eprints@whiterose.ac.uk
https://eprints.whiterose.ac.uk/

Zinc Stable Isotope Fractionation Mechanisms during Adsorption on and Substitution in Iron (Hydr)oxides

Xinran Yan[1], Wei Li[2], Chuanwei Zhu[3], Caroline L. Peacock[4], Yizhang Liu[3], Hui Li[5], Jing Zhang[6], Mei Hong[7], Fan Liu[1], Hui Yin[1,*]

[1]Key Laboratory of Arable Land Conservation (Middle and Lower Reaches of Yangtse River), Ministry of Agriculture and Rural affairs, College of Resources and Environment, Huazhong Agricultural University, Wuhan 430070, China

[2]Key Laboratory of Surficial Geochemistry, Ministry of Education, School of Earth Sciences and Engineering, Nanjing University, Nanjing 210023, China

[3]State Key Laboratory of Ore Deposit Geochemistry, Institute of Geochemistry, Chinese Academy of Sciences, Guiyang 550002, China

[4]School of Earth and Environment, University of Leeds, Leeds, LS2 9JT, UK

[5]Department of Crop and Soil Sciences, North Carolina State University, Raleigh, North Carolina 27695, USA

[6]Beijing Synchrotron Radiation Facility, Institute of High Energy Physics, Chinese Academy of Sciences, Beijing 100039, China

[7]College of Grassland, Resources and Environment, Inner Mongolia Agricultural University, Hohhot 010018, China

Corresponding Authors:

Tel.: +86 27 87280271. Fax: +86 27 87288618.

Email: yinhui666@mail.hzau.edu.cn.

ABSTRACT: Zn isotope fingerprint is widely used as a proxy of various environmental geochemical processes, so it is crucial to determine which are the mechanisms responsible for isotopic fractionation. Iron (Fe) (hydr)oxides greatly control the cycling and fate and thus isotope fractionation factors of Zn in terrestrial environments. Here, Zn isotope fractionation and related mechanisms during adsorption on and substitution in three FeOOH polymorphs are explored. Results demonstrate that heavy Zn isotopes are preferentially enriched onto solids, with almost similar isotopic offsets ($\Delta^{66/64}Zn_{\text{solid-solution}}$=0.25-0.36‰) for goethite, lepidocrocite and feroxyhyte. This is consistent with the same average Zn-O bond lengths for adsorbed Zn on these solids as revealed by Zn K-edge X-ray absorption fine structure spectroscopy. In contrast, at an initial Zn/Fe molar ratio of 0.02, incorporation of Zn into goethite and lepidocrocite by substituting for lattice Fe preferentially sequesters light Zn isotopes with $\Delta^{66/64}Zn_{\text{substituted-stock solution}}$ of -1.52 ± 0.09‰ and -1.18 ± 0.15‰, while Zn-substituted feroxyhyte (0.06±0.11‰) indicates almost no isotope fractionation. This is closely related to the different crystal nucleation and growth rates during the Zn-doped FeOOH formation processes. These results provide direct experimental evidence of incorporation of isotopically light Zn into Fe (hydr)oxides, and improve our understanding of Zn isotope fractionation mechanisms during mineral-solution interface processes.

KEYWORDS: metal (hydr)oxides, metal isotope fractionation, interface reactions, isomorphous substitution, adsorption, X-ray absorption fine structure spectroscopy
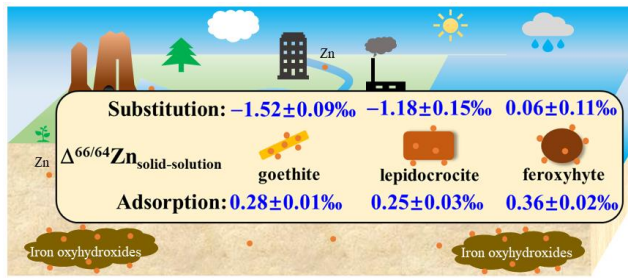
SYNOPSIS

Structures, crystal nucleation and growth rates of FeOOH polymorphs affect Zn isotope fractionation during adsorption and substitution.

TOC Art

Substitution: $-1.52\pm0.09‰$ $-1.18\pm0.15‰$ $0.06\pm0.11‰$

$\Delta^{66/64}\mathrm{Zn}_{solid-solution}$ goethite   lepidocrocite   feroxyhyte

Adsorption: $0.28\pm0.01‰$ $0.25\pm0.03‰$ $0.36\pm0.02‰$

Iron oxyhydroxides   Iron oxyhydroxides

1. INTRODUCTION

Though an essential micronutrient at low concentrations, zinc (Zn) is toxic at high concentrations.[1] In terrestrial environments, high concentrations of Zn mainly result from intensive anthropogenic activities, including those associated with mining, smelting, chemical industries, agriculture, scrap disposal, combustion of domestic wastes and processing of domestic waste waters as well as runoff from urban surfaces, but Zn also originates from natural sources, such as natural weathering and hydrothermal emissions.[2] As a result of these anthropogenic activities and natural processes, Zn-containing particles are released and subsequently weathered, and the release of dissolved Zn to soils as a result of weathering processes is probably the greatest source of Zn in the environment.[3] Once released and in order to better understand and predict Zn behavior, recent research is increasingly focused on the isotopic signatures associated with Zn biogeochemical cycling because these can fingerprint the different Zn sources, processes and pathways that release Zn and control Zn mobility and fates.[4-8] The isotopic features of different Zn sources may be blurred however, by the biogeochemical processes and pathways that control its behavior, creating a "black box" of isotope signals that are extremely difficult to disentangle.[12] In particular once Zn is released into the environment, Zn isotopes might be fractionated by a series of solid-solution interfacial reactions, such as adsorption, substitution and coprecipitation with minerals,[9-13] and mineral dissolution[4, 14-16]. It is thus critical to investigate Zn isotope fractionation during these interfacial processes in order to use Zn isotopic signals to trace and track Zn in contaminated and natural systems.

In surficial environments, iron (Fe) (hydr)oxides play an important role in mediating the geochemical behavior and fate of metals in soils, sediments and waters through adsorption and isomorphous substitution. While metal isotope fractionation during adsorption onto mineral surfaces is well-studied,[9, 10, 18-21] few studies have been conducted on fractionation during incorporation into Fe (hydr)oxide structures[17]. During adsorption processes, mineral crystalline structure, distribution of charge within the crystal lattices, presence of organic or mineral coatings and solution chemistry may govern metal binding mechanisms and thus metal isotope fractionation direction and

115 magnitude.[19, 22] Previous work has reported that heavy Zn isotopes are preferentially

116 adsorbed onto Fe (hydr)oxides with the fractionation magnitude for goethite much

117 smaller than that for ferrihydrite.[19] Other studies however, observe negative Zn isotope

118 fractionations induced by adsorption onto Fe (hydr)oxides at acidic conditions.[18, 23, 24]

119 During Zn adsorption onto quartz and amorphous silica, the fractionation magnitude

120 for the former is much smaller than that for the latter, probably due to the different

121 surface structural characteristics (e.g., structural disorder).[25] These results suggest that

122 there may be different Zn isotope fractionation during adsorption onto different FeOOH

123 polymorphs, which yet remains to be investigated.

124 Metal isotope fractionation induced by incorporation into the mineral structures

125 involves much complex mechanisms. This process is generally interpreted from a

126 kinetic fractionation effect, in which light isotopes are preferentially incorporated into

127 the mineral structure owing to faster diffusion rates than heavy isotopes.[26-28] Isotope

128 fractionation can also be caused by preferential attachment of one specific metal species

129 onto the host mineral primary nuclei growth sites after fast aqueous isotopic exchange.[29]

130 Crystal nucleation and growth mechanisms can also be particularly important. Isotope

131 fractionation of Cd during incorporation into goethite for example, is probably related

132 to the ferrihydrite dissolution-goethite recrystallization mechanism.[17] Last but not least,

133 host mineral crystallization rate may affect the fractionation magnitude.[26, 28, 30] Despite

134 the role that incorporation into Fe (hydr)oxides plays in controlling Zn mobility and

135 fate, the isotope fractionation of Zn during this process and the mechanism(s)

136 responsible for governing Zn isotope behavior are unknown.

137 Here we aim to determine Zn isotope fractionation mechanisms during

138 incorporation and adsorption with FeOOH polymorphs, including goethite (Goe),

139 lepidocrocite (Lep) and feroxyhyte (Fero). The different structures of these FeOOH

140 polymorphs may mean different crystal nucleation and growth mechanisms, metal

141 uptake amounts[31] and metal binding mechanisms[32]. It is therefore possible that

142 incorporation and adsorption of Zn with these FeOOH polymorphs may induce

143 different isotope fractionations. Zn K-edge X-ray absorption fine structure

144 spectroscopy (XAFS) and spherical aberration-corrected scanning transmission

electron microscopy are used to determine Zn binding mechanisms in substituted and adsorbed Fe (hydr)oxide minerals. The mineral crystal nucleation and growth mechanisms, crystallization rates and Zn binding mechanisms are then coupled to the Zn isotope fractionation during incorporation in and adsorption on these FeOOH polymorph minerals.

2. MATERIALS AND METHODS

All reagents were used as received without further purification and detailed information is presented in Text S1. The $ZnCl_2$ ($\geq$98.0%, Sinopharm Chemical Reagent Co., Ltd, China) used in Zn-substituted FeOOH polymorph synthesis had a Zn isotope composition ($\delta^{66/64}Zn$) of -0.21 ± 0.05‰, while that of the 1000 mg·L^{-1} $Zn(NO_3)_2$ bulk solution (Guobiao Testing & Certification Co., Ltd, China) used in Zn adsorption experiments was -2.40 ± 0.05‰.

