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pristine and polluted  
Amazon  
environments

J. G. Levine et al.

This discussion paper is/has been under review for the journal Atmospheric Chemistry and Physics (ACP). Please refer to the corresponding final paper in ACP if available.

# Isoprene chemistry in pristine and polluted Amazon environments: Eulerian and Lagrangian model frameworks and the strong bearing they have on our understanding of surface ozone and predictions of rainforest exposure to this priority pollutant

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– a very strong source of isoprene – that could not be reproduced with then-current atmospheric chemistry models. The incompatibility of measured and modelled OH and isoprene was further demonstrated for the Borneo rainforest by Pugh et al. (2010, 2011) and Stone et al. (2011). Whilst heterogeneous chemistry is not considered in this study, isoprene also interacts with air pollution and the climate system by acting as a precursor to secondary biogenic organic aerosol, with an aerosol yield of a few percent (e.g. Carlton et al., 2009; Chen et al., 2015).

The specific difficulty Lelieveld et al. (2008) and Butler et al. (2008) faced was explaining observations of simultaneously high isoprene- and OH concentrations, made during the Guyanas Atmosphere–Biosphere exchange and Radicals Intensive Experiment with a Learjet campaign of 2005 (GABRIEL; see Atmospheric Chemistry and Physics Special Issue 88). High isoprene emissions were expected to sustain high isoprene concentrations whilst suppressing OH concentrations (due to the rapid reaction between the two). It was the unexpectedly high OH concentrations that led Lelieveld et al. (2008) to speculate that the chemistry models were missing a mechanism by which some of the OH initially consumed in isoprene oxidation was “recycled”. Meanwhile, Butler et al. (2008) explored the role that the physical separation, or “segregation”, of air masses containing isoprene emissions could play in resolving the apparent paradox, as have Pugh et al. (2011) since; see later. Here, we confront two atmospheric chemistry models, and modelling frameworks (Eulerian and Lagrangian), with recent airborne measurements of atmospheric composition above the Amazon rainforest, made during the South American Biomass Burning Analysis campaign of 2012 (SAMBBA; see Darbyshire and Johnson, 2013). Building on the recent studies of Squire et al. (2014, 2015), we compare the abilities of these models subject to two chemical mechanisms – one including an OH “recycling” mechanism that should prove effective in pristine Amazon environments. In the Lagrangian model, motivated by studies such as Kuhn et al. (2010), we also explore the impact that the spatial resolution of trace gas emissions has on our ability to capture the atmospheric chemistry, and

hence the impact of ground-level O<sub>3</sub> on the health of the rainforest, in polluted plumes downwind of the city of Manaus.

Squire et al. (2014) explored the impacts that possible future changes in isoprene emissions – stemming from changes in atmospheric CO<sub>2</sub>, the physical climate (e.g. surface air temperatures), and anthropogenic land use – could have on tropospheric O<sub>3</sub>. They did so using a global Eulerian chemistry-climate model, the UK Met Office Unified Model (UM; Hewitt et al., 2011) coupled to the UK Chemistry and Aerosol model (UKCA; O'Connor et al., 2014), jointly referred to as UM-UKCA. We use the same Eulerian model here, building on much of Squire et al. (2014, 2015)'s work as outlined in the next section. Squire et al. (2014) employed isoprene emissions calculated using parameterisations based on the Model of Emissions of Gases and Aerosols from Nature (MEGAN; Guenther et al., 2006), with vegetation simulated offline using the Sheffield Dynamic Global Vegetation Model (SDGVM; Beerling et al., 1997; Beerling and Woodward, 2001) as described by Lathièrè et al. (2010). Before exploring the impact of changes in isoprene emissions, Squire et al. (2014) demonstrated that UM-UKCA showed some skill at reproducing recent observations of tropospheric O<sub>3</sub> when employing present day emissions: they compared their simulated profiles of O<sub>3</sub> with sonde profiles from the Southern Hemisphere ADditional OZonesondes network (SHADOZ; Thompson et al., 2003). Notably, however, this network did not offer measurements of tropospheric O<sub>3</sub> above the Amazon rainforest – globally, responsible for almost half of all biogenic NMVOC emissions (Guenther et al., 1995) and the greatest source of isoprene (see, e.g., Fig. 2 of Squire et al., 2014). Squire et al. (2015) then explored the sensitivity their projections of future tropospheric O<sub>3</sub> showed to the chemical mechanism they employed. However, they did not explore the impact of this mechanism on their ability to reproduce present day observations.

Here, we test the ability of (i) a nudged version of UM-UKCA and (ii) a Lagrangian model, the Cambridge Tropospheric Trajectory model of Chemistry And Transport (CiT-TyCAT; Pugh et al., 2012), to simulate SAMBBA measurements above the Amazon. In each model, we carry out (otherwise identical) integrations employing two of the four

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flown with the UK's Facility for Airborne Atmospheric Measurements' (FAAM) BAe-146 in September and October 2012 (the dry season and "biomass burning season"). We focus, however, on five flights aimed at characterising the "background" composition of the atmosphere in that region – the influence of biogenic emissions in the absence of biomass burning: B735, B744, B745, B749 and B750. Some portions of these flights still intercepted air influenced by biomass burning upwind, but considerably less than those flights specifically aiming to probe this influence. Chemically, we focus on measured concentrations of five species central to gas-phase tropospheric chemistry: O<sub>3</sub>, NO, NO<sub>2</sub>, isoprene (C<sub>5</sub>H<sub>8</sub>) and carbon monoxide (CO). We have no measurements of OH concentration with which to compare our simulated OH concentrations. However, the comparison of measured and modelled O<sub>3</sub> concentrations is nonetheless valuable in the context of isoprene oxidation and its impact on OH, since: OH is derived from O<sub>3</sub> (by O<sub>3</sub> photolysis and reaction of the resulting excited O(<sup>1</sup>D) oxygen atoms with water vapour); and, if a model were to reproduce measurements of OH concentration but not O<sub>3</sub> concentration, it would be simulating OH well but for the wrong reason(s).

### 2.1.1 SAMBBA flight tracks (and the histories of air parcels encountered)

The thick black lines in Fig. 1 illustrate the flight tracks in longitude and latitude of flights B735, B744, B745, B749 and B750. (The variation in altitude during each flight is illustrated in subsequent figures comparing measured and modelled trace gas concentrations.) Superimposed on each of these panels in Fig. 1, are 7 day back-trajectories describing the calculated routes by which the air sampled at points spaced one minute apart along each flight track reached those locations; the bottom right panel illustrates the 7 day back-trajectories calculated from the (arbitrary) site we explore downwind of Manaus; the "x" marks the site (61.0° W, 3.1° S) and the dot marks Manaus. The trajectories are coloured according to their pressures, providing some indication of where air parcels might have picked up emissions (e.g. at pressures greater than 800–900 hPa, corresponding to altitudes of less than roughly 1–2 km). The trajectory calculations are described later (Sect. 2.2.2) but, for the time being, we note two points: the divergence

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of air parcels backwards in time (i.e. with increasing distance from the flight tracks), which offers the Lagrangian model an opportunity to capture diversity in initial composition that may be lost or less well resolved in the Eulerian model; and the general consistency of colour along any one trajectory bound for a flight track, indicating that many air parcels have spent much of the 7 days prior to arrival on the flight track at approximately the same pressure. This implies that air parcels encountered in low-altitude portions of the flights have often spent much of the last 7 days at low altitudes and, thus, exposed to trace gas emissions (and dry deposition) in the boundary layer; air parcels encountered in high-altitude portions of the flights, on the other hand, have often spent much of the last 7 days well above the boundary layer and hence exempt from these influences. The trajectories bound for the site downwind of Manaus, meanwhile, appear to descend over the course of 7 days from origins in the mid troposphere to the site in the boundary layer (900 hPa) and are thereby predominantly exposed to trace gas emissions (and dry deposition) in the latter half of their journeys.

### 2.1.2 O<sub>3</sub> measurements

O<sub>3</sub> was measured at 0.1 Hz by a Thermo Environmental UV absorption photometer model 49C, traceable to the UK National Physical Laboratory primary ozone standard with an uncertainty of 2 %, and a precision of 1 ppb.

### 2.1.3 NO and NO<sub>2</sub> measurements

NO<sub>x</sub> was measured from the aircraft using a high sensitivity NO<sub>x</sub> chemiluminescence system built by Air Quality Design, Inc. The instrument has a dual channel architecture for independent quantification of NO and NO<sub>2</sub>, with each channel having a sample flow of 1 L min<sup>-1</sup>. NO is measured in one channel by an established chemiluminescence technique (Lee et al., 2009), with NO<sub>2</sub> quantified in a second channel by photolytic conversion to NO using blue light emitting diodes centred at 395 nm. The 395 nm wavelength has a specific affinity for NO<sub>2</sub> photolytic conversion to NO, giving high analyte

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selectivity within the channel. Work by Pollack et al. (2010) evaluated the relative high  $\text{NO}_2$  affinity for conversion of  $\text{NO}_2$  to  $\text{NO}$  using 395 nm blue light LEDs. They highlighted the low probability of other species within the gaseous chemical matrices such as nitrous acid (HONO), being affected by the 395 nm light, so in turn reducing possible non  $\text{NO}_2$  species interfering with the measurement.  $\text{NO}_x$  was then quantified by ozonation of the subsequent total  $\text{NO}$  present in the reaction vessel after conversion with  $\text{NO}_2$  derived from the difference between  $\text{NO}_x$  and  $\text{NO}$  mixing ratios.

The instrument was calibrated by adding a small flow (5 sccm) of known  $\text{NO}$  concentration (5 ppmv – Air Liquide) into the ambient sample flow, resulting in around 10 ppbv of  $\text{NO}$ . The conversion efficiency of the  $\text{NO}_2$  converter was measured during each calibration by gas phase titration of the  $\text{NO}$  to  $\text{NO}_2$  by addition of  $\text{O}_3$ . In flight calibrations were always carried out above the boundary layer, thus ensuring low and stable background levels of  $\text{NO}_x$ . Typically calibrations are carried out at the beginning and end of a flight, with sensitivities and conversion efficiency interpolated between the two and applied to all data. Detection limits for the 10 s averaged data were  $\sim 10$  pptv for  $\text{NO}$  and 15 pptv for  $\text{NO}_2$  with approximate total errors at 1 ppbv being 10 and 15 % for  $\text{NO}$  and  $\text{NO}_2$  respectively.

### 2.1.4 CO measurements

$\text{CO}$  was measured at 1 Hz by an Aerolaser VUV fluorescence analyser model AL5002 (Gerbig et al., 1999). The instrument was calibrated in-flight using an air standard, traceable to the World Meteorological Organisation  $\text{CO}$  scale X2004, with a 2 % uncertainty, and 3 ppb precision.

### 2.1.5 $\text{C}_5\text{H}_8$ measurements

Isoprene was measured using an on-board proton transfer reaction mass spectrometer containing a quadrupole detector (PTR-MS, Ionicon Analytik GmbH, Innsbruck, Austria). During the SAMBBA campaign, the instrument measured a range of hydrocar-

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bons and oxygenated hydrocarbons with a typical cycle time of 3–5 s. Isoprene mixing ratios were determined using a dynamically-diluted calibrated gas standard ( $\sim 500$  ppb in nitrogen, uncertainty  $\pm 5\%$ , Apel-Reimer, Boulder CO). For the purposes of this comparison we have applied a 15-point smoothing function to the high frequency data to give an approximately 1 min moving averaged mixing ratio. The mean limit of detection for isoprene under these conditions was 110 ppt. The overall measurement uncertainty is estimated to be  $\pm 15\%$ . Full instrumental, operational and calibration details are described in Murphy et al. (2010).

Additionally, whole air samples (WAS) were collected on flights B735 and B749, and subsequently analysed to measure isoprene mixing ratios. The WAS system, described in greater detail by Lidster et al. (2014), comprises sixty four canisters with fused silica deactivated inner surfaces, each with three litre internal volume. Individual canisters were filled at operator-determined times using a double-headed metal bellows pump (all stainless steel components) to a final pressure of up to 40 psi and shipped back to the UK for analysis within one month of collection. Analysis was performed using a dual channel gas chromatograph with flame ionisation detectors, described in detail by Hopkins et al. (2011), which was calibrated using a certified standard supplied by the National Physical Laboratory (Ozone precursors mix, cylinder number D641613). Detection limits were in the single parts per trillion range with typical calculated uncertainties of between 3 and 20 %.

