



This is a repository copy of *Drought versus heat: What's the major constraint on Mediterranean green roof plants?*.

White Rose Research Online URL for this paper:
<http://eprints.whiterose.ac.uk/101229/>

Version: Accepted Version

Article:

Savi, T., Dal Borgo, A., Love, V.L. orcid.org/0000-0001-7128-3018 et al. (3 more authors) (2016) Drought versus heat: What's the major constraint on Mediterranean green roof plants? *Science of the Total Environment*, 566-56. pp. 753-760. ISSN 0048-9697

<https://doi.org/10.1016/j.scitotenv.2016.05.100>

Article available under the terms of the CC-BY-NC-ND licence
(<https://creativecommons.org/licenses/by-nc-nd/4.0/>)

Reuse

This article is distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs (CC BY-NC-ND) licence. This licence only allows you to download this work and share it with others as long as you credit the authors, but you can't change the article in any way or use it commercially. More information and the full terms of the licence here: <https://creativecommons.org/licenses/>

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

1 **Drought versus heat: what's the major constraint on Mediterranean green**
2 **roof plants?**

3
4 Tadeja SAVI^{a*}, Anna DAL BORGO^a, Veronica L. LOVE^{a,b}, Sergio ANDRI^c, Mauro TRETACH^a,
5 and Andrea NARDINI^a

6
7 a) Dipartimento di Scienze della Vita, Università di Trieste, Via L. Giorgieri 10, 34127 Trieste, Italia

8 b) Department of Landscape, University of Sheffield, Western Bank, Sheffield, South Yorkshire, S10
9 2TN, United Kingdom

10 c) Harpo seic verdepensile, Via Torino 34, 34123 Trieste, Italia

11

12 * Corresponding author: tsavi@units.it

13

14 Anna Dal Borgo: dalborgo.anna@gmail.com

15 Veronica Lee Love: vllove1@sheffield.ac.uk

16 Sergio Andri: s.andri@seic.it

17 Mauro Tretiach: tretiach@units.it

18 Andrea Nardini: nardini@units.it

19

20

21

1 **ABSTRACT**

2 Green roofs are gaining momentum in the arid and semi-arid regions due to their multiple benefits as
3 compared with conventional roofs. One of the most critical steps in green roof installation is the
4 selection of drought and heat tolerant species that can thrive under extreme microclimate conditions.
5 We monitored the water status, growth and survival of 11 drought-adapted shrub species grown on
6 shallow green roof modules (10 and 13 cm deep substrate) and analyzed traits enabling plants to cope
7 with drought (symplastic and apoplastic resistance) and heat stress (root membrane stability). The
8 physiological traits conferring efficiency/safety to the water transport system under severe drought
9 influenced plant water status and represent good predictors of both plant water use and growth rates
10 over green roofs. Moreover, our data suggest that high substrate temperature represents a stress factor
11 affecting plant survival to a larger extent than drought per se. In fact, the major cause influencing
12 seedling survival on shallow substrates was the species-specific root resistance to heat, a single and
13 easy measurable trait that should be integrated into the methodological framework for screening and
14 selection of suitable shrub species for roof greening in the Mediterranean.

15

16 **Keywords:** drought resistance, heat resistance, shallow substrate depths, shrub species, water status,
17 mortality

18 **INTRODUCTION**

19 Green roofs are engineered ecosystems representing an effective strategy to address some of the most
20 challenging environmental issues in urban areas (Castleton et al., 2010; Berardi et al., 2014). In
21 particular, green roofs have the potential to mitigate the quantity and quality of storm-water runoff,
22 provide thermal insulation to buildings with related energy savings, extend the roof lifespan, mitigate
23 the ‘urban heat island’, and provide space and habitats for urban biodiversity (Castleton et al., 2010;
24 Madre et al., 2014; Benvenuti and Bacci, 2010; Cao et al., 2014; Vijayaraghavan and Raja, 2014).
25 Extensive green roofs, characterized by shallow substrate, reduced weight and low maintenance costs,

1 represent an innovative, energy-saving solution (Van Mechelen et al., 2014; Price et al., 2011). Over
2 the last decades, the urban areas covered by green roofs has substantially increased in North and
3 Central Europe and in temperate and sub-tropical regions worldwide (Castleton et al., 2010; Madre et
4 al., 2014; Berardi et al., 2014; Thuring and Grant, 2015). More recently, research has focused
5 on the implementation of green roofs in Mediterranean regions, where high temperatures and
6 prolonged drought significantly challenge plant survival in these artificial habitats (Olivieri et
7 al., 2013; Benvenuti and Bacci, 2010; Raimondo et al., 2015; Rayner et al., 2015).

8 A fundamental question addressed by Mediterranean green roof research is how to increase
9 water retention capacity while keeping the substrate depth at a minimum. In fact, reducing
10 substrate depth to limit installation costs apparently contrasts with the need to maximize the
11 amount of water available to vegetation, and to minimize temperature extremes. In fact,
12 another important aim of recent studies has been the selection of drought tolerant species that
13 can survive the extreme green roof conditions in these hot and arid regions. There is evidence
14 that targeted substrate amendments with hydrogel, peat, and biochar, or modifications to the
15 layering design (substrate particle size, drainage panels etc.), have the potential to enhance the
16 moisture retention properties of green roofs, thus increasing the volume of water available and
17 improving plant water status and survival (Savi et al., 2013; Cao et al., 2014; Savi et al.,
18 2014; Vijayaraghavan and Raja, 2014; Raimondo et al., 2015). Several criteria have been
19 proposed to optimize species' selection for green roofs, but these are mainly based on
20 ecological or morpho-anatomical approaches (Lundholm, 2006; Caneva et al., 2015; Van
21 Mechelen et al., 2014; Rayner et al., 2015). Moreover, most screening studies have been
22 focused on succulents or herbaceous species (Benvenuti and Bacci, 2010; Price et al., 2011;
23 Van Mechelen et al., 2014; Rayner et al., 2015), while studies on shrubs as potential growth
24 forms for green roof vegetation are still limited. Indeed, shrubs are generally characterized by
25 a higher capacity in stomatal control of transpiration than herbaceous plants (Galmés et al.,

1 2007; Farrell et al., 2013) and should be taken into serious consideration when selecting
2 potential species assemblages for Mediterranean green roofs. Moreover, a selection process
3 based on an ecophysiological approach might be more effective, at least when functional traits
4 enabling plants to cope with stress factors, like drought and high temperature, are properly
5 analyzed and quantified.

6 Plant tolerance to drought stress is commonly quantified in terms of symplastic and apoplastic
7 vulnerability to dehydration. The former is generally correlated to the water potential
8 inducing loss of cell turgor (Ψ_{tp} , Bartlett et al., 2012). Low Ψ_{tp} values allow drought-adapted
9 plants to maintain cell turgor, stomatal aperture, and positive carbon gain even under low soil
10 water availability and/or high atmospheric evaporative demand. On the other hand, apoplastic
11 vulnerability to water stress is generally quantified in terms of xylem vulnerability to
12 embolism formation. In fact, intense or prolonged drought can affect the root-to-leaf water
13 transport by causing the breakage of water columns in xylem conduits (Tyree and Sperry,
14 1989), potentially leading to plant desiccation and death (Nardini et al., 2014b). Xylem
15 hydraulic vulnerability is generally quantified in terms of P50 i.e., the xylem water potential
16 inducing 50% loss of hydraulic conductivity (Choat et al., 2012), with species displaying
17 lower P50 generally performing better under drought stress (Nardini et al., 2013) than species
18 with relatively higher P50 values.

