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1 **Household's preferences for waste sorting systems: The role of values, socioeconomic**  
2 **characteristics, and contexts**

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9  
10 **Abstract:** Efficient Waste Management Systems (WMS) depend on citizens' willingness to sort  
11 waste and to cooperate for its improved provision. Thus, it is essential to understand what attributes  
12 individuals value the most and what drives WMS preference heterogeneity. In this paper we  
13 investigate how individuals value WMS attributes (e.g., the number of waste sorting categories, the  
14 introduction of textile sorting, and frequency of collection per week) and how the local contexts,  
15 socioeconomics and environmental values shape their preferences. We collected data from a  
16 representative Italian sample and analysed them using hybrid mixed discrete choice models. We  
17 found that individuals are willing to pay for an improved WMS, even if this implies an increased  
18 effort, from their part, and storage space. In addition, we found that the extent to which individuals  
19 support WMS improvements strongly depends on their environmental values, experiences (shaped  
20 through their local context) and socioeconomic factors. Our results suggest that to increase recycling  
21 rates and citizen satisfaction, there is a need for context-specific WMS designs that acknowledge the  
22 observed heterogeneity of preferences.

23  
24 **Keywords:** Waste sorting systems, Circular Economy, willingness to pay (WTP), hybrid choice  
25 models

27        **1. Introduction**

28        A circular economy is increasingly seen as a pillar for achieving Global Sustainability Goals. Several  
29        governmental bodies, including the European Union, are putting strong circular economy targets at  
30        the top of their policy agenda (Camana et al., 2021). To achieve these goals, we need a waste  
31        management system (WMS) that efficiently recovers garbage from the source to treatment facilities  
32        and, thus, a system in which end consumers recycle their waste (Di Foggia and Beccarello, 2021).  
33        Thus, the individuals' willingness to sort waste constitutes a key pillar for an efficient WMS.  
34        Therefore, understanding people's preferences, and thus, willingness to participate in the provision  
35        of this quasi-public good, is essential for the success of a circular economy.

36        Because waste sorting is a time, space and effort consuming activity, households' motivation for  
37        cooperating with the WMS has been widely studied in the literature (Berglund, 2006; Gilli et al.,  
38        2018; Czajkowski et al., 2019; Aprile and Fiorillo, 2019). It has been found that social norms,  
39        personal values, environmental contexts, and demographics are among the main drivers of  
40        individuals' willingness to recycle (Degli Antoni and Vitucci Marzetti, 2019; D'Amato, 2016;  
41        Vassanadumrongdee and Kittipongvises, 2018). Moreover, recent literature argues that economic  
42        incentives play a small role in recycling when compared to individual and social norms and values  
43        (Berglund, 2006; Degli Antoni and Vitucci Marzetti, 2019). Consequently, understanding how values  
44        and socioeconomics shape a households' willingness to contribute to WMS is essential to its efficient  
45        design.

46        Using Italy as a case study, in this paper, we uncover how environmental values, socioeconomic and  
47        local factors shape the willingness to cooperate for an improved WMS by examining households'  
48        stated willingness to pay (WTP) for improved WMS attributes. We use a hybrid mixed logit to  
49        disentangle the impact of socioeconomics on waste sorting through its impact on environmental  
50        values and through its impact on other specific traits of the demographic group. In addition, we

51 analyze how the local recycling rate, population density, and type of collection system, which are  
52 factors that impact a household's environment, shape the preferences for an improved WMS. To the  
53 best of our knowledge, this issue has not been empirically studied in the literature.

54 Italy constitutes an interesting case study as the recycling rate varies widely across regions, with high  
55 recycling rates (around 71%) in the north and medium-low rates in the south (around 54%)<sup>1</sup>. As in  
56 Italy the WMS is managed at the local level, this heterogeneity has been linked to differences in the  
57 economic and social development across regions, which in turn translates to a lower capacity to  
58 secure collection and treatment of waste (Cerqueti et al., 2021; Di Foggia and Beccarello, 2021;  
59 Lombardi et al., 2021; Chakraborty et al., 2022; Romano et al., 2022). Our study analyses the  
60 attributes households living in low- and high-recycling rate municipalities value the most.

61 In sum, our research contributes to the stream of literature on efficient WMS designs as it allows to  
62 disentangle the main factors affecting the heterogeneity of households' preferences for improved  
63 WMS designs. Therefore, considering different contexts and socioeconomics, our results give  
64 important insights to policymakers on how to increase households' wellbeing and willingness to  
65 recycle.

66 The rest of the paper is organized as follows. Section 2 reviews households' willingness to pay for  
67 improved WMS and their motivations to recycle. Section 3 introduces the data and experimental  
68 design used in this study. Section 4 presents the methodology used. Section 5 presents the results,  
69 whilst section 6 discusses them. Finally, section 7 concludes.