2.1 Preparation of Zn Substituted FeOOH Polymorphs

Zinc-substituted FeOOH polymorphs with an initial Zn/Fe molar ratio of 0.02 were synthesized according to our previous study.[31] For Zn-substituted goethite (Goe), firstly 90 mL of 5 M NaOH solution was added to 50 mL of 1 M $FeCl_3 \cdot 6H_2O$ and 20 mM $ZnCl_2$ mixture in an acid-cleaned 1 L Teflon bottle. Subsequently, the obtained suspension was diluted to 1 L with ultrapure water (18.2 MΩ·cm) under stirring until the pH was adjusted to >13, and aged for 60 h at 70 ^{o}C. For Zn-substituted lepidocrocite (Lep), 16 g $FeCl_2 \cdot 4H_2O$, 22.4 g $(CH_2)_6N_4$, 5.6 g $NaNO_2$ and 0.22 g $ZnCl_2$ solids were added to 560 mL ultra-pure water in an acid-cleaned 1 L Teflon bottle (the pH of the suspension was ~6.2) and then put in water bath kettle under stirring at 60 ^{o}C for 3 h. For Zn-substituted feroxyhyte (Fero), 5 M NaOH solution was added to 300 mL of 0.1 M $FeCl_2 \cdot 4H_2O$ and 2 mM $ZnCl_2$ mixture to adjust the suspension pH to 8 under vigorously stirring in an acid-cleaned 1 L Teflon bottle, then 30% H_2O_2 was added to the solution. The obtained green suspension gradually transformed into a reddish brown precipitate and no further bubbles formed after ~1 h.

At the end of each synthesis, 50 mL of suspension was withdrawn under vigorously stirring, and the solid and solution were separated by centrifugation. The solution was kept for further analysis and labeled as "supernatant" in order to

differentiate it from that obtained during Zn adsorption experiments. The obtained solids (named as Zn2Goe, Zn2Lep and Zn2Fero, respectively) were treated with 50 mL 2 M HNO_3 solution for 0.5 h to remove adsorbed Zn species on the mineral surfaces (This part of Zn was labelled as "Adsorbed").[33] The remaining solids were named as Zn2Goe_n, Zn2Lep_n and Zn2Fero_n.

In order to monitor the crystal formation processes, independent synthetic experiments were conducted. 25 mL suspension was withdrawn at predetermined time intervals after heating of the initial reactant mixtures during the Zn-substituted Goe and Lep synthesis or upon the addition of H_2O_2 into the initial reactants during Zn-substituted Fero synthesis. The final pH values of the suspensions were 12.65 ± 0.05, 5.32 ± 0.05 and 2.22 ± 0.05, respectively. The suspensions were immediately centrifuged, and as-obtained solids were thoroughly washed and then freeze dried for further use.

2.2 Sample Characterization

Pure Goe, Lep and Fero samples were synthesized as described above without the addition of Zn. The purity of the obtained solid samples was confirmed by powder X-ray diffraction (XRD) analysis (Figure S1). Quantitative phase analysis or Rietveld structure refinement of the intermediate solids during the synthesis of Zn-doped FeOOH polymorphs was conducted using TOPAS software version 4.2 (Bruker-AXS).[17] The specific surface areas of Goe, Lep and Fero were determined to be 46, 168 and 116 $m^2 \cdot g^{-1}$ by multipoint BET modelling of the N_2 physical adsorption data, while the points of zero charge (PZCs) of these samples were measured to be ~9.7, ~8.5 and ~9.4 respectively by adopting a Zeta potential method[17] (Figure S2). The sample morphologies were probed by electron microscopy (JEM-2010 HT, Japan) (Figure S3). The atomic images of Zn2Goe_n and Zn2Lep_n were obtained on a JEM-NEOARM spherical aberration-corrected scanning transmission electron microscopy at 200 kV (JEOL, Japan). Energy dispersive X-ray spectroscope (EDS) quantitative analysis of single crystals for each mineral was also conducted.

The Fe and Zn concentrations in the solutions and solids were measured by flame atomic adsorption spectrometry (FAAS, Agilent Technologies 200 series AA) or inductively coupled plasma-optical emission spectrometry (ICP-OES, Agilent 5110).

205 The limit of detection for Zn by FAAS is 4.76 $\mu g \cdot L^{-1}$ with an uncertainty of 0.4%,

206 while for ICP-OES is 9 $\mu g \cdot L^{-1}$ with an uncertainty of 2.0%.

2.3 Adsorption Experiments

208 The minerals were hydrated in 0.05 M KNO_3 solution for 24 h before Zn addition.

209 For kinetic adsorption experiments, 91.7 or 152.9 μM Zn was reacted with 1 $g \cdot L^{-1}$ Goe,

210 or 0.5 $g \cdot L^{-1}$ Lep and Fero at pH 7±0.05 for 48 h, during which suspension aliquots were

211 withdrawn at predetermined intervals. Adsorption edge experiments were conducted

212 between pH 4.0-8.0. Adsorption isotherms were conducted with Zn initial

213 concentrations of 0-305.8 μM for Goe or 0-611.5 μM for Lep and Fero at pH 7±0.05.

214 The initial Zn concentrations were designed to obtain similar Zn coverages on these

215 solids and to prevent Zn precipitation. The suspension pH was maintained by the

216 addition of 1 M HNO_3 or KOH solution. All adsorption experiments were conducted in

217 Teflon tubes. A reaction time of 24 h was adopted by assuming that both adsorption

218 and isotope fractionation equilibrium are approached according to literature.[19, 34]

219 At the end of experiments, solids and supernatants were separated through 0.2 μm

220 cellulose membranes. To remove dissolved Zn the solids were immediately washed

221 with background electrolyte and sequentially ultrapure water, the pHs of which were

222 adjusted to be consistent with the adsorption experiments.[19] The washed solids were

223 collected with membranes, sealed with Kapton tape and then stored at 4 °C within 24 h

224 for further Zn K-edge XAFS analysis. The Zn concentrations in the supernatants and

225 solids after digestion were measured by FAAS. The obtained Zn-loaded solids were

226 labeled as ZnmGoe_pHn, ZnmLep_pHn and ZnmFero_pHn, in which m is the initial

227 Zn concentration in $mg \cdot L^{-1}$ and n is the reaction pH. Replicated experiments were

228 carried out 2-3 times to ensure reproducibility.

229 A Zn adsorbed ferrihydrite (Fh) standard, Zn20Fh_pH7.5, was also prepared by

230 reacting 0.31 mM Zn with 0.5 $g \cdot L^{-1}$ Fh (synthesized according to a previous study)[17]

231 for 24 h at pH 7±0.05 in 0.05 M KNO_3 solution.

2.4 Purification and Measurement of Zn Isotopes by MC-ICP-MS

233 Zinc-containing solids were digested using 12 M HCl and 15 M HNO_3 solutions

234 until dry and then redissolved in 2 M HCl solution. About 3 μg Zn was weighted for

235 Zn purification. After drying, 2 mL of 2 M HCl was added, and the solution was

236 transferred into a 15 mL polypropylene centrifuge tube. Samples were then purified on

237 columns containing 3 mL of pre-cleaned 100–200 mesh AG MP-1M (Bio-Rad, USA)

238 anion-exchange resin.[35, 36] After the adsorption of metals onto the column, 30 mL of

239 2 M HCl and 12 mL of 0.3 M HCl were passed through the columns, respectively. Zn

240 was eluted using 12 mL of 0.012 M HCl. The solution was evaporated to dryness at

241 110 °C and dissolved in 3 mL of 1% HNO_3. Additionally, 0.5 mL of the final solution

242 was used for Zn measurement to monitor the recovery, and the residue was used for Zn

243 isotope analysis. Satisfactory recoveries were obtained for the unprocessed and

244 processed samples (> 98%).

245   Zinc isotope ratios were measured using a Thermo Scientific Neptune plus MC-

246 ICP-MS at the State Key Laboratory of Crust–Mantle Evolution and Mineralization at

247 Nanjing University. Instrumental mass bias was corrected using a coupled method of

248 sample-standard bracketing (SSB) and Cu doping. More details are provided in Text

249 S2. Sample Zn isotope ratios were reported in standard delta notation in per mil units

250 relative to IRMM 3702 Zn solution according to Eq. 1:

$$\delta^{66/64}Zn=\left[\frac{(^{66}Zn/^{64}Zn)_{sample}}{(^{66}Zn/^{64}Zn)_{std}} - 1\right]\times1000 \quad (1)$$

251   Since a mass dependent fractionation law applies to all samples (Figure S4), only

252 $\delta^{66/64}Zn$ was reported. CAGS-1 and the new AA-ETH Zn isotope standard solutions

253 were used as internal laboratory secondary reference materials, and the $\delta^{66/64}Zn_{IRMM\,3702}$

254 values were −0.85±0.05‰ (n=6) and −0.01±0.05‰ (n=6) respectively, agreeing well

255 with previously reported values.[35, 37] An in-house sulphide standard (BCR-2) was used

256 to monitor potential Zn isotope fractionation during purification, and the analyses of

257 BCR-2 yielded $\delta^{66/64}Zn_{JMC-Lyon}$ of 0.28±0.02‰ (n=2), consistent with reported values.[38]

258 The long-term reproducibility of 0.05‰ was used for data measured with 2SD value of

259 < 0.05‰. The Zn isotope fractionation ($\Delta^{66/64}Zn_{solid-solution}$) between solid and aqueous

260 phases is defined as Eq. 2:

$$\Delta^{66/64}Zn_{solid-solution}=\delta^{66/64}Zn_{solid} - \delta^{66/64}Zn_{solution} \quad (2)$$

2.5 Zn K-edge XAFS Data Collection and Analysis

Zinc K-edge XAFS spectra for Zn-containing samples along with aqueous Zn(NO$_3$)$_2$ were collected on beamline 1W2B at BSRF at room temperature with a pair of Si(111) monochromators in fluorescence or transmission mode. The Zn metal foil (E$_0$ = 9659 eV) was used for energy calibration. The data processing and analysis were performed using the IFEFFIT software.[39] The parameters used for background removal were: E$_0$ = 9667 eV, k-weight = 2 and R$_{bkg}$ = 1.1. Structural parameters (R, CN, and σ^2) were obtained by fitting the experimental k^2-weighted data to the standard equation.[40] An amplitude reduction factor (S$_0^2$) of 0.85 was used.[20]

3. RESULTS AND DISCUSSION

3.1 Zn Isotope Fractionation during Adsorption on FeOOH Polymorphs

Similar Zn adsorption kinetics, pH adsorption edges and adsorption isotherms are observed for these FeOOH polymorphs (Figure S5 and Table S1). The maximum Zn adsorption densities on Goe, Lep and Fero obtained by Langmuir fitting are 2.47, 2.62 and 4.12 μmol·m^{-2}, respectively (Table S2).

