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1 **Spatially and seasonally resolved estimate of the ratio of organic matter to organic carbon**

2

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21

22 **Highlights:**

- 23 1. Simple method to estimate spatially and seasonally resolved OM/OC.
- 24 2. The OM/OC ratio can be estimated from satellite-derived NO<sub>2</sub> concentrations.
- 25 3. Parameterization developed from Aerosol Mass Spectrometer measurements.
- 26 4. OM/OC is lower in urban areas and higher in rural areas.
- 27 5. OM/OC is lower in winter and higher in summer.

28

29 **Keywords:** Organic aerosol, Organic carbon, NO<sub>2</sub>, AMS, OMI, GEOS-Chem

30

31 **Abstract**

32 Particulate organic matter (OM) is of interest for air quality and climate research, but the  
33 relationship between ambient OM and organic carbon (OC) remains ambiguous both in  
34 measurements and in modeling. We present a simple method to derive an estimate of the  
35 spatially and seasonally resolved global, lower tropospheric, ratio between OM and OC. We  
36 assume ambient NO<sub>2</sub> concentrations as a surrogate for fresh emission which mostly determines  
37 the continental scale OM/OC ratio. For this, we first develop a parameterization for the OM/OC  
38 ratio using the primary organic aerosol (POA) fraction of total OM estimated globally from  
39 Aerosol Mass Spectrometer (AMS) measurements, and evaluate it with high mass resolution  
40 AMS data. Second, we explore the ability of ground-level NO<sub>2</sub> concentrations derived from the

41 OMI satellite sensor to serve as a proxy for fresh emissions that have a high POA fraction, and  
42 apply NO<sub>2</sub> data to derive ambient POA fraction. The combination of these two methods yields an  
43 estimate of OM/OC from NO<sub>2</sub> measurements. Although this method has inherent deficiencies  
44 over biomass burning, free-tropospheric, and marine environments, elsewhere it offers more  
45 information than the currently used global mean OM/OC ratios. The OMI-derived global  
46 OM/OC ratio ranges from 1.3 to 2.1 (μg/μgC), with distinct spatial variation between urban and  
47 rural regions. The seasonal OM/OC ratio has a summer maximum and a winter minimum over  
48 regions dominated by combustion emissions. This dataset serves as a tool for interpreting organic  
49 carbon measurements, and for evaluating modeling of atmospheric organics. We also develop an  
50 additional parameterization for models to estimate the ratio of primary OM to OC from  
51 simulated NO<sub>x</sub> concentrations.

52

## 53 **1. Introduction**

54 Organic aerosols (OA) are a major constituent of fine particulate mass which affects air quality,  
55 visibility, and climate. Primary OA (POA) are mostly produced by combustion sources, while  
56 secondary OA (SOA) form through oxidation and partitioning of volatile organic compounds  
57 from anthropogenic and biogenic sources. Ambient OA is a complex mix of thousands of  
58 different organic molecules that introduces difficulties in analytical measurements, leaving large  
59 measurement uncertainty (e.g., Kanakidou et al., 2005; Jimenez et al., 2009). OA contains  
60 organic carbon (OC) as its major constituent and other elements such as oxygen, hydrogen and  
61 nitrogen, which together with OC constitute the total organic aerosol mass. Characterization of

62 OA requires spatially and seasonally resolved information about the ratio of OA with OC  
63 (OA/OC, also commonly written as OM/OC, where OM is “organic mass”, a synonym of OA).

64 Traditional ground-based impaction or filter-based instruments routinely measure OC using  
65 different analytic methods (e.g., Hand et al., 2012), but not OM due to difficulties in  
66 characterizing different components of OM. Other specific techniques that measure OM directly,  
67 such as Fourier Transform Infrared spectroscopy (FTIR) (Russell et al., 2009) and solvent  
68 extraction techniques (El-Zanan et al., 2005), are not used extensively. A wide range of literature  
69 is available on the methods to determine the OM/OC ratio, and it is a broadly debated issue. A  
70 common practice to interpret OC measurements is through the use of a continental mean value  
71 for the OM/OC ratio (such as 1.4 (Grosjean and Friedlander 1975; White and Roberts, 1977), 1.6  
72 (Malm et al., 1994), 1.4-2.1 (Turpin and Lim, 2001), and 1.8 (Hand et al., 2012) for North  
73 America). The spatial and seasonal variation in the OM/OC ratio is often neglected in the  
74 interpretation of measurement data due to insufficient information (e.g., Hand et al., 2012).

75 These global-mean values are frequently used in aerosol models to convert between POA and  
76 OM (e.g., Park et al., 2003). This is because most global and regional models simulate POA as  
77 OC. Despite notable developments for simulating the oxidative aging of OM (Simon et al.,  
78 2012), most models do not yet readily predict the OM/OC ratio. A spatially and seasonally  
79 varying estimate of the OM/OC ratio should help interpret simulated organic aerosols.