## 2.2 UM-UKCA and CiTTyCAT models

### 2.2.1 UM-UKCA (Eulerian model)

We start from the setup of UM-UKCA, employing present day boundary conditions, that Squire et al. (2014) demonstrated had some skill at reproducing recent tropospheric  $O_3$  observations (sonde profiles from the SHADOZ network; see their Fig. 3). This setup, similar to that described by Telford et al. (2010), was comprised of the Hadley Centre Global Environment Model version 3 – Atmosphere only (HadGEM3-A r2.0) at

UM version 7.3 (Hewitt et al., 2011) and UKCA TropChem (O'Connor et al., 2014). For full details, the reader is referred to Squire et al. (2014). Here, we simply note that the model was run in “climate mode” – at a relatively low spatial resolution, N48 L60 (3.75° longitude × 2.5° latitude; 60 hybrid height levels stretching from the surface to around 84 km) – and employed the standard tropospheric chemistry mechanism, CheT. This is the setup that Squire et al. (2015) subsequently used in their “BASE CheT” experiment. Their “BASE CheT2” experimental setup was identical except for employing the updated CheT2 chemistry. Here, we carry out two integrations with UM-UKCA based on Squire et al. (2015)’s BASE CheT and BASE CheT2 experiments: UM-UKCA (CheT) and UM-UKCA (CheT2), respectively.

Our integrations differ from Squire et al. (2015)’s in four ways: we employ different trace gas emissions, as outlined in Sect. 2.4; we nudge UM-UKCA towards European Centre for Medium-range Weather Forecasts (ECMWF) ERA Interim analyses, as described by Telford et al. (2009); we run the model for just less than 9.5 months (8 months from 00:00 UT 2 January 2012 to spin the model up from BASE CheT and BASE CheT2 start dumps, and a further 40 days from 00:00 UT 2 September 2012 to cover the SAMBBA campaign period); and we output the concentrations of all chemical species at points spaced one minute apart along each of the five SAMBBA flights, using Telford et al. (2013)’s flight track code. Note that we also output the concentrations of all chemical species simulated at the times and locations of the air parcel trajectories 7 days previously, based on the back-trajectory calculations mentioned in Sect. 2.1.1. These data are used to initialise the integrations with CiTTYCAT. Likewise, when modelling the composition of air downwind of Manaus, we provide UM-UKCA with an artificial “flight track” to output the concentrations of chemical species simulated at 6 hourly intervals (throughout September 2012) at a boundary layer site (61.0° W, 3.1° S, 900 hPa) roughly 100 km downwind (1° west) of Manaus, and at the locations of the corresponding air parcels 7 days previously – based on further back-trajectory calculations.

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## 2.2.2 CiTTYCAT (Lagrangian model)

CiTtYCAT r4.2.1 (Pugh et al., 2012) is a Lagrangian model of atmospheric chemistry and transport, stemming from the Cambridge Tropospheric Trajectory model of Chemistry And Transport (Wild et al., 1996). This is not the first time that CiTTYCAT has been used to simulate atmospheric chemistry and composition over a tropical rainforest: Pugh et al. (2010), as briefly referred to in the introduction, tested the performance of the model in two-box mode (two boxes, to account for the nocturnal collapse of the boundary layer and development of a residual layer above it), confronting it with measurements made during the OP3 campaign at Danum Valley, Malaysian Borneo. We use the model in single trajectory mode (moving a single model box along one trajectory at a time) many times over as we loop over all back-trajectories bound for (a) the arrival points spaced one minute apart on the five SAMBBA flights, and (b) the receptor site downwind of Manaus at 6 hourly intervals throughout September 2012. The single trajectory mode has been used extensively in previous studies of long range transport (see, e.g., Wild et al., 1996; Evans et al., 2000; Real et al., 2007, 2008). Note that the treatment of transport constitutes the main difference between CiTTYCAT and UM-UKCA: transport in the Lagrangian framework is described by discrete trajectories (series of times and locations) calculated offline, as opposed to fluxes between adjacent model boxes in a fixed 3-D Eulerian grid.

The back-trajectories, illustrated in Fig. 1, are calculated using ROTRAJ (Methven, 1997) in conjunction with ECMWF ERA Interim analyses, as previously outlined by Pugh et al. (2012). The analysed wind fields, available at 6 hourly intervals (00:00 UT, 06:00 UT, 12:00 UT and 18:00 UT) are interpolated linearly in space and time. The location of each trajectory is then calculated by integrating the interpolated wind velocities with respect to time according to the fourth order Runge–Kutta method (Methven, 1997) and recorded every 6 h together with the air temperature and specific humidity, which are also interpolated in space and time from the analyses.

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To ensure the transport in the two models is broadly consistent, we use the same analyses to calculate the trajectories as we use to nudge UM-UKCA (see above). However, two key differences remain. Firstly, the trajectory calculations exploit the full-resolution of the analysed winds (roughly  $0.7^\circ \times 0.7^\circ$ ) whilst UM-UKCA is nudged towards these winds following degradation to the resolution of its Eulerian grid ( $3.75^\circ \times 2.5^\circ$  in “climate mode”). The transport in CiTTYCAT is therefore more finely resolved and should yield greater structure in the composition of air it simulates along each flight track, and downwind of Manaus, particularly when combined with high resolution trace gas emissions. The transport in CiTTYCAT, however, only includes convection as captured by the analyses (i.e. large-scale convection) whilst UM-UKCA explicitly adds updrafts and downdrafts associated with convection on smaller scales, following Gregory and Rowntree (1990) and Gregory and Allen (1991). CiTTYCAT therefore lacks a certain amount of vertical mixing. Some mixing within the boundary layer is included implicitly, as the addition of emissions (conversion from mass fluxes to enhancements in concentration) depends on a length scale associated with the height of the boundary layer, but no ventilation of the boundary layer or exchange with the free troposphere is included. We focus first on the simulation of independent air parcels – with no vertical (or horizontal) mixing – to explore the influence of contrasting air parcel histories on the chemistry ensuing therein. However, we subsequently explore the sensitivity of some of our results to a simple treatment of diffusive vertical mixing.

The treatment of diffusive vertical mixing, described by Pugh et al. (2012), comprises relaxation towards background composition at rates specified by free-troposphere and boundary-layer diffusion coefficients,  $\kappa_{FT}$  and  $\kappa_{BL}$ . These yield relaxation timescales of  $\tau_{FT} = D^2 / (2\kappa_{FT})$  and  $\tau_{BL} = BLH^2 / (2\kappa_{BL})$ , where  $D$  is a free tropospheric depth parameter and  $BLH$  is boundary layer height. Pugh et al. (2012) suggest  $\kappa_{FT}$  should typically take values of between  $0.5 \text{ m}^2 \text{ s}^{-1}$  (under stable conditions) and  $1.5 \text{ m}^2 \text{ s}^{-1}$  (under more turbulent ones); Pisso et al. (2009) reported slightly lower values of  $0.3\text{--}1.0 \text{ m}^2 \text{ s}^{-1}$ . Typically,  $\kappa_{BL} = 10\kappa_{FT}$  whilst  $D$  takes values of roughly  $200\text{--}500 \text{ m}$ . We explore the impact of mixing subject to three different combinations of  $\kappa_{FT}$ ,  $\kappa_{BL}$  and  $D$  (Mix1, Mix2 and



coefficients in the two models (Squire et al., 2014, 2015), and we use 3-D fields of precipitation, output from UM-UKCA every 20 min timestep, to drive the wet deposition in CiTTyCAT. This is in addition to initialising the composition of air parcels in CiTTyCAT with the concentrations of species simulated in UM-UKCA (subject to the same chemical mechanism) as described at the end of the last section.

### 2.3 CheT and CheT2 chemical mechanisms

The standard tropospheric chemistry mechanism, CheT, includes 56 chemical tracers and 165 photochemical reactions, of which 16 tracers and 44 reactions comprise the MIM (Pöschl et al., 2000). It is the result of a systematic reduction of version 2 of the Master Chemical Mechanism (MCM; Jenkin et al., 1997), in which species are lumped together based on their structure, for example all hydroxyperoxy radicals as “ISO<sub>2</sub>”. CheT2 differs only with respect to isoprene oxidation, with 24 tracers and 59 reactions in place of the previous 16 and 44 respectively, and is traceable to MCM version 3.2 (MCMv3.2). The differences, reflecting the updates compiled by Jenkin (2012) for the UK Met Office, are as follows:

1. Changes to the chemistry of first generation isoprene nitrates (ISON): NO<sub>x</sub> is regenerated from ISON in CheT by photolysis, or conversion to second generation nitrates (NALD), followed by reaction with OH; in CheT2, the yield of NO<sub>x</sub> from ISON is increased in line with the measurements of Perring et al. (2009) by increasing the rate of ISON photolysis and adding a further ISON + OH → NO<sub>2</sub> channel; CheT2 also includes the addition of O<sub>3</sub>-initiated ISON degradation (Lockwood et al., 2010).
2. The inclusion, as mentioned in Sect. 1, of a route by which OH initially consumed in isoprene oxidation may be efficiently regenerated at low ambient NO<sub>x</sub> concentrations: the formation of hydroperoxy-aldehydes (HPALDS) from hydroperoxy radicals (ISO<sub>2</sub>) and their subsequent rapid release of OH (Peeters et al., 2009; Crouse et al., 2011).

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3. The inclusion of the formation of isoprene epoxydiols (IEPOX) from the oxidation of isoprene hydroxyl-hydroperoxides (ISOOH); Paulot et al. (2009) identified these as a potential source of secondary organic aerosols.

4. A reduction in the yield of peroxyacetylic nitric anhydride (MPAN) from isoprene oxidation relative to that adopted in CheT; see Jenkin (2012) for details.

In this study, however, we are less concerned with the differences between the two mechanisms, which have already been explored at length (see, e.g., Archibald et al., 2010a, b; Squire et al., 2015), than we are with their relative abilities to reproduce observations of atmospheric composition above the Amazon rainforest – and the latter subject to different model frameworks (Eulerian and Lagrangian) and trace-gas emissions.

### 2.4 Trace gas emissions

The trace gas emissions are comprised of: anthropogenic emissions taken from EDGAR version 4.2 (<http://edgar.jrc.ec.europa.eu>); and biogenic emissions calculated with the Organising Carbon and Hydrology In Dynamic Ecosystems land surface model (ORCHIDEE), with the exception of NO<sub>2</sub> emissions from soils that are taken from the Global Emissions Inventory Activity (GEIA; Yienger and Levy, 1995). The annual total emission of each species, globally, is given in Table 3, including its breakdown into anthropogenic and biogenic components.

We employ EDGAR 4.2 emissions of NO<sub>2</sub>, CO and NMVOCs from all sectors apart from “Non-road transportation” (1A3a + c + d + e in the nomenclature of the Intergovernmental Panel on Climate Change; IPCC), since the latter includes aircraft emissions that are difficult to implement in the Lagrangian model; CH<sub>4</sub> is treated as a constant field (1.76 ppmv). We adopt the most recent emissions available, which correspond to the year 2008. Available at a spatial resolution of up to 0.1° × 0.1° globally, these are capable of resolving a city of approximately 10 km × 10 km in the tropics, such as Manaus. We note, however, that they do not include any seasonality; we expect the

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seasonality to be relatively low in the tropics. The NMVOC emissions come lumped together as a single carbon flux. We derive emissions of ethane ( $C_2H_6$ ), propane ( $C_3H_8$ ), formaldehyde (HCHO), acetone ( $CH_3C(O)CH_3$ ) and acetaldehyde ( $CH_3CHO$ ) from this using the IPCC (2002)'s speciation of industrial- and biomass burning emissions; see their Table 4.7(b). This speciation is crude: we assume, for example, that “ketones” are entirely comprised of  $CH_3C(O)CH_3$ , and “other aldehydes” solely  $CH_3CHO$ . However, our priority is to start from anthropogenic emissions of sufficient spatial resolution to resolve the city of Manaus, and EDGAR 4.2 is unique in providing emissions of this resolution, globally. The CiTTYCAT integrations employing “High res” emissions exploit their full  $0.1^\circ \times 0.1^\circ$  resolution. For use in UM-UKCA, and the CiTTYCAT integrations employing “UKCA res” emissions, the emissions are degraded to  $3.75^\circ \times 2.5^\circ$ ; see Table 2 and accompanying text.