Analysis of the isotopic compositions of the solutions and solids shows that heavy Zn isotopes are preferentially enriched on the mineral surfaces. As the reaction pH raises from 6 to 8, the proportion of Zn adsorbed (*f*) increases from 19.0 to 91.4% for Goe, from 18.1 to 92.7% for Lep and from 20.6 to 96.8% for Fero, while correspondingly, the $\delta^{66/64}$Zn$_{solution}$ value decreases from –2.42±0.05‰ to –2.62±0.05‰, from –2.50±0.05‰ to –2.77±0.05‰ and from –2.40±0.08‰ to –2.61±0.05‰ (Table S3). When the isotopic compositions are plotted as a function of *f* (Figure 1), it is clear that experimental $\delta^{66/64}$Zn values in the solutions and solids linearly decrease as *f* increases. The data are fitted with both an equilibrium (Eq. 3) and Rayleigh model (Eq. 4):

$$\delta^{66/64}Zn_{solution} = \frac{\delta^{66/64}Zn_{stock\ solution} - 1000 \cdot f(\alpha_{solid-solution} - 1)}{1 + f(\alpha_{solid-solution} - 1)} \quad (3)$$

$$\delta^{66/64}Zn_{solution} = (1000 + \delta^{66/64}Zn_{stock\ solution}) \cdot (1 - f)^{(\alpha_{solid-solution} - 1)} - 1000 \quad (4)$$

286 Where $\alpha_{\text{solid-solution}}$ represents the estimated isotope fractionation factor between

287 adsorbed and aqueous Zn, and $\delta^{66/64}\text{Zn}_{\text{stock solution}}$ denotes the estimated value for stock

288 solution. The equilibrium model applies to systems in which forward and backward

289 reactions occur at similar rates, while the Rayleigh model applies to homogeneous

290 reactant pools where light isotopes are continuously and preferentially removed.[41]

291   The fits generated using the equilibrium model agree much better with the

292 experimental data than those using the Rayleigh model. This demonstrates that Zn

293 isotope fractionation during adsorption onto the FeOOH polymorphs results from an

294 equilibrium fractionation mechanism. The derived fractionation factors ($\alpha_{\text{solid-solution}}$) are

295 1.00028± 0.00001, 1.00025±0.00003 and 1.00036±0.00002 for Goe, Lep and Fero,

296 respectively. The isotope fractionation between adsorbed and dissolved Zn can be

297 calculated according to the isotope fractionation factor (Eq. 5):

$$\Delta^{66/64}\text{Zn}_{\text{solid-solution}} \cong 1000 \times \ln\boldsymbol{\alpha}_{\text{solid-solution}} \qquad (5)$$

298   The $\Delta^{66/64}\text{Zn}_{\text{solid-solution}}$ are thus calculated to be 0.28±0.01‰, 0.25±0.03‰ and

299 0.36±0.02‰ for Goe, Lep and Fero, respectively. This suggests that Zn adsorption on

300 these three FeOOH polymorphs induces a similar isotope fractionation.
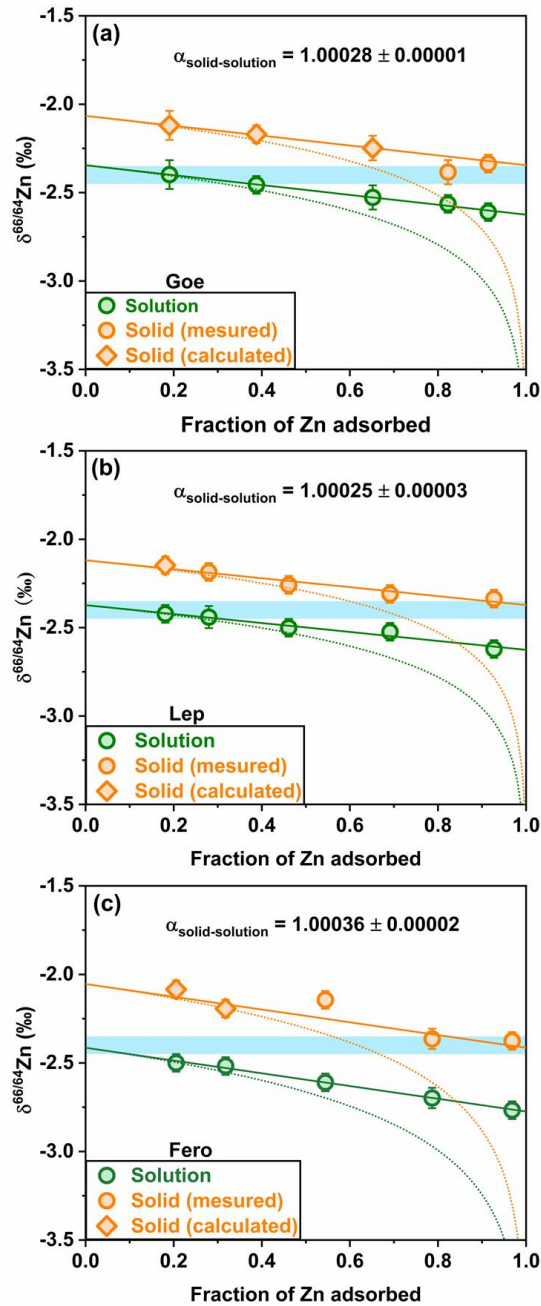
Figure 1. Zn isotopic compositions of solution and solid phases as a function of adsorbed Zn fraction (*f*) during adsorption onto Goe (a), Lep (b) and Fero (c). The lines and dashed curves represent the theoretical $\delta^{66/64}$Zn values calculated using the equilibrium model and the Rayleigh model, respectively. Several $\delta^{66/64}$Zn values of the solids (diamonds) were calculated according to the mass balance: $\delta^{66/64}$Zn$_{solid}$ = ($\delta^{66/64}$Zn$_{stock\ solution}$ − $\delta^{66/64}$Zn$_{solution}$×(1 − *f*))/*f*. The $\delta^{66/64}$Zn of the bulk Zn solution used for adsorption experiments is −2.40±0.05‰.

3.2 Zinc Binding Environments in the Zn-containing Samples

Both Zn K-edge X-ray absorption near edge structure (XANES) and extended X-ray absorption fine structure (EXAFS) spectra were used to analyze the Zn binding mechanisms in typical Zn-containing samples. XANES spectra of all samples display broadening and/or splitting of the main edge at ~9672 eV (Figure 2A), due to the contribution of second Fe neighbors.[42] A well-defined peak at ~9688 eV is observed for aqueous $Zn(NO_3)_2$ and Zn-substituted samples Zn2Goe_n and Zn2Lep_n, suggesting that Zn is predominantly in octahedral (^{VI}Zn) geometry in the substituted samples and predominantly tetrahedral (^{IV}Zn) complexes in the adsorbed samples.[24, 42-44] Linear combination fitting of the adsorbed samples using Zn20Fh_pH7.5 and Zn2Goe_n as endmembers, in which exclusively ^{IV}Zn or ^{VI}Zn exists respectively,[19, 44-46] demonstrates that Zn adsorbed Fero samples contain only ^{IV}Zn, Zn adsorbed Lep samples contain 17-27% ^{VI}Zn, while Zn adsorbed Goe samples contain 31-40% ^{VI}Zn (Table S4). Previous studies have demonstrated that Zn is adsorbed on goethite as octahedral[19] but also tetrahedral complexes.[23, 47] The varying proportions of ^{IV}Zn and ^{VI}Zn may be related to the mineral characteristics and solution chemistry.
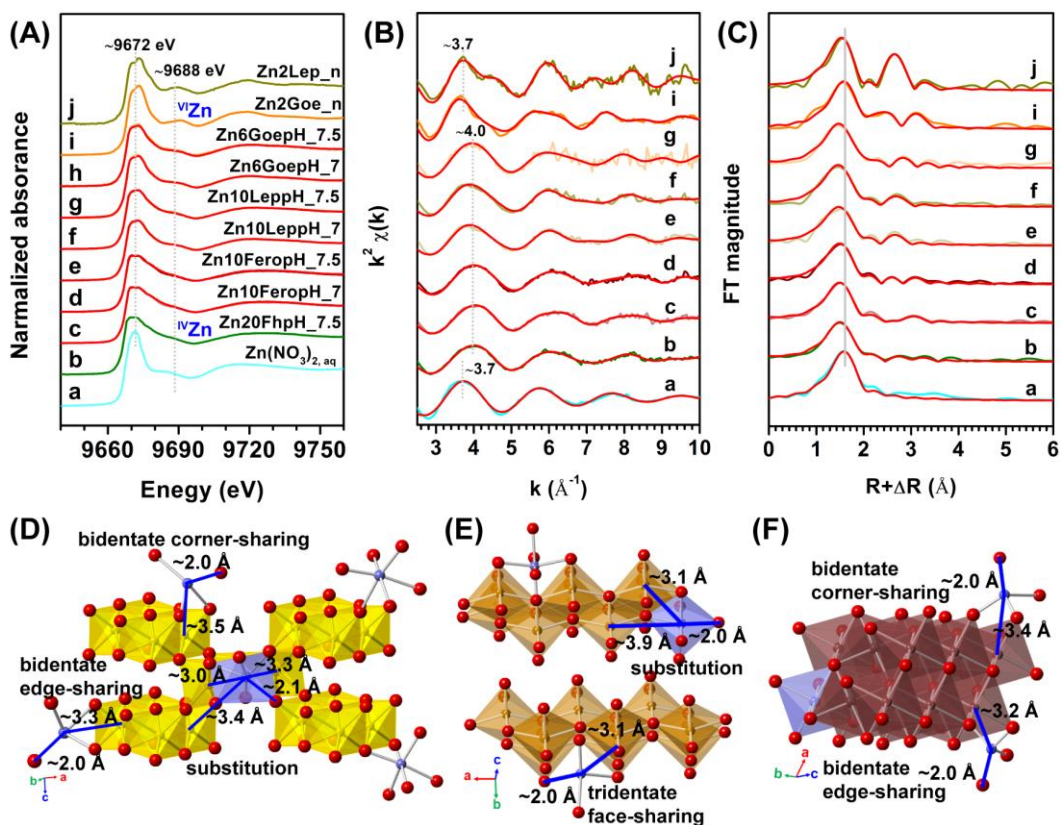
Figure 2. Zinc K-edge XANES (A), k^2-weighted EXAFS (B) and the corresponding