80 The OM/OC ratio is directly related to the O/C ratio in the organic mass, since the contribution  
81 of non-oxygen elements to the OM/OC ratio is generally small, and both ratios increase with  
82 chemical aging of OA (Aiken et al., 2008; Pang et al., 2006). OM/OC has seasonal and spatial  
83 variation depending on the sources of POA and SOA and their degree of aging. The Aerosol

84 Mass Spectrometer (AMS) offers quantitative determination of the size-resolved submicron OM  
85 at high temporal resolution through mass spectrometry (Jimenez et al., 2003; Canagaratna et al.,  
86 2007). Factor analysis of the AMS spectra can differentiate POA such as hydrocarbon-like OA  
87 (HOA) and several types of oxygenated OA (OOA) (Zhang et al., 2005a) that are typically SOA  
88 surrogates (Zhang et al., 2007, 2011). POA has lower OM/OC than SOA, and fresh SOA has  
89 lower OM/OC than aged SOA (Aiken et al., 2008).

90 The oxidative aging of organics may be indirectly estimated through different proxies. Both  $\text{NO}_x$   
91 and submicron POA are mainly emitted from combustion processes.  $\text{NO}_x$  is oxidized in the  
92 atmosphere with a timescale of about a day. Over regional scales OA is dominated by SOA, and  
93 anthropogenic SOA is also formed with a timescale of about a day and in amounts much larger  
94 than the originally emitted POA (DeCarlo et al., 2010). Globally most SOA may be due to  
95 anthropogenic enhancement of biogenic SOA, which is also thought to have a similar timescale  
96 of formation (Goldstein et al., 2009; Spracklen et al., 2011). Once formed, all types of SOA  
97 appear to age with a characteristic timescale of about 1-2 days, increasing the OM/OC ratio  
98 (Jimenez et al., 2009). Given the similar spatial emission patterns and timescales, we explore  
99 ambient  $\text{NO}_2$  concentrations as a surrogate for the POA/OA fraction and thus OM/OC ratios.

100 Here, we introduce a parameterization for OM/OC based on the POA fraction estimated from  
101 AMS measurements (section 2.1), test global ground-level  $\text{NO}_2$  and  $\text{NO}_x$  concentrations as a  
102 proxy for POA and thus OM/OC (sections 2.2 and 2.3), and develop a gridded dataset of the  
103 seasonally varying OM/OC ratio (section 2.4). We describe the spatially and seasonally varying  
104 satellite-derived OM/OC in section 3.

105

106 **2. Materials and methods**

107 **2.1. Primary OA fraction of the AMS data to predict OM/OC**

108 We first explore a method to derive OM/OC. For this, we use the POA fraction of the AMS data  
109 as a proxy for combustion emissions (where combustion emissions can then be determined from  
110 the measurements of co-emitted species such as NO<sub>x</sub>). Aiken et al. (2008) demonstrate a method  
111 to use high-resolution time-of-flight AMS ambient OA measurements to directly quantify the  
112 OM/OC ratio. Aiken et al. (2008) report a high correlation for OM/OC versus O/C ( $r = 0.998$ ),  
113 quantify the OM/OC ratio for urban POA as  $\sim 1.3$ , and find that for OOA the OM/OC ratio varies  
114 from 1.9 to 2.4. Per the timescale discussion above,  $f_{\text{POA}}$  (the ratio of POA, determined from  
115 factor analysis of AMS spectra (Zhang et al., 2011), to the total measured OA) is expected to be  
116 inversely related to OM/OC. Therefore, we parameterize OM/OC from AMS using  $f_{\text{POA}}$ ,

117 
$$\text{OM/OC} = 1.3 \times f_{\text{POA}} + 2.1 \times (1 - f_{\text{POA}}) \quad (1)$$

118 This method assumes that POA has an OM/OC ratio of 1.3 and OOA has a value of 2.1.  
119 However it should be noted that the OM/OC values for OOA can range from 1.9 to 2.4, and  
120 therefore our assumption introduces an uncertainty to the OM/OC estimates from the equation 1.  
121 An independent evaluation with direct OM/OC estimates from the AMS quantifies this  
122 uncertainty ( $\pm 0.2 \mu\text{g}/\mu\text{gC}$ ) as described below. We collected nine published high mass  
123 resolution campaign-mean AMS-estimated  $f_{\text{POA}}$  and AMS-measured OM/OC data from field  
124 campaigns to evaluate this parameterization. Figure 1 shows the scatter plot of AMS OM/OC  
125 and  $f_{\text{POA}}$ . It includes a representation of a typical urban POA with OM/OC ratio of 1.3 ( $f_{\text{POA}} = 1$ ).  
126 We plotted equation 1 over this scatter plot (black line). The scatter points (OM/OC) are

127 approximately within  $\pm 0.2$  ( $\mu\text{g}/\mu\text{gC}$ ) from the solid black line. This confirms the validity of the  
128 predicted relation in equation 1. Moreover, the high correlation ( $r = - 0.91$ ) supports the linear  
129 relation between AMS  $f_{\text{POA}}$  and OM/OC in the continental boundary layer across a variety of  
130 regions. However, we caution that there are few free-tropospheric, biomass burning or marine  
131 boundary layer observations in this dataset. OM/OC values can exceed 2.1 outside of the  
132 continental boundary layer (Sun et al., 2009) and from biomass burning (Turpin and Lim, 2001).  
133 Thus, this parameterization may underestimate OM/OC in those regions. Although the relation in  
134 equation 1 offers another estimate (indirect) of OM/OC over AMS locations, this relation is  
135 developed only to generate OM/OC from  $\text{NO}_x$  measurements (section 3). More reliable methods  
136 to estimate OM/OC from AMS are already available (for example, publications listed in the  
137 Figure 1 caption).