## 70 **2. Literature review**

71 Stated preference (SP) surveys have received increasing attention in recent years as a method to  
72 understand households' intention to recycle and cooperate for WMS improvements, due to their low  
73 costs and simplicity compared to randomised control trials (Czajkowski et al., 2019). Examples

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<sup>1</sup> Statistics obtained for the Superior Institute for Environmental Protection and Research available at:  
<https://www.catasto-rifuti.isprambiente.it/index.php?pg=nazione>

74 include Ferreira and Cunha Marques (2015), Song et al. (2019), Vassanadumrongdee and  
75 Kittipongvises (2018), and Kayamo (2020), who used contingent valuation methods (CVM) to  
76 understand households' willingness to pay for the introduction of waste recycling in Portugal,  
77 Bangladesh, Thailand and Ethiopia, respectively.

78 Discrete choice experiments (DCE) have gained popularity within SP methods in estimating  
79 preferences for WMS improvements. Unlike other SP approaches, DCE allows valuating multiple  
80 attributes and calculating the trade-offs between them. Overall, besides costs, the most popular  
81 WMS attributes valued in the literature are the collection frequency and the number of waste-  
82 sorting categories (Karousakis and Birol, 2008; Czajkowski et al., 2014, Benyam et al., 2020). In  
83 general, higher collection frequency has been linked to higher WTP levels, whilst mixed results have  
84 been found with respect to the number of sorting categories.

85 Because waste-sorting behaviour does not seem to follow an economic logic, recent literature has  
86 studied the main factors influencing willingness to cooperate on recycling. For example, studies  
87 have addressed the role of certain demographic characteristics on recycling intention and behavior,  
88 such as age, gender, education level, income, and household characteristics (Vassanadumrongdee  
89 and Kittipongvises, 2019; Benyam et al., 2020; Romano et al., 2022). However, the impact of  
90 demographics on willingness to sort waste has received mixed results and, in general, was found to  
91 have a small impact when compared to psychological and institutional factors such as values and  
92 attitudes (Aprile and Fiorillo, 2019; Romano et al 2022).

93 Overall, the main psychological factors influencing willingness to recycle include habits, attitudes,  
94 perceptions, values, and social and subjective norms (Massarutto et al., 2019, Czajkowskia et al.,  
95 2019, Agovino et al., 2019; Degli Antoni and Marzetti, 2019 ). Examples of such factors include  
96 environmental values, trust, peer-pressure, perceived environmental benefits and social rewards.  
97 Considering the importance of psychological factors on recycling behaviour, a stream of research

98 has aimed to include these factors in estimations of recycling intention. Some studies have directly  
99 used indicators of these unobservable psychological factors to measure their impact. For example,  
100 Aprile and Fiorillo (2019) used indicators of environmental concerns to empirically measure their  
101 impact on recycling behaviour in Italy. Others have measured these psychological factors using  
102 more sophisticated factor analysis techniques. For instance, Vassanadumrongdee and Kittipongvises  
103 (2019) used factor analysis to estimate the influence of latent traits such as perception of  
104 inconvenience, pro-environmental attitudes, and subjective norms towards waste-sorting on WTP  
105 estimates for WSM improvement, finding significant and strong impacts. In addition, latent class  
106 models (LCM) have also been used for this purpose; examples include Massarutto et al. (2019) and  
107 Nainggolan et al. (2019). Nevertheless, LCM, though useful, is disadvantageous in relying too much  
108 on the researchers' intuition about the drivers of the observed class heterogeneity (Weller et al.,  
109 2020).

110 In conclusion, although the individuals' willingness to cooperate for an improved WMS has been  
111 systematically assessed in the literature, a gap remains in understanding the role of local contexts in  
112 shaping those preferences. In addition, to the best of our knowledge, previous research on the  
113 impact of socioeconomics on WMS preferences has not analyzed whether socioeconomics exert an  
114 influence through their association with environmental values or through their association with  
115 some other traits common to the demographic group. The present work aims to fill these gaps.  
116 Finally, to measure environmental awareness values we use the hybrid modelling technique, which  
117 despite having been recognized as a robust method to include latent traits in WTP estimates (Ben-  
118 Akiva et. Al, 2002), its use has been very limited in the literature on WMS preferences.

### 119 **3. Data and experimental design**

120 Our research is built upon a national stated preference survey of Italian households carried out  
121 online in April 2020 with the support of a professional market research company specialized in

122 opinion surveys. Because the contract with the market research company contemplated all  
123 questionnaires compiled, there is a null non-response rate. We collected 605 questionnaires. After  
124 cleaning the data by removing protesters, we analyzed the data of 547 respondents from 272  
125 municipalities.

126 The survey was developed based on widely-accepted guidelines for DCE (Riera et al. 2012; Johnston  
127 et al. 2017) and it was comprised of three parts. In the first section, we collected information on the  
128 WMS. In the second section, we described the attributes and levels of the DCM. Subsequently, we  
129 presented 12 choice situations. In the final section we collected attitudinal indicators of pro-  
130 environmental values and socio-demographic characteristics.

131 In Italy, WMS is managed at the local level (municipality). In all regions, households pay a municipal  
132 waste management tariff (TARI<sup>2</sup>) which comprises fixed costs (determined based on the costs of the  
133 service, i.e., investments and depreciations) and variable costs (to finance waste transport, collection  
134 recycling, and disposal) related to the amount of waste produced. For households, the TARI is  
135 calculated considering the surface of the house and the family size. In some municipalities, the  
136 quantity of residual waste produced and recycled is also considered. Each municipality sets the  
137 TARI with respect to the quantity and quality of waste produced. Thus, municipalities with a higher  
138 average TARI tend to have lower recycling rates (Confcommercio, 2022; Lombardi et al., 2021).  
139 This was reflected in our data, as a significant negative correlation was found between the municipal  
140 recycling rate and the TARI paid<sup>3</sup>.