327 Fourier transformed spectra (FTs, C) of typical Zn adsorbed, substituted samples, and

328 aqueous $Zn(NO_3)_2$ standard ($Zn(NO_3)_{2, aq}$), overlaid with the best fits. The experimental

329 data are displayed as colored lines and the best fits are shown as red lines. The adsorbed

330 samples were named as ZnmGoe_pHn, ZnmLep_pHn and ZnmFero_pHn, in which m

331 is the initial Zn concentration in $mg \cdot L^{-1}$ and n is the reaction pH. The Zn substituted

332 samples were labeled as ZnmGoe_n and ZnmLep_n, in which m is the initial Zn/Fe

333 molar ratio. During linear combination fitting of the Zn K-edge XANES spectra for the

334 Zn adsorbed samples, Zn adsorbed ferrihydrite sample (Zn20Fh_pH7.5) and Zn

335 substituted goethite sample (Zn2Goe_n) were used as tetrahedral (^{IV}Zn) and octahedral

336 (^{VI}Zn) endmembers. A schema representing the local environments of Zn adsorbed onto

337 and incorporated into goethite (D), lepidocrocite (E) and feroxyhyte (F).

338   In the k^2-weighted EXAFS spectra (Fig. 2B), there is a shift in the first oscillation

339 of the Zn adsorbed samples to high k (~4.0 Å^{-1}) relative to that of the aqueous $Zn(NO_3)_2$

340 and Zn substituted samples (~3.7 Å^{-1}). This shift indicates a change of Zn-O first shell

341 coordination from octahedral to tetrahedral,[48] further confirming the XANES analysis.

342 In the Fourier transformed spectra (Fig. 2C), the obvious peaks beyond the first Zn-O

343 shell indicate the formation of Zn inner-sphere complexes on the mineral surfaces or

344 incorporation into the lattices. Shell-by-shell EXAFS fitting (Table S4) shows an

345 average Zn-O distance of 2.07±0.01 Å for aqueous $Zn(NO_3)_2$, and of 1.99±0.01 Å for

346 ^{IV}Zn in Zn adsorbed ferrihydrite, which agree well with previous literature.[9, 19, 47, 49]

347 Similar average Zn-O distances of 1.98-1.99 Å are observed for the Zn-adsorbed

348 samples. These distances are closer to those for tetrahedral Zn. This is consistent with

349 the linear combination fitting which shows that Zn adsorbed on Fero, Lep and Goe is

350 predominantly tetrahedral but with an increasing amount of octahedral Zn, respectively

351 (Table S4). Similar results were previously observed for Zn-O distances in hydrozincite

352 (~2.00 Å),[9] Zn adsorbed on manganite (1.98-2.04 Å)[50] and on todorokite at pH6-8

353 (2.00-2.05 Å).[9] Two Zn-Fe distances of 3.12-3.25 Å and 3.35-3.49 Å are also observed

354 for Zn adsorbed Fh, Fero and Goe. This indicates the formation of bidentate edge- and

355 corner-sharing complexes on the mineral surfaces (Fig. 2D,F).[43, 19, 49] In contrast, only

356 a Zn-Fe distance of 3.11-3.13 Å is detected for Zn adsorbed Lep, which can be assigned

357 to tridentate face-sharing complexes (Fig. 2E).[19] For Zn-substituted Goe, an average

358 Zn-O distance of 2.07 ± 0.01 Å and three Zn-Fe distances at 3.02 ± 0.03 Å, 3.33 ± 0.06 Å

359 and 3.49 ± 0.05 Å are observed. These distances correspond well to those observed for

360 Zn substituted goethite.[51] For Zn-doped Lep, a Zn-O distance of 2.01 ± 0.01 Å and two

361 Zn-Fe distances of 3.11 ± 0.02 Å and 3.90 ± 0.08 Å support the incorporation of Zn into

362 the mineral lattices according to the mineral crystal structure.

3.3 Isotope Fractionation Mechanisms during Zn Adsorption on FeOOH Polymorphs

365 Our results clearly suggest that heavy Zn isotopes are preferentially partitioned

366 onto the FeOOH mineral surfaces, which is in good agreement with previous studies.[10,

367 19, 20, 25] It was previously reported that heavy Zn isotopes are enriched on Fe, Si, Al and

368 Mn (hydr)oxides surfaces during adsorption processes, owing to the formation of inner-

369 sphere complexes with shorter Zn-O lengths compared to aqueous Zn.[19, 20, 25, 52, 53]

370 Generally, heavier metal isotopes are preferentially concentrated in stiffer bonding

371 environments, e.g. coordinated to highly covalent bonds, with lower coordination

372 number and shorter bond lengths.[12, 41] The Zn K-edge XAFS analysis shows that Zn

373 adsorbed onto Goe, Lep and Fero form inner-sphere complexes with Zn-O bond lengths

374 ranging from 1.98-1.99 Å that are much shorter than that of octahedral Zn in solution

375 $(2.07\pm0.01$ Å) (Table S4). These bond length differences are probably responsible for

376 adsorption-induced enrichment of heavy Zn isotopes onto these FeOOH polymorphs.

377 The fact that these Zn-O bond lengths are almost the same for the different FeOOH

378 polymorphs (Table S4), also explains why we observe a similar $\Delta^{66/64}Zn_{solid-solution}$ of

379 0.25-0.36‰ for Goe, Lep and Fero. Though Zn K-edge XANES linear combination

380 fitting shows that Zn is predominantly adsorbed as ^{IV}Zn on Fero, Lep and Goe but with

381 increasing proportions of ^{VI}Zn in the latter two, EXAFS analysis, which fits the average

382 bonding environment, gives essentially the same Zn-O distances. As such our results

383 indicate that having only ^{IV}Zn (feroxyhyte) or a mixture of ^{IV}Zn-^{VI}Zn

384 (goethite/lepidocrocite) does not significantly modify the recorded $\Delta^{66/64}Zn_{solid-solution}$

385 (difference < 0.1‰), which may be ascribed to the weaker effect of ^{VI}Zn than ^{IV}Zn on

386 isotope fractionation magnitude during adsorption on Fe (hydr)oxides.[44]

It is noteworthy that although our study confirms the enrichment of heavy Zn isotopes during adsorption on Fe (hydr)oxides as revealed by some previous studies,[19] others observed negative $\Delta^{66/64}Zn_{\text{solid-solution}}$ during Zn adsorption on goethite at acid conditions (e.g., pH5.3-6.1)[18, 23, 24]. In contrast, positive $\Delta^{66/64}Zn_{\text{solid-solution}}$ is observed at higher pHs (6-8) here and in previous study[19]. We therefore hypothesize that pH and Zn solution speciation may play important roles in Zn isotope fractionation direction and magnitude.[18, 23] For example, relative to $Zn(H_2O)_6^{2+}$, $Zn(OH)(H_2O)_5^+$ and $Zn(OH)_2(H_2O)_4$, the proportions of which are increased at higher pH, enrich relatively heavy isotopes.[54] Speciation calculations in the current study however, show that aqueous $Zn(H_2O)_6^{2+}$ is the dominant species (~95%) with only limited $ZnNO_3(H_2O)_5^+$ (5%) over pH 4-8 (Figure S6). This suggests that Zn aqueous speciation contributes little to the Zn isotope fractionation during adsorption onto the FeOOH polymorphs in the present study. Thus it is possible that different isotope fractionation mechanisms dominate at low and high pH conditions during Zn adsorption onto goethite but confirmation of this requires further study.

**3.4 Crystal Formation Processes of Zn-Substituted FeOOH Polymorphs**

In order to better understand the Zn isotope fractionation during substitution into the FeOOH polymorph minerals, we first discuss the crystal growth processes. Powder XRD analysis of intermediate solids during Zn-substituted Goe synthesis shows the formation of goethite (ICSD 71870) via a ferrihydrite precursor (ICSD 158475) (Figure 3a,S7a).[17, 27, 55] Specifically, the diffraction peaks of goethite appear within 3 h, and then gradually increase with time and stay almost unchanged after 12 h. The ferrihydrite formed at 1 h has a coherent scattering domain (CSD) size of 2.6±0.4 nm (Figure 3a, Table S5). Quantitative phase analysis of the 3 h and 6 h solids shows 61±10% and 13±5% ferrihydrite, and the precursor is almost completely transformed to goethite at 12 h. The average CSD sizes of the goethite particles slightly increase from 27.0±0.2 nm at 12 h to 28.5±0.2 nm at 60 h. Owing to the high reaction pH (~12.7) and the reactivity of ferrihydrite/goethite towards metals, almost all the initial Zn is retained on the intermediate solids with constant Zn/Fe molar ratios of 1.8-1.9%.

Unlike the Zn-doped Goe system, there is probably no ferrihydrite precursor

during the Zn-doped Lep synthesis under the current experimental conditions (pH~5.3). XRD patterns show the appearance of weak crystalline lepidocrocite (ICSD 108876) after reaction for 1 min but without ferrihydrite (Figure 3b,S7b). This is consistent with previous studies showing that lepidocrocite dominates upon Fe^{2+} oxidation at pH>5 while ferrihydrite dominates at pH < 5.[56] As the reaction progresses, a gradual increase in the XRD peak intensity indicates increasing mineral crystallinity within 1 h. The particle CSD size increases from 3.3±0.1 nm at 1 min to 7.6±0.1 nm at 1 h, and then remains constant (Table S5). This may indicate that Lep is formed through a direct nucleation and crystal growth mechanism. Concurrently, owing to the finite particle size, at 1 min the solid has a Zn/Fe ratio of 1.9±0.0%. As the crystal grows and particle size increases, some of the Zn may be released back into solution and at 30 min the solid Zn/Fe ratio decreases to 1.3%. The solids from 1-3 h have Zn/Fe ratios of 1.5-1.6%, probably owing to Zn re-adsorption onto the solid.

