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Low but contrasting neutral genetic differentiation shaped by winter temperature in European great tits

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1	Low but contrasting neutral genetic differentiation shaped by winter temperature in
2	European great tits
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65 Population genetic structure

70 Abstract

Gene flow is usually thought to reduce genetic divergence and impede local adaptation by homogenising gene pools between populations. However, evidence for local adaptation and phenotypic differentiation in highly mobile species, experiencing high levels of gene flow, is emerging. Assessing population genetic structure at different spatial scales is thus a crucial step towards understanding mechanisms underlying intraspecific differentiation and diversification. Here, we studied the population genetic structure of a highly mobile species – the great tit Parus major – at different spatial scales. We analysed 884 individuals from 30 sites across Europe including 10 close-by sites (< 50 km), using 22 microsatellite markers. Overall we found a low but significant genetic differentiation among sites ($F_{ST} = 0.008$). Genetic differentiation was higher, and genetic diversity lower, in south-western Europe. These regional differences were statistically best explained by winter temperature. Overall, our results suggest that great tits form a single patchy metapopulation across Europe, in which genetic differentiation is independent of geographical distance and gene flow may be regulated by environmental factors via movements related to winter severity. This might have important implications for the evolutionary trajectories of sub-populations, especially in the context of climate change, and calls for future investigations of local differences in costs and benefits of philopatry at large scales.

91 Introduction

Gene flow is generally thought to impede local adaptation by introducing locally maladapted genotypes into populations exchanging individuals. Consequently, micro-evolutionary processes at small scales are predicted to be rare in highly mobile organisms with high gene flow over large spatial scales, due to spatial genetic homogenisation. However, evidence for genetic differentiation and local adaptation at small scales despite high levels of gene flow at large scales has recently started to accumulate in different taxa (e.g. mammals: Musiani et al., 2007; marine invertebrates: Sanford & Kelly, 2011; birds: Charmantier et al., 2015; fish: Junge et al., 2011; trees: Savolainen, Pyhäjärvi & Knürr, 2007). This evidence suggests that dispersal is not a diffusion-like movement process and that ultimately gene flow may vary in space.

Spatial variation in gene flow is probably common, especially in relation to environmental factors in highly mobile species. High mobility and long distance dispersal facilitate spatial spread and the colonization of new habitats (Nathan et al., 2003). As a consequence, highly mobile species are likely to experience a large set of environmental conditions that may shape locally adaptive processes. In addition, high mobility combined with the ability to cross physical barriers such as seas or mountains may minimize the influence of geographical factors. Increased mobility may also reduce the impact of historical factors on gene flow by homogenising gene pools, increasing local population size and counteracting genetic drift (Slatkin, 1987). In this case, environmental factors may become the main force shaping gene flow (e.g. Pilot et al., 2006). Assessing gene flow between populations at small and large spatial scales in highly mobile species and the links between gene flow and environmental factors is crucial to understand the ecological mechanisms leading to intraspecific differentiation and diversification. When dispersal movements and immigration rate do not provide reliable estimates of gene flow, such as in highly mobile

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species, a population genetic approach may help investigating patterns of gene flow atdifferent spatial scales (Nathan et al., 2003).

The great tit Parus major, a widespread passerine bird across Eurasia (Snow & Perrins, 1998), is a particularly interesting biological model to address such questions. This species is considered to be an "evolutionary winner", given its ability to colonize and rapidly adapt to new habitats. Its rapid spread across Europe since the last glaciation period (Kvist et al., 2003; Pavlova et al., 2006) suggests high dispersal ability and gene flow among sub-populations (Caswell, Lensink & Neubert, 2003; Pilot et al., 2006 but see Peterson & Denno, 1998). Conversely, long-term monitoring studies provide evidence for small-scale local adaptation (Garant et al., 2005; Postma & van Noordwijk, 2005) with a considerable fraction of individuals dispersing over short distances (e.g. Verhulst, Perrins & Riddington, 1997). Thus, although great tits are considered highly mobile and forming a homogeneous population across Europe, microevolutionary processes linked with limited gene flow occur at small scales and with it, the detection of subtle fine-scale genetic structures (Björklund, Ruiz & Senar, 2010; Garroway et al., 2013; Van Bers et al., 2012). These conflicting observations call for investigating genetic differentiation using microsatellite markers at different spatial scales in this species. Indeed microsatellite markers generating multi-locus diploid genotypes provide an ideal resolution to study recent or ongoing micro-evolutionary processes occurring both at small and large scales (e.g. Wang, 2010).

Moreover, the environmental heterogeneity over the species' range combined with its colonisation history provides excellent conditions to study the influence of environmental factors on population genetic structure in this species. Indeed, phylogeographic studies based on mitochondrial DNA (mtDNA) suggest that other tit species colonized Europe from different glacial refugia, each harbouring distinct mitochondrial lineages and forming secondary contact zones within Europe (Kvist et al., 2004; Päckert, Martens & Sun, 2010; Pentzold et al., 2013). In contrast, all western-European great tits share a common haplotype, suggesting that they originate from a single glacial refugium located in southern Europe (Kvist et al., 2007; Kvist et al., 1999; Pavlova et al., 2006, Fig. S1, Table S1). Therefore genetic differentiation in great tits estimated with microsatellites that evolve faster than mtDNA and are more powerful to detect recent and local micro-evolutionary processes among populations, are less likely to result from past genetic discontinuities across different glacial refugia as is the case for many other species (e.g. Hewitt, 2000; Kvist et al., 1999).

Using 22 microsatellite markers, we investigated population genetic diversity and structure, as well as the scale of genetic differentiation, in great tits by sampling 30 sites across Europe including 10 close-by (i.e. up to 50 km) sites. We expected the genetic differentiation to be correlated with the geographical distance either at small or large scales: the studied geographical scale should allow us to determine at which scale isolation-by-distance would occur in great tits. In addition, a signal of historical range expansion from the South to the North should result in decreased genetic diversity with increasing latitude. In a second step, we explored the influence of environmental factors on the observed genetic diversity and differentiation patterns, focusing on factors that can be expected to affect individual movement. In particular, temperature may strongly shape genetic differentiation among populations by acting on both dispersal movements (e.g. Parn et al., 2012) and establishment success (i.e. survival and reproductive success after settlement) of long-distance immigrants (e.g. Van Doorslaer et al., 2009). Three different patterns may thus be predicted in relation to temperature. First, because temperature can be positively correlated with survival and population density (Ahola et al., 2009; Garant et al., 2004; Parn et al., 2012) that increase dispersal propensities (Forsman & Monkkonen, 2003; Matthysen, 2005), genetic diversity could increase and genetic differentiation decrease with increasing temperature. Second, a negative relationship between temperature and dispersal propensities may be expected in the case of partial migration (e.g. Nilsson et al., 2006). In this case, temperature should relate to

environmental conditions during winter, triggering partial migration and favouring dispersal in general or the establishment of migrants in non-natal breeding areas. Genetic diversity should consequently decrease while genetic differentiation should increase with temperature (e.g. Miller et al., 2012). Third, if the establishment success of immigrants is linked to adaptation to temperature, we predicted that genetic differentiation should increase with the difference of temperature between sites.

Material and Methods

Species description, sampling and genotyping

The great tit is a hole-nesting passerine bird that readily breeds in nest boxes, providing easy access to breeding pairs. In this study, all individuals from all but one site (FI.TU, see Table S2) were breeding adults caught in nest boxes during the nestling period. Thirty woodland sites across Europe were sampled between 2005-2010 (Fig.1, Table S2), 10 of which were within a range of 50 km on the island of Gotland (57°10'N, 18°20'E). Overall, our studied populations fell along a south-west – northeast gradient (Fig.1). Either blood or feather samples were obtained. Most sites were sampled once, except when the sample size was too low for statistical analysis (in 10 sites). In this case, samples of two consecutive years were pooled. The number of sampled individuals per site ranged from 18 to 47 with an average of 29.

DNA was extracted with magnetic beads (MagneSil Blue, Promega AG, Dübendorf, Switzerland) and genotyped at 22 microsatellite loci (Table S3, Saladin & Richner, 2012). These 22 microsatellite markers were developed using individuals from CH.BE, a site in the geographical centre of our sampling scheme. For details on the PCR protocols and allele

scoring procedure, see Saladin & Richner, 2012). Twelve individuals with missing alleles or atypical profiles at different loci were excluded from all analyses. None of the individuals shared the same multilocus genotype indicating that none of the individuals was sampled twice. Overall, 884 individuals were analysed. Allelic dropout, scoring errors and null alleles were checked for each locus per site with MICRO-CHECKER (Van Oosterhout et al., 2004). Among all loci, no evidence for allelic dropout was detected and only one locus in one sampling site showed scoring errors. Moreover, null alleles were randomly distributed, and present at only 19 (i.e. 2.9%) locus × site combinations. Genotypic linkage disequilibrium and departure from Hardy-Weinberg equilibrium (HWE) were tested with probability tests per locus per site. In addition, departure from HWE for the overall population, i.e. across loci and sites, was tested using a multisample score test. All tests were performed using GENEPOP on the web (Rousset, 2008). P-values for multiple tests were corrected with a sequential Bonferroni procedure (Rice, 1989).

Genetic diversity and differentiation among sites

To assess genetic diversity at each sampling site, both the observed and unbiased expected heterozygosity (H_0 and H_E) were calculated using GENALEX v6 (Peakall & Smouse, 2006). In addition, the mean allelic richness per site (A_R) based on 18 individuals, corresponding to the smallest number of individuals sampled in a given site, was estimated with FSTAT v2.9.3 (Goudet, 1995). Genetic differentiation among sites was quantified using pairwise and global $F_{\rm ST}$ calculated in FSTAT with 10,000 permutations to assess significance. Because $F_{\rm ST}$ estimates may be strongly affected by the polymorphism of the markers used (Meirmans & Hedrick, 2011), standardized estimators G''_{ST} and D were calculated with GENODIVE 2.0B27 (Meirmans & Van Tienderen, 2004).

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	215	To test for a spatial pattern of genetic differentiation among sites, two methods were
	216	used: (i) a principal coordinate analysis (PCoA), based on codominant genotypic distance
	217	among sites with a standardized covariance matrix, using GENALEX 6.5 and (ii) a neighbour-
	218	joining (NJ) phenogram based on Nei's genetic distance between sites, using PHYLIP v3.68
)	219	(Felsenstein, 2008). The presence of genetic clusters was also tested using two methods. First,
<u> </u>	220	an individual-based Bayesian cluster analysis was implemented in STRUCTURE v2.2 (Pritchard,
	221	Stephens & Donnelly, 2000). Ten runs of an admixture model with correlated allele
}	222	frequencies among sites and LOCPRIOR were performed for each value of putative
)	223	population number (K) between 1 and 40 with a burn-in of 50,000 iterations followed by
	224	100,000 iterations in the Markov chain. The most likely number of genetically different
) - -	225	populations was determined from the posterior probability of the data for a given K and the
) ,	226	ΔK (Evanno, Regnaut & Goudet, 2005). To test for a potential bias due to the inclusion of 10
;)	227	close-by sites from Gotland, the PCoA and STRUCTURE analyses were run once using
)	228	individuals from all 30 sites and once using individuals from 21 sites including only a single
- 	229	site from Gotland (SE.OG). Since the results did not qualitatively differ (Fig.S2-S6, Table
))	230	S4), we presented only the results based on 30 sites. In addition, assignment probabilities of
, }	231	individuals to their original site (P _A) were calculated using a discriminant analysis of principal
)	232	components (DAPC - Jombart, Devillard & Balloux, 2010) in R 3.0.1 (R CORE TEAM, 2013).
2	233	Second, the clustering of sites into groups was investigated by a K-Means clustering using an
, - -	234	analysis of molecular variance (AMOVA) with 40 independent Markov chains with 50,000
j ,	235	iterations each assuming 2 to 15 clusters with GENODIVE. The most likely number of clusters
;)	236	was determined from the smallest bayesian information criterium (BIC). Furthermore, genetic
)	237	differentiation was quantified between groups and among sampling sites within groups using
- 	238	an AMOVA with 10,000 permutations to assess significance using GENODIVE. Additionally,
, ,	239	within-group global F_{ST} values were calculated and compared with 10,000 permutations using
}	240	FSTAT.
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To test for the presence of isolation-by-distance patterns, a decomposed pairwise regression analysis (DPR) was conducted in R to account for potential between-site differences in the gene flow-drift equilibrium (Koizumi, Yamamoto & Maekawa, 2006). Briefly, DPR first detects outlier sites based on the distribution of residuals from the overall regression between genetic and geographical distances. In a second step, genetic distances are regressed against geographical distances for each site against all other non-outlier sites to obtain a regression intercept and slope per site. The intercept and slope of the decomposed regressions measures genetic differentiation to other populations and isolation-by-distance (IBD) respectively for each site (see Koizumi et al., 2006 for details).

