



UNIVERSITY OF LEEDS

This is a repository copy of *The variation of productivity and its allocation along a tropical elevation gradient: a whole carbon budget perspective*.

White Rose Research Online URL for this paper:
<http://eprints.whiterose.ac.uk/103760/>

Version: Supplemental Material

Article:

Malhi, Y, Girardin, C, Goldsmith, G et al. (15 more authors) (2017) The variation of productivity and its allocation along a tropical elevation gradient: a whole carbon budget perspective. *New Phytologist*, 214 (3). pp. 1019-1032. ISSN 0028-646X

<https://doi.org/10.1111/nph.14189>

© 2016, Wiley. This is the peer reviewed version of the following article: "Malhi, Y., Girardin, C. A. J., Goldsmith, G. R., Doughty, C. E., Salinas, N., Metcalfe, D. B., Huaraca Huasco, W., Silva-Espejo, J. E., del Aguilla-Pasquell, J., Farfán Amézquita, F., Aragão, L. E. O. C., Guerrieri, R., Ishida, F. Y., Bahar, N. H. A., Farfan-Rios, W., Phillips, O. L., Meir, P. and Silman, M. (2016), The variation of productivity and its allocation along a tropical elevation gradient: a whole carbon budget perspective. *New Phytol.* doi:10.1111/nph.14189" which has been published in final form at <http://doi.org/10.1111/nph.14189>. This article may be used for non-commercial purposes in accordance.

Reuse

Unless indicated otherwise, fulltext items are protected by copyright with all rights reserved. The copyright exception in section 29 of the Copyright, Designs and Patents Act 1988 allows the making of a single copy solely for the purpose of non-commercial research or private study within the limits of fair dealing. The publisher or other rights-holder may allow further reproduction and re-use of this version - refer to the White Rose Research Online record for this item. Where records identify the publisher as the copyright holder, users can verify any specific terms of use on the publisher's website.

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

***New Phytologist* Supporting Information**

Article title: The variation of productivity and its allocation along a tropical elevation gradient: a whole carbon budget perspective

Authors: Yadvinder Malhi, Cécile A.J. Girardin, Gregory R. Goldsmith, Christopher E. Doughty, Norma Salinas, Daniel B. Metcalfe, Walter Huaraca Huasco, Javier E. Silva-Espejo, Jhon del Aguilla-Pasquell, Filio Farfán Amézquita, Luiz E. O. C. Aragão, Rossella Guerrieri, Françoise Yoko Ishida, Nur Bahar, William Farfan-Rios, Oliver L. Phillips, Patrick Meir, Miles Silman

Article acceptance date: 12 July 2016

The following Supporting Information is available for this article:

Fig. S1 Relationship between foliar nutrient concentrations (nitrogen and phosphorus) and elevation.

Fig. S2 Relationship between primary productivity and temperature.

Fig. S3 Relationship between primary productivity and elevation without influential plot.

Appendix S1 Detailed methods.