Similar to that of Zn-doped Lep formation, Fero crystal (ICSD 291515) is directly formed by Fe^{2+} oxidation without ferrihydrite formation (Figure 3c,S7c).[56] The Fero crystal nucleation and growth however, occur at much higher rates than those of Lep, e.g., for Fero crystal nucleation and growth are almost complete in the first minute, as evidenced by the almost unchanged XRD peak intensities and CSD sizes of the intermediate solids (9.8-11.5 nm) as the reaction goes on (Table S5). The Zn/Fe molar ratios in these intermediates slightly decrease from 1.0±0.0% in the 1 min solid to 0.8±0.0% in the 60 min solid, probably owing to the competition by protons for active sites at a pH of ~2.2.
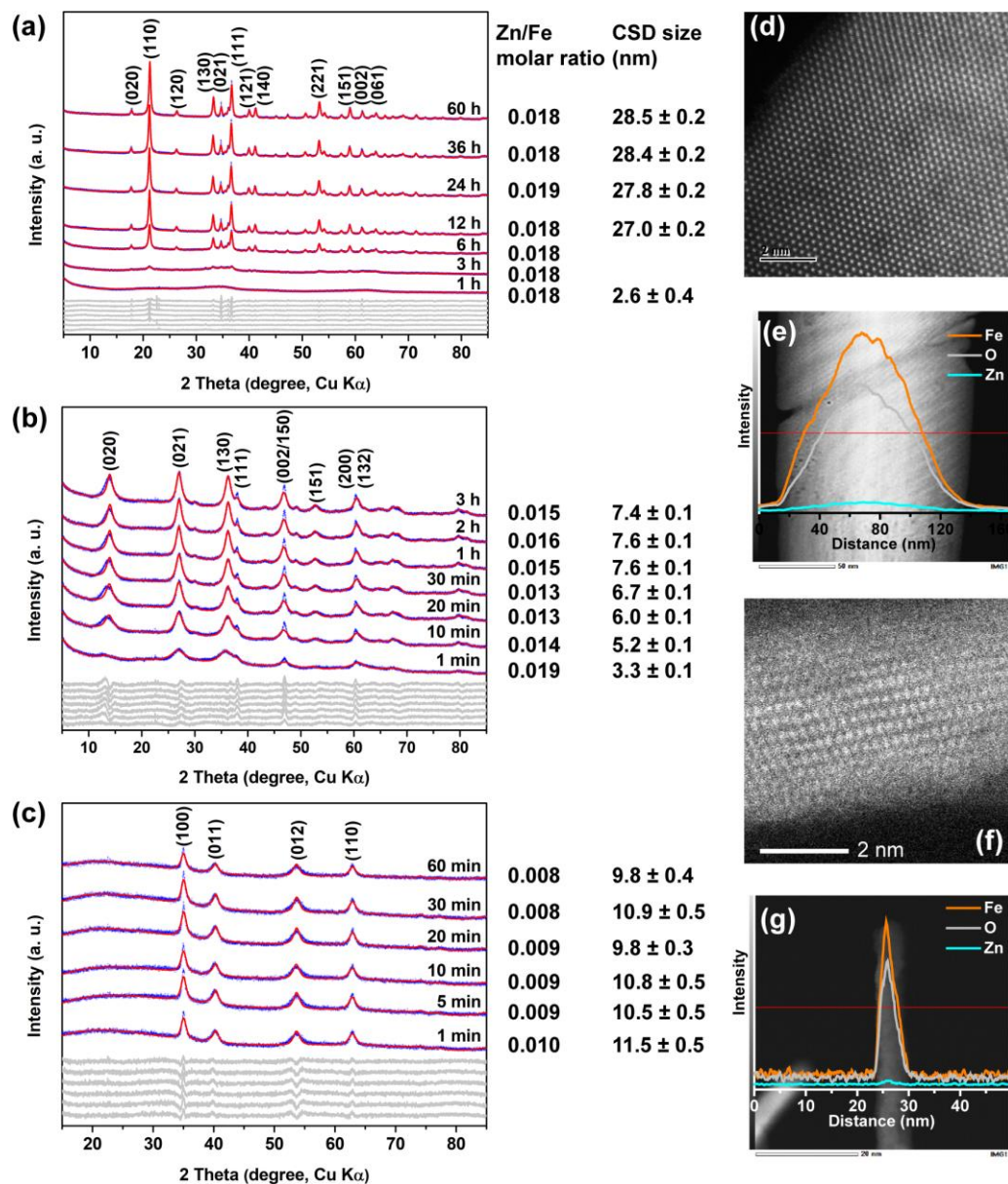
Figure 3. Powder XRD patterns, final Zn/Fe molar ratios and coherent scattering domain (CSD) sizes of the reaction intermediates during the synthesis of Zn-substituted goethite (a), lepidocrocite (b) and feroxyhyte (c). The XRD experimental patterns (blue crosses) are overlaid with the best Rietveld structure refinement or quantitative phases analysis results (red lines), based on the structure models of goethite (ICSD 71810), ferrihydrite (ICSD 158475), lepidocrocite (ICSD 108876) and feroxyhyte (ICSD 291515). The difference patterns are shown as gray lines at the bottom. All experiments were conducted at an initial Zn/Fe molar ratio of 0.02. Atomic resolution images and EDS line scans of Zn-substituted goethite sample, Zn2Goe_n (d and f), and Zn-substituted lepidocrocite sample, Zn2Lep_n (e and g).

450   The final molar ratios of Zn/Fe in Zn2Goe_n, Zn2Lep_n and Zn2Fero_n are

451   1.9±0.0%, 1.3±0.0% and 0.4±0.0%, respectively. Zn cations removed by HNO_3

452   treatment of Zn2Goe, Zn2Lep and Zn2Fero samples are probably those adsorbed on

453   these mineral surfaces. This can be confirmed by several lines of evidence. Firstly, only

454   2.0%, 7.2% and 7.3% of the total Fe are removed from these samples (Figure S8).

455   Secondly, powder XRD patterns of these samples before and after HNO_3 treatment are

456   almost the same (Figure S9). Thus, the solids obtained after HNO_3 treatment are Zn-

457   substituted minerals. Atomic images of Zn2Goe_n and Zn2Lep_n show the uniform

458   distribution of Zn in the crystal lattices (Figure 3d,f). This is further confirmed by the

459   EDS line scan of single crystals for each mineral that clearly shows the strong

460   correlations between Fe, Zn and O (Figure 3e,g). Furthermore, the EDS analysis gives

461   an average Zn/Fe molar ratio of 2.1±0.1% for Zn2Goe_n and 1.0±0.1% for Zn2Lep_n

462   (Figure S10), which agree well with the wet chemical analysis results.

463   **3.5 Zn Isotope Fractionation during Substitution for Fe in FeOOH Polymorphs**

464   At the end of Zn2Goe synthesis, Zn isotope compositions in the supernatant

465   ($\delta^{66/64}Zn_{supernatant}$) and in the solid ($\delta^{66/64}Zn_{Zn2Goe}$) are 1.27±0.05‰ and −0.15±0.07‰,

466   respectively (Figure 4,S8). The latter is identical to the Zn stock solution (-

467   0.21±0.05‰), due to the fact that almost all Zn is retained in Zn2Goe. After 2 M HNO_3

468   treatment, $\delta^{66/64}Zn$ for Zn incorporated into goethite ($\delta^{66/64}Zn_{Zn2Goe\_n}$) is −0.16±0.05‰,

469   while $\delta^{66/64}Zn$ for Zn in HNO_3 solution ($\delta^{66/64}Zn_{Adsorbed}$) is −0.07±0.05‰ (Figure 4,S8).

470   The $\delta^{66/64}Zn$ of Zn2Lep and the corresponding supernatant are −0.28±0.08‰ and

471   0.40±0.05‰, while $\delta^{66/64}Zn_{Zn2Lep\_n}$ and $\delta^{66/64}Zn_{Adsorbed}$ decrease to −0.45±0.05‰ and

472   0.12±0.08‰, respectively. In contrast, $\delta^{66/64}Zn$ of Zn2Fero and the corresponding

473   supernatant are −0.09±0.05‰ and −0.25±0.05‰ respectively, while $\delta^{66/64}Zn_{Zn2Fero\_n}$

474   decreases to −0.24±0.05‰ and $\delta^{66/64}Zn_{Adsorbed}$ increases to 0.18±0.05‰.
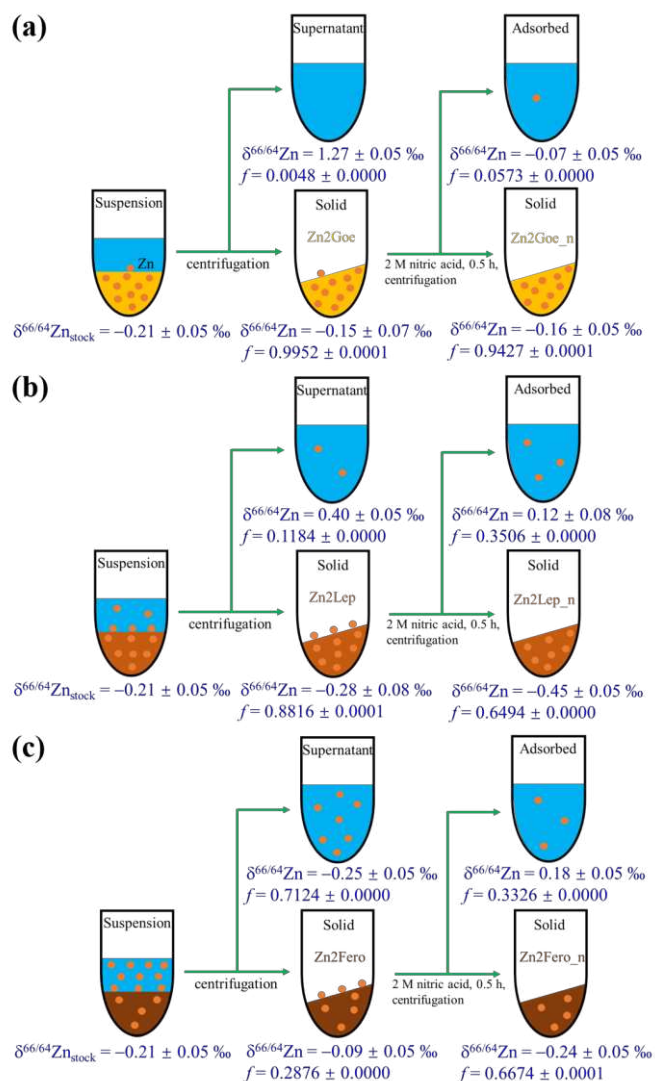
Figure 4. Schematic illustration of the experimental protocols for Zn isotope fractionation analysis during Zn substitution in (a) Goe, (b) Lep and (c) Fero. The Zn isotope composition and the fraction of Zn (f) at each step were presented. The $\delta^{66/64}Zn$ of ZnCl$_2$ used for Zn-substituted FeOOH polymorphs synthesis is -0.21 ± 0.05‰. It should be noted that the different f cannot be summed to obtain 100% and each step considers a new total population (100%), which is divided into two parts.

