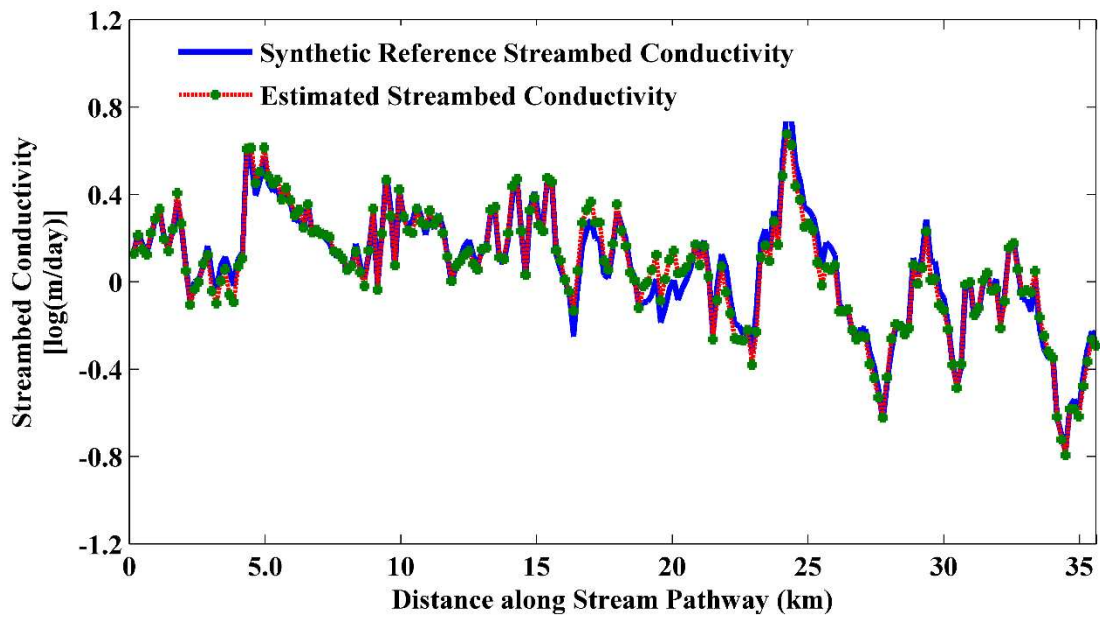
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2 **Figure 1.** Regional stream-aquifer system of the South Platte River Basin in northeastern Colorado is shown in  
 3 panel (a). The alluvial aquifer and local study area are shown in panel (b).



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2 **Figure 2.** Model Boundaries, stream reach, observation wells, and stream gauge within the study area. The  
 3 simulated groundwater hydraulic head is shown for December 2006.