141 Table 1 portrays the descriptive statistics from the respondents in our sample and from their  
142 municipality, including general contextual variables (recycling rate and population density) and  
143 attributes of the WMS (TARI paid in 2020, type and frequency of collection), compared with the  
144 Italian census data, when applicable. In terms of WMS attributes, the average TARI paid in 2020 by

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<sup>2</sup> In Italian *Tassa sui rifiuti*.

<sup>3</sup> The correlation size was -0.17, the Pearson's product moment correlation coefficient was -4.05.

145 sample's respondents was 274€, but the exact amount varied widely across respondents as the  
 146 standard deviation was 147€. In terms of socioeconomics, our sample was representative of the  
 147 Italian population with the exception that it was slightly more educated.

148 **Table 1: Socio-demographic characteristics of respondents compared to Italian population,**  
 149 **WMS attributes and context-specific variables**

	Variable	Sample range/Categories	Sample distribution			Italian population
			N	Mean	s.d.	Mean
Sociodemographic	Gender (Categorical)	0 if Male	284	0.520		0.482
		1 if Female	263	0.480		0.518
	Age (continuous)	Age in years. 18 years lower limit, 65 years higher limit		43.40	12.2	
		18-24 years		0.084		0.082
		25-34 years		0.188		0.127
		35-44 years		0.221		0.154
		45-54 years		0.269		0.192
		55-64 years		0.223		0.166
	Bachelor (Categorical)	> 65 years		0.015		0.276
		0 if educational attainment is high-school or lower	323	0.590		0.821
Household size (Categorical)	1 if educational attainment is bachelor's degree or higher	224	0.410		0.179	
	0 if household size is lower than 3 family members	100	0.182		0.58	
Income (Categorical)	1 if household size is equal or higher than 3 family members	449	0.820		0.41	
	0 if annual income after tax is lower than 40,000 EUR	490	0.896		0.923	
Political orientation (Categorical)	1 if annual income after tax is higher or equal than 40,000 EUR	57	0.104		0.077	
	0 if center or right	344	0.629			
WMS attributes	Actual Collection Frequency (continuous)	1 if left	203	0.371		
		1 time per week	67	0.122		
	2 times per week	121	0.221			
	3 times per week	101	0.185			
	4 times per week	37	0.068			
	5 times per week	66	0.121			
	6 times per week	109	0.199			
Type of Collection (Categorical)	7 times per week	46	0.084			
	0 if collection type is door-to-door or mixed collection	299	0.547			
TARI (Continuous)	1 if collection type is curbside collection	248	0.453			
	TARI paid 2020. 30 euros lower limit, 1,000 euros higher limit		274	147		
Recycling rate (continuous)	Proportion of waste recycled in 2020 per municipality.					
	10% lower limit, 87% higher limit (ISPRA, 2020)		56%	17%		
Population density (continuous)	1000 habitants per Km <sup>2</sup> in municipality. 0.1 lower limit,					
	12.1 higher limit (ISTAT, 2022)		2.41	2.26		

Notes: Source of descriptive statistics for Italian population: Istituto Nazionale di Statistica (2020)

150  
 151 Following a review of the literature and after performing a focus group with participants recruited  
 152 from the general population, we included four attributes in our choice modeling experiment:



153 additional waste sorting categories, the introduction of textile sorting<sup>4</sup>, frequency of collection per  
 154 week, and costs in the form of a percentual increase in the TARI. A summary of attributes and  
 155 levels, which were also selected based on the literature and focus group discussion, is presented in  
 156 table 2. Following purely utilitarian views, there is no rationale for individuals to positively evaluate  
 157 the first two attributes, i.e., additional waste-sorting categories and the introduction of textile  
 158 collection, as they both imply an increase in effort, time, and space. Thus, for these variables, it is  
 159 key to understand how non-utilitarian factors influence the willingness to pay and engage in  
 160 recycling. In this investigation, we focused on environmental values as the rationale for sorting waste  
 161 is linked to environmental sustainability targets.

162 **Table 2: Attributes and levels of experiment**

Attribute	Description	Levels
Categories of waste	Number of additional waste sorting categories	<i>No more categories</i> <i>+1</i> : 1 additional category <i>+2</i> : 2 additional categories
Textile collection	Introduction of textile collection	<i>Yes</i> <i>No</i>
Collection frequency	Number of times per week the collection of one or more categories of waste is carried out	<i>3 times</i> <i>5 times</i> <i>7 times</i>
Costs	Percentage increase of the annual waste management tariff (TARI)	<i>+ 3%</i> : increase by 3% <i>+ 5%</i> : increase by 5% <i>+ 7%</i> : increase by 7% <i>+ 10%</i> : increase by 10%