According to the isotope signals of different Zn pools during these processes, we can also calculate the Zn isotope compositions ($\delta^{66/64}Zn_{substituted}$) of the Zn-substituted FeOOH polymorph minerals (Eq. 6).

$$\delta^{66/64}Zn_{\text{stock solution}}$$

$$= f_1 \times \delta^{66/64}Zn_{\text{supernatant}} + f_2 \times \delta^{66/64}Zn_{\text{adsorbed}} \tag{6}$$

$$+ f_3 \times \delta^{66/64}Zn_{\text{substituted}}$$

where f_1, f_2 and f_3 refer to the fractions of Zn in supernatant, adsorbed on the mineral surfaces, and incorporated into the crystal lattices, respectively. Based on Eq. 6, the $\delta^{66/64}Zn_{\text{substituted}}$ values are calculated to be −0.22±0.05‰, −0.46±0.09‰ and −0.14±0. 10‰ for Zn2Goe_n, Zn2Lep_n and Zn2Fero_n respectively, which are consistent with the measured values. In these systems, adsorption and substitution processes both lead to the distribution of Zn between solution and solid and thus isotope fractionation.

If we assume that the two isotope fractionation factors involved here, for adsorption and incorporation, follow an equilibrium regime, the Zn isotope fractionation during incorporation into the lattices of these minerals ($\Delta^{66/64}Zn_{\text{substituted-stock solution}}$) can be calculated according to the isotope mass balance (Eq. 7):

$$\Delta^{66/64}Zn_{\text{solid-supernatant}} = f_2/(f_2 + f_3) \times \Delta^{66/64}Zn_{\text{adsorbed-supernatant}} \tag{7}$$

$$+ f_3/(f_2 + f_3) \times \Delta^{66/64}Zn_{\text{substituted-stock solution}}$$

where $\Delta^{66}Zn_{\text{solid-supernatant}}$ refers to Zn isotope fractionation between supernatant and solid at the end of the FeOOH formation, and equals -1.42±0.09‰, -0.68±0.09‰ and 0.16±0.07‰ for Zn2Geo, Zn2Lep and Zn2Fero, respectively. Thus, the $\Delta^{66/64}Zn_{\text{substituted-stock solution}}$ values are calculated to be −1.52±0.09‰, −1.18±0.15‰ and 0.06±0.11‰ for Zn2Goe_n, Zn2Lep_n and Zn2Fero_n respectively. These results clearly indicate that Zn substitution enriches light isotopes in goethite and lepidocrocite but almost no isotope fractionation occurs in feroxyhyte at the experimental conditions.

3.6 Zn Isotope Fractionation Mechanisms during Substitution for Fe in FeOOH Polymorphs

About >94% Zn is retained in the Zn2Goe_n solid by substituting for lattice Fe, and this results in a substantial negative isotope fractionation (−1.52±0.09‰) (Figures 4a,S8A). EXAFS analysis shows that Zn substitutes for lattice Fe in Zn2Geo_n with a Zn-O bond length the same as that (2.07±0.01 Å) for aqueous Zn. This suggests that

508   the enrichment of light Zn isotopes in Zn2Geo_n is not related to the Zn coordination

509   environment but rather is probably related to the goethite formation process. Goethite

510   generally forms from Fh dissolution-Goe recrystallization processes.[27, 55] In tandem

511   with the Fh formation, lighter Zn isotopes are preferentially incorporated in and/or

512   adsorbed onto the solid due to a kinetic effect, in which lighter isotopes diffuse faster

513   than heavier ones (e.g., within 1 h, Figure 3a). This results in the enrichment of heavy

514   Zn isotopes in the solution with $\delta^{66/64}Zn_{supernatant} > 1.27\pm0.05‰$. Subsequently, Fh

515   particles slowly dissolve, releasing into the solution soluble Fe as nucleus to form the

516   Goe and Zn species as suitable growth units (e.g., 1-12 h, Figure 3a). This process may

517   also be kinetically controlled, such that isotopically light Zn is released into the

518   surrounding solution. It was previous reported that in the early stages of proton- and

519   oxalate-promoted dissolution of finely powdered biotite granite, isotopically light Zn

520   was released following a kinetic isotope fractionation.[16] Further, the Zn released is

521   expected to be relatively isotopically light, because heavy Zn isotopes are preferentially

522   adsorbed on/incorporated in the ferrihydrite by adopting a tetrahedral geometry.[49]

523   Notwithstanding this, the released relatively light Zn isotopes are probably in

524   tetrahedral coordination.[19, 42, 49] After their release, a transformation from tetrahedral to

525   octahedral would occur and then the octahedral Zn is directly attached onto the goethite

526   growth unit. Alternatively, the released Zn tetrahedra can be immediately attached onto

527   the goethite growth unit and then transform to octahedra. Although all these kinetically

528   controlled isotope fractionation processes that are possibly involved during Zn

529   substitution in goethite can contribute to the enrichment of light Zn isotopes in the

530   goethite lattices, kinetic effects often occur in the first few hours of metal-mineral

531   interactions.[22] After this initial window, Zn adsorbed on the mineral surfaces and that

532   remaining in the supernatant exchange and reach adsorption/desorption equilibrium and

533   isotope fractionation equilibrium. Thus the $1.27\pm0.05‰$ Zn isotope signal of the

534   supernatant at the end of Zn-doped goethite synthesis is possibly the result of a complex

535   series of kinetic and equilibrium processes. Further, the calculated large

536    $\Delta^{66/64}Zn_{substituted\text{-}stock\ solution}$ ($-1.52\pm0.09‰$) suggests that the isotope signal of Zn-

537   substituted goethite is predominantly inherited from the Zn associated with the

ferrihydrite precursor. The exact mechanisms of Zn isotope fractionation during incorporation into goethite however, requires further investigations at varying pH and Zn/Fe molar ratios.

A Zn isotope fractionation of −1.18±0.15‰ is derived for Zn incorporation into Lep structure and is probably also caused by a complex suite of mineral nucleation and growth processes (Figure 3b). At the initial stage (< 1 min), finite Lep particles directly nucleate from solution,[56] and almost all the Zn is retained on the solids (Figure 3b). As the Lep crystals grow larger with time (1-60 min), part of the Zn is incorporated into the lattices, while part of the Zn is released back into solution, especially in the first 9 mins. The Zn released is probably that previously adsorbed on the solid surfaces during 0-1 min and is probably relatively heavy according to the adsorption experiments. This decreases the solid $\delta^{66/64}$Zn. After crystal growth is complete at ~1 h, Zn adsorption-desorption reactions dominate and probably approach equilibrium, and as-induced isotope exchanges occur simultaneously. During this stage, relatively heavy isotopes are adsorbed onto the Zn2Lep_n surfaces. This may consequently counteract the negative isotope fractionation during Zn incorporation into the lattices, resulting in the $\delta^{66/64}$Zn of Zn2Lep that is indistinguishable from that of the stock solution (Figure 4b, S8B).

The $\delta^{66/64}$Zn value for Zn2Fero is higher than the stock solution (Figure 4c, S8C), probably owing to the adsorption of heavy Zn isotopes onto the solids. Removal of the adsorbed Zn on the mineral surfaces by acid washing decreases the $\delta^{66/64}$Zn value of Zn2Fero_n so that it is indistinguishable from the stock solution. The negligible Zn isotope fractionation (0.06±0.11‰) in Zn2Fero_n is probably closely related to the fast crystal nucleation and growth that is almost complete in 1 min (Figure 3c). This is consistent with previous studies showing that the expression of kinetic isotope effects should be prevented in the solids when produced at extremely rapid precipitation rates.[28] Later (1-60 min), relatively heavy Zn isotopes in the residual solution pool are specifically adsorbed onto the solid surfaces, while at the same time, protons strongly compete for adsorption sites and drive the release of relatively light Zn isotopes into solution.[16]

568　　　　The different $\Delta^{66/64}Zn_{\text{substituted-stock solution}}$ induced by Zn substitution for lattice Fe in

569　these FeOOH polymorphs may be strongly related to their different nucleation and

570　growth rates, in addition to their different formation pathways. Several previous studies

571　have examined the effects of precipitation rates on Fe isotope fractionation. In general,

572　typically small or almost no Fe isotope fractionation is observed when precipitation

573　occurs either very fast or very slowly (e.g., hundreds of days).[26, 30] Significant isotope

574　fractionations occur however, when the precipitation progresses over time scales of

575　hours to weeks.[26, 28] In the present study, the time scales for Zn2Goe_n and Zn2Lep_n

576　nucleation and growth range from 1 h to 12 h, while that for Zn2Fero_n is within 1 min

577　(Figure 3). The calculated $\Delta^{66/64}Zn_{\text{substituted-stock solution}}$ values for these Zn substituted

