



This is a repository copy of *Could CO₂-induced changes to C₄ grass flammability aggravate savanna woody encroachment?*.

White Rose Research Online URL for this paper:

<https://eprints.whiterose.ac.uk/181227/>

Version: Accepted Version

Article:

Raubenheimer, S.L., Simpson, K. orcid.org/0000-0001-6673-227X, Carkeek, R. et al. (1 more author) (2021) Could CO₂-induced changes to C₄ grass flammability aggravate savanna woody encroachment? *African Journal of Range & Forage Science*, 39 (1). pp. 82-95. ISSN 1022-0119

<https://doi.org/10.2989/10220119.2021.1986131>

This is an Accepted Manuscript of an article published by Taylor & Francis in *African Journal of Range & Forage Science* on 24th November 2021, available online:
<https://www.tandfonline.com/doi/abs/10.2989/10220119.2021.1986131>

Reuse

Items deposited in White Rose Research Online are protected by copyright, with all rights reserved unless indicated otherwise. They may be downloaded and/or printed for private study, or other acts as permitted by national copyright laws. The publisher or other rights holders may allow further reproduction and re-use of the full text version. This is indicated by the licence information on the White Rose Research Online record for the item.

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>



Could CO₂-induced changes to C₄ grass flammability aggravate savanna woody encroachment?

Journal:	<i>African Journal of Range & Forage Science</i>
Manuscript ID	TARF-2021-0079.R1
Manuscript Type:	Original Article-research notes, research papers
Keywords:	elevated CO ₂ , flammability, fire regimes, Climate Change, C ₄ grasses
Abstract:	<p>Fire plays an important role in maintaining the savanna tree-grass balance by limiting the recruitment of heat-sensitive tree seedlings. However, fire behaviour may change under increasing CO₂ concentrations, due to altered flammability of the grassy layer. Here, we determined the effect of predicted future CO₂ concentrations, and how it interacts with water-availability, on grass flammability and traits influential to flammability, and uncovered the physiological mechanisms underpinning these responses. Using the widespread C₄ savanna grass, <i>Themeda triandra</i>, as a model, we found that improved water-use efficiency under elevated CO₂ (800ppm) resulted in a larger (greater aboveground biomass) but wetter (higher moisture content) grass fuel load, that cured at a slower rate under drought conditions. These changes were associated with increased time to ignition, reduced flaming times and predicted rate of spread. We modelled the effect of altered grass flammability on fire behaviour at a national level (South Africa), finding large-scale reductions in fire spread under elevated CO₂, mitigating the converse effects of predicted increases in aridity, and marginal increases in fireline intensity. CO₂-induced reductions in fire frequency, spread or intensity could have serious implications for savanna vegetation dynamics, possibly exacerbating the woody encroachment already seen in these ecosystems across the world</p>

SCHOLARONE™
Manuscripts

1 **2. Introduction**

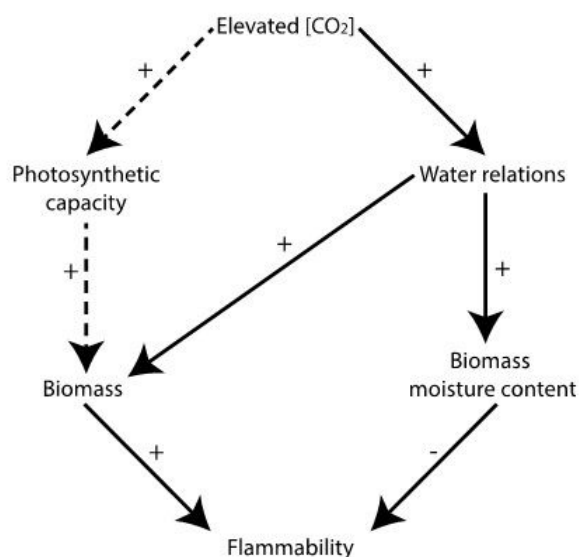
2 In tropical savannas, fire is a highly frequent disturbance that is essential in
3 maintaining the balance of the two life forms present – C₄ grasses and C₃ savanna
4 trees (van Langevelde et al., 2003; Bond et al., 2005; Sankaran et al., 2005). Regular
5 fire maintains an open canopy that allows shade-intolerant C₄ grasses to thrive by
6 limiting the recruitment and maturation of savanna trees (Bond & Midgley, 2001). The
7 grassy understory fuels fire in these ecosystems, and therefore, it is the flammability
8 of the grass community that determines fire behavior (Bond and van Wilgen, 1996;
9 Beckage et al., 2009; Prior et al., 2017). The key plant traits that determine savanna
10 grass flammability are the amount of aboveground biomass and its moisture content
11 (Simpson et al., 2016), as these dictate how much fuel is available and how easy it is
12 to ignite and spread fire. However, both of these traits may be altered under increasing
13 atmospheric CO₂ concentrations. Elevated CO₂ (eCO₂) can enhance C₄ grass
14 productivity, through stimulating photosynthetic capacity (Ward et al., 1999) or by
15 improving water relations in these seasonally-dry habitats (Quirk et al., 2019). Related
16 to the improved water relations is the maintenance of higher biomass moisture
17 contents into the dry season when savanna fire activity peaks. The effect of eCO₂ on
18 savanna grass flammability is largely unknown (Manea et al., 2015), but is crucial in
19 predicting how fire behavior may alter in the future, and thus the continued co-
20 existence of trees and grasses in these ecosystems.

21 Plant flammability is a key determinant of fire activity worldwide (Bond and van
22 Wilgen, 1996; Beckage et al., 2009; Prior et al., 2017). This is true for savannas where
23 the strong influence of grass flammability on fire behavior is evidenced by striking
24 examples of altered fire behavior alongside changes in the flammability of the grass
25 layer (D'Antonio, 2000; Platt and Gottschalk, 2001; Davies and Nafus, 2012,
26 Setterfield et al., 2013). Savanna grasses differ substantially in their flammability
27 (Simpson et al., 2016). This inter-specific variation means that even within a
28 landscape, spatial differences in grass community composition, and thus flammability,
29 can mediate fire behavior. For example, changes in the flammability of the grass layer
30 alone were responsible for fires becoming less intense on the approach to the interface
31 between savanna and tropical forest in Gabon (Cardoso et al., 2018). Alternatively,
32 having a similarly flammable grass layer can result in convergent fire regimes in
33 savannas with drastically different climatic conditions (e.g. Longleaf pine savanna in
34 the USA vs. Eucalypt savanna in Australia; Archibald et al., 2018).

35 The flammability of a plant depends on chemical and physical plant traits that
36 together contribute to biomass ignitability (the ability to ignite and spread fire),
37 sustainability (the time that combustion is maintained), and combustibility (burning
38 intensity; Rothermel, 1972; van Wilgen et al., 2010; Pausas et al., 2017).
39 Above-ground biomass drives combustibility and sustainability with high biomass
40 species burning more intensely and for longer and producing high predicted fire spread
41 rates (Simpson et al., 2016). Biomass moisture content determines how plant material
42 absorbs heat energy, with high moisture levels translating into delayed ignition, lower
43 combustion, and reduced rates of fire spread (Rothermel, 1972; Pyne, 1976). The
44 moisture in a plant acts as a sink for heat energy, reducing the heat energy available
45 for ignition (Chuvienco et al., 2004). For example, higher fuel moisture contents resulted
46 in reduced combustion and lower rates of fire spread in savanna fire trials in South
47 Africa (Govender et al., 2006). The contribution of these two grass traits to savanna
48 fire behavior varies across rainfall gradients, with low-rainfall savanna fires being
49 biomass (fuel) limited and high-rainfall savanna fires being limited by fuel moisture
50 (Alvarado et al., 2019).

51 Rising atmospheric CO₂ has the potential to alter savanna grass flammability
52 because it influences both the amount and the moisture content of biomass (Fig. 1).
53 eCO₂ can directly affect grass productivity through the stimulation of photosynthetic
54 capacity. C₄ grasses may show enhanced carbon assimilation under eCO₂ despite
55 their carbon concentrating mechanism, and therefore a greater biomass production,
56 although this response appears to be variable (Wand et al., 1999). Of greater
57 importance however, is the indirect impact of CO₂-induced improved water relations
58 on increased savanna grass biomass. eCO₂ reduces stomatal conductance (G_{st}) and
59 transpiration, while allowing photosynthetic capacity to be maintained (i.e. enhanced
60 water use efficiency (E), WUE; Atkin et al., 1999; Saralabai et al., 1997). High
61 temperatures and irradiance levels, combined with seasonal rainfall, means that
62 savanna grasses are susceptible to water limitation (Osborne and Sack, 2012).
63 Therefore, improved water relations under eCO₂ allows the effects of water limitation
64 to be delayed or reduced, and assimilation (and therefore growth) to be maintained for
65 longer (Reyes-Fox et al., 2014). As well as increasing biomass, this may also result in
66 higher fuel moisture contents, due to plant water potential being maintained higher for
67 longer into the dry season. Experiments that have investigated the effects of eCO₂ on
68 C₄ grasses have found evidence for both of these predicted changes. For example,

69 Quirk, Bellasio and colleagues (2019) found that the C₄ savanna grass, *Eragrostis*
 70 *curvula*, had more aboveground biomass that also had higher relative water contents
 71 when grown under eCO₂. How increases in biomass quantity and moisture content
 72 under eCO₂ will interact to drive changes in grass flammability is unclear due to their
 73 opposing impacts.



74
 75 **Figure 1.** Potential impacts of elevated atmospheric CO₂ concentrations ([CO₂]) on
 76 savanna grass traits and flammability. Dashed lines represent smaller or less certain
 77 effects.

78 CO₂-induced changes to grass flammability could have significant impacts on
 79 savanna community composition and structure, particularly if grass flammability is
 80 reduced. Savannas have experienced a significant reduction in fire activity over the
 81 past few decades (Andela et al., 2017;) **mostly due to increased fuel moisture levels**
 82 **(Zubkova et al., 2019)**, and this pattern could be further exacerbated through future
 83 reductions in grass flammability. Less frequent fire is both a cause and consequence
 84 of rapidly increasing tree densities in savannas: fewer fires allow more tree individuals
 85 to escape the 'fire trap' and establish (Smit et al., 2010), and more trees reduce fire
 86 activity by shading out the grass understory. This positive feedback may be intensified
 87 in the future by the CO₂-fertilisation of C₃ savanna trees **(Bond and Midgley, 2000;**
 88 **Kgope et al., 2010)**. On the other hand, eCO₂ could increase grass flammability and
 89 fire activity, and so limit woody plant establishment via increased seedling mortality.

90 The aim of this study is to determine how savanna grass flammability is altered
 91 under eCO₂. Here we used plants of a typical C₄ grass species that is widespread in

92 South African savannas, *Themeda triandra*, grown under ambient and predicted future
93 [CO₂], to test the impacts of eCO₂ on different flammability components and their
94 underlying plant traits. Various authors have predicted increases in fire activity as the
95 climate becomes warmer and drier (e.g. Flannigan et al., 2009; Fargeon et al., 2020;
96 Smith et al., 2020), but these not do consider the effects of CO₂ on vegetation. To
97 determine the role of water availability on an eCO₂ effect, plants were also subjected
98 to either a well-watered or a water-limited treatment before traits and flammability
99 measurements were taken. We also undertook a modelling exercise to see how
100 flammability changes may impact on country-wide fire behaviour.

101 3. Materials and methods

102 3.1. Study species

103 *Themeda triandra* Forssk. (Poaceae, Panicoideae, Andropogoneae; hereafter simply
104 referred to as *Themeda*) is one of the most widespread and ecologically dominant C₄
105 grass species in paleotropical savannas (Snyman et al., 2013; Dunning et al., 2017).
106 This perennial, tussock-forming grass has a close association with recurrent fire: it is
107 reliant on periodic burning to persist (Snyman et al., 2013) and is lost from a system
108 when both fire and grazing are excluded (Danckwerts, 1993). It recovers rapidly after
109 fire by resprouting from basal meristems (Bond et al., 2003; Morgan and Lunt, 1999)
110 and has been shown to plastically adjust traits relating to post-fire recovery in order to
111 maximise fitness under different fire regimes (Simpson et al., 2019).

112 In this study, *Themeda* plants were collected from a grassland site in the Eastern
113 Cape of South Africa (coordinates: -33.278, 26.489). To simulate a typical field
114 scenario in terms of biotic interactions and plant densities, three *Themeda* tussocks
115 (each consisting of 8 tillers) were planted along with one seedling of the common C₃
116 savanna trees, *Vachellia karroo* Hayne., in large 25L plastic pots (31x40cm). *V. karroo*
117 seedlings grown with the *Themeda* were all similar in size (2.5±0.4g). Only data from
118 the focal species, *Themeda*, is discussed in this manuscript. Soil was obtained from a
119 nearby savanna where both *Themeda* and *V. karroo* grow naturally (Fort Beaufort,
120 Eastern Cape) classified as Loam soil and considered to be representative of local
121 savanna soil nutrient and texture (19 / 8 / 70 % clay / silt / sand respectively).

122 3.2. Experimental conditions

123 The study plants were subjected to experimental conditions at the Rhodes
124 University Elevated CO₂ Facility (RUECF; Grahamstown, Eastern Cape) from mid-
125 growing season (December) until mid-dry season (June). This facility consists of 3m-
126 diameter open-top chambers (OTCs) with clear-plastic sides (for a full description of
127 the OTC construction; see Baso et al., 2021). CO₂ concentrations within 12 of the
128 OTCs was set at one of two levels: ambient (400ppm) or elevated (800ppm; reflecting
129 the Representative Concentration Pathway (RCP)-8.5 predictions for the turn of the
130 21st century; Pachauri et al., 2014). CO₂ concentrations were measured with open-
131 path CO₂ analysers every second (GMP343, Vaisala, Finland), and maintained using
132 a proportional–integral–derivative procedure to control injection of CO₂ (2873
133 proportional valve, Burkert, Germany). The OTCs were ventilated via a perforated,
134 circular diffuser supplied by a 3-phase fan, positioned at the plant canopy level. The
135 precision of the CO₂ treatment control within each OTC was good (>70% of
136 measurements within 1 standard deviation of the target; details given in Table S1).
137 Additional sensors were used to monitor solar energy (ES2, Delta-T Devices, Ltd.),
138 rainfall (RG2 Rain gauge sensor, Delta-T Devices, Ltd). Environmental data for the
139 experiment is summarized in Table S2. OTCs led to + 4°C increases in temperature,
140 and this was not affected by CO₂ treatment.

141 In addition to the CO₂ treatment (n=30 for both CO₂ concentrations; spread across
142 12 OTCs), plants were also subjected to one of two different water-availability
143 treatments to simulate savanna conditions in wet and dry seasons. Plants were either
144 well-watered (daily watering) or exposed to a series of four-week drought cycles,
145 separated by two-day watered recovery periods. Soil water content (5TM volumetric
146 water content sensors, Decagon Devices, Pullman, WA) and soil water potential
147 (MPS-6 soil water potential sensors, Decagon Devices) were monitored and recorded
148 every five seconds in a subset of pots from both treatments to ensure watering
149 treatments were conducted correctly (n=8 per treatment combination). Plants were
150 grown under these conditions for 24 weeks during the growing season (December to
151 June, ending at the beginning of the cold season). The first drought cycle was
152 implemented after 8 weeks of growth to increase establishment success. After 24
153 weeks, the mean soil water content was 24 ± 5.6 and 9 ± 6.2% at the time of harvest
154 for watered and drought treatments, respectively. Soil water content did not differ
155 significantly between CO₂ treatment and are thus grouped for the average values
156 presented.

157 3.2. Water and performance measurements

158 In order to determine a mechanism that could explain differences in plant traits
159 and flammability at the end of the experiment, the growth and performance of
160 *Themeda* plants were monitored during the treatments. Measures of photosynthetic
161 activity and water relations, including both AC_i curves and instantaneous gas
162 exchange measurements, were examined by individually enclosing leaves of a subset
163 of plants in the chamber of an infrared gas analyser (Li-Cor 6400, Li-Cor Biosciences,
164 Lincoln, NE, USA). Leaf temperature and leaf-to-air vapour pressure deficit were kept
165 below 1.5 kPa, block temperature was held constant at 30°C and photosynthetically
166 active radiation at 1500 mol m⁻² s⁻¹. Instantaneous measurements of carbon
167 assimilation rate (A), G_{st} , WUE (A/G_{st}), and E were conducted on a subset of
168 individuals ($n=8$) for all treatments. Leaf water potential (leaf ψ) was measured on a
169 clipped leaf after gas exchange measurements using a Schölander pressure chamber.
170 Gas exchange measures were taken at three time points through the experiment (8
171 weeks apart) each in the final week of a four-week drought cycle, with the final time
172 point including leaf water potential measures.

173 3.3. Plant traits influential to flammability

174 After 24 weeks of growth under the experimental conditions, all tussocks from
175 each pot were destructively harvested in order to measure traits influential to
176 flammability. The two key traits (aboveground biomass quantity and moisture content)
177 were measured on the harvested shoot biomass. Aboveground dry biomass (fuel load;
178 in g) was measured after oven-drying the harvested shoot biomass at 60°C to a
179 constant dry weight. Biomass moisture content (in g g⁻¹) was calculated by dividing the
180 difference between fresh and dry shoot biomass by the dry shoot biomass. Another
181 trait we thought would be influential to flammability and vegetation modelling (Scheiter
182 et al., 2012) is leaf curing rate. This indicates the speed at which dry, dead biomass
183 (that is highly flammable) accumulates. In order to measure leaf curing rate, and how
184 it is affected by CO₂, the length of the cured section of leaves was measured twice
185 weekly on the same six marked leaves per individual during the experiment, in both
186 aCO₂ and eCO₂ treatments. Length of cured leaf was deemed a suitable measure of
187 curing as we observed *Themeda* leaves to cure from the tip down rather than in a
188 series of leaves. This was only done on plants in the drought treatment, as those in

189 the watered treatment did not senesce. A curing rate (in mm day⁻¹) was determined
190 from a linear relationship fitted to the length of the cured leaf section over time.

191 Two other grass traits important to flammability, although to a lesser extent than
192 biomass quantity and moisture content, were also measured. The surface area to
193 volume ratio (SA:V) of grass leaves influences flaming duration, with high values
194 associated with low flaming duration (Simpson et al., 2016). Leaf SA:V was calculated
195 on five leaves per individual by dividing leaf surface area (determined by scanning the
196 leaves in a flat-bed scanner and using Image-J software to measure an area of pre-
197 defined colour) by leaf volume. The latter was defined by multiplying leaf area by leaf
198 thickness, which was measured using digital callipers (accurate to 0.01 mm) for five
199 leaves and averaged. Biomass density, a measure of fuel load per unit volume,
200 contributes to grass combustibility, such that plants with high biomass density combust
201 most rapidly (Simpson et al 2016; Gao and Schwilk, 2018; Wigley et al., 2021). After
202 wet biomass was recorded, biomass density (in g cm⁻¹) was measured as the vertical
203 biomass distribution for each plant. For this, the relationship between the logged
204 cumulative dry biomass and vertical height (vertical height intervals of 2, 10, 20 or 30
205 cm depending upon plant height) for each individual was determined. The biomass of
206 each grass clump was separated into the described height intervals and weighed after
207 oven-drying at 60°C. Linear models were fitted to the relationship between cumulative
208 dry biomass and vertical height. The slope of this relationship was used as a proxy for
209 biomass density, with higher values indicating more densely packed biomass.