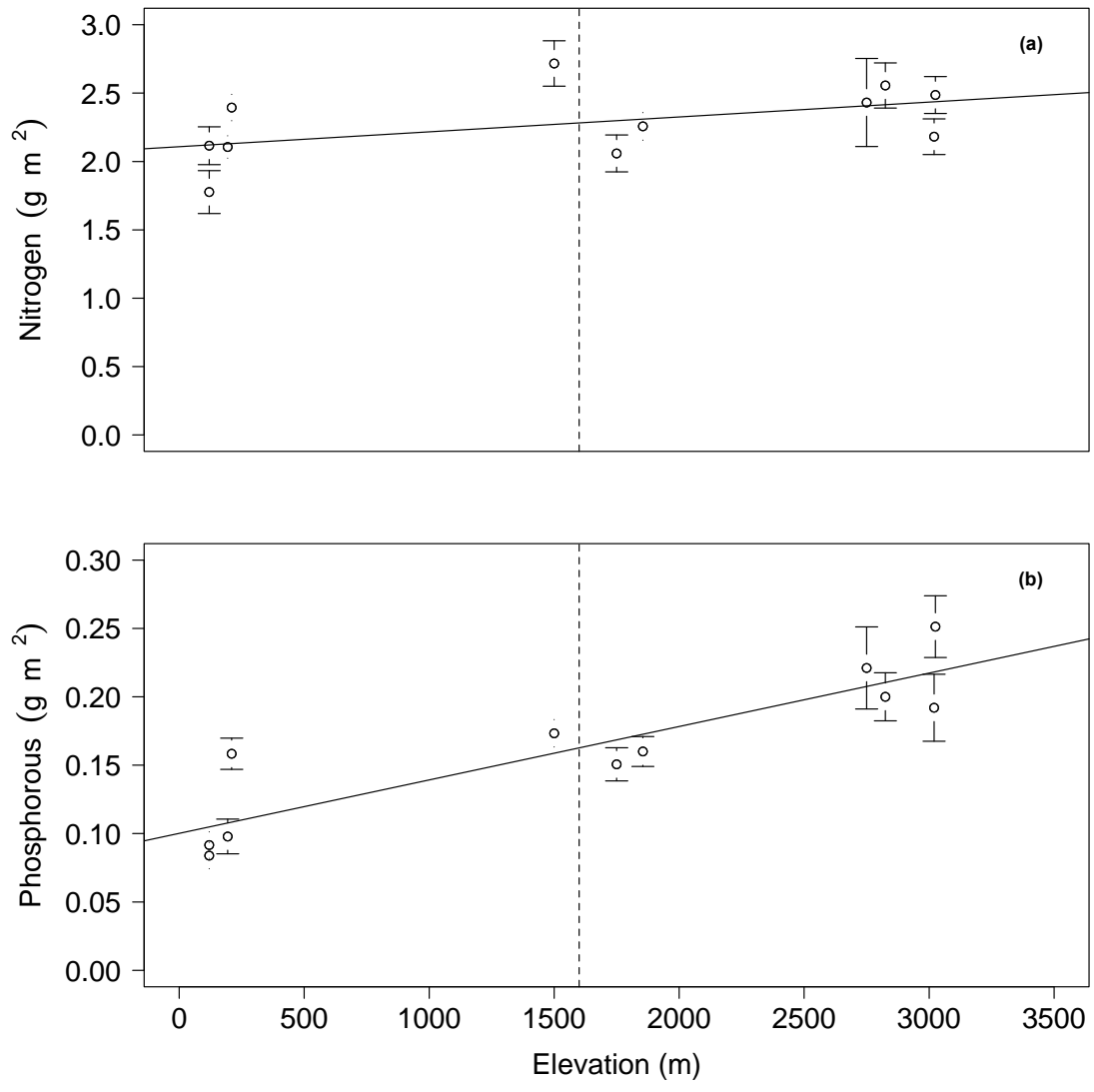


Fig. S1 Variation in foliar (a) nitrogen and (b) phosphorus concentration along a 3300 m tropical montane elevation transect in Peru. The vertical dashed line indicates the submontane-cloud forest ecotone. Error bars indicate standard errors.