578　FeOOH polymorphs are also comparable to those for Fe(III) precipitated at similar

579　precipitation rates.[26, 30] As the initial crystal nucleation and growth rates for Zn2Goe_n

580　and Zn2Lep_n are relatively low, diffusion gradient-controlled Zn incorporation into

581　the growing solids results in the large negative isotope fractionations observed. But for

582　Zn2Fero_n, diffusion gradients in the liquid boundary layer around the primary

583　feroxyhyte nuclei may limit isotope exchange between the residual Zn in this layer and

584　the bulk aqueous pool and finally lead to no isotope fractionation upon the rapid Zn

585　retention by the solids.[26]

586　**4. Environmental Implications**

587　　　　Various mineral-solution interfacial reactions, including adsorption and

588　substitution, contribute to a "black box" of isotope signals in both experimental and

589　environmental systems. In particular, adsorption and substitution with iron (hydr)oxides

590　play important roles in mediating element geochemical cycling, mobility and fate,

591　including the isotope fractionation factors of Zn in terrestrial environments. Here

592　preferential adsorption of heavy Zn isotopes onto FeOOH polymorph surfaces is

593　confirmed at circumneutral pH conditions. Moreover, the present study shows for the

594　first time that isotopically light Zn is incorporated into Fe (hydr)oxide structures. It is

595　recently reported that secondary minerals (e.g., Fe and Mn (hydr)oxides) during

596　pedogenesis enrich light Zn and Cu isotopes as internalized species, due to the

597　incorporation of these metals into Fe (hydr)oxides.[4, 13] Similarly, the enrichment of light

598 Zn isotopes in acid mine drainage precipitates, jarosite and goethite, relative to the

599 drainage is also attributed to Zn incorporation into the solids.[8] Here we provide the first

600 experimental evidence for these field observations. We observe larger fractionation

601 factors ($\Delta^{66/64}Zn_{substituted-stock\ solution}$= -1.5‰ to -1.2‰) however, for Zn substituted

602 goethite and lepidocrocite than the theoretical maximum of -0.7‰ for Zn incorporation

603 into Fe (hydr)oxides based on available global soil data[4] and the range of −0.35‰ to

604 −0.08‰ for Zn-substituted goethite[8]. This is probably caused by the different crystal

605 nucleation and growth rates of these Zn-substituted Fe (hydr)oxides under experimental

606 and field conditions[28] and highlights the important role that nucleation and growth

607 processes play in Zn isotope fractionation with Fe (hydr)oxides. Further experimental

608 studies conducted over a range of environmentally relevant conditions (e.g.,

609 circumneutral pH, micromolar concentrations of Zn and Fe, and low temperature) and

610 with other minerals are necessary to constrain the theoretical range of Zn isotope

611 fractionation during interactions with Fe (hydr)oxides, and further determine the most

612 important processes responsible for Zn isotope fractionation in experimental and

613 natural systems. As an additional implication of this work, the different Zn isotope

614 fractionation directions observed between adsorption onto and incorporation into Fe

615 (hydr)oxides might be used to identify the crystal chemistry of Zn in these minerals

616 according to their Zn isotope compositions. Conclusively, these results reveal possible

617 Zn isotope fractionation mechanisms during mineral-solution geochemical processes

618 pertinent to Earth's critical zone and provide a mechanistic framework towards source

619 tracing and process tracking Zn in contaminated and natural environments.

620 **ASSOCIATED CONTENT**

621 **Supporting Information**

622 Reagent information; details of Zn isotope ratio measurement; powder XRD, TEM,

623 EDS and zeta potential analyses of obtained minerals; Zn macroscopic adsorption

624 kinetics, adsorption edge, isotherms, and Zn isotope compositions in solid and aqueous

625 phases and mass balance; isotope and chemical compositions of solutions and solids

obtained during the Zn substitution experiments; Rietveld structure refinement results of intermediates during Zn-doped FeOOHs formation; linear combination fitting analysis of Zn K-edge XANES spectra and structure parameters derived from Zn K-edge EXAFS fitting; Zn species calculation.

Notes

The authors declare no competing financial interest.

REFERENCES

1. Jennings, A. A., Analysis of worldwide regulatory guidance values for the most commonly regulated elemental surface soil contamination. *J. Environ. Manage.* **2013,** *118*, 72-95.
2. Rout, G. R.; Das, P., Effect of metal toxicity on plant growth and metabolism: I. Zinc. In *Sustainable Agriculture*, Lichtfouse, E.; Navarrete, M.; Debaeke, P.; Véronique, S.; Alberola, C., Eds. Springer Netherlands: Dordrecht, 2009; pp 873-884.
3. Agency for Toxic Substances and Disease Registry (ATSDR). 2005. Toxicological profile for zinc. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service.
4. Little, S. H.; Munson, S.; Prytulak, J.; Coles, B. J.; Hammond, S. J.; Widdowson, M., Cu and Zn isotope fractionation during extreme chemical weathering. *Geochim. Cosmochim. Acta* **2019,** *263*, 85-107.
5. Liu, S. A.; Liu, P. P.; Lv, Y. W.; Wang, Z. Z.; Dai, J. G., Cu and Zn isotope fractionation during oceanic alteration: Implications for oceanic Cu and Zn cycles. *Geochim. Cosmochim. Acta* **2019,** *257*, 191-205.
6. Zhang, T.; Sun, R. Y.; Liu, Y.; Chen, L.; Zheng, W.; Liu, C. Q.; Chen, J. B., Copper and Zinc isotope signatures in scleratinian corals: Implications for Cu and Zn cycling in modern and ancient ocean. *Geochim. Cosmochim. Acta* **2022,** *317*, 395-408.
7. Ma, L.; Wang, W.; Xie, M. W.; Wang, W. X.; Evans, R. D., Using Zn isotopic signatures for source identification in a contaminated estuary of Southern China.

Environ. Sci. Technol. **2020,** *54,* (8), 5140-5149.

8.   Liu, Y. H.; Gao, T.; Xia, Y. F.; Wang, Z. R.; Liu, C. S.; Li, S. H.; Wu, Q. Q.; Qi, M.; Lv, Y. W., Using Zn isotopes to trace Zn sources and migration pathways in paddy soils around mining area. *Environ. Pollut.* **2020,** *267.*

9.   Wang, Z.; Kwon, K. D.; Peacock, C.; Mo, X.; Gou, W.; Feng, X.; Li, W., Zn stable isotope fractionation during adsorption onto todorokite: A molecular perspective from X-ray absorption spectroscopy and density functional theory. *Geochim. Cosmochim. Acta* **2022,** *327,* 116-136.

10. Guinoiseau, D.; Gelabert, A.; Moureau, J.; Louvat, P.; Benedetti, M. F., Zn isotope fractionation during sorption onto kaolinite. *Environ. Sci. Technol.* **2016,** *50,* (4), 1844-52.

11. Mavromatis, V.; González, A. G.; Dietzel, M.; Schott, J., Zinc isotope fractionation during the inorganic precipitation of calcite – Towards a new pH proxy. *Geochim. Cosmochim. Acta* **2019,** *244,* 99-112.

12. Komárek, M.; Ratié, G.; Vaňková, Z.; Šípková, A.; Chrastný, V., Metal isotope complexation with environmentally relevant surfaces: Opening the isotope fractionation black box. *Crit. Rev. Environ. Sci. Technol.* **2022,** *52,* (20), 3573-3603.

13. Quantin, C.; Guinoiseau, D., The use of stable isotopes in soil science: Metals. In *Reference Module in Earth Systems and Environmental Sciences*, Elsevier: 2022; pp 1-8.

14. Fernandez, A.; Borrok, D. M., Fractionation of Cu, Fe, and Zn isotopes during the oxidative weathering of sulfide-rich rocks. *Chem. Geol.* **2009,** *264,* (1-4), 1-12.

15. Opfergelt, S.; Cornelis, J. T.; Houben, D.; Givron, C.; Burton, K. W.; Mattielli, N., The influence of weathering and soil organic matter on Zn isotopes in soils. *Chem. Geol.* **2017,** *466,* 140-148.

16. Weiss, D. J.; Boye, K.; Caldelas, C.; Fendorf, S., Zinc isotope fractionation during early dissolution of biotite granite. *Soil Sci. Soc. Am. J* **2014,** (78), 171–179.

17. Yan, X. R.; Zhu, M. Q.; Li, W.; Peacock, C. L.; Ma, J. Y.; Wen, H. J.; Liu, F.; Zhou, Z. B.; Zhu, C. W.; Yin, H., Cadmium isotope fractionation during adsorption and substitution with iron (oxyhydr)oxides. *Environ. Sci. Technol.* **2021,** *55,* (17), 11601-11611.

18. Pokrovsky, O. S.; Viers, J.; Freydier, R., Zinc stable isotope fractionation during its adsorption on oxides and hydroxides. *J. Colloid Interf. Sci.* **2005,** *291,* (1), 192-200.

19. Juillot, F.; Maréchal, C.; Ponthieu, M.; Cacaly, S.; Morin, G.; Benedetti, M.; Hazemann, J. L.; Proux, O.; Guyot, F., Zn isotopic fractionation caused by sorption on goethite and 2-Lines ferrihydrite. *Geochim. Cosmochim. Acta* **2008,** *72,* (19), 4886-4900.

20. Gou, W.; Li, W.; Ji, J.; Li, W., Zinc isotope fractionation during sorption onto Al oxides: Atomic level understanding from EXAFS. *Environ. Sci. Technol.* **2018,** *52,* (16), 9087-9096.

21. Wasylenki, L. E.; Howe, H. D.; Spivak-Birndorf, L. J.; Bish, D. L., Ni isotope fractionation during sorption to ferrihydrite: Implications for Ni in banded iron formations. *Chem. Geol.* **2015,** *400,* 56-64.

22. Wasylenki, L. E.; Swihart, J. W.; Romaniello, S. J., Cadmium isotope fractionation

during adsorption to Mn oxyhydroxide at low and high ionic strength. *Geochim. Cosmochim. Acta* **2014,** *140*, 212-226.