210 3.3. Flammability measurements

211 Leaf-scale flammability was measured as the culmination of three components;
212 ignitability (time to ignition), sustainability (amount of time that combustion is
213 sustained), and combustibility (mass loss rate) (Anderson, 1970). Leaf-scale
214 measurements were used as they have been found to be relatively good proxies of
215 plant-scale flammability in savanna grasses (Simpson et al., 2016), despite these
216 scales of flammability traits being decoupled in other plant groups (Alam et al., 2020).
217 These three components were measured at the end of the experiment by exposing
218 leaf subsets of 0.2g (±0.001 g) cut into 2 cm segments to temperatures of ≈ 400°C
219 with a flame directed at a metal plate on which the leaves were placed. Flammability
220 tests were carried out in a fume cupboard with a constant vertical wind speed of 0.2 m
221 s⁻¹ to ensure consistent conditions for all samples (Simpson et al, 2016). All samples

222 ignited and burned. The temperature of the metal plate was monitored with a
223 thermocouple and was maintained at $400 \pm 30^{\circ}\text{C}$. Flaming tests were filmed at 25
224 frames s^{-1} and analysis of these videos allowed for the calculation of (i) Ignitability -
225 time to ignition (the time from application to first flaming), (ii) Sustainability - flaming
226 time (the time from ignition to extinction), and (iii) Combustibility (leaf mass divided by
227 flaming time).

228 The relationship between plant-scale combustibility and plant biomass (with plant
229 architecture intact) for *Themeda* has been previously established. Simpson et al
230 (2016) burned whole tussocks of *Themeda* (obtained from the same site as this study)
231 and quantified the strong linear relationship between maximum combustion rate and
232 aboveground biomass (maximum combustion rate = $0.163 + 0.656 \times \text{plant biomass}$;
233 $P < 0.001$, $n = 7$, $R^2 = 0.84$). This relationship is based on partially- and fully-cured grass
234 tussocks (as the relationship did not differ between them). We used this relationship
235 to estimate maximum combustion rate for the whole grass tussocks in this study.

236 The fire spread rate (ROS), representative of community-scale flammability, was
237 estimated from plant trait values from the experimental plants with Rothermel's (1972)
238 surface fire spread model. Models were implemented using the `ros()` function in the
239 Rothermel package (Vacchiano and Ascoli, 2014) in the R language and environment
240 for statistical computing (R Core Team, 2021). Fire behaviour was simulated for each
241 grass tussock through parameterization of the model with the following measured plant
242 traits according to the methods of Simpson et al. (2016): Leaf SA:V ($\text{m}^2 \text{m}^{-3}$), fuel
243 moisture content (%), fuel bed depth (cm – plant height), and fuel load (t ha^{-1} - total
244 biomass divided by cover area for each tussock). Uniform weather and topography
245 (site slope = 0%) were assumed for the simulation and all individuals were considered
246 to be one-hour fuels. Dead fuel moisture of extinction was set at 60% for all simulations
247 as this is the maximum possible value for one-hour fuels (Simpson et al., 2016). Mid-
248 flame wind speed was set at 3.6 km h^{-1} (mean wind speed in Grahamstown in July,
249 the peak of the fire season; Simpson et al., 2016). A value of 8.8 kJ g^{-1} was used for
250 fuel heat content (kJ kg^{-1} – effective heat of combustion) for all simulations based on
251 the mean measured value for *Themeda* by Simpson et al. (2016).

252 We also estimated a landscape-scale measure of fireline intensity for *Themeda*
253 based on traits under $a\text{CO}_2$ and $e\text{CO}_2$. Fire line intensity was predicted using the
254 methods of Byram (1959). Fire line intensity (*FI*; in $\text{KJ m}^{-2} \text{s}^{-1}$) was calculated as $I =$

255 Hwr; where H = heat yield of the fuel (kJ g^{-1} ; 8.8 kJ g^{-1} following Simpson et al., 2016),
256 w = weight of available fuel (g m^{-2}), and r = ROS (ms^{-1}).

257 3.4. Predicting future fire behaviour in South Africa

258 To estimate the effects of CO_2 -induced changes in grass flammability on fire
259 behaviour over a larger scale, we estimated ROS (Rothermel, 1972) and predicted FI
260 (Byram, 1959) under $a\text{CO}_2$ and $e\text{CO}_2$ for *Themeda* across South Africa for the month
261 of July (i.e. the month of peak fire activity) based on a continuous ground cover of the
262 widespread grass. This was done for South Africa due to the large portion of the
263 country consisting of grassland and savannah biomes and the distribution of *Themeda*
264 throughout the country. Although this is a simplistic approach, a general view on how
265 CO_2 -mediated changes to savanna grass flammability may change fire behaviour at a
266 larger spatial scale is valuable. ROS and FI were modelled under 4 scenarios: (A)
267 under current ambient conditions, (B) under RCP8.5 predictions of soil moisture
268 changes, (C) under RCP8.5 predictions of soil moisture changes with the effects of
269 $e\text{CO}_2$ on plant physiology and flammability incorporated, and (D) as in (C) but with the
270 addition of an adjusted soil moisture scenario calculated according to measured
271 experimental soil water savings associated with $e\text{CO}_2$. The countrywide version of the
272 models were parameterised as done in 3.3 but with values for slope, wind speed and
273 fuel moisture content varying spatially. A gridded dataset at a resolution of 5km was
274 constructed for these variables. Slope values were calculated from the altitude values
275 in the World Slope Model dataset using the terrain() function of package 'raster'
276 (Hijmans et al., 2019). Gridded mean annual wind speed data was used from the
277 Global Wind Atlas version 3.0 (Badger et al., 2019). Fuel moisture content in each grid
278 cell was calculated for each CO_2 scenario (400 ppm or 800 ppm) using soil water
279 content values based on either ambient or RCP8.5 projection scenarios for 2070
280 (RCP8.5 CCSM4, CM5A, CNRM-CM5, MIROC-ESM, and MPI-ESM-LR projections).
281 The RCP8.5 CCSM4 projection is presented in the results and alternative projections
282 are presented in the supplementary materials. Soil water content was calculated from
283 measures of precipitation (CHELSA database, Karger et al., 2017), wilting points and
284 field capacity (IGBP Global Soil Data Task, 2000) according to the methods described
285 by Trabucco and Zomer, 2010). Soil moisture content was then converted into plant
286 moisture content using linear relationships developed during the experiment between
287 measured soil water content (%) and leaf moisture content ($R^2 = 0.62$ and 0.46 for 400

288 ppm and 800 ppm treatments, respectively; see Figure S.1). Adjusted soil water
289 scenario used in (D) were adjusted according to experimentally determined soil water
290 savings ($24.08 \pm 4.26\%$ soil water saving). This soil water saving was calculated from
291 the change in measured soil water content over a 22 day period during the final 4 week
292 drought cycle (8 soil sensors per treatment). Soil water savings were equal to the
293 difference between the mean water loss at aCO_2 and at eCO_2 , divided by that at
294 aCO_2 . The % difference was then used to add 24.08% of the value in each grid cell
295 to the cell value. These gridded datasets were merged to the same extent and
296 resolution using package 'raster' (Hijmans et al., 2019), before models were run for
297 each grid cell for both aCO_2 and eCO_2 .

298 3.5. Statistical analysis

299 The data obtained were analyzed using the R language and platform (R Core Team
300 2020). To determine the influence of the treatments (both CO_2 and water-availability)
301 on *Themeda* traits and flammability, linear mixed-effects models were fitted using the
302 nlme R package (Pinheiro et al., 2021). The fixed effects were the CO_2 treatment
303 (either aCO_2 or eCO_2) and water treatment (watered or drought) and the OTC
304 identifying number (1-12) was included as a random effect (to account for any
305 differences arising from the position or functioning of the different OTCs). Statistics
306 presented in the results show the effect of CO_2 treatment on individual groups.
307 Principal component analyses (PCAs) were conducted using the FactoMineR R
308 package (Lê et al., 2008) to display overall trait differences between plants grown at
309 aCO_2 and eCO_2 under the two water-availability treatments. Separate PCAs were run
310 for plant traits (leaf curing rate, biomass moisture, and biomass quantity) and
311 flammability traits (time to ignition, sustainability, maximum combustion rate, ROS, and
312 FI).

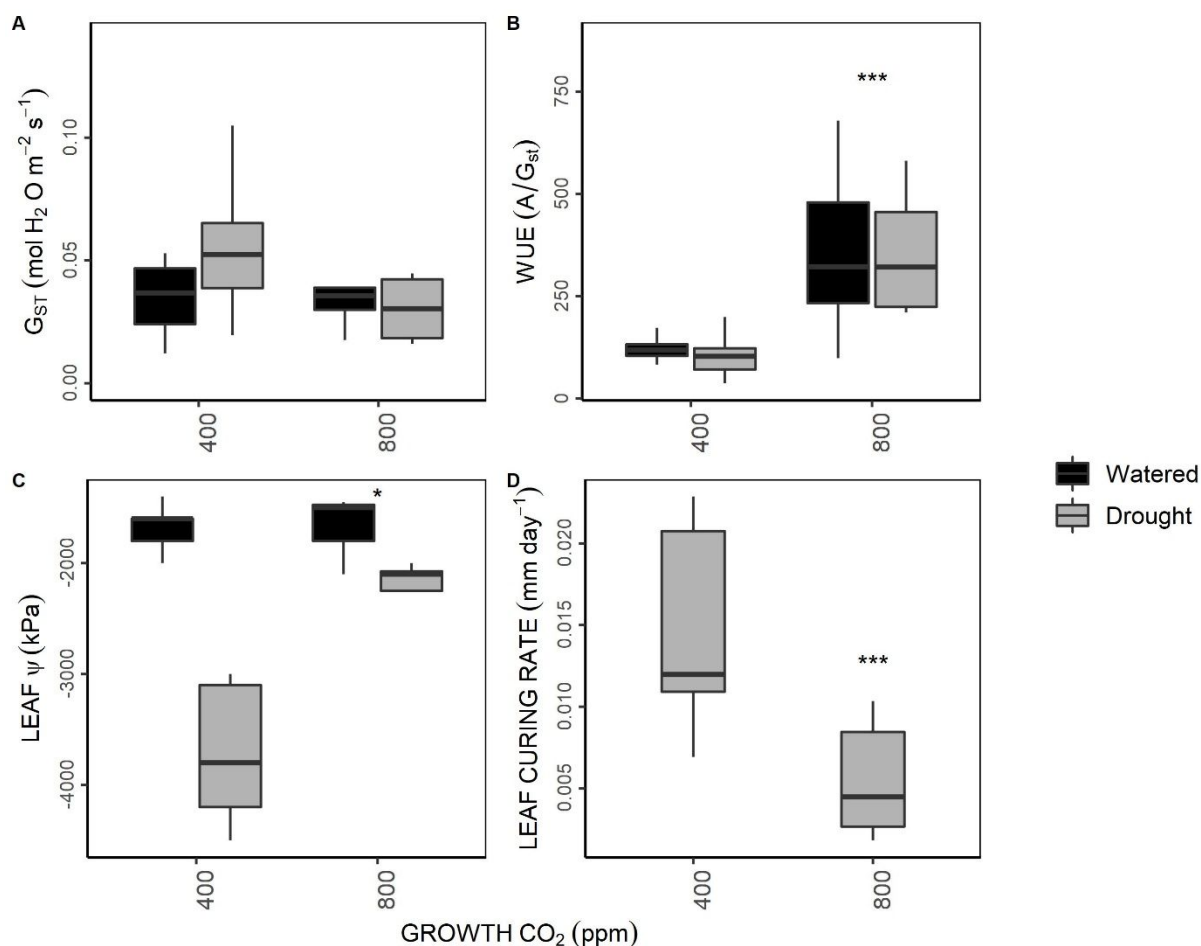
313 4. Results

314 4.1. Plant performance under eCO_2

315 G_{st} was similar in the two watering treatments at aCO_2 , but under drought conditions,
316 eCO_2 enhanced water savings and caused an average 30% reduction in G_{st} , although
317 this effect was not significant ($T(6)=-0.7$, $p=0.53$). The lower G_{st} in combination with
318 generally but non-significant higher photosynthetic rates (on average 52% higher;
319 $T(8)=-1.04$, $p=0.33$) led to significant increases in WUE (Fig 2B; $T(8)=4.71$, $p<0.01$)

320 and less negative leaf ψ (Fig 2C; T(6)=1.81, $p < 0.01$) in individuals exposed to drought.
 321 A knock-on effect of improved water relations under eCO₂ was a significant 58%
 322 reduction in leaf curing rates in grasses exposed to water limitation (Fig 2D; T(8)=
 323 5.57, $p < 0.001$).

324



325

326 **Figure 2.** Stomatal conductance (G_{st}) [A], water use efficiency (WUE) [B], leaf water
 327 potential ($\text{leaf } \psi$) [C], and leaf curing rate [D] of individual *Themeda triandra* tussocks
 328 exposed to different CO₂ (400 or 800 ppm), water (watered or drought) treatments.
 329 Mean soil water content was 24 ± 5.6 and $9 \pm 6.2\%$ at the time the plants were
 330 harvested under the watered and drought treatments, respectively. Box plots show
 331 means, interquartile ranges and minimum/maximum values. * represents a significant
 332 effect of eCO₂ ($p < 0.1$; *, $p < 0.05$; ***, $p < 0.001$).

333 4.2. Plant flammability under eCO₂

334 Differential plant productivity and water relations under eCO₂ resulted in
 335 changes in traits influential to flammability. *Themeda* plants grown at eCO₂ had an

336 average 14% higher aboveground biomass in watered conditions ($T(8)=0.63$, 0.54),
 337 and 38% higher under drought conditions ($T(8)=2.48$, $p<0.05$; Table 2). **eCO₂**
 338 **completely negated the adverse effects of drought to allow for similar levels of**
 339 **productivity when water was limited.** eCO₂ had a significant effect on biomass moisture
 340 content in the drought treatment when soil moisture contents was low (<10%; Figure
 341 S.1). When soil moisture content is <10%, biomass moisture content is significantly
 342 higher under eCO₂ (33% under aCO₂ vs 47% eCO₂; $T(8)=4.49$, $p<0.01$). Biomass
 343 density and leaf SA:V were unaffected by CO₂ or water treatments (Table 2).

344 CO₂-induced changes in flammability traits had cascading effects on multiple
 345 aspects of **flammability**. At the leaf level, higher moisture contents under eCO₂ resulted
 346 in lower ignitability (a delay in ignition), with time to ignition increasing by 20%
 347 ($T(8)=1.32$, $p<0.05$) and 8% ($T(8)=0.98$, $p=0.28$) under well-watered and drought
 348 treatments respectively (Table 1). The length of time that flaming was sustained was
 349 generally, but not significantly, lower in grasses grown at eCO₂ ($T(8)=-1.71$, $p=0.09$
 350 and $T(8)=-1.42$, $p=0.16$ in watered and drought treatments, respectively). Estimated
 351 maximum plant scale combustion rate was significantly higher with eCO₂ under
 352 drought conditions ($T(8)=2.61$, $p<0.05$).

353 **When predicting fire behaviour**, higher grass biomass moisture content
 354 overrode the effect of increased aboveground biomass at eCO₂ on fire behaviour, to
 355 generate lower predicted rates of fire spread and fire intensity (Table 1). Fire spread
 356 models parameterised with *Themeda* traits under the different treatments showed a
 357 31% decline in predicted fire spread rate under well-watered conditions ($T(8)=-1.76$,
 358 $p=0.43$), and a 13% decline ($T(8)=-2.75$, $p<0.01$) under water-limitation. Similarly,
 359 there was a 12% decline in predicted fire line intensity under well-watered conditions
 360 ($T(8)=-0.74$, $p=0.46$), and a 4% decline under water-limitation ($T(8)=-0.28$, $p=0.78$).

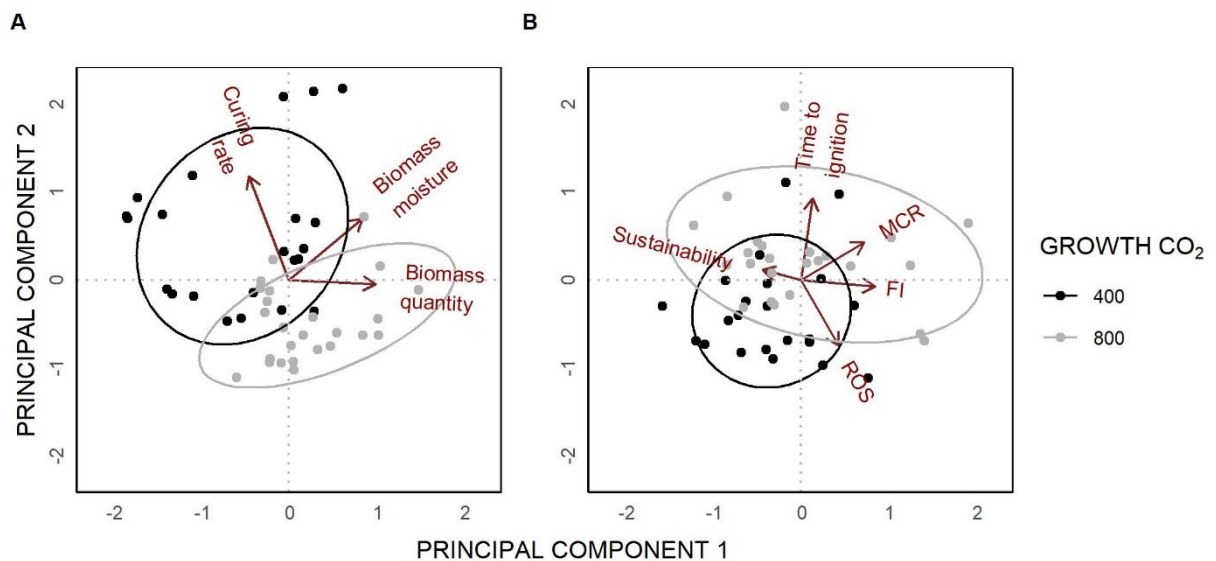
361 **Table 1.** Fuel load per tussock, biomass moisture content, biomass density,
 362 leaf surface area:volume ratio, time to ignition, sustainability, combustibility, estimated
 363 maximum combustion rate, predicted rate of fire spread (*ROS*), and predicted fireline
 364 intensity (*FI*) of individual *Themeda* tussocks exposed to different CO₂ (400 or 800
 365 ppm), and water (watered or drought) treatments (n=30 for each). Showing mean ±
 366 S.E. * represents a significant effect of eCO₂ ($p < 0.1$; *, $p < 0.05$; ***, $p < 0.001$).