19 Water availability aside, high temperatures can also pose serious limitations to plant
20 performance on green roofs. Heat stress can alter both membrane stability and enzymatic
21 function and thus affects photosynthesis and respiration, altering carbon gain, growth, and
22 secondary metabolism at the root and shoot levels (Wahid et al., 2007; Huang et al., 2012;
23 Vile et al., 2012). Most importantly, shallow green roof substrates potentially expose root
24 systems to temperature extremes that largely surpass those experienced by plants in natural

1 soils. In fact, the root system is generally more vulnerable to heat stress compared to the shoot
2 (Kuroyanagi and Paulsen, 1988). The co-occurrence of both drought and heat stress over
3 green roofs poses important challenges to plant life, frequently leading to foliage desiccation,
4 plant die-back, and ultimately death (Allen et al., 2010; Price et al., 2011; Nardini et al.,
5 2013; Rayner et al., 2015), and also complicates the identification of key physiological traits
6 allowing to predict plant performance on green roofs installed in arid regions.

7 To the best of our knowledge, a comparative study of physiological traits conferring
8 resistance to drought and heat stress has never been coupled to the monitoring of plant
9 performance on extensive green roofs. In this study, we contribute to this literature gap, by
10 analyzing the performance in terms of growth and survival of 11 Mediterranean shrub species,
11 established on shallow green roof experimental modules, as related to several indicators of
12 their physiological vulnerability to water stress and high temperatures. We monitored plant
13 water status, leaf symplastic resistance to drought and stem vulnerability to xylem embolism,
14 as well as root resistance to heat stress. We aimed at understanding which functional traits
15 underlie plant performance and survival on Mediterranean green roofs. Our main hypothesis
16 was that plant physiological traits conferring efficiency/safety to the water transport system
17 under severe drought, as well as root resistance to heat stress, significantly influence the
18 overall plant performance and survival. Moreover, on the basis of the results, we propose a
19 methodological framework for screening and selection of suitable shrub species for roof
20 greening in the Mediterranean.

21

22 **2. MATERIALS AND METHODS**

23 **2.1. Study area and experimental set-up**

1 The study was carried out between 2013 and 2015 on the experimental green roof installed on
2 the rooftop of the Dept. of Life Sciences, University of Trieste (NE Italy; 45° 39'40''N,
3 13°47'40''E). Trieste lies on the upper Adriatic coast and it is characterized by a sub-
4 Mediterranean climate, with mild winters and relatively warm, dry summers. Mean annual
5 temperatures in the period 1994-2015 (www.osmer.fvg.it) averaged 15.7°C (highest 25.1°C in
6 July, lowest 7.0°C in January). Maximum daily temperatures frequently exceed 30°C in
7 summer, while in winter the minimum values drop under 0°C only occasionally. Mean annual
8 rainfall is 869 mm, with relatively dry periods in July and January-February. Snow events are
9 rare. The prevalent wind (Bora) blows from E-NE.

10 The experimental extensive green roof was composed of 10 modules, each covering an area of
11 2.5 m². Modules were built with a six-layer system by SEIC (Harpo Spa, Trieste, Italy),
12 consisting of: a waterproof/root resistant membrane, a moisture retention layer, a drainage
13 layer, a filter membrane, and substrate (dry bulk density=848 kg m⁻³, organic matter=2.9%,
14 pH=6.8, water content at saturation=0.44 g g⁻¹). The experimental modules were filled with
15 10 (D-10) or 13 (D-13) cm deep substrate (5 modules per depth). Each module had an
16 independent discharge for excess water runoff, and was equipped with a temperature sensor
17 (TT-500, Tecno.el srl, Roma, Italy) installed at the maximum substrate depth and a soil
18 moisture content sensor (WC, EC-5, Decagon Devices, Pullman, WA, USA), both recording
19 values at 1 h time intervals. In April 2013, the modules were vegetated with 11 woody species
20 belonging to the Mediterranean and sub-Mediterranean flora (Pignatti, 2002). In particular,
21 both evergreen (Cistus salvifolius L., Ligustrum vulgare L., Phillyrea angustifolia L., Pistacia
22 lentiscus L., Salvia officinalis L.) and deciduous species (Cotinus coggygia Scop., Emerus
23 majus Mill., Paliurus spina-christi Mill., Prunus mahaleb L., Pyrus pyraster Burgsd.,
24 Spartium junceum L., Conti et al., 2008) were used (Table 1). The species were selected on

1 the basis of the characteristics of their natural habitat (shallow soils with low water storage
2 capacity, high temperatures) and their known tolerance to drought stress (Pignatti, 2002;
3 Choat et al., 2012; Raimondo et al., 2015; Savi et al., 2015). The 2-3 year-old potted plants
4 were provided by either a public (Regional Forestry Service, Tarcento, Italy) or a private
5 nursery (Vita Verde, Bologna, Italy) in small pots of approximately 0.5 L. The substrate of
6 the pots was carefully removed and the roots gently rinsed with water. Four individuals per
7 species were randomly transplanted in each experimental module at a minimum distance of
8 20 cm between individuals, and abundantly irrigated. Moreover, 10 individuals per species
9 were transplanted in 2 liters pots filled with the same green roof substrate (13 cm deep), and
10 maintained nearby experimental modules for additional physiological measurements (see
11 below). During the study period, both potted plants and those growing on green roof modules
12 received natural rainfall and additional emergency irrigation only during severe drought
13 (about 25 mm over the whole summer season).

14 Microclimatic parameters (i.e., wind, air temperature and humidity, irradiance, precipitations)
15 during the study period were recorded by a weather station installed near the modules (Savi et
16 al., 2015; S1).

17

18 **2.2. Plant water status**

19 Plant water status was assessed in terms of pre-dawn (Ψ_{pd}) and minimum (Ψ_{min}) water
20 potential, and leaf conductance to water vapor (g_L). Measurements were performed on two
21 subsequent sunny days in June 2014 (high water availability) and August 2014 (dry period).
22 During these days the substrate water potential ($\Psi_{substrate}$) recorded in D-10 modules was -
23 0.04 ± 0.02 MPa and -2.40 ± 0.65 MPa, while in D-13 modules was -0.06 ± 0.02 MPa and -

1 **3.90±1.67 MPa for June and August, respectively.** At 0500 h, at least three leaves per species
2 (one leaf from each of three randomly selected individuals) and per substrate depth were
3 detached, wrapped in cling-film, and inserted in plastic bags. Leaves were immediately
4 transported in the laboratory and their Ψ_{pd} was measured with a pressure chamber (mod.
5 1505D, PMS Instruments, Albany, OR, USA). On the same days, g_L was measured at midday
6 on at least three leaves per species and per substrate depth using a porometer (SC1, Decagon
7 Devices, Pullman, WA, USA). After g_L measurements, leaves were sampled and transported
8 to the laboratory for Ψ_{min} determination as described above. **The water status measurements**
9 **were performed on mature, fully expanded, healthy, and undamaged leaves.**

10

11 **2.3. Physiological traits**

12 Leaf water potential isotherms (PV-curves) were measured in July 2014 to evaluate the
13 symplastic drought tolerance of the study species (Lenz et al., 2006). At least three leaves per
14 species were detached in the morning from different potted individuals and rehydrated for 30
15 min while wrapped in cling film. The initial leaf water potential (Ψ_{leaf}) was measured with the
16 pressure chamber, followed by fresh weight measurements (FW). Leaves were left
17 dehydrating on the bench and sequential measurements of Ψ_{leaf} and FW were performed until
18 the relationship between $1/\Psi_{leaf}$ and cumulative water loss became linear. PV-curve
19 elaboration (Tyree and Hammel, 1972) led to the extrapolation of the osmotic potential at full
20 turgor (π_0) and the water potential at turgor loss point (Ψ_{tlp}).

