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1 **Live fast, die old: no evidence of reproductive senescence or costs**
2 **of mating in a damselfly (Odonata: Zygoptera)**

3
4
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17
18 Running head: No reproductive senescence in a damselfly

19 **Summary**

- 20 1. Recent examples of actuarial senescence in wild insect populations have challenged the
21 long-held assumption that the brevity of wild insect lifespans precludes senescence.
- 22 2. We investigate age-related patterns in mating behaviour in adults of a short-lived
23 damselfly, *Coenagrion puella*, and the implications of this mating. Using capture
24 histories for 1,033 individuals over two field seasons, we conduct both pooled and
25 stratified analyses of variations in breeding activity.
- 26 3. Pooled analyses suggest that there is strong age-related variation in the probability of
27 being present at the mating rendezvous. However, no age-related variation was observed
28 in the probability of mating. Stratified approaches confirmed a general pattern of age-
29 related declines in survival probability, but provided only equivocal evidence of an effect
30 of age on transition between temporary breeding states. Mating males and females
31 showed greater survival than non-mating individuals, possibly as a consequence of higher
32 body condition. Older males that were not currently breeding were less likely to
33 commence breeding on the next day, but showed no patterns in breeding cessation.
34 Overall, transitions between both breeding states declined with age, suggesting that males
35 that breed tend to continue breeding while those that do not breed continue to be
36 unsuccessful. Female mating rates were consistently high across all ages with no age-
37 related decline apparent.
- 38 4. While previous research has demonstrated actuarial senescence in this population, as does
39 this study, we find little evidence of either age-related declines in reproductive behaviour
40 or breeding-related declines in survival, which might indicate functional senescence or
41 costs of mating respectively. Indeed, the greater survival in mating individuals of both
42 sexes suggests that variations in individual quality may mediate both reproductive success
43 and longevity.

44 5. Contrary to recent studies, we found no compelling evidence for reproductive senescence
45 or a cost of mating in an important and well-studied model odonate. The possible link
46 between condition and ageing suggests that individual quality needs to be taken into
47 account when studying senescence. We recommend the use of multistrata models for the
48 future investigation of these phenomena.

49

50 **Keywords:** Ageing, condition, insect, natural population, reproduction, sex, multistrata
51 model.

52 **Introduction**

53 Senescence is considered to be a ubiquitous aspect of an organism's life history, whether it
54 manifests as an age-related increase in mortality ("actuarial senescence") or a decline in
55 performance ("functional senescence"). The putative ubiquity of senescence is believed to be
56 a consequence of reduced selection pressures acting on phenotypic traits later in life (Rose
57 1991). Nevertheless some studies have failed to uncover evidence of senescence (Promislow
58 1991), the relationship between age and mortality rates varies markedly between species
59 (Jones *et al.* 2014), and there is evidence that certain clonal taxa may avoid senescence
60 altogether (e.g. *Hydra*, Martínez 1998) while other taxa exhibit "negative senescence"
61 (Vaupel *et al.* 2004). While it has been proposed that senescence would be rare in wild
62 populations due to high natural rates of mortality and thus most individuals die before they
63 deteriorate in condition (Rose 1991; Kirkwood & Austad 2000), there is ample evidence of
64 both actuarial and functional senescence in long-lived wild vertebrates, and several recent
65 studies have demonstrated similar effects in short-lived invertebrates (Nussey *et al.* 2013).

66

67 The majority of previous studies of senescence have emphasised actuarial senescence
68 (Ricklefs 2010; Nussey *et al.* 2013), but this focus neglects other important components of
69 individual fitness (Partridge & Barton 1996). Age-related declines in the functioning of the
70 reproductive system ("reproductive senescence") have been demonstrated in several taxa and
71 in both sexes. For example, pup weight, pup survival, litter size, and litter frequency decline
72 with age in female meerkats (Sharp & Clutton-Brock 2010), laying date is delayed and clutch
73 size reduced in female swans (McCleery *et al.* 2008), and lamb production and weaning
74 success decline with age in bighorn ewes (Bérubé, Festa-Bianchet & Jorgenson 1999). Older
75 female cockroaches produce fewer and smaller clutches than younger cockroaches in the
76 laboratory (Moore & Moore 2001) and older male antler flies mate at a lower rate than

77 younger antler flies in the field (the only comprehensive description of reproductive
78 senescence in a wild insect population under natural conditions, and a study which was
79 limited to males, Bonduriansky & Brassil 2002; Bonduriansky & Brassil 2005). A recent
80 study of males of the damselfly *Hetaerina americana* compared “mature” and “young” males
81 (defined loosely according to wing flexibility) and showed that males in the mature age class
82 were more likely to copulate (González-Tokman, González-Santoyo & Córdoba-Aguilar
83 2013). However, that study did not investigate female reproductive senescence, observed
84 only 63 copulations from 239 males, and provided no quantitative measure of age. Age-
85 related declines in female egg loads (e.g. in female parasitoids, Mangel & Heimpel 1998) and
86 male sperm loads (e.g. in male ferrets, Wolf *et al.* 2000) represent declines in reproductive
87 function which may also cause conflict between the sexes.

88

89 While reproductive senescence involves age-related declines in one or more components of
90 reproduction, the “costs of mating” can be defined as reproduction-related declines in
91 survival or function (Daly 1978). The idea that reproduction can reduce longevity has a long
92 history (see Kirkwood 2001 for a review) and it is now clear that the specific aspect of
93 reproduction (if any) that can affect mortality varies from system to system. For example,
94 laboratory selection experiments to enhance longevity in *Drosophila melanogaster* (Meigen)
95 produced long lived individuals with relatively low early fecundity (Zwaan, Bijlsma &
96 Hoekstra 1995); egg laying increases age-related mortality in rose-hip flies, while mating
97 itself has no effect (Roitberg 1989); by contrast, the act of mating has been shown to reduce
98 female survival, for example, via the action of male seminal proteins in *Drosophila*
99 (Chapman *et al.* 1995) or genital damage in *Callosobruchus maculata* (Crudgington & Siva-
100 Jothy 2000).

101

102 While there have been a number of studies documenting age-related variations in
103 performance, such studies are complicated by the issue of variation in individual quality
104 (Vaupel & Yashin 1985; McCleery *et al.* 2008). High quality individuals may not only live
105 longer but also exhibit greater reproduction, leading to a positive relationship between age
106 and reproduction as poor-quality individuals are lost from the population (Vaupel & Yashin
107 1985; Abrams 1993; Williams & Day 2003). Such an outcome can arise even if there is an
108 individual cost of reproduction, with variation in condition masking effects of the cost of
109 reproducing.

110

111 Field studies of reproductive senescence in natural populations of insects are rare
112 (Bonduriansky & Brassil 2005), yet essential if we are to obtain a complete picture of
113 senescence in the wild. Here we present a detailed analysis of the relationship between
114 reproduction and ageing in the damselfly *Coenagrion puella* (L) (Odonata: Zygoptera) which
115 has previously been shown to exhibit actuarial senescence in the form of an age-related
116 increase in daily mortality (Sherratt *et al.* 2010). Sherratt *et al.* (2010) analysed capture
117 histories independent of the reproductive status of the individuals in question, i.e. whether
118 they were seen alone or they were engaged in tandem or in copula. Therefore, age-related
119 changes in the rate of reproduction were not investigated in this earlier study, and neither
120 were the effects of reproduction on mortality. Nevertheless, there is good reason to believe
121 that such effects might arise. In other species of odonates for example, a number of age-
122 related changes in physiology have been documented including transitions between colour
123 morphs (Sanchez-Guillen, Van Gossum & Cordero Rivera 2005), increases in muscle mass
124 during adult maturation (Marden, Fitzhugh & Wolf 1998), increases in flight-related thermal
125 sensitivity (Marden 1995), a transition from territorial to sneaker mating strategies (Forsyth
126 & Montgomerie 1987), and increased wing tatter (Banks & Thompson 1985). To date, none

127 of these changes has been linked directly either to age-related patterns in mortality or
128 reproductive function. We therefore tested whether there is an age-related decline in mating
129 behaviour (characterised by a decline in the daily rate of onset of mating and an increase in
130 the daily likelihood of cessation of mating, Figure 1A) and/or an effect of mating status on
131 daily survival (a "cost of mating", characterised by reduced survival when mating compared
132 to non-mating, Figure 1B).