138 Spracklen et al. (2011) and Zhang et al. (2007) collected AMS POA and OOA data for 47  
139 approximately month-long observation campaigns over 37 locations. However, this dataset does  
140 not have reliable high mass resolution OM/OC measurements to evaluate equation 1.  
141 Nevertheless, this dataset has reliable POA fraction data for a broad range of locations. The  
142 prediction model in equation 1 can be applied to  $f_{\text{POA}}$  estimates of these 47 campaign mean  
143 dataset to generate an indirect OM/OC dataset. Figure 2 shows these global AMS measurements  
144 of OA (top left), POA (top right),  $f_{\text{POA}}$  (bottom left), and the derived OM/OC ratio (bottom  
145 right).  $f_{\text{POA}}$  is high over regions with fresh anthropogenic emissions. The OM/OC ratio typically  
146 ranges from 1.7 to 2.1.

147 We also considered the OM/OC ratio as estimated through a multiple regression analysis of  
148 IMPROVE data (Simon et al., 2011). However, the OM/OC ratio estimated from AMS and



149 IMPROVE measurements differs from each other ( $r = 0.4$ ,  $RMSE = 0.13 \mu\text{g}/\mu\text{gC}$ , number of  
150 points = 13), and the difference is not well understood. The O/C from the AMS is a more direct  
151 measurement, whereas the OM from IMPROVE is based on the measured OC plus extra mass  
152 needed to get closure between the total characterized mass and the total measured mass (Aiken et  
153 al., 2008; Simon et al., 2011). The inaccuracy of OC and inorganic mass measurements by  
154 IMPROVE filter techniques, and the subtraction of two large numbers introduce errors in the  
155 OM/OC calculation (Simon et al., 2011). In addition, filter measurements tend to suffer loss of  
156 semivolatile species (especially during warm seasons) and condensation of volatile organic  
157 compounds. Real time AMS measurements are less influenced by gas-particle partitioning of  
158 semivolatile species. Thus we focus exclusively on AMS data for the remainder of this  
159 manuscript.

## 160 **2.2. Nitrogen oxides as a proxy for primary OA fraction**

161 Primary organic aerosol has major sources from local combustion processes. Several studies  
162 reported the high correlation of AMS-estimated POA with nitrogen oxide species (e.g., Zhang et  
163 al. 2005b; Sun et al., 2011; Ge et al., 2012). Therefore, the fractionation of ambient OA to  
164 primary and oxygenated components can be determined indirectly from the measurements of  
165  $\text{NO}_x$ . Given the correlation between POA and  $\text{NO}_x$ , here we explore the relationship between  
166 AMS POA fraction measurements, and a spatially coincident modeled  $\text{NO}_x$  climatology to  
167 develop a proxy for the POA fraction and then the OM/OC ratio. We use the GEOS-Chem  
168 (<http://geos-chem.org>) global three-dimensional chemical transport model to test the correlations  
169 between AMS-estimated  $f_{\text{POA}}$  and co-emitted combustion related species. GEOS-Chem includes  
170 a detailed simulation of oxidant-aerosol chemistry as described in the Appendix.

171 We tested the spatial correlation between AMS-estimated  $f_{\text{POA}}$  and coincident GEOS-Chem  
172 simulations of  $\text{NO}_x$  ( $r = 0.73$ ),  $\text{NO}_2$  ( $r = 0.74$ ), hydrophobic OC ( $r = 0.42$ ), BC ( $r = 0.57$ ), and CO  
173 ( $r = 0.15$ ). The significant correlation for  $\text{NO}_x$  supports the use of this species as a proxy for the  
174 AMS POA fraction. Therefore, we assume that the typical continental POA fraction is a function  
175 of  $\text{NO}_x$  concentrations.

### 176 **2.3. Ambient primary OA fraction estimate from satellite-derived nitrogen dioxide** 177 **concentration**

178 Given the significant correlation between modeled nitrogen oxides and observed POA fraction,  
179 we further compare AMS  $f_{\text{POA}}$  with satellite retrievals of  $\text{NO}_2$  which offer finer spatial  
180 information than the model. Advancements in satellite remote sensing over the last decade yield  
181 accurate retrievals of global  $\text{NO}_2$  column concentrations at moderate spatial resolution (Boersma  
182 et al., 2011; Bucsela et al., 2013). The  $\text{NO}_2$  column concentrations are closely related to local  
183  $\text{NO}_x$  emission (Martin et al., 2003a) and ground-level  $\text{NO}_2$  concentrations (Lamsal et al., 2008).  
184 We relate these  $\text{NO}_2$  concentrations with the POA fraction at ground-level. We begin with  $\text{NO}_2$   
185 columns retrieved by Bucsela et al. (2013) for years 2005-2008 from the Dutch-Finnish built  
186 Ozone Monitoring Instrument (OMI) sensor aboard the NASA Aura satellite. We retain column  
187 data that have cloud fraction  $< 0.3$ , solar zenith angle  $< 78^\circ$  and near-nadir viewing angle (scan  
188 positions 7 to 53 out of 60).