Parameter	CO ₂ (ppm)	Watered	Drought
-----------	--------------------------	---------	---------

2.1. Plant traits			
	400	15.37 ± 9.87	10.54 ± 5.1
Fuel load per tussock (g)	800	17.48 ± 6.55	16.98 ± 8.63 *
	400	54.27 ± 24.01	50.15 ± 18.59
Biomass moisture content (%)	800	56.28 ± 19.7	52.6 ± 19.15
	400	-	32.53 ± 5.48
Biomass moisture content when soil moisture content <10% (%)	800	-	47.0 ± 9.22 **
	400	1.64 ± 1.27	1.36 ± 0.74
Biomass density (g cm⁻²)	800	1.90 ± 0.83	1.70 ± 0.97
	400	88.22 ± 34.97	89.85 ± 33.24
Leaf SA:V ratio	800	84.14 ± 18.37	88.14 ± 31.70
2.2. Flammability measures			
	400	2.47 ± 0.16	2.74 ± 1.43
Time to ignition (s)	800	3 ± 0.17 *	2.98 ± 1.16
	400	65.47 ± 25.65	62.62 ± 28.05
Sustainability (s)	800	58.41 ± 24.71	57.31 ± 21.39
	400	0.0036 ± 0.0014	0.0040 ± 0.0018
Combustibility (g s⁻¹)	800	0.0042 ± 0.002	0.0041 ± 0.0016
	400	10.33 ± 6.57	7.08 ± 3.35
Estimated maximum combustion rate (gs⁻¹)	800	11.81 ± 4.35	11.30 ± 5.66 *
2.3. Fire behaviour			
	400	2.44 ± 0.82	2.13 ± 0.78
ROS (m min⁻¹)	800	1.86 ± 0.93 *	1.88 ± 0.86 *
	400	77.9 ± 56.73	53.26 ± 32.76
FI (KJ m⁻² s⁻¹)	800	68.47 ± 50.47	51.18 ± 37.94

367 A principle components analysis (PCA) was used to examine the variance in traits
 368 associated with flammability according to CO₂ treatment. We discuss the PCA based
 369 on the droughted individuals here (so to include leaf curing rate in the analysis, which
 370 was only measured on droughted individuals); the PCA based on the well-watered
 371 individuals can be found in the supplementary information (Fig S.2). The first two
 372 principal components accounted for 88.35% (Fig. 3A) and 63.75% (Fig. 3B) of the total
 373 variance respectively. The PCA shows plant traits related to flammability to separate

374 out according to CO₂ treatment (Fig. 3A). Knock-on effects of lower leaf curing rate
 375 were seen in the higher time to ignition and lower ROS at eCO₂ (Fig. 3B). The higher
 376 biomass quantity led to higher estimates of maximum combustion rate (MCR) and FI
 377 with eCO₂ (Fig. 3B). Table 2 shows the contribution of CO₂ treatment to leaf-scale
 378 flammability parameters used in Figure 3. CO₂ treatment significantly predicted leaf
 379 curing rate (T(8)=-5.57, p<0.001), biomass quantity (T(8)=2.61, p<0.05), estimated
 380 MCR (T(8)=2.61, p<0.05), and ROS (T(8)=-1.89, p<0.05).



381 **Figure 3.** Principal components analysis of C₄ grass (*Themeda triandra*) traits
 382 associated with flammability: [A] leaf curing rate, biomass moisture, and biomass
 383 quantity. [B] sustainability, ignitability, maximum combustion rate (MCR), predicted
 384 rate of spread (ROS), and predicted fireline intensity (FI). All plants used in the PCA
 385 were exposed to drought.
 386

387 **Table 2.** The contribution of CO₂ treatment to leaf-scale flammability parameters of C₄
 388 grasses exposed to drought used in the principle component analysis in Fig. 3A (Leaf
 389 curing rate, biomass moisture, biomass quantity) and Fig. 3B (Time to ignition,
 390 sustainability, estimated maximum rate of combustion MCR, predicted rate of spread
 391 ROS, and predicted fireline intensity FI) as determined by linear mixed-effects models.
 392 Estimate values represent estimates of the slope, showing standard error (SE) and p
 393 values. * represents a significant effect of eCO₂ (p < 0.1; *, p < 0.05; ***, p < 0.001).

Model	Parameter	Effect of CO ₂ treatment		
		Estimate	SE	P value

A. Plant traits	Leaf curing rate	-0.009	0.003	<0.001 **
	Biomass moisture	4.63	6.18	0.5
	Biomass quantity	6.45	2.47	<0.05 *
B. Flammability	Time to ignition	0.96	0.45	0.09
	Sustainability	-12.36	17.35	0.52
	<i>MCR</i>	4.24	1.63	<0.05 *
	<i>ROS</i>	-0.43	0.23	<0.05 *
	<i>FI</i>	24.26	13.96	0.15

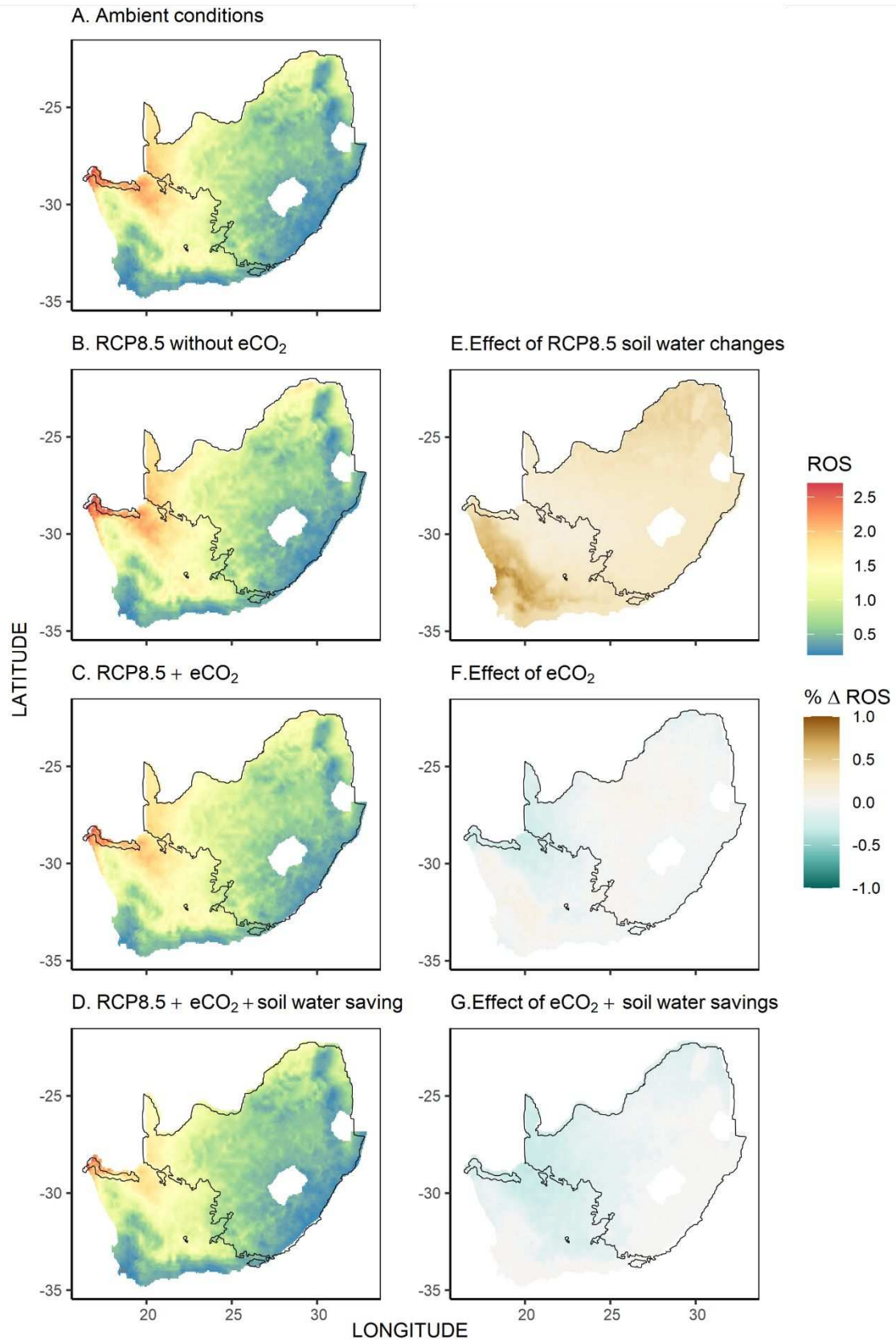
394

395 **4.3. Changes in fire behaviour under eCO₂ across South Africa**

396 We predicted *ROS* (Fig. 4) and *FI* (Fig. 5) across South Africa for the month of July
397 for a variety of climatic projection scenarios, based on a continuous ground cover of
398 the widespread grass *Themeda*. Future soil moisture scenarios in Fig.4 and 5 are
399 according to the RCP8.5 CNRM-CM5 projection. Alternative projections are presented
400 in Fig. S3-10. We acknowledge that one species will not represent community level
401 flammability, but instead this approach provides a basic estimation of a dominant
402 species. When *ROS* is modelled for *Themeda* across South Africa at ambient
403 environmental conditions, *ROS* is predicted to be higher in the more arid zones in the
404 north-west of the country and lower along the coastal regions (Fig. 4A). Similar
405 predictions are seen for *FI* (Fig. 5A). Accounting for predicted changes in soil water
406 content by the year 2070 but excluding the effects of eCO₂ on the physiology and
407 flammability of the grasses, our models predict *ROS* to increase in general across the
408 country, with the most notable changes being seen in the western regions (Fig. 4B).
409 *FI* changes are predicted to be more moderate than those of *ROS* and show increases
410 only in the western regions of South Africa (Fig. 5B). Incorporating the effects of eCO₂
411 on the physiology and flammability of the grasses, we predict *ROS* to decrease
412 relevant to ambient conditions (Fig. 4C) and *FI* (Fig. 5C) to increase slightly in certain
413 areas along the west coast and central interior and to moderate the effects of
414 reductions in soil water availability on *ROS* across much of the country. Further
415 accounting for the effects of eCO₂, we predict both *ROS* and *FI* to decrease under an
416 adjusted soil water scenario where soil water contents are adjusted according to
417 experimentally measured soil water savings associated with eCO₂ (Fig. 4D and 5D).

418 We acknowledge that soil water savings would not be uniform across soil types but
419 show that CO₂ associated soil water savings should be considered in predictions of
420 future fire patterns.
421

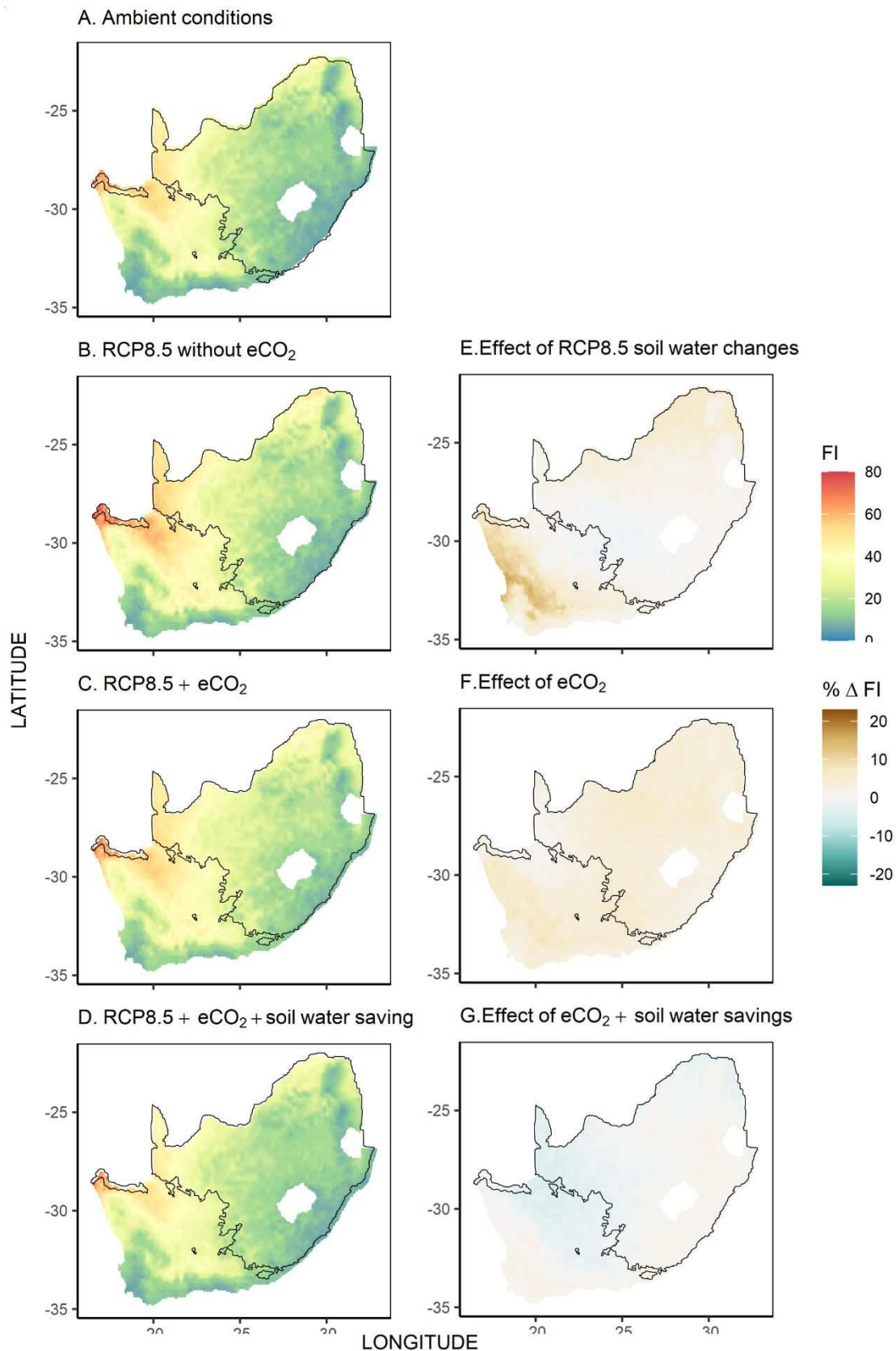
For Peer Review Only



422

423 **Figure 4.** Rate of fire spread (ROS; m min⁻¹) (Rothermel, 1972) is predicted for
 424 *Themeda triandra* grasses under ambient conditions [A], under predicted future soil
 425 moisture conditions (for the year 2070 according to the RCP 8.5 CNRM-CM5

426 projection scenario) without the effects of eCO₂ on plant physiology and flammability
427 [B], under predicted future soil moisture conditions with the effects of eCO₂ on plant
428 physiology and flammability incorporated, and under predicted future soil moisture
429 conditions adjusted according to experimentally determined soil water savings (24.08
430 ± 4.26% lower soil water loss) associated with eCO₂ and with the effects of eCO₂ on
431 plant physiology and flammability incorporated [D]. Predictions are based on the
432 flammability of the widespread savanna grass, *Themeda triandra*, grown under aCO₂
433 (400ppm) and eCO₂ (800ppm). The area outlined in black represents that currently
434 classified as savanna or grassland biomes (South African National Biodiversity
435 Institute, 2012). Plots using alternative projection scenarios are presented in Fig. S.3-
436 S.6.
437



438

439 **Figure 5.** Fireline intensity (FI ; $\text{KJ m}^{-2} \text{s}^{-1}$) (based on Byram's FI ; Byram 1959) is
 440 predicted for *Themeda triandra* grasses under ambient conditions [A], under predicted
 441 future soil moisture conditions (for the year 2070 according to the RCP 8.5 CNRM-

442 CM5 projection scenario) without the effects of eCO₂ on plant physiology and
443 flammability [B], under predicted future soil moisture conditions with the effects of
444 eCO₂ on plant physiology and flammability incorporated, and under predicted future
445 soil moisture conditions adjusted according to experimentally determined soil water
446 savings (24.08 ± 4.26% lower soil water loss) associated with eCO₂ and with the
447 effects of eCO₂ on plant physiology and flammability incorporated [D]. Predictions are
448 based only on the flammability of the widespread savanna grass, *Themeda triandra*,
449 grown under aCO₂ (400ppm) and eCO₂ (800ppm). The area outlined in black
450 represents that currently classified as savanna or grassland biomes (South African
451 National Biodiversity Institute, 2012). Plots using alternative projection scenarios are
452 presented in Fig. S.7-S.10.

453

454 4. Discussion

455 Using the widespread, ecologically-dominant, C₄ savanna, *Themeda triandra*, as a
456 model species, we determined the effect of predicted future CO₂ concentrations on
457 flammability-influential traits and uncovered the physiological mechanisms
458 underpinning these responses. By investigating the effects of water availability
459 alongside CO₂ enrichment, we find a notable impact of eCO₂ on grass biomass
460 quantity, moisture content and curing rates, with knock-on effects on flammability.
461 Grasses took longer to ignite, sustained flaming for shorter times and spread fire at a
462 lower rate on average when grown under eCO₂. We also modelled how these
463 flammability effects might influence fire behaviour in South African grass-dominated
464 ecosystems, and found much of this area would experience lower fire spread rates
465 under eCO₂, but that future fireline intensity trends would be variable. These results
466 suggest that future declines in grass flammability, and therefore fire activity, may
467 benefit savanna trees at the expense of grasses, and contribute to the already-
468 increasing trend of tree densities in savannas (Stevens et al., 2017).

469 Grass biomass was elevated under eCO₂, as has been found elsewhere for
470 *Themeda*, and other C₄ savanna grass species (Wand et al., 1999 and references
471 therein; Manea and Leishman 2019; Quirk et al., 2019). The mechanism responsible
472 for this enhanced productivity is primarily due to improved water relations under eCO₂,
473 as demonstrated by the larger biomass response in the drought treatment. We also
474 found that the photosynthetic capacity of *Themeda* was improved at 800ppm CO₂,

475 suggesting that the photosynthesis of this species is not carbon-saturated under $a\text{CO}_2$.
476 This adds to a growing body of research that C_4 grasses can respond to $e\text{CO}_2$ through
477 improved CO_2 assimilation rates (Wand et al., 1999). Better water relations under
478 $e\text{CO}_2$ also had the effect of increasing leaf moisture content - *Themeda* plants were
479 able to maintain higher biomass moisture contents across the range of soil moistures
480 under $e\text{CO}_2$ (Fig. S1). This allowed plants to remain physiologically-active and
481 productive for longer into dry-down, and delay the onset of leaf senescence.