21 To assess species-specific vulnerability to drought-induced xylem embolism, stem
22 vulnerability curves (VCs) of the 11 study species were measured using the air injection
23 method in summer 2015 (Ennajeh et al., 2011; Cochard et al., 2013). Potted plants were

1 abundantly irrigated and after 24 h were cut under water at the root collar. The stem was re-
2 cut under water several times at both ends to the final length, corresponding to 1.5 times the
3 maximum vessel length, as estimated with the air-injection method (Jacobsen et al., 2012), to
4 avoid possible artefacts due to the presence of xylem conduits open at both sample ends
5 (Ennajeh et al., 2011). The basal end was connected to a tubing system and flushed with a
6 perfusion solution (10 mM KCl) filtered at 0.2 μm for 30 min, under a pressure (P) of 0.18
7 MPa. The stem was then inserted through a 10 cm long double-ended pressure chamber and
8 perfused with the reference solution at low pressure (5 kPa). The diameter of the tubing
9 connected to the sample was large enough to allow the escape of air bubbles originating from
10 the sample during pressurization. The flow (F) was measured by collecting effluent with pre-
11 weighed vials filled with absorbent material over 1-min intervals (F_{max} , average of five
12 measurements). The pressure in the chamber was progressively increased by 0.5 MPa
13 intervals and F was measured after 5 min equilibration at each pressure level. The percentage
14 loss of hydraulic conductivity (PLC) was calculated as $\text{PLC}=1-(F/F_{\text{max}})\times 100$. At least three
15 individuals per species were analyzed and PLC data corresponding to each applied pressure
16 were averaged in a single VC. As a reference parameter indicating species-specific
17 vulnerability to xylem embolism (Choat et al., 2012), the value of xylem pressure inducing
18 50% loss of stem hydraulic conductivity (P50) was calculated from VCs.

19

20 **2.4. Plant growth and mortality**

21 In May 2013, the diameter at the root collar (D_i , calculated as the mean of two measurements
22 taken at 90° angles), of all transplanted individuals was measured with a digital caliper
23 (Absolute Coolant-Proof, Mitutoyo, Aurora, IL, USA). The diameter was re-measured in
24 September 2014 (D_f) and the relative diameter increment was calculated as: $G=(D_f/D_i)-1\times 100$.

1 The aim of these measurements was to estimate the species' growth rate after two years of
2 establishment on the D-10 or D-13 modules.

3 Drought survival of the study species growing in the two substrate depths was estimated in
4 September 2014 and 2015 on the basis of visual assessments. Desiccated plants without green
5 foliage and vital buds were considered dead (D). Individuals with turgid and healthy leaves,
6 even if limited in number, were classified as living plants (L). Species-specific mortality rates
7 (M) for each category of substrate depth was calculated as the ratio between dead plants and
8 the number of all planted individuals (D/D+L).

9

10 **2.5. Root vulnerability to heat stress**

11 On the basis of the significant differences found in substrate temperature and plant mortality
12 between D-10 and D-13 modules (see Results), a laboratory experiment was performed in
13 September 2015 to evaluate species-specific vulnerability of roots to heat stress. Root cell
14 membrane stability at high temperatures was estimated with electrolyte leakage tests. Four
15 potted plants per species were gently uprooted to collect about 200 mg (fresh weight) of fine
16 roots (diameter < 1 mm), which were rinsed with water and placed in two tubes (100 mg each)
17 containing 1.5 ml of deionized water. The tubes were shaken for 1 h at laboratory temperature
18 to eliminate remaining debris and ions entrapped in the root cortex apoplast (apparent free
19 space, Bernstein and Nieman, 1960). The solution was afterward discarded and 1.5 ml of
20 fresh deionized water was added to the samples. One tube per plant was incubated for 30
21 minutes in a bath containing water at 45°C (T, treatment), while the second tube was kept at
22 lab temperature (C, control). After the heat stress treatments, all samples were allowed to
23 reach room temperature, and the initial electrical conductivity (C_i) of the solution was

1 measured (Twin Cond B-173, Horiba, Kyoto, Japan). Both T and C samples were then
2 subjected to 3 freezing-thawing cycles (1 min in liquid N₂ followed by 30 min at room
3 temperature) and the final electrical conductivity was measured (C_f). The relative leakage
4 ratio was calculated as: $REL=(C_i/C_f)\times 100$. The root cell membrane vulnerability to heat stress
5 was estimated as: $\Delta REL=REL_T-REL_C$.

6

7 **2.6. Statistical analysis**

8 Statistical significance of differences and correlations was tested on the basis of unpaired
9 Student's t-test and Pearson product-moment correlation. **A multiple regression analysis was**
10 **carried out to evaluate physiological traits as potential predictors of plant growth and**
11 **mortality (R v. 3.2.2; www.r-project.org).** All results were considered statistically significant
12 at $P\leq 0.05$. Means are reported \pm standard error of the mean (SEM).

13

14 **3. RESULTS**

15 **Table 2** reports the values of functional traits derived from PV-curves and stem VCs
16 elaboration, as well as growth rates (G) assessed two years after planting. The overall mean
17 Ψ_{tlp} and π_0 of the study species were -1.92 ± 0.15 MPa and -1.42 ± 0.12 MPa, respectively. The
18 species with the lowest (more negative) values of Ψ_{tlp} and π_0 was *P. lentiscus*, while the
19 highest values were recorded for *S. junceum*. P50 values ranged between -1.55 MPa in *P.*
20 *lentiscus* (high vulnerability to drought-induced xylem dysfunction) and -5.00 MPa in *L.*
21 *vulgare* (high resistance to embolism). Over two growing seasons, the diameter at the root
22 collar increased by 60% and 84% in plants growing on 10 and 13 cm deep substrate,
23 respectively. The G of *P. pyrastrer* individuals was not assessed due to high mortality in this

1 species (see below). Interestingly, G was not correlated to P50, but a positive and significant
2 correlation emerged with symplastic drought tolerance (Table 4, S2). Indeed the lowest G was
3 recorded in *P. lentiscus* and the highest in *S. junceum*. A positive correlation was also
4 observed between Ψ_{tp} or π_0 and plant water status as recorded in June and August, in both D-
5 10 and D-13 modules (Table 4). Overall, species characterized with lower Ψ_{tp} and π_0 showed
6 more negative Ψ_{pd} and Ψ_{min} , as well as lower g_L values. For example, in June *S. junceum* had
7 the most favorable water status, while the lowest values of Ψ_{pd} , Ψ_{min} , and g_L were again found
8 in *P. lentiscus*. Unfortunately, it was not possible to measure the g_L for *S. junceum* due to its
9 small and drought-deciduous leaves (Pignatti, 2002). In August, *P. angustifolia* experienced
10 the least favorable water status, reaching a Ψ_{min} of -4.2 MPa ($\Psi_{\text{tp}}=-2.49$ MPa) and a g_L of
11 about $110 \text{ mmol m}^{-2} \text{ s}^{-1}$ (the lowest after that of *P. lentiscus*).

12 Overall, the results point to a slightly more favorable water status in plants grown on 10 than
13 on 13 cm deep substrate. In particular, the mean Ψ_{min} for all shrubs recorded in June was
14 found to be -1.16 ± 0.07 and 1.39 ± 0.10 MPa for D-10 and D-13 plants, respectively ($P=0.08$).
15 Moreover, the Ψ_{pd} in *P. mahaleb* and *P. spina-christi* was about 0.3 MPa more negative in
16 plants grown on deeper substrate ($P<0.05$). Nevertheless, plants classified as on the basis of
17 complete desiccation of their aerial portion were about 44% in D-10 modules and only 20% in
18 D-13 ones ($P<0.05$), with notable differences among species (Fig. 1). **Plants' mortality**
19 **measured at the end of summer 2014 (data not shown) was not statistically different of data**
20 **recorded in 2015.** The lowest mortality rate was recorded for *P. angustifolia* (no dead plants
21 in D-13), while the highest rates were found in *P. pyraeaster* (average $M=71.1\%$) and *P.*
22 *lentiscus* (average $M=62.5\%$). No striking correlations were highlighted between M and plant
23 water status, as well as Ψ_{tp} and π_0 . Surprisingly, a highly significant relationship ($P<0.01$) was

1 observed between M and P50 in plants growing on 10 cm deep substrate but not in those
2 growing on 13 cm (Table 4, S2).