163  
 164 To measure environmental attitudes, we used the New Ecological Paradigm (NEP) scale (Dunlap et  
 165 al., 2000). The construct validity of NEP has been widely tested with positive results, making it one  
 166 of the most popular measures of environmental values and attitudes (Matsiori et al., 2020). Lately, it  
 167 has also been used to estimate the impact of pro-environmental attitudes on WTP estimates (see  
 168 Tyllianakis and Ferrini (2021), Bartczak (2015), and Faccioli et al. (2020), among others). The revised  
 169 NEP scale consists of 15 items and uses a 5-point Likert response scale (see table A.1 in the

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<sup>4</sup> At the time of the research, in Italy textile materials were thrown into undifferentiated waste; they were not recycled. As of January 1 2022, through legislative decree 116/2020, the obligation to separate textile waste came into effect, put forward by three years the implementation of one of the decrees contained in the "Package of directives on the circular economy" adopted by the European Union in 2018. However, the entire supply chain is still not truly operational and functional.

170 appendix). After performing exploratory factor analysis, item 6 of the revised NEP was deleted from  
 171 our study as it did not provide a significant loading to the latent construct in our sample (see table  
 172 A.2 in appendix).  
 173 Finally, an efficient design (Rose and Bliemer, 2009) generated with Ngene software (ChoiceMetrics,  
 174 2018), was used to generate the 12 choice situations. The priors were obtained through a pilot survey  
 175 involving 100 respondents. The design of the pilot was also made through an efficient design, using  
 176 priors from the literature. In addition, although cost levels were presented as a percentual increase of  
 177 the TARI, we also calculated a family estimate of the value in euros and presented it in the choice  
 178 cards alongside the percentage increase. An example of the choice card is portrayed in figure 1.

179 **Figure 1: Example of choice card for respondents that paid a TARI equal to 200 EUR in**  
 180 **2020**

Please indicate whether you prefer system A or system B, or neither.			
	System A	System B	Neither
Categories of waste	No more categories	+2	I prefer the current WMS
Collection of textile waste	No	Yes	
Collection frequency	7 times	5 times	
Costs	7% (14 EUR)	3% (6 EUR)	
I CHOOSE:	<input type="radio"/>	<input type="radio"/>	<input type="radio"/>

181

#### 182 4. Methodology

183 Our methodological approach is built upon the use of hybrid choice models, which have gained  
 184 increasing recognition for assessing the impact of latent variables (LVs) in WTP estimates such as  
 185 attitudes, perceptions, and values (Walker, 2001). Hybrid models allow to estimate and include LVs  
 186 in the utility function, and thus to compute their impact on WTP values. In addition, they are  
 187 considered superior to two-step factor analytical approaches, as they simultaneously estimate the  
 188 LVs and the coefficients of the utility function, avoiding endogeneity and omitted-variable biases  
 189 (Vij and Walker, 2016). Furthermore, the LVs are estimated using a multiple causes and indicators  
 190 (MIMIC) approach, which allows uncovering how socioeconomic and other observable variables

191 shape LVs. Therefore, the MIMIC approach generates a valuable framework for making  
 192 predictions.

193 Hybrid choice models comprise three components. The first is the discrete choice part, which  
 194 models individual's utility based on the random utility theory (Walker, 2001). The second and third  
 195 are the measurement and structural equations of the MIMIC framework, which model the LVs.  
 196 Below we explain in detail each of these components in the context of our case study.

#### 197 4.1. Discrete choice model

198 Following the random utility theory, the utility that an individual  $i$  obtains for the proposed changes  
 199 on the WMS  $j$  and the status quo  $s$  in each choice situation  $k$  can be modelled as:

$$200 U_{ijk} = \theta c_{jk} + \beta X_{jk} + \alpha EA_i X_{jk} + \delta S_i X_{jk} + \varepsilon_{ijk} + \epsilon_i \quad (1)$$

$$201 U_{isk} = \theta c_{sk} + \beta X_{sk} + \alpha EA_i X_{sk} + \delta S_i X_{sk} + \varepsilon_{isk} + \gamma_i$$

202 Where  $c_{jk}$  and  $c_{sk}$  are the costs of the proposed improvements in the WMS and of the status, thus  
 203  $\theta$  is the marginal utility of income.  $X_{jk}$  and  $X_{sk}$  are vectors of the attributes of the alternatives (i.e.,  
 204 frequency of collection, introduction of textile sorting, and additional sorting categories).  $\beta$  is the  
 205 vector of baseline coefficients of the WMS attributes  $X$ .  $EA_i$  is the latent variable *environmental*  
 206 *awareness* (EA) measured via the NEP scale, thus  $\alpha$  is a vector of coefficients measuring EA's impact  
 207 on the preferences for WMS's attributes  $X$ . Similarly,  $S_i$  is a vector of socioeconomics and  
 208 contextual characteristics of the individuals, and  $\delta$  is a vector of coefficients measuring the impact of  
 209 demographics and the surroundings on the preferences for  $X$ .  $\varepsilon_{isk}$  are independent but identically  
 210 distributed (IID) Gumbel error components reflecting all unobservable idiosyncratic factors  
 211 influencing individual's preferences. Finally,  $\epsilon_i$  and  $\gamma_i$  are error terms  $N(0, \sigma_{panel}^2)$  with a twofold  
 212 function. First,  $\epsilon_i$  and  $\gamma_i$  capture the correlation effect among observations of the same individual,  
 213 so called pseudo-panel effect (Hess et al., 2008; Cantillo et al., 2007). Secondly,  $\epsilon_i$  and  $\gamma_i$  capture the