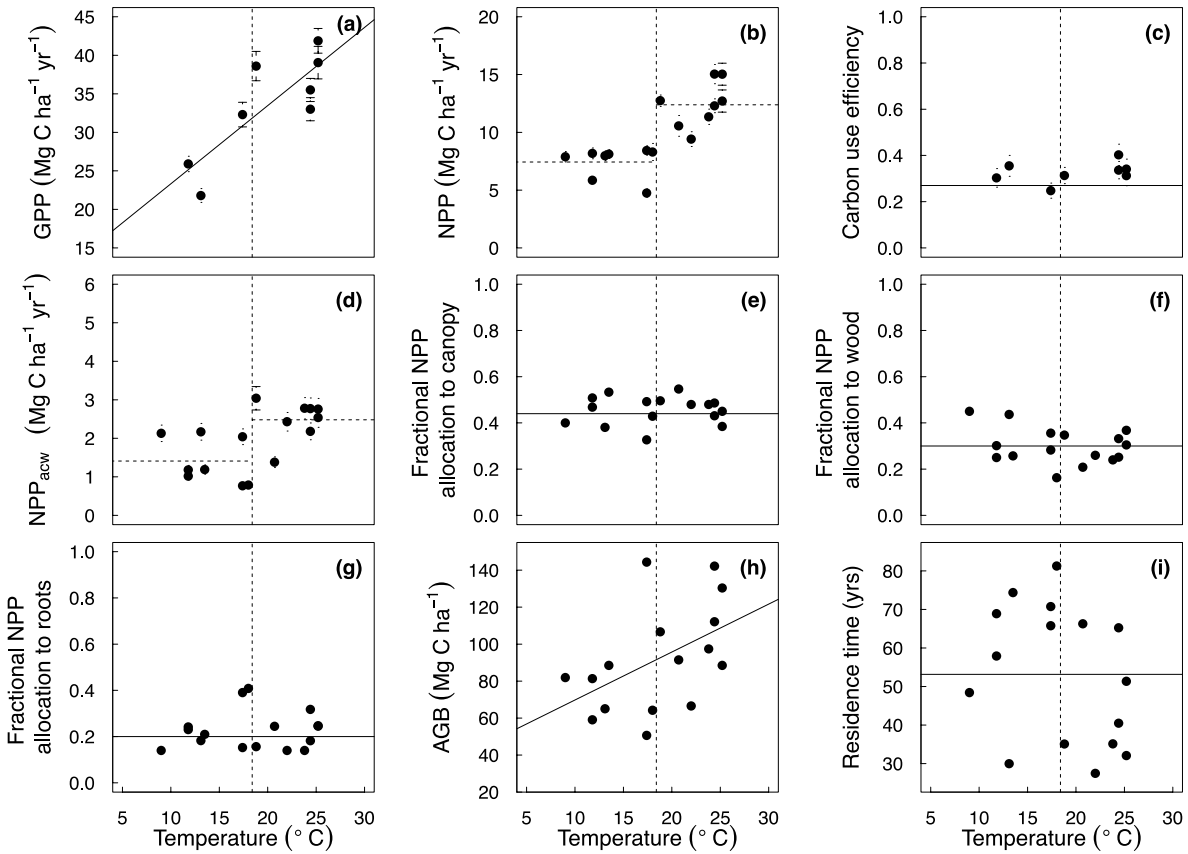


Fig. S2 As Figure 3 in the main text, but with all variables plotted against temperature instead of elevation. Variation in carbon cycle characteristics along the 3300 m tropical montane elevation transect, including (a) gross primary productivity (*GPP*), (b) net primary productivity (*NPP*), (c) carbon use efficiency, the fraction NPP/GPP , (d) aboveground coarse woody *NPP* (NPP_{acw}), (e) fractional *NPP* allocation to canopy components, (f) fractional *NPP* allocation to woody components, (g) fractional *NPP* allocation to roots, (h) above-ground live biomass (*AGB*) and (i) woody residence time. The best model fit (according to AIC) is shown when significant, either a single horizontal line or slope, or two lines split at 1600 m (cloud base). The vertical dashed line indicates the temperature at the submontane forest-cloud forest ecotone. Error bars indicate standard errors.