3 Data on soil temperature at the maximum substrate depth revealed marked differences
4 between the two categories of substrate depth. In particular, the temperatures recorded on a
5 representative warm, summer day (mean air temperature=29.6°C) ranged between 26.5 and
6 43.6°C in 10 cm deep substrate, while the range was 29.3–39.2°C for the 13 cm deep
7 substrate (Fig. 2A). The average daily thermal regime of the substrate in July (the hottest
8 month) was about 15°C in D-10 and only 10°C in D-13 modules. Moreover, the maximum
9 temperature peak was usually delayed by 2 hours in deeper modules (2000 h) if compared to
10 the shallower ones (1800 h). A highly significant difference ($P<0.001$) was observed in terms
11 of absolute daily maximum substrate temperatures reached during the study period between
12 D-10 ($43.8\pm 0.49^\circ\text{C}$) and D-13 ($39.4\pm 0.68^\circ\text{C}$) modules (Fig. 2B).

13 Figure 3 summarizes the results of experiments designed to estimate the root vulnerability to
14 heat stress. Cell membrane sensitivity to high temperatures, estimated as ΔREL , ranged from
15 about 6% (low vulnerability to heat stress) to about 22% (high vulnerability to heat stress), as
16 recorded in *C. coggygria* and *P. pyraister*, respectively. ΔREL was found to be significantly
17 correlated with plant mortality in both 10 ($P=0.02$) and 13 ($P=0.001$) cm deep modules.

18

19 **4. DISCUSSION**

20 Our results provide experimental evidence that species-specific functional traits are useful and
21 reliable proxies of plant performance on green roofs installed in Mediterranean-climate
22 regions. In particular, our data suggest that traits conferring resistance to drought and high

1 substrate temperatures represent the essential trademarks of plant species to be used for roof
2 greening in warm and dry climates.

3 Our study was focused on the analysis of traits conferring symplastic and apoplastic drought
4 tolerance, in terms of maintenance of positive turgor and efficient root-to-leaf pathway, both
5 of which ensure maintenance of gas exchange rates and plant survival under drought
6 conditions. The wide spectrum of Ψ_{tlp} , π_0 , and P50 values recorded in the study species
7 support the hypothesis that Mediterranean plants are flexible in their adaptation to drought
8 and in fact display a range of different hydraulic strategies (Galmés et al., 2007; Nardini et
9 al., 2014a).

10 Both Ψ_{tlp} and π_0 are considered reliable indicators of drought tolerance (Bartlett et al., 2012).
11 In fact, our data show that Ψ_{tlp} sets the limit that can be reached by Ψ_{pd} and Ψ_{min} .
12 Progressively more negative Ψ_{tlp} allowed some species to reach and tolerate more negative
13 Ψ_{pd} and Ψ_{min} , thus extending the time interval for maintenance of stomatal aperture,
14 photosynthetic carbon gain, and growth (Sack and Holbrook, 2006; Lenz et al., 2006). The
15 highly significant positive correlation between Ψ_{tlp} or π_0 and g_L further points to symplastic
16 drought resistance as a good predictor of plant water use over green roofs. In fact, low g_L
17 values displayed by species with low Ψ_{tlp} translates into low evapotranspiration rates and a
18 more conservative water use, which represents a desirable feature of plants selected for green
19 roofs to be installed in drought-prone regions (Savi et al., 2015). Similarly, low water use
20 under drought conditions has been recently reported for granite outcrop shrubs capable to
21 tolerate substantial Ψ_{leaf} drop under drought (Farrell et al., 2013).

22 Plants with more negative π_0 or Ψ_{tlp} also displayed significantly lower growth rates in both 10
23 and 13 cm deep modules (Table 4, S2). Low growth rates in these species might arise as a
24 consequence of both limited g_L and reduced carbon gain, and osmoregulation processes

1 involving substantial carbon investment. The reduction of π_0 , driven by active accumulation
2 of compatible solutes in cells, protects membranes during stress and preserves metabolic
3 functionality, but requires high energetic costs (Lenz et al., 2006; Dichio et al., 2009; Bartlett
4 et al., 2012) at the expense of plant growth. In any case, low growth rates translate into the
5 development of small-sized vegetation, representing a desirable characteristic for extensive
6 green roofs due to associated reduction of installation load and maintenance costs (Caneva et
7 al., 2015; Berardi et al., 2014; Savi et al., 2014).

8 An overall more favorable water status (albeit only marginally significant, $P=0.12$) was
9 recorded in plants growing on D-10 than on D-13 modules. As an example, Ψ_{pd} measured for
10 *P. spina-christi* in both June and August was significantly higher in D-10 than in D-13
11 modules. In a recent experiment by some of us, it was shown that reduced substrate depth
12 may translate into less severe plant water stress, as a likely consequence of reduced plant
13 biomass, coupled to faster recovery of hydration of substrate and water retention layer during
14 rainfalls (Savi et al., 2015). The results of the present experiment support these conclusions,
15 as shrubs growing on 13 cm deep substrate showed an overall tendency to grow faster when
16 compared to the individuals growing on 10 cm, and also displayed lower water potentials.

17 Even if the water status of plants grown on D-10 modules was more favorable, the recorded
18 mortality rate exceeded 40% in these modules, while it was less than 20% in D-13 modules.
19 In fact, for *E. majus* 73% of the plants established on shallow substrate died, while a 100%
20 survival rate of the same species was observed in deeper substrate. Moreover, an overall high
21 M (62.5 %) was observed for *P. lentiscus*, despite the high symplastic resistance to drought of
22 this species (low Ψ_{tip} and π_0). **This result suggests that a difference of only 3 cm in substrate
23 depth can produce significant effects on plant performance.** Our data are consistent with
24 recent studies, reporting improved plant survival in green roof installations with deep

1 substrates than in shallower ones (Dunnett et al., 2008; Razzaghmanesh et al., 2014; Zhang et
2 al., 2014). However, our mortality data, coupled to measurements of plant water status and
3 analysis of functional traits related to species-specific drought resistance, suggest that water
4 stress is not the only, and probably not the major cause of plant failure on Mediterranean
5 green roofs.