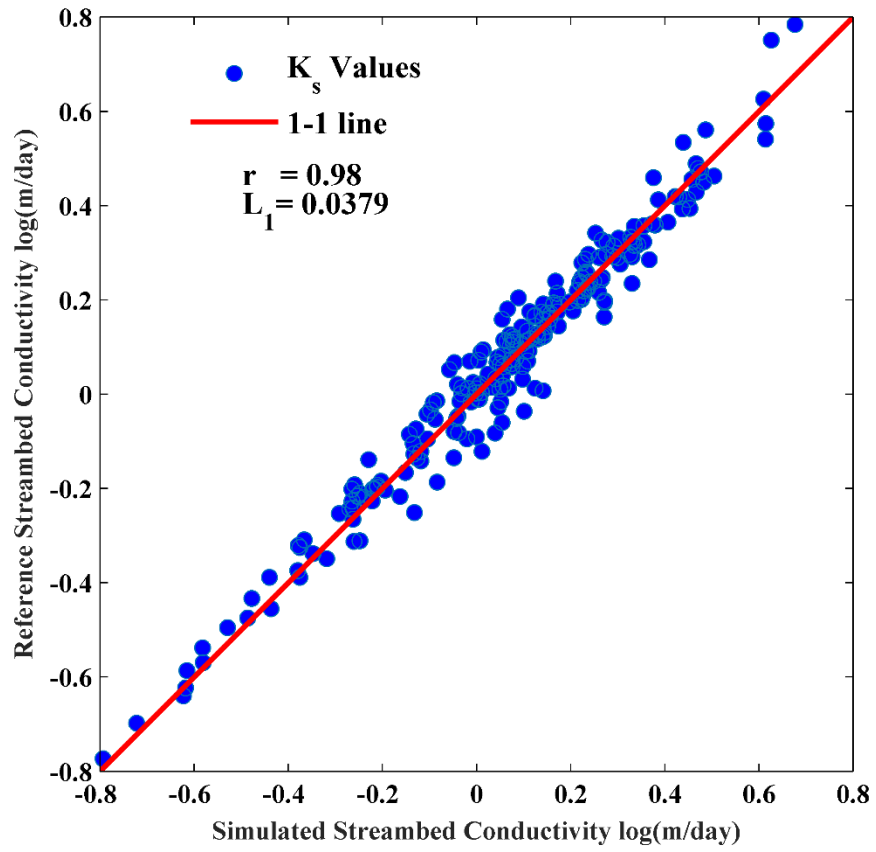


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2 **Figure 3.** Comparison between the reference streambed conductivity field and the mean of the updated ensemble of  
 3 streambed conductivity field using synthetic hydraulic head data only.

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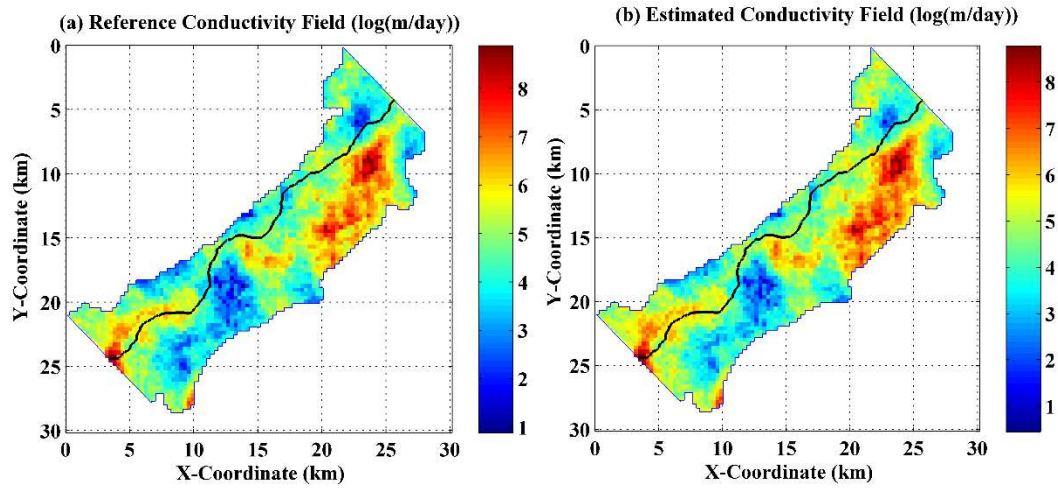


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3 **Figure 4.** Scatter plot comparing the reference streambed  $K_s$  field values and the mean of the updated ensemble of  
4 streambed  $K_s$  values estimated by assimilating synthetic hydraulic head data.