133

134 **Methods**

135 *Data collection*

136 A mark-recapture study of *C. puella* was conducted over two consecutive breeding seasons at
137 an isolated (nearest neighbouring wetland >1 km), small pond (maximum dimensions 32 m
138 long, 14 m wide, and 1.5 m deep; perimeter 75 m) in southern England (50°57'39"N,
139 0°58'41"W). Full details of the study can be found in Thompson *et al.* (2011), which also
140 contain details of sources for climate (temperature, precipitation and sunshine hours) and
141 ectoparasitic mite data. Fieldwork commenced 11th May 2005 and 17th May 2006 before the
142 major flight period for the species (June-mid July) and continued until the end of July (29th
143 2005 and 30th July 2006). Three to six people surveyed the site for *C. puella* every day from
144 09:30 hours local time until 15:30-17:30 hours when no further flight activity was evident.
145 All unmarked individuals were caught and marked on first encounter, ensuring that days
146 since marking was a good approximation of days since reproductive maturity. The actual age
147 since emergence of newly marked adults may vary due to slight variations in the immature
148 period (the time between emergence from the natal water body and the return to the water
149 body as a reproductively mature adult, estimated as 13.2 (SE 0.22) days for males and 16.5
150 (SE 0.34) days for females; Banks & Thompson 1985), but this is unlikely to have significant
151 impacts on our results (for discussion, see Sherratt *et al.* 2010; Sherratt *et al.* 2011).

152

153 The average copulation duration of *C. puella* is 28 minutes, and males subsequently engage
154 in contact mate-guarding during oviposition site selection and egg-laying for an average of 84
155 minutes (Banks & Thompson 1985), making it likely that individuals engaging in
156 reproductive activity will be seen around the pond. Over the two field seasons, a total of 620
157 males and 429 females were captured and marked. Of these individuals 1,049 individuals,
158 841 (80.2%) were resighted at the study site. If an individual was seen repeatedly on a given
159 day but on at least one occasion it was seen mating then that individual would be recorded as
160 a mating individual for that day.

161

162 Capture and resighting histories for these 1,049 individuals, as singles and/or mating pairs,
163 were analysed using two methods. First, a pooled analysis was used to visualise the
164 reproduction-related patterns in the data. This approach produces straightforward results
165 based on treating all animals as if they emerged at the same time to look at what proportion of
166 individuals of each age are (i) absent from the mating rendezvous, and (ii) engaged in mating
167 behaviour. However, the pooled approach ignores individual differences (e.g. parasite
168 burden) and temporal effects (e.g. daily fluctuations in weather). Therefore, stratified models
169 were fitted using Program MARK. This “multistrata” approach allows one to control for
170 variables such as resighting probability, parasite burdens, and daily climate, while measuring
171 their effects. In particular, multistrata models allow: (i) the comparison of age-related trends
172 in survival between individuals in different states (here we compare currently breeding
173 individuals to currently non-breeding individuals), and (ii) the investigation of rates of
174 switching between states (for example, how likely is an individual to switch from not
175 breeding on one day to breeding on the next at differing ages). See below for details of
176 multistrata models.

177

178 *Pooled analysis*

179 Capture histories were pooled for each sex in each year to give a single dataset including both
180 years and both sexes. Death was conservatively assumed to have occurred on the day
181 following the final sighting. This allows the calculation, for demonstrably living individuals,
182 of (i) the proportion of individuals of both sexes at the mating rendezvous (“presence”), and
183 (ii) the proportion of present individuals that are seen mating (“mating”) on that day. First,
184 variation in the proportion of individuals present, and the proportion of individuals mating,
185 was analysed using a generalised linear mixed effects model (GLMM) in the nlme package
186 (Pinheiro *et al.* 2013) in R (R Development Core Team 2013) with sex (fixed factor), year
187 (random factor), age (covariate) and sex*age as predictors. A binomial error term was applied
188 with a response variable composed of successes and failures (present vs absent, or mating vs
189 non-mating) in two separate columns. To examine trends in more detail, data were analysed
190 separately for each sex using GLMMs with year as a random effect and age as a covariate.
191 Overdispersion was tested for by calculating the ratio of the residual deviance (Rdev) to the
192 residual degrees of freedom (Rdf) and checking that the sum of the squared residuals were χ^2
193 distributed (Venables & Ripley 2002). Overdispersion was found in models for presence in
194 the combined dataset (Rdev/Rdf=7.842, p<0.001) and for males (Rdev/Rdf=5.757, p<0.001)
195 and females (Rdev/Rdf=9.710, p<0.001) individually, and the model for mating in the
196 combined dataset (Rdev/Rdf=1.355, p=0.023). Models for mating in males
197 (Rdev/Rdf=1.141, p=0.387) and females (Rdev/Rdf=1.308, p=0.257) were not significantly
198 overdispersed. The four overdispersed models were refitted with quasi-binomial errors using
199 the glmmPQL function in MASS (Venables & Ripley 2002).

200

201 *Stratified analysis*

202 *Multistrata models*

203 The stratified models fitted using MARK (White & Burnham 1999) permit the inclusion of a
204 wide array of extrinsic (e.g. temperature, sunshine) and intrinsic (e.g. sex, mite load, age)
205 parameters to explain variation in ϕ (daily survival) and p (daily re-sighting probability, see
206 Table 1 for a full list of terms). The relationship between these parameters has been
207 established in a previous analysis (Sherratt *et al.* 2010). Here we extend Sherratt *et al.*'s
208 models to include the role of mating status using multistrata models. Briefly, multistrata
209 models build on standard MARK models by decomposing the ϕ (survival) parameter into a
210 stratum-specific survival (specifically the probability of surviving to time t_i given the stratum
211 at t_{i-1} , designated “S”) and a probability of moving between strata (designated “ ψ ”). This
212 approach makes use of an individual-based capture history that records not simply the
213 presence of an individual, but the state (or stratum) in which the animal is found. Figure 2
214 provides a schematic of how a certain capture history might arise from different combinations
215 of events, as well as the parameters that describe those patterns. We use multistrata models
216 to investigate the effects of different breeding states on rates of senescence, and how the
217 likelihood of switching between breeding states changes with age. Overall probabilities of an
218 individual occupying a given state can be calculated as the stationary states of the stochastic
219 matrices describing the transition probabilities at each time step. The results of these models
220 describe the effects of being in a particular breeding state on the response variables, but this
221 does not imply that particular individuals always occupy a given stratum: the states
222 themselves are transient but permit an insight into how particular states influence population
223 parameters as a whole.

224

225 *Model selection*

226 Multistrata models were constructed in the program MARK through the RMark package
227 (Laake 2010) in R (R Development Core Team 2013). These fitted models were then
228 compared using Akaike's information criterion corrected for small sample sizes (AICc). In
229 these models, capture histories were coded such that an individual that was not sighted on a
230 given day received a "0" for that day, an individual observed to be mating at any stage during
231 that day was given "M" ("mating") and an individual not observed to be mating but present at
232 the pond was given "N" ("non-mating").