189 We use the daily coincident ratio of column density to ground-level concentration simulated with  
190 the GEOS-Chem model to derive daily ground-level  $\text{NO}_2$  concentrations at a spatial resolution of  
191  $0.1^\circ \times 0.1^\circ$  following Lamsal et al. (2008, 2010). Ground-level  $\text{NO}_2$  concentrations derived from

192 this approach exhibit significant temporal ( $r = 0.30 - 0.96$ , mean  $r = 0.69$ ) and spatial ( $r = 0.78$ )  
193 correlation versus in situ measurements at 307 sites across North America (Lamsal et al., 2013).

194 We sampled the long-term monthly mean OMI-derived ground-level  $\text{NO}_2$  mixing ratio (at a  
195 spatial resolution of  $0.1^0 \times 0.1^0$ ) with the spatially coincident AMS-estimated POA fraction.  
196 Figure 3 shows the scatter plot of OMI  $\text{NO}_2$  versus AMS  $f_{\text{POA}}$ . Similar to model comparisons  
197 mentioned above, the correlation for these two species is significant ( $r = 0.73$ ). Excluding  
198 outliers (New York City, summer 2001; New York City, winter 2004; and Vancouver, Canada,  
199 August 2001 campaigns) increases it further up ( $r = 0.87$ ). This supports the utility of OMI  
200 nitrogen dioxide measurements to predict the fractionation of ambient OA.

201 A linear relation can be used to derive ambient  $f_{\text{POA}}$  from OMI  $\text{NO}_2$  data given their high  
202 correlation. We used the correlation statistics obtained for a reduced major axis regression  
203 between the AMS  $f_{\text{POA}}$  and OMI  $\text{NO}_2$  ( $r = 0.73$ , slope = 11.32, offset = -0.77,  $N = 47$ ) to fit a  
204 line. We then compared  $f_{\text{POA}}$  estimated from OMI  $\text{NO}_2$  (using the linear fit) against 47 AMS  $f_{\text{POA}}$   
205 data, and obtained an RMSE of 0.31 ( $r = 0.73$ , slope = 1.0, offset = 0,  $N = 47$ ). However, we note  
206 that a fast increase of  $f_{\text{POA}}$  at low  $\text{NO}_2$  concentrations is more consistent with Figure 3. OMI  $\text{NO}_2$   
207 is almost constant for  $f_{\text{POA}}$  between 0 and 0.2, and increasing thereafter. Therefore we test a non  
208 linear relation here.

209 The ground-level  $f_{\text{POA}}$  and the OMI  $\text{NO}_2$  data (ppb) follow,

$$210 \quad f_{\text{POA}} = ((\text{OMI } \text{NO}_2 - 0.2) / 15)^{0.5} \quad (2)$$

211 We reproduced the  $f_{\text{POA}}$  values from OMI  $\text{NO}_2$  data using equation 2, and compared with 47  
212 AMS  $f_{\text{POA}}$ , and obtained a high correlation ( $r = 0.85$ , slope = 1.11, offset = -0.02, RMSE = 0.25,

213 N = 47). The RMSE of 0.31 for a linear fit exceeds the RMSE of 0.25 for the non-linear fit.  
214 Therefore, equation 2 is a good approximation of the relation between the fractionation of  
215 ambient OA versus NO<sub>2</sub> concentrations over a broad region.

#### 216 **2.4. Global OM/OC from satellite-derived and modeled nitrogen oxides**

217 Having predicted (equations 1 and 2) and evaluated the relation for AMS OM/OC versus AMS  
218 f<sub>POA</sub> (section 2.1), and AMS f<sub>POA</sub> versus OMI NO<sub>2</sub> (section 2.3), it is straightforward to derive  
219 seasonally varying maps of the global OM/OC ratio. Equation 1 and 2 leads to an empirical  
220 OM/OC prediction model as,

$$221 \quad \text{OM/OC} = 2.1 - 0.8 \times ((\text{OMI NO}_2 - 0.2) / 15)^{0.5} \quad (3)$$

222 We use monthly mean OMI-derived ground-level NO<sub>2</sub> concentrations (ppb) for 2005-2008 to  
223 produce a global climatology of the OM/OC ratio using equation (3). For NO<sub>2</sub> < 0.2, the OM/OC  
224 ratio is 2.1 which is consistent with a ratio over regions which lack primary emission. We  
225 imposed a lower limit of 1.3 to avoid unrealistic (f<sub>POA</sub> exceeding 1) values for extremely high  
226 NO<sub>2</sub> conditions. A cross validation of the OMI-derived long-term monthly mean OM/OC  
227 product with the 47 AMS- derived OM/OC (from AMS f<sub>POA</sub> using equation 1) data shows an  
228 RMSE of 0.06 μg/μgC (r = 0.85, slope = 1.11, offset = -0.21, N = 47).