482 As fire tends to show threshold behaviour, responding to biological and
483 geophysical factors in non-linear ways (Archibald et al., 2017), even small changes in
484 the flammability of a system can have large implications on fire characteristics (Cox
485 and Durrett, 1988; Archibald et al., 2012). Therefore the modest CO_2 -induced changes
486 we see in grass flammability here, could result in considerable alterations to grass-
487 fuelled fire regimes. The larger but wetter fuel load produced by *Themeda* plants under
488 $e\text{CO}_2$ (particularly under water limitation) took longer to ignite, and burned at a faster
489 rate once alight for a shorter period of time. In addition, $e\text{CO}_2$ -grown plants also took
490 longer to senesce, with grasses staying greener for longer under dry conditions. The
491 only other paper that has investigated grass flammability under $e\text{CO}_2$, Manea et al.,
492 (2015), similarly found increased fuel loads in woodland understory grasses.
493 Conversely, that study surprisingly found lower leaf moisture content under $e\text{CO}_2$,
494 which the authors could not explain. A potential impact of the flammability changes
495 found here is a reduction in fire occurrence under future CO_2 concentrations. If $e\text{CO}_2$
496 increases moisture biomass retention, more heat energy is required for ignition and
497 fire spread (Trollope, 1978; Gill and Moore, 1996; Alessio et al., 2008; Plucinski and
498 Anderson, 2008), resulting in fewer successful ignitions and smaller fires. As natural
499 fire regimes largely coincide with the dry season in South African grass-dominated
500 ecosystems, the slower curing rate of *Themeda* under $e\text{CO}_2$ may result in a delay in
501 plants being sufficiently highly flammable to ignite and spread fire, and therefore the
502 start of the fire season. Once alight however, the larger average *Themeda* biomass
503 under $e\text{CO}_2$ resulted in higher estimated maximum combustion rates, although only
504 significantly under water-limited conditions. This suggests that despite the higher
505 moisture content of biomass under $e\text{CO}_2$, it should still combust equally or better than
506 under ambient CO_2 levels, and that fire will continue to typically remove all
507 aboveground grass biomass.

508 Future fire behaviour in grass-dominated systems is also determined by the ability
509 of grasses to resprout after fire, and therefore recover aboveground biomass quickly
510 to fuel subsequent fires, and the decomposition rate of grass biomass. Previous
511 studies have found that the resprouting abilities of tropical grasses, including
512 *Themeda*, are unaffected by eCO₂ (Kgope et al., 2010; Manea and Leishman., 2019),
513 however grass decomposition may decline (Manea et al., 2015). Therefore, grass-
514 driven changes in fire behaviour under eCO₂ are likely to result from changes in grass
515 flammability alone, and not from changes in grass resprouting ability.

516 We predicted landscape-scale fire behaviour across South African grass-
517 dominated ecosystems based on *Themeda* trait data grown under different CO₂
518 treatments and on predicted changes in soil water availability. This simplistic approach
519 resulted in the prediction that the rate of fire spread will increase over much of the
520 country's grasslands and savannas as soil water availability declines, but that the
521 effects of eCO₂ on plant physiology and flammability in *Themeda* will mitigate this
522 effect (and even reverse this trend in some areas). Greater increases in biomass
523 moisture content under these conditions, due to enhanced water use efficiency, drives
524 these patterns, and suggests that eCO₂ may cause fire in low-rainfall savannas to be
525 increasingly fuel-moisture limited (Alvarado et al., 2019). We took our predictions of
526 the effects of eCO₂ on fire behaviour one step further by adjusting soil moisture
527 scenarios by mean soil water savings measured during our experiments. Soil water
528 savings led to further declines in predicted rate of spread. We predict fireline intensity
529 to increase both with predicted declines in soil moisture and with the slightly
530 heightened biomass accumulation associated with eCO₂. This effect was, however,
531 reduced by eCO₂-associated soil water savings. These fire behavior models only use
532 one fuel – *Themeda* – and therefore do not account for variation in fuels, which would
533 include many herbaceous and woody plant species. Another caveat to this approach
534 is that differences in soil fertility are not accounted for, which can mediate plant CO₂-
535 responses (Reich et al., 2014). However, community-level flammability tends to be
536 driven by the most flammable species present (van Altena et al., 2012; de Magalhães
537 and Schwilk, 2012; Wyse et al., 2018), and grassy ecosystems are typically dominated
538 by a few key species (Edwards et al., 2010). By using a highly flammable and widely-
539 dominant species here, we can gain some insight into future changes in grass-fuelled
540 fire behaviour. We only consider CO₂-induced changes to fire behaviour in South
541 Africa here. However, when the extent of grass fuelled fires is considered (>80% of

542 annual burnt area is grassy; van der Werf et al., 2006), changes in the flammability of
543 grass communities could have considerable impacts upon fire regimes at a global
544 scale. The simplistic approach of our models mean they are not suitable for strict
545 prediction of fire behaviour but do shed light on the short falls of current fire modeling
546 efforts and highlight the importance of CO₂ effects on both plant physiology and water
547 relations as well as associated soil water savings.

548 Changes in fire regime, caused by increasing CO₂ concentrations, may directly
549 impact on the grass communities that fuel fire. Fire regimes may act as a selection
550 pressure or an ecological filter for certain traits associated with persistence in fire-
551 prone landscapes (Keeley et al., 2011). For grasses, traits related to survival through
552 fire and recovery afterwards, via resprouting or from seed, are linked to fire frequency
553 (Forrestel et al., 2014). For example, less frequent fire is associated with more
554 resprouting grass species and fewer seeder species (Simpson et al., 2021). Therefore,
555 changes in fire frequency may result in altered grass communities in terms of the traits
556 present. Similarly, grass lineages are strongly sorted along fire-frequency gradients,
557 with some lineages having a low tolerance of fire and others being reliant on fire for
558 survival (Uys et al., 2004; Visser et al. 2012; Lehmann et al., 2019). However, savanna
559 grasses, including *Themeda*, are plastic in their fire-related traits (Simpson et al.,
560 2019), and so may be able to tolerate fire-frequency changes and prevent associated
561 alterations in grass community composition. Fire intensity appears to be less important
562 than fire frequency in influencing grass traits (Uys et al., 2004; Peláez et al., 2013;
563 Simpson et al., 2021), and so CO₂-induced changes in fire intensity are less likely to
564 impact on grass communities in terms of the traits and species present.

565 Indirect effects of altered fire regime under eCO₂, mediated by changes in
566 relative competitive ability of savanna trees and grasses, are much more likely to alter
567 grass communities than direct effects. Studies in African savannas show that fire
568 reduces woody biomass (Trapnell, 1959; van Wyk, 1971; Bond *et al.*, 2005), or at least
569 alters woody pant demographics, by causing either top-kill or mortality to individuals
570 in vulnerable, small size classes (Higgins et al., 2007). Therefore, reductions in fire
571 frequency lowers mortality in fire-sensitive tree species and allows more individuals to
572 escape the 'fire trap' and reach maturity (Bond and Midgley, 2001; Smit et al., 2010).
573 Increases in savanna tree abundance would decrease C₄ grass biomass due to
574 heightened competition and shading (Angassa, 2005; Ratajczak et al., 2012), and
575 therefore decreases fuel for fire. This positive feedback loop can ultimately lead to fire

576 becoming rare or absent, once a certain tree-cover threshold is reached and the grass
577 layer is too small and discontinuous to spread fire. In southern Africa, this threshold is
578 ~40% tree cover, above which, burnt area declines rapidly (Archibald et al., 2009).
579 Therefore eCO₂-induced changes to grass flammability, and fire behaviour, may
580 contribute to altered savanna dynamics that threaten the existence of this biome.

581 **6. Conclusions**

582 **This research sheds light on an understudied aspect of fire in a changing climate.**
583 Future changes in atmospheric CO₂ are predicted to alter grass flammability with
584 serious implications for savanna vegetation dynamics. Improved water relations and
585 productivity together resulted in a larger but wetter grass fuel load that took longer to
586 ignite and spread fire at a lower rate. **The relative influence of fuel moisture versus**
587 **biomass appears to vary spatially, but with the effects of moisture changes being more**
588 **important in general. We suggest that fire will be less frequent but possibly more**
589 **intense at future CO₂ levels, although predicted changes in fireline intensity are**
590 **marginal. We highlight the importance of including the effects of eCO₂ on both plant**
591 **physiology and associated soil moisture savings in fire modelling.** At a large spatial
592 scale, reductions in fire spread may result in smaller fires and reduced fire frequency
593 in grass-dominated systems. Therefore eCO₂-induced changes in grass flammability
594 and fire frequency will likely exacerbate the trend in woody thickening of savannas into
595 the future.

596

597 **References**

- 598 Alam MA, Wyse SV, Buckley HL, et al. 2020. Shoot flammability is decoupled from
599 leaf flammability, but controlled by leaf functional traits. *Journal of Ecology* 108: 641–
600 653.
- 601 Albini F. 1978. Estimating wildfire behaviour and effects. Department of Agriculture,
602 Forest Service, *Intermountain Forest and Range Experiment Station. USDA For Serv*
603 *Gen Tech Rep* 30.
- 604 Alessio GA, Peñuelas J, Llusà J, Ogaya R, Estiarte M, De Lillis M. 2008. Influence of
605 water and terpenes on flammability in some dominant Mediterranean species.
606 *International Journal of Wildfire* 17(2): 274–86.

- 607 Alvarado ST, Andela N, Silva TSF, Archibald S. 2020. Thresholds of fire response to
608 moisture and fuel load differ between tropical savannas and grasslands across
609 continents. *Global Ecology and Biogeography*. 29(2): 331–44.
- 610 Andela N, Morton DC, Giglio L, Chen Y, van der Werf GR, Kasibhatla PS, DeFries RS,
611 Collatz GJ, Hantson S, Kloster S, Bachelet D, Forrest M, Lasslop G, Li F, Mangeon S,
612 Melton JR, Yue C, Randerson JT. 2017. A human-driven decline in global burned area.
613 *Science*, 356, 6345: 1356-1362.
- 614 Angassa A. 2005. The ecological impact of bush encroachment on the yield of grasses
615 in Borana rangeland ecosystem. *African Journal of Ecology*, 43: 14-20.
- 616 Archibald S, Lehmann CER, Belcher CM, Bond WJ, Bradstock RA, Daniau A-L,
617 Dexter K, Forrester EJ, Greve M, He T, Higgins SI, Hoffman WA, Lamont BB,
618 McGlenn DJ, Moncrieff GR, Osborne CP, Pausas JG, Price O, Ripley BS, Rogers
619 BM, Schwilk DW, Simon MF, Turetsky MR, van der Werf GR, Zanne AE. 2018.
620 Biological and geophysical feedbacks with fire in the Earth System. *Environmental*
621 *Research Letters*, 13: 033003
- 622 Archibald S, Staver AC, Levin SA. 2012. Evolution of human-driven fire regimes in
623 Africa. *Proceedings of the National Academy of Science*, 109: 847–852.
- 624 Atkin OK, Schortemeyer M, McFarlane N, Evans JR. 1999. The response of fast- and
625 slow-growing *Acacia* species to elevated atmospheric CO₂: An analysis of the
626 underlying components of relative growth rate. *Oecologia* 120(4): 544–54.
- 627 Badger J, Bauwens I, Casso P, Davis N, Hahmann A, Bo Krohn Hansen S, Ohrbeck
628 Hansen B, Heathfield D, Knight O, Lacave O, Lizcano G, Bosch I Mas A, Mortensen
629 N, Olsen B, Onninen M, Potter Van Loon A, Volker P. 2019. Global Wind Atlas 3.0
630 Global Wind Atlas 3.0, a free, web-based application developed, owned and operated
631 by the Technical University of Denmark (DTU) using funding provided by the Energy
632 Sector Management Assistance Program and The World Bank.
- 633 Baso NC, Coetzee JA, Ripley BS, Hill MP. 2021. The effects of elevated atmospheric
634 CO₂ concentration on the biological control on invasive aquatic weeds. *Aquatic*
635 *Botany*, 170, 103348.

- 636 Beck HE, Zimmermann NE, McVicar TR, Vergopolan N, Berg A, Wood
637 EF. 2018. Present and future Köppen-Geiger climate classification maps at 1-km
638 resolution. *Nature Scientific Data*, 5: 180214
- 639 Beckage B, Platt WJ, Gross LJ. 2009. Vegetation, fire, and feedbacks: A disturbance-
640 mediated model of savannas. *American Naturalist* 174(6): 805–18.
- 641 Beerling DJ, Osborne CP. 2006. The origin of the savanna biome. *Global Change*
642 *Biology* 12(11): 2023–31.
- 643 Bond WJ, Midgley GF. 2000. A proposed CO₂-controlled mechanism of woody plant
644 invasion in grasslands and savannas. *Global Change Biology*. 6(8): 865–9.
- 645 Bond WJ, Midgley GF, Woodward FI. 2003. What controls South African vegetation—
646 climate or fire? *South African Journal of Botany*, 69(1), 79-91.
- 647 Bond WJ, Midgley JJ. 1995. Kill thy neighbour: An individualistic argument for the
648 evolution of flammability. *Oikos* 73(1): 79–85.
- 649 Bond WJ, Midgley JJ. 2001. Ecology of sprouting in woody plants: the persistence
650 niche. *Trends in Ecology and Evolution* 16: 45– 51.
- 651 Bond WJ, van Wilgen BW. 1996. Why and how do ecosystems burn? *Fire and Plants*
652 16–33.
- 653 Bond, W.J., 2008. What limits trees in C₄ grasslands and savannas? *Annual Review*
654 *of Ecology, Evolution, and Systematics* 39: 641-659.
- 655 Bradstock RA. 2019. A biogeographic model of fire regimes in Australia: Current and
656 future implications. *Global Ecology and Biogeography* 19(2): 145–58.
- 657 Buitenwerf R, Bond WJ, Stevens N, Trollope WSW. 2012. Increased tree densities in
658 South African savannas: >50 years of data suggests CO₂ as a driver. *Global Change*
659 *Biology* 18(2): 675–84.
- 660 Byram GM. 1959. Combustion of forest fuels. In: Davis KP (ed) *Forest fire: control and*
661 *use. McGraw-Hill, New York*, 61–89

- 662 Cardoso AW, Oliveras I, Abernethy KA, Jeffery KJ, Lehmann D, Edzang NJ,
663 McGregor I, Belcher CM, Bond WJ, Malhi YS. 2018. Grass species flammability, not
664 biomass, drives changes in fire behavior at tropical forest-savanna transitions.
665 *Frontiers in Forests and Global Change*, 1: 6
- 666 Chuvieco E, Aguado I, Dimitrakopoulos AP. 2004. Conversion of fuel moisture content
667 values to ignition potential for integrated fire danger assessment. *Canadian Journal of*
668 *Forest Research* 34(11): 2284–93.
- 669 Cox JT, Durrett R. 1988. Limit theorems for the spread of epidemics and forest fires.
670 *Stochastic Processes and their Application*, 30: 171–191.
- 671 Curtis PS, Wand SJE, Midgley GF, Jones MH. 1999. Elevated atmospheric CO₂
672 concentration: a meta-analytic test of current theories and perceptions. *Global*
673 *Change Biology* 5: 723–41.
- 674 Dale VH, Joyce LA, McNulty S, Neilson RP, Ayres MP, Flannigan MD. 2001. Climate
675 change and forest disturbances. *BioScience* 51(9): 723–34.
- 676 Danckwerts T. 1993. Reserve carbon and photosynthesis: Their role in regrowth of
677 *Themeda triandra*, a widely distributed subtropical graminaceous species. *Functional*
678 *Ecology*, 634-641.
- 679 D'Antonio CM, Vitousek PM. 1992. Biological Invasions by Exotic Grasses, the Grass
680 / Fire Cycle. *Annual Review of Ecology and Systematics* 23: 63-87.
- 681 de Magalhães RMQ, Schwilk DW. 2012. Leaf traits and litter flammability: Evidence
682 for non-additive mixture effects in a temperate forest. *Journal of Ecology* 100(5): 1153–
683 63.
- 684 Dorigo W, Preimesberger W, Reimer C; Van der Schalie R; Pasik A; De Jeu R, Paulik
685 C. 2020. Soil moisture gridded data from 1978 to present v201912.0.0. Copernicus
686 Climate Change Service (C3S) Climate Data Store (CDS).
- 687 Dunning T, Liabot A, Olofsson JK, Smith EM, Simpson KJ, Lundgren MR, Addicott E,
688 Gallagher RV, Chu Y, Pennington RT, Christin P, Lehman CER. 2017. The recent and
689 rapid spread of *Themeda triandra*. *Botany Letters* 164:4: 327-337

- 690 Eamus D, Palmer AR. 2007. Is climate change a possible explanation for woody
691 thickening in arid and semi-arid regions? *Research Letters in Ecology* 1–5.
- 692 Edwards EJ, Osborne CP, Stromberg C.A. 2010. The origins of C4 grasslands:
693 integrating evolutionary and ecosystem science. *Science* 328: 587–59.
- 694 Engelbrecht F, Adegoke J, Bopape MJ, Naidoo M, Garland R, Thatcher M, et al.
695 Projections of rapidly rising surface temperatures over Africa under low mitigation.
696 *Environmental Response Letters* 2015;10(8).
- 697 Fargeon H, Pimont F, Martin-StPaul N, De Caceres M, Ruffault J, Barbero R, Dupuy
698 JL. 2020. Projections of fire danger under climate change over France: where do the
699 greatest uncertainties lie?. *Climatic Change* 160(3):479-93.
- 700 Flannigan M, Stocks B, Turetsky M, Wotton M. 2009. Impacts of climate change on
701 fire activity and fire management in the circumboreal forest. *Global change biology*
702 15(3):549-60.
- 703 Forrestel EJ, Donoghue MJ, Smith MD. 2014. Convergent phylogenetic and functional
704 responses to altered fire regimes in mesic savanna grasslands of North America and
705 South Africa. *New Phytologist* 203.
- 706 Gao X, Schwilk DW. 2018. Grass canopy architecture influences temperature
707 exposure at soil surface. *Fire* 1(3): 35.
- 708 Gill AM, Moore PHR. 1996. Ignitibility of Leaves of Australian Plants. *A Contract*
709 *Report to the Australian Flora Foundation* 1–34.
- 710 Govender N, Trollope WSW, Van Wilgen BW. 2006. The effect of fire season, fire
711 frequency, rainfall and management on fire intensity in savanna vegetation in South
712 Africa. *Journal of Applied Ecology* 43(4): 748–58.
- 713 Grigulis K, Lavorel S, Davies ID, Dossantos A, Lloret F, Vilà M. 2005. Landscape-
714 scale positive feedbacks between fire and expansion of the large tussock grass,
715 *Ampelodesmos mauritanien* in Catalan shrublands. *Global Change Biology* 11(7):
716 1042–53.