251 Testing for the influence of environmental factors on differences among sites

To investigate potential mechanisms underlying differences in genetic diversity, the relationships between indices of genetic diversity per site and the following environmental factors, which may be expected to influence individuals' movements, were tested: (i) geographical location (latitude and longitude); (ii) vegetation type (deciduous or coniferous trees; excluding SP.MU and ES.KI, where birds were sampled in orange tree plantations or mixed areas); (iii) temperature and (iv) minimal distance to the sea. Latitude, longitude and minimal distance to the sea were obtained using GOOGLE EARTH v5.2.1. Using the position along a southwest – northeast axis as a geographical location did not affect the results, and thus only results including latitude and longitude are reported. Temperatures were obtained from the European photovoltaic geographical information system (Huld et al., 2006). The measures based on temperature were (i) average daily temperature per month, (ii) temperature variance per year, (iii) difference between the most extreme annual temperatures and (iv) average temperature during autumn-winter (September - February) and spring-summer (March- August). In addition to the indices of genetic diversity per site, we calculated an

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estimate of effective population size (N_e) with the linkage disequilibrium method using a threshold of 0.05 for the exclusion of rare alleles in $N_FESTIMATOR v2$ (Do et al., 2014) and the kinship coefficient of Loiselle et al. (1995) averaged per site with GENODIVE. The relationships between genetic diversities (A_R and H_E), assignment probabilities, kinship coefficients, effective population sizes and environmental factors were tested using linear models since all indices were normally distributed (residuals were checked for normality and homoscedasticity). Because the environmental factors were correlated with each other (correlation coefficients ranging from 0.31 to 0.86, all P < 0.001, results not shown), Akaike's information criterion (AIC) values of models including each factor separately were compared in order to identify the environmental factor(s) that best explained the data using the package AICmodavg (Mazerolle, 2015) in R. The best models included the model with the smallest AIC and all models with a difference in AIC (Δ AIC) to this model of less than 2 (Burnham, Anderson & Huyvaert, 2011). Once the best models were identified, the significance of the effects retained was assessed with an F test.

In a second step, the influence of the following environmental factors on genetic differentiation among sampling sites was tested: (i) geographical distance between sites, (ii) mean geographical location of sites, (iii) absolute difference in average daily temperature between sites and (iv) mean of the average daily temperature of sites. Because previous analyses showed that genetic diversity was best explained by temperatures in autumn-winter (see Results section), only the difference in average autumn-winter temperatures between sites (hereafter called autumn-winter temperature difference) and the mean of the average daily temperature in autumn-winter of sites (hereafter called mean autumn-winter temperature) were tested in the analyses of genetic differentiation. Similarly, only the latitude was retained here to characterize geographical location for analyses on genetic differentiation since site latitude and longitude were correlated in our study (i.e. sites were distributed along a south-west / north-east axis). The difference between values for the two sites in pairwise

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comparisons provides a measure of the environmental contrast between sites, whereas the mean value gives a measure of the position of the pair of sites in each pairwise comparison along the environmental gradient considered (geographical position or winter severity). The genetic differentiation between sites was calculated for each pair of sites and summarized in a pairwise matrix; the same approach was used for the differences and mean values of the environmental factors between sites. Correlations between levels of pairwise genetic differentiation based on either F_{ST} , G''_{ST} or D and pairwise differences in environmental factors were investigated with Mantel tests (or partial Mantel tests when more than two matrices were compared) with 10,000 permutations using the package vegan (Oksanen et al., 2011) in R. By homogenising the genetic composition of connected populations, gene flow should reduce both the mean level and the variability of genetic differentiation between populations (Hutchison & Templeton, 1999). Consequently, a factor affecting gene flow should be correlated with both the level of genetic differentiation and the absolute values of residuals of the linear regression between the factor and the level of genetic differentiation (hereafter called residual pairwise F_{ST} , G''_{ST} or D respectively) Hutchison & Templeton, 1999). Therefore, the correlation between matrices of environmental factors and their residual pairwise genetic differentiation was also tested.

Results

Genetic diversity and equilibrium

No evidence for linkage disequilibrium at any locus in any site or departure from HWE was found after correction for multi-comparisons. Pooling all sites, a significant deviation from HWE was observed (score-test: P < 0.001), suggesting the existence of sub-populations. The number of alleles per locus ranged from 4 to 41 with an average of 16 alleles across loci.

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Mean allelic richness per site ranged from 6.32 to 7.66 (Table S2). Expected heterozygosity varied between 0.60 and 0.68 and the number of effective alleles between 3.94 and 4.92 (Table S2). F_{IS} per site ranged from -0.049 to 0.047 (Table S2), but no F_{IS} value differed significantly from zero after correcting for multiple tests, as expected under within-site HWE.

Genetic differentiation among sampling sites

Genetic differentiation among sampling sites across Europe was low, but significant (global $F_{\text{ST}} = 0.008$, $G''_{\text{ST}} = 0.024$, D = 0.016, all P < 0.001). Pairwise F_{ST} ranged from -0.004 to 0.040 (Table S3). Out of 435 pairwise $F_{\rm ST}$ comparisons, 147 (i.e. 33.8%) were significantly different from zero after sequential Bonferoni correction. Interestingly, the majority of significant comparisons (134 out of 147, i.e. 91.1%) involved six (out of seven) sampling sites located in the south-western part of Europe, i.e. below 47°N (CH.BE, FR.MO, SP.MU, SP.FR, SP.MA and PO.CO), indicating different levels of genetic differentiation between northern and southern sites (Fig. 2). FR.RO was the only site located in the southern region for which pairwise F_{ST} values were non-significant. Results of both the PCoA analysis and NJ phenogram based on Nei's genetic distance were congruent with the observed pairwise F_{ST} pattern for six out of the seven southern sites (Fig. 3). The PCoA accounted for 62% of the total genetic variation on the first 3 axes (26.5%, 18% and 17.5% respectively). Independently of the axes considered, PO.CO, SP.MU, SP.MA, SP.FR, FR.MO were identified as being rather distinct from all other sites (i.e. outside the 50% and close to the 95% limit of the confidence interval; Fig. 3a-b). These south-western sites were also differentiated from each other, except SP.MA and SP.FR, which also showed lower pairwise F_{ST} values. Only CH.BE, which had relatively low F_{ST} values, was not identified as a differentiated site by the PCoA and the NJ phenogram analyses. Furthermore the central cluster was randomly distributed on each PCoA axis, in particular with no clumping of the 10 close-by sampling sites located on

Gotland (Fig. 3a-b), which was confirmed on the NJ phenogram. In fact, populations on Gotland showed similar levels of differentiation among themselves as among the other sites from northern Europe (Fig. 3c). Depending on the method used, some of the northern sites appeared differentiated from the central cluster (e.g. SE.LO, Fig. 2b and 3a; SE.SA, Fig. 3a-c; or NE.LA, Fig. 3a) suggesting that they could be distinct from the central cluster yet less differentiated than the south-western sites. Overall, the results indicate that (i) genetic differentiation among sampling sites was low (Fig. 2 and 3); (ii) many sites (including close-by ones) presented similar and low levels of genetic differentiation without spatial structure (e.g. a centred star-like pattern; Fig. 3c); and (iii) at least five southern sites were differentiated from the central cluster and differentiated from each other, except SP.MA and SP.FR (Fig. 2 and 3).

STRUCTURE identified three genetic clusters (K=3) following the Evanno correction (Fig. S4-S5). Two of these clusters were mainly associated with the four Iberian sites, where the Portuguese site (PO.CO) was further distinct from all Spanish sites (SP.MU, SP.MA and SP.FR), however no individual was fully assigned to either cluster (Fig. S6). All other sites were predominantly assigned to a third cluster except for CH.BE, which showed evidence for introgression from south-western Europe. Concordantly, the AMOVA based K-means clustering identified two groups: one comprising the four Iberian sites and CH.BE and a second including all other sites (all northern sites and the two sites in France). The AMOVA using south-western (i.e. below 47° latitude: PO.CO, SP.MU, SP.MA, SP.FR, FR.MO, FR.RO, CH.BE) and northern (above 47° latitude) sites as grouping variable suggested low but significant genetic differentiation between these groups ($F_{\text{group-total}} = 0.002$, P < 0.001) and among sites within groups ($F_{\text{sites-group}} = 0.008$, P < 0.001). In addition, the differentiation was higher within southern sites than other sites (global $F_{ST} = 0.016$ and 0.005, $G''_{ST} = 0.052$ and 0.014, D = 0.034 and 0.009, for southern sites and other sites, respectively; P < 0.001).

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Excluding CH.BE, FR.MO and FR.RO, did not change qualitatively the results of the hierarchical AMOVA and the level of differentiation, suggesting that the observed clustering was mainly driven by the four Iberian sites, which are more differentiated than the other south-western sites. Interestingly, the weak overall differentiation among the northern sites did not result from differentiation between specific sampling sites since 19 sites had to be excluded one after the other (starting from the sites with the highest mean pairwise F_{ST} value and going downwards) for the overall differentiation to become non-significant (results not detailed). Moreover, differentiation among the close-by sites on Gotland (with distance ranging from 3 to 50 km) was not lower than among other northern sites (global $F_{ST} = 0.006$ and 0.004 respectively, P = 0.646; Fig. 5a).

Finally, the DPR analysis identified FR.MO (the only urban site) as an outlier, since the model excluding this site had a lower AIC (-94.78) and higher R^2 (0.17) values, although other models (either comprising all sites or with additional outliers) could not be excluded ($\Delta AIC < 1.28$). Overall, the DPR divided sampling sites into five groups (see Fig. 1 for location, Table S2): (1) two southern sites (FR.MO and FR.RO) showed a significant atypical negative IBD pattern and significant differentiation from other sites; (2) the four Iberian sites (SP.MU, SP.MA, SP.FR and PO.CO) and CH.BE showed no significant IBD but significant differentiation from other sites; (3) ten northern sites in Fennoscandia showed both significant differentiation from other sites and an IBD pattern; (4) nine northern sites from different locations showed no differentiation from other sites but significant IBD; and (5) four central sites (UK.WY, UK.CA, BE.CE, BE.BO) showed no differentiation and no IBD. Interestingly, all but two close-by sites on Gotland showed both significant differentiation from other sites and an IBD pattern.

Exploring the influence of environmental factors on genetic differences among sites

Models including latitude, longitude, and the variance and difference in daily temperature were retained for none of the five indices (allelic richness A_R , expected heterozygosity H_E , assignment probability P_A , kinship coefficient and effective population size N_e , $\Delta AIC > 2$ in all cases; Table S5). Conversely, models with average daily temperature for months September to January, and consequently average autumn-winter temperature, were among the models best explaining the data for P_A , A_R , H_E ($\Delta AIC < 2$ in all cases; Table S5). For kinship coefficient, models with average daily temperature for months August, September and December were among the models best explaining the data ($\Delta AIC < 2$) but not the model with average autumn-winter temperature despite a relative low AIC ($\Delta AIC < 2.5$). The model including vegetation type was the only best model in explaining the data for the effective population size. Allelic richness decreased ($F_{1,28} = 6.90$, P = 0.014, $R^2 = 0.20$) while assignment probabilities and kinship coefficients increased ($F_{1,28} = 10.57$, P = 0.003, $R^2 =$ 0.27; $F_{1,28} = 17.04$, P < 0.001, $R^2 = 0.36$ respectively) with increasing average autumn-winter temperature (Fig. 4). Expected heterozygosity and effective population size were not correlated with average autumn-winter temperature ($F_{1,28}=0.81$, P = 0.38; $F_{1,24} = 0.56$, P =0.46 respectively, Fig. 4). Effective population size was similar for coniferous and deciduous forests ($F_{1,22} = 0.03$, P = 0.87). Models with other factors were retained for part of the indices only: temperatures in summer months (July to August) for AR and HE, temperature in February for H_E , average spring-summer temperature for H_E , vegetation type for A_R and distance to the sea for H_E (Table S5). However, allelic richness was similar in coniferous and deciduous forests ($F_{1,26} = 0.08$, P = 0.77), and expected heterozygosity was not correlated with spring-summer temperature or distance to the sea ($F_{1,28} < 2.5$, P > 0.12). Based on these results, only the average autumn-winter temperature was retained among temperature measures for the analyses of genetic differentiation.