229 Similarly, we create a prediction model from the GEOS-Chem NO<sub>x</sub> concentrations,

$$230 \quad \text{OM/OC} = 2.1 - 0.8 \times ((\text{NO}_x - 0.5) / 18)^{0.6} \quad (4)$$

231 The GEOS-Chem  $\text{NO}_x$  estimated OM/OC also has significant agreement ( $r = 0.74$ , slope = 1.08,  
232 offset = -0.15, RMSE =  $0.08 \mu\text{g}/\mu\text{gC}$ ) with the AMS data. Hence, model simulations of  $\text{NO}_x$  also  
233 can be used as a proxy for OM/OC.

### 234 **3. Discussion of the global OM/OC ratio**

235 Figure 4 shows seasonal maps of the predicted OM/OC ratio based on the OMI-derived  $\text{NO}_2$   
236 concentration climatology. Maximum OM/OC ratios above 1.9 are found in regions with low  
237  $\text{NO}_x$  emissions. Urban and industrial regions of the Northern Hemisphere tend to have lower  
238 OM/OC ratios. Fresh urban OA emission from transportation leads to low values for OM/OC in  
239 Eastern North America. The southeast U.S. with biogenic SOA production has high OM/OC  
240 values. Biomass burning regions have high predicted OM/OC ratios which are coincidentally  
241 consistent with expectations; future work should extend this parameterization to explicitly  
242 represent the enhanced OM/OC from fires. Biases due to soil  $\text{NO}_x$  emissions are not apparent,  
243 due to the diffuse nature of the source. Annual mean values for urban ( $\sim 1.5 \mu\text{g}/\mu\text{gC}$ ) and rural  
244 ( $\sim 2.0 \mu\text{g}/\mu\text{gC}$ ) regions are broadly consistent with the values recommended by Turpin and Lim  
245 (2001). FTIR spectroscopy methods yield annual mean values of 1.9-2.0  $\mu\text{g}/\mu\text{gC}$  over both  
246 Pittsburg (Polidori et al. 2008) and a rural site in Hungary (Kiss et al. 2002), which close to our  
247 regional estimate of  $1.8 \mu\text{g}/\mu\text{gC}$ .

248 The OM/OC ratio has a winter minimum and summer maximum over polluted regions. The high  
249 summer OM/OC in this dataset reflects lower  $\text{NO}_x$  and  $f_{\text{POA}}$  in summer compared to winter. In  
250 general, high temperatures and sunlight in summer lead to high oxidant availability and enhanced  
251 photochemistry. The resulting oxidative aging of POA and enhanced SOA formation increase the

252 summertime OM/OC ratio compared to winter. The lower winter OM/OC is due to lower oxygen  
253 content in the aged OA. Simon et al., (2011) also find similar seasonality for the OM/OC ratio  
254 despite differences in magnitude. Xing et al. (2013) used chemical analysis of organic  
255 compounds extracted from PM<sub>2.5</sub> to find OM/OC ratios for 14 Chinese cities in summer of  
256  $1.75 \pm 0.13 \mu\text{g}/\mu\text{gC}$  compared with our estimate of  $\sim 1.8 \mu\text{g}/\mu\text{gC}$ , and in winter of  
257  $1.59 \pm 0.18 \mu\text{g}/\mu\text{gC}$  compared with our estimate of  $\sim 1.6 \mu\text{g}/\mu\text{gC}$ . Hence, we conclude that the  
258 global annual mean values for urban and rural regions, and its seasonal variation over urban  
259 regions are broadly consistent with the values recommended by several other studies.

#### 260 **4. Conclusions**

261 We developed a simple technique to estimate the spatial and seasonal variation of the global  
262 OM/OC ratio. OMI NO<sub>2</sub> observations were used to represent air mass age and scaled to AMS  
263 observations. The resultant dataset appears to generally represent the OM to OC ratio in most of  
264 the continental boundary layer. Underestimates are expected in biomass burning, marine and free  
265 tropospheric environments. Nonetheless the dataset should serve as an improvement over the  
266 commonly used global values for OM/OC. This work would benefit from more AMS  
267 measurements of the OM/OC ratio across multiple seasons and regions. Ongoing efforts to  
268 explicitly simulate oxidative aging of OA should ultimately yield a more complete representation  
269 of OM/OC.

#### 270 **Appendix**

271 The GEOS-Chem (version 9-01-03; <http://geos-chem.org>) is a global three-dimensional chemical  
272 transport model driven with assimilated meteorological data from the Goddard Earth Observing

273 System (GEOS-5) at the NASA Global Modeling Assimilation Office (GMAO). We conduct our  
274 simulations at a spatial resolution of  $2^0 \times 2.5^0$  for the years 2005-2008. The lowest layer of the  
275 model is approximately 130 meters with a total of 47 levels from the Earth's surface to the top of  
276 the atmosphere. We use full mixing of species below the mixed layer depth, with a correction to  
277 the GEOS-5 predicted mixed layer depth (Heald et al., 2012; Walker et al., 2012). We use a  
278 timestep of 15 minutes for dynamical processes (transport and convection), and a timestep of 60  
279 minutes for chemical processes and emissions.