- 717 Hijmans RJ. 2015. Introduction to the 'raster' package (version 2.3-24). *R-CRAN*
718 *Project* 1–27.
- 719 Hughes F, Vitousek P, Tunison T. 2014. Alien Grass Invasion and Fire In the Seasonal
720 Submontane Zone of Hawai. *Ecological Society of America* 72(2): 743–7.
- 721 Karger DN, Conrad O, Böhner J, Kawohl T, Kreft H, Soria-Auza RW, Zimmermann
722 NE, Linder HP, Kessler M. 2017. Climatologies at high resolution for the earth's land
723 surface areas. Scientific data. Sep 5;4(1): 1-20.
- 724 Keeley JE, Pausas JG, Rundel PW, Bond WJ, Bradstock RA. 2011. Fire as an
725 evolutionary pressure shaping plant traits. *Trends in Plant Science*, 16: 406-411.
- 726 Knapp AK, Hamerlynck EP, Owensby CE. 2020. Photosynthetic and Water Relations
727 Responses to Elevated CO₂ in the C₄ Grass *Andropogon gerardii*. *The University of*
728 *Chicago Press* 154(4): 459–66.
- 729 Lê, S, Josse J, Mazet F. 2008. Package FactoMineR. *Journal of Statistical Software*
730 25.1: 1-18.
- 731 Lehmann CER, Griffith DM, Simpson KJ, Anderson TM, Archibald S, Beerling DJ,
732 Bond WJ, Denton E, Edwards EJ, Forrester EJ, Fox DL, Georges D, Hoffman WA,
733 Kluyver T, Mucina L, Pau S, Ratnam J, Salamin N, Santini B, Smith MD, Spriggs EL,
734 Westley R, Still CJ, Strömberg CAE, Osborne CP. 2019. Functional diversification
735 enabled grassy biomes to fill global climate space. *BioRxiv*, 583625.
- 736 Manea A, Grootemaat S, Leishman MR. 2015. Leaf flammability and fuel load increase
737 under elevated CO₂ levels in a model grassland. *International Journal of Wildland Fire*,
738 24: 819–827.
- 739 Manea A, Leishman M. 2019. The resprouting response of co-occurring temperate
740 woody plant and grass species to elevated [CO₂]: An insight into woody plant
741 encroachment of grasslands. *Austral Ecology*, 44: 917–926
- 742 Morgan JW, Lunt ID. 1999. Effects of time-since-fire on the tussock dynamics of a
743 dominant grass (*Themeda triandra*) in a temperate Australian grassland. *Biological*
744 *Conservation*, 88(3): 379-386

- 745 Mutch RW. 1970. Wildland Fires and Ecosystems--A Hypothesis. *Ecology* 51(6):
746 1046-1051.
- 747 Pausas JG, Keeley JE, Schwilk DW. 2017. Flammability as an ecological and
748 evolutionary driver. *Journal of Ecology* 105(2): 289–97.
- 749 Peláez DV, Andrioli RJ, Elia OR, Bontti EE, Tomas MA, Blazquez
750 FR. 2013. Response of grass species to different fire frequencies in semi-arid
751 rangelands of central Argentina. *Rangeland Journal*, 35: 385– 392.
- 752 Pinheiro J, Bates D, DebRoy S, Sarkar D. 2021. Linear and Nonlinear Mixed Effects
753 Models. *R package version*, 3(57): 1-89.
- 754 Plucinski MP, Anderson WR. 2008. Laboratory determination of factors influencing
755 successful point ignition in the litter layer of shrubland vegetation. *International Journal*
756 *of Wildfire* 17(5): 628–37.
- 757 Prior LD, Murphy BP, Williamson GJ, Cochrane MA, Jolly WM, Bowman DMJS. 2017.
758 Does inherent flammability of grass and litter fuels contribute to continental patterns
759 of landscape fire activity? *Journal of Biogeography* 44(6):1225–38.
- 760 Pyne SJ, Andrews PL, Laven RD. 1996. Introduction to Wildland Fire. *John Wiley and*
761 *Sons, New York*.
- 762 Quirk J, Bellasio C, Johnson DA, Beerling DJ. 2019. Response of photosynthesis,
763 growth and water relations of a savannah-adapted tree and grass grown across high
764 to low CO₂. *Annals of Botany*, Volume 124, Issue 1: 77–90.
- 765 Ratajczak Z, Nippert JB, Collins SL. 2012. Woody encroachment decreases diversity
766 across North American grasslands and savannas. *Ecology*, 93: 697-703.
- 767 Reich PB, Hobbie SE, Lee TD. 2014. Plant growth enhancement by elevated CO₂
768 eliminated by joint water and nitrogen limitation. *Nature Geoscience*, 7: 920–924.
- 769 Reyes-Fox M, Steltzer H, Trlica MJ, McMaster GS, Andales AA, LeCain DR, Morgan
770 JA. 2014. Elevated CO₂ further lengthens growing season under warming conditions.
771 *Nature*, 510.7504: 259-262.

- 772 Rothermel RC. 1972. A mathematical model for predicting fire spread. *United States*
773 *Department of Agriculture. Forest Service Research Paper* 46.
- 774 Sankaran M, Hanan NP, Scholes RJ, Ratnam J, Augustine DJ, Cade BS. 2005.
775 Determinants of woody cover in African savannas. *Nature* 438(7069): 846–9.
- 776 Saralabai VC, Vivekanandan M, Babu RS. 1997. Plant responses to high CO₂
777 concentration in the atmosphere. *Photosynthetica*, 33(1): 7-37.
- 778 Scheiter S, Higgins SI, Osborne CP, Bradshaw C, Lunt D, Ripley BS, Taylor LL,
779 Beerling DJ. Fire and fire-adapted vegetation promoted C₄ expansion in the late
780 Miocene. *New Phytologist*. 2012 Aug;195(3):653-66.
- 781 Scholes RJ, Archer SR. 1997. Tree-grass interactions in savannas. *Annual review of*
782 *Ecology and Systematics*, 28(1): 517-544.
- 783 Schwilk DW. 2003. Flammability is a niche construction trait: canopy architecture
784 affects fire intensity. *American Naturalist* 162(6): 725–33.
- 785 Sionit N, Patterson DT. 1984. Responses of C₄ grasses to atmospheric CO₂
786 enrichment. *Oecologia*, 65(1): 30-34.
- 787 Simpson KJ, Jardine EC, Archibald S, Forrester EJ, Lehmann CER, Thomas GH,
788 Osborne CP. 2021. Resprouting grasses are associated with less frequent fire than
789 seeders. *New Phytologist*, 230: 832-844.
- 790 Simpson KJ, Olofsson JK, Ripley BS, Osborne CP. 2019. Frequent fires prime plant
791 developmental responses to burning. *Proceedings of the Royal Society, B*: 286 1315.
- 792 Simpson KJ, Ripley BS, Christin PA, Belcher CM, Lehmann CER, Thomas GH. 2016.
793 Determinants of flammability in savanna grass species. *Journal of Ecology* 104(1):
794 138–48.
- 795 Smit IPJ, Asner GP, Govender N, Kennedy-Bowdoin T, Knapp DE, Jacobson J. 2010.
796 Effects of fire on woody vegetation structure in African savanna. *Ecological*
797 *Applications* 20(7): 1865–75.

- 798 Smith AJ, Jones MW, Abatzoglou JT, Canadell JG, Betts RA. 2020. Climate Change
799 Increases the Risk of Wildfires: September 2020. *ScienceBrief*.
- 800 Snyman HA, Ingram LJ, Kirkman K P. 2013. *Themeda triandra*: a keystone grass
801 species. *African Journal of Range & Forage Science* 30(3), 99-125.
- 802 South African National Biodiversity Institute. 2012. Vegetation Map of South Africa,
803 Lesotho and Swaziland vector geospatial dataset. *Biodiversity GIS website*. Accessed
804 10 June 2021.
- 805 Staver AC, Botha J, Hedin L. 2017. Soils and fire jointly determine vegetation structure
806 in an African savanna. *New Phytologist* 216(4): 1151–60.
- 807 Stevens N, Lehmann CE, Murphy BP, Durigan G. 2017. Savanna woody
808 encroachment is widespread across three continents. *Global Change Biology* 23: 235-
809 244
- 810 Task G.S.D. 2000. Global gridded surfaces of selected soil characteristics (IGBP-DIS).
811 ORNL DAAC.
- 812 Trabaud L. 1976. Inflammabilité et combustibilité des principales espèces des
813 garrigues de la région méditerranéenne. *Oecologia Plantarum* 11: 117-136.
- 814 Trabucco A, Zomer R. 2010. Global soil water balance geo spatial database. CGIAR
815 Consortium for Spatial Information, Published Online, available from the CGIAR-CSI
816 Geo Portal at <https://cgiarcsi.community>,
- 817 Trapnell CG. 1959. Ecological results of woodland and burning experiments in
818 Northern Rhodesia. *Journal of Ecology* 47(1): 129.
- 819 Trollope WSW. 1978. Die gedrag van vuur - 'n voorlopige ondersoek. *Annual*
820 *Congresses of the Grassland Society of Southern Africa* 13(1):123–8.
- 821 Uys RG, Bond WJ, Everson TM. 2004. The effect of different fire regimes on plant
822 diversity in southern African grasslands. *Biological Conservation*, 118: 489–499.
- 823 Vacchiano G, Ascoli D. 2015. An implementation of the rothermel fire spread model in
824 the R programming language. *Fire Technology* 51(3): 523–35.

- 825 van Altena C, van Logtestijn RSP, Cornwell WK, Cornelissen JHC. 2012. Species
826 composition and fire: Non-additive mixture effects on ground fuel flammability.
827 *Frontiers in Plant Science* 3(APR):1–10.
- 828 van der Werf GR, Randerson JT, Giglio L, Collatz GJ, Kasibhatla PS, Arellano JrAF.
829 2006. Interannual variability in global biomass burning emissions from 1997 to 2004.
830 *Atmosphere Chemistry and Physics*, 6: 3423–3441
- 831 van Langevelde FVV, Claudius ADM, Vijver VD, Kumar, L. Van De Koppel J, Ridder,
832 NV, Van Andel J, Skidmore AK, Hearne JW, Stroosnijder L, Bond WJ, Prins HHT.
833 2003. Effects of fire and herbivory on the stability of savanna ecosystems. *Ecology*
834 84(2): 337–50.
- 835 Van Wilgen BW, Higgins KB, Bellstedt DU. 1990. The role of vegetation structure and
836 fuel chemistry in excluding fire from forest patches in the fire-prone fynbos shrublands
837 of South Africa. *Journal of Ecology* 78(1): 210–22.
- 838 Van Wyk P. 1971. Veld burning in the Kruger National Park. An interim report of some
839 aspects of research. *Proceedings of the 11th Tall Timbers Fire Ecology Conference*
840 T:9–31.
- 841 Visser V, Woodward FI, Freckleton RP, Osborne CP. 2012. Environmental factors
842 determining the phylogenetic structure of C₄ grass communities. *Journal of*
843 *Biogeography*, 39: 232–246.
- 844 Walker BH, Ludwig D, Holling CS, Peterman RM. 1981. Stability of Semi-Arid Savanna
845 Grazing Systems. *British Ecological Society* 69(2): 473–98.
- 846 Wigley BJ, Charles-Dominique T, Hempson GP, Stevens N, te Beest M, Archibald S,
847 Bond WJ, Bunney K, Coetsee C, Donaldson J, Fidelis A. 2021. Corrigendum to: A
848 handbook for the standardised sampling of plant functional traits in disturbance-prone
849 ecosystems, with a focus on open ecosystems. *Australian Journal of Botany*, 69(2):
850 110-110.
- 851 Wyse SV, Perry GLW, Curran TJ. 2018. Shoot-Level Flammability of Species Mixtures
852 is Driven by the Most Flammable Species: Implications for Vegetation-Fire Feedbacks
853 Favouring Invasive Species. *Ecosystems* 21, 886–900.

854 Zubkova M, Boschetti L, Abatzoglou JT, Giglio L. 2019. Changes in fire activity in
855 Africa from 2002 to 2016 and their potential drivers. *Geophysical research letters*,
856 46(13):7643-7653.

For Peer Review Only

1 **Supplementary materials**

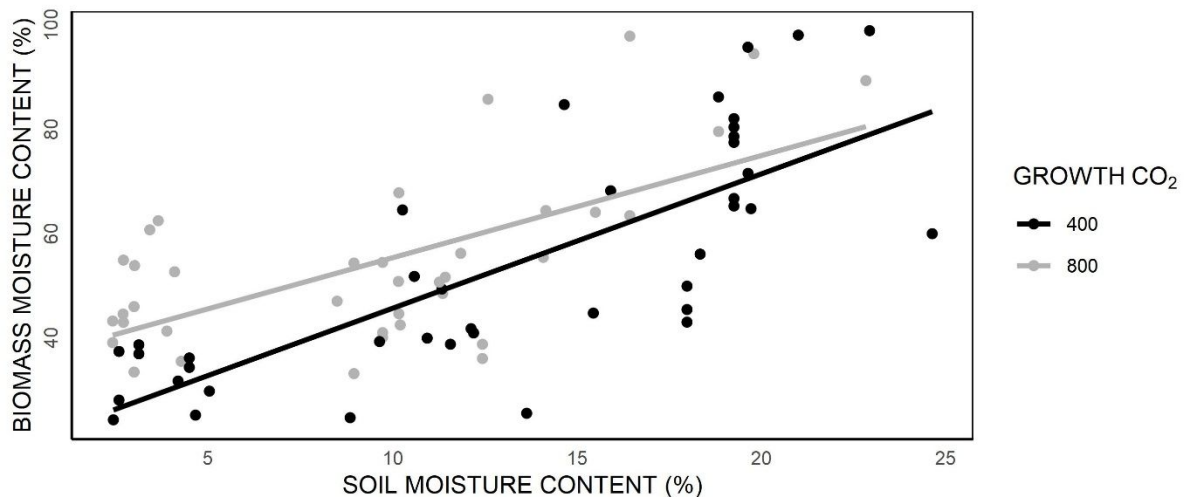
2 **Table S.1.** % of the time where measured [CO₂] was within the standard deviation
3 (SD), within 50 ppm, and within 100 ppm of the target [CO₂].

Target (ppm)	CO ₂ % S.D.	% within ppm	% within 50 ppm	% within 100
400	72.82	92.85	99.66	
800	70.89	38.42	66.09	

4
5 **Table S.2.** Environmental data for the experimental period in the open-top chambers.
6 Data relating to temperature, relative humidity, precipitation, solar inclination, azimuth,
7 elevation, hour angle, declination, and air mass experienced during the experimental
8 period are shown.

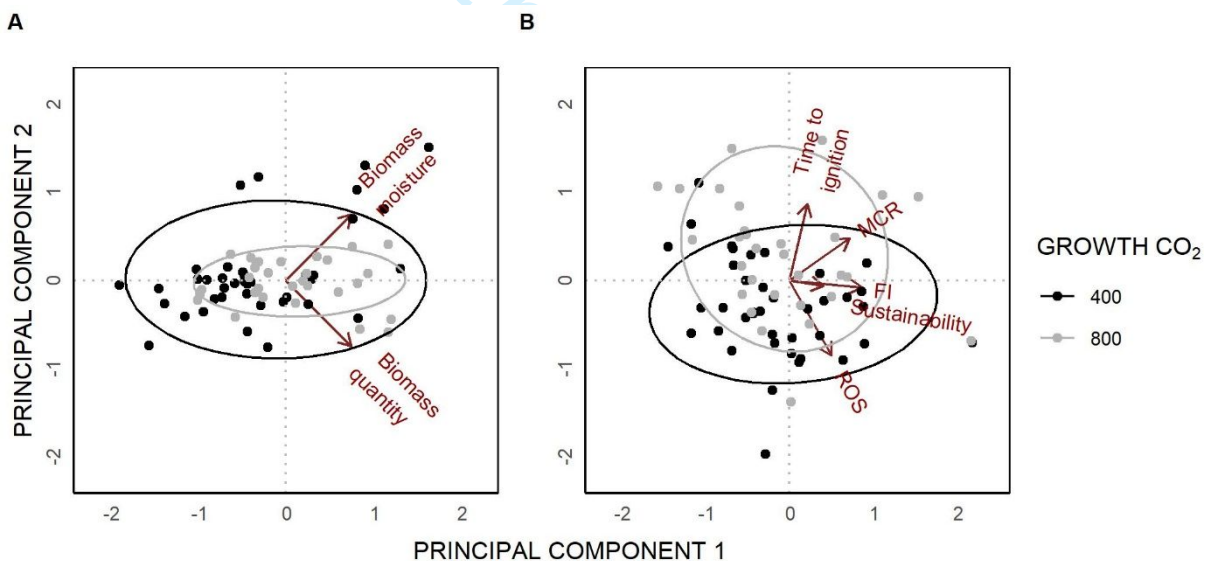
Environmental variable	Mean	Minimum	Maximum
Temperature (°C)	19.25	0	42.92
Relative Humidity (%)	77.93	5.12	100.00
Daily precipitation (mm)	0.87		
Monthly precipitation (mm)	18.53		
Total accumulated precipitation (mm)	111.2		
Solar inclination (W/m ²)	663.6		
Solar azimuth (°)	179.96		
Solar elevation (°)	4.94		
Solar hour angle (°)	-0.0002		
Solar declination (°)	-0.12		
Solar air mass	5.42		