214 correlation of the alternatives proposing a change in the WMS  $j = (1,2)$ , thus relaxing the  
 215 assumption of Independence from Irrelevant Alternatives (IIA) (Walker, 2001).  
 216 We estimate a random parameters logit in WTP space<sup>5</sup> (Train and Weeks, 2005) to capture  
 217 unobserved preference heterogeneity. We propose a negative log-normal distribution for the cost  
 218 coefficient  $\theta \sim \text{Lognormal}(\mu_\theta, \sigma_\theta^2)$  and a normal distribution for the baseline coefficients  $\beta$   
 219  $N(\mu_\beta, \sigma_\beta^2)$ , whilst the interaction coefficients ( $\alpha$  and  $\delta$ ) are considered constant.

#### 220 4.2. MIMIC framework: structural and measurement equations

221 The latent variable EA is measured following the MIMIC approach, composed of measurement and  
 222 structural equations. The measurement equations relate the weight that the latent variable EA has in  
 223 explaining the observable NEP indicators ( $NEP_{ir}$ ), specifically it portrays that each indicator  $r$  can  
 224 be explained by the latent variable  $EA_i$ :

$$225 \quad NEP_{ir} = \gamma_r EA_i + e_{ir} \quad (2)$$

226 Where  $\gamma_r$  reflects the degree in which the latent variable  $EA_i$  explains each of the observable  $r$  NEP  
 227 indicators.  $e_{ir}$  is an error term following a normal distribution. As the NEP indicators are measured  
 228 with a 5-point Likert scale, equation (2) is estimated using an ordered regression.

229 On the other hand, the structural equation measures the influence of different observable variables  
 230 such as socioeconomics characteristics on the latent variable EA:

$$231 \quad EA_i = \eta S_i + \xi_i \quad (3)$$

232 Where  $\eta$  is a vector measuring the impact of each observable variable  $S_i$  on  $EA_i$ , whilst  $\xi_i$  is an error  
 233 term following a normal distribution. Through (3) we can, therefore, measure the impact of  
 234 demographics and contextual variables  $S_i$  on individual's environmental attitudes. Therefore, to  
 235 estimate the overall impact of  $S_i$  on the preferences for each WMS's attribute  $X$ , we need to

---

<sup>5</sup> To get estimations in WTP space we multiply equation (1) by the cost coefficient.

236 consider both its direct impact, through  $\delta$  in (1), and its indirect impact, through  $\eta$  in (3) and  $\alpha$  in  
237 (1).

238 Equations 1-3 are estimated using simulated maximum likelihood method (Walker, 2002). We used  
239 the package APOLLO in R (Hess and Palma, 2019). One thousand random Halton draws were used  
240 in the estimations.

## 241 **5. Results**

242 To understand the average preferences in our case-study, we first estimate random-parameters logit  
243 without interaction effects. The results are portrayed in table 3. In line with previous investigations  
244 (i.e., Czajkowski et al., 2014; Massarutto et al., 2019; Benyam et al., 2020), a higher frequency of  
245 waste collection is positively valued by respondents. In our case, individuals are, on average, willing  
246 to pay 1.08€ to increase the weekly frequency of collection by one day. Our results also show that  
247 respondents have an average WTP of 3.68€ and 3.77€ to increase, respectively, by one and two the  
248 number of sorting categories. In addition, the introduction of textile collection is the attribute that  
249 individuals, on average, value the most, with a WTP of 6.54€. The positive valuation of extra  
250 waste-sorting categories and of textile collection supports the thesis that non-utilitarian factors play a  
251 key role in explaining WTP estimates concerning WMS improvements. On the other hand, our  
252 results show that, on average, respondents hold lower preferences from the status quo.

253 Concerning preference heterogeneity, the high unobserved heterogeneity of textile sorting and  
254 frequency of collection is noteworthy, with a standard deviation of 7.77€ and 3.58€, respectively. In  
255 contrast, the standard deviation of the addition of one or two sorting categories is lower and less  
256 significant. Interestingly, the attribute that was absent in the Italian WMS when the survey was  
257 collected, i.e., textile collection, had the highest preference heterogeneity. We believe that this may  
258 be caused by the lack of experience of individuals toward this attribute, causing skepticism or over-  
259 excitement on its environmental benefits and feasibility.