6 Xylem hydraulic vulnerability as estimated in terms of P50 was correlated with Ψ_{pd} and g_L
7 measured in June in the shallow modules (D-10). This result indeed suggests that high
8 resistance to stem hydraulic dysfunction (more negative P50) may allow plants to tolerate
9 lower Ψ_{leaf} while maintaining positive safety margins (calculated as P50–seasonal minimum
10 Ψ_{leaf}) towards massive embolism formation (Choat et al., 2012; Nardini et al., 2014a). The
11 reduced Ψ_{leaf} enhances the driving force for the water movement in the root-to-leaf pathway,
12 enabling the plant to absorb water at lower $\Psi_{substrate}$. A very interesting result was the lack of
13 correlation between P50 and M in D-13 modules, while such relationship was highly
14 significant in shallow modules ($P < 0.01$). In particular, the highest mortality was observed for
15 species characterized by low P50 values, i.e. *P. lentiscus* (P50=-1.55 MPa) and *P. pyraster*
16 (P50=-1.70 MPa). This is in accordance with recent studies reporting correlations between
17 tree die-back and species-specific P50 in natural habitats characterized by extremely shallow
18 limestone soils (Nardini et al., 2012). On the other hand, the lowest M was recorded for *C.*
19 *cogygria* (P50=-3.88 MPa), known to be a drought resistant species colonizing limestone
20 cliffs and degraded areas (Pignatti, 2002). More than 50% of the tested species showed almost
21 complete survival on D-13 modules, suggesting that just 3 cm of deeper substrate might
22 significantly enhance the chances of plant survival. Aside from P50, however, no significant
23 correlations were found between M and other physiological traits related to drought
24 resistance. The trend towards improved plant growth/survival on deeper substrates has been

1 related to the higher volume of available water to vegetation, or to the mitigation of
2 temperature extremes ensured by deep substrates compared to shallow ones (Dunnett et al.,
3 2008; Price et al., 2011; Razzaghmanesh et al., 2014). Surprisingly enough, to the best of our
4 knowledge, a clear demonstration of the relative importance of drought versus heat stress in
5 driving plant mortality over green roofs is still lacking.

6 In our study, the 3 cm difference in substrate depth translated into an increase of saturated
7 water content by 30% in D-13 versus D-10. However, as discussed above, plant water status
8 was overall more favorable in D-10 than in D-13. On the other hand, we observed that both
9 minimum and maximum temperatures, as well as daily thermal regime recorded at the
10 maximum substrate depths, were significantly different in D-10 and D-13 modules. The 25%
11 deeper substrate led to a 4.4°C difference in the absolute temperature peak reached during
12 summer. In particular, the temperature in D-10 modules frequently exceeded 42°C, while it
13 was constantly below such critical threshold in modules that were just 3 cm deeper. The
14 temperatures recorded in our study are in accordance with those reported for a 15 cm deep
15 green roof established in Mediterranean climate (Olivieri et al., 2013) and slightly higher (by
16 about 3°C) of those measured under 10 cm deep substrate layer under subtropical climate
17 conditions (Simmons et al., 2008). On the basis of the maximum temperature peak reached in
18 D-10 modules, the species-specific root vulnerability to heat stress (ΔREL) was estimated
19 after a 45°C treatment. Interestingly ΔREL was correlated to plant mortality in both D-10 and
20 D-13 modules, thus suggesting that high substrate temperature represents a stress factor
21 affecting plant survival on green roofs to a larger extent than drought per se (Fig. 3; S2). In
22 fact, several authors have reported that both chronic and abrupt heat stress can reduce root
23 growth and limit nutrient and water uptake, since roots are often more sensitive to heat stress
24 than shoots, Huang et al., 2012). High temperatures at the root level may adversely affect

1 respiration and cell membrane stability, as well as modulate levels of hormones and primary
2 and secondary metabolites, with a consequent effect on root-to-shoot signaling (Kuroyanagi
3 and Paulsen, 1988; Wahid et al., 2007; Huang et al., 2012). Moreover, the effects of high
4 temperature and water deficit stress, both of which characterize green roof ecosystems, are
5 globally additive (Vile et al., 2012) and their combined effect is known to be even more
6 deleterious for plant life in both natural and semi-natural ecosystems (Allen et al., 2010; Price
7 et al., 2011; Nardini et al., 2013).

8

9 **5. CONCLUSION**

10 Our data highlight the importance of plant physiological traits conferring resistance against
11 both drought and high substrate temperatures as proxies to be taken into account when
12 selecting species for roof greening in the Mediterranean-climate regions. In fact, drought-
13 tolerant species had also lower water needs and growth rates, while the ability to survive in
14 harsh microclimate conditions was significantly correlated to the resistance of the root system
15 to heat stress. It has been demonstrated that reducing soil temperature while maintaining air
16 temperature relatively high improve the growth and the functional status of both roots and
17 shoots, ensuring plant survival (Kuroyanagi and Paulsen, 1988; Price et al., 2011; Huang et
18 al., 2012). One of the main targets in green roof research is reducing substrate depth, to limit
19 installation weight and costs (Cao et al., 2014). However, our results show that such a
20 strategy might contrast with the need to minimize temperature extremes in the substrate and
21 assure plant survival. Future experiments should test possible solutions to increase albedo on
22 green roof systems with shallow substrates. In this light, the optimal design for green roofs in
23 arid-prone areas should include a carefully selected drought resistant vegetation, able to save
24 water and tolerate extreme below-ground temperatures.

1

2 **FUNDING INFORMATION**

3 The present study was supported by the Fondo Europeo di Sviluppo Regionale POR FESR n.
4 54/2009/C. Love V.L. was supported by EU and Regione Friuli Venezia Giulia (Fondo
5 Sociale Europeo, Programma Operativo Regionale 2007-2013) in the frame of the project
6 S.H.A.R.M. (Supporting Human Assets of Research and Mobility).

7

8 **ACKNOWLEDGEMENTS**

9 Plant material was kindly provided by Servizio gestione forestale e produzione legnosa,
10 Direzione centrale risorse rurali, agroalimentari e forestali, Friuli Venezia Giulia (Vivaio
11 Pascual Tarcento). We are grateful to Luca Grizzo, Roberto Alberti and Sabrina Grizon for
12 technical assistance during hydraulic analysis.

13

14 **CONFLICT OF INTEREST**

15 We declare that there are no conflicts of interest including any financial, personal or other
16 relationships with other people or organizations that could influence this work.

17

18 **REFERENCES**

19 **Allen CD, Macalady AK, Chenchouni H, et al. 2010.** A global overview of drought and
20 heat-induced tree mortality reveals emerging climate change risks for forests. *Forest Ecology*
21 *and Management* **259**: 660–684.

- 1 **Bartlett MK, Scoffoni C, Sack L. 2012.** The determinants of leaf turgor loss point and
2 prediction of drought tolerance of species and biomes: A global meta-analysis. *Ecology*
3 *Letters* **15**: 393–405.
- 4 **Benvenuti S, Bacci D. 2010.** Initial agronomic performances of Mediterranean xerophytes in
5 simulated dry green roofs. *Urban Ecosystems* **13**: 349–363.
- 6 **Berardi U, GhaffarianHoseini A, GhaffarianHoseini A. 2014.** State-of-the-art analysis of
7 the environmental benefits of green roofs. *Applied Energy* **115**: 411–428.
- 8 **Bernstein L, Nieman RH. 1960.** Apparent Free Space of Plant Roots. *Plant Physiology* **35**:
9 589–598.
- 10 **Caneva G, Kumbaric A, Savo V, Casalini R. 2015.** Ecological approach in selecting
11 extensive green roof plants: a data-set of Mediterranean plants. *Plant Biosystems* **149**: 374-
12 383.
- 13 **Cao CTN, Farrell C, Kristiansen PE, Rayner JP. 2014.** Biochar makes green roof
14 substrates lighter and improves water supply to plants. *Ecological Engineering* **71**: 368–374.
- 15 **Castleton HF, Stovin V, Beck SBM, Davison JB. 2010.** Green roofs; building energy
16 savings and the potential for retrofit. *Energy and Buildings* **42**: 1582–1591.
- 17 **Choat B, Jansen S, Brodribb TJ, et al. 2012.** Global convergence in the vulnerability of
18 forests to drought. *Nature* **491**: 752–5.
- 19 **Cochard H, Badel E, Herbette S, Delzon S, Choat B, Jansen S. 2013.** Methods for
20 measuring plant vulnerability to cavitation: a critical review. *Journal of Experimental Botany*
21 **64**: 4779–4791.
- 22 **Conti F, Abbate G, Alessandrini A, Blasi C. 2005.** An annotated checklist of the Italian