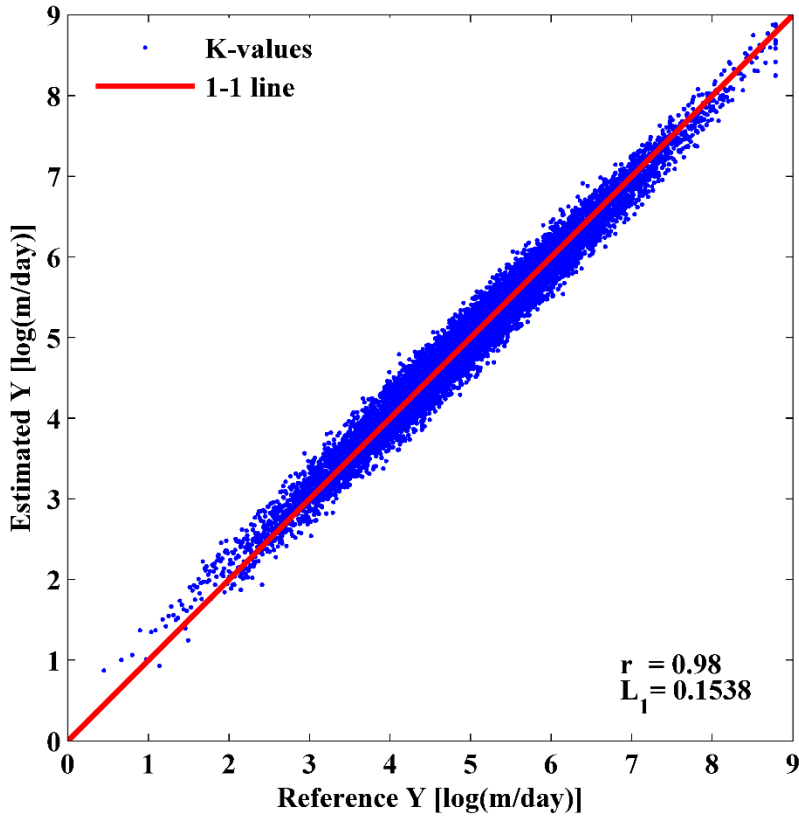
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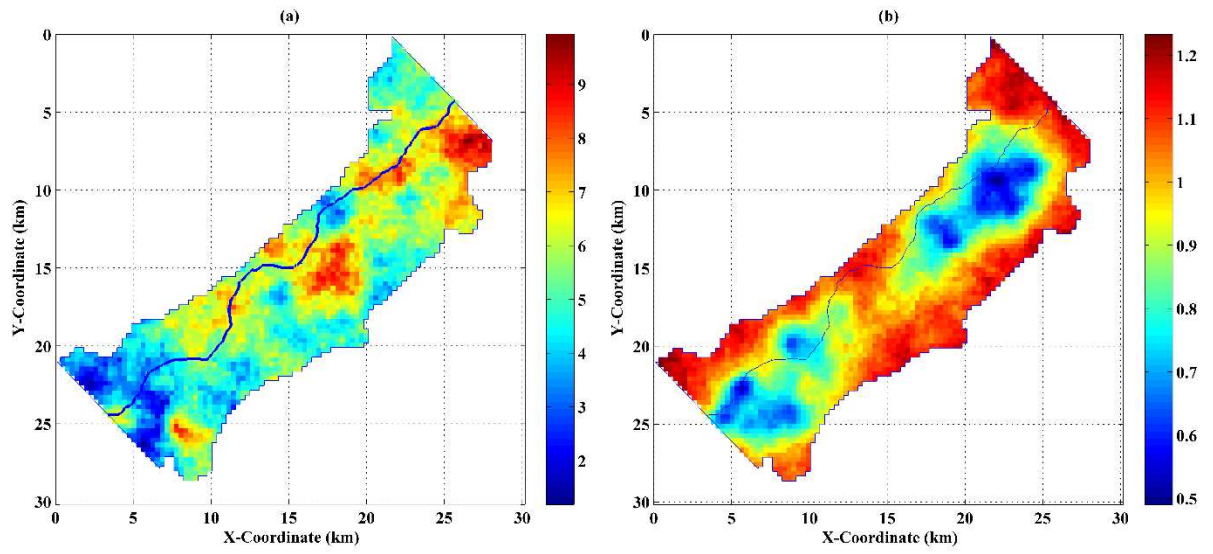
2 **Figure 5.** Comparison between (a) the reference aquifer K field and (b) the mean of the updated ensemble of aquifer  
 3 K fields, by assimilating synthetic hydraulic head data.