All pairwise genetic differentiation indices increased with geographical distance
between sites, autumn-winter temperature difference between sites and mean autumn-winter

 temperature of the two sites in pairwise comparisons, and decreased with mean latitude of the two sites (Table 2; Fig. 5). Each environmental factor explained 36 to 57% of the variation in pairwise genetic differentiation. Furthermore, both mean autumn-winter temperature and latitude, but not geographical distance or autumn-winter temperature difference, were correlated with their respective residual pairwise genetic differentiation (Table 2). This suggests that genetic differentiation is mainly driven by site characteristics (latitude, mean autumn-winter temperature) rather than environmental contrast between sites. Mean autumn-winter temperature remained significantly correlated with genetic differentiation after correcting for latitude (partial Mantel test: $r_M = 0.31$, P = 0.019), whereas mean latitude was not correlated with genetic differentiation anymore after correcting for mean autumn-winter temperature (partial Mantel test: $r_M = -0.03$, P = 0.534). This suggests that mean autumn-winter temperature was the best predictor of genetic differentiation among the tested environmental effects.

431 Discussion

432 Biological relevance of the observed genetic differentiation

The low but significant global genetic differentiation based on microsatellite markers suggests extensive gene flow among great tit populations across Europe. Nevertheless, the overall deviation from Hardy-Weinberg equilibrium, the absence of inbreeding within sites (as revealed by heterozygosity) and the overall population differentiation support a Wahlund effect, i.e. a substructure among sites. Individual-based clustering methods failed to characterise discrete genetic groups, yet found some indication for substructure among southwestern sites. This is consistent with the high proportion of the genetic variance (> 98%)observed within populations (e.g. Chen et al., 2007; Latch et al., 2006). We are nevertheless

441 confident about the validity of the significant global genetic differentiation given the 442 relatively large sample sizes and because none of the analyses suggested a bias in both global 443 and pairwise genetic differentiation due to variation in sample size among sites or being 444 associated by specific loci and sites.

In general, a significant IBD supports the biological relevance of low genetic differentiation among populations (e.g. F_{ST} values around 0.003), especially in species characterised by large population sizes and high gene flow such as birds (e.g. Prochazka et al., 2011) or marine fish (e.g. Purcell et al., 2006). But low genetic differentiation even in absence of IBD may also reflect heterogeneity in gene flow affecting ongoing microevolutionary processes in highly mobile organisms. This is illustrated by the case of a physically isolated island population of great tits, where immigrants from the mainland can be easily identified (Postma & van Noordwijk, 2005). In this population, direct (i.e. observed movements of individuals) and indirect (i.e. genetic, based on microsatellite markers) measures of gene flow were compared. The genetic differentiation between resident and immigrant individuals was low but significant (F_{ST} =0.007; Postma et al., 2009). Consistent with a higher immigration rate in the western part (43%) compared to the eastern part (13%) of the study island, a low but significant genetic differentiation was found between the two parts (F_{ST} =0.011; Postma et al., 2009). Because mainland individuals lay larger clutches, immigration was shown to impede local adaptation in the western but not the eastern part of the island (Postma & van Noordwijk, 2005). Using similar microsatellite markers in the present study, we also found comparable levels of genetic differentiation between populations, supporting the biological implications of our findings. Lastly, using a restricted set of microsatellite markers, we retrieved a comparable level of genetic differentiation between two sites (NE.HO and UK.WY; $F_{ST} = 0.005$) as has been observed with several thousand SNP markers for the same sites (van Bers *et al.* 2012; $F_{ST} = 0.010$). The slightly higher level of genetic differentiation in their study could be due to the inclusion of some highly divergent outlier loci. Another study

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467 also using the same SNP set further identified cryptic genetic differentiation within the 468 UK.WY site, which was similarly driven by few (<1%) markers (Garroway et al., 2013). Thus 469 our microsatellite data set seems to be suitable to accurately calculate population genetic 470 estimates that resemble average genome wide patterns (i.e.Van Bers et al., 2012), whereas 471 few genomic regions may exist that underlie patterns of local adaptation (Garroway et al., 472 2013; Van Bers et al., 2012).

Our analyses revealed higher genetic differentiation in south-western compared to northern European sites. This finding suggests decreased gene flow between south-western and northern Europe as well as within south-western Europe. Subsequent generalisations towards other southern European populations need to be done with caution since our sampling design focused only on south-western populations. A similar pattern was reported for different passerine species as well as for plants and mammals (Hewitt, 2000; Kvist et al., 2004; Pentzold et al., 2013; Prochazka et al., 2011) and is generally interpreted as the result of postglacial recolonisation. In the present case, the higher divergence of southern populations compared to northern ones could be due to the fact that both groups may have derived from different glacial refugia (Hewitt, 2000). Such scenario has been suggested for other tit species, for which distinct glacial refugia may have existed in the Mediterranean region (Kvist et al., 2004) and across Europe (Pentzold et al., 2013). However, for several reasons, the genetic differentiation observed in great tits using microsatellite markers seems unlikely to result from the occurrence of one or several genetic lineages that have recolonized northern Europe from distinct refugia. First, the presence of several glacial refugia would have led to the existence, at least in southwestern populations, of genetic variations specific to the multiple refugia causing a higher genetic diversity within the Iberian Peninsula (Pentzold et al., 2013; Prochazka et al., 2011). In contrast, the Iberian Peninsula harboured a level of allelic richness at microsatellite markers that was comparable to all other sites (7.26 and 7.32 alleles respectively). Similarly, phylogenetic studies showed in great tits a homogeneous

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mitochondrial diversity from northern to southern Europe (Fig S1), which is consistent with a colonisation from a single refugium and the absence of strong geographical barriers to dispersal (Kvist et al., 2003; Kvist et al., 1999; Pavlova et al., 2006). Second, a rapid post-glacial range expansion from a single refugium is likely to result in lower genetic diversity within the colonized range as opposed to the ancestral refugium (Antoniazza et al., 2014; Pavlova et al., 2006). In contrast, Iberian populations had a slightly lower allelic richness per site compared to all other sites (6.71 \pm 0.29 and 7.11 \pm 0.29 alleles respectively, Table S2). Interestingly haplotype diversity was lower in all south-western populations than in the north-eastern populations in coal tits (Pentzold et al., 2013) suggesting that a lower genetic diversity in southern regions could have arisen long time ago. However such pattern was not detected with mitochondrial DNA in great tits (Pavlova et al., 2006). Therefore, the observed patterns of genetic differentiation at microsatellite loci among great tit populations are unlikely to result from post-glacial recolonization processes from one or several refugia but rather represent other historical and/or recent processes. Further studies using genetic modelling approaches combined with increased genomic coverage are however necessary to elucidate the factors underlying the pattern observed here.

How could gene flow be shaped by temperature?

Latitude and (autumn-winter) temperature were significantly correlated with both the level of genetic differentiation among populations and its level of variation in contrast to the geographical distance and the difference of temperature that explained only the level of genetic differentiation among populations. Moreover, only temperature was significantly associated with the level of genetic differentiation after taking into account latitude. Finally, temperature but not latitude explained the decrease of genetic diversity from the South to the North. The effect of temperature on different components of the genetic variation suggests a Page 23 of 58

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518	strong relationship between temperature and neutral genetic structure among great tit
519	populations. However we cannot exclude that temperature is correlated with additional
520	environmental factors such as photoperiod or irradiance cues (De Frenne et al., 2013) and
521	then the correlation between temperature and genetic differentiation is a by-product of the
522	effect of environmental factors on genetic variation that we did not measure here. Nonetheless
523	the relationship between temperature and neutral genetic structure suggests that genetic
524	differentiation, and hence gene flow, may be related to winter local movements and partial
525	migration (Nilsson, Alerstam & Nilsson, 2008; Nowakowski & Vähätalo, 2003). This could
526	also be associated with winter severity: food availability may be especially restricted in
527	northern Europe (Newton, 2011 but see Nilsson et al., 2008; Nowakowski & Vähätalo, 2003)
528	when insect abundances are lowest and great tits become mainly granivorous (Vel'ky, Kanuch
529	& Kristin, 2011). Great tits are considered to be resident in southern and western Europe, but
530	partial migrants in northern Europe, as shown in particular by captures at migratory passage
531	sites in autumn and spring (Nowakowski & Vähätalo, 2003; Poluda, 2011; Gosler, 2002). Part
532	of the birds (especially juveniles) may move during winter over short to long distances (up to
533	> 1000 km; Nilsson et al., 2008; Nowakowski & Vähätalo, 2003). In spring, these migrants
534	may either stay on the wintering grounds or return to their natal region to breed more or less
535	close to their natal site (Gosler, 2002; Nilsson et al., 2008; Nowakowski & Vähätalo, 2003).
536	Partial migration could therefore generate on average longer dispersal distances, associated
537	with higher variance, in the northern compared to southern European populations (see Orell et
538	al., 1999). Although part of the immigrant individuals (often around 50% of local breeders in
539	monitored populations) may originate from the surroundings of study areas (e.g. Verhulst et
540	al., 1997), differences in immunological, behavioural and/or life-history traits between
541	potential immigrants (i.e. not previously captured in the population) and locally born
542	individuals (e.g. Postma & van Noordwijk, 2005; Snoeijs et al., 2004) may support the
543	existence of long-distance immigration in great tits. Because obtaining additional information

544 on the origin of immigrant individuals in the field is highly challenging, this hypothesis 545 however remains difficult to test.

Interestingly, similar genetic structures across Europe have been found in other small passerine species, i.e. for the bluethroat (Luscinia svecica; Johnsen et al., 2006) and the pied flycatcher (Ficedula hypoleuca; Lehtonen et al., 2009). In the latter case, no large-scale differentiation was observed in north-eastern Europe but small-scale differentiation was found in southern Europe. Because the pied flycatcher is an obligatory migratory species, wintering in Sub-Saharan Africa, the lower genetic differentiation of northern sites cannot be explained by differences in winter movements linked to winter severity. Nevertheless, lower philopatry and local recruitment rates, and thus higher dispersal rates, have been suggested in northern compared to southern sites for several migratory species, including the pied flycatcher (Lehtonen et al., 2009) and the barn swallow (Balbontin et al., 2009). In these species dispersal may be linked to other environmental factors such as e.g. habitat stability, fragmentation or elevation. Both here and in the study by Lehtonen et al. (2009), southern populations were sampled in specific habitats, including high elevation sites (great tits: SP.MA, SP.FR and CH.BE > 500 m.a.s.l.; pied flycatchers: Lehtonen et al., 2009), urban environment (FR.MO) or plantations (SP.MU), in contrast to northern sites located mainly in temperate lowland forests. In southern Europe, stable habitat heterogeneity, niche specialisation or high temperature may promote local adaptation (e.g.Husby, Visser & Kruuk, 2011). This could increase local genetic differentiation and select against dispersal to a higher degree than in the northern regions (Van Doorslaer et al., 2009), where the availability of large and/or homogeneous habitat patches may reduce dispersal costs (Travis & Dytham, 1999) in both migratory and sedentary species. Individuals of the southern populations may therefore be less prone to accept breeding in new sites, leading to lower gene flow. Consequently, intraspecific differentiation might be more likely than neutral differentiation in southern sites (e.g. Johnsen et al., 2006; Lehtonen et al., 2011).

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571 Conclusion

Non-random dispersal and genetic structure in great tits have previously been investigated at small scales, providing evidence for local adaptation (i.e. within a few km; Garant et al., 2005; Garroway et al., 2013; Postma et al., 2009; Postma & van Noordwijk, 2005). Here, we compared populations across Europe and found low but significant genetic differentiation among populations. This differentiation was unrelated to geographical distance between sites but was influenced by geographic location and environmental factors, in particular autumn-winter temperature. This might have important implications for the evolutionary trajectories of great tit populations and other species showing similar patterns. The northern populations may represent a single large population in which gene flow drives demographic and evolutionary processes. In this case, habitat choice and assortative mating may play a central role in local adaptation processes (e.g. Postma & van Noordwijk, 2005). In contrast, the southern populations may be more isolated and experience stronger genetic drift and/or higher selective pressures (e.g. Lehtonen et al., 2011). Studying potentially ongoing intraspecific diversification may be particularly relevant in these populations.

The association between genetic differentiation and winter severity may have further implications in the context of climate change. If the increase of winter temperatures favours increased philopatry in northern populations (e.g. Van Vliet, Musters & Ter Keurs, 2009), the latter may reach a gene flow-drift equilibrium. As a consequence, increased genetic differentiation and IBD could arise, favouring neutral genetic differentiation and/or local adaptation. Conversely, southern populations may become extinct if genetic adaptation or phenotypic plasticity fail to allow to adapt sufficiently fast (Visser, 2008; Boeye et al., 2013). Alternatively, an increase of philopatry among northern populations, induced by warmer winters could intensify competition especially during the breeding season, leading to a

population decline (Kokko, 2011 but see Stenseth et al., 2015). And southern populations may persist if climate change combined with habitat fragmentation select for less emigration but larger dispersal distances (Boeye et al., 2013; Fronhofer et al., 2014). If global warming results in population extinction, proportionally more genetic diversity would be lost in the South than in the North of Europe. Because most studies on great tits have been conducted in north-central Europe, further work is needed to assess both the large-scale variation of philopatry, its relation to local and regional winter partial migration movements and its consequence in terms of gene flow between populations.