- 1 vascular flora. Roma: Palombi.
- 2 **Dichio B, Margiotta G, Xiloyannis C, Bufo SA, Sofo A, Cataldi TRI. 2009.** Changes in
3 water status and osmolyte contents in leaves and roots of olive plants (*Olea europaea* L.)
4 subjected to water deficit. *Trees* **23**: 247–256.
- 5 **Dunnett N, Nagase A, Hallam A. 2008.** The dynamics of planted and colonising species on a
6 green roof over six growing seasons 2001-2006: Influence of substrate depth. *Urban*
7 *Ecosystems* **11**: 373–384.
- 8 **Ennajeh M, Nouri M, Khemira H, Cochard H. 2011.** Improvement to the air-injection
9 technique to estimate xylem vulnerability to cavitation. *Trees* **25**: 705–710.
- 10 **Farrell C, Szota C, Williams NSG, Arndt SK. 2013.** High water users can be drought
11 tolerant: Using physiological traits for green roof plant selection. *Plant and Soil* **372**: 177–
12 193.
- 13 **Galmés J, Flexas J, Savé R, Medrano H. 2007.** Water relations and stomatal characteristics
14 of Mediterranean plants with different growth forms and leaf habits: responses to water stress
15 and recovery. *Plant and Soil* **290**: 139–155.
- 16 **Huang B, Rachmilevitch S, Xu J. 2012.** Root carbon and protein metabolism associated
17 with heat tolerance. *Journal of Experimental Botany* **63**: 3455–3465.
- 18 **Jacobsen AL, Pratt RB, Tobin MF, Hacke UG, Ewers FW. 2012.** A global analysis of
19 xylem vessel length in woody plants. *American Journal of Botany* **99**: 1583–1591.
- 20 **Kuroyanagi T, Paulsen GM. 1988.** Mediation of high-temperature injury by roots and
21 shoots during reproductive growth of wheat. *Plant, Cell and Environment* **11**: 517-523.
- 22 **Lenz TI, Wright IJ, Westoby M. 2006.** Interrelations among pressure-volume curve traits

- 1 across species and water availability gradients. *Physiologia Plantarum* **127**: 423–433.
- 2 **Lundholm JT. 2006.** Green roofs and facades: a habitat template approach. *Urban Habitats*
3 **4**: 87–101.
- 4 **Madre F, Vergnes A, Machon N, Clergeau P. 2014.** Green roofs as habitats for wild plant
5 species in urban landscapes: First insights from a large-scale sampling. *Landscape and Urban*
6 *Planning* **122**: 100–107.
- 7 **Van Mechelen C, Dutoit T, Kattge J, Hermy M. 2014.** Plant trait analysis delivers an
8 extensive list of potential green roof species for Mediterranean France. *Ecological*
9 *Engineering* **67**: 48–59.
- 10 **Nardini A, Battistuzzo M, Savi T. 2013.** Shoot desiccation and hydraulic failure in
11 temperate woody angiosperms during an extreme summer drought. *New Phytologist* **200**:
12 322–329.
- 13 **Nardini A, Lo Gullo MA, Trifilò P, Salleo S. 2014a.** The challenge of the Mediterranean
14 climate to plant hydraulics: Responses and adaptations. *Environmental and Experimental*
15 *Botany* **103**: 68–79.
- 16 **Nardini A, Savi T, Novak M. 2014b.** Droughts, Heat Waves and Plant Hydraulics: Impacts
17 and Legacies. *Agrochimica* **58**: 146–161.
- 18 **Olivieri F, Di Perna C, D’Orazio M, Olivieri L, Neila J. 2013.** Experimental measurements
19 and numerical model for the summer performance assessment of extensive green roofs in a
20 Mediterranean coastal climate. *Energy and Buildings* **63**: 1–14.
- 21 **Pignatti S. 2002.** *Flora d’Italia*. Bologna: Edagricole.
- 22 **Price JG, Watts SA, Wright AN, Peters RW, Kirby JT. 2011.** Irrigation lowers substrate

1 temperature and enhances survival of plants on green roofs in the southeastern United States.
2 HortTechnology **21**: 586–592.

3 **Raimondo F, Trifilò P, Lo Gullo MA, Andri S, Savi T, Nardini A. 2015.** Plant
4 performance on Mediterranean green roofs: interaction of species-specific hydraulic strategies
5 and substrate water relations. AoB plants **7**: plv007.

6 **Rayner JP, Farrell C, Raynor KJ, Murphy SM, Williams NSG. 2015.** Plant establishment
7 on a green roof under extreme hot and dry conditions: The importance of leaf succulence in
8 plant selection. Urban Forestry and Urban Greening **15**: 6–14.

9 **Razzaghamanesh M, Beecham S, Kazemi F. 2014.** The growth and survival of plants in
10 urban green roofs in a dry climate. Science of the Total Environment **476-477**: 288–297.

11 **Sack L, Holbrook NM. 2006.** Leaf Hydraulics. Annual Review of Plant Biology **57**: 361–
12 381.

13 **Savi T, Andri S, Nardini A. 2013.** Impact of different green roof layering on plant water
14 status and drought survival. Ecological Engineering **57**: 188–196.

15 **Savi T, Marin M, Boldrin D, Incerti G, Andri S, Nardini A. 2014.** Green roofs for a drier
16 world: Effects of hydrogel amendment on substrate and plant water status. Science of the
17 Total Environment **490**: 467–476.

18 **Savi T, Boldrin D, Marin M, et al. 2015.** Does shallow substrate improve water status of
19 plants growing on green roofs ? Testing the paradox in two sub-Mediterranean shrubs.
20 Ecological Engineering **84**: 292–300.

21 **Simmons MT, Gardiner B, Windhager S, Tinsley J. 2008.** Green roofs are not created
22 equal: the hydrologic and thermal performance of six different extensive green roofs and

- 1 reflective and non-reflective roofs in a sub-tropical climate. *Urban Ecosystems* **11**: 339–348.
- 2 **Thuring C, Grant G. 2015.** The biodiversity of temperate extensive green roofs – a review
3 of research and practice. *Israel Journal of Ecology and Evolution*: 1–14.
- 4 **Tyree MT, Hammel HT. 1972.** The measurement of the turgor pressure and the water
5 relations of plants by the pressure-bomb technique. *Journal of Experimental Botany* **23**: 267–
6 282.
- 7 **Tyree MT, Sperry JS. 1989.** Cavitation and Embolism. *Ann. Rev. Plant. Phys. Mol. Bio.* **40**:
8 19–38.
- 9 **Vijayaraghavan K, Raja FD. 2014.** Design and development of green roof substrate
10 to improve runoff water quality: Plant growth experiments and adsorption. *Water Research*
11 **63**: 94–101.
- 12 **Vile D, Pervent M, Belluau M, et al. 2012.** Arabidopsis growth under prolonged high
13 temperature and water deficit: Independent or interactive effects? *Plant, Cell and*
14 *Environment* **35**: 702–718.
- 15 **Wahid A, Gelani S, Ashraf M, Foolad M. 2007.** Heat tolerance in plants: An overview.
16 *Environmental and Experimental Botany* **61**: 199–223.
- 17 **Zhang H, Lu S, Wu J, Jiang Y, Lu Y, Zhao H. 2014.** Effect of substrate depth on 18 non-
18 succulent herbaceous perennials for extensive green roofs in a region with a dry spring.
19 *Ecological Engineering* **71**: 490–500.

20 LEGENDS

21 **Table 1.** List of the 11 study species and relative family, growth form (Ch, chamaephyte,
22 partially woody shrubs; NP, nano-phanerophyte, low woody shrubs; P, phanerophyte, small

1 trees with shrub-like growth habit), leaf habit (E, evergreen; D, deciduous), habitat and
2 altitude range. Maquis, evergreen sclerophyllous shrubland; garigue, low, soft-leaved
3 scrubland.