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**Figure 6.** Scatter plot comparing the reference aquifer K field values and the mean of updated ensemble of aquifer K values, by assimilating synthetic hydraulic head data.

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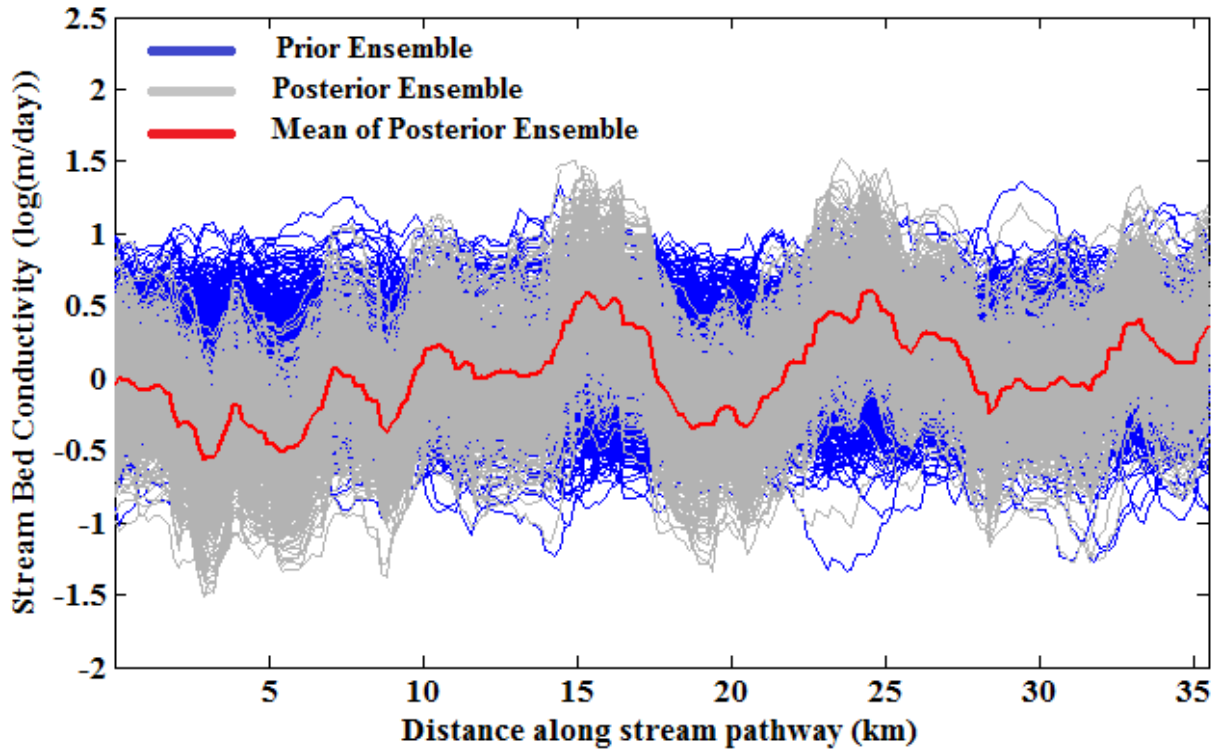


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3 **Figure 7.** Panel (a) shows the ensemble mean of the hydraulic conductivity ( $K_f$ ) posterior ensemble ( $\log(\text{m/day})$ ).  
4 Panel (b) shows the posterior standard deviation of  $K_f$  field at each local cell.