233

234 In the first stage of modelling, we used a set of models testing for an effect of breeding
235 stratum (mating vs non-mating, designated as "strat" in the models) in each of p , S , and ψ .
236 All eight potential combinations of "dot" (i.e. floating intercept) and strata terms were used.
237 We then built on the models constructed by Sherratt *et al.* (2010), who demonstrated that
238 resighting probability (p) was best explained using climate variables (collectively designated
239 "c") and mite loads ("m"), and that survival (S) was best explained using age ("x") and mite
240 loads. To limit the potentially enormous number of models that could be explored, we held
241 constant the resighting and survival portions of the models as: $p(c+m+strat)$ and
242 $S(m+x+strat+x*strat)$ in favour of investigating the factors influencing the transition
243 probabilities between breeding states. In each model due to non-linear effects, age (x) and
244 mites (m) are accompanied by the corresponding quadratic terms (x^2 and m^2), which we omit
245 from the notation to enhance clarity. Removal of the age term from the survival model
246 resulted in a substantial decrease in explanatory power (as indicated by the difference in the
247 Akaike Information Criterion, $\Delta AICc=16.5$ compared with best model) confirming the
248 importance of age even when accounting for the variation due to breeding state. Climate data
249 were added as a unit (i.e. all three variables together) and replaced time in the resighting (p)
250 models, since it has been demonstrated that daily variations in resighting are caused by

251 meteorological variations (Sherratt *et al.* 2010). The inclusion of strata in the survival term
252 enables us to test for a difference in survival between breeding states (a “cost of mating”)
253 using the main effect, and a difference in the nature of the age vs. survival relationship
254 between breeding states using the interaction.

255

256 Finally, we explore the probability of transition between mating states in more detail to test
257 for the presence of reproductive senescence. In estimating the breeding state and sex-specific
258 transition probability ψ in a given season, the following variables were allowed to vary in all
259 their permutations: (i) mite load (number of ectoparasitic mites counted at first capture, and
260 quadratic, collectively designated as “m”) as an individual covariate, (ii) body size (length of
261 the right forewing, and quadratic, collectively designated “s”) as an individual covariate, (iii)
262 age (days since first capture, and quadratic, collectively designated “x”), and (iv) climate (the
263 daily temperature, precipitation and sunshine hours, collectively designated “c”) as time-
264 varying covariates. Mite numbers are unlikely to increase after emergence and the impacts of
265 mites are fixed after engorgement so the mite count on first capture is representative of
266 parasitic burden (Hassall *et al.* 2010). All variables had some effect on actuarial senescence
267 in a previous study (Sherratt *et al.* 2010) and the inclusion of age specifically allows us to test
268 for age-related variation in mating behaviour transitions (reproductive senescence). The four
269 variables (mites, size, age, and climate) give a total of 15 permutations, leading to 60 models
270 across the two seasons and two sexes. This count excludes the $\psi(\cdot)$ model which was already
271 tested against $\psi(\text{strat})$ above. Model averaging was then performed on the best-fitting models
272 ($\Delta\text{AICc} < 4$) to produce estimates for each parameter.

273

274 To provide a metric of reproductive success that is comparable to previous studies (Banks &
275 Thompson 1985; Banks & Thompson 1987), we calculated inter-mating interval (IMI) which

276 gives the number of days between observed matings. For female *C. puella* IMI strongly
277 correlates with clutch size, and the total number of clutches produced by a female explains
278 90% of total lifetime egg production while clutch size *per se* explains only 10% (Banks &
279 Thompson 1987). While the interpretation of IMI is more complex for males, as its
280 relationship with fitness will be determined by sperm competition and female egg loads at the
281 time of copulation, IMI still indicates a level of mating frequency that is at least partially
282 correlated with fitness. IMI for each age was calculated as the reciprocal of the equilibrium
283 value (based on the transition (ψ) matrix) of the proportion of individuals mating. Multistrata
284 recapture histories, data for the cohort analysis, individual covariates and climate data are all
285 available to download (Hassall *et al.* 2015a) along with the R code for the analysis (Hassall *et*
286 *al.* 2015b).

287

288 **Results**

289 *Pooled analysis*

290 Males were significantly more likely to be present at the mating rendezvous, but females
291 were significantly more likely to be seen mating. There was a negative main effect of age in
292 predicting both the presence of individuals at the mating rendezvous and the proportion of
293 individuals mating (Table 2). As the significant sex*age interaction complicates the
294 interpretation of the models, further GLMMs were constructed to analyse the sexes
295 separately. Older males were significantly less likely to be present at the mating rendezvous,
296 but of those males present older males were significantly more likely to mate. Older females
297 were no less likely to be present at the mating rendezvous, but of those animals present there
298 was a slight tendency for older females to mate less (Table 2). However, as can be seen from
299 Figure 3, these patterns are subtle.

300

301 *Stratified analysis – Costs of mating*

302 The addition of breeding stratum to simple models consistently improved model fit for
303 survival across both sexes and both years, suggesting a difference between breeders and non-
304 breeders in patterns of survival (Table 3, see Table S1 and Table S2 for more extensive
305 model sets). Age-related declines in survival varied more strongly between the two years of
306 the study than between the sexes, with individuals showing lower rates of actuarial
307 senescence in 2005 than in 2006. In both sexes and both years there is evidence that
308 individuals who were observed to be mating exhibited slower declines in survival than
309 individuals that were observed not mating, an effect which is clearest in 2006 (as can be seen
310 from the non-overlapping 95% confidence intervals around the survivorship curves in Figure
311 4B and D). These results run counter to the range of predictions of the cost of mating
312 hypothesis (shown in Figure 1A-C). However, the situation in 2005 is more complex. The
313 average survival probabilities for females in 2005 were consistently higher in mating
314 individuals than non-mating individuals, although females of intermediate and older ages (5-
315 20 days) showed considerable overlap in confidence intervals (Figure 4A). For males,
316 younger mating individuals have significantly greater survival until day 7 and then the 95%
317 confidence intervals for the survivorship curves begin to overlap, suggesting more rapid late-
318 life senescence in mating males and negligible senescence in non-mating males (Figure 4C).
319 Model-averaged parameters for models with $\Delta AICc < 4$ showed that being in the non-breeding
320 stratum had a negative effect in both sexes in both years, and this effect was significant and
321 negative in males in 2006 and females in both years (see Table S4). Age had a significant
322 negative relationship with survival in non-breeding males in 2005, but not breeding males.
323 Age had a non-significant, negative relationship with survival in both non-breeding males
324 and breeding males in 2005. There were no significant relationships between age and survival
325 in females when separated into breeding strata (Table S4).

326

327 *Stratified analysis – Reproductive senescence*

328 Breeding status (i.e. mating or non-mating) was also found to be important in the probability
329 of transitioning between mating states (ψ), as demonstrated by AICc (Table 3). When
330 individual and time-varying covariates were included in models, clear differences arose
331 between those parameters that were selected to explain transitions ψ in females and those
332 selected for males (Table 4). No model for female transition probability ψ with $\Delta AICc < 2$
333 contained age, suggesting that the reproductive behaviour was independent of the age of the
334 female, as was found in the pooled analysis. Figure 5A and B show that females exhibited a
335 very low probability of transition from mating to non-mating (breeding cessation) and a very
336 high probability of transition from non-mating to mating (breeding commencement). This
337 pattern of transition probabilities results in very little variation in IMIs for females (ranging
338 between 1.05 and 1.15 days, see Figure 6) that appears independent of age. Males, on the
339 other hand, showed a more complex relationship with different patterns in each year: in 2005
340 males under 10 days since capture showed a high likelihood (40-50% of stopping breeding on
341 any given day, while once they had stopped breeding they were decreasingly likely to start as
342 they aged (as indicated by the gradual decay in the white dotted line and dark shaded area in
343 Figure 5C). In 2006, non mating males were quite likely to start breeding if they were not
344 already, although this tendency decays with age as observed in 2005. However, in 2006
345 mating males were more likely to continue mating, as shown by the consistently low
346 probability of transition for mating males in Figure 5D.