As stated above, biogenic emissions of  $C_5H_8$ , HCHO,  $CH_3C(O)CH_3$  and  $CH_3CHO$  are calculated with ORCHIDEE. This includes parameterisations based on Guenther et al. (1995) and Lathièrè et al. (2006), modified according to Guenther et al. (2012) and more recent findings to take into account the progress of our knowledge in this field (see Messina et al., 2015). ORCHIDEE is forced with 2012 National Centers for Environmental Prediction meteorological analyses (NCEP v5.3) from the Climatic Research Unit of the US National Centre for Atmospheric Research. These daily (24 h average) emissions are used at full spatial resolution ( $0.5^\circ \times 0.5^\circ$ ) in the “High res” integrations with CiTTYCAT but, just as for the anthropogenic emissions, degraded to  $3.75^\circ \times 2.5^\circ$  for use in UM-UKCA and the CiTTYCAT integrations employing “UKCA res” emissions. We apply a diurnal cycle – the same in both models, based on the division of each 24 h period into 20 min intervals – to the emissions of  $C_5H_8$  but not HCHO,  $CH_3C(O)CH_3$  or  $CH_3CHO$ . The  $NO_2$  emissions from soils are taken from GEIA dataset, soilNOXmn1.1a (Yienger and Levy, 1995). In view of the uncertainty in these, they are used in both models, in all integrations, at a resolution of  $3.75^\circ \times 2.5^\circ$ .

Figure 2 illustrates the total “UKCA res” emissions of each species (anthropogenic + biogenic) on the 1 January and 1 July 2012. Recall, only the species including a bio-



tions of the flight (around 12.5, 13.5, 14.5 and 15.5 UT; 08:30, 09:30, 10:30 and 11:30 LT). This is consistent with the air parcels during those periods generally not encountering the boundary layer (altitudes of less than 1000–2000 m) during the previous 7 days; see Fig. 1 and accompanying text. Since these air parcels are subject to neither emissions nor dry deposition, the small differences between the green and dashed-blue lines are due to intervening chemistry and/or wet deposition. The  $[O_3]$  simulated with CiTTYCAT also shows much more structure in these regions than that simulated with UM-UKCA. The initialisation is responsible for this additional structure – the result of divergent air parcel histories (see Sect. 2.1.1 and, again, Fig. 1) – note that it is largely retained in CiTTYCAT over the course of 7 days but almost entirely lost in UM-UKCA. Compared to the measurements, CiTTYCAT shows too much structure (perhaps due to the lack of explicit mixing in the model) whilst UM-UKCA shows too little (presumably due to mixing on the scale of its  $3.75^\circ \times 2.5^\circ$  grid); see discussion in Sect. 2.2.2 of the differences in transport, especially mixing, between the two models. A combination of the two models, such as the simple treatment of diffusive vertical mixing described earlier (relaxation of CiTTYCAT towards UM-UKCA), would likely yield best agreement with the measurements in this respect; see later. We note, however, that on two flights in which the measurements of  $[O_3]$  at high altitudes show less structure (B744 and B749), UM-UKCA yields much better agreement with the measurements; see Figs. S1 and S3 in the Supplement.

At low altitudes (0–1500 m), UM-UKCA tends to overestimate measured  $[O_3]$  whilst CiTTYCAT generally shows good agreement; the exception is the period shortly after 13:00 UT in which UM-UKCA shows good agreement and CiTTYCAT underestimates the measurements. Generally, CiTTYCAT simulates lower  $[O_3]$  than UM-UKCA in these low altitude regions and yields better agreement with measured  $[O_3]$  here. Similar behaviour is observed on flights B749 and B750 (see Figs. S3 and S4). We note, however, that this does not extend to all low-altitude portions of flights B744 and B745 (see Figs. S1 and S2); on these two flights, at altitudes towards the top end of the 0–1500 m range, possibly bordering on the free troposphere, CiTTYCAT starts to yield

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chemical mechanism and an increase in the resolution of trace gas emissions. The switch from CheT to CheT2 chemistry, for example, may be expected to reduce the  $[C_5H_8]$  simulated in both models on account of including an OH “recycling” mechanism (see Sect. 2.3).

Finally, we compare modelled and measured  $[CO]$  in Fig. 7. Note that the high/low  $[CO]$  values at approximately one hour intervals correspond to the “span/zero” in-flight calibrations of the VUV fluorescence CO monitor (see Sect. 2.1.4). UM-UKCA generally underestimates the measurements, simulating around 100 ppbv throughout the flight; this is understood to be linked to a general high bias in  $[OH]$ , and hence a low bias in CO lifetime, in this version of the model (see, e.g., Telford et al., 2013). Meanwhile, CiTTYCAT simulates near identical concentrations to UM-UKCA in the mid-to-high altitude portions of the flight, but considerably higher  $[CO]$  at low altitudes. These higher concentrations overestimate the measurements – possibly the result of overestimating  $[C_5H_8]$ , and hence overestimating the production of CO from  $C_5H_8$  oxidation. A lack of boundary-layer ventilation in CiTTYCAT, and hence a lack of exchange between air with high  $[CO]$  in the boundary layer and free tropospheric air of lower  $[CO]$ , could again contribute. The low bias in  $[CO]$  simulated with UM-UKCA extends across all five flights (see Figs. S1–S4); the difference in behaviour between the two models clearly extends to flights B749 and B750 (Figs. S3 and S4) but is less pronounced on flights B744 and B745 (Figs. S1 and S2).

### 3.2 Comparing chemical mechanisms (CheT and CheT2)

We find that the chemical mechanism employed has a negligible effect on the  $[O_3]$ , and only a modest effect on the  $[NO]$ ,  $[NO_2]$ ,  $[C_5H_8]$  and  $[CO]$ , we simulate for flight B735 in UM-UKCA and CiTTYCAT (see Fig. 8). The impact of switching from CheT to CheT2 for  $[NO]$  and  $[NO_2]$  is largely limited to the low altitude portions of the flight in which  $C_5H_8$  is encountered. This is consistent with the direct changes to the chemistry being limited to ones relating to ISON (the lumped species comprised of first generation isoprene nitrates); see Sect. 2.3. To first order, we would expect the increase in the

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We observe this behaviour across all five case study flights, and it is this central finding, which appears to result from the choice of model framework alone, on which we mean to focus from here on and speculate is the result of differences in vertical mixing.

### 3.4 Exploring sensitivity to vertical mixing in CiTTYCAT

To explore the impact of introducing a simple treatment of diffusive vertical mixing in the Lagrangian framework, we return to flight B735, employing “UKCA res” emissions and CheT chemistry. We explore three formulations of mixing, or relaxation, as outlined in Sect. 2.2.2 and Table 1: Mix1, Mix2 and Mix3. These formulations differ with respect to the timescales on which the concentrations of species simulated in CiTTYCAT are relaxed towards background concentrations in the free troposphere ( $\tau_{FT}$ ) and boundary layer ( $\tau_{BL}$ ), increasing from Mix1 to Mix3 commensurate with increasingly stable conditions; see, again, Table 1. The three formulations can be caricatured as follows: Mix1 (solid red line) applies rapid relaxation in both regions; Mix2 (solid blue line) applies slower relaxation in both regions; and Mix3 (solid green line) applies the same, relatively slow relaxation in the boundary layer as Mix2 but still slower relaxation in the free troposphere. The picture regarding [NO] is not simple to summarise. However, for [O<sub>3</sub>], [NO<sub>2</sub>], [C<sub>5</sub>H<sub>8</sub>] and [CO], the inclusion of this relaxation, subject to all three formulations (Mix1–3), brings the concentrations simulated with CiTTYCAT, originally with no mixing (dotted green lines), much closer in line with those simulated with UM-UKCA (dashed blue lines). It would therefore appear that vertical mixing (or the lack thereof) has potential to explain some of the differences observed between the two models/model frameworks.

Of course, the close agreement between UM-UKCA and CiTTYCAT, on including mixing in the latter, is only to be expected for [O<sub>3</sub>], since we relax the [O<sub>3</sub>] simulated with CiTTYCAT towards monthly mean values simulated with UM-UKCA (see Sect. 2.2.2). We do likewise for other species of intermediate lifetimes: CO, C<sub>2</sub>H<sub>6</sub>, C<sub>3</sub>H<sub>8</sub> and PAN. Meanwhile, the concentrations of all short-lived species, including C<sub>5</sub>H<sub>8</sub>, are relaxed towards zero concentrations – characteristic of free tropospheric air. The most rapid

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relaxation, Mix1, yields the best agreement between modelled and measured  $[C_5H_8]$ . Indeed, the agreement is excellent. However, Mix1 yields the worst agreement between modelled and measured  $[O_3]$  in the low altitude portions of the flight – in these regions, considerably worse than the simulation without mixing, consistently overestimating the measurements by 15–20 ppbv in absolute terms, and close to 100 % in relative terms. The slower relaxations, Mix 2 and Mix 3, yield somewhat higher  $[C_5H_8]$  – greater than that measured but a significant improvement over that simulated with CiTTyCAT without mixing, and better in two out of three portions of the flight than that simulated with UM-UKCA. Mix2 and Mix3, meanwhile, yield better agreement between modelled and measured  $[O_3]$  in the low altitude parts of the flight. Furthermore, Mix3 starts to retain some of the structure in  $[O_3]$  simulated in the high altitude portions of the flight that is almost entirely absent from the simulation with UM-UKCA but present, at least to some extent, in the measurements; see earlier discussion in Sect. 3.1. Judging by the structure present in the  $[O_3]$  measured in these parts of the flight, there could be justification for still slower relaxation in the free troposphere. The mixing formulation could be iterated further, however, of the three formulations applied here (exploring the range of parameter values suggested in the literature; see Sect. 2.2.2 and Table 1), we judge Mix3 to be the most appropriate overall.

### 3.5 Modelling atmospheric chemistry downwind of Manaus

We now move from what has predominantly been an exploration in the spatial domain – in other words, comparing modelled and measured trace gas concentrations on flight tracks – to an exploration of the temporal domain. We simulate  $[O_3]$  at a single site downwind of Manaus over an extended period of time, the month in which all five case-study SAMBBA flights took place. Recall, back-trajectories were calculated from a boundary layer site (61.0° W, 3.1° S, 900 hPa) roughly 100 km downwind (1° west) of Manaus at 6 hourly intervals throughout September 2012. The choice of site is arbitrary; we are not aware of any measurements of  $[O_3]$  at this site or comparable sites, during the period in question, with which to compare the  $[O_3]$  we simulate –

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this is a purely model-based exploration of the consequences of simulating boundary layer [ $O_3$ ] with an Eulerian model run in “climate mode” (UM-UKCA) and a Lagrangian model capable of exploiting very high resolution anthropogenic emissions and retaining compositional structure (CiTTyCAT).

The structure that Lagrangian modelling can generate in [ $O_3$ ], and retain over a period of time, is of particular interest in an exploration of the exposure to boundary-layer [ $O_3$ ] of the rainforest downwind of an isolated source such as Manaus. Plant exposure to ozone is often expressed in terms of accumulated exposure above a given threshold. For crop species, the most commonly used metric remains the total number of hour-averages over 40 ppbv accumulated over a 3 month growing season, AOT40 (UNECE, 2010), although there is growing awareness that plant effects are more closely linked to the stomatal ozone dose than concentration exposure, such as the phytotoxic  $O_3$  dose above a threshold flux  $Y$  ( $POD_Y$ ; see, e.g., LRTAP Convention, 2010). For tropical, long-lived, wild plants, such as those in the rainforest downwind of Manaus, the “growing season” may not be the appropriate accumulation time (Ainsworth et al., 2012). Below, we discuss an AOT40-like exposure metric based on a 30 day time series, and refer to the metric as “AOT40” to remind readers of the non-standard accumulation period we use. Our study is not designed to calculate annual exposure metrics, but simply to highlight the sensitivity of exposure metrics to the modelling method used (where measurements are not available). In view of the sensitivity that our simulations with CiTTyCAT show to the inclusion of a simple treatment of diffusive vertical mixing (relaxation towards background composition; see previous section), we simulate the [ $O_3$ ] downwind of Manaus with and without relaxation formulations, Mix1–3. We start, as before, by using CheT chemistry but employ High res emissions ( $0.1^\circ \times 0.1^\circ$ ) in an attempt to resolve the episodic influence of anthropogenic emissions from the city 100 km upwind. We compare the results of these integrations with the [ $O_3$ ] simulated with UM-UKCA in “climate mode” (employing anthropogenic emissions at  $3.75^\circ \times 2.5^\circ$ ) in the top left of Fig. 11; the corresponding “box and whisker” plots of the absolute min-

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imum, absolute maximum, median, and 25th and 75th percentile values of simulated  $[O_3]$  are included in the top right.