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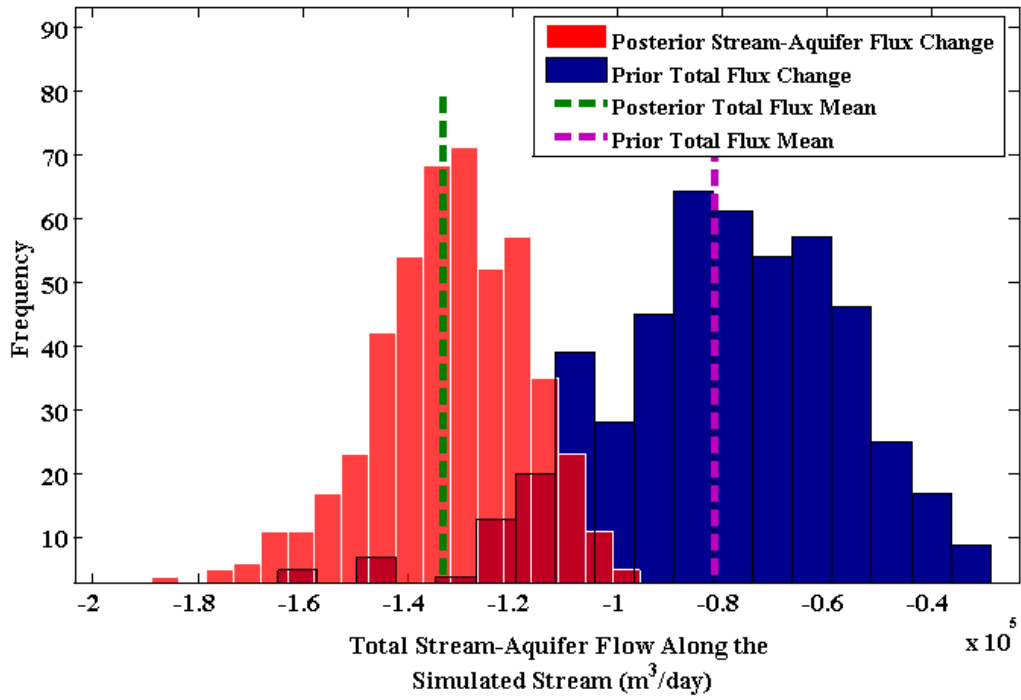


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3 **Figure 8.** Prior and posterior ensembles of streambed K. The mean of the posterior ensemble is highlighted with red  
4 color.

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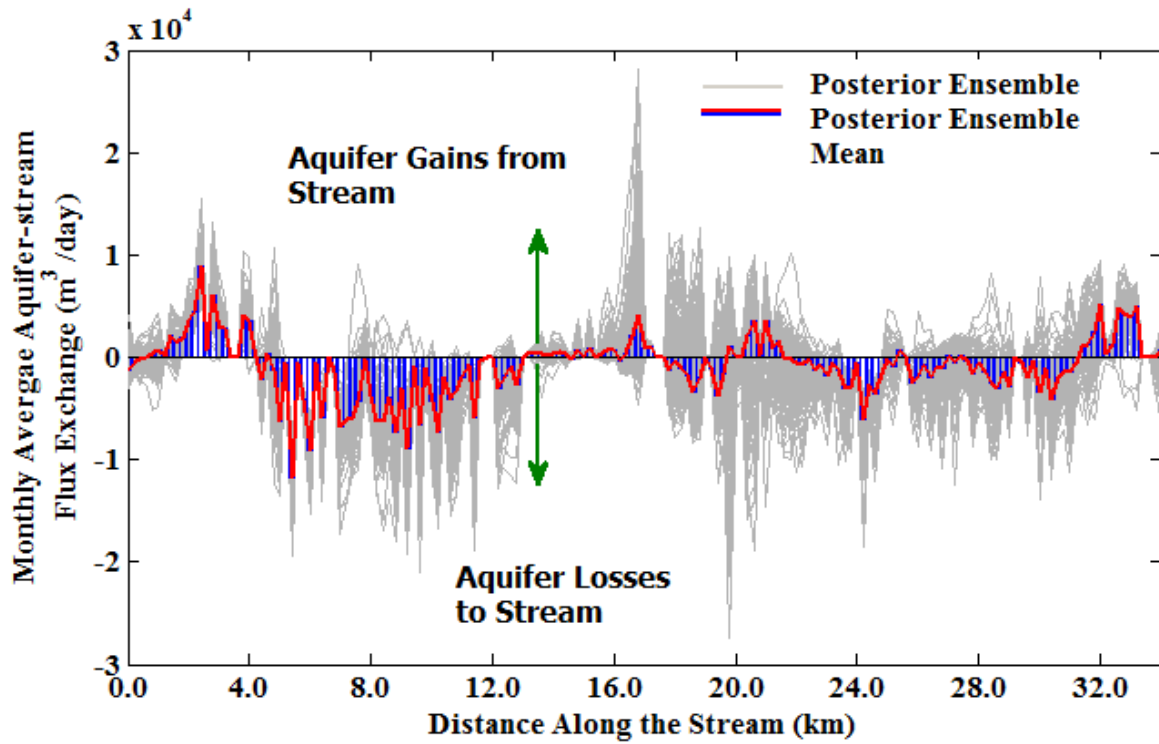
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3 **Figure 9.** Shows the prior and posterior ensembles of stream-aquifer total flux exchange along the simulated stream.  
4 The means of the prior and posterior ensembles are highlighted. Negative flow rates indicate groundwater leaving  
5 the aquifer to the stream, while the positive flow rates indicate that the aquifer receives water from the stream.