9



10

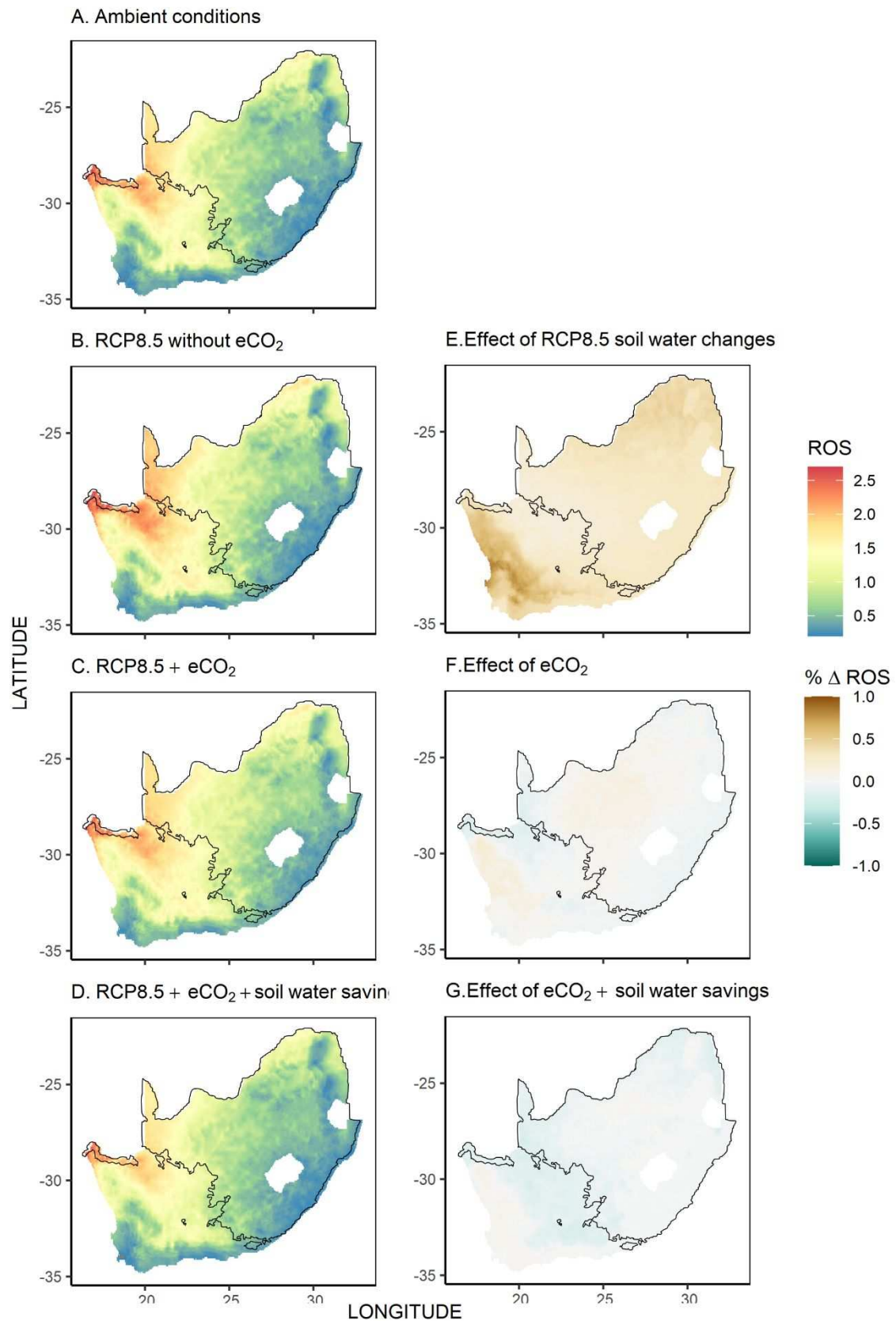
11 **Figure S.1.** The relationship between biomass moisture content and soil moisture
 12 content for *Themeda triandra* grasses exposed to different CO₂ (400 or 800 ppm)
 13 treatments. 400ppm: $y=257.74x+19.789$, $R^2=0.60$, $p<0.001$. 800ppm:
 14 $y=218.83x+32.474$, $R^2=0.46$, $p<0.001$.



15

16 **Figure S.2.** Principal components analysis of C₄ grass (*Themeda triandra*) traits
 17 associated with flammability: [A] biomass moisture, and biomass quantity. [B]
 18 sustainability, ignitability, maximum combustion rate (*MCR*), predicted rate of spread
 19 (*ROS*), and predicted fireline intensity (*FI*). All plants used in the PCA were well
 20 watered.

21

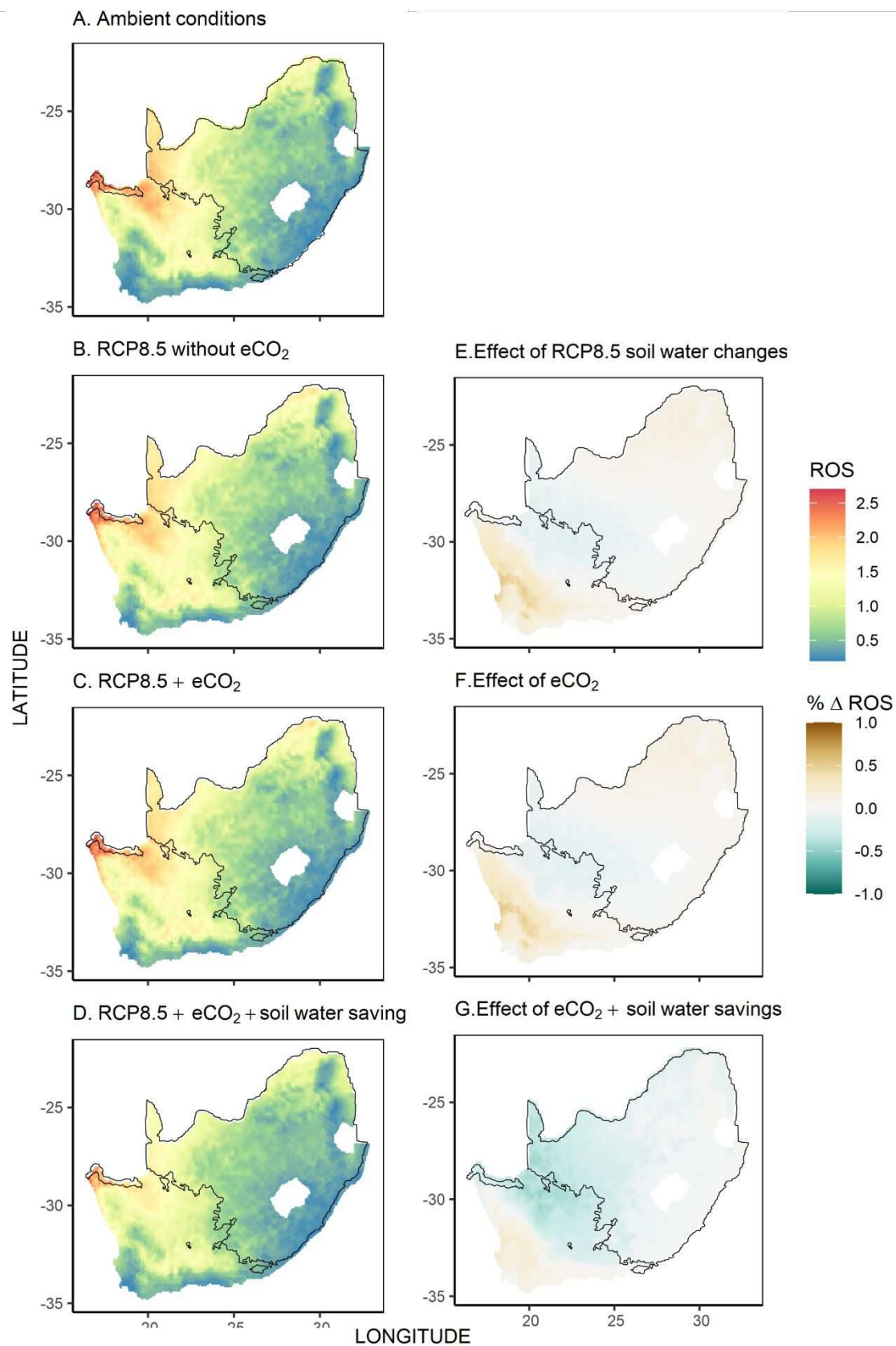


22

23 **Figure S.3.** Rate of fire spread (ROS; m min⁻¹) (Rothermel, 1972) is predicted for
 24 *Themeda triandra* grasses under ambient conditions [A], under predicted future soil

3

25 moisture conditions (for the year 2070 according to the RCP 8.5 CCSM4 projection
26 scenario) without the effects of eCO₂ on plant physiology and flammability [B], under
27 predicted future soil moisture conditions with the effects of eCO₂ on plant physiology
28 and flammability incorporated, and under predicted future soil moisture conditions
29 adjusted according to experimentally determined soil water savings (24.08 ± 4.26%
30 soil water saving) associated with eCO₂ and with the effects of eCO₂ on plant
31 physiology and flammability incorporated [D]. Predictions are based only on the
32 flammability of the widespread savanna grass, *Themeda triandra*, grown under aCO₂
33 (400ppm) and eCO₂ (800ppm). The area outlined in black represents that currently
34 classified as savanna or grassland biomes (South African National Biodiversity
35 Institute, 2012).

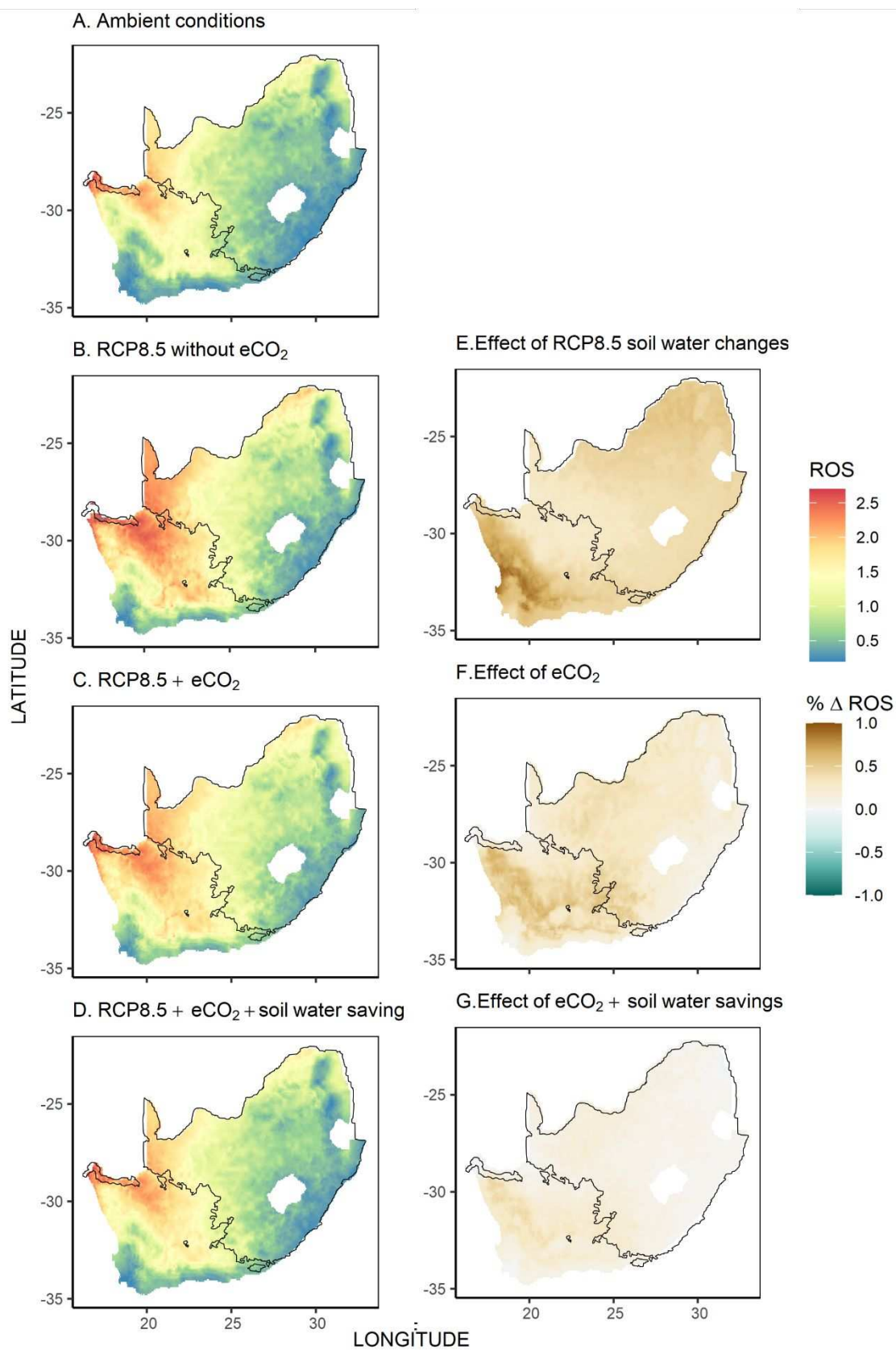


36

37 **Figure S.4.** Rate of fire spread (ROS; m min⁻¹) (Rothermel, 1972) is predicted for38 *Themeda triandra* grasses under ambient conditions [A], under predicted future soil

39 moisture conditions (for the year 2070 according to the RCP 8.5 CM5A projection
40 scenario) without the effects of eCO₂ on plant physiology and flammability [B], under
41 predicted future soil moisture conditions with the effects of eCO₂ on plant physiology
42 and flammability incorporated, and under predicted future soil moisture conditions
43 adjusted according to experimentally determined soil water savings (24.08 ± 4.26%
44 soil water saving) associated with eCO₂ and with the effects of eCO₂ on plant
45 physiology and flammability incorporated [D]. Predictions are based only on the
46 flammability of the widespread savanna grass, *Themeda triandra*, grown under aCO₂
47 (400ppm) and eCO₂ (800ppm). The area outlined in black represents that currently
48 classified as savanna or grassland biomes (South African National Biodiversity
49 Institute, 2012).

50

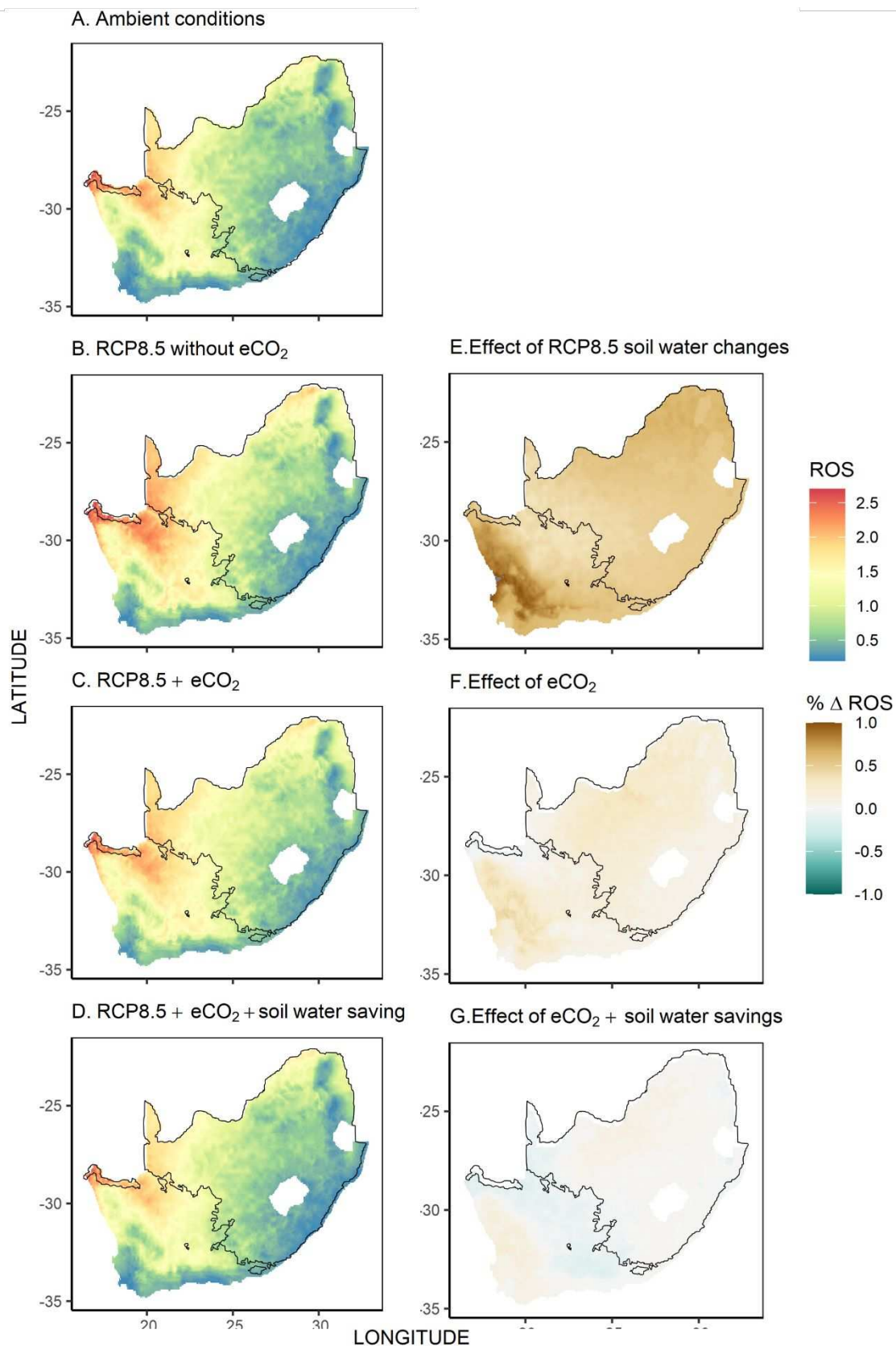


51

52 **Figure S.5.** Rate of fire spread (ROS; m min⁻¹) (Rothermel, 1972) is predicted for
 53 *Themeda triandra* grasses under ambient conditions [A], under predicted future soil

7

54 moisture conditions (for the year 2070 according to the RCP 8.5 MIROC-ESM
55 projection scenario) without the effects of eCO₂ on plant physiology and flammability
56 [B], under predicted future soil moisture conditions with the effects of eCO₂ on plant
57 physiology and flammability incorporated, and under predicted future soil moisture
58 conditions adjusted according to experimentally determined soil water savings (24.08
59 ± 4.26% soil water saving) associated with eCO₂ and with the effects of eCO₂ on plant
60 physiology and flammability incorporated [D]. Predictions are based only on the
61 flammability of the widespread savanna grass, *Themeda triandra*, grown under aCO₂
62 (400ppm) and eCO₂ (800ppm). The area outlined in black represents that currently
63 classified as savanna or grassland biomes (South African National Biodiversity
64 Institute, 2012).