4

5 **Table 2.** Leaf water potential at turgor loss point (Ψ_{tp} , MPa), osmotic potential at full turgor
6 (π_0 , MPa), and water potential inducing 50% loss of stem hydraulic conductivity (P50, MPa)
7 of the 11 Mediterranean and sub-Mediterranean woody species. The relative diameter
8 increment (G, %) as estimated 2 years after planting in 10 cm (D-10) and 13 cm (D-13) thick
9 experimental modules is also reported.

10

11 **Table 3.** Pre-dawn (Ψ_{pd}) and minimum (Ψ_{min}) leaf water potential (MPa), and leaf
12 conductance to water vapor (g_L , $\text{mmol m}^{-2} \text{s}^{-1}$) as recorded for the 11 study species in 10 cm
13 (D-10) and 13 cm (D-13) experimental modules in June (high water availability) and in
14 August (limited water availability) 2014.

15

16 **Table 4.** Correlation matrices reporting the coefficient r and P value (as asterisks, Pearson
17 product moment correlation) for correlations between pairs of traits: water potential at turgor
18 loss point (Ψ_{tp}), osmotic potential at full turgor (π_0), water potential inducing 50% loss of
19 stem hydraulic conductivity (P50), plant mortality (M), relative diameter increment (G), pre-
20 dawn and minimum water potentials (Ψ_{pd} , Ψ_{min}), and leaf conductance to water vapor (g_L), as
21 measured in 10 and 13 cm deep green roof modules. *, $P \leq 0.05$; **, $P \leq 0.01$; ***, $P \leq 0.001$.

22

1 **Fig. 1.** Plant mortality (M, %) of the 11 study species growing in 10 cm (D-10, black
2 columns) and 13 cm (D-13, gray columns) deep green roof modules. The average plant
3 mortality calculated for 10 or 13 cm thick substrate (n=11) is also reported. * indicates
4 statistically significant difference between experimental categories (Student's t-test, P<0.05).

5

6 **Fig. 2.** A) Temperature course (°C) recorded at the maximum substrate depth in 10 cm (D-10,
7 closed circles) and 13 cm (D-13, open circles) on a representative warm summer day. The
8 average thermal regime of the substrate in July (the hottest month) is also reported. B) The
9 absolute maximum substrate temperature reached during the study period in D-10 (black
10 columns) and D-13 (gray columns). * indicates statistically significant difference between
11 experimental categories (Student's t-test, P<0.05).

12

13 **Fig. 3.** Relationship between root vulnerability to heat stress (Δ REL, %) and plant mortality
14 (M, %) as measured in September 2015 in 10 cm (D-10, closed circles) and 13 cm (D-13,
15 open circles) experimental modules. The correlation coefficient r and P value (Pearson
16 product moment correlation) are reported.

17

18

19

20

21

1

Species	Family	Growth form	Leaf habit	Habitat and altitude range
<i>Cistus salviifolius</i> L.	Cistaceae	NP	E	<i>Maquis</i> and oak forests, on acid or limestone soil; 0-1200 m
<i>Cotinus coggygria</i> Scop.	Anacardiaceae	NP	D	Brush and cliffs, on limestone soil; 0-900 m
<i>Emerus majus</i> Mill.	Fabaceae	NP	D	Brush and woodlands; 0-1650 m
<i>Ligustrum vulgare</i> L.	Oleaceae	NP	E	Termophilic deciduous forest margins and brush; 0-1300 m
<i>Paliurus spina-christi</i> Mill.	Rhamnaceae	P	D	Woodland and bush, on steep slopes; 0-500 m
<i>Phillyrea angustifolia</i> L.	Oleaceae	P	E	Hot and very dry <i>maquis</i> and <i>garigues</i> ; 0-600 m
<i>Pistacia lentiscus</i> L.	Anacardiaceae	P	E	<i>Maquis</i> developed on the coast; 0-700 m
<i>Prunus mahaleb</i> L.	Rosaceae	P	D	Thermophilous submediterranean forests; 0-1900 m
<i>Prunus spinosa</i> L.	Rosaceae	P	D	Deciduous forests, brush, hedges; 0-1500 m
<i>Pyrus pyraster</i> Burgsd.	Rosaceae	P	D	Deciduous forests, on nutrient-rich soil; 0-1400 m
<i>Salvia officinalis</i> L.	Lamiaceae	Ch	E	<i>Garigue</i> , rocky meadows arid cliffs; on limestone soil; 0-300 m
<i>Spartium junceum</i> L.	Fabaceae	P	D	Sun exposed brush; 0-2000 m

2

3

4 **Table 1**

5

6

7

8

9

10

11

1

2

Species	Ψ_{lp}	π_0	P50	Growth, %	
	-MPa	-MPa	-MPa	D-10	D-13
<i>C. salviifolius</i>	1.64 ± 0.14	1.28 ± 0.05	4.40	59.3	128.5
<i>C. coggygria</i>	1.89 ± 0.22	1.32 ± 0.18	3.9	81.1	87.0
<i>E. majus</i>	1.90 ± 0.17	1.44 ± 0.17	2.76	47.8	103.4
<i>L. vulgare</i>	1.75 ± 0.12	1.15 ± 0.09	5.00	74.6	106.1
<i>P. spina-christi</i>	2.02 ± 0.1	1.51 ± 0.03	2.13	30.4	34.9
<i>P. angustifolia</i>	2.49 ± 0.02	1.78 ± 0.16	2.7	41.3	25.0
<i>P. lentiscus</i>	2.69 ± 0.15	2.23 ± 0.08	1.6	0.0	15.9
<i>P. mahaleb</i>	2.15 ± 0.12	1.55 ± 0.14	5.0	34.4	48.5
<i>P. pyrastrer</i>	2.32 ± 0.29	1.68 ± 0.28	1.7	x	x
<i>S. officinalis</i>	1.26 ± 0.04	1.03 ± 0.02	2.51	122.2	72.9
<i>S. junceum</i>	1.02 ± 0.16	0.69 ± 0.14	3.66	202.6	219.1