347

348 Male breeding transitions ψ were best explained by models including age. The probabilities
349 of transitions from non-mating to mating (breeding commencement) show monotonic
350 declines with age from 0.319 to 0.100 in 2005 and 0.558 to 0.175 in 2006 (Figure 5). The

351 probabilities of transitions from mating to non-mating (breeding cessation) were complex and
352 varied between years. In 2005, males showed an initial increase in the probability of
353 breeding cessation between the ages of 1 and 6 days since capture. From 6 days, the
354 probability declined monotonically until 21 days after first capture, the maximum longevity
355 of males in 2005. This leads to a unimodal pattern of IMI variation with age (Figure 6). In
356 2006, males showed relatively little variation in patterns of breeding cessation over ages 1 to
357 20 days since first capture. Only 6 out of 351 males that were marked in 2006 survived to 20
358 days after first capture, so the slight increase following that age is based on very few
359 individuals. In 2005, therefore, males follow the predictions of reproductive senescence until
360 age 6 days (Figure 1C). Beyond this age, both transition probabilities decline showing that
361 individuals are more likely to remain in whatever breeding state they occupy. In 2006, most
362 individuals are breeding (as shown by a relatively high commencement rate and a low
363 cessation rate). However, over time the commencement rate declines, as predicted under
364 reproductive senescence, while the cessation rate remains low. Again, this suggests that older
365 individuals are more likely to remain in whatever breeding state they occupy. Details of
366 model-averaged parameters for models with $\Delta AICc < 4$ can be seen in Table S4. Absolute
367 probabilities of breeding, calculated from the stationary states of the transition matrices, can
368 be seen in Figure S1 with data in Table S5.

369

370 **Discussion**

371 There has been a substantial increase in the reporting of senescence from wild populations,
372 spanning a wide range of taxa (Nussey *et al.* 2013; Jones *et al.* 2014). We use a wild
373 population of a model insect, the damselfly *C. puella*, to evaluate age-related patterns in
374 breeding (reproductive senescence) and breeding-related patterns in mortality (costs of
375 mating). Using two complementary forms of analysis (pooled and stratified) we find that

376 while survival declines with age, breeding behaviour occurs at rates that are largely
377 independent of age. Females exhibit high frequencies of breeding irrespective of their age.
378 Males exhibit an age-related decline in the probability of switching from non-breeding to
379 breeding, indicative of reproductive senescence, although patterns of breeding cessation are
380 more complex. In males and females, the breeding state was associated with higher rates of
381 survival in both years, contrary to the simplest interpretation of the cost of mating hypothesis.
382 Body condition may play an important role mediating the effect here (see below). Indeed,
383 breeding state in 2006 was associated with negligible senescence, suggesting that those
384 animals that are capable to find a mate and copulate exhibit mortality rates that are
385 independent of their age. In males in both seasons, both transition probabilities
386 (commencement and cessation of breeding) decline with age, suggesting that males are more
387 likely to remain in whatever breeding state they occupied on the previous day. Taken
388 together, these findings suggest that individuals that breed successfully are of greater
389 individual quality and, as a result, are more likely to survive and breed in the future.

390

391 In a previous manuscript, we described age-related declines in survival in the same
392 population of *C. puella* (Sherratt *et al.* 2010) and these effects of age on survival remain
393 when accounting for breeding state. However, it appears that whatever processes are
394 underlying those declines in survival have only a weak effect on an individual's probability of
395 mating and this effect is only present in males. Decoupling of the age-dependent patterns in
396 mortality and reproduction may result from the scramble mating system in this species, which
397 is common to many other invertebrates (Thornhill & Alcock 1983). In scramble mating there
398 is a focus on mate finding ability, with a limited role for physical condition compared to
399 territorial or lekking species. Furthermore, since male odonates tend to surround the water
400 body that acts as a breeding rendezvous, there is as much chance of encountering a female

401 and mating for a low quality male as there is for a high quality male. As such, while poor
402 quality males may be more susceptible to predation and less capable of prey capture, it is
403 likely that the effect of functional senescence on mating success is negligible due to the low
404 correlation between male performance and mating in this system. Similarly for females,
405 which are operationally the rarer sex, mating is almost guaranteed on arrival at a mating
406 rendezvous to the point that a proportion of females in many species of odonates have
407 evolved to mimic males in order to reduce harassment (Van Gossum, Sherratt & Cordero-
408 Rivera 2008).

409

410 *Reproductive senescence*

411 Reproductive senescence has only previously been documented in natural populations of one
412 species of invertebrate, *Protophila litigata* (Bonduriansky) (Diptera: Piophilidae)
413 (Bonduriansky & Brassil 2002; Bonduriansky & Brassil 2005); a second study using the field
414 cricket *Teleogryllus commodus* (Walker) demonstrated reproductive senescence, but was
415 conducted under semi-natural conditions with enclosures (Zajitschek *et al.* 2009). Both
416 studies also demonstrated actuarial senescence. However, having demonstrated the presence
417 of actuarial senescence in *C. puella* (Sherratt *et al.* 2010) and further confirmed it here, we
418 now find no evidence of reproductive senescence in the same population. While Banks and
419 Thompson (1987) demonstrated that reproductive capacity declines with age in female *C.*
420 *puella* (inter-clutch interval increases as females age), their analysis considered only a subset
421 of females aged >5 days since first capture. Incorporating all females in the population, we
422 demonstrate no evidence for a decline in the proportion of females mating (Figure 3B, Table
423 2), no evidence for a strong effect of age on transition between breeding states (Figure 5,
424 Table 4), and negligible age-related change in IMI (Figure 6). Elsewhere, Banks and
425 Thompson (1985) provided evidence for a positive relationship between mating rates and age

426 in male *C. puella*. We demonstrate more complex patterns of age-related IMI variation in
427 males, with a unimodal relationship in 2005 and negligible change in IMI in 2006 (Figure 6).

428

429 We have found little evidence for reproductive senescence in *C. puella*, a species that is
430 effectively non-territorial. In looking more broadly for evidence of reproductive senescence
431 in odonates, work on territorial species is of relevance, particularly work investigating which
432 factors influence success in territorial disputes. With territorial success comes a thousand-
433 fold increase in mating success associated with territorial over non-territorial behaviour
434 (Plaistow & Siva-Jothy 1996), providing a potential source of age-related variation in
435 breeding. However, the effects of age on the outcome of territorial contests is inconsistent
436 (Suhonen, Rantala & Honkavaara 2008). One mechanism by which age can influence
437 competitive/reproductive success is through the gradual attrition of fat reserves after maturity.
438 Adult males build large reserves of fat prior to entering their territorial state, which are then
439 burnt during territorial contests. The amount of fat remaining determines the outcome of
440 those contests, meaning that older individuals are at a disadvantage (Marden & Waage 1990;
441 Plaistow & Siva-Jothy 1996). Where age does have an effect, it is the younger males who
442 displace older males from territories (Forsyth & Montgomerie 1987).

443

444 It is likely that there is no such role for fat in non-territorial species, given the lack of pre-
445 reproductive mass gain in non-territorial species (Anholt, Marden & Jenkins 1991).
446 However, given the fact that males do not invest in energetic reserves, it is surprising that
447 there is no age-related decline in reproductive activity. Harassment experienced by tandem
448 pairs and greater risk of predation make this a very energetically-costly behaviour (Córdoba-
449 Aguilar 2009). Despite this, in a previous study *C. puella* was shown to exhibit an increase in
450 mating rates with age from 1-6 days since first marking (Banks & Thompson 1985). This

451 relationship is absent from our pooled analysis (Figure 3), and evidence from the stratified
452 analysis suggests that, if anything, the probability of breeding cessation increases slightly
453 during this time (Figure 5). Not only do those individuals that are expending energy in these
454 activities exhibit greater mating success but they also survive better.