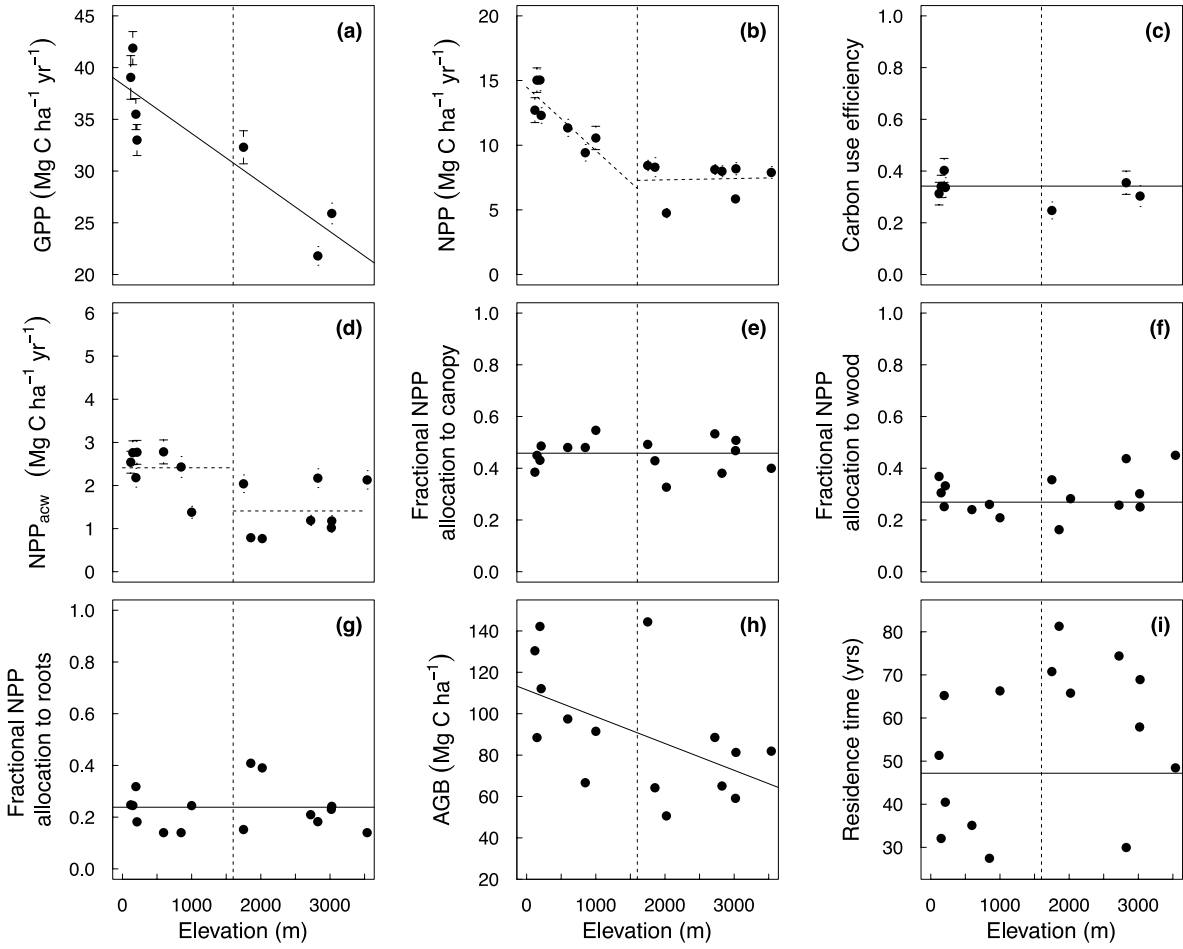


Fig. S3 As Figure 3 in the main text, but with the sometimes influential plot SPD-02 removed. The main difference compared to Figure 3 in the main text is in the plot of NPP.

Appendix S1

Methods S1

This section gives a more detailed explanation of the measurements and data analysis procedures.

Our approach in the 1 ha GEM plots is to quantify the major components of the autotrophic carbon cycle at multiple sites. Herein, “autotrophic” implies focusing on the plant processes of photosynthesis, productivity, autotrophic respiration and allocation, rather than heterotrophic processes such as decay and soil organic matter respiration. We adopt the field protocol of the GEM network (<http://gem.tropicalforests.ox.ac.uk>). These methods are described in detail in an online manual on the GEM website and are summarised here only briefly. Methods and descriptions of individual pairs of plots along our gradients have recently been published, but there has been no systematic, cross-site comparison of annual averages along the entirety of our lowland dataset.

Here we summarise the protocols and the primary differences among sites based on the site-specific papers (Girardin et al. 2014, Araujo-Murakami et al. 2014, da Costa et al. 2014, Doughty et al. 2014, Metcalfe et al. 2010, Rocha et al. 2014, Huaraca Huasco et al. 2014, del Aguila-Pasquel et al., 2014; Malhi et al., 2014).

Site characteristics and disturbance history

All sites included in this analysis show little evidence of anthropogenic disturbance of the forest community structure, hosting mixed-age tree communities with little net increment in biomass.

SPD-02 was lightly logged at some point several decades ago, as evidenced by the presence of a large tree stump within the plot. The montane sites are likely impacted by occasional landslips and more rare landslides (Clark et al 2014)

Above ground net primary productivity

Measured above-ground net primary productivity (NPP_{AG}) components included:

Above ground coarse wood net primary productivity ($NPP_{ACW \geq 10}$): All trees ≥ 10 cm DBH were censused to determine growth rate of existing, surviving trees and rate of recruitment of new trees. Stem biomass was calculated using an allometric equation for tropical moist forests, employing diameter, height, and wood density data (Chave et al., 2005). To convert biomass values into carbon, we assumed that dry stem biomass is 47.3% carbon (Martin & Thomas, 2011). Where tree height data were not available, height was estimated from an allometric equation appropriate for each region (Feldpausch et al., 2011).

Above ground coarse wood net primary productivity ($NPP_{ACW \leq 10}$): All trees < 10 cm DBH (details on minimum size provided in Table S1) were censused in subplots within each site to estimate the contribution of smaller stems to NPP_{ACW} .

Figure S3. As Figure 3, but with the sometimes influential plot SPD-03 removed. The main differences is in the plot of NPP.

Branch turnover net primary productivity ($NPP_{\text{branch turnover}}$): The turnover of branches, where trees shed branches and grow new ones, can generate a significant component of woody NPP

that is not accounted for by the static tree allometries used above. Branches > 2 cm diameter (excluding those fallen from dead trees) were surveyed along fixed transects; small branches were cut to include only the transect-crossing component, removed and weighed. Larger branches had their dimensions taken (diameter at 2 or more points) and all were assigned a wood density value according to their decomposition class. Details of decomposition status and surface area formulas are available in the RAINFOR-GEM network field manual (www.gem.tropicalforests.ox.ac.uk).

Litterfall net primary productivity ($NPP_{\text{litterfall}}$): Annual values of $NPP_{\text{litterfall}}$ were estimated from dead organic material <2 cm diameter collected in 0.25 m² (50 cm × 50 cm) litter traps placed at 1 m above the ground at the centre of each of 25 subplots established within each plot. Litter is separated into its components, oven dried at 80 °C to constant mass and weighed. Leaf carbon content was estimated to be 49.2% (Tambopata and montane plots) and 53.2% (Allpahuayo) carbon based on direct measurements on sampled leaves.