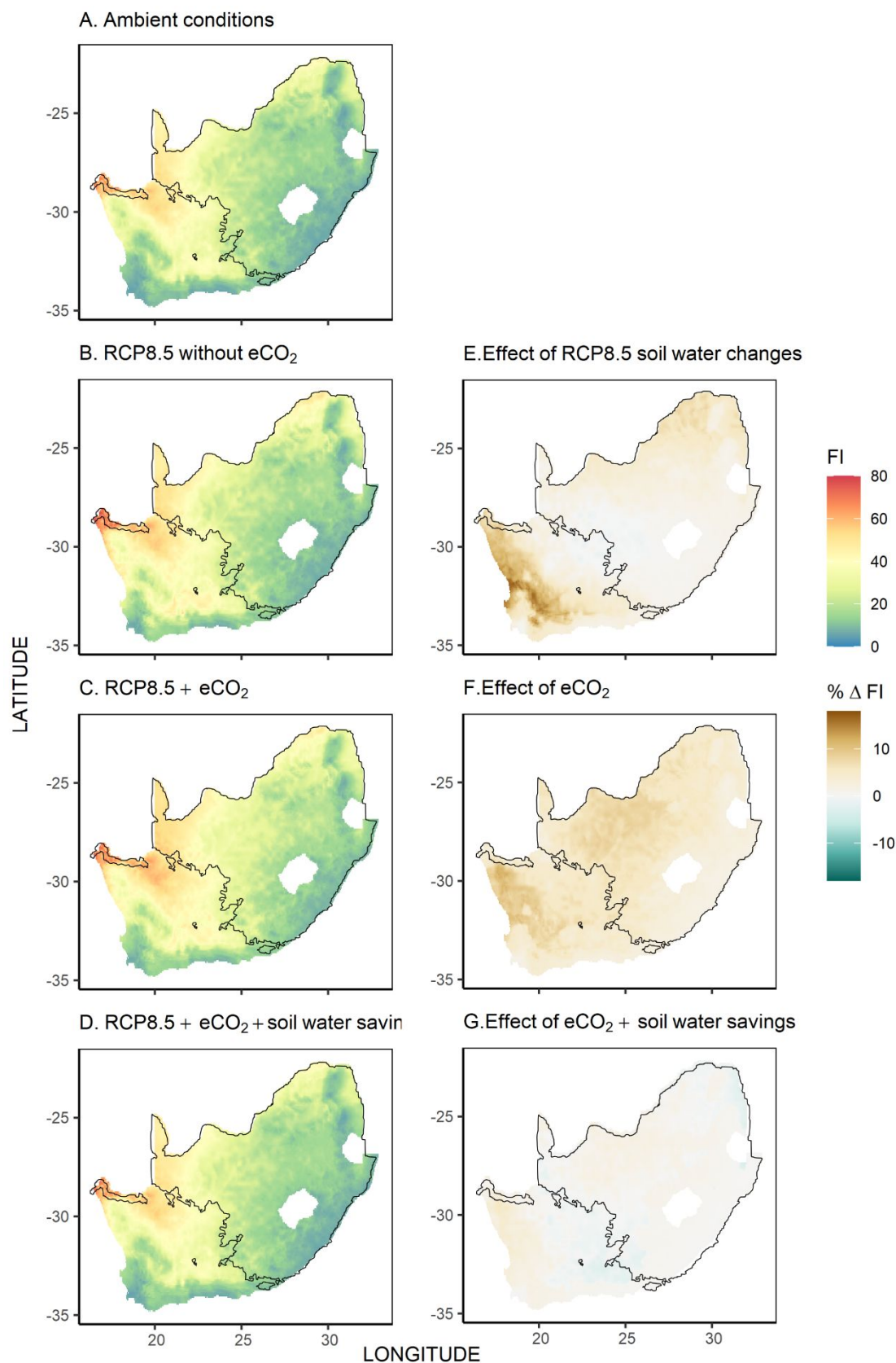


65

66 **Figure S.6.** Rate of fire spread (ROS; m min⁻¹) (Rothermel, 1972) is predicted for67 *Themeda triandra* grasses under ambient conditions [A], under predicted future soil

68 moisture conditions (for the year 2070 according to the RCP 8.5 MPI-ESM-LR
69 projection scenario) without the effects of eCO₂ on plant physiology and flammability
70 [B], under predicted future soil moisture conditions with the effects of eCO₂ on plant
71 physiology and flammability incorporated, and under predicted future soil moisture
72 conditions adjusted according to experimentally determined soil water savings (24.08
73 ± 4.26% soil water saving) associated with eCO₂ and with the effects of eCO₂ on plant
74 physiology and flammability incorporated [D]. Predictions are based only on the
75 flammability of the widespread savanna grass, *Themeda triandra*, grown under aCO₂
76 (400ppm) and eCO₂ (800ppm). The area outlined in black represents that currently
77 classified as savanna or grassland biomes (South African National Biodiversity
78 Institute, 2012).

79

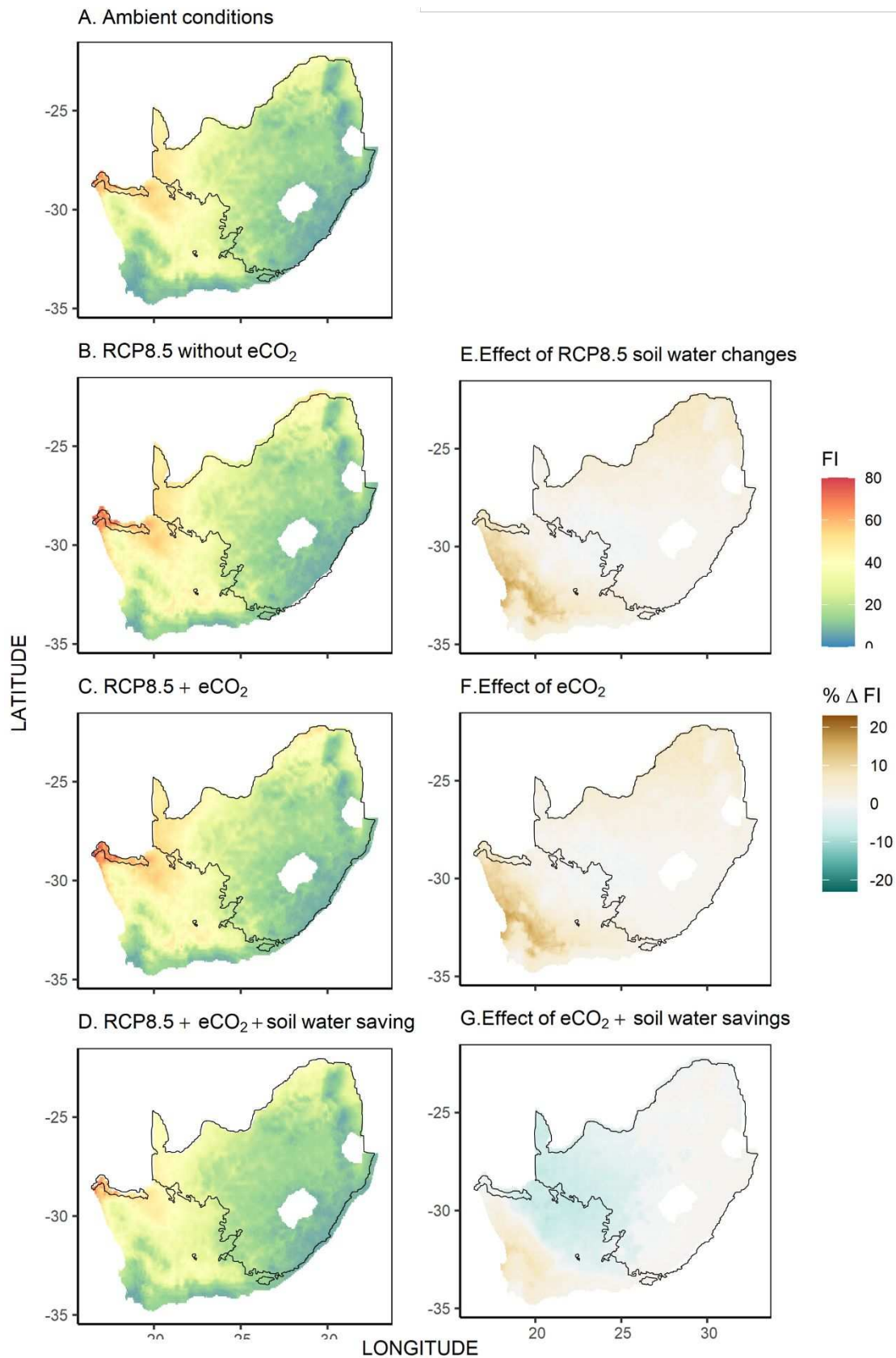


80

81 **Figure S.7.** Fireline intensity (FI ; $\text{KJ m}^{-2} \text{s}^{-1}$) (based on Byram's FI ; Byram 1959) is
 82 predicted for *Themeda triandra* grasses under ambient conditions [A], under predicted
 83 future soil moisture conditions (for the year 2070 according to the RCP 8.5 CCSM4

84 projection scenario) without the effects of eCO₂ on plant physiology and flammability
85 [B], under predicted future soil moisture conditions with the effects of eCO₂ on plant
86 physiology and flammability incorporated, and under predicted future soil moisture
87 conditions adjusted according to experimentally determined soil water savings (24.08
88 ± 4.26% soil water saving) associated with eCO₂ and with the effects of eCO₂ on plant
89 physiology and flammability incorporated [D]. Predictions are based only on the
90 flammability of the widespread savanna grass, *Themeda triandra*, grown under aCO₂
91 (400ppm) and eCO₂ (800ppm). The area outlined in black represents that currently
92 classified as savanna or grassland biomes (South African National Biodiversity
93 Institute, 2012).

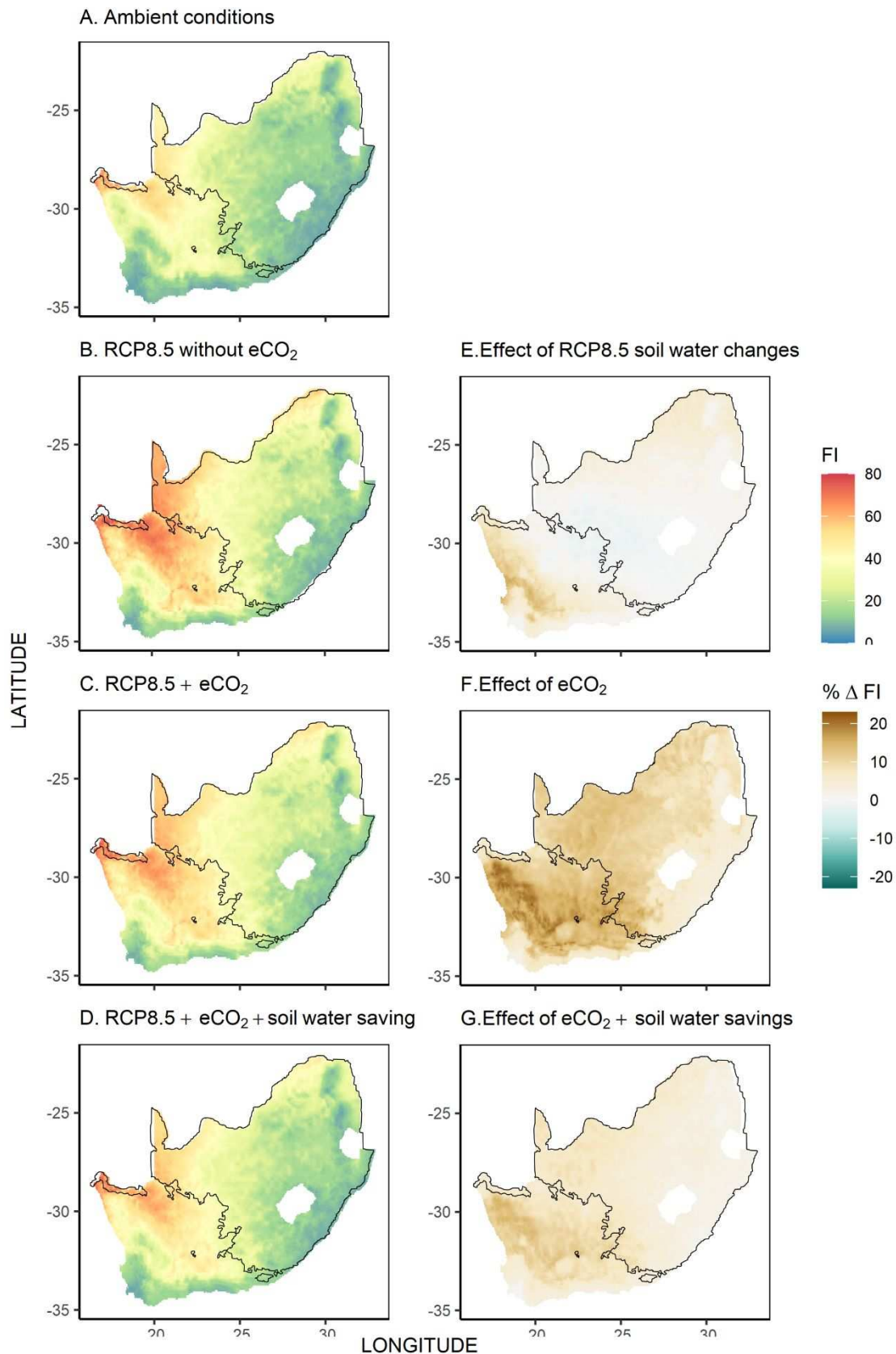
For Peer Review Only



94

95 **Figure S.8.** Fireline intensity (FI ; $\text{KJ m}^{-2} \text{s}^{-1}$) (based on Byram's FI ; Byram 1959) is96 predicted for *Themeda triandra* grasses under ambient conditions [A], under predicted

97 future soil moisture conditions (for the year 2070 according to the RCP 8.5 CM5A
98 projection scenario) without the effects of eCO₂ on plant physiology and flammability
99 [B], under predicted future soil moisture conditions with the effects of eCO₂ on plant
100 physiology and flammability incorporated, and under predicted future soil moisture
101 conditions adjusted according to experimentally determined soil water savings (24.08
102 ± 4.26% soil water saving) associated with eCO₂ and with the effects of eCO₂ on plant
103 physiology and flammability incorporated [D]. Predictions are based only on the
104 flammability of the widespread savanna grass, *Themeda triandra*, grown under aCO₂
105 (400ppm) and eCO₂ (800ppm). The area outlined in black represents that currently
106 classified as savanna or grassland biomes (South African National Biodiversity
107 Institute, 2012).



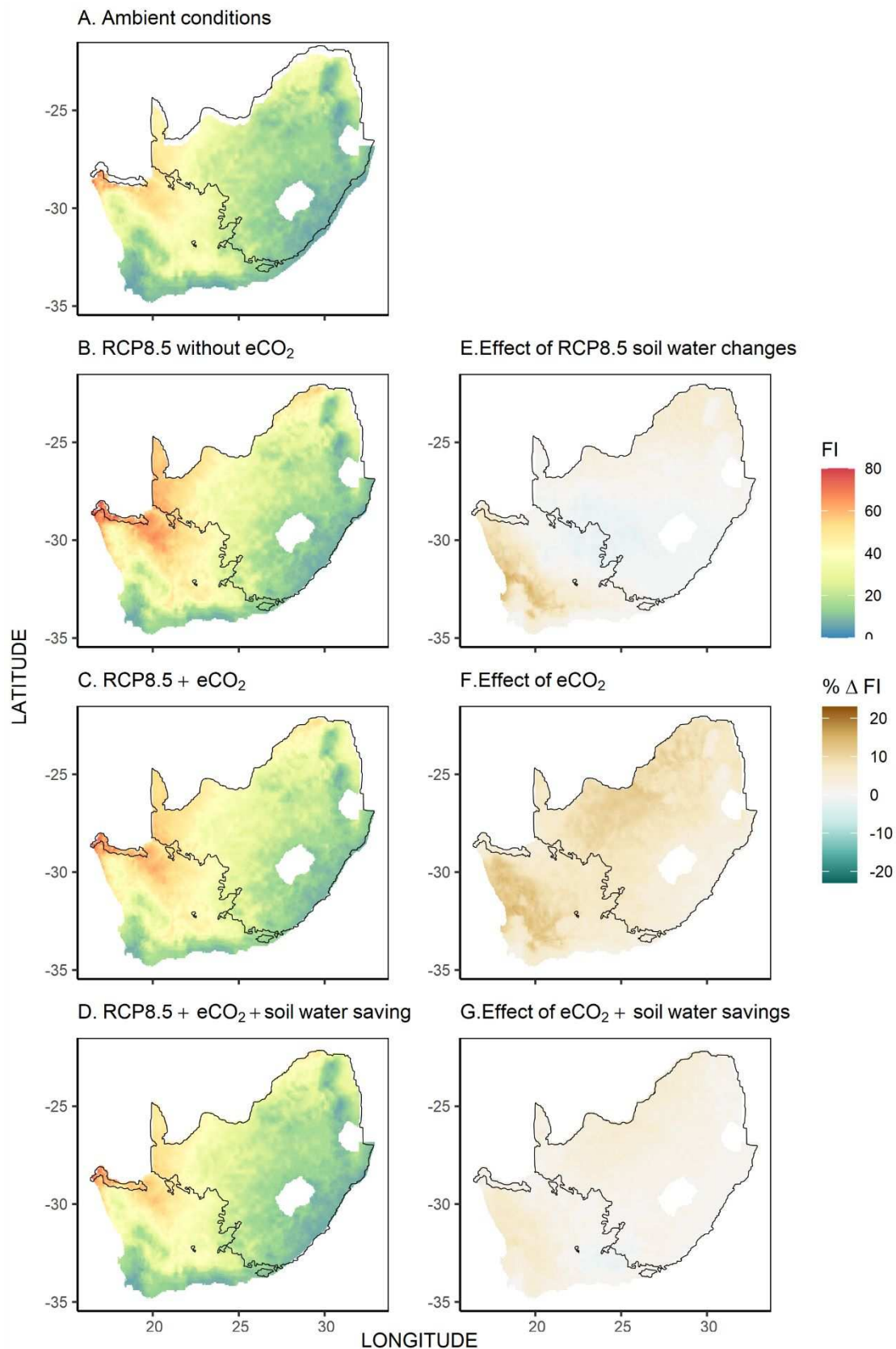
108

109 **Figure S.9.** Fireline intensity (FI ; $\text{KJ m}^{-2} \text{s}^{-1}$) (based on Byram's FI ; Byram 1959) is
 110 predicted for *Themeda triandra* grasses under ambient conditions [A], under predicted
 111 future soil moisture conditions (for the year 2070 according to the RCP 8.5 MIROC-

15

112 ESM projection scenario) without the effects of eCO₂ on plant physiology and
113 flammability [B], under predicted future soil moisture conditions with the effects of
114 eCO₂ on plant physiology and flammability incorporated, and under predicted future
115 soil moisture conditions adjusted according to experimentally determined soil water
116 savings (24.08 ± 4.26% soil water saving) associated with eCO₂ and with the effects
117 of eCO₂ on plant physiology and flammability incorporated [D]. Predictions are based
118 only on the flammability of the widespread savanna grass, *Themeda triandra*, grown
119 under aCO₂ (400ppm) and eCO₂ (800ppm). The area outlined in black represents that
120 currently classified as savanna or grassland biomes (South African National
121 Biodiversity Institute, 2012).