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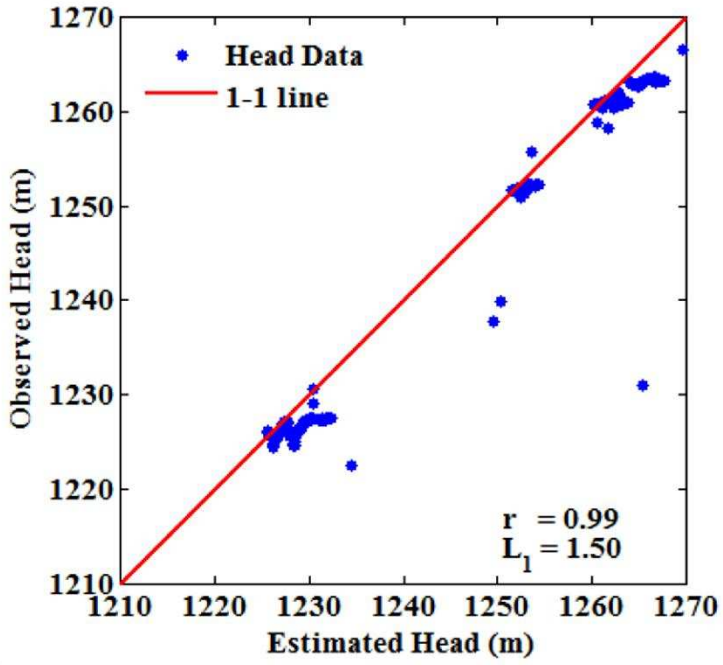
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**Figure 10.** Shows the posterior ensembles of stream-aquifer flux exchange. The mean of the posterior ensemble is highlighted with red color. Negative flow rates indicate groundwater leaving the aquifer to the stream, while the positive flow rates indicate that the aquifer receives water from the stream.