In the absence of vertical mixing, the  $[O_3]$  simulated with CiTTYCAT (dotted green line) exhibits much more structure than that simulated with UM-UKCA (dashed blue line), frequently exceeding 50 ppbv and exceeding 75 ppbv on seven occasions. Depending on the speed of relaxation imposed, this structure is suppressed to a greater or lesser extent, and the simulations with CiTTYCAT (solid red, blue and green lines) can generate more or less variability in the time series than UM-UKCA. Mixing formulation, Mix3, judged in the last section to yield best agreement between modelled and measured  $[O_3]$  over a range of altitudes, including specifically low altitudes, yields a distribution of  $[O_3]$  that has a higher median value than UM-UKCA (34.0 cf. 31.7 ppbv), a higher 75th percentile (38.3 cf. 34.9 ppbv), and a higher absolute maximum (54.3 cf. 41.7 ppbv). The  $[O_3]$  simulated with UM-UKCA never exceeds 50 ppbv in this period, whilst that simulated with CiTTYCAT, subject to Mix3, does so five times. This shift towards higher  $[O_3]$  leads to an 8 % increase in mean  $[O_3]$  simulated throughout the month, from 32.5 to 35.1 ppbv. Note that the “AOT40” we calculate over just 30 days increases by a factor of almost 40, from 22.6 to 863 ppbv h.

In the bottom left of Fig. 11, we explore the effects on the simulation of  $[O_3]$  with CiTTYCAT, subject to Mix3, of changing from CheT to CheT2 chemistry and/or degrading the resolution of the emissions to that used in UM-UKCA ( $3.75^\circ \times 2.5^\circ$ ); the corresponding box and whisker plots are included in the bottom right. Only modest differences arise. The switch to CheT2 chemistry yields substantially different  $[O_3]$  only on two days (21 and 24 September); compare the green and red lines. Meanwhile, degrading the resolution of the emissions has little effect throughout; compare the solid and dotted lines. Our earlier findings appear to hold irrespective of the chemistry and emissions employed: CiTTYCAT (Mix3) yields much more structure in  $[O_3]$  than UM-UKCA, exceeding 50 ppbv on four or five occasions as opposed to none; and, whilst the mean  $[O_3]$  increases by approximately 10 %, or 3 ppbv, “AOT40” increases by a factor of roughly 40–60; see Table 4. Whilst our study was not designed to calculate flux-based

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included neither explicit vertical mixing in the boundary layer (due to convection) nor ventilation of the boundary layer and exchange with air in the free troposphere. It is to the latter that we attribute the Lagrangian model's overestimation of  $[\text{C}_5\text{H}_8]$  in the boundary layer and  $[\text{O}_3]$  in the free troposphere; we believe they are the result of a lack of mixing with, and hence dilution by, lower  $[\text{C}_5\text{H}_8]$  air above and lower  $[\text{O}_3]$  air below respectively (see Sect. 3.1 and, earlier, Sect. 2.2.2).

The simple approach to diffusive vertical mixing that we later introduced into the Lagrangian model has been used in previous studies; see Pugh et al. (2012) and the references contained therein. It comprised: the relaxation of species of “intermediate lifetimes” towards 3-D monthly mean concentrations simulated in the Eulerian model; and the relaxation of short-lived species (all other species besides  $\text{CH}_4$ ) towards zero concentrations (characteristic of the free troposphere). Recall that  $[\text{CH}_4]$  was fixed in the model. The relaxation was applied on timescales  $\tau_{\text{BL}}$  and  $\tau_{\text{FT}}$  in the boundary layer and free troposphere respectively; see Sect. 2.2.2 and Table 1 for more details. Having explored just three combinations of  $\tau_{\text{BL}}$  and  $\tau_{\text{FT}}$  spanning literature values, we cannot claim to have fully optimised this simple treatment of mixing. Predictably, however, all three formulations brought the Lagrangian simulations more closely in line with their Eulerian counterparts; see Sect. 3.5. Moreover, one formulation, Mix3 ( $\tau_{\text{BL}} = 27.8$  h, subject to a boundary layer height of 1000 m, and  $\tau_{\text{FT}} = 69.4$  h) yielded the desirable combination of: reasonable agreement with measurements of both  $[\text{O}_3]$  and  $[\text{C}_5\text{H}_8]$  at low altitudes; and structure in the simulated  $[\text{O}_3]$  reminiscent of that exhibited by the measurements. It was predominantly this formulation that we subsequently applied to a simulation of boundary layer  $[\text{O}_3]$ .

We should note at this point that the  $[\text{O}_3]$  simulated by this combination of Lagrangian and Eulerian models (subject to Mix3) did not show as good agreement with the measurements at low altitudes as that simulated with the Lagrangian model alone: the Lagrangian model simulated 10–15 ppbv lower  $[\text{O}_3]$  than the combination of models (and the Eulerian model on its own), in closer agreement with the measurements. It may have done so, however, for the wrong reason. If the lack of vertical mixing in the

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Lagrangian model contributed to its overestimation of  $[O_3]$  in the free troposphere, it likely also contributed to its simulation of low  $[O_3]$  in the boundary layer; the lack of ventilation of the boundary layer, and exchange with free tropospheric air aloft, would have neglected mixing with, and enrichment by, higher  $[O_3]$  air above. This raises the question whether something else might be amiss in the model(s), leading to an overestimation of boundary layer  $[O_3]$  in this environment. One possibility is that the dry deposition of  $O_3$  to the rainforest is underestimated. Hardacre et al. (2015) recently highlighted the crude treatment of dry deposition to tropical forests in current global chemistry-climate models. In both the Lagrangian model and the Eulerian model, we have employed a 1 m dry deposition velocity ( $V_d$ ) of  $0.5 \text{ cm s}^{-1}$  to the forest. This compares favourably with the measurements of Rummel et al. (2007) in the Amazon, which yielded a “mean midday maximum”  $O_3$   $V_d$  of  $0.5 \text{ cm s}^{-1}$  in the dry season. However, such measurements in the Amazon are sparse and the  $V_d$  of  $O_3$  could vary from one region to another: more  $O_3$   $V_d$  measurements are called for.

We used the Lagrangian model, the Eulerian model, and the combination of the two models, to simulate  $[O_3]$  at a site in the boundary layer, approximately 100 km downwind of Manaus, over a period of a month; see Sect. 3.5. The choice of site was arbitrary, but chosen in anticipation of demonstrating the need to employ very high resolution anthropogenic emissions (capable of resolving a city of the order of  $10 \text{ km} \times 10 \text{ km}$ ) to correctly capture the chemistry and composition downwind. To our surprise, our simulations again proved relatively insensitive to the resolution of emissions (and chemical mechanism) employed, consistent with our earlier findings when simulating the SAMBBA measurements. However, again, they demonstrated a high degree of sensitivity to the model framework; see top left and top right of Fig. 11. At one extreme, the Lagrangian model yielded very high variability in  $[O_3]$  at this site, frequently exceeding 50 ppbv and sometimes exceeding 75 ppbv. At the other extreme, the Eulerian model simulated far less variability in  $[O_3]$ , never exceeding 50 ppbv and yielding a mean  $[O_3]$  of 32.5 ppbv and an “AOT40” of around 20 ppbv h (over 30 days). Between these extremes, the combination of Lagrangian and Eulerian models (subject



for Atmospheric Science, University of Cambridge, for their generous welcome whilst working remotely there as a visiting researcher.

## References

- Ainsworth, E. A., Yendrek, C. R., Sitch, S., Collins, W. J., and Emberson, L. D.: The effects of tropospheric ozone on net primary productivity and implications for climate change, *Annu. Rev. Plant Biol.*, 63, 637–61, 2012.
- Archibald, A. T., Cooke, M. C., Utembe, S. R., Shallcross, D. E., Derwent, R. G., and Jenkin, M. E.: Impacts of mechanistic changes on HO<sub>x</sub> formation and recycling in the oxidation of isoprene, *Atmos. Chem. Phys.*, 10, 8097–8118, doi:10.5194/acp-10-8097-2010, 2010a.
- Archibald, A. T., Jenkin, M. E., and Shallcross, D. E.: An isoprene mechanism intercomparison, *Atmos. Environ.*, 44, 5356–5364, doi:10.1016/j.atmosenv.2009.09.016, 2010b.
- Arneth, A., Monson, R. K., Schurgers, G., Niinemets, Ü., and Palmer, P. I.: Why are estimates of global terrestrial isoprene emissions so similar (and why is this not so for monoterpenes)?, *Atmos. Chem. Phys.*, 8, 4605–4620, doi:10.5194/acp-8-4605-2008, 2008.
- Avnery, S., Mauzerall, D., Liu, J., and Horowitz, L.: Global crop yield reductions due to surface ozone exposure: 1. Year 2000 crop production losses and economic damage, *Atmos. Environ.*, 45, 2284–2296, 2011.
- Beerling, D. and Woodward, F.: *Vegetation and the Terrestrial Carbon Cycle: The First 400 Million Years*, Cambridge University Press, Cambridge, UK, p. 405, 2001.
- Beerling, D., Woodward, F., Lomas, M., and Jenkins, A.: Testing the responses of a dynamic global vegetation model to environmental change: a comparison of observations and predictions, *Global Ecol. Biogeogr.*, 6, 439–450, 1997.
- Butler, T. M., Taraborrelli, D., Brühl, C., Fischer, H., Harder, H., Martinez, M., Williams, J., Lawrence, M. G., and Lelieveld, J.: Improved simulation of isoprene oxidation chemistry with the ECHAM5/MESSy chemistry-climate model: lessons from the GABRIEL airborne field campaign, *Atmos. Chem. Phys.*, 8, 4529–4546, doi:10.5194/acp-8-4529-2008, 2008.
- Cain, M., Methven, J., and Highwood, E. J.: Quantification of chemical and physical processes influencing ozone during long-range transport using a trajectory ensemble, *Atmos. Chem. Phys.*, 12, 7015–7039, doi:10.5194/acp-12-7015-2012, 2012.

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Carlton, A. G., Wiedinmyer, C., and Kroll, J. H.: A review of Secondary Organic Aerosol (SOA) formation from isoprene, *Atmos. Chem. Phys.*, 9, 4987–5005, doi:10.5194/acp-9-4987-2009, 2009.

Chen, Q., Farmer, D. K., Rizzo, L. V., Pauliquevis, T., Kuwata, M., Karl, T. G., Guenther, A., Allan, J. D., Coe, H., Andreae, M. O., Pöschl, U., Jimenez, J. L., Artaxo, P., and Martin, S. T.: Submicron particle mass concentrations and sources in the Amazonian wet season (AMAZE-08), *Atmos. Chem. Phys.*, 15, 3687–3701, doi:10.5194/acp-15-3687-2015, 2015.

Crounse, J. D., Paulot, F., Kjaergaard, H. G., and Wennberg, P. O.: Peroxy radical isomerization in the oxidation of isoprene, *Phys. Chem. Chem. Phys.*, 13, 13607–13613, 2011.

Crowther, R., Law, K., Pyle, J., Bekki, S., and Smit, H.: Characterising the effect of large-scale model resolution upon calculated OH production using MOZAIC data, *Geophys. Res. Lett.*, 29, 1613, doi:10.1029/2002GL014660, 2002.

Darbyshire, E. and Johnson, B.: The South American Biomass Burning Analysis (SAMBBA) Field Experiment, September–October 2012, Brazil: Summary of research flights (available on request from the authors), University of Manchester, Manchester, UK, and UK Met Office, Exeter, UK, 2013.

Dodge, M.: Combined use of modeling techniques and smog chamber data to derive ozone-precursor relationships, in: Proceedings of the International Conference on Photochemical Oxidant Pollution and its Control, United States Environmental Protection Agency, 12–17 September 1976, Raleigh, North Carolina, USA, 881–889, 1977.

Dragani, R., Redaelli, G., Visconti, G., Mariotti, A., Rudakov, V., MacKenzie, A. R., and Stefanutti, L.: High resolution stratospheric tracer fields reconstructed with lagrangian techniques: a comparative analysis of predictive skill, *J. Atmos. Sci.*, 59, 1943–1958, 2002.

European Commission, Joint Research Centre (JRC)/Netherlands Environmental Assessment Agency (PBL): Emission Database for Global Atmospheric Research (EDGAR), release version 4.0, available at: <http://edgar.jrc.ec.europa.eu> (last access: January 2015), 2009.