455

456 *Costs of mating*

457 In a previous study of *C. puella* mating there was no relationship between longevity and the
458 rate of clutch production in females, which suggests that there is no cost of mating (Banks &
459 Thompson 1987) as found here. Daily mating rate was positively correlated with longevity in
460 younger male *C. puella* (≤ 6 days since first marking), then negatively correlated with age in
461 older males (>6 days old) (Banks & Thompson 1985). Our results from 2006 suggest a
462 similar nonlinear response where breeding individuals survive better until age 12 days, at
463 which point the survivorship curves of breeding and non-breeding strata cross such that non-
464 breeders survive better beyond 12 days (Figure 4B). However, this is not present in the 2005
465 data (Figure 4A). Nevertheless, our finding of consistently higher survival in mating males is
466 consistent with the positive association of mating and longevity reported by Banks and
467 Thompson.

468

469 Mating produces a number of stressors that could conceivably increase mortality rates in both
470 sexes. Females suffer an energetic cost via harassment during scramble competition among
471 males (Bots *et al.* 2009). There are also costs associated with increased predation during
472 mating and oviposition, especially when the female submerges to oviposit (Zeiss, Martens &
473 Rolff 1999). Males also experience an increase in predation during mating, although this
474 may not be as high as in females (Michiels & Dhondt 1990). In both sexes, age is associated
475 with wing wear (Marden & Waage 1990; Siva-Jothy & Tsubaki 1994), although the effect of

476 insect wing wear on survival and flight is unclear (e.g. Hedenström, Ellington & Wolf 2001).
477 Despite the costs associated with mating, we show that mating individuals of both sexes tend
478 to exhibit greater survival probabilities over the following 24 hours than non-mating
479 individuals.

480

481 Selection on individual quality could explain both the greater survivorship in breeding
482 individuals and the absence of clear reproductive senescence in older individuals (Abrams
483 1993; Williams & Day 2003). Variation in some aspect of underlying quality may result in
484 individuals that persist in populations due to high survivorship and also breed at later ages,
485 since poor quality individuals are lost at early ages (Vaupel & Yashin 1985; Abrams 1993;
486 Williams & Day 2003). However, it is unclear how to quantify independently individual
487 quality in this species. As we discuss above, fat reserves have been identified as an important
488 measure of quality in territorial odonates (Marden & Waage 1990; Plaistow & Siva-Jothy
489 1996), but this may not play a role in species such as *C. puella* that exhibit scramble
490 competition (Stoks 2000). Arguably size could be a measure of condition, but previous work
491 has demonstrated stabilising selection on mating success with peak mating rates at
492 intermediate body sizes (Banks & Thompson 1985; Stoks 2000). Previous studies have
493 suggested that chance encounters with females, in conjunction with fluctuations in local
494 weather conditions, may form the primary determinants of male mating success (Banks &
495 Thompson 1985) and that weather may be the primary determinant of female reproductive
496 success (Thompson 1990).

497

498 In general, it has been observed that individual quality is poorly defined in the majority of
499 studies (Wilson & Nussey 2010) and it is unclear in this system why there appears to be a
500 correlation between success in multiple life history traits. Fluctuating asymmetry (FA)

501 correlates with mating rates in some species (e.g. Harvey & Walsh 1993; Cordoba-Aguilar
502 1995) but not in others (e.g. Leung & Forbes 1997; Carchini *et al.* 2000). FA is proposed to
503 be an indicator of developmental stability, or the ability of an individual to weather
504 environmental insults during ontogeny. As such, it is considered to be a measure of
505 individual quality. A complicating factor is the role of parasitic mites which have been
506 shown to affect fitness in this species (Thompson *et al.* 2011). Lifetime mating success in *C.*
507 *puella* has been shown to correlate with FA (Harvey & Walsh 1993), but this may result from
508 parasitism by mites affecting both mating and FA (Bonn *et al.* 1996). In this study,
509 survivorship of both sexes in both field seasons peaked at an intermediate number of mites,
510 indicated by a positive relationship with mites as a main effect, combined with a negative
511 relationship with mites as a quadratic term. Intermediate mite infestations were also related
512 to peak transitions in females in 2005 and 2006, and to peak transitions in males in 2005. In
513 2006, male peak transitions showed the opposite trend, with a dip in transition rates at
514 intermediate mite numbers. However, while mites were present in the best-fit model for ψ in
515 2006, the main and quadratic parameters were not significantly different from zero.

516

517 It is worth noting the additional benefits afforded by a transition-based approach to studying
518 senescence. To a certain extent this approach simply provides an alternative method of
519 visualizing (and conceptualizing) the relationship between past and future breeding and the
520 effect of age on commencement/cessation of breeding, although we believe this to be an
521 important contribution in itself. However, the ability to decompose breeding probability into
522 those animals who begin to breed and those who stop breeding allows additional insights. It
523 may be, for example, that in a territorial system animals are unable to breed until they
524 displace an existing male from a territory. Such territorial success could be age-dependent
525 and so $p(N \rightarrow B)$ increases with age. However, once a territory has been established that male

526 will remain in the territory until it expends its resources or senesces, so $p(B \rightarrow N)$ will start
527 high and decline with age. Such a pattern would manifest as an age-invariant trend in
528 $p(\text{breeding})$ if the two trends averaged one another out. Furthermore, the transitions between
529 breeding states will vary depending upon the reproductive skew of a population. We have
530 examined a damselfly system which operates based on scramble competition and so there is
531 reasonably low reproductive skew compared with territorial mating systems. However, the
532 application of the transition-based analysis to other systems could yield insights into the
533 drivers of skew and the factors predicting individual success in territorial systems (i.e.
534 covariates in the $\psi()$ term in MRR models).

535

536 Studies of senescence in natural populations of insects are relatively rare (Bonduriansky &
537 Brassil 2002; Sherratt *et al.* 2010; Sherratt *et al.* 2011). Male antler flies exhibit decreased
538 daily mating rates with increasing age (Bonduriansky & Brassil 2002), as expected from
539 theory (Bonduriansky *et al.* 2008). In contrast, we present an example of a system with male-
540 biased mortality rates (Sherratt *et al.* 2011) but where there is little evidence of a cost of
541 mating or reproductive senescence. This may result from a scramble mating system that does
542 not necessarily reward animals in good condition with more matings. Based on this study,
543 we recommend that future studies of senescence take into account breeding status using
544 stage-structured or multistrata models, since different types of individual may senesce at
545 different rates.

546

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556

557 **Data accessibility**

558 – Multistrata capture histories: figshare doi: 10.6084/m9.figshare.1422087

559 – R scripts: figshare doi: 10.6084/m9.figshare.1422127

560

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730

731

732

733 **Tables**

734 Table 1 – Abbreviations for terms used in the models.

735

Abbreviation	Term	Definition
<i>Model components</i>		
p	resighting probability	The probability of an individual being sighted
S	survivorship	The stratum-specific probability of an individual surviving
ψ	transition probability	The probability of an individual moving between strata (in this analysis: (i) breeding, and (ii) non-breeding)
<i>Covariates</i>		
.	"dot"	This indicates a fixed value for a given parameter
t	time	The day of the study.
x	age	The age of an individual on day t
m	mite burden	The number of ectoparasitic mites an individual was carrying on first capture
c	climate	Three climate variables: daily temperature, daily precipitation and daily hours of sunshine.
s	size	Length of the right fore wing
strat	breeding stratum	The breeding status of an individual ("breeding" or "non-breeding")

736

737 Table 2 – Generalised linear mixed effects models from the “pooled” analysis showing how the
 738 proportion of damselflies present at a mating rendezvous (“presence”), and the proportion of
 739 damselflies mating (“mating”) varies with age. In both models year is a random effect. All models for
 740 presence and the mating model including both males and females are fitted with quasibinomial errors
 741 to take into account overdispersion (see text for details). Parameter estimates for the “Sex” term are
 742 for males compared to females as the baseline, such that a positive value indicates that the response is
 743 higher in males. Significant terms are highlighted in **bold**.