280 GEOS-Chem contains a detailed simulation of  $\text{HO}_x$ - $\text{NO}_x$ -VOC-ozone-aerosol chemistry (Bey et  
281 al., 2001; Park et al., 2004). Mao et al. (2010) describe the GEOS-Chem chemical mechanism in  
282 detail. The simulation of secondary inorganic ions is directly coupled with gas phase chemistry  
283 (Park et al., 2004). Aerosol-gas interactions are simulated through heterogeneous chemistry  
284 (Jacob, 2000) with updated aerosol uptake of  $\text{N}_2\text{O}_5$  (Evans and Jacob, 2005) and  $\text{HO}_2$  (Thornton  
285 et al., 2008), aerosol extinction effects on photolysis rates (Martin et al., 2003b), the partitioning  
286 of aerosols from gas using the ISORROPIA II thermodynamic scheme (Fontoukis and Nenes,  
287 2007) as implemented by Pye et al. (2009), and gas-aerosol partitioning of semivolatile products  
288 of VOC oxidation (Henze et al., 2006, 2008; Liao et al., 2007).

289 Anthropogenic  $\text{NO}_x$  emission inventories are from the EDGAR v32-FT2000 global inventory for  
290 2000 (Olivier et al., 2005), with regional overwrites over the U.S. (Environmental protection  
291 Agency-National Emissions Inventory 2005), Canada (CAC; <http://www.ec.gc.ca/pdb/cac/>),  
292 Mexico (BRAVO; Kuhns et al., 2005), Europe (EMEP; <http://www.emep.int/>), and East Asia  
293 (Zhang et al., 2009). Emissions are scaled to subsequent years on the basis of energy statistics  
294 (van Donkelaar et al., 2008) and after 2006 with OMI  $\text{NO}_2$  data (Lamsal et al., 2011). Other

295 major non-anthropogenic land NO<sub>x</sub> emissions include biomass burning emissions (GFED-3; Mu  
296 et al., 2011) and soil emissions (Yienger and Levy, 1995; Wang et al., 1998).

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612 **Figure Captions**

613 Figure 1: Scatter plot of Aerosol Mass Spectrometer (AMS) measured OM/OC ratio, and the  
614 corresponding estimated primary organic aerosol fraction ( $f_{\text{POA}}$ ) from several field campaigns  
615 around the globe. The dots and the corresponding numbers on the plot represent AMS field  
616 campaign mean values for (1) Whistler, Canada (Sun et al., 2009); (2) Long Island (Zhou et al.,  
617 in preparation); (3) CARES T1 (Setyan et al., 2012); (4) Pearl River Delta, China (Huang et al.,  
618 2011); (5) SOAR 1 (Docherty et al., 2011); (6) New York City (Sun et al., 2011); (7) CalNex,  
619 Los Angeles (Hayes et al., 2013); (8) Beijing, China (Huang et al., 2010); (9) Fresno, California  
620 (Ge et al., 2012); and (10) typical urban POA (Jimenez et al., 2009). The solid black line is  
621 equation 1 (also in inset).

622 Figure 2: Aerosol Mass Spectrometer measured organic aerosol (OA), estimated primary OA  
623 (POA), POA fraction ( $f_{\text{POA}}$ ), and predicted organic mass (OM) to organic carbon (OC) ratio  
624 using equation 1 (data source: Spracklen et al. 2011 and Zhang et al. 2007).

625 Figure 3: Scatter plot of estimated primary organic aerosol fraction ( $f_{\text{POA}}$ ) from the Aerosol Mass  
626 Spectrometer (AMS) versus ground-level  $\text{NO}_2$  concentrations derived from the OMI satellite  
627 instrument. The circled dots and the corresponding numbers represent outliers. The dots/numbers  
628 from 1 to 3 represent mean values for the New York City, Summer 2001; New York City,  
629 Winter 2004; and Vancouver, Canada, August 2001 campaigns respectively. The solid black line  
630 is equation 2 (also in inset).

631 Figure 4: Seasonal OM/OC ratio estimated from the prediction model (equation 3) using the  
632 OMI-derived ground-level  $\text{NO}_2$  concentration. Seasons are defined as December-January-