Evans, M. J., Shallcross, D. E., Law, K. S., Wild, J. O. F., Simmonds, P. G., Spain, T. G., Berrisford, P., Methven, J., Lewis, A. C., McQuaid, J. B., Pilling, M. J., Bandy, B. J., Penkett, S. A., and Pyle, J. A.: Evaluation of a Lagrangian box model using field measurements from EASE (Eastern Atlantic Summer Experiment) 1996, *Atmos. Environ.*, 34, 3843–3863, 2000.

Gerbig, C., Schmitgen, S., Kley, D., and Volz-Thomas, A.: An improved fast-response vacuum-UV resonance fluorescence CO instrument, *J. Geophys. Res.*, 104, 1699–1704, doi:10.1029/1998jd100031, 1999.

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Gregory, D. and Allen, S.: The effect of convective downdraughts upon NWP and climate simulations, in: Nineth Conference on Numerical Weather Prediction, Denver, Colorado, 122–123, 1991.

Gregory, D. and Rowntree, P. R.: A massflux convection scheme with representation of cloud ensemble characteristics and stability dependent closure, *Mon. Weather Rev.*, 118, 1483–1506, 1990.

Guenther, A., Hewitt, C., Erickson, D., Fall, R., Geron, C., Graedel, T., Harley, P., Klinger, L., Lerdau, M., McKay, W., Pierce, T., Scholes, B., Steinbrecher, R., Tallamraju, R., Taylor, J., and Zimmerman, P.: A global model of natural volatile organic compound emissions, *J. Geophys. Res.*, 100, 8873–8892, 1995.

Guenther, A., Karl, T., Harley, P., Wiedinmyer, C., Palmer, P. I., and Geron, C.: Estimates of global terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and Aerosols from Nature), *Atmos. Chem. Phys.*, 6, 3181–3210, doi:10.5194/acp-6-3181-2006, 2006.

Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., and Wang, X.: The Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): an extended and updated framework for modeling biogenic emissions, *Geosci. Model Dev.*, 5, 1471–1492, doi:10.5194/gmd-5-1471-2012, 2012.

Hardacre, C., Wild, O., and Emberson, L.: An evaluation of ozone dry deposition in global scale chemistry climate models, *Atmos. Chem. Phys.*, 15, 6419–6436, doi:10.5194/acp-15-6419-2015, 2015.

Hewitt, C. N., Lee, J. D., MacKenzie, A. R., Barkley, M. P., Carslaw, N., Carver, G. D., Chappell, N. A., Coe, H., Collier, C., Commane, R., Davies, F., Davison, B., DiCarlo, P., Di Marco, C. F., Dorsey, J. R., Edwards, P. M., Evans, M. J., Fowler, D., Furneaux, K. L., Gallagher, M., Guenther, A., Heard, D. E., Helfter, C., Hopkins, J., Ingham, T., Irwin, M., Jones, C., Karunaharan, A., Langford, B., Lewis, A. C., Lim, S. F., MacDonald, S. M., Mahajan, A. S., Malpass, S., McFiggans, G., Mills, G., Misztal, P., Moller, S., Monks, P. S., Nemitz, E., Nicolas-Perea, V., Oetjen, H., Oram, D. E., Palmer, P. I., Phillips, G. J., Pike, R., Plane, J. M. C., Pugh, T., Pyle, J. A., Reeves, C. E., Robinson, N. H., Stewart, D., Stone, D., Whalley, L. K., and Yin, X.: Overview: oxidant and particle photochemical processes above a south-east Asian tropical rainforest (the OP3 project): introduction, rationale, location characteristics and tools, *Atmos. Chem. Phys.*, 10, 169–199, doi:10.5194/acp-10-169-2010, 2010.

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- Hewitt, H. T., Copsey, D., Culverwell, I. D., Harris, C. M., Hill, R. S. R., Keen, A. B., McLaren, A. J., and Hunke, E. C.: Design and implementation of the infrastructure of HadGEM3: the next-generation Met Office climate modelling system, *Geosci. Model Dev.*, 4, 223–253, doi:10.5194/gmd-4-223-2011, 2011.
- 5 Hopkins, J. R., Jones, C. E., and Lewis, A. C.: A dual channel gas chromatograph for atmospheric analysis of volatile organic compounds including oxygenated and monoterpene compounds, *J. Environ. Monitor.*, 13, 2268–2276, 2011.
- IPCC, Climate change 2001: The scientific basis. Contribution of Working Group 1 to the Third Assessment Report of the Intergovernmental Panel on Climate Change, edited by: Houghton, J. T., Ding, Y., Griggs, D. J., Noguer, M., van der Linden, P. J., Dai, X., Maskell, K., and Johnson, C. A., Cambridge University Press, Cambridge, UK, and New York, USA, doi:10.1002/joc.763, 2002.
- 10 Jenkin, M.: Review of the atmospheric chemistry of isoprene and evaluation of mechanisms for global modelling, Tech. Rep., UK Met Office, Atmospheric Chemistry Services, Oakhampton, Devon, UK, 2012.
- 15 Jenkin, M. E., Saunders, S. M., and Pilling, M. J.: The tropospheric degradation of volatile organic compounds: a protocol for mechanism development, *Atmos. Environ.*, 31, 81–104, 1997.
- Krol, M., Molemaker, M., and de Arellano, J.: Effects of turbulence and heterogeneous emissions on photochemically active species in the convective boundary layer, *J. Geophys. Res.*, 20 105, 6871–6884, 2000.
- Kuhn, U., Ganzeveld, L., Thielmann, A., Dindorf, T., Schebeske, G., Welling, M., Sciare, J., Roberts, G., Meixner, F. X., Kesselmeier, J., Lelieveld, J., Kolle, O., Ciccioli, P., Lloyd, J., Trentmann, J., Artaxo, P., and Andreae, M. O.: Impact of Manaus City on the Amazon Green Ocean atmosphere: ozone production, precursor sensitivity and aerosol load, *Atmos. Chem. Phys.*, 10, 9251–9282, doi:10.5194/acp-10-9251-2010, 2010.
- 25 Lathière, J., Hauglustaine, D. A., Friend, A. D., De Noblet-Ducoudré, N., Viovy, N., and Folberth, G. A.: Impact of climate variability and land use changes on global biogenic volatile organic compound emissions, *Atmos. Chem. Phys.*, 6, 2129–2146, doi:10.5194/acp-6-2129-2006, 2006.
- 30 Lathière, J., Hewitt, C. N., and Beerling, D. J.: Sensitivity of isoprene emissions from the terrestrial biosphere to 20th century changes in atmospheric CO<sub>2</sub> concentration, climate, and land use, *Glob. Change Biol.*, 24, GB1004, doi:10.1029/2009GB003548, 2010.

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- Lee, J. D., Moller, S. J., Read, K. A., Lewis, A. C., Mendes, L., and Carpenter, L. J.: Year-round measurements of nitrogen oxides and ozone in the tropical North Atlantic marine boundary layer, *J. Geophys. Res.*, 114, D21302, doi:10.1029/2009JD011878, 2009.
- 5 Lelieveld, J., Butler, T. M., Crowley, J. N., Dillon, T. J., Fischer, H., Ganzeveld, L., Harder, H., Lawrence, M. G., Martinez, M., Taraborrelli, D., and Williams, J.: Atmospheric oxidation capacity sustained by a tropical forest, *Nature*, 452, 737–740, doi:10.1038/nature06870, 2008.
- Lidster, R. T., Hamilton, J. F., Lee, J. D., Lewis, A. C., Hopkins, J. R., Punjabi, S., Rickard, A. R., and Young, J. C.: The impact of monoaromatic hydrocarbons on OH reactivity in the coastal UK boundary layer and free troposphere, *Atmos. Chem. Phys.*, 14, 6677–6693, doi:10.5194/acp-14-6677-2014, 2014.
- 10 Lockwood, A. L., Shepson, P. B., Fiddler, M. N., and Alaghmand, M.: Isoprene nitrates: preparation, separation, identification, yields, and atmospheric chemistry, *Atmos. Chem. Phys.*, 10, 6169–6178, doi:10.5194/acp-10-6169-2010, 2010.
- LRTAP Convention: Chapter 3 of the LRTAP Convention Manual of Methodologies for Modelling and Mapping Effects of Air Pollution, edited by: Mills, G., Pleijel, H., Büker, P., Braun, S., Emberson, L., Harmens, H., Simpson, D., Grünhage, L., Karlsson, P., Danielsson, H., Bermejo, V., and Gonzalez-Fernandez, I., available at: <http://icpvegetation.ceh.ac.uk/> (last access: 19 January 2015), 2010.
- 15 Martin, S. T., Andreae, M. O., Althausen, D., Artaxo, P., Baars, H., Borrmann, S., Chen, Q., Farmer, D. K., Guenther, A., Gunthe, S. S., Jimenez, J. L., Karl, T., Longo, K., Manzi, A., Müller, T., Pauliquevis, T., Petters, M. D., Prenni, A. J., Pöschl, U., Rizzo, L. V., Schneider, J., Smith, J. N., Swietlicki, E., Tota, J., Wang, J., Wiedensohler, A., and Zorn, S. R.: An overview of the Amazonian Aerosol Characterization Experiment 2008 (AMAZE-08), *Atmos. Chem. Phys.*, 10, 11415–11438, doi:10.5194/acp-10-11415-2010, 2010.
- 20 Methven, J.: Offline trajectories: calculation and accuracy, Tech. Rep. UK Univ. Global Atmos. Modell. Program, Program, Dep. of Meteorol., Univ. of Reading, Reading, UK, 1997.
- Methven, J., Arnold, S. R., O'Connor, F. M., Barjat, H., Dewey, K., Kent, J., and Brough, N.: Estimating photochemically produced ozone throughout a domain using flight data and a Lagrangian model, *J. Geophys. Res.*, 108, 4271, doi:10.1029/2002JD002955, 2003.
- 25 Murphy, J. G., Oram, D. E., and Reeves, C. E.: Measurements of volatile organic compounds over West Africa, *Atmos. Chem. Phys.*, 10, 5281–5294, doi:10.5194/acp-10-5281-2010, 2010.
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- O'Connor, F. M., Johnson, C. E., Morgenstern, O., Abraham, N. L., Braesicke, P., Dalvi, M., Folberth, G. A., Sanderson, M. G., Telford, P. J., Voulgarakis, A., Young, P. J., Zeng, G., Collins, W. J., and Pyle, J. A.: Evaluation of the new UKCA climate-composition model – Part 2: The Troposphere, *Geosci. Model Dev.*, 7, 41–91, doi:10.5194/gmd-7-41-2014, 2014.
- 5 Pacifico, F., Folberth, G. A., Sitch, S., Haywood, J. M., Rizzo, L. V., Malavelle, F. F., and Artaño, P.: Biomass burning related ozone damage on vegetation over the Amazon forest: a model sensitivity study, *Atmos. Chem. Phys.*, 15, 2791–2804, doi:10.5194/acp-15-2791-2015, 2015.
- Paulot, F., Crouse, J. D., Kjaergaard, H. G., Kuerten, A., St Clair, J. M., Seinfeld, J. H., and Wennberg, P. O.: Unexpected epoxide formation in the gas-phase photooxidation of isoprene, *Science*, 325, 730–733, 2009.
- 10 Peeters, J., Nguyen, T. L., and Vereecken, L.: HO<sub>x</sub> radical regeneration in the oxidation of isoprene, *Phys. Chem. Chem. Phys.*, 11, 5935–5939, 2009.
- Perring, A. E., Bertram, T. H., Wooldridge, P. J., Fried, A., Heikes, B. G., Dibb, J., Crouse, J. D., Wennberg, P. O., Blake, N. J., Blake, D. R., Brune, W. H., Singh, H. B., and Cohen, R. C.: Airborne observations of total RONO<sub>2</sub>: new constraints on the yield and lifetime of isoprene nitrates, *Atmos. Chem. Phys.*, 9, 1451–1463, doi:10.5194/acp-9-1451-2009, 2009.
- 15 Pissso, I., Real, E., Law, K. S., Legras, B., Bousseres, N., Attie, J. L., and Schlager, H.: Estimation of mixing in the troposphere from Lagrangian trace gas reconstructions during longrange pollution plume transport, *J. Geophys. Res.-Atmos.*, 114, D19301, doi:10.1029/2008JD011289, 2009.
- Pollack, I. B., Lerner, B. M., and Ryerson, T. B.: Evaluation of ultraviolet light-emitting diodes for detection of atmospheric NO<sub>2</sub> by photolysis – chemiluminescence, *J. Atmos. Chem.*, 65, 111–125, 2010.
- 25 Pöschl, U., von Kuhlmann, R., Poisson, N., and Crutzen, P.: Development and intercomparison of condensed isoprene oxidation mechanisms for global atmospheric modeling, *J. Atmos. Chem.*, 37, 29–52, 2000.
- Pugh, T. A. M., MacKenzie, A. R., Hewitt, C. N., Langford, B., Edwards, P. M., Furneaux, K. L., Heard, D. E., Hopkins, J. R., Jones, C. E., Karunaharan, A., Lee, J., Mills, G., Misztal, P., Moller, S., Monks, P. S., and Whalley, L. K.: Simulating atmospheric composition over a South-East Asian tropical rainforest: performance of a chemistry box model, *Atmos. Chem. Phys.*, 10, 279–298, doi:10.5194/acp-10-279-2010, 2010.
- 30