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619	Table 1. Decomposed pairwise regression (DPR) of the genetic differentiation with
620	geographic distance for each sampling site. Intercepts indicate the level of differentiation of
621	sites, and slopes indicate isolation-by-distance (IBD). FR.MO was identified as an outlier site
622	and was therefore excluded to calculate the pairwise regressions of other sites. Significant
623	values are indicated in bold.

	Intercep	pt ± SE		Slope	± SE					
Site	(10) ⁻²)	Р	(10	-6)	Р	\mathbb{R}^2	Genetic differentiation pattern		
FR.MO	2.57	0.26	0.000	-3.87	1.71	0.032	0.159	Negative IBD, differentiated		
FR.RO	0.95	0.27	0.002	-3.75	1.77	0.044	0.147	sites		
SP.MU	1.79	0.21	0.000	0.07	1.06	0.950	0.000			
PO.CO	1.70	0.22	0.000	0.34	0.97	0.729	0.005			
SP.MA	1.28	0.23	0.000	-0.24	1.09	0.828	0.002	No IBD, differentiated sites		
SP.FR	1.21	0.21	0.000	-0.58	0.99	0.561	0.013			
CH.BE	0.61	0.22	0.010	1.73	1.80	0.344	0.035			
SE.SA	0.92	0.14	0.000	2.65	1.13	0.027	0.175			
SE.LO	0.87	0.12	0.000	4.35	1.01	0.000	0.418			
SE.BO	0.52	0.12	0.000	2.86	1.02	0.010	0.232			
FI.TU	0.50	0.17	0.006	2.36	1.11	0.043	0.148			
NO.DA	0.44	0.13	0.002	2.59	1.02	0.017	0.199	IBD, differentiated sites		
SE.OG	0.42	0.11	0.001	3.50	0.92	0.001	0.355	IBD, differentiated sites		
SE.JA	0.35	0.11	0.005	5.31	0.93	0.000	0.558			
SE.GE	0.33	0.12	0.009	2.51	0.97	0.016	0.203			
SE.SI	0.30	0.13	0.035	3.34	1.08	0.005	0.269			
SE.BI	0.26	0.11	0.021	2.61	0.86	0.005	0.264			
NE.LA	0.43	0.24	0.081	6.43	2.27	0.009	0.235			
FI.KO	0.20	0.17	0.251	2.38	0.94	0.018	0.197			
SE.ET	0.10	0.12	0.407	3.23	0.99	0.003	0.292			
NE.HO	0.07	0.21	0.754	4.33	2.02	0.042	0.150			
SE.DT	0.01	0.13	0.944	4.13	1.06	0.001	0.369	IBD, undifferentiated sites		
NE.WE	0.00	0.17	0.994	3.94	1.55	0.018	0.199			
ES.KI	0.00	0.14	0.997	2.14	0.97	0.036	0.159			
PL.PU	-0.16	0.25	0.520	4.30	1.98	0.039	0.154			
HU.PI	-0.21	0.29	0.490	5.21	2.24	0.028	0.172			
UK.WY	0.45	0.25	0.089	0.80	1.96	0.687	0.006			
BE.BO	0.38	0.22	0.098	2.83	2.03	0.175	0.070	No IBD, undifferentiated sites		
BE.CE	0.27	0.17	0.127	2.76	1.57	0.090	0.106	ino IBD, undifferentiated sites		
UK.CA	0.22	0.23	0.331	2.84	1.83	0.132	0.085			
All	0.45	0.06	0.000	3.21	0.40	0.000	0.130	IBD, differentiated sites		

Table 2: Effects of environmental factors on the genetic differentiation between sampling sites across Europe and its variation based on Mantel tests

625 (r_M). Genetic distance was measured as pairwise F_{ST} , G''_{ST} and D and their variation was investigated using the residuals of a linear regression

between each environmental factor and the respective pairwise genetic distances. See text for details. Significant correlations are indicated in bold.

Response variable:		, ST	Residuals on $F_{\rm ST}$		G" _{ST}		Residuals on G''_{ST}		D		Residuals on D		
Explanatory variable:	r _M	Р	r _M	Р	r _M	Р	r _M	Р	r _M	Р	r _M	Р	
Mean autumn- winter temperature	0.57	<0.001	0.17	0.022	0.57	<0.001	0.16	0.025	0.57	<0.001	0.15	0.02	
Latitude	-0.50	0.002	-0.23	0.010	-0.50	0.002	-0.22	0.010	-0.51	0.001	-0.22	0.00	
Geographic distance	0.36	<0.001	0.06	0.205	0.37	0.001	0.07	0.188	0.37	<0.001	0.07	0.16	
Difference in autumn-winter temperature	0.39	0.002	0.10	0.125	0.39	0.002	0.10	0.120	0.39	0.002	0.10	0.1	

Figure legends:

Figure 1. Location of the 30 sampling sites across Europe. The inset shows the 10 sampling sites on the island of Gotland, Sweden. The dashed line shows the 47° latitude. IBD analysis treats all populations into a single quantity assuming that all local populations have similar characteristics. In contrast, the DPR analysis extracts the elements of individual local population from the information on an entire metapopulation and identifies five groups differing in relative strengths of gene flow and genetic drift patterns (i.e. different patterns of genetic differentiation and IBD summarised by different colours).

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Figure 2. Heatmap of the pairwise F_{ST} values between all sites. Sites are ordinated by pairwise F_{ST} values. Black bars highlight the sites located below the 47° latitude (i.e. southwestern sites).

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Figure 3. (a, b) Principal coordinate analysis (PCoA) contrasting axes 1 vs. 2 (a) and 1 vs. 3 (b), and (c) NJ phenogram based on Nei's genetic distance, with bootstrap values of specific clusters. (a,b) On the PCoA plots, the smallest and largest ellipses represent the 50% and 95% confidence intervals respectively; black dots represent the five sites identified as satellites, grey dots potential other satellites and white dots non-differentiated sites. (c) On the NJ phenogram, the three grey circles indicate identified clusters and the five sites identified as e indicateu n oo... satellites are indicated in bold.

- **Figure 4.** Relationships between average autumn-winter temperature and (a) latitude and (b-f)
- 655 different population indices: assignment probability (b), allelic richness (c), mean pairwise
- 656 kinship (d), unbiased expected heterozygosity (e), effective population size (f).

For peer perieu

2 3 4	658	Figure 5. Relationship between pairwise F_{ST} values and (a) geographic distance between sites
4 5 6	659	and (b) mean autumn-winter temperature of the two sites in pairwise comparisons. White
7 8	660	dots: pairwise F_{ST} values between northern sites; grey dots: pairwise F_{ST} values between one
9 10	661	northern and one south-western site; black dots: pairwise F_{ST} values between south-western
11 12	662	sites.
13 14	663	
15 16 17	664	
$\begin{array}{c} 17\\ 18\\ 19\\ 20\\ 21\\ 22\\ 23\\ 24\\ 25\\ 26\\ 27\\ 28\\ 29\\ 30\\ 31\\ 32\\ 33\\ 34\\ 35\\ 37\\ 38\\ 39\\ 40\\ 41\\ 42\\ 43\\ 44\\ 50\\ 51\\ 52\\ 53\\ 45\\ 56\\ 57\\ 58\\ 59\\ 60\\ \end{array}$	665	

- Ahola MP, Laaksonen T, Eeva T, Lehikoinen E. 2009. Great tits lay increasingly smaller
 clutches than selected for: a study of climate- and density-related changes in
 reproductive traits. *Journal of Animal Ecology* 78: 1298-1306.
- Antoniazza S, Kanitz R, Neuenschwander S, Burri R, Gaigher A, Roulin A, Goudet J.
 2014. Natural selection in a postglacial range expansion: the case of the colour cline in
 the European barn owl. *Molecular Ecology* 23: 5508-5523.
 - Balbontin J, Moller AP, Hermosell IG, Marzal A, Reviriego M, de Lope F. 2009.
 Geographic patterns of natal dispersal in barn swallows *Hirundo rustica* from
 Denmark and Spain. *Behavioral Ecology and Sociobiology* 63: 1197-1205.
- Björklund M, Ruiz I, Senar JC. 2010. Genetic differentiation in the urban habitat: the great
 tits (*Parus major*) of the parks of Barcelona city. *Biological Journal of the Linnean Society* 99: 9-19.
- Boeye J, Travis JMJ, Stoks R, Bonte D. 2013. More rapid climate change promotes
 evolutionary rescue through selection for increased dispersal distance. *Evolutionary Applications* 6: 353-364.
- Burnham KP, Anderson DR, Huyvaert KP. 2011. AIC model selection and multimodel
 inference in behavioral ecology: some background, observations, and comparisons. *Behavioral Ecology and Sociobiology* 65: 23-35.
- Caswell H, Lensink R, Neubert MG. 2003. Demography and dispersal: life table response
 experiments for invasion speed. *Ecology* 8: 1968-1978.

Charmantier A, Doutrelant C, Dubuc-Messier G, Fargevieille A, Szulkin M. 2015. Mediterranean blue tits as a case study of local adaptation. *Evolutionary Applications*: n/a-n/a.

3
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5 6 7
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8 9 10 11
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46
47
48
49
50
51
52
52 53
53 54
55
56
57
58
59
60

690 Chen C, Durand E, Forbes F, Francois O. 2007. Bayesian clustering algorithms
691 ascertaining spatial population structure: a new computer program and a comparison
692 study. *Molecular Ecology Notes* 7: 747-756.

693 De Frenne P, Graae BJ, Rodriguez-Sanchez F, Kolb A, Chabrerie O, Decocq G, De Kort
694 H, De Schrijver A, Diekmann M, Eriksson O, Gruwez R, Hermy M, Lenoir J,

- 695 Plue J, Coomes DA, Verheyen K. 2013. Latitudinal gradients as natural laboratories
 696 to infer species' responses to temperature. *Journal of Ecology* 101: 784-795.
- bo C, Waples RS, Peel D, Macbeth GM, Tillett BJ, Ovenden JR. 2014. NEESTIMATOR
 v2: re-implementation of software for the estimation of contemporary effective
 - 699 population size (Ne) from genetic data. *Molecular Ecology Resources* 14: 209-214.
- For a simulation study. *Molecular Ecology* 14: 26112620.
 - Felsenstein J. 2008. PHYLIP (Phylogeny Inference Package) Version 3.6.8. In: author Dbt,
 ed. Seattle: Departement of Genome Sciences, University of Washington.
 - Forsman JT, Monkkonen M. 2003. The role of climate in limiting European resident bird
 populations. *Journal of Biogeography* 30: 55-70.
- Fronhofer EA, Stelz JM, Lutz E, Poethke HJ, Bonte D. 2014. Spattially correlated
 extinctions select for less emigration but larger dispersal distances in the spider mite
 Tetranychus urticae. Evolution 68: 1838-1844.
- Garant D, Kruuk LEB, McCleery RH, Sheldon BC. 2004. Evolution in a changing
 environment: A case study with great tit fledging mass. *American Naturalist* 164:
 E115-E129.
 - Garant D, Kruuk LEB, Wilkin TA, McCleery RH, Sheldon BC. 2005. Evolution driven
 by differential dispersal within a wild bird population. *Nature* 433: 60-65.

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19 20
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Garroway CJ, Radersma R, Sepil I, Santure AW, De Cauwer I, Slate J, Sheldon BC.
2013. Fine-scale genetic structure in a wild bird population: the role of limited
dispersal and environmentally based selection as causal factors. *Evolution* 67: 34883500.

Gosler AG. 2002. Great tit, Parus major. In: Wernham C, Siriwardena G, Toms M, Marchant

JH, Clark JA and Baillie S, eds. *The Migration Atlas: Movements of the Birds of Britian and Ireland*. London: T & AD Poyser. 602-605.

- Goudet J. 1995. FSTAT (Version 1.2): A computer program to calculate F-statistics. *Journal of Heredity* 86: 485-486.
- Hewitt G. 2000. The genetic legacy of the Quaternary ice ages. *Nature* 405: 907-913.
- Huld TA, Suri M, Dunlop ED, Micale F. 2006. Estimating average daytime and daily
 temperature profiles within Europe. *Environmental Modelling & Software* 21: 16501661.
- Husby A, Visser ME, Kruuk LEB. 2011. Speeding up microevolution: The effects of
 increasing temperature on selection and genetic variance in a wild bird population.
 PLoS Biology 9: 9.
- Hutchison DW, Templeton A, R. 1999. Correlation of pairwise genetic and geographic
 distance measures: inferring the relative influences of gene flow and drift on the
 distribution of genetic variability. *Evolution* 53: 1898-1914.
- Johnsen A, Andersson S, Garcia Fernandez J, Kempenaers B, Pavel V, Questiau S,
 Raess M, Rindal E, Lifjeld JT. 2006. Molecular and phenotypic divergence in the
 bluethroat (*Luscinia svecica*) subspecies complex. *Molecular Ecology* 15: 4033-4047.
- Jombart T, Devillard S, Balloux F. 2010. Discriminant analysis of principal components: a
 new method for the analysis of genetically structured populations. *BMC Genetics* 11:
 Article No.: 94.