Leaf Area Index (LAI): Canopy images recorded with a digital camera and hemispherical lens near the centre of each of the 25 subplots, at a standard height of 1 m, and during overcast conditions LAI estimated using “true LAI” output from the CANEYE program (INRA, Avignon, France) which accounts for clumping of foliage. Leaves were separated into sunlit and shaded fractions using the following equation: $F_{\text{sunlit}} = (1 - \exp(-K \cdot \text{LAI})) / K$ where K is the light extinction coefficient, and F_{sunlit} is the sunlit leaf fraction (Doughty & Goulden, 2008). The model assumptions are randomly distributed leaves, and $K = 0.5 / \cos(Z)$, where Z is the solar zenith angle, which was set to 30°.

Loss to leaf herbivory ($NPP_{\text{herbivory}}$): Loss to leaf herbivory is the fraction of NPP_{canopy} lost to herbivory prior to litterfall. At Tambopata, Wayqecha, Esperanza and San Pedro, leaves collected in the 25 litterfall traps in each plot were scanned prior to being dried and the leaf area calculated using image analysis software (ImageJ, NIH, USA). The fractional herbivory (H) for each leaf was calculated as: $H = (A_{\text{nh}} - A_h) / A_{\text{nh}}$, where A_h is the area of each individual leaf including the damage incurred by herbivory and A_{nh} is the leaf area prior to herbivory. The mean values of H were calculated across all leaves collected both per litterfall trap and per plot. Data on leaf herbivory are further explored by Metcalfe et al (2014). The mean herbivory fraction observed at Wayqecha was assigned to all montane plots, that observed at SPD-02 assigned to the sub-montane plots (Pantiacolla), and that observed at Tambopata also assigned to Allpahuayo.

Belowground net primary productivity

Below-ground net primary productivity (NPP_{BG}) components consisted of fine and coarse roots
NPP:

Coarse root net primary productivity ($NPP_{\text{coarse root}}$): Due to potential damage to the trees, this is not measured directly. Instead, $NPP_{\text{coarse root}}$ is estimated as 0.21 ± 0.03 of above-ground woody productivity, based on the published values of coarse root biomass to above ground biomass (Jackson et al., 1996; Cairns et al., 1997). Details and the range of the root:shoot ratio are available in the RAINFOR-GEM network field manual (www.gem.tropicalforests.ox.ac.uk).

Fine root net primary productivity ($NPP_{\text{fine root}}$): Sixteen ingrowth cores (mesh cages 12 cm diameter, installed to 30 cm depth) were installed in each plot. Cores were extracted and roots were manually removed from the soil samples in four 10 min time steps and the pattern of cumulative extraction over time was used to predict root extraction beyond 40 min. Root-free soil was then re-inserted into the ingrowth core. Collected roots were thoroughly rinsed, oven dried at 80°C to constant mass, and weighed. An additional correction factor was applied for fine roots not collected within 40 min. A further correction was applied for unmeasured roots below 30 cm depth according to fine root biomass profiles extrapolated to the observed soil depth (or to 1 m in the case of the deep soiled plots in the lowlands).

Autotrophic and heterotrophic respiration

Total Soil CO₂ efflux (R_{soil}): Total soil CO₂ efflux was measured using a closed dynamic chamber method, employed at the centre of each of the 25 sub-plots every 20 min, with an infra-red gas analyser and soil respiration chamber (EGM-4 IRGA and SRC-1 chamber, PP Systems, Hitchin, UK) sealed to a permanent collar in the soil. Soil surface temperature (T260 probe, Testo Ltd., Hampshire, UK) and moisture (Hydrosense probe, Campbell Scientific Ltd., Loughborough, UK) were recorded at each point after efflux measurement.

Soil CO₂ efflux partitioned into autotrophic ($R_{\text{rhizosphere}}$) and heterotrophic (R_{soilhet}) components:

The autotrophic and heterotrophic components of soil respiration were quantified using a partitioning experiment similar to that described in Metcalfe et al. (2007). The partitioning experiment allows estimation of the relative contributions of surface organic litter, rhizosphere

and soil organic matter to total soil CO₂ efflux. At four points at each corner of the plot, plastic tubes of 12 cm diameter were placed; three tubes with short collars (10 cm depth) allowing both heterotrophic and rhizosphere respiration, three tubes with longer collars (40 cm depth) with no windows to exclude both roots and mycorrhizae.

Canopy respiration (R_{leaves}): Leaf gas exchange measurements of R_{dark} were performed for at least 20 trees using infra-red gas analysers. To obtain the leaves, one branch each from sunlit and shaded portions of canopy trees were randomly selected and immediately re-cut under water to restore hydraulic connectivity for subsequent gas exchange measurement. The leaves were fully darkened for 30 min prior to measuring R_{dark} . To scale to whole-canopy respiration, mean dark respiration for sunlit and shade-lit leaves were multiplied by the respective estimated fractions of total LAI.