		Presence					Mating			
		DF	Estimate	SE	t	P	Estimate	SE	Z	p
Females and males	(Intercept)	136	0.298	0.121	2.469	0.015	1.458	0.282	5.168	<0.001
	Sex	136	1.346	0.191	7.029	<0.001	-2.277	0.120	-18.959	<0.001
	Age	136	-0.015	0.010	-1.456	0.148	-0.016	0.008	-1.994	0.049
	Sex*Age	136	-0.089	0.021	-4.298	<0.001	0.060	0.014	4.330	<0.001
Females	(Intercept)	71	0.298	0.135	2.209	0.030	1.344	0.146	9.227	<0.001
	Age	71	-0.015	0.012	-1.303	0.197	-0.012	0.006	-2.008	0.045
Males	(Intercept)	64	1.642	0.129	12.761	<0.001	-0.802	0.348	-2.307	0.021
	Age	64	-0.105	0.015	-6.831	<0.001	0.043	0.009	4.563	<0.001

744

745 Table 3 – Results of MARK models showing that breeding status (strat) is important in determining
 746 resighting, survival and breeding status transitions in both sexes in both years. For the full model fits
 747 see Table S1.

			Male 2005		Male 2006		Female 2005		Female 2006	
Model			$\Delta AICc$	w_i						
p(strat)	S(strat)	$\psi(\text{strat})$	0.00	0.73	0.00	0.63	0.00	0.85	0.00	0.86
p(strat)	S(.)	$\psi(\text{strat})$	1.99	0.27	39.77	0.00	7.13	0.02	25.34	0.00
p(strat)	S(strat)	$\psi(\cdot)$	66.43	0.00	1.05	0.37	4.16	0.11	12.20	0.00
p(.)	S(strat)	$\psi(\cdot)$	69.03	0.00	152.14	0.00	126.35	0.00	261.29	0.00
p(strat)	S(.)	$\psi(\cdot)$	69.61	0.00	39.32	0.00	7.33	0.02	30.04	0.00
p(.)	S(strat)	$\psi(\text{strat})$	71.05	0.00	32.29	0.00	61.11	0.00	3.73	0.13
p(.)	S(.)	$\psi(\cdot)$	77.52	0.00	220.52	0.00	150.76	0.00	313.61	0.00
p(.)	S(.)	$\psi(\text{strat})$	79.53	0.00	92.86	0.00	82.36	0.00	24.16	0.00

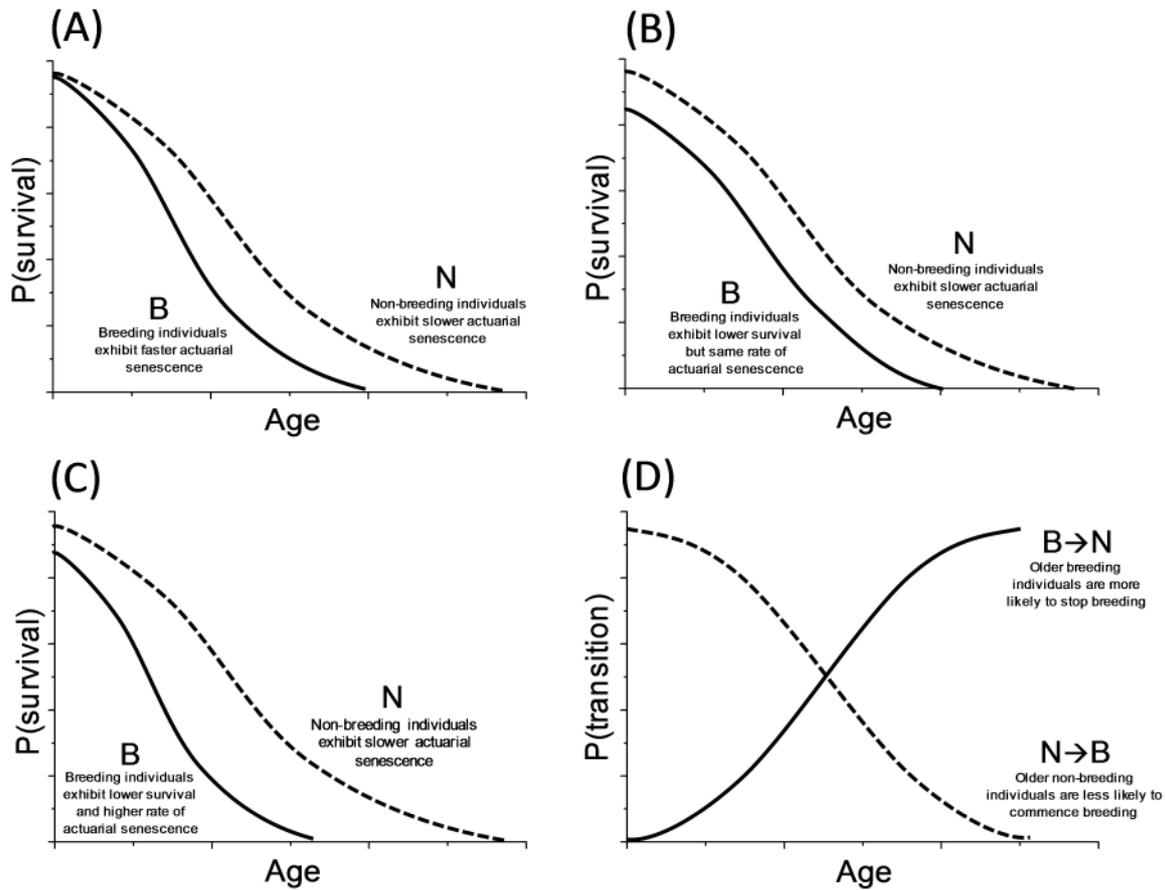
748

749 Table 4 – Results of MARK models incorporating breeding status (strat), climate data (c, comprising
750 mean daily temperature (°C), daily precipitation (mm) and hours of sunshine), individual mite burden
751 (m), size (s, indicated by fore wing length) and age (x). All models containing “x”, “s”, or “m”, also
752 contain the respective quadratic terms that have been omitted from the table for clarity. Only models
753 with $\Delta AICc < 4$ are shown – see Table S3 for the total model set. All models contained the terms
754 $p(c+m+strat)$, $S(x*strat+m+strat)$.

Sex	Year	Psi	df	logLik	AICc	delta	weight
Female	2005	strat+m	19	-1218.265	2475.670	0.000	0.511
		strat+m+s	20	-1218.231	2477.723	2.053	0.183
		strat+m+x+x*strat	23	-1215.737	2479.139	3.469	0.090
	2006	strat+c+s	22	-2723.190	5490.998	0.000	0.379
		strat+c	20	-2725.348	5491.209	0.211	0.341
		strat+c+m+s	24	-2722.822	5494.378	3.380	0.070
		strat+c+s+x+x*strat	26	-2720.791	5494.443	3.445	0.068
		strat+c+m	22	-2725.011	5494.639	3.642	0.061
		strat+c+x+x*strat	24	-2722.984	5494.703	3.705	0.059
	Male	2005	strat+c+s+x+x*strat	26	-1736.061	3525.473	0.000
strat+c+m+s+x+x*strat			27	-1735.157	3525.770	0.297	0.327
strat+c+x+x*strat			24	-1738.844	3526.840	1.367	0.191
strat+c+m+x+x*strat			26	-1737.602	3528.554	3.081	0.081
2006		strat+c+m+s+x+x*strat	27	-2558.061	5170.994	0.000	0.246
		strat+c+m+s	23	-2562.217	5171.069	0.075	0.237
		strat+c+s+x+x*strat	25	-2560.726	5172.201	1.207	0.135
		strat+c+s	21	-2564.922	5172.375	1.381	0.124