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9 10	
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58	
59	

60

740	Junge C, Vøllestad LA, Barson NJ, Haugen TO, Otero J, Sætre G-P, Leder EH,
741	Primmer CR. 2011. Strong gene flow and lack of stable population structure in the
742	face of rapid adaptation to local temperature in a spring-spawning salmonid, the
743	European grayling (Thymallus thymallus). Heredity 106: 460-471.
744	Koizumi I, Yamamoto S, Maekawa K. 2006. Decomposed pairwise regression analysis of
745	genetic and geographic distances reveals a metapopulation structure of stream-

746 dwelling Dolly Varden charr. *Molecular Ecology* **15:** 3175-3189.

747 Kokko H. 2011. Directions in modelling partial migration: how adaptation can cause a 748 population decline and why the rules of territory acquisition matter. *Oikos* 120: 1826749 1837.

- Kvist L, Arbabi T, Päckert M, Orell M, Martens J. 2007. Population differentiation in the
 marginal populations of the great tit (Paridae: *Parus major*). *Biological Journal of the Linnean Society* 90: 201-210.
- Kvist L, Martens J, Higuchi H, Nazarenko AA, Valchuk OP, Orell M. 2003. Evolution
 and genetic structure of the great tit (*Parus major*) complex *Proceedings of Royal Society of London Serie B: Biological sciences* 270: 1447-1454.
- Kvist L, Ruokonen M, Lumme J, Orell M. 1999. The colonization history and present-day
 population structure of the european great tit (*Parus major major*). *Heredity* 82: 495502.
- Kvist L, Viiri K, Dias PC, Rytkonen S, Orell M. 2004. Glacial history and colonization of
 Europe by the blue tit *Parus caeruleus*. *Journal of Avian Biology* 35: 352-359.
- Latch EK, Dharmarajan G, Glaubitz JC, Rhodes OE. 2006. Relative performance of
 Bayesian clustering software for inferring population substructure and individual
 assignment at low levels of population differentiation. *Conservation Genetics* 7: 295 302.

765	Lehtonen P, Laaksonen T, Artemyev AV, Belskii E, Berg PR, Both C, Buggiotti L, Bureš
766	S, Burgess M, Bushuev AV, Krams I, Moreno J, Mägi M, Nord A, Potti J,
767	Ravussin P-A, Sirkiä P, Sætre G-P, Winkel W, Primmer CR. 2011. Candidate
768	genes for colour and vision exhibit signals of selection across the pied flycatcher
769	(Ficedula hypoleuca) breeding range. Heredity.
770	Lehtonen P, Laaksonen T, Artemyev AV, Belskii E, Both C, Bureš S, Bushuev AV,
771	Krams I, Moreno J, Mägi M, Nord A, Potti J, Ravussin P-A, Sirkiä P, Sætre G-P,
772	Primmer CR. 2009. Geographic patterns of genetic differentiation and plumage
773	colour variation are different in the pied flycatcher (Ficedula hypoleuca). Molecular
774	<i>Ecology</i> 18: 4463-4476.
775	Loiselle BA, Sork VL, Nason J, Graham C. 1995. Spatial genetic-structure of a tropical
776	understory shrub, Psychtria officinalis (Rubiaceae) American Journal of Botany 82:
777	1420-1425.
778	Matthysen E. 2005. Density-dependent dispersal in birds and mammals. Ecography 28: 403-
779	416.
780	Mazerolle MJ. 2015. AICcmodavg: Model selection and multimodel inference based on
781	(Q)AIC(c). <i>R package version 2.0-3</i> .
782	Meirmans PG, Hedrick PW. 2011. Assessing population structure: F-ST and related
783	measures. Molecular Ecology Resources 11: 5-18.
784	Meirmans PG, Van Tienderen PH. 2004. GenoType amd GenoDive: two programs for the
785	analysis of genetic diversity of asexual organisms Molecular Ecology Notes 4: 792-
786	794.
787	Miller MP, Mullins TD, Parrish JW, Jr., Walters JR, Haig SM. 2012. Variation in
788	migratory behavior influences regional genetic diversity and structure among
789	american kestrel populations (Falco sparverius) in North America. Journal of
790	<i>Heredity</i> 103: 503-514.

2 3 4	791	Musiani M, Leonard JA, Cluff D, Gates C, Mariani S, Paquet PC, Vilà C, Wayne RK.
5	792	2007. Differentiation of tundra/taiga and boreal coniferous forest wolves: genetics,
7 8	793	coat colour and association with migratory caribou. <i>Molecular Ecology</i> 16.
9 10	794	Nathan R, Perry G, Cronin JT, Strand AE, Cain ML. 2003. Methods for estimating long-
11 12	795	distance dispersal. Oikos 103: 261-273.
13 14 15	796	Newton I. 2011. Obligate and facultative migration in birds: ecological aspects. Journal of
16 17	797	Ornithology.
18 19	798	Nilsson ALK, Alerstam T, Nilsson JA. 2008. Diffuse, short and slow migration among blue
20 21	799	tits. Journal of Ornithology 149: 365-373.
22 23 24	800	Nilsson ALK, Linstrom A, Jonzen N, Nilsson SG, Karlsson L. 2006. The effect of climate
24 25 26	801	change on partial migration - the blue tit paradox. Global Change Biology 12: 2014-
27 28	802	2022.
29 30	803	Nowakowski JK, Vähätalo A. 2003. Is the great tit Parus major an irruptive migrant in
31 32	804	North-east Europe? Ardea 91: 231-244.
33 34 35	805	Oksanen J, Blanchet FG, Kindt R, Legendre P, O'Hara RB, Simpson GL, Solymos P, H.
36 37	806	Stevens MH, Wagner H. 2011. vegan: Community Ecology Package R package
38 39	807	version 1.17-6: <u>http://CRAN.R-project.org/package=vegan</u> .
40 41	808	Orell M, Lahti K, Koivula K, Rytkonen S, Welling P. 1999. Immigration and gene flow in
42 43 44	809	a northern willow tit (Parus montanus) population. Journal of Evolutionary Biology
44 45 46	810	12: 283-295.
47 48	811	Päckert M, Martens J, Sun Y-H. 2010. Phylogeny of long-tailed tits and allies inferred from
49 50	812	mitochondrial and nuclear markers (Ayes: Passeriformes, Aegithalidae). Molecular
51 52	813	Phylogenetics and Evolution 55: 952-967.
53 54 55	814	Parn H, Ringsby TH, Jensen H, Saether B-E. 2012. Spatial heterogeneity in the effects of
56 57	815	climate and density-dependence on dispersal in a house sparrow metapopulation.
58 59 60	816	Proceedings of the Royal Society B-Biological Sciences 279: 144-152.

	Biological Journal of the Linnean Society
817	Pavlova A, Rohwer S, Drovetski SV, Zink RM. 2006. Different post-pleistocene histories
818	of Eurasian parids. Journal of Heredity 97: 389-402.
819	Peakall R, Smouse PE. 2006. GENALEX 6: genetic analysis in Excel. Population genetic
820	software for teaching and research. Molecular Ecology Notes 6: 288-295.
821	Pentzold S, Tritsch C, Martens J, Tietze DT, Giacalone G, Lo Valvo M, Nazarenko AA,
822	Kvist L, Paeckert M. 2013. Where is the line? Phylogeography and secondary
823	contact of western Palearctic coal tits (Periparus ater: Aves, Passeriformes, Paridae).
824	Zoologischer Anzeiger 252: 367-382.
825	Peterson MA, Denno RF. 1998. The influence of dispersal and diet breadth on patterns of
826	genetic isolation by distance in phytophagous insects. The American Naturalist 152:
827	428-446.
828	Pilot M, Jędrezejewski W, Branicki W, Sidorovich VE, Jędrezejewska B, Stachura K,
329	Funk SM. 2006. Ecological factors influence population genetic structure of
830	European grey wolves. <i>Molecular Ecology</i> 15: 4533-4553.
31	Poluda AM. 2011. Spatial-Temporal Characteristics of Migrations of the Great Tits Parus
32	major (Aves, Passeriformes, Paridae) in Ukraine. Vestnik Zoologii 45: 343-357.
33	Postma E, Den Tex R-J, Van noordwijk AJ, Mateman AC. 2009. Neutral markers mirror
334	small-scale quantitative genetic differentiation in an avian island population.
335	Biological Journal of the Linnean Society 97: 867-875.
836	Postma E, van Noordwijk AJ. 2005. Gene flow maintains a large genetic difference in
837	clutch size at a small spatial scale. <i>Nature</i> 433: 65-68.
838	Pritchard JK, Stephens M, Donnelly P. 2000. Inference of population structure using
839	multilocus genotype data. Genetics 155: 945-959.
840	Prochazka P, Stokke BG, Jensen H, Fainova D, Bellinvia E, Fossoy F, Vikan JR, Bryja
341	J, Soler M. 2011. Low genetic differentiation among reed warbler Acrocephalus
842	scirpaceus populations across Europe. Journal of Avian Biology 42: 103-113.
	Biological Journal of the Linnean Society

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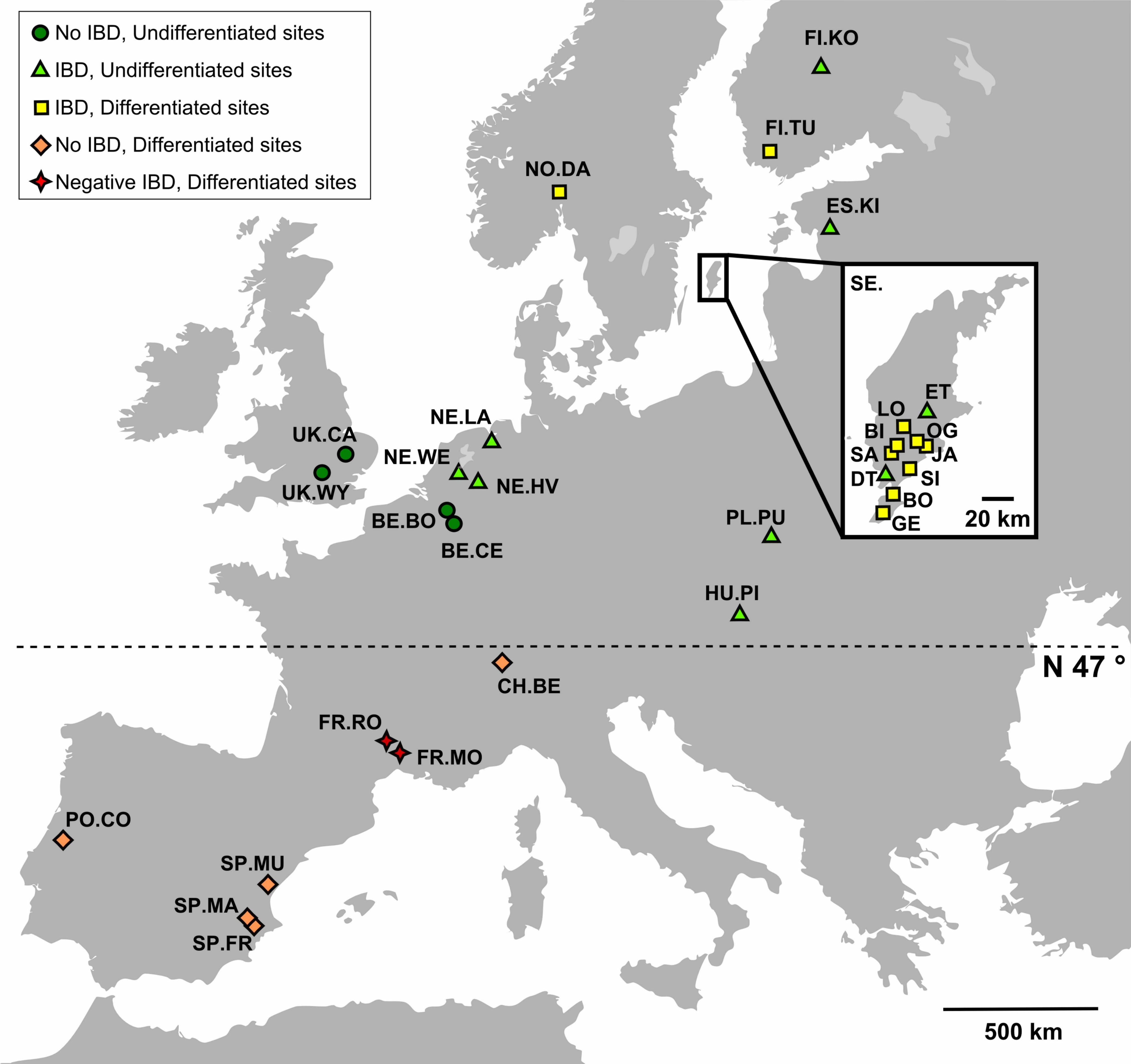
Purcell JFH, Cowen RK, Hughes CR, Williams DA. 2006. Weak genetic structure
indicates strong dispersal limits: a tale of two coral reef fish. *Proceedings of the Royal Society B-Biological Sciences* 273: 1483-1490.