260 **Table 3: Random parameters logit results in WTP space**

Attribute	Mean coefficients	Standard deviations
ASC status quo	-3.34*** (0.63)	3.85*** (0.14)
+1 category	3.68*** (0.52)	0.59 (0.37)
+2 category	3.77*** (0.51)	1.36** (0.55)
Textile collection	6.54*** (0.49)	7.77*** (0.68)
Frequency	1.08*** (0.16)	3.58*** (0.27)
Costs	-2.43*** (0.15)	1.60*** (0.08)
Pseudo-panel		3.17*** (0.2)
Number of individuals	547	
Number of observations	6564	
Log-likelihood	-4,642.46	
AIC	9310.93	
BIC	9399.19	
McFadden's R2	0.34	

Notes: Standard deviations in parenthesis  
 \*\* and \*\*\* indicate significance levels at 5% and 1%, respectively

261  
 262  
 263 Although the results in table 3 display the overall picture of WMS's preferences, they do not address  
 264 the cause of the heterogeneity of preferences for WMS's attributes, nor to what extent context-  
 265 specific and socioeconomic variables are influencing WTP estimates. For this purpose, we estimated  
 266 a hybrid random-parameters logit with interaction effects (see equations 1-3). Results for the discrete  
 267 choice component are portrayed in table 4 and for the MIMIC component in table 5. The  
 268 description of the covariates can be found in table 1. It is worth mentioning that the covariates  
 269 *Population density* and *Recycling rate* were centered at zero to avoid any spurious impact on the  
 270 alternative specific constant.  
 271 Table 4 shows that the Environmental Awareness (EA) construct measured with the NEP scale has  
 272 a significant and positive impact on WTP for improvements in WMS. In particular, individuals with  
 273 high EA positively value increases in sorting categories. The marginal impact of EA on the WTP  
 274 for 1 and 2 additional categories are 0.8€ and 1.5€, respectively. Textile sorting is the attribute that

275 high EA individuals value the most, with a marginal WTP of 2.43€. Finally, regarding the frequency  
276 of collection, WTP also increases with EA values, with a marginal WTP of 0.45€.

277 We now turn to table 5 to understand the impact of demographics on EA values. Curbside  
278 collection was not included as a predictor of EA, as theoretically there should not be a link between  
279 this covariate and EA. Consistent with the literature (Dunlap, 2000), we found significant evidence  
280 suggesting that female and political left-wing individuals have higher EA. In addition, in our case  
281 study, older and higher-income individuals have higher EA values.

282 Our results also suggest that individuals living in high recycling rate municipalities develop higher  
283 EA values. We recognize that there may be simultaneity in the causality of this relationship as  
284 individuals with higher EA values may also tend to recycle more. Nevertheless, as we use  
285 municipality data, we can assume that this variable is exogenous to the individual. In contrast,  
286 although not significant, individuals living in high population density cities with a household size  
287 equal to or bigger than 3 have higher EA values. Finally, an unexpected, though not significant,  
288 result emerges from the negative correlation of a bachelor's degree in the construct EA.

289 We turn back to the discrete choice estimates (table 4) to understand the impact of socio-  
290 demographics and environmental (context-specific) variables on WTP estimates. In doing so, it is  
291 important to bear in mind that these variables have a two-way impact on preferences: 1) There is an  
292 indirect impact through their influence on EA values, 2) there is a direct impact capturing other  
293 traits specific-to the demographic group. These direct and indirect values are calculated and  
294 portrayed in table A.3 of the appendix section.

295 In general, we find that the indirect impact of demographics on WMS preferences is lower than their  
296 direct impact. Therefore, our results portray that, in most cases, the influence of socio-demographic  
297 characteristics on WMS preferences through EA values is low when compared to other traits

298 specific to the group. This result is in line with Pienaar et al. (2013), who found that the relationship  
299 between socio-demographics and environmental values, measured by the NEP scale, is weak.  
300 Nevertheless, we find strong evidence suggesting that socio-demographics and context-specific  
301 variables shape individuals' preferences for WMS attributes. Focusing on direct estimates (table 4  
302 coefficients), we find that, despite having higher EA values, women and older individuals are less  
303 eager to increase household waste sorting categories. In addition, families with three or more  
304 members and high-income households have a higher willingness to pay for increasing waste sorting  
305 categories.

306 Our results also demonstrate that textile collection is less valued by bigger households and  
307 individuals from left-wing parties. However, it is important to bear in mind that these individuals  
308 also report higher EA values. In fact, for left-wing individuals, the positive indirect impact on WTP  
309 for textile collection through EA is almost as high as its negative direct impact. On the other hand,  
310 turning to the frequency of collection, we find that high income and left-wing individuals, who also  
311 have higher EA values, have higher preferences for this attribute.

312 An unexpected result arises from the lower WTP estimates of extra sorting categories and textile  
313 introduction from individuals with a bachelor's degree compared to their counterparts. Nevertheless,  
314 despite having an overall lower WTP for WMS attributes, individuals with a university degree have  
315 higher WTP to change from the status quo situation.

316 Turning to the local context variables *Curbside collection*, *Population density* and *Recycling rate*, we find that  
317 they have an important influence in shaping preferences for WMS improvements. In particular,  
318 individuals with a curbside collection scheme have lower willingness to increase their waste sorting  
319 categories and introduce textile collection in comparison to individuals with a door-to-door  
320 collection scheme. On the other hand, the willingness to change the WMS with respect to the status  
321 quo is lower for individuals with a curbside collection scheme.