The mean R_{leaves} measured for sun leaves and shade leaves was applied to the sun and shade fractions respectively. The estimation of sun and shade fractions is described in the section on LAI above.

The wet season respiration mean was applied to all months with > 100 mm rain; for the dry season months, measured dry season respiration was linearly scaled by the soil moisture saturation to allow for more continuous variation of leaf respiration. To account for daytime light inhibition of leaf dark respiration, we apply an inhibition factor: 67% of daytime leaf dark respiration, 33% of total leaf dark respiration (Malhi et al., 2009). These were calculated by applying the Atkin et al. (2000) equations for light inhibition of leaf respiration to a plot in Tapajós forest in Brazil (Malhi et al., 2009; Lloyd et al., 2010). In recognition of the substantial uncertainty in this estimate, we assigned a 30% error to the multiplying factor.

Above ground live wood respiration (R_{stems}): Bole respiration was measured using a closed dynamic chamber method, from 25 trees distributed evenly throughout each plot at 1.3 m height with an IRGA (EGM-4) and soil respiration chamber (SRC-1) connected to a permanent collar, sealed to the tree bole surface. To estimate plot-level stem respiration, tree respiration per unit bole area was multiplied by bole surface area (SA in m^2) for each tree, estimated with the following equation (Chambers et al., 2004): $\log_{10}(\text{SA}) = -0.105 - 0.686 \log(\text{DBH}) + 2.208 \log(\text{DBH})^2 - 0.627 \log(\text{DBH})^3$, where DBH is bole diameter in cm at 1.3 m height. Finally, for all 25 trees together, we regressed mean annual bole respiration against total annual growth. In recognition of the substantial uncertainty in this estimate, we assigned a 30% error to the multiplying factor.

Coarse root respiration ($R_{\text{coarse roots}}$): A substantial amount of wood respiration may occur in or near the root core immediately below the bole, but this has rarely been measured and is not included in our soil respiration partitioning methodology. In addition, even small coarse roots are too slow-growing to be present in 3-monthly ingrowth cores. We therefore estimate this term separately. This component of respiration was not measured directly but estimated by multiplying above-ground live wood respiration by 0.21 (same ratio used in these studies to estimate coarse root biomass and growth – see above). To our knowledge, there are no available data on below-ground coarse root surface area for tropical forests, so a mass-based approach was used. In recognition of the substantial uncertainty in this estimate, we assigned a 50% error (± 0.10) to the multiplying factor. Malhi et al. (2009) did not account for this term, but it seems appropriate to include it for a more complete description of the below-ground carbon budget.

Summations

NPP is then calculated as:

$$\text{NPP} = \text{NPP}_{\text{ACW} \geq 10 \text{ cm}} + \text{NPP}_{\text{ACW} < 10 \text{ cm}} + \text{NPP}_{\text{litterfall}} + \text{NPP}_{\text{branch turnover}} + \text{NPP}_{\text{herbivory}} + \text{NPP}_{\text{fineroot}} + \text{NPP}_{\text{coarse root}} \quad (1)$$

Total autotrophic respiration is estimated as

$$R_a = R_{\text{leaf}} + R_{\text{stem}} + R_{\text{rhizosphere}} + R_{\text{coarse root}} \quad (2)$$

In plant-level autotrophic steady state conditions (and on annual timescales or longer where there is little net non-structural carbohydrate storage), gross primary productivity (GPP), the carbon taken up via photosynthesis, should be approximately equal to plant carbon expenditure (PCE), the amount of carbon used for NPP and autotrophic plant respiration (R_a). Note that the autotrophic steady state condition does not require the total plot carbon cycle to be in equilibrium, the plot can still be gaining or losing biomass or soil carbon stocks, as long as there is no substantial accumulation or loss of non-structural carbohydrates. Hence, we estimated GPP as

$$\text{GPP} = \text{NPP} + R_a \quad (3)$$

We calculated the Carbon Use Efficiency (CUE) as the proportion of total GPP invested in NPP rather than R_a :

$$\text{CUE} = \text{NPP} / \text{GPP} = \text{NPP} / (\text{NPP} + R_a) \quad (4)$$

The above-ground NPP estimation neglects several small NPP terms, such as NPP lost as volatile organic emissions, non-measured litter trapped in the canopy, or dropped from ground flora below the litter traps. At a site in central Amazonia volatile emissions were found to be a minor component of the carbon budget ($0.13 \pm 0.06 \text{ Mg C ha}^{-1} \text{ year}^{-1}$; Malhi et al. 2009), and a close comparison with flux tower data and near-closure of the carbon budget in three well-studied Brazilian Amazonian sites, suggests that the other neglected NPP terms are relatively minor (Malhi et al. 2009). For below-ground NPP, the allocation to root exudates and to mycorrhizae is neglected. In effect, we treat root exudation and transfer to mycorrhizae as rhizosphere autotrophic respiration rather than as NPP, which could potentially impact our CUE numbers. Given that these exudates are labile and rapidly respired by mycorrhizae and soil microfauna in the rhizosphere, in terms of carbon cycling this exudate NPP term is very similar to fine root autotrophic respiration.