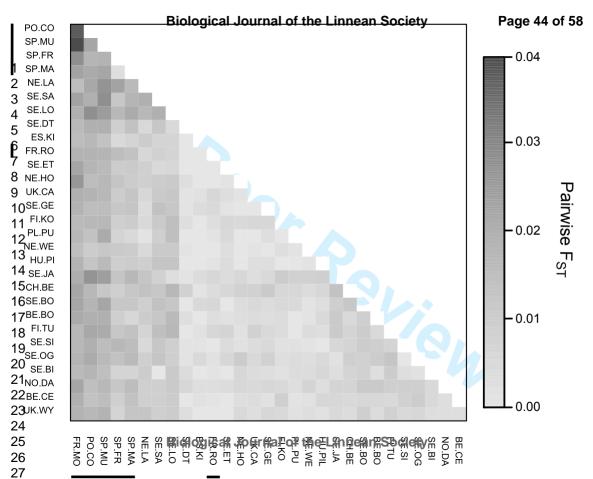
R Core Team. 2013. R: A language and environment for statistical computing. R Foundation for Statistical Computing. Vienna, Austria: R Foundation for Statistical Computing.

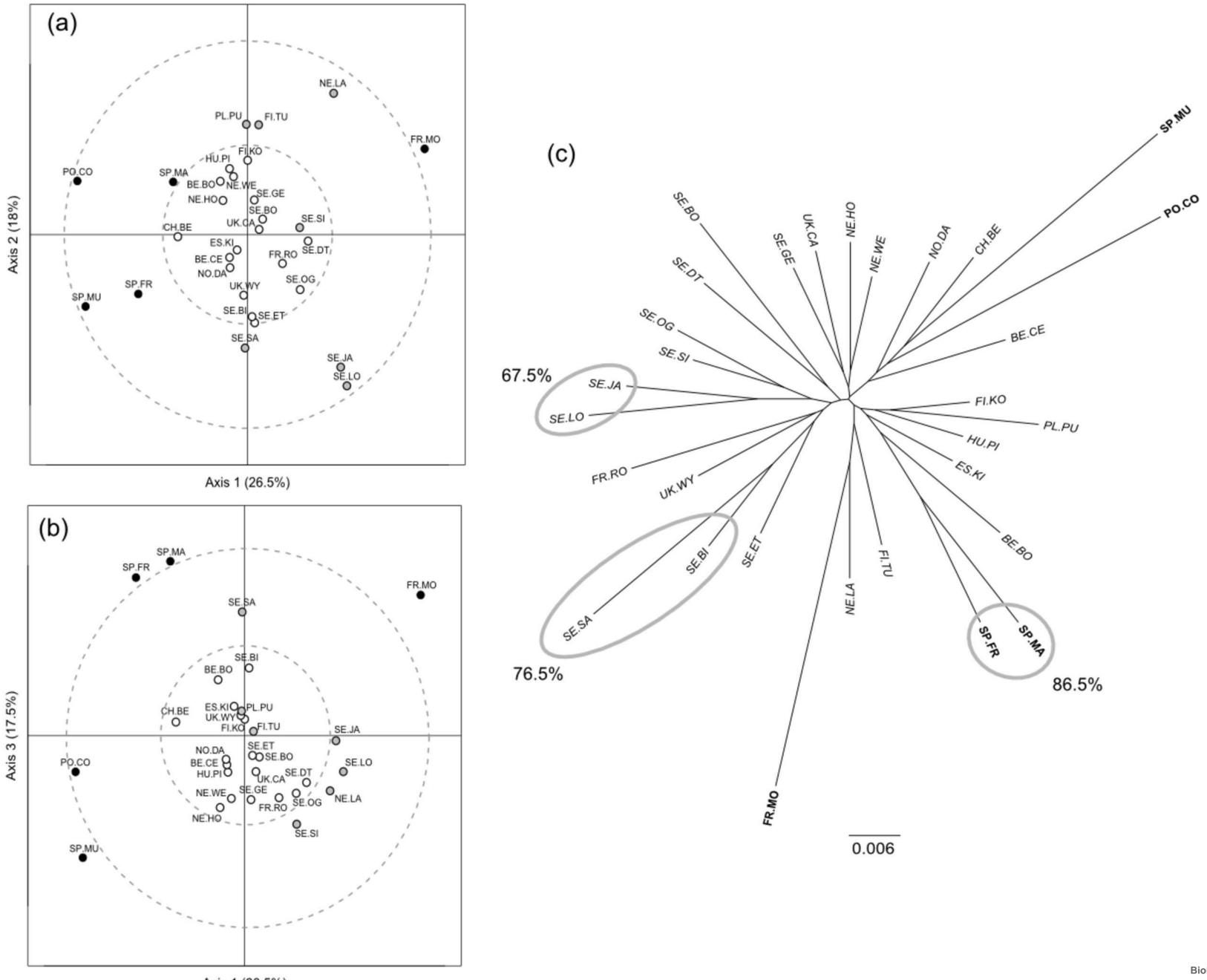
Rice W. 1989. Analyzing tables of statistical tests. *Evolution* **43:** 223-225.

- Rousset F. 2008. GENEPOP ' 007: a complete re-implementation of the GENEPOP software
 for Windows and Linux. *Molecular Ecology Resources* 8: 103-106.
 - Saladin V, Richner H. 2012. A set of 48 microsatellite loci for the great tit *Parus major*including 15 novel markers. *Molecular Ecology Resources* 12: 185-189.
 - 853 Sanford E, Kelly MW. 2011. Local Adaptation in Marine Invertebrates. *Annual Review of*854 *Marine science* 3: 509-535.
 - 855 Savolainen O, Pyhäjärvi T, Knürr T. 2007. Gene flow and local adaptation in Trees.
 856 Annual Review of Ecology, Evolution and Systematics 38: 595-619.
 - 857 Slatkin M. 1987. Gene flow and the geographic structure of natural populations. *Science* 236:
 858 787-792.
- Snoeijs T, Van de Casteele T, Adriaensen F, Matthysen E, Eens M. 2004. A strong
 association between immune responsiveness and natal dispersal in a songbird. *Proceedings of Royal Society of London Serie B: Biological sciences* 271: S199-S201.
- 862 Snow DW, Perrins CM. 1998. *The Birds of the Western Palearctic*. Oxford University Press:
 863 Oxford, New York.
- Stenseth NC, Durant JM, Fowler MS, Matthysen E, Adriaensen F, Jonzen N, Chan K-S,
 Liu H, De Laet J, Sheldon BC, Visser ME, Dhondt AA. 2015. Testing for effects of
 climate change on competitive relationships and coexistence between two bird
 species. *Proceedings of the Royal Society B-Biological Sciences* 282.

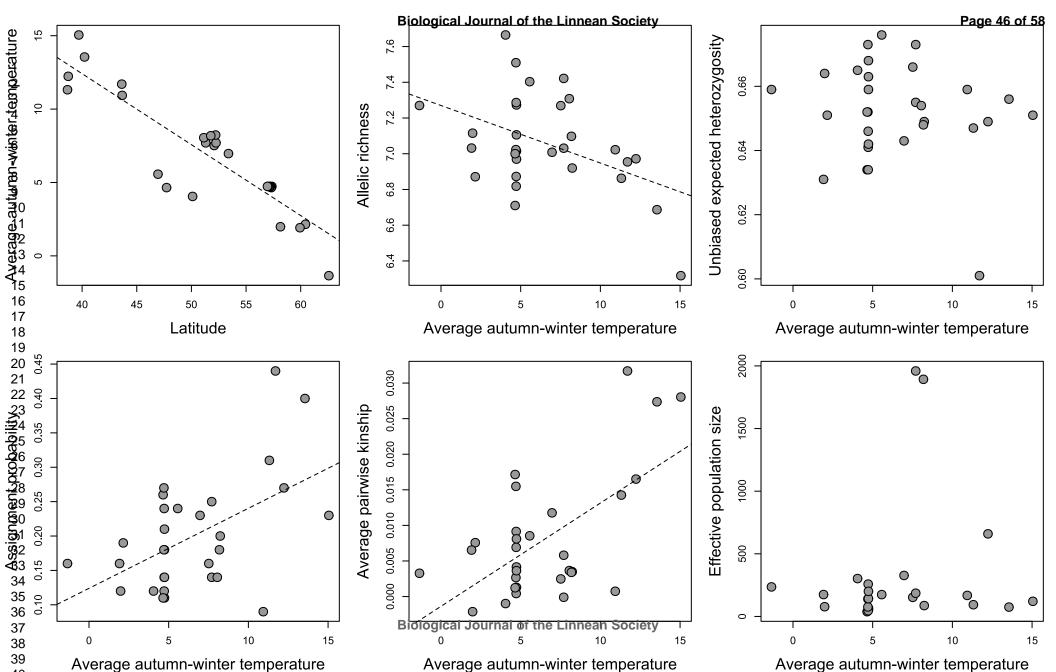
868	Travis JMJ, Dytham C. 1999. Habitat persistence, habitat availability and the evolution of
869	dispersal. Proceedings of Royal Society of London Serie B: Biological sciences 266:
870	723-728.
871	Van Bers NEM, Santure AW, Van Oers K, De Cauwer I, Dibbits BW, Mateman C,
872	Crooijmans RPMA, Sheldon BC, Visser ME, Groenen MAM, Slate J. 2012. The
873	design and cross-population application of a genome-wide SNP chip for the great tit
874	Parus major. Molecular Ecology Resources 12: 753-770.
875	Van Doorslaer W, Vanoverbeke J, Duvivier C, Rousseaux S, Jansen M, Jansen B,
876	Feuchtmayr H, Atkinson D, Moss B, Stoks R, De Meester L. 2009. Local
877	adaptation to higher temperatures reduces immigration success of genotypes from a
878	warmer region in the water flea Daphnia. Global Change Biology 15: 3046-3055.
879	Van Oosterhout C, Hutchinson WF, Wills DPM, Shipley P. 2004. MICRO-CHECKER:
880	software for identifying and correcting genotyping errors in microsatellite data.
881	Molecular Ecology Notes 4: 535-538.
882	Van Vliet J, Musters CJM, Ter Keurs W. 2009. Changes in migration behaviour of
883	blackbirds Turdus merula from the Netherlands. Bird study 56: 276-281.
884	Vel'ky M, Kanuch P, Kristin A. 2011. Food composition of wintering great tits (Parus
885	major): habitat and seasonal aspects. Folia Zoologica 60: 228-236.
886	Verhulst S, Perrins CM, Riddington R. 1997. Natal dispersal of great tits in a patchy
887	environment. Ecology 78: 864-872.
888	Visser ME. 2008. Keeping up with a warming world; assessing the rate of adaptation to
889	climate change. Proceedings of the Royal Society B-Biological Sciences 275: 649-659.
890	Wang IJ. 2010. Recognizing the temporal distinctions between landscape genetics and
891	phylogeography. Molecular Ecology 19: 2605-2608.
892	

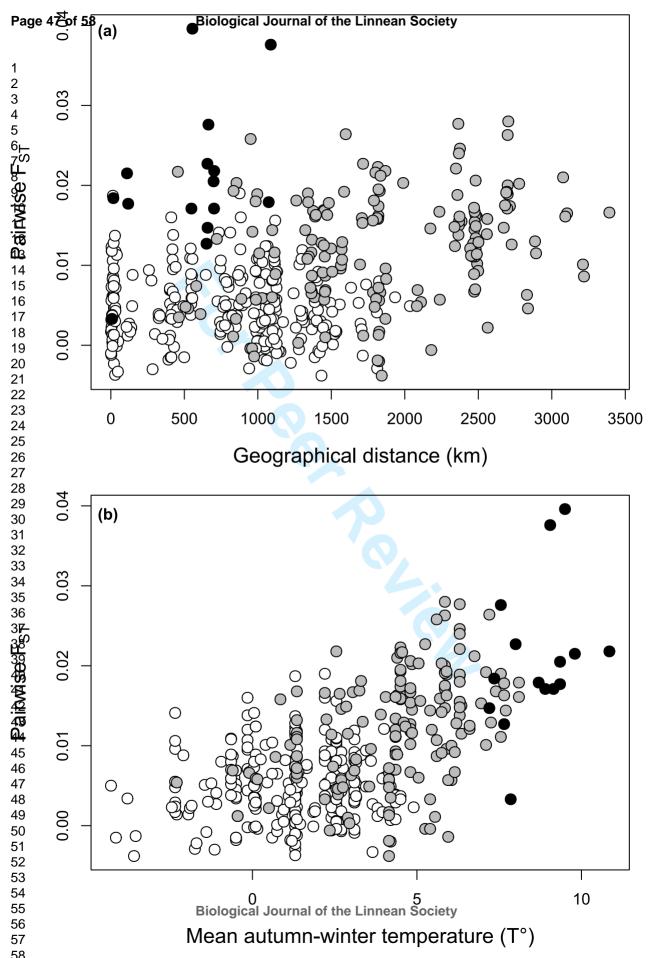






Axis 1 (26.5%)





Low but contrasting neutral genetic differentiation shaped by winter temperature in European great tits

Supplementary Information by Lemoine et al.