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- Pugh, T. A. M., MacKenzie, A. R., Langford, B., Nemitz, E., Misztal, P. K., and Hewitt, C. N.: The influence of small-scale variations in isoprene concentrations on atmospheric chemistry over a tropical rainforest, *Atmos. Chem. Phys.*, 11, 4121–4134, doi:10.5194/acp-11-4121-2011, 2011.
- 5 Pugh, T. A. M., Cain, M., Methven, J., Wild, O., Arnold, S. R., Real, E., Law, K. S., Emmer-  
son, K. M., Owen, S. M., Pyle, J. A., Hewitt, C. N., and MacKenzie, A. R.: A Lagrangian  
model of air-mass photochemistry and mixing using a trajectory ensemble: the Cambridge  
Tropospheric Trajectory model of Chemistry And Transport (CiTTyCAT) version 4.2, *Geosci.*  
*Model Dev.*, 5, 193–221, doi:10.5194/gmd-5-193-2012, 2012.
- 10 Real, E., Law, K. S., Weinzierl, B., Fiebig, M., Petzold, A., Wild, O., Methven, J., Arnold, S.,  
Stohl, A., Huntrieser, H., Roiger, A., Schlager, H., Stewart, D., Avery, M., Sachse, G., Brow-  
ell, E., Ferrare, R., and Blake, D.: Processes influencing ozone levels in Alaskan forest fire  
plumes during long-range transport over the North Atlantic, *J. Geophys. Res.-Atmos.*, 112,  
D10S41, doi:10.1029/2006JD007576, 2007.
- 15 Real, E., Law, K. S., Schlager, H., Roiger, A., Huntrieser, H., Methven, J., Cain, M., Holloway, J.,  
Neuman, J. A., Ryerson, T., Flocke, F., de Gouw, J., Atlas, E., Donnelly, S., and Parrish, D.:  
Lagrangian analysis of low altitude anthropogenic plume processing across the North At-  
lantic, *Atmos. Chem. Phys.*, 8, 7737–7754, doi:10.5194/acp-8-7737-2008, 2008.
- Rummel, U., Ammann, C., Kirkman, G. A., Moura, M. A. L., Foken, T., Andreae, M. O., and  
20 Meixner, F. X.: Seasonal variation of ozone deposition to a tropical rain forest in southwest  
Amazonia, *Atmos. Chem. Phys.*, 7, 5415–5435, doi:10.5194/acp-7-5415-2007, 2007.
- Sillman, S. and He, D. Y.: Some theoretical results concerning  $O_3$ - $NO_x$ -VOC chemistry and  
 $NO_x$ -VOC indicators, *J. Geophys. Res.*, 107, 4659, doi:10.1029/2001JD001123, 2002.
- 25 Sitch, S., Cox, P. M., Collins, W. J., and Huntingford, C.: Indirect radiative forcing of climate  
change through ozone effects on the land-carbon sink, *Nature*, 448, 791–794, 2007.
- Squire, O. J., Archibald, A. T., Abraham, N. L., Beerling, D. J., Hewitt, C. N., Lathière, J.,  
Pike, R. C., Telford, P. J., and Pyle, J. A.: Influence of future climate and cropland expan-  
sion on isoprene emissions and tropospheric ozone, *Atmos. Chem. Phys.*, 14, 1011–1024,  
doi:10.5194/acp-14-1011-2014, 2014.
- 30 Squire, O. J., Archibald, A. T., Griffiths, P. T., Jenkin, M. E., Smith, D., and Pyle, J. A.: Influence  
of isoprene chemical mechanism on modelled changes in tropospheric ozone due to climate  
and land use over the 21st century, *Atmos. Chem. Phys.*, 15, 5123–5143, doi:10.5194/acp-  
15-5123-2015, 2015.

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- Stone, D., Evans, M. J., Edwards, P. M., Commane, R., Ingham, T., Rickard, A. R.,  
Brookes, D. M., Hopkins, J., Leigh, R. J., Lewis, A. C., Monks, P. S., Oram, D., Reeves, C. E.,  
Stewart, D., and Heard, D. E.: Isoprene oxidation mechanisms: measurements and modelling  
of OH and HO<sub>2</sub> over a South-East Asian tropical rainforest during the OP3 field campaign,  
5 *Atmos. Chem. Phys.*, 11, 6749–6771, doi:10.5194/acp-11-6749-2011, 2011.
- Telford, P., Braesicke, P., Morgenstern, O., and Pyle, J.: Reassessment of causes of ozone  
column variability following the eruption of Mount Pinatubo using a nudged CCM, *Atmos.*  
*Chem. Phys.*, 9, 4251–4260, doi:10.5194/acp-9-4251-2009, 2009.
- Telford, P. J., Lathièrre, J., Abraham, N. L., Archibald, A. T., Braesicke, P., Johnson, C. E., Mor-  
10 genstern, O., O'Connor, F. M., Pike, R. C., Wild, O., Young, P. J., Beerling, D. J., Hewitt, C. N.,  
and Pyle, J.: Effects of climate-induced changes in isoprene emissions after the eruption  
of Mount Pinatubo, *Atmos. Chem. Phys.*, 10, 7117–7125, doi:10.5194/acp-10-7117-2010,  
2010.
- Telford, P. J., Abraham, N. L., Archibald, A. T., Braesicke, P., Dalvi, M., Morgenstern, O.,  
15 O'Connor, F. M., Richards, N. A. D., and Pyle, J. A.: Implementation of the Fast-JX Photo-  
lysis scheme (v6.4) into the UKCA component of the MetUM chemistry-climate model (v7.3),  
*Geosci. Model Dev.*, 6, 161–177, doi:10.5194/gmd-6-161-2013, 2013.
- Thompson, A., Witte, J., McPeters, R., Oltmans, S., Schmidlin, F., Logan, J., Fujiwara, M.,  
Kirchhoff, V., Posny, F., Coetzee, G., Hoegger, B., Kawakami, S., Ogawa, T., Johnson, B.,  
20 Vomel, H., and Labow, G.: Southern Hemisphere Additional Ozonesondes (SHADOZ)  
1998–2000 tropical ozone climatology – 1. Comparison with Total Ozone Mapping  
Spectrometer (TOMS) and ground-based measurements, *J. Geophys. Res.*, 108, 8238,  
doi:10.1029/2001JD000967, 2003.
- UNECE: Mapping Manual: UNECE Convention on Long-range Transboundary Air Pollution.  
25 Chapter III Mapping Critical Levels for Vegetation, 52, 2010.
- Wild, O., Law, K. S., McKenna, D. S., Bandy, B. J., Penkett, S. A., and Pyle, J. A.: Photochemical  
trajectory modeling studies of the North Atlantic region during August 1993, *J. Geophys.*  
*Res.*, 101, 29269–29288, 1996.
- WHO: Air Quality Guidelines for Europe, Tech. Rep., World Health Organization, Copenhagen,  
30 Denmark, 2000.
- WMO: Scientific assessment of ozone depletion: 1995 Global Ozone Research and Monitoring  
Project, Geneva, Switzerland, 1995.

Yienger, J. J. and Levy II, H.: Global inventory of soil-biogenic  $\text{NO}_x$  emissions, J. Geophys. Res., 100, 11447–11464, 1995.

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15, 24251–24310, 2015

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**Table 1.** Parameters subject to which mixing is explored in formulations, Mix1–3:  $\kappa_{\text{FT}}$  and  $\kappa_{\text{BL}}$  are free-troposphere and boundary-layer diffusion coefficients;  $D$  is a free tropospheric depth parameter.  $\tau_{\text{FT}}$  and  $\tau_{\text{BL}}$  are the resulting free-troposphere and boundary-layer relaxation timescales, the latter when subject to a boundary layer height (BL height) of 1000 m.

Formulation	$\kappa_{\text{FT}}$ ( $\text{m}^2 \text{s}^{-1}$ )	$\kappa_{\text{BL}}$ ( $\text{m}^2 \text{s}^{-1}$ )	$D$ (m)	$\tau_{\text{FT}}$ (h)	$\tau_{\text{BL}}$ (h)
Mix1	1.5	15.0	200	3.7	9.3
Mix2	0.5	5.0	200	11.1	27.8
Mix3	0.5	5.0	500	69.4	27.8

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**Table 2.** Main features of the eight CiTTYCAT model setups, including: the UM-UKCA integration used to initialise the composition of air parcels; the chemical mechanism employed; the trace gas emissions used; and the inclusion, or otherwise, of mixing subject to formulations Mix1–3.

CiTtYCAT model setup	Initialisation	Chemistry	Emissions	Mixing
1. CiTtYCAT (CheT)	UM-UKCA (CheT)	CheT	UKCA res	None
2. CiTtYCAT (CheT2)	UM-UKCA (CheT2)	CheT2	UKCA res	None
3. CiTtYCAT (CheT, HRE)	UM-UKCA (CheT)	CheT	High res	None
4. CiTtYCAT (CheT2, HRE)	UM-UKCA (CheT2)	CheT2	High res	None
5. CiTtYCAT (CheT, Mix1–3)	UM-UKCA (CheT)	CheT	UKCA res	Mix1–3
6. CiTtYCAT (CheT2, Mix1–3)	UM-UKCA (CheT2)	CheT2	UKCA res	Mix1–3
7. CiTtYCAT (CheT, HRE, Mix1–3)	UM-UKCA (CheT)	CheT	High res	Mix1–3
8. CiTtYCAT (CheT2, HRE, Mix1–3)	UM-UKCA (CheT2)	CheT2	High res	Mix1–3

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**Table 3.** Global trace gas emission totals (employed in all integrations), including the total anthropogenic and biogenic contributions.

Tg (species) yr <sup>-1</sup>	NO <sub>2</sub>	CO	HCHO	C <sub>2</sub> H <sub>6</sub>	C <sub>3</sub> H <sub>8</sub>	CH <sub>3</sub> C(O)CH <sub>3</sub>	CH <sub>3</sub> CHO	C <sub>5</sub> H <sub>8</sub>
Anthropogenic	99.0	875	2.37	9.85	9.46	4.35	6.54	0.0
Biogenic	17.8	0.0	4.21	0.0	0.0	39.4	14.3	401
Total	117	875	6.57	9.85	9.46	43.7	20.8	401

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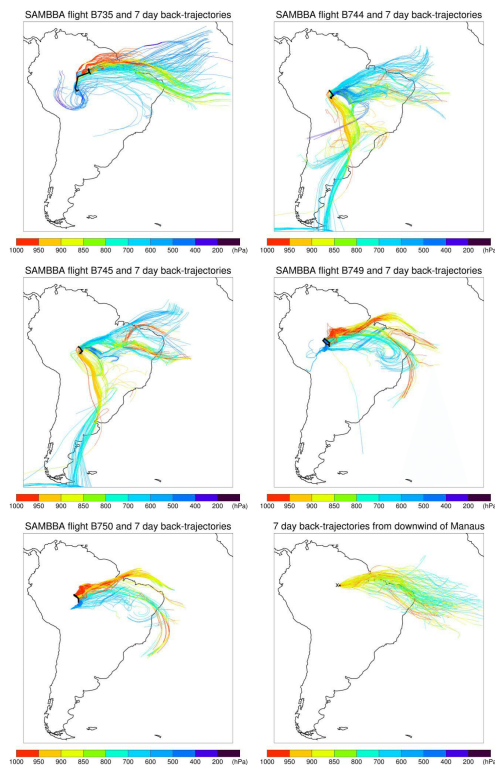
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**Table 4.** Metrics regarding the  $[O_3]$  simulated downwind of Manaus in UM-UKCA and CiTTyCAT, subject to CheT and CheT2 chemical mechanisms; all integrations with CiTTyCAT employ mixing formulation Mix3, and those labelled “HRE” employ High res emissions. See text for a discussion of the accumulation time for the reported AOT40-like metric, labelled “AOT40”.