Statistics and error analysis

A key consideration was assignment and propagation of uncertainty in our measurements. There were two primary types of uncertainty. Firstly, there was uncertainty associated with the spatial

heterogeneity of the study plot and the limited number of samples. Examples include the variability among litter traps, or among fine root ingrowth cores. Secondly, there was uncertainty associated with either unknown error in measurement, or error in scaling measurements to the plot level. Examples of unknown biases included the possibility of soil-derived CO₂ in the transpiration stream affecting the stem CO₂ efflux measurements, and uncertainties in scaling include the allometry of scaling of bole stem CO₂ efflux to whole tree stem respiration, or leaf dark respiration to whole canopy dark respiration. Here we assumed that most NPP terms were measured fairly precisely and sampled without large biases; hence the NPP component measurements were dominated by sampling uncertainty, which could be reliably estimated, assuming a normal distribution. On the other hand, some of the main autotrophic respiration terms were dominated by systematic uncertainty. This systematic uncertainty can be very hard to reliably quantify; here, in each case we made an explicit and conservative estimate of the systematic uncertainty of key variables. Key sources of systematic uncertainty applied are detailed above and in the site-specific papers.

All estimated fluxes reported in this study are in Mg C ha⁻¹ year⁻¹ (or month⁻¹ for seasonal fluxes) and all reported errors are ± 1 s.e.; error propagation was carried out for all combination quantities using standard rules of quadrature (Hughes & Hase, 2010), assuming that uncertainties were independent and normally distributed. One Mg C ha⁻¹ year⁻¹ is equal to 100 g C m⁻² year⁻¹, or 0.264 μmol C m⁻² s⁻¹.

References

- Araujo-Murakami A, Doughty CE, Metcalfe DB, Silva-Espejo JE, Arroyo L, Heredia JP, Flores M, Sibling R, Mendizabal LM, Pardo-Toledo E, et al. 2014.** The productivity, allocation and cycling of carbon in forests at the dry margin of the Amazon forest in Bolivia. *Plant Ecology & Diversity* 7: 1–15.
- Atkin OK, Evans JR, Ball MC, Lambers H, Pons TL. 2000.** Leaf respiration of snow gum in the light and dark. Interactions between temperature and irradiance. *Plant Physiology* 122: 915–923.
- Cairns MA, Brown S, Helmer EH, Baumgardner GA. 1997.** Root biomass allocation in the world's upland forests. *Oecologia* 111: 1–11.
- Chambers JQ, Tribuzy ES, Toledo LC, Crispim BF, Higuchi N, Santos JD, Araújo AC, Kruijt B, Nobre AD, Trumbore SE. 2004.** Respiration from a tropical forest ecosystem: partitioning of sources and low carbon use efficiency. *Ecological Applications* 14: 72–88.
- Chave J, Andalo C, Brown S, Cairns MA, Chambers JQ, Eamus D, Fölster H, Fromard F, Higuchi N, Kira T, et al. 2005.** Tree allometry and improved estimation of carbon stocks and balance in tropical forests. *Oecologia* 145: 87–99.
- Clark KE, West AJ, Hilton RG, Asner GP, Quesada CA, Silman MR, Saatchi SS, Farfan-Rios W, Martin RE, Horwath AB, et al. 2016.** Storm-triggered landslides in the Peruvian Andes and implications for topography, carbon cycles, and biodiversity. *Earth Surface Dynamics*, 4, 47-70.
- da Costa ACL, Metcalfe DB, Doughty CE, de Oliveira AAR, Neto GFC, da Costa MC, Silva Junior J de A, Aragão LEOC, Almeida S, Galbraith DR, et al. 2014.** Ecosystem respiration and net primary productivity after 8–10 years of experimental through-fall reduction in an eastern Amazon forest. *Plant Ecology & Diversity* 7: 1–18.
- del Aguila-Pasquel J, Doughty CE, Metcalfe DB, Silva-Espejo JE, Girardin CAJ, Gutierrez JAC, Navarro-Aguilar GE, Quesada CA, Hidalgo CG, Huaymacari JMR, et al. 2014.**

The seasonal cycle of productivity, metabolism and carbon dynamics in a wet aseasonal forest in north-west Amazonia (Iquitos, Peru). *Plant Ecology & Diversity* 7: 1–13.