List of supplementary informations

A. Current knowledge about the geographical repartition of mtDNA haplotypes in European great tits

Figure S1. Haplotype network based on the mitochondrial control region (578 bp) for 15 sites in Europe.

Table S1. Origin of the 103 sequences from Genbank of the mitochondrial control region.

B. Description of microsatellite loci, sampling sites and genetic diversity

 Table S2. Description of sampling sites and average genetic diversity indices per site.

Table S3. Characteristics of microsatellite loci developed on individuals from CH.BE

C. Testing for a potential bias due to the inclusion of 10 close-by sites from Gotland

Figure S2. Principal coordinate analysis (PCoA) contrasting axes 1 vs. 2 when (a) 30 populations and (b) 21 populations are included.

Figure S3. Principal coordinate analysis (PCoA) contrasting axes 1 vs. 3 when (a) 30 populations (a) and (b) 21 populations are included.

Figure S4. (a) Mean (\pm SD) of estimated posterior likelihood and (b) estimation of ΔK over 10 STRUCTURE runs for successive K values when 30 (i.e. all) populations are included in the analysis.

Figure S5. (a) Mean (\pm SD) of estimated posterior likelihood and (b) estimation of ΔK over 10 STRUCTURE runs for successive K values when 21 (i.e. only 1 population from Gotland is included) populations are included in the analysis.

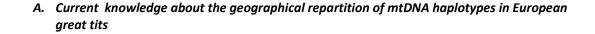
Figure S6. Assignment plots for K = 3 based on a sampling including a) 30 populations and b) 21 populations.

Table S4. Probability assignments of STRUCTURE to cluster 1 and 2 by sites for K = 3 when 21 and 30 sites are included in the analysis.

D. The effect of environmental factors on population indices

Table S5. Comparison of models testing the effect of environmental factors on indices of genetic diversity per site and other parameters

E. References



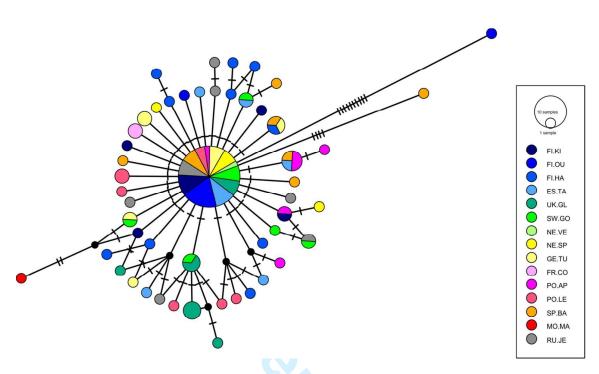


Fig S1. Haplotype network based on 103 available sequences from Genbank of the mitochondrial control region (578 bp) for 15 sites in Europe performed with the software POPART (Leigh and Bryant, 2015). Colours are indicative of the locations where the haplotypes were found. The circle sizes are proportional to the haplotype frequency.

Table S2. Origin of the 103 se	equences from Genbar	nk of the mitochondria	ll control region (578 bp)

Country, Site	Abb.	N _{Ind.}	N _{Haplot.}
Finland, Kilpisjärvi	FI.KI	8	5
Finland, Oulu	FI.OU	9	3
Finland, Harjavalta	FI.HA	10	10
Estonia, Tartu	ES.TA	9	6
United Kingdom	UK.GL	10	5
Sweden, Gotland	SW.GO	8	6
Netherlands, Veluwe	NE.VE	1	1
Netherlands, Speuld	NE.SP	5	3
Germany, Tübingen	GE.TU	8	5
France, Corsica	FR.CO	2	1
Portugal, Apostica	PO.AP	6	5
Portugal, Leiria	PO.LE	8	6
Spain, Barcelona	SP.BA	9	7
Morocco, Marrakesch	MO.MA	1	1
Russia, Jekaterinburg	RU.JE	9	7

Abb. is the abbreviation used in the haplotype network; $N_{Ind.}$ Number of individuals sampled by site and N_{Haplot} : Number of haplotypes sampled by site.

B. Description of microsatellite loci, sampling sites and genetic diversity

 Table S2. Description of sampling sites and average genetic diversity indices per site.

Sampling sites with country of origin and geographical name of the site: Abb.: abbreviation names used in the text; Veg.: dominant vegetation: P = coniferous trees, D = deciduous trees or O = orange tree, - = unavailable information; T°Dec: mean daily winter temperature in December (in °C); Sea: shortest distance to the sea (in km); Year: year of sampling; n, number of adults sampled; AR: mean allelic richness based on 18 individuals; AE: number of effective alleles; AP: number of private alleles; HO: observed heterozygosity; HE: unbiased expected heterozygosity; FIS: inbreeding coefficient and PA: probability of assignment of individuals to their original sampling site.

	Country, Site	Abb.	Latitude (N)	Longitude (E)	Veg.	T° _{Dec}	Sea	Year	n	A _R	A _E	A _P	Ho	Η _E	F _{IS}	PA
1	Finland, Konnesi	FI.KO	62°34'36.01	26°20'38.00	Р	-6	214	2009	47	7.27	4.86	0.27	0.654	0.659	0.008	0.1
2	Estonia, Kilingi Nõmme	ES.KI	58°08'47.25	24°57'06.45	М	-2.3	27	2007	31	7.12	4.66	0	0.655	0.664	0.012	0.1
3	Finland, Turku	FI.TU	60°26'04.13	22°10'21.36	Р	-1.6	65	2009	40	6.87	4.44	0.09	0.634	0.651	0.026	0.1
4	Sweden, Etelhem	SE.ET	57°21'30.48	18°31'57.54	D	1.3	10	2007/2008	26	7.01	4.59	0.09	0.645	0.668	0.035	0.1
5	Sweden, Lojsta	SE.LO	57°18'40.30	18°22'42.07	D	1.3	13	2008	32	6.71	4.06	0.14	0.619	0.634	0.024	0.2
6	Sweden, Sigdes	SE.SI	57°11'00.51	18°26'42.40	D	1.4	2	2007/2008	33	7.27	4.76	0.05	0.696	0.663	-0.049	0.1
7	Sweden, Oggesanget	SE.OG	57°12'26.71	18°25'19.06	D	1.3	5	2007/2008	33	6.87	4.37	0.05	0.627	0.646	0.03	0.1
8	Sweden, Jaksarve	SE.JA	57°12'33.05	18°24'41.25	D	1.3	6	2007/2008	27	7.29	4.5	0.14	0.625	0.641	0.026	0.2
9	Sweden, Binge	SE.BI	57°11'30.11	18°22'01.41	D	1.3	6	2007/2008	24	7.51	4.92	0.18	0.689	0.673	-0.026	0.2
10	Sweden, Södra Alva	SE.SA	57°11'23.69	18°20'33.99	Р	1.3	7	2007/2008	21	7.02	4.53	0.05	0.628	0.652	0.038	0.2
11	Sweden, Drive Through	SE.DT	57°07'43.75	18°18'52.02	Р	1.3	5	2007/2008	18	6.82	4.17	0	0.629	0.634	0.009	0.1
12	Sweden, Botarve	SE.BO	57°01'26.45	18°17'00.34	D	1.3	2	2007/2008	21	7.11	4.38	0	0.634	0.642	0.012	0.
13	Sweden, Gervalds	SE.GE	56°57'29.27	18°10'18.57	D	1.3	0.5	2008	32	6.97	4.47	0.09	0.658	0.659	0.003	0.
14	Poland, Puszcza	PL.PU	50°06'34.70	20°25'24.89	D	-1.2	471	2009	20	7.66	4.89	0.09	0.689	0.665	-0.037	0.
15	Hungrary, Pilis Mountains	HU.PI	47°43'44.00	19°00'40.00	D	-1.1	434	2010	35	7	4.65	0.09	0.656	0.652	-0.006	0.
16	Norway, Daeli	NO.DA	59°55'59.99	10°33'00.00	D	-2.6	5	2007	32	7.03	4.78	0.09	0.631	0.631	0.001	0.
17	Netherlands, Lauwersmeer	NE.LA	53°23'26.20	06°13'34.19	D	3.1	10	2008	39	7.01	4.82	0.09	0.656	0.643	-0.021	0.
18	Netherlands, Hoge Veluwe	NE.HO	52°04'59.99	05°49'60.00	D	3.5	9	2008	27	7.27	4.58	0.23	0.65	0.666	0.025	0.
19	Netherlands, Westerheide	NE.WE	52°15'08.78	05°11'14.29	D	3.8	5	2009	30	7.03	4.51	0.05	0.635	0.655	0.031	0.
20	Belgium, Brasschaat	BE.BO	51°18'29.41	04°31'42.91	D	3.8	64	2008	30	7.42	4.6	0.32	0.647	0.673	0.04	0.
21	Belgium, Boshoek	BE.CE	51°07'59.99	04°31'60.00	D	4.1	72	2007	30	7.31	4.63	0.14	0.624	0.654	0.046	0.
22	Switzerland, Bern	CH.BE	46°56'32.57	07°18'46.39	Р	1.4	306	2009	40	7.4	4.86	0.27	0.698	0.676	-0.032	0.
23	United Kingdom, Cambridge	UK.CA	52°13'01.06	00°02'58.87	D	4.9	65	2009	29	6.92	4.34	0.09	0.636	0.649	0.019	0
24	United Kingdom, Wytham	UK.WY	51°46'23.02	-01°20'18.00	D	4.9	95	2008	29	7.1	4.35	0.14	0.639	0.648	0.013	0.
25	France, Montpellier	FR.MO	43°37'44.64	-03°52'06.93	D	7.7	12	2010	18	6.95	3.94	0.05	0.573	0.601	0.047	0.
26	France, Rouvière	FR.RO	43°40'00.01	-03°40'00.00	D	7	26	2010	19	7.02	4.44	0.05	0.641	0.659	0.027	0.
27	Spain, Sagunto	SP.MU	39°42'01.04	-00°15'00.00	0	11.3	6	2007	25	6.32	4.22	0.05	0.655	0.651	-0.006	0.
28	Spain, Sierra Mariola	SP.MA	38°43'59.99	-00°33'00.00	Р	8.3	36	2005/2006	30	6.97	4.47	0.05	0.647	0.649	0.003	0.
29	Spain, Font Roja	SP.FR	38°39'45.18	-00°32'45.56	D	7.4	30	2005/2006	30	6.86	4.4	0.05	0.638	0.647	0.014	0.
30	Portugal, Coimbra	PO.CO	40°13'19.78	-08°27'11.64	D	10.4	35	2006	36	6.69	4.49	0.14	0.662	0.656	-0.009	C

 Table S3. Characteristics of microsatellite loci developed on individuals from CH.BE: Locus name, repeat type and motif, species for which a locus has been described initially, as well as intra site variation in the number of alleles found, the smallest allele size and number of sites with null alleles at a specific locus.