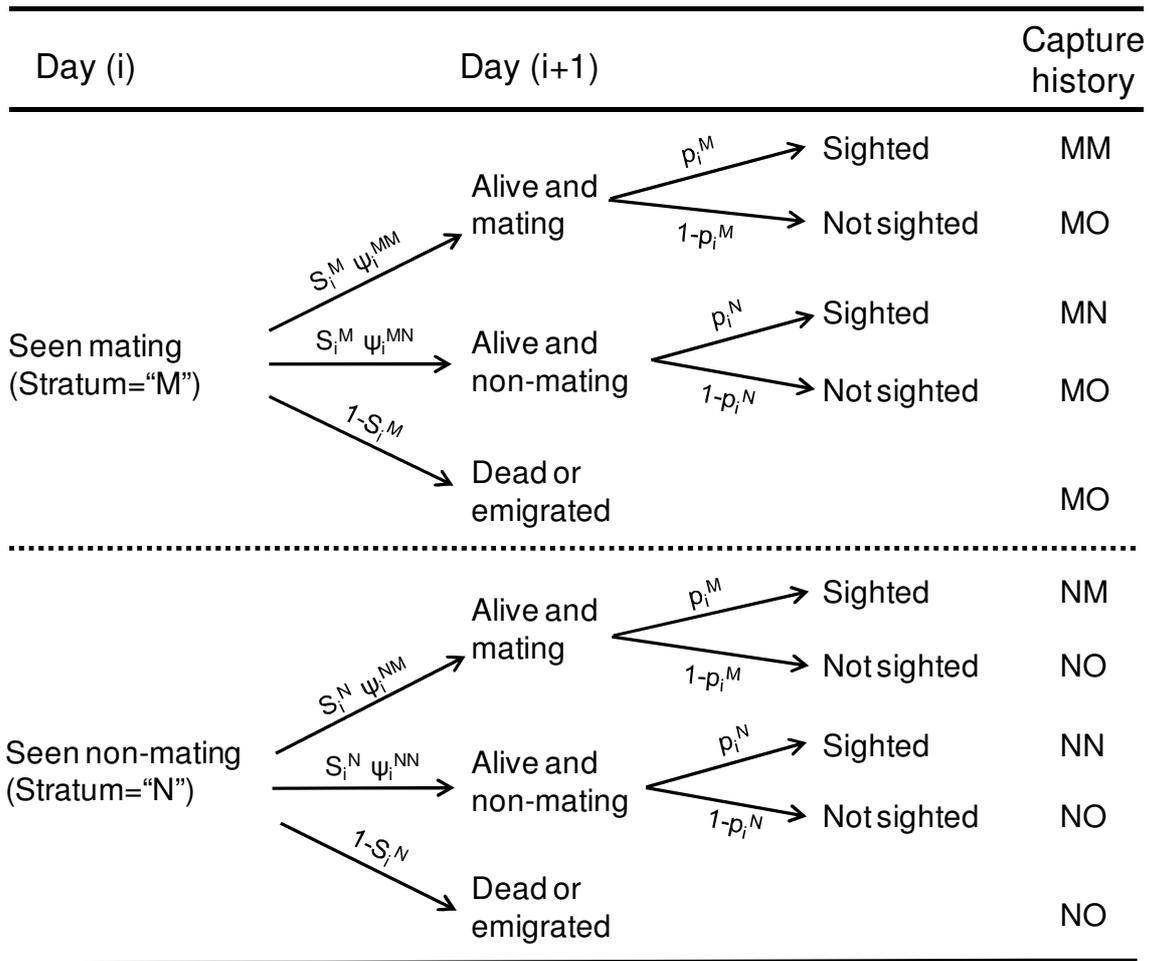
strat+c+m+x+x*strat	26	-2560.343	5173.495	2.501	0.071
strat+c+m	22	-2564.497	5173.577	2.583	0.068
strat+c+x+x*strat	24	-2562.658	5174.007	3.013	0.055
strat+c	20	-2566.900	5174.282	3.288	0.048

755



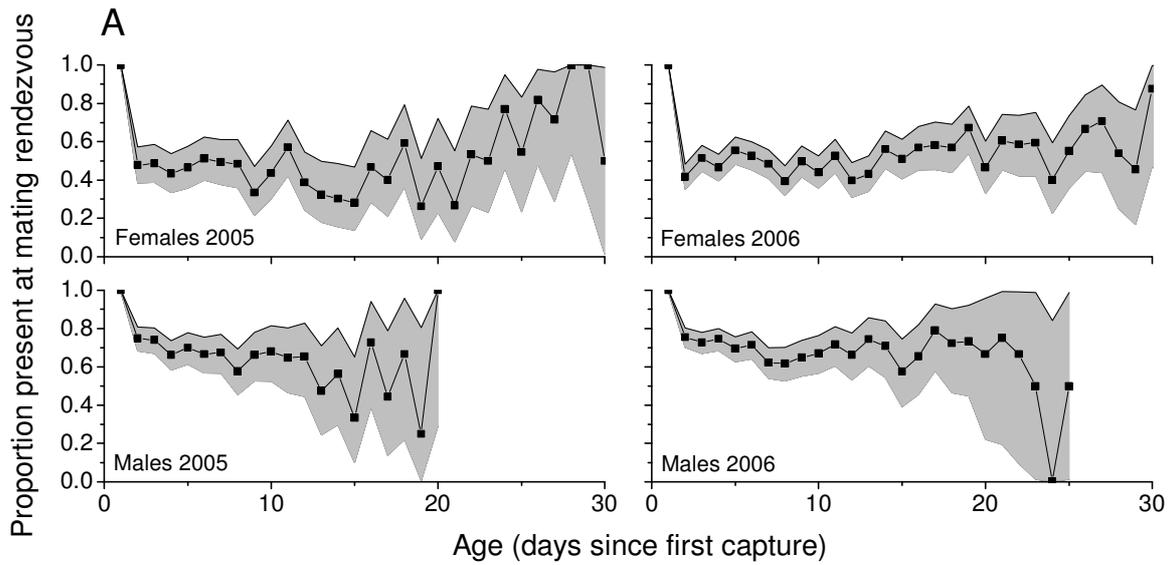
757

758 Figure 1 – Predictions of three different models of costs of sex (A-C) and reproductive senescence
 759 (D). (A) Breeding animals senesce at a greater rate than non-breeding animals, (B) breeding animals
 760 incur a fixed cost of breeding but senesce at the same rate as non-breeding animals, (C) breeding
 761 animals pay a fixed cost and senesce at a greater rate than non-breeding animals, and (D) older
 762 individuals are less likely to begin breeding (N→B) and more likely to stop breeding (B→N)
 763 ("reproductive senescence").