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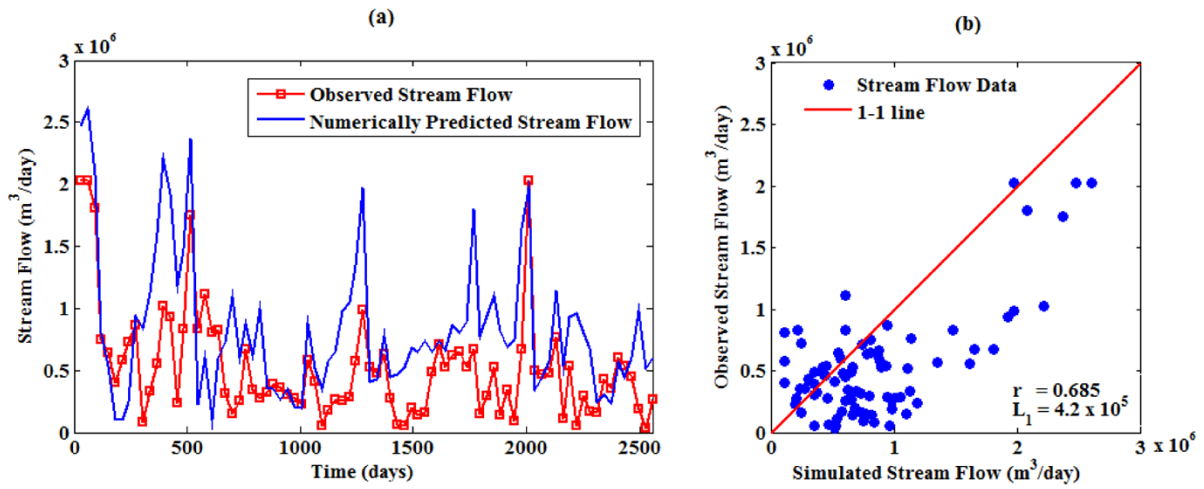
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**Figure 11.** Scatter Plot that compares between the observed hydraulic head measurements and the simulated hydraulic head obtained by simulating the  $K_f$  and  $K_s$ .

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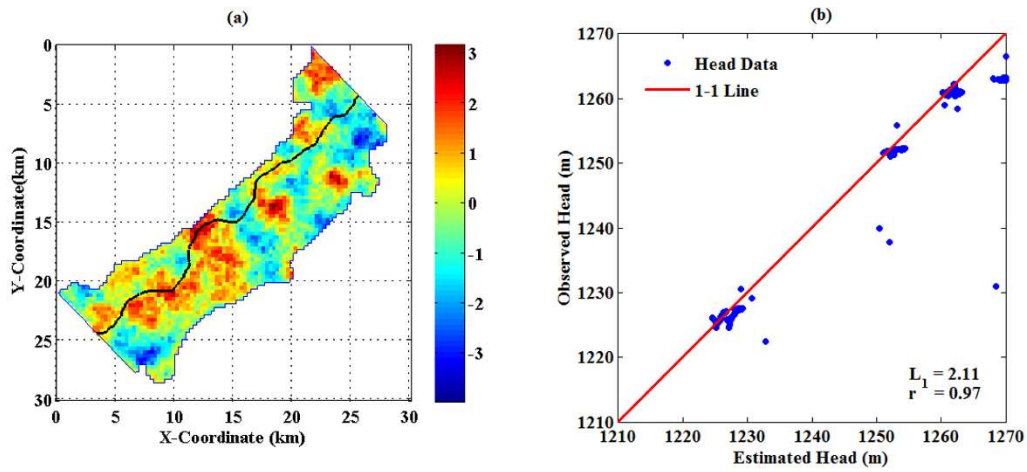
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3 **Figure 12.** Panel (a) shows the simulated and observed stream flow gauges, (b) Scatter Plot that compares between  
4 the observed stream flow and the simulated stream flow obtained by simulating the estimated  $K_f$  and  $K_s$ .

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3 **Figure 13.** Panel (a) shows the spatial variability of difference between the calibrated aquifer  $K_a$  field when  
4 streambed conductivity is calibrated and aquifer  $K_a'$  when streambed conductivity calibration is disregarded, (b)  
5 Scatter Plot that compares between the observed hydraulic head measurements and the simulated hydraulic head  
6 obtained by simulating  $K_a'$ .