633 February (DJF), March-April-May (MAM), June-July-August (JJA), and September-October-  
634 November (SON).

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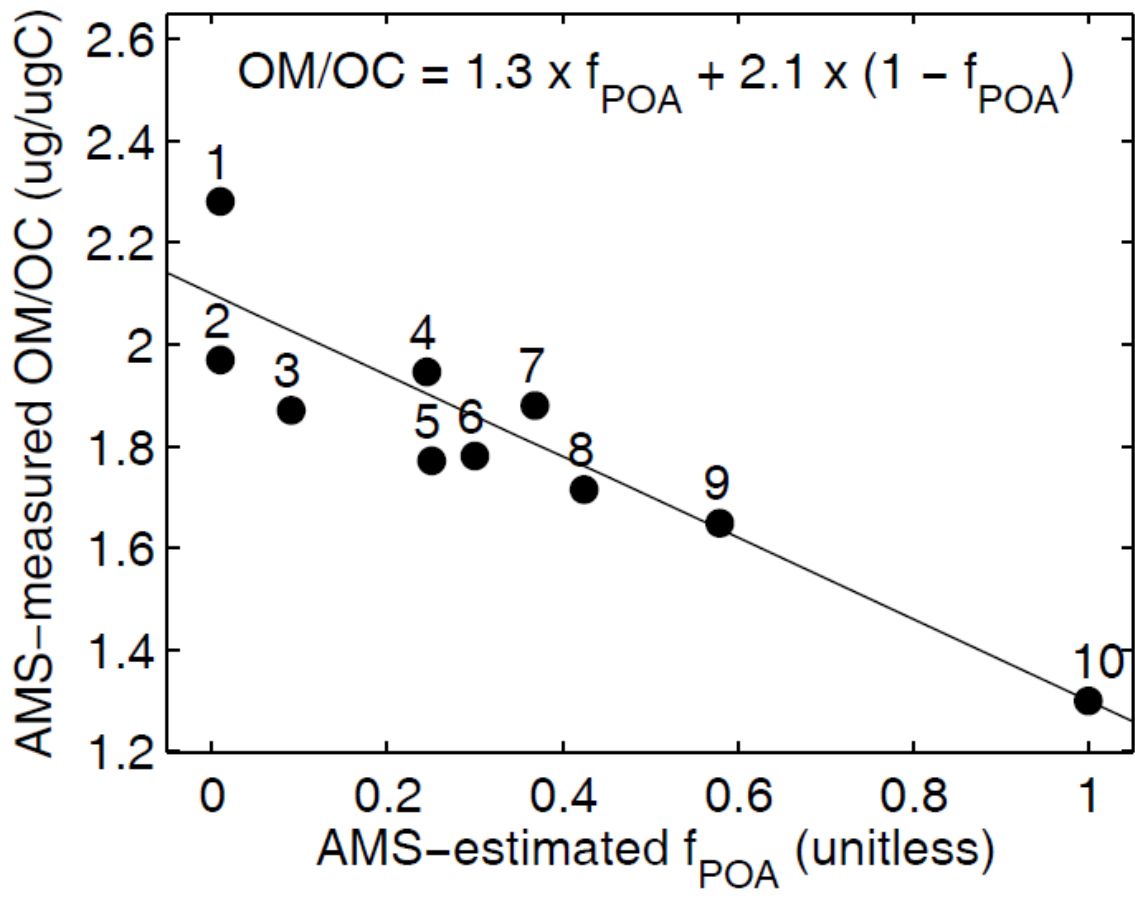
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651 Figure 1.