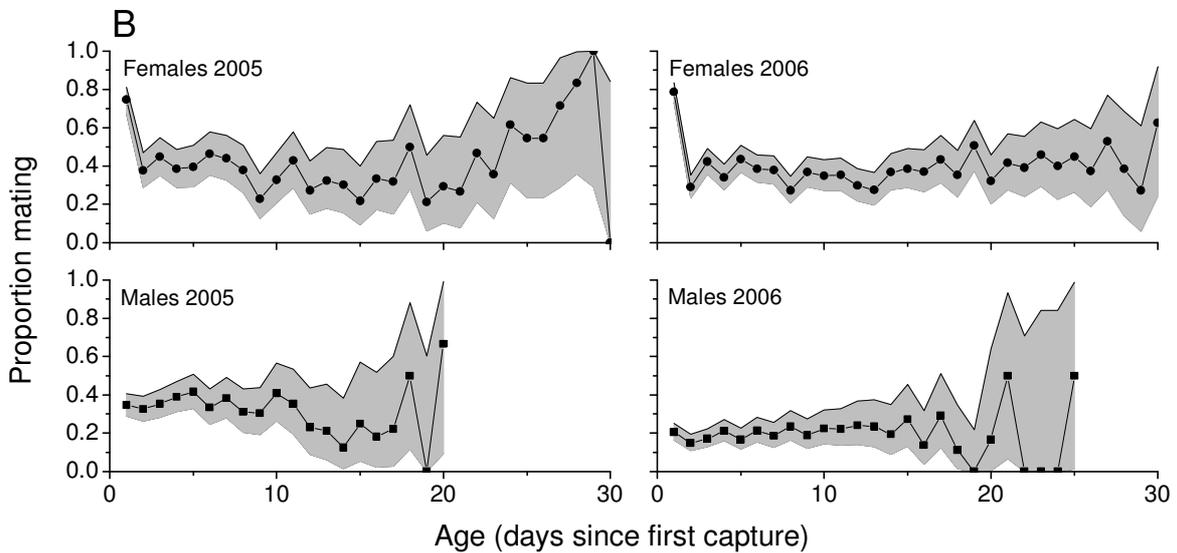


765

766 Figure 2 – An illustration of the parameters used in multistrata models (S=survival, p=resighting
 767 probability, ψ =transition probability between strata) and the capture histories that they describe
 768 (M=mating, N=non-mating, 0=absent). Note the same capture history (the sequence of interactions
 769 between the researcher and the individual animal) can arise from multiple sequences of events.
 770 Stratified modelling allows us to estimate the probability of events even when we do not directly
 771 observe an animal. Illustration based on Figure 1 in Nichols *et al.* (1994).

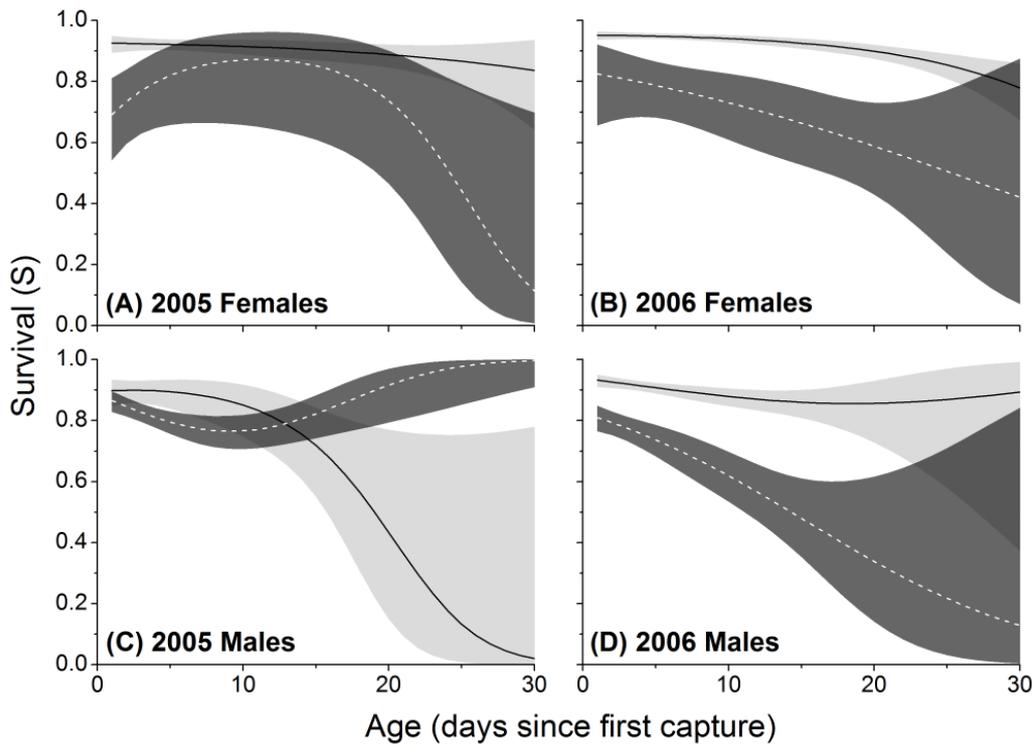


772



773

774 Figure 3 – (A) The proportion of living individuals present at the mating rendezvous, and (B) the
 775 proportion of living individuals that were observed mating in the pooled analysis. Shaded areas
 776 represent 95% Clopper-Pearson confidence intervals for the proportions.



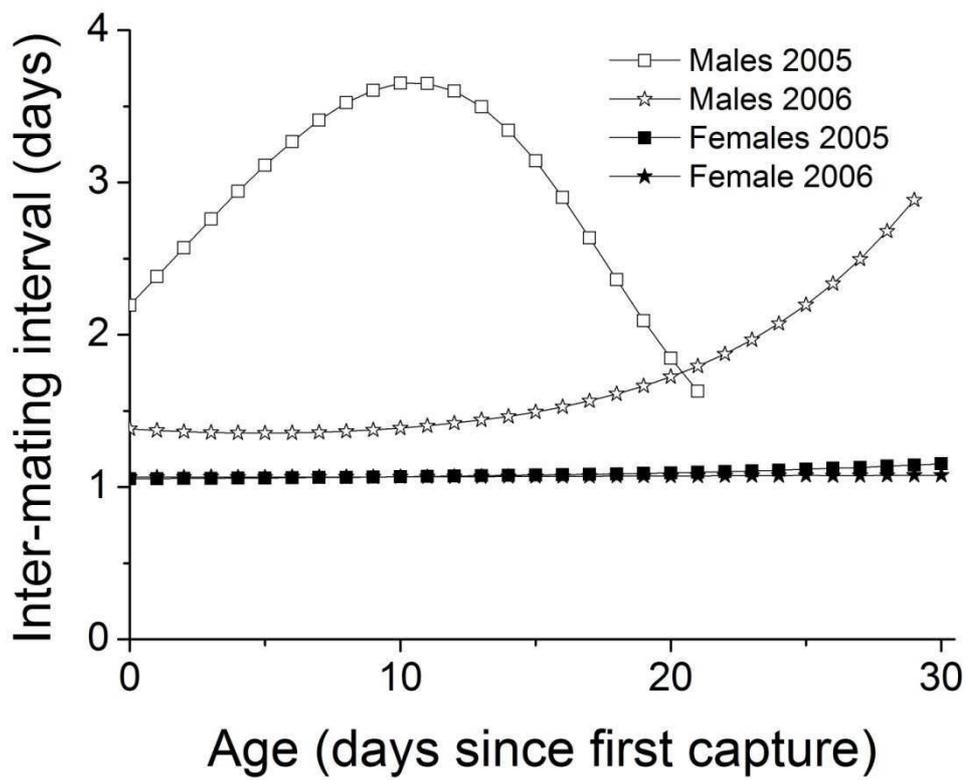
777

778 Figure 4 – Variation in age-related changes in daily survival (S) in the damselfly *Coenagrion puella*:
 779 (A) females in 2005, (B) females in 2006, (C) males in 2005, and (D) males 2006. Parameter values
 780 are taken from best fit models in Table 3. Dotted lines with black 95% confidence regions are non-
 781 breeding individuals, and solid lines with grey error regions are breeding individuals.



782

783 Figure 5 – Variation in age-related probability of transition between breeding states (Ψ) in the
 784 damselfly *Coenagrion puella*: (A) females in 2005, (B) females in 2006, (C) males in 2005, and (D)
 785 males 2006. Parameter values are taken from best fit models in Table 3. Dotted lines with black 95%
 786 confidence regions are non-breeding individuals, and solid lines with grey error regions are breeding
 787 individuals. The transition probability represents the daily probability of changing to the other
 788 breeding state (e.g. stopping breeding or starting breeding).



789 Figure 6 – Age-dependent change in inter-mating interval (time between clutches fertilised or laid) in
 790 *Coenagrion puella*.
 791