Simulated $[O_3]$ metric	UM-UKCA		CiTTyCAT (Mix3)			
	CheT	CheT2	CheT, HRE	CheT2, HRE	CheT	CheT2
No. of times $[O_3] > 50$ ppbv	0	0	5	5	4	5
Mean $[O_3]$ (ppbv)	32.5	32.5	35.1	35.9	35.0	36.0
Increase in mean $[O_3]$ relative to UM-UKCA (subject to same chemistry)	–	–	2.6 ppbv +8 %	3.4 ppbv +11 %	2.5 ppbv +8 %	3.5 ppbv +11 %
“AOT40” (ppbv h)	22.6	17.5	863	1081	889	1091
Increase in “AOT40” relative to UM-UKCA (subject to same chemistry)	–	–	841 ppbv h 38×	1063 ppbv h 62×	866 ppbv h 39×	1073 ppbv h 62×



**Figure 1.** Flight tracks (black lines), and 7 day back-trajectories arriving on these at one minute intervals (coloured according to pressure), for SAMBBA flights B735, B744, B745, B749 and B750; see text for details. Bottom right: 7 day back-trajectories arriving at a site roughly 100 km downwind (1° west) of Manaus at 6 hourly intervals throughout September 2012 (coloured according to pressure); “x” marks the site (61.0° W, 3.1° S) and the black dot marks Manaus.

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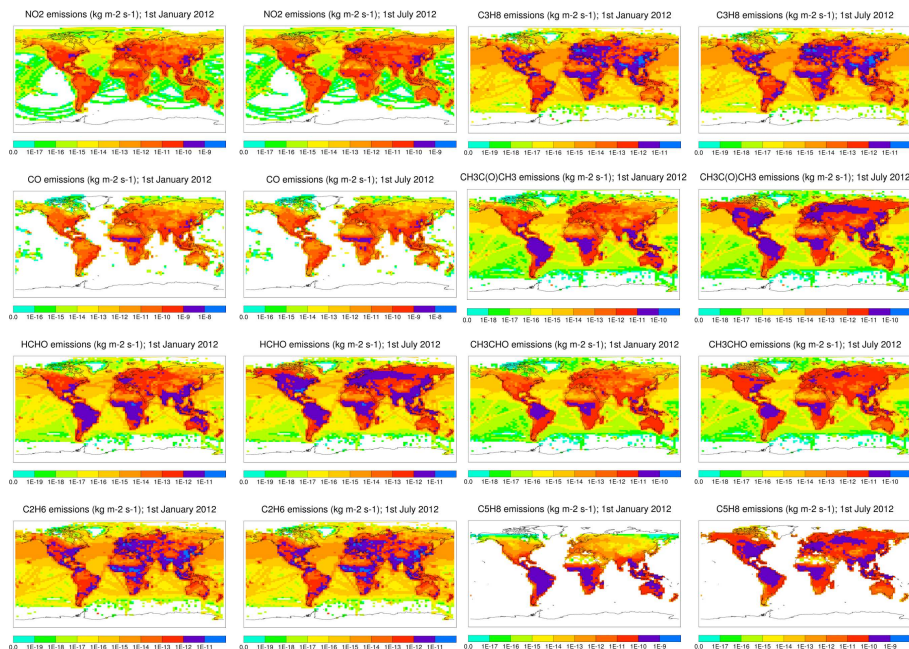
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**Figure 2.** Total (anthropogenic + biogenic) trace gas emissions at  $3.75^\circ$  longitude  $\times$   $2.5^\circ$  latitude on 1 January and 1 July 2012, employed in UM-UKCA (run in “climate mode”) and “UKCA res” integrations with CiTTyCAT.

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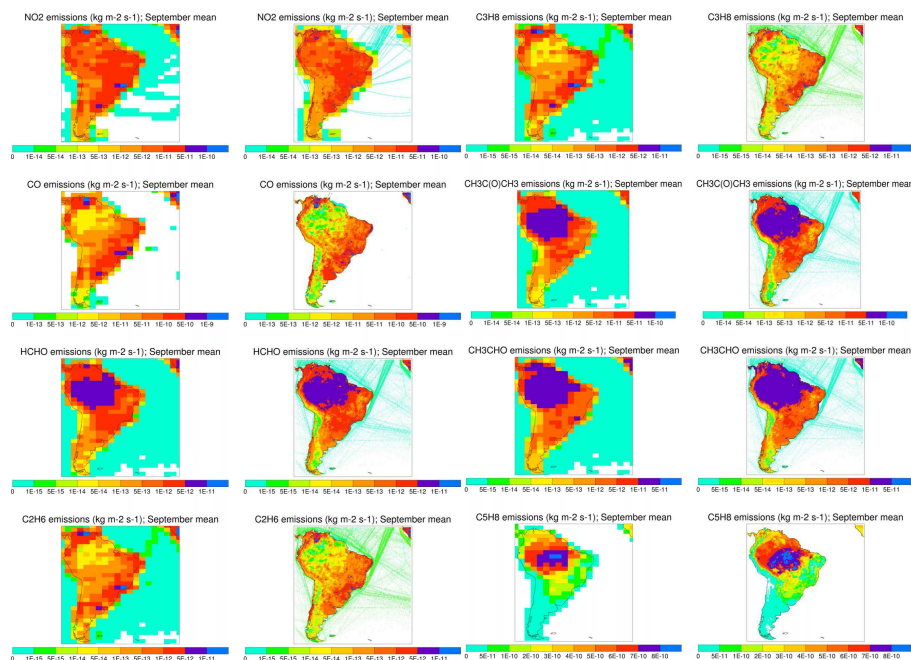
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**Figure 3.** Total (anthropogenic + biogenic) trace gas emissions employed in CiTTyCAT “UKCA res” integrations (left) and “High res” integrations (right); all emissions are September 2012 monthly means, focussing on South America.

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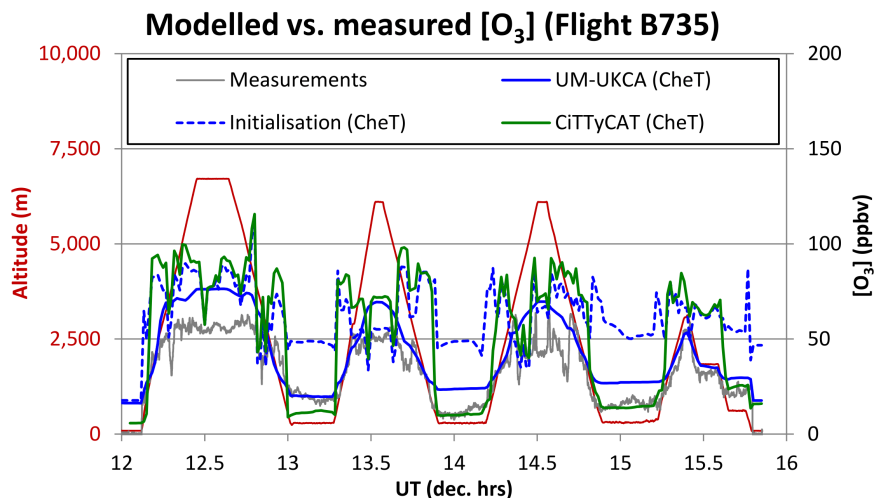
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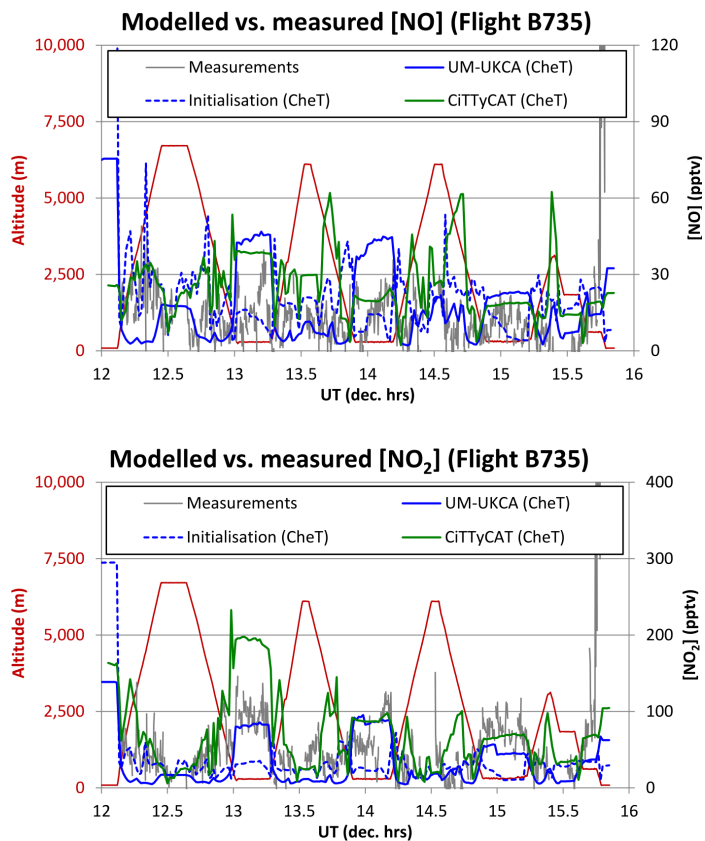


**Figure 4.**  $[O_3]$  measured on SAMBBA flight B735 and simulated in UM-UKCA (CheT) and CiTTYCAT (CheT). The  $[O_3]$  with which CiTTYCAT is initialised is also illustrated; see text for details.

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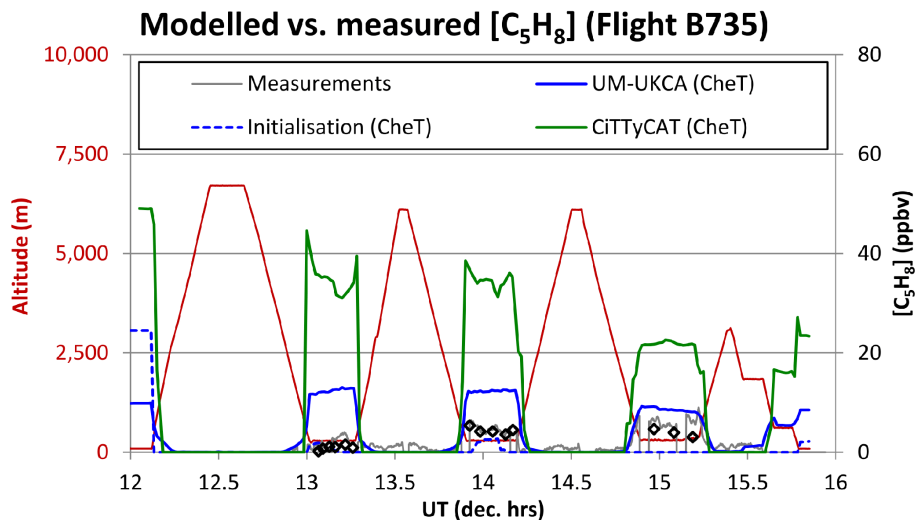


**Figure 5.** [NO] and [NO<sub>2</sub>] measured on SAMBBA flight B735 and simulated in UM-UKCA (CheT) and CiTTYCAT (CheT). The [NO] and [NO<sub>2</sub>] with which CiTTYCAT is initialised are also illustrated; see text for details.