322 Population density and recycling rate are significant predictors of preferences for textile collection  
323 and frequency of collection, partly explaining the significant heterogeneity of these attributes  
324 (measured by the standard deviation). First, we find that an increase in the recycling rate by 1% and  
325 by 1000 habitants per Km<sup>2</sup> in the population density, decrease WTP for a more frequent WMS by  
326 0.27€ and 0.02€, respectively. Second, our results also show that individuals living in more densely  
327 populated areas and with lower recycling rate have higher WTP for the introduction of textile  
328 sorting. A decrease in the recycling rate by 1% and an increase of population density by 1000  
329 habitants per Km<sup>2</sup> increases WTP for textile introduction by 0.08€ and 0.41€, respectively.  
330 Finally, we also find the standard deviation of our baseline estimates slightly decreases, portraying  
331 that some of the heterogeneity of preferences is captured by the analysis of covariates, i.e., values,  
332 socioeconomics, and the local context.

Attributes	Means	Standard deviations	Interactions									
			LV		Socioeconomic demographics					Local contexts		
			NEP	Female	Household size	Age	Bachelor	Political left	High Income	Population density	Recycling rate	Curbside collection
<b>ASC status quo</b>	10.18*** (3.39)	1.78** (0.69)	-0.22 (0.67)	-3.87*** (1.27)	1.66 (1.49)	-0.10* (0.06)	-5.19*** (1.21)	0.19 (1.02)	1.44 (1.60)	0.02 (0.29)	0.01 (0.06)	-5.21*** (1.07)
<b>+1 category</b>	5.07 (3.32)	0.82*** (0.20)	0.80** (0.39)	-3.26*** (0.95)	3.02** (1.19)	-0.03 (0.06)	-2.3** (1.06)	1.96 (1.20)	1.01 (1.18)	0.39 (0.24)	-0.03 (0.03)	-1.52* (0.79)
<b>+2 category</b>	9.12*** (3.2)	2.92*** (0.68)	1.50*** (0.50)	-1.45* (0.88)	1.71 (1.28)	-0.14** (0.05)	-3.36*** (1.17)	0.65 (1.06)	2.25* (1.32)	0.22 (0.25)	0.01 (0.04)	-1.62* (0.9)
<b>Textile collection</b>	10.81*** (2.52)	5.79*** (0.49)	2.43*** (0.30)	-0.99 (1.02)	-1.45* (0.83)	0.03 (0.04)	-5.16*** (0.89)	-1.59** (0.66)	-0.51 (1.2)	0.41** (0.18)	-0.08** (0.04)	-3.30*** (0.81)
<b>Frequency</b>	-0.51 (0.66)	2.24*** (0.28)	0.45*** (0.11)	0.08 (0.26)	0.01 (0.01)	0.01 (0.01)	-0.07 (0.27)	1.08*** (0.17)	1.01*** (0.38)	-0.27*** (0.04)	-0.02*** (0.01)	0.01 (0.20)
<b>Costs</b>	-2.71*** (0.14)	1.81*** (0.14)										
<b>Pseudo-panel</b>		3.41*** (0.24)										
<b>Number of individuals</b>							547					
<b>Number of observations</b>							6564					
<b>Log-likelihood (Whole model)</b>							-13260.55					
<b>Log-likelihood (Choice component)</b>							-4608.97					
<b>McFadden's R2 (Choice component)</b>							0.35					

Notes: Standard deviations in parenthesis

\*, \*\* and \*\*\* indicate significance levels at 10%, 5% and 1%, respectively

Recycling rate and population density were centered to zero

336

337 **Table 5: MIMIC coefficients of hybrid mixed logit results in WTP space with all**  
338 **interactions**

Measurement equations		Structural equations	
NEP 1	0.43*** (0.10)	Female	0.22*** (0.10)
NEP 2 (r)	1.44*** (0.13)	Age	0.02*** (0.00)
NEP 3	1.92*** (0.20)	Recycling rate	0.01* (0.00)
NEP 4 (r)	0.38*** (0.10)	Population density	0.03 (0.02)
NEP 5	3.01*** (0.33)	Household size	0.16 (0.13)
NEP 7	1.71*** (0.16)	Bachelor	-0.10 (0.09)
NEP 8 (r)	1.60*** (0.15)	High Income	0.34** (0.17)
NEP 9	1.08*** (0.13)	Political orientation	0.45*** (0.09)
NEP 10 (r)	1.29*** (0.13)		
NEP 11	0.91*** (0.12)		
NEP 12 (r)	1.61*** (0.15)		
NEP 13	1.88*** (0.19)		
NEP 14 (r)	1.21*** (0.13)		
NEP 15	2.93*** (0.31)		

339 Notes: Standard deviation in parenthesis  
340 \*, \*\* and \*\*\* indicate significance levels at 10%, 5% and 1%, respectively  
341 (r) indicates that the scale has been reversed  
342

343 **6. Discussion**

344 Our study highlights the importance of non-utilitarian factors and social contexts in shaping citizens’  
345 preferences for WMS, thus supporting previous research on the motivations of recycling behaviour  
346 (Massarutto et al., 2019; Czajkowski et al., 2019). Nevertheless, by analysing how the local  
347 environment and environmental awareness shape individuals’ willingness to cooperate for an  
348 improved WMS, our research makes a significant departure from the current stream of literature.