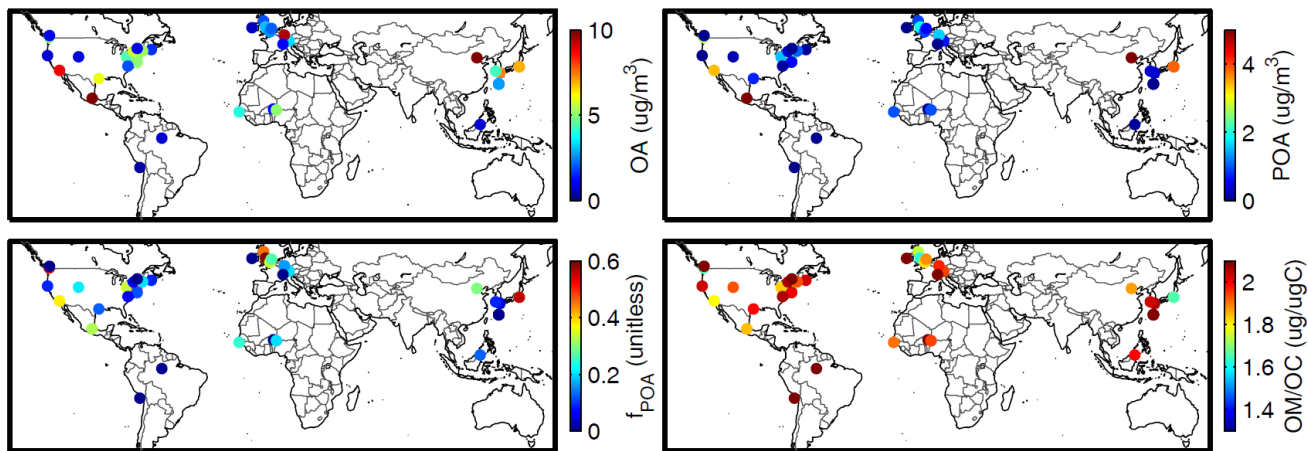
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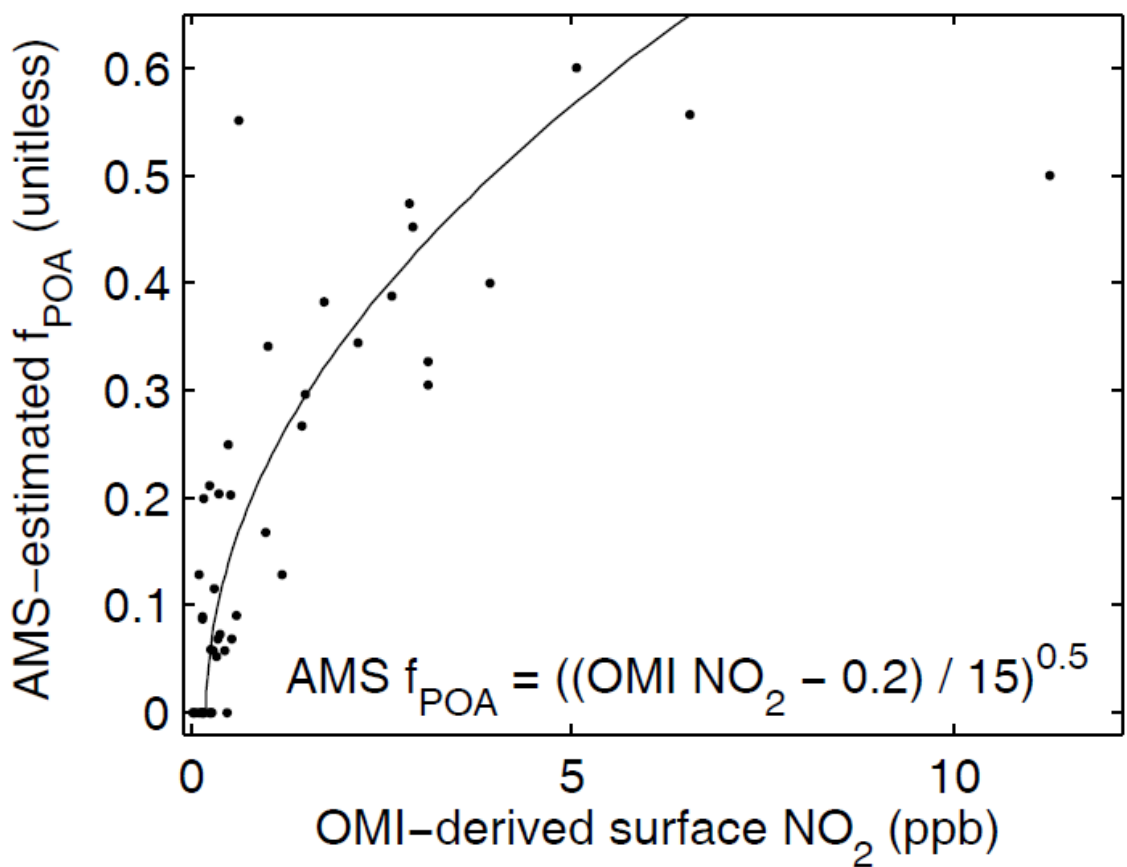
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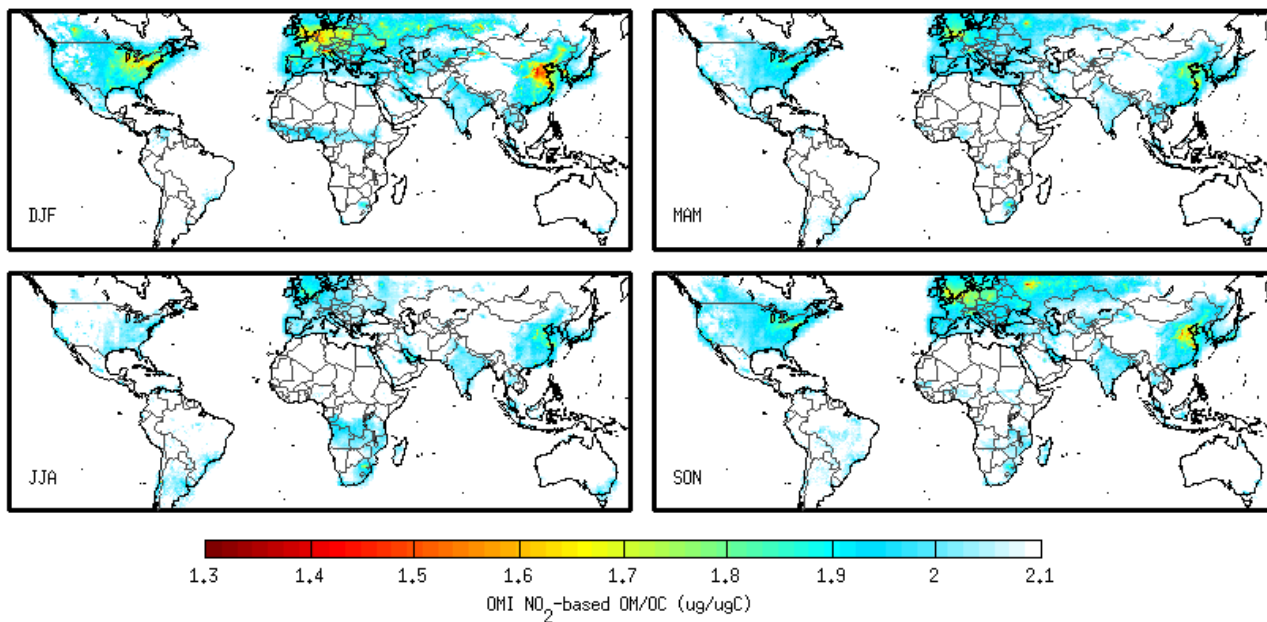
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678 Figure 4.

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