Doughty CE, Goulden ML. 2008. Seasonal patterns of tropical forest leaf area index and CO₂ exchange. *Journal of Geophysical Research* 113: G00B06.

Doughty CE, Metcalfe DB, da Costa MC, de Oliveira AAR, Neto GFC, Silva JA, Aragão LEOC, Almeida SS, Quesada CA, Girardin CAJ, et al. 2014. The production, allocation and cycling of carbon in a forest on fertile terra preta soil in eastern Amazonia compared with a forest on adjacent infertile soil. *Plant Ecology & Diversity* 7: 1–13.

Feldpausch TR, Banin L, Phillips OL, Baker TR, Lewis SL, Quesada CA, Affum-Baffoe K, Arets EJMM, Berry NJ, Bird M, et al. 2011. Height-diameter allometry of tropical forest trees. *Biogeosciences* 8: 1081–1106.

Girardin CAJ, Espejo JES, Doughty CE, Huasco WH, Metcalfe DB, Durand-Baca L, Marthews TR, Aragão LEOC, Farfán-Rios W, García-Cabrera K, et al. 2014. Productivity and carbon allocation in a tropical montane cloud forest in the Peruvian Andes. *Plant Ecology & Diversity* 7: 107–123.

Huaraca Huasco W, Girardin CAJ, Doughty CE, Metcalfe DB, Baca LD, Silva-Espejo JE, Cabrera DG, Aragão LEOC, Davila AR, Marthews TR, et al. 2014. Seasonal production, allocation and cycling of carbon in two mid-elevation tropical montane forest plots in the Peruvian Andes. *Plant Ecology & Diversity* 7: 1–18.

Hughes I, Hase T 2010. *Measurements and Their Uncertainties*. Oxford, UK: Oxford University Press.

Jackson RB, Canadell J, Ehleringer JR, Mooney HA, Sala OE, Schulze ED 1996. A global analysis of root distributions for terrestrial biomes. *Oecologia* 108: 389–411.

Lloyd J, Patino S, Paiva RQ, Nardoto GB, Quesada CA, Santos AJB, Baker TR, Brand WA, Hilke I, Gielmann H, et al. 2010. Optimisation of photosynthetic carbon gain and within-canopy gradients of associated foliar traits for Amazon forest trees. *Biogeosciences* 7:

1833–1859.

Malhi Y, Amézquita FF, Doughty CE, Silva-Espejo JE, Girardin CAJ, Metcalfe DB, Aragão LEOC, Huaraca-Quispe LP, Alzamora-Taype I, Eguiluz-Mora L, et al 2014.

The productivity, metabolism and carbon cycle of two lowland tropical forest plots in south-western Amazonia, Peru. *Plant Ecology & Diversity* 7: 1–21.

Malhi Y, Aragão LEOC, Metcalfe DB, Paiva R, Quesada CA, Almeida S, Anderson L, Brando P, Chambers JQ, da Costa ACL, et al. 2009. Comprehensive assessment of carbon productivity, allocation and storage in three Amazonian forests. *Global Change Biology* 15: 1255–1274.

Martin AR, Thomas SC 2011. A reassessment of carbon content in tropical trees (ed Chave J). *PLoS ONE* 6: e23533.

Metcalfe DB, Asner GP, Martin RE, Silva-Espejo JE, Huaraca Huasco W, Farfán Amézquita FF, Carranza-Jimenez L, Galiano Cabrera DF, Durand Baca L, Sinca F, et al. 2014. Herbivory makes major contributions to ecosystem carbon and nutrient cycling in tropical forests. *Ecology Letters* 17: 324–332.

Metcalfe DB, Meir P, Aragao LEOC, Lobo-do-Vale R, Galbraith D, Fisher RA, Chaves MM, Maroco JP, da Costa ACL, de Almeida SS, et al. 2010. Shifts in plant respiration and carbon use efficiency at a large-scale drought experiment in the eastern Amazon. *New Phytologist* 187: 608–621.

Metcalfe DB, Williams M, Aragao LEOC, da Costa ACL, de Almeida SS, Braga AP, Gonçalves PHL, de Athaydes J, Junior S, Malhi Y, et al. 2007. A method for extracting plant roots from soil which facilitates rapid sample processing without compromising measurement accuracy. *New Phytologist* 174: 697–703.

Rocha W, Metcalfe DB, Doughty CE, Brando P, Silvério D, Halladay K, Nepstad DC, Balch JK, Malhi Y. 2014. Ecosystem productivity and carbon cycling in intact and annually burnt forest at the dry southern limit of the Amazon rainforest (Mato Grosso, Brazil). *Plant Ecology & Diversity* 7: 1–16.