349 We found a significant impact of local contexts (curbside collection, recycling rate, and population  
350 rate) on individuals' WTP for WMS attributes, highlighting the importance of experiences and local  
351 context in shaping preferences. First, we found that individuals with a curbside collection scheme  
352 value less the introduction of extra-sorting categories and textile collection than those with a door-  
353 to-door collection scheme. In a curbside scheme, the collection point is further away from the  
354 household, thus the introduction of more categories involves extra-effort. Therefore, the collection  
355 scheme and number of categories should be assessed jointly on WMS designs, as they interact with  
356 individuals' utility. In addition, we also found that individuals with a curbside collection scheme have  
357 a higher willingness to change the WMS with respect to the status quo, thus, probably signaling  
358 dissatisfaction with the service. In fact, Lombardi et al. (2020) found that introducing a door-to-door  
359 collection scheme was linked to a higher recycling rate in Italian municipalities.

360 Another relevant finding arises from the interaction of the recycling rate with the studied WMS  
361 attributes. We found that individuals living in low recycling rate municipalities have higher WTP for  
362 increasing the frequency of waste collection and introducing textile sorting. This shows citizens'  
363 willingness to cooperate to increase the recycling rate at low-efficiency levels. Therefore, our results  
364 demonstrate that low-recycling rate municipalities should focus on improving frequency of waste  
365 collection. This strategy increases overall recycling efficiency and household satisfaction with the  
366 system.

367 Finally, we also found that population density is an important predictor of WTP for WMS attributes;  
368 individuals living in municipalities with low population density have higher preferences for a more  
369 frequent collection system. This is coherent with the literature, which evidences that individuals  
370 living in rural areas have higher nature-relatedness values (Bashan et al., 2021) and, thus, higher  
371 preferences for sustaining a cleaner environment near their households. In contrast, our results also  
372 report that the introduction of textile collection, an attribute which was not present in the Italian

373 WMS when the study was conducted, is more valued by citizens living in denser municipalities. As  
374 the introduction of textile collection in Italy would constitute an innovative WSM policy, our results  
375 are consistent with the evidence reporting that individuals living in more dense cities have a higher  
376 willingness to implement transformative and novel policies and ideas (Florida et al., 2017; Kunkel et  
377 al., 2022).

378 We also evaluate how socioeconomics impact WMS preferences by analysing the magnitude and  
379 direction of its indirect impact through its influence on environmental values, its direct impact, and  
380 other demographic-specific characteristics. We found that socioeconomics is a relevant predictor of  
381 WTP for WMS attributes, as previously analysed in the literature (Romano et al., 2022). Moreover,  
382 we found that its indirect impact through environmental values is low when compared to its direct  
383 one, which captures other demographic-specific traits influencing decisions. Further investigation is  
384 needed to uncover how socioeconomics impact WMS preference directly.

385 Finally, as environmental awareness values are a significant predictor of WTP estimates,  
386 improvements in WMS should be accompanied by public campaigns aimed at increasing awareness  
387 of its environmental benefits. Higher environmental awareness will lead to higher support of  
388 households for an improved WMS and, thus, increase overall wellbeing through a higher recycling  
389 rate and satisfaction with the system.

390 We have uncovered how local contexts, environmental values, and demographics influence  
391 preferences for WMS. Understanding the drivers of WMS preference heterogeneity is key for  
392 assessing the wellbeing and distributional impacts of any WMS policy. Therefore, our results are very  
393 relevant to effectively improve citizens' welfare and increase the WMS's efficiency.

## 394 **7. Conclusion**

395 Using Italy as a case study, we examined citizens' preferences for an improved WMS. Overall, we  
396 found that individuals positively value improvements in the WMS, including an increase in waste-

397 sorting categories, a more frequent and systematic waste collection, and the introduction of textile  
398 recycling. This shows citizens' enthusiasm to cooperate for a more efficient and sustainable WMS.  
399 We have also analysed how local contexts, environmental awareness values, and demographics shape  
400 individuals' preferences. We found that environmental awareness increases WTP for an improved  
401 WMS. Thus, environmental education campaigns may be an effective policy to increase WMS  
402 cooperation. In addition, we have uncovered how local contexts influence WMS preferences,  
403 highlighting the importance of context-specific WMS designs. Finally, we have found that  
404 socioeconomic variables have a significant impact on WMS preferences; they exert both an indirect  
405 impact, which emerges through the impact of demographics on environmental values, and a direct  
406 impact, which reflects other specific traits influencing decisions. Our results show that the direct  
407 impacts are generally higher than the indirect ones, pointing out that socioeconomics impact WMS  
408 preferences through specific traits that are not captured by environmental values. Future research  
409 could analyse the means by which demographics influence WTP for WMS.

410

#### 411 **CRedit** author statement

412 **Tatiana Cantillo:** Conceptualization, Methodology, Formal analysis, Investigation, Data curation,  
413 Writing - original draft, Writing - review & editing, Visualisation. **Sandra Notaro:**  
414 Conceptualization, Methodology, Formal analysis, Investigation, Data curation, Writing - original  
415 draft, Writing - review & editing, Visualisation, Supervision. **Nicolao Bonini:** Writing - review &  
416 editing, Funding acquisition. **Constantinos Hadjichristidis:** Writing - review & editing, Funding  
417 acquisition

418

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