3

4

5 Table 2

6

7

8

9

10

11

12

13

14

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15

Species	Ψ_{pd} , -MPa				Ψ_{min} , -MPa				gL, mmol m ⁻² s ⁻¹			
	June		August		June		August		June		August	
	D-10	D-13	D-10	D-13	D-10	D-13	D-10	D-13	D-10	D-13	D-10	D-13
<i>C. salviifolius</i>	0.57 ±0.11	0.56 ±0.24	1.19 ±0.39	1.33 ±0.08	1.20 ±0.11	1.35 ±0.09	2.03 ±0.33	2.43 ±0.12	527.9 ±155.2	493.0 ±58.1	151.6 ±28.4	210.4 ±76.1
<i>C. coggygria</i>	0.20 ±0.02	0.15 ±0.02	1.29 ±0.12	1.06 ±0.03	1.10 ±0.07	1.13 ±0.06	2.17 ±0.11	2.24 ±0.05	425.6 ±16.6	466.4 ±21.9	203.5 ±22.5	216.7 ±38.0
<i>E. majus</i>	0.80 ±0.14	0.61 ±0.09	0.59 ±0.04	1.43 ±0.53	1.30 ±0.02	1.55 ±0.18	1.25 ±0.15	2.57 ±0.39	81.3 ±9.5	339.9 ±75.8	157.3 ±42.9	182.2 ±120.4
<i>L. vulgare</i>	0.56 ±0.05	0.78 ±0.22	0.65 ±0.05	1.84 ±0.64	1.32 ±0.09	1.28 ±0.21	1.83 ±0.53	2.76 ±0.28	338.7 ±110.3	226.8 ±32.8	325.7 ±82.7	168.1 ±132.7
<i>P. spina-chris</i>	0.88 ±0.05	1.14 ±0.1	1.34 ±0.07	1.84 ±0.02	1.30 ±0.12	1.42 ±0.12	2.57 ±0.29	2.99 ±0.34	189.2 ±25.7	340.3 ±107.3	242.0 ±104.4	228.9 ±102.2
<i>P. angustifolia</i>	0.88 ±0.31	1.05 ±0.05	2.80 ±0.8	2.12 ±1.2	1.13 ±0.3	2.03 ±0.37	4.20 ±0.75	3.62 ±1.53	164.7 ±41.3	111.7 ±12.8	108.8 ±41.4	176.8 ±25.0
<i>P. lentiscus</i>	1.30 ±0.02	1.44 ±0.07	1.98 ±0.08	1.75 ±0.65	2.20 ±0.02	2.34 ±0.29	3.71 ±0.36	3.37 ±0.31	95.5 ±15.6	231.5 ±54.8	66.4 ±26.8	154.5 ±60.0
<i>P. mahaleb</i>	0.54 ±0.1	0.58 ±0.12	0.97 ±0.05	1.25 ±0.03	1.20 ±0.2	1.34 ±0.25	2.06 ±0.11	2.29 ±0.07	435.8 ±10.5	435.9 ±24.9	212.8 ±40.8	212.8 ±49.8
<i>S. officinalis</i>	0.73 ±0.05	0.64 ±0.06	0.74 ±0.04	0.80 ±0.02	1.06 ±0.14	0.86 ±0.05	1.68 ±0.12	1.85 ±0.7	468.5 ±183.2	475.9 ±133.5	389.9 ±68.4	468.0 ±151.6
<i>S. junceum</i>	0.27 ±0.09	0.25 ±0.03	0.71 ±0.21	0.59 ±0.24	0.54 ±0.07	0.60 ±0.03	1.23 ±0.26	2.36 ±0.19	x	x	x	x

Table 3

1

2

3

(A)	Ψ_{pd}				Ψ_{min}				gL			
	June		August		June		August		June		August	
	D-10	D-13	D-10	D-13	D-10	D-13	D-10	D-13	D-10	D-13	D-10	D-13
Ψ_{tlp}	0.67 *	0.71 *	0.73 *	-0.15	0.76 **	0.94 ***	0.82 **	0.76 **	0.67 *	0.67 *	0.8 **	0.7 *
π_0	0.78 **	0.77 **	0.7 *	-0.067	0.86 **	0.96 ***	0.81 **	0.72 *	0.67 *	0.56	0.82 **	0.56
P50	0.73 *	0.61	0.43	-0.12	0.42	0.45	0.45	0.43	0.67 *	0.31	0.28	-0.17
M-10	-0.7 *	-0.57	0.14	-0.47	-0.57	-0.33	0.013	-0.24	-0.64	-0.21	-0.11	0
M-13	-0.2	-0.06	0.08	-0.15	-0.24	0.076	-0.009	0.4	0.33	0.4	0.22	0.64
G-10	0.67 *	x	0.48	x	0.81 **	x	0.61	x	0.64	x	0.79 **	x
G-13	x	0.73 *	x	-0.11	x	0.74 **	x	0.51	x	0.5	x	0.08
(B)	Growth		Mortality									
	D-10	D-13	D-10	D-13								
Ψ_{tlp}	0.89 ***	0.83 **	-0.22	-0.12								
π_0	0.89 ***	0.84 **	-0.35	-0.2								
P50	0.24	0.43	-0.73 **	-0.42								

4

5

6 Table 4

7

8

9

10

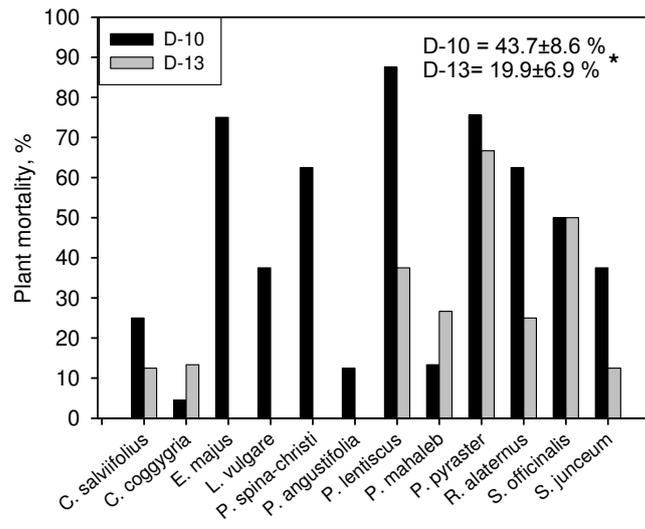
11

12

1

2

3



4

5 Fig. 1

6

7

8

9

10

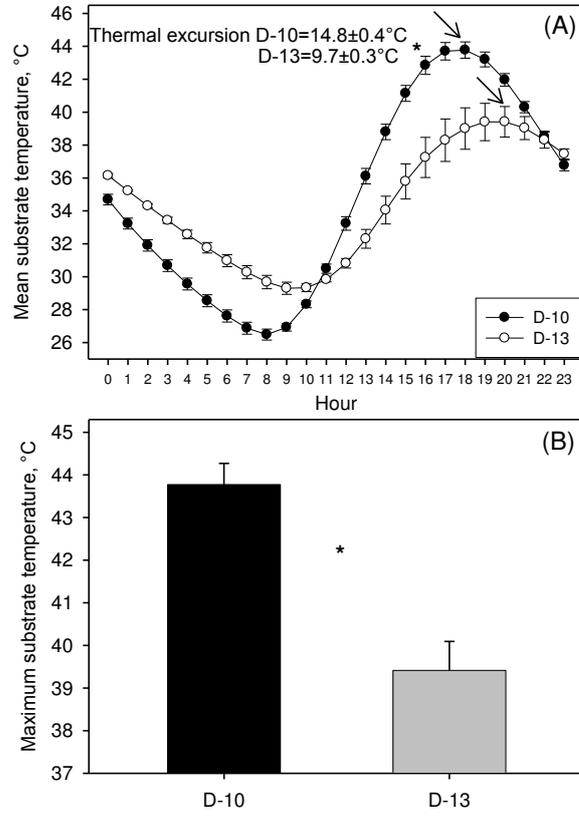
11

12

13

1

2



3

4 Fig. 2

5

6

7

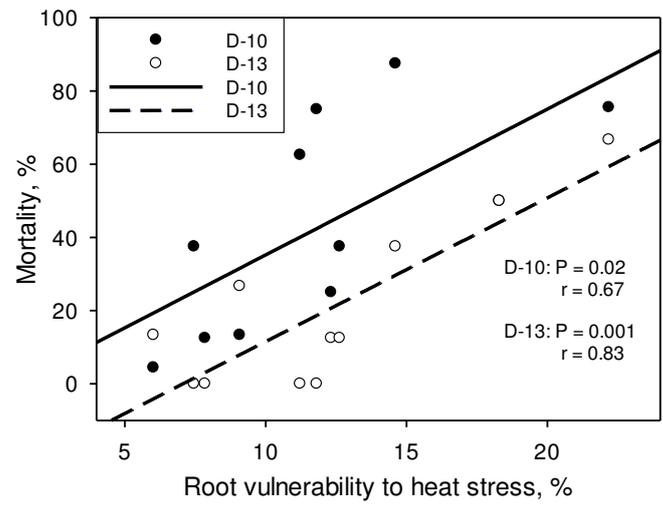
8

9

1

2

3



4

5 **Fig. 3**

6

7

8