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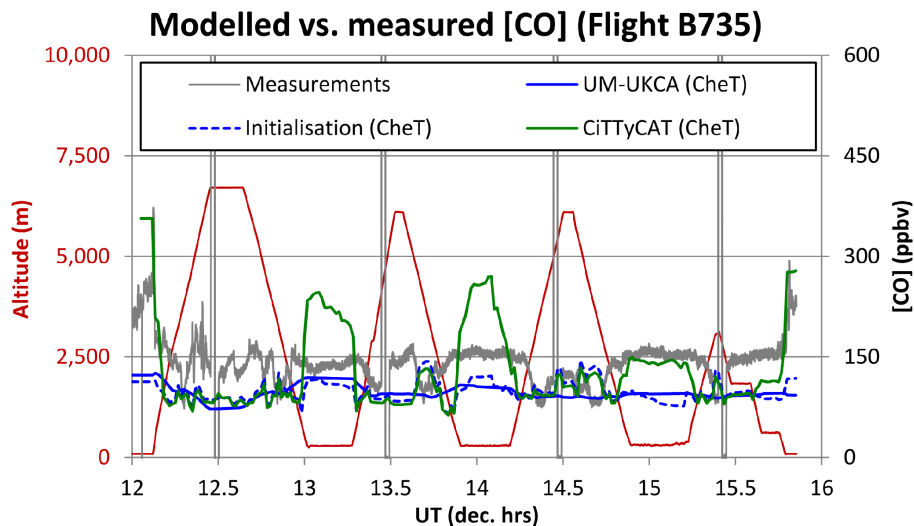


**Figure 6.**  $[C_5H_8]$  measured on SAMBBA flight B735 and simulated in UM-UKCA (CheT) and CiTTYCAT (CheT). The  $[C_5H_8]$  with which CiTTYCAT is initialised is also illustrated; the open black diamonds correspond to the  $[C_5H_8]$  measurements based on whole air samples.

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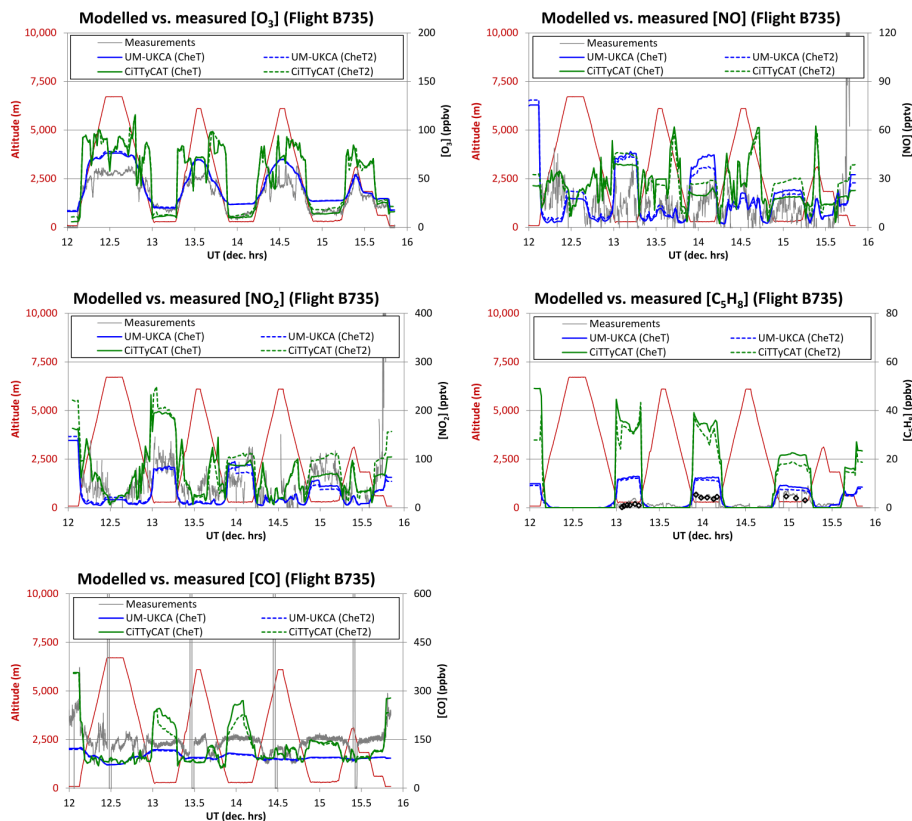


**Figure 7.** [CO] measured on SAMBBA flight B735 and simulated in UM-UKCA (CheT) and CiTTYCAT (CheT). The [CO] with which CiTTYCAT is initialised is also illustrated; see text for details.

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**Figure 8.**  $[O_3]$ ,  $[NO]$ ,  $[NO_2]$ ,  $[C_5H_8]$  and  $[CO]$  measured on SAMBBA flight B735 and simulated in UM-UKCA and CiTTYCAT, subject to CheT and CheT2 chemical mechanisms (see Sect. 2.3 for more details); the open black diamonds correspond to  $[C_5H_8]$  measurements based on whole air samples.

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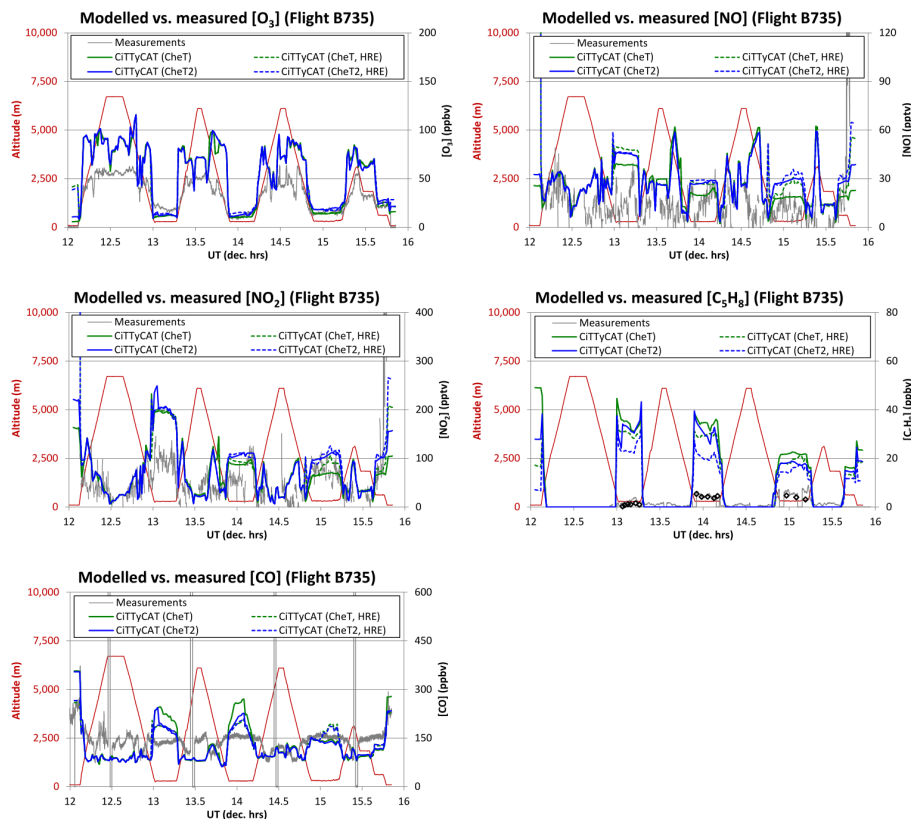
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**Figure 9.**  $[O_3]$ ,  $[NO]$ ,  $[NO_2]$ ,  $[C_5H_8]$  and  $[CO]$  measured on SAMBBA flight B735 and simulated in CiTTYCAT, subject to CheT and CheT2 chemical mechanisms (see Sect. 2.3) and both “UKCA res” emissions and “High res” emissions (HRE; see Sect. 2.4 for more details); the open black diamonds correspond to  $[C_5H_8]$  measurements based on whole air samples.

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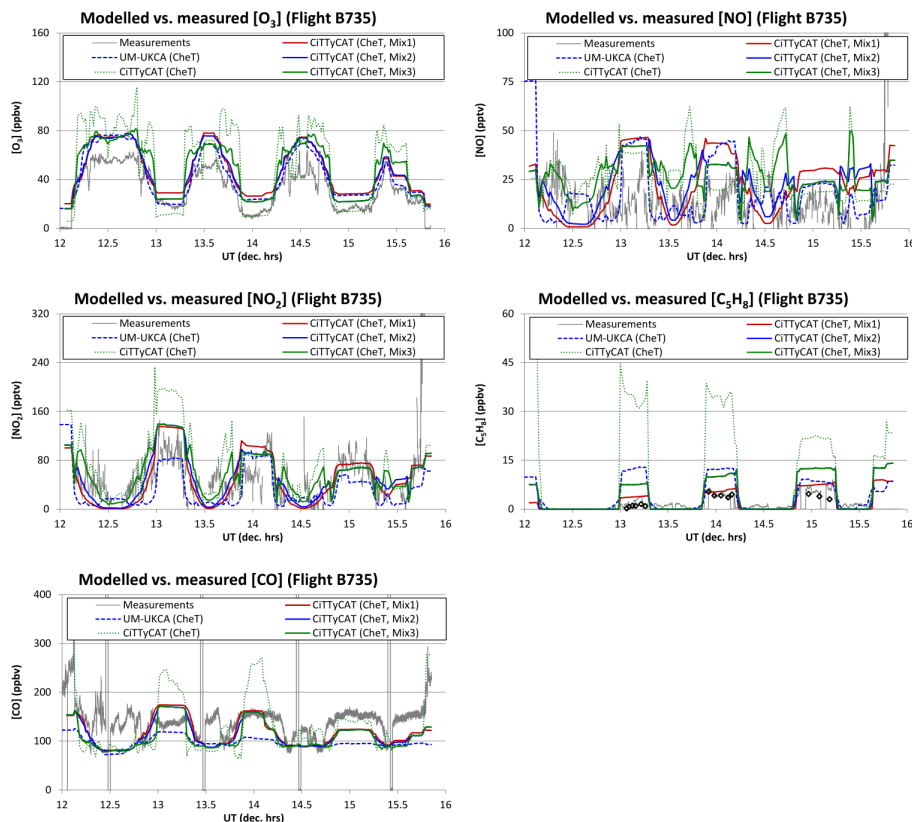
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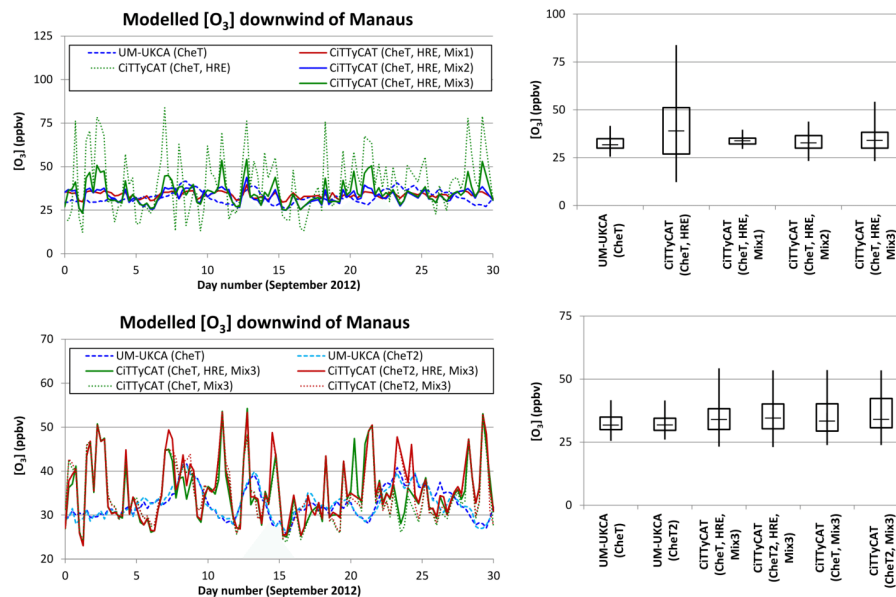
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**Figure 10.**  $[O_3]$ ,  $[NO]$ ,  $[NO_2]$ ,  $[C_5H_8]$  and  $[CO]$  measured on SAMBBA flight B735 and simulated in UM-UKCA (CheT) and CiTTyCAT (CheT), the latter subject to no mixing and, subsequently, three formulations of simple diffusive vertical mixing (relaxation towards background values) as outlined in Sect. 2.2.2 and Table 1; the open black diamonds correspond to  $[C_5H_8]$  measurements based on whole air samples.

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**Figure 11.** Left: [O<sub>3</sub>] simulated approximately 100 km downwind of Manaus in UM-UKCA and CiTTYCAT; the top panel compares CiTTYCAT integrations, all employing the CheT chemical mechanism but differing with respect to mixing formulation (Mix1–3; see Sect. 2.2.2 and Table 1 for more details), whilst the bottom panel compares CiTTYCAT integrations, all employing mixing formulation, Mix3, but differing with respect to the chemical mechanism (CheT or CheT2) and/or the resolution of trace gas emissions employed (HRE = High res ems). Right: the corresponding “box and whisker” plots of the minimum, maximum, median, and first- and third quartile [O<sub>3</sub>] values. The UM-UKCA (CheT) and CiTTYCAT (CheT, HRE, Mix3) runs are included in both top and bottom panels.