Locus	Repeat type	Repeat motif	Original species Described in			Smallest allele size (bp)	Presence of null alleles	
PmaC25	trinucleotide	(CAT)11	Parus major	Saladin et al 2003	8-14	312	3	
PmaD105	tetranucleotide	(GTCT)3 (ATCT)12	Parus major	Saladin et al 2003	5-10	376	0	
PmaD22	tetranucleotide	(CTAT)15 (CCAT)12	Parus major	Saladin et al 2003	10-17	370	2	
PmaGAn27	trinucleotide	(CAT)16	Parus major	Saladin et al 2003	11-20	186	0	
PmaGAn30	dinucleotide	(GA)10	Parus major	Saladin et al 2003	4-6	304	0	
PmaTAGAn71	tetranucleotide	(TAGG)6(TAGA)11	Parus major	Saladin et al 2003	6-10	156	0	
PmaTAGAn86	tetranucleotide	(TAGA)21	Parus major	Saladin et al 2003	9-15	139	2	
PmaTGAn33	tetranucleotide	(GATA)14GAT(GATA)8	Parus major	Saladin et al 2003	13-22	254	1	
PmaTGAn42	tetranucleotide	(TCCA)15	Parus major	Saladin et al 2003	5-9	248	1	
PmaTGAn45	trinucleotide	(TGA)10	Parus major	Saladin et al 2003	5-9	296	1	
PmaCAn2	dinucleotide	(CA)16	Parus major	Saladin et al 2012	8-20	106	0	
PmaGAn11	dinucleotide	(GA)9	Parus major	Saladin et al 2003	2-5	102	0	
PmaGAn13	dinucleotide	(GA)11(CA)14	Parus major	Saladin et al 2012	5-9	404	1	
PmaGAn22	dinucleotide	(GA)15GG(GA)11	Parus major	Saladin et al 2012	3-8	347	1	
PmaGAn31	dinucleotide	(GA)13	Parus major	Saladin et al 2003	3-6	83	1	
PmaGAn35	dinucleotide	(GA)19	Parus major	Saladin et al 2012	6-13	178	1	
PmaGAn40	dinucleotide	(GA)10	Parus major	Saladin et al 2003	3-6	410	0	
PmaTAGAn89	tetranucleotide	(TAGA)3TGA(TAGA)13	Parus major	Saladin et al 2012	7-13	200	3	
TG01-124	dinucleotide	(AT)11	Taeniopygia guttata	Dawson et al 2010	1-4	398	1	
TG05-046	dinucleotide	(AT)8 (A)4 (AT)6 (A)9 (AT)2	Taeniopygia guttata	Dawson et al 2010	1-2	340	0	
Tgu07	dinucleotide	(AC)14, in total 44xAC	Taeniopygia guttata	Slate et al 2007	4-8	93	1	
TG08-024	dinucleotide	(AT)4 AG (AT)2 AA (AT)3 AA (AT)5	Taeniopygia guttata	Dawson et al 2010	2-4	241	0	

C. Testing for a potential bias due to the inclusion of 10 close-by sites from Gotland

To test for a potential bias due to the inclusion of 10 close-by sites from Gotland, the PCoA and Structure analyses were run once using individuals from all 30 sites and once using individuals from 21 sites including only a single site from Gotland (SE.OG).

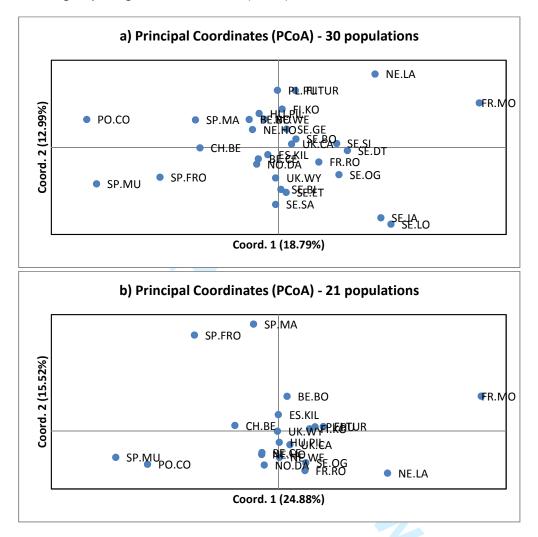


Figure S2. Principal coordinate analysis (PCoA) contrasting axes 1 vs. 2 when (a) 30 populations and (b) 21 populations are included.

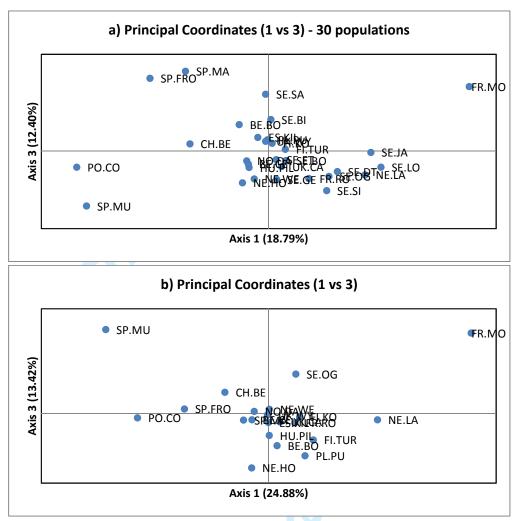


Figure S3. Principal coordinate analysis (PCoA) contrasting axes 1 vs. 3 when (a) 30 populations (a) and (b) 21 populations are included.

Figure S4. (a) Mean (\pm SD) of estimated posterior likelihood and (b) estimation of ΔK over 10 STRUCTURE runs for successive K values when 30 (i.e. all) populations are included in the analysis.

Figure S5. (a) Mean (\pm SD) of estimated posterior likelihood and (b) estimation of Δ K over 10 STRUCTURE runs for successive K values when 21 (i.e. only 1 population from Gotland is included) populations are included in the analysis.

For both analyses, the best K identified using 10 STRUCTURE runs was K = 3.

Figure S6. Assignment plots for K = 3 based on a sampling including a) 30 populations and b) 21 populations (i.e. only one out of 10 populations from Gotland). Populations are indicated in Figure 1.

Table S4. Probability assignments of STRUCTURE to cluster 1 and 2 by sites for K = 3 when 21 and 30 sites are included in the analysis. Populations are indicated in Figure 1.

-	-		-				
	21 9	sites	30 sites				
Site	Cluster1	Cluster2	Cluster1	Cluster2			
BE_BO	0.83	0.05	0.83	0.04			
BE_CE	0.73	0.16	0.76	0.17			
ES_KI	0.84	0.07	0.87	0.05			
FI_KO	0.78	0.12	0.76	0.11			
FI_TU	0.91	0.05	0.92	0.04			
FR_MO	0.93	0.05	0.94	0.02			
FR_RO	0.88	0.11	0.93	0.06			
HU_PI	0.80	0.14	0.80	0.13			
NE_HO	0.76	0.19	0.80	0.15			
NE_LA	0.84	0.14	0.88	0.09			
NE_WE	0.77	0.17	0.76	0.18			
NO_DA	0.74	0.24	0.75	0.24			
PD_PU	0.78	0.15	0.77	0.15			
PO_CO	0.21	0.78	0.21	0.77			
SE_OG	0.89	0.06	0.95	0.03			
SP_FR	0.44	0.03	0.45	0.01			
SP_MA	0.47	0.02	0.43	0.02			
SP_MU	0.47	0.15	0.46	0.25			
UK_CA	0.89	0.08	0.94	0.04			
UK_WY	0.84	0.13	0.89	0.09			
CH_BE	0.66	0.13	0.62	0.31			
SE_BI	-	-	0.84	0.02			
SE_BO	-	-	0.82	0.05			
SE_DT	-	-	0.91	0.05			
SE_ET	-	-	0.89	0.06			
SE_GE	-	-	0.88	0.05			
SE_JA	-	-	0.97	0.02			
SE_LO	-	-	0.94	0.05			
SE_SA	-	-	0.85	0.01			
SE_SI	-	-	0.91	0.04			

The correlations between probability assignments to cluster 1 when 21 and 30 sites are included approximated both 0.99 with and without CH.BE. The correlations between probability assignments to cluster 2 when 21 and 30 sites are included approximated 0.98 and 0.95 respectively without and with CH.BE.

Overall, the results of the analyses including 30 or 21 sites did not qualitatively differ.

D. The effect of environmental factors on population indices

Table S5. Comparison of models testing the effect of environmental factors on indices of genetic diversity per site and other parameters (A_R : allelic richness, H_E : unbiased expected heterozygosity, P_A : assignment probability, Kinship and N_e : effective population size). Best models (i.e. with the smallest AIC or with a Δ AIC lower than 2) are indicated in bold. Temperatures are the average daily temperature for each month.

Explanatory variable	A _R			HE		P _A		Kinship			N _e				
	AIC	ΔΑΙΟ	wi	AIC	Δ AIC	wi	AIC	Δ AIC	wi	AIC	Δ AIC	wi	AIC	Δ AIC	wi
Latitude	9.68	5.05	0.01	-160.69	2.38	0.04	-66.21	2.69	0.04	-203.13	6.63	0.01	402.36	28.52	0.00
Longitude	9.59	4.95	0.01	-160.96	2.11	0.04	-66.61	2.3	0.05	-200.43	9.34	0.00	399.84	26	0.00
Temperature in January	5.44	0.8	0.09	-161.5	1.55	0.05	-68.27	0.64	0.11	-206.67	3.09	0.05	401.74	27.9	0.00
Temperature in February	7.78	3.15	0.03	-161.1	1.95	0.04	-66.75	2.15	0.05	-204.06	5.71	0.01	401.43	27.6	0.00
Temperature in March	8.53	3.9	0.02	-161	2.06	0.04	-66.68	2.22	0.05	-204.04	5.72	0.01	401.58	27.75	0.00
Temperature in April	9.13	4.5	0.01	-160.88	2.18	0.04	-64.69	4.22	0.02	-202.63	7.13	0.01	401.86	28.03	0.00
Temperature in May	10.19	5.56	0.01	-160.82	2.24	0.04	-63.45	5.45	0.01	-201.37	8.39	0.00	402.17	28.34	0.00
Temperature in June	8.09	3.45	0.02	-161.3	1.78	0.05	-65.08	3.82	0.02	-205.18	4.58	0.02	402.6	28.76	0.00
Temperature in July	6.52	1.88	0.05	-162	1.12	0.07	-65.51	3.4	0.03	-207.22	2.54	0.07	402.54	28.7	0.00
Temperature in August	4.98	0.35	0.11	-162.2	0.91	0.08	-66.66	2.25	0.05	-209.5	0.3	0.21	402.58	28.75	0.00
Temperature in September	4.63	0	0.13	-161.7	1.41	0.06	-67.84	1.06	0.09	-209.8	0	0.25	402.5	28.67	0.00
Temperature in October	6.48	1.84	0.05	-161.4	1.69	0.05	-67.63	1.27	0.08	-207.6	2.16	0.08	402.28	28.44	0.00
Temperature in November	5.71	1.08	0.08	-161.4	1.7	0.05	-67.75	1.15	0.09	-207.24	2.53	0.07	401.96	28.13	0.00
Temperature in December	4.89	0.25	0.12	-161.5	1.6	0.05	-68.9	0	0.15	-207.9	1.9	0.09	402.06	28.22	0.00
Variance of temperature	10.4	5.76	0.01	-160.69	2.38	0.04	-62.89	6.01	0.01	-196.68	13.08	0.00	400	26.16	0.00
Difference of temperature	10.13	5.5	0.01	-160.72	2.35	0.04	-62.71	6.19	0.01	-196.48	13.29	0.00	399.84	26	0.00
Mean Spring Summer temperature	7.95	3.31	0.03	-161.3	1.82	0.05	-65.88	3.02	0.03	-205.45	4.31	0.03	402.37	28.54	0.00
Mean Autumn Winter temperature	5.82	1.18	0.07	-161.4	1.66	0.05	-68.07	0.84	0.1	-207.29	2.48	0.07	402	28.16	0.0
Vegetation	5.12	0.49	0.10	-148.13	14.93	0.00	-53.2	15.71	0.00	-184.97	24.79	0.00	373.83	0	1.0
Distance to sea	6.72	2.09	0.05	-163.1	0.00	0.12	-59.98	8.92	0.00	-195.18	14.58	0.00	402.54	28.71	0.0

E. References

Dawson, D.A., Horsburgh, G.J., Kupper, C., Stewart, I.R.K., Ball, A.D., Durrant, K.L., Hansson, B., Bacon, I., Bird, S., Klein, A., Krupa, A.P., Lee, J.-W., Martin-Galvez, D., Simeoni, M., Smith, G., Spurgin, L.G., Burke, T., 2010. New methods to identify conserved microsatellite loci and develop primer sets of high cross-species utility - as demonstrated for birds. Mol. Ecol. Resour. 10, 475-494.

Leigh, J.W., Bryant, D., 2015. popart: full-feature software for haplotype network construction. Methods in Ecology and Evolution 6, 1110-1116.

Saladin, V., Bonfils, D., Binz, T., Richner, H., 2003. Isolation and charcterization of 16 microsatellite loci in the Great Tit Parus major. Molecular Ecology Notes 3, 520-522.

Saladin, V., Richner, H., 2012. A set of 48 microsatellite loci for the great tit Parus major including 15 novel markers. Mol. Ecol. Resour. 12, 185-189.

2, 1&.. , 2007. Simple S , s: a new resource for C Slate, J., Hale, M.C., Birkhead, T.R., 2007. Simple sequence repeats in zebra finch (Taeniopygia guttata) expressed sequence tags: a new resource for evolutionary genetic studies of passerines. Bmc Genomics 8.