23. Weiss, D.; Northover, G.; Hanif, M.; García-España, E.; Vilar, R.; Arnold, T.; Markovic, T.; Wissuwa, M.; Delgado, E., Isotope fractionation of zinc in the paddy rice soil-water environment and the role of 2'deoxymugineic acid (DMA) as zincophore under Zn limiting conditions. *Chem. Geol.* **2021,** *577*.

24. Aucour, A.-M.; Bedell, J.-P.; Queyron, M.; Tholé, R.; Lamboux, A.; Sarret, G., Zn speciation and stable isotope fractionation in a contaminated urban wetland soil–Typha latifolia system. *Environ. Sci. Technol.* **2017,** *51*, (15), 8350-8358.

25. Nelson, J.; Wasylenki, L.; Bargar, J. R.; Brown, G. E.; Maher, K., Effects of surface structural disorder and surface coverage on isotopic fractionation during Zn(II) adsorption onto quartz and amorphous silica surfaces. *Geochim. Cosmochim. Acta* **2017,** *215*, 354-376.

26. Skulan, J. L.; Beard, B. L.; Johnson, C. M., Kinetic and equilibrium Fe isotope fractionation between aqueous Fe(III) and hematite. *Geochim. Cosmochim. Acta* **2002,** *66*, (17), 2995-3015.

27. Clayton, R. E.; Hudson-Edwards, K. A.; Malinovsky, D.; Andersson, P., Fe isotope fractionation during the precipitation of ferrihydrite and transformation of ferrihydrite to goethite. *Mineral. Mag.* **2005,** *69*, (5), 667-676.

28. Balci, N.; Bullen, T. D.; Witte-Lien, K.; Shanks, W. C.; Motelica, M.; Mandernack, K. W., Iron isotope fractionation during microbially stimulated Fe(II) oxidation and Fe(III) precipitation. *Geochim. Cosmochim. Acta* **2006,** *70*, (3), 622-639.

29. Guinoiseau, D.; Galer, S. J. G.; Abouchami, W., Effect of cadmium sulphide precipitation on the partitioning of Cd isotopes: Implications for the oceanic Cd cycle. *Earth and Planet. Sci. Lett.* **2018,** *498*, 300-308.

30. Johnson, C. M.; Skulan, J. L.; Beard, B. L.; Sun, H.; Nealson, K. H.; Braterman, P. S., Isotopic fractionation between Fe(III) and Fe(II) in aqueous solutions. *Earth Planet. Sci. Lett.* **2002,** *195*, (1), 141-153.

31. Hu, B.; Yan, X.; Wang, W.; Li, Y.; Li, H.; Hong, M.; Liu, F.; Yin, H., Iron oxyhydroxide polytype (γ-, δ- and β-FeOOH) structures govern Zn mobility. *Chem. Geol.* **2022**, 121167.

32. Manceau, A.; Nagy, K. L.; Spadini, L.; Ragnarsdottir, K. V., Influence of anionic layer structure of Fe-oxyhydroxides on the structure of Cd surface complexes. *J. Colloid Interface Sci.* **2000,** *228*, (2), 306-316.

33. Liu, L.; Wang, X. M.; Zhu, M. Q.; Ma, J. Y.; Zhang, J.; Tan, W. F.; Feng, X. H.; Yin, H.; Liu, F., The speciation of Cd in Cd–Fe coprecipitates: Does Cd substitute for Fe in goethite structure? *ACS Earth Space Chem.* **2019,** *3*, (10), 2225-2236.

34. Gou, W.; Li, W.; Siebecker, M. G.; Zhu, M.; Li, L.; Sparks, D. L., Coupling molecular-scale spectroscopy with stable isotope analyses to investigate the effect of Si on the mechanisms of Zn–Al LDH formation on Al oxide. *Environ. Sci. Technol.* **2022,** *56*, (19), 13829-13836.

35. Zhu, C.; Liao, S.; Wang, W.; Zhang, Y.; Yang, T.; Fan, H.; Wen, H., Variations in Zn and S isotope chemistry of sedimentary sphalerite, Wusihe Zn-Pb deposit, Sichuan Province, China. *Ore Geol. Rev.* **2018,** *95*, 639-648.

36. Pallavicini, N.; Engström, E.; Baxter, D. C.; Öhlander, B.; Ingri, J.; Rodushkin, I., Cadmium isotope ratio measurements in environmental matrices by MC-ICP-MS. *J. Anal. At. Spectrom.* **2014,** *29,* (9), 1570-1584.

37. Archer, C.; Andersen, M. B.; Cloquet, C.; Conway, T. M.; Dong, S.; Ellwood, M.; Moore, R.; Nelson, J.; Rehkämper, M.; Rouxel, O.; Samanta, M.; Shin, K.; Sohrin, Y.; Takano, S.; Wasylenki, L., Inter-calibration of a proposed new primary reference standard AA-ETH Zn for zinc isotopic analysis. *J. Anal. Atom. Spectrom.* **2017,** *32,* (2), 415-419.

38. Wang, Z.-Z.; Liu, S.-A.; Liu, J.; Huang, J.; Xiao, Y.; Chu, Z.-Y.; Zhao, X.-M.; Tang, L., Zinc isotope fractionation during mantle melting and constraints on the Zn isotope composition of Earth's upper mantle. *Geochim. Cosmochim. Acta* **2017,** *198,* 151-167.

39. Ravel, B.; Newville, M., ATHENA, ARTEMIS, HEPHAESTUS: data analysis for X-ray absorption spectroscopy using IFEFFIT. *J. Synchrotron Radiat.* **2005,** *12,* 537-541.

40. Kelly, S. D.; Hesterberg, D.; Ravel, B., Analysis of Soils and Minerals Using X-ray Absorption Spectroscopy. In *Methods of Soil Analysis Part 5—Mineralogical Methods*, Soil Science Society of America: 2008; pp 387-463.

41. Schauble, E. A., Applying stable isotope fractionation theory to new systems. *Rev. Mineral. Geochem.* **2004,** *55,* (1), 65-111.

42. Waychunas, G. A.; Fuller, C. C.; Davis, J. A.; Rehr, J. J., Surface complexation and precipitate geometry for aqueous Zn(II) sorption on ferrihydrite: II. XANES analysis and simulation. *Geochim. Cosmochim. Acta* **2003,** *67,* (5), 1031-1043.

43. Cismasu, A. C.; Levard, C.; Michel, F. M.; Brown, G. E., Properties of impurity-bearing ferrihydrite II: Insights into the surface structure and composition of pure, Al- and Si-bearing ferrihydrite from Zn(II) sorption experiments and Zn K-edge X-ray absorption spectroscopy. *Geochim. Cosmochim. Acta* **2013,** *119,* 46-60.

44. Aucour, A.; Bedell, J.; Queyron, M.; Magnin, V.; Testemale, D.; Sarret, G., Dynamics of Zn in an urban wetland soil–plant system: Coupling isotopic and EXAFS approaches. *Geochim. Cosmochim. Acta* **2015,** *160,* 55-69.

45. Frierdich, A. J.; Catalano, J. G., Controls on Fe(II)-activated trace element release from goethite and hematite. *Environ Sci. Technol.* **2012,** *46,* (3), 1519-1526.

46. Frierdich, A. J.; Scherer, M. M.; Bachman, J. E.; Engelhard, M. H.; Rapponotti, B. W.; Catalano, J. G., Inhibition of trace element release during Fe(II)-activated recrystallization of Al-, Cr-, and Sn-substituted goethite and hematite. *Environ. Sci. Technol.* **2012,** *46,* (18), 10031-10039.

47. Trivedi, P.; Axe, L.; Tyson, T. A., An analysis of zinc sorption to amorphous versus crystalline iron oxides using XAS. *J. Colloid Interface Sci.* **2001,** *244,* (2), 230-238.

48. Roberts, D. R.; Ford, R. G.; Sparks, D. L., Kinetics and mechanisms of Zn complexation on metal oxides using EXAFS spectroscopy. *J. Colloid Interf. Sci.* **2003,** *263,* (2), 364-376.

49. Waychunas, G. A.; Fuller, C. C.; Davis, J. A., Surface complexation and precipitate geometry for aqueous Zn(II) sorption on ferrihydrite I: X-ray absorption extended fine structure spectroscopy analysis. *Geochim. Cosmochim. Acta* **2002,** *66,* (7), 1119-1137.

50. Bochatay, L.; Persson, P., Metal ion coordination at the water–manganite (γ-

MnOOH) interface: II. An EXAFS study of zinc(II). *J. Colloid Interf. Sci.* **2000,** *229*, (2), 593-599.

51. Kaur, N.; Gräfe, M.; Singh, B.; Kennedy, B., Simultaneous incorporation of Cr, Zn, Cd, and Pb in the goethite structure. *Clays Clay Miner.* **2009,** *57*, (2), 234-250.

52. Bryan, A. L.; Dong, S.; Wilkes, E. B.; Wasylenki, L. E., Zinc isotope fractionation during adsorption onto Mn oxyhydroxide at low and high ionic strength. *Geochim. Cosmochim. Acta* **2015,** *157*, 182-197.

53. Balistrieri, L. S.; Borrok, D. M.; Wanty, R. B.; Ridley, W. I., Fractionation of Cu and Zn isotopes during adsorption onto amorphous Fe(III) oxyhydroxide: Experimental mixing of acid rock drainage and ambient river water. *Geochim. Cosmochim. Acta* **2008,** *72*, (2), 311-328.

54. Fujii, T.; Moynier, F.; Blichert-Toft, J.; Albarède, F., Density functional theory estimation of isotope fractionation of Fe, Ni, Cu, and Zn among species relevant to geochemical and biological environments. *Geochim. Cosmochim. Acta* **2014,** *140*, 553-576.

55. Liu, J.; Sheng, A.; Li, X.; Arai, Y.; Ding, Y.; Nie, M.; Yan, M.; Rosso, K. M., Understanding the importance of labile Fe(III) during Fe(II)-catalyzed transformation of metastable iron oxyhydroxides. *Environ. Sci. Technol.* **2022,** *56*, (6), 3801-3811.

56. Cornell, R. M.; Schwertmann, U., Formation. In *The iron oxides: structure, properties, reactions, occurrences, and uses*, Second ed.; John Wiley & Sons: 2003; pp 345-364.