For Peer Review Only



122

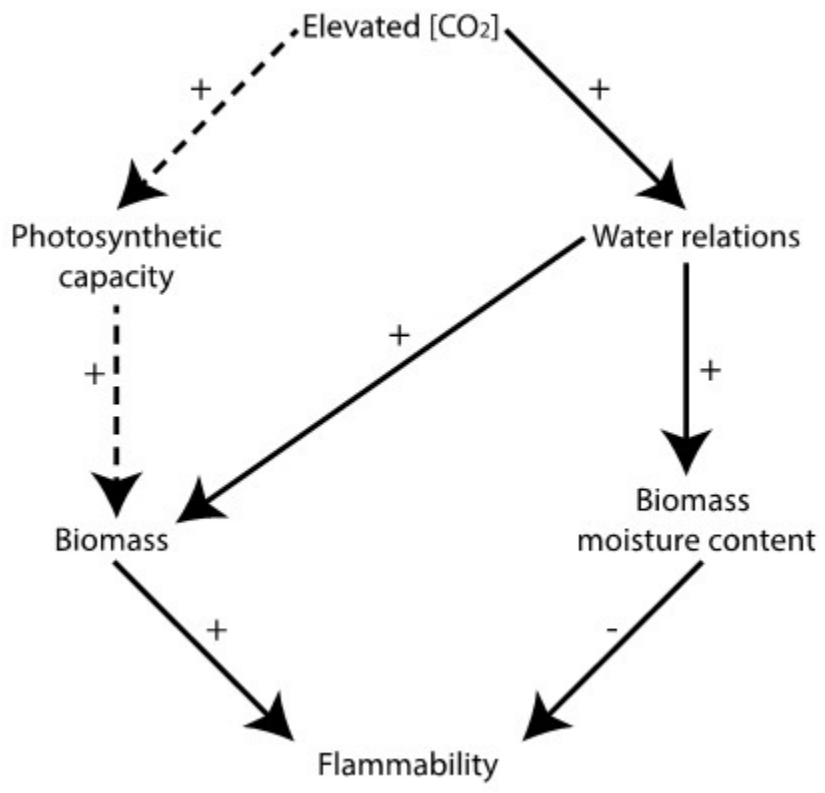
123 **Figure S.10.** Fireline intensity (FI ; $\text{KJ m}^{-2} \text{s}^{-1}$) (based on Byram's FI ; Byram 1959) is
 124 predicted for *Themeda triandra* grasses under ambient conditions [A], under predicted
 125 future soil moisture conditions (for the year 2070 according to the RCP 8.5 MPI-ESM-

17

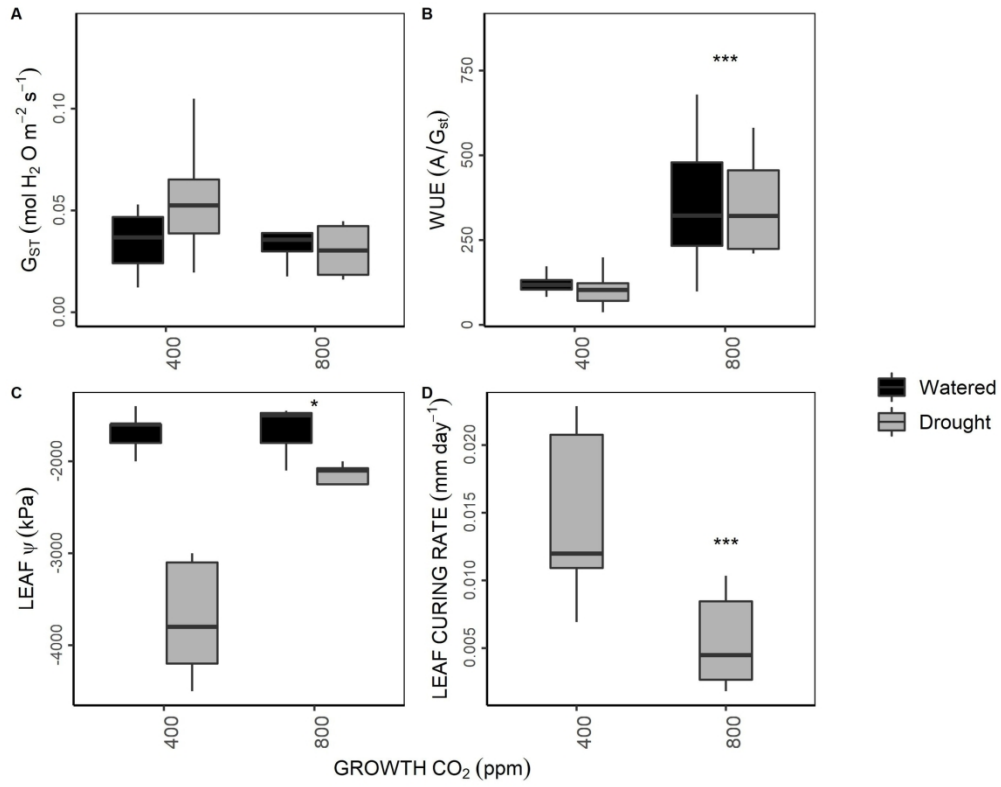
126 LR projection scenario) without the effects of eCO₂ on plant physiology and
127 flammability [B], under predicted future soil moisture conditions with the effects of
128 eCO₂ on plant physiology and flammability incorporated, and under predicted future
129 soil moisture conditions adjusted according to experimentally determined soil water
130 savings (24.08 ± 4.26% soil water saving) associated with eCO₂ and with the effects
131 of eCO₂ on plant physiology and flammability incorporated [D]. Predictions are based
132 only on the flammability of the widespread savanna grass, *Themeda triandra*, grown
133 under aCO₂ (400ppm) and eCO₂ (800ppm). The area outlined in black represents that
134 currently classified as savanna or grassland biomes (South African National
135 Biodiversity Institute, 2012).

136

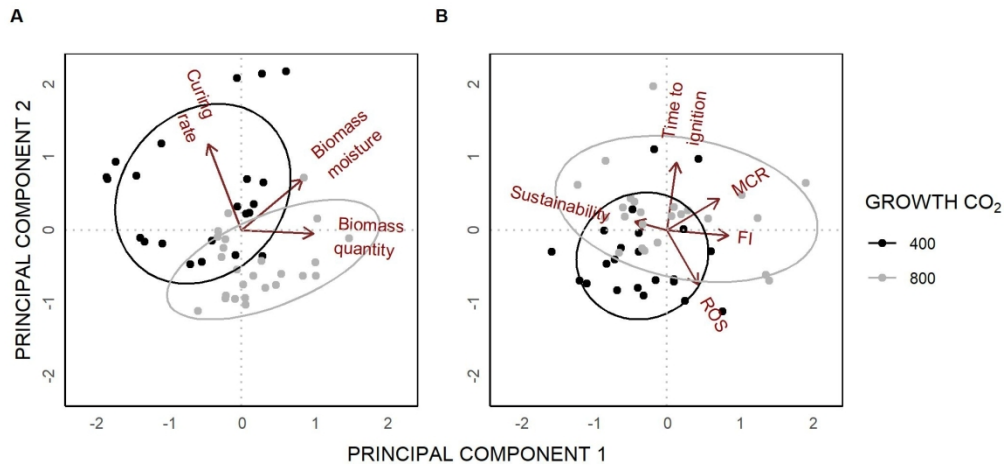
137



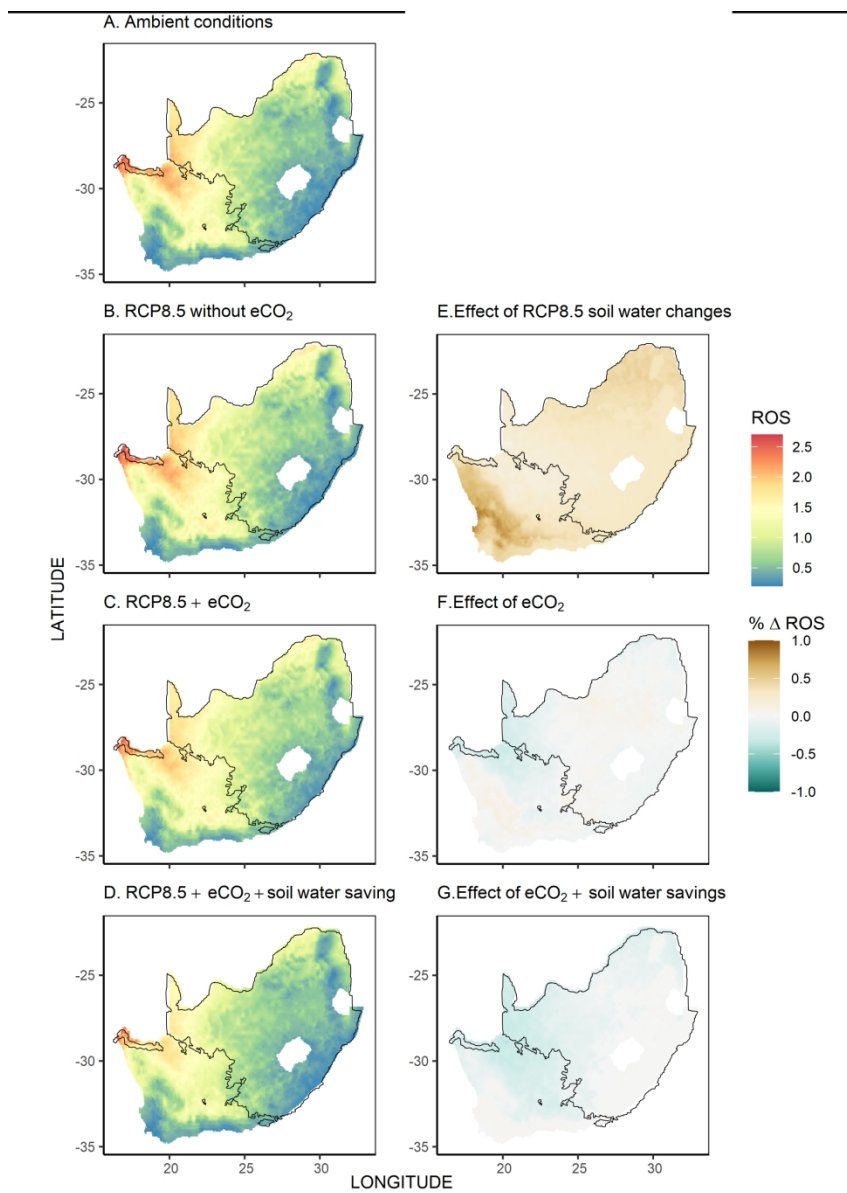
83x80mm (137 x 136 DPI)



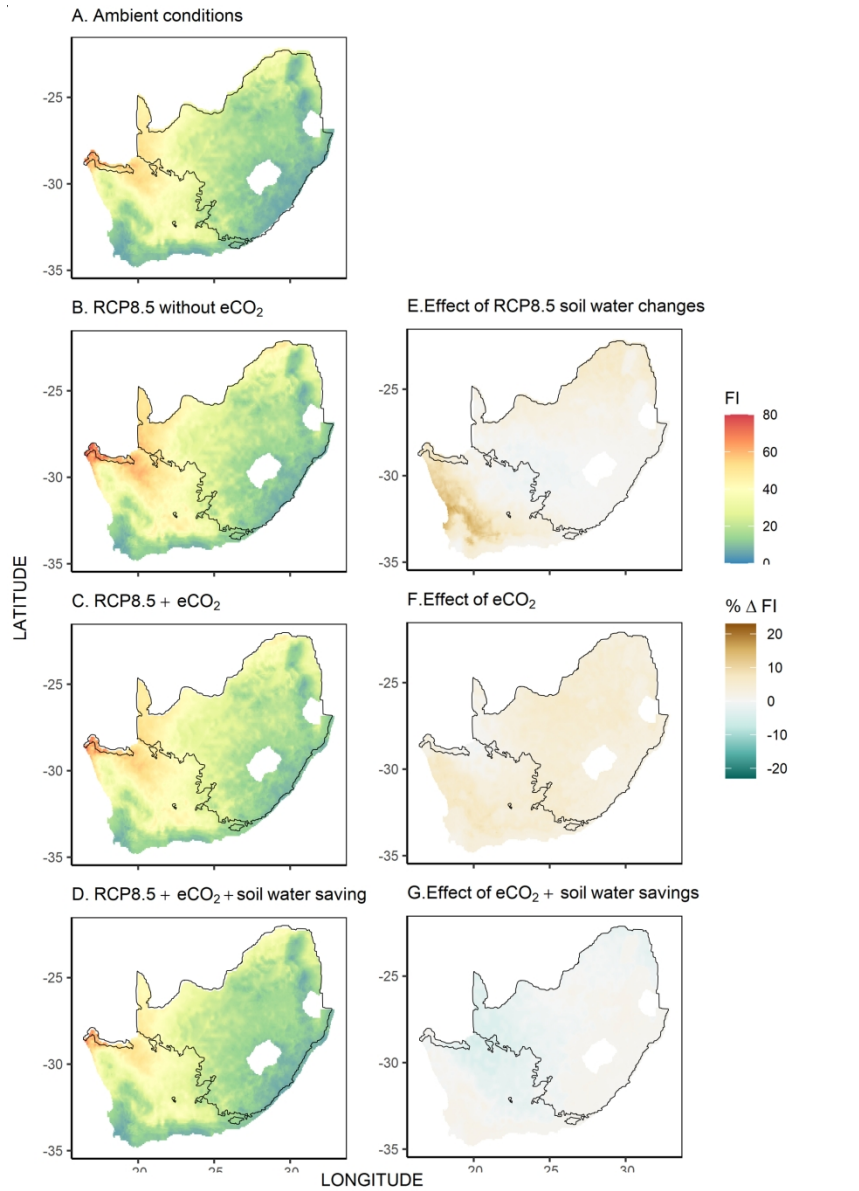
175x139mm (300 x 300 DPI)



175x89mm (300 x 300 DPI)



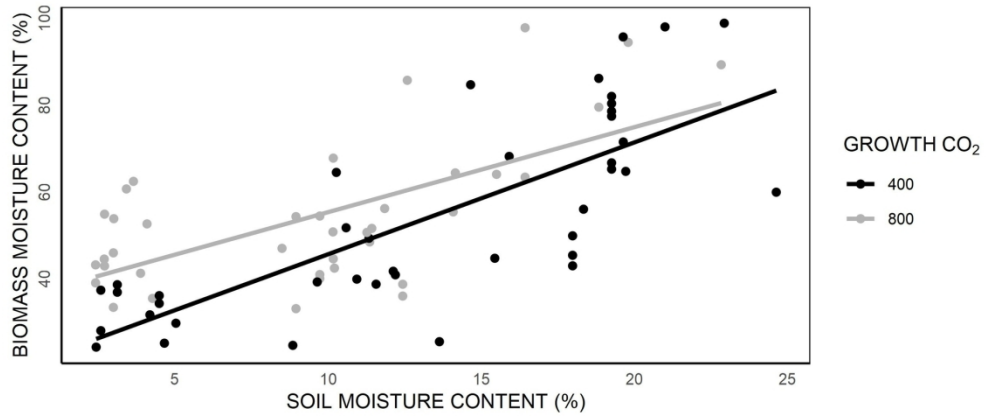
446x628mm (130 x 130 DPI)



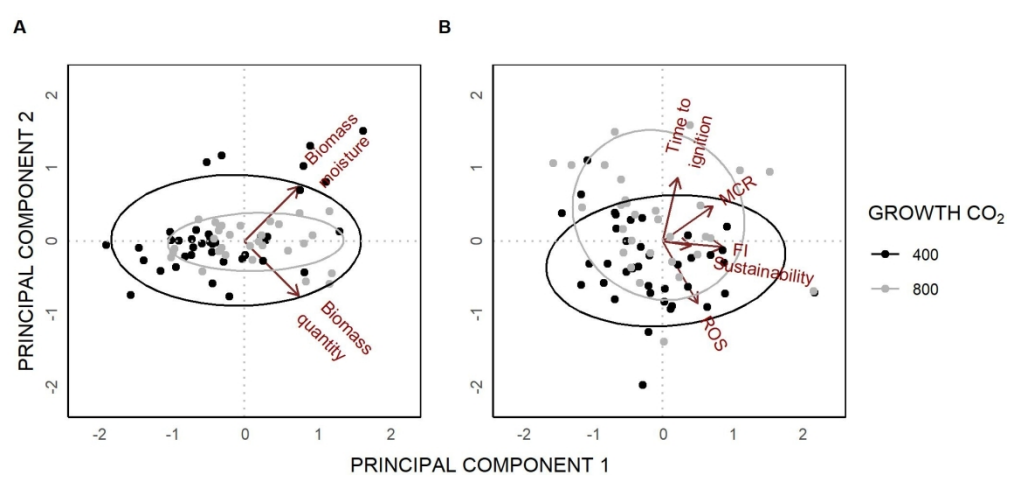
447x628mm (130 x 130 DPI)

Parameter	CO ₂ (ppm)	Watered	Drought
2.1. Plant traits			
Fuel load per tussock (g)	400	15.37 ± 9.87	10.54 ± 5.1
	800	17.48 ± 6.55	16.98 ± 8.63 *
Biomass moisture content (%)	400	54.27 ± 24.01	50.15 ± 18.59
	800	56.28 ± 19.7	52.6 ± 19.15
Biomass moisture content when soil moisture content <10% (%)	400	-	32.53 ± 5.48
	800	-	47.0 ± 9.22 **
Biomass density (g cm⁻²)	400	1.64 ± 1.27	1.36 ± 0.74
	800	1.90 ± 0.83	1.70 ± 0.97
Leaf SA:V ratio	400	88.22 ± 34.97	89.85 ± 33.24
	800	84.14 ± 18.37	88.14 ± 31.70
2.2. Flammability measures			
Time to ignition (s)	400	2.47 ± 0.16	2.74 ± 1.43
	800	3 ± 0.17 *	2.98 ± 1.16
Sustainability (s)	400	65.47 ± 25.65	62.62 ± 28.05
	800	58.41 ± 24.71	57.31 ± 21.39
Combustibility (g s⁻¹)	400	0.0036 ± 0.0014	0.0040 ± 0.0018
	800	0.0042 ± 0.002	0.0041 ± 0.0016
Estimated maximum combustion rate (gs⁻¹)	400	10.33 ± 6.57	7.08 ± 3.35
	800	11.81 ± 4.35	11.30 ± 5.66 *
2.3. Fire behaviour			
ROS (m min⁻¹)	400	2.44 ± 0.82	2.13 ± 0.78
	800	1.86 ± 0.93 *	1.88 ± 0.86 *
FI (KJ m⁻² s⁻¹)	400	77.9 ± 56.73	53.26 ± 32.76
	800	68.47 ± 50.47	51.18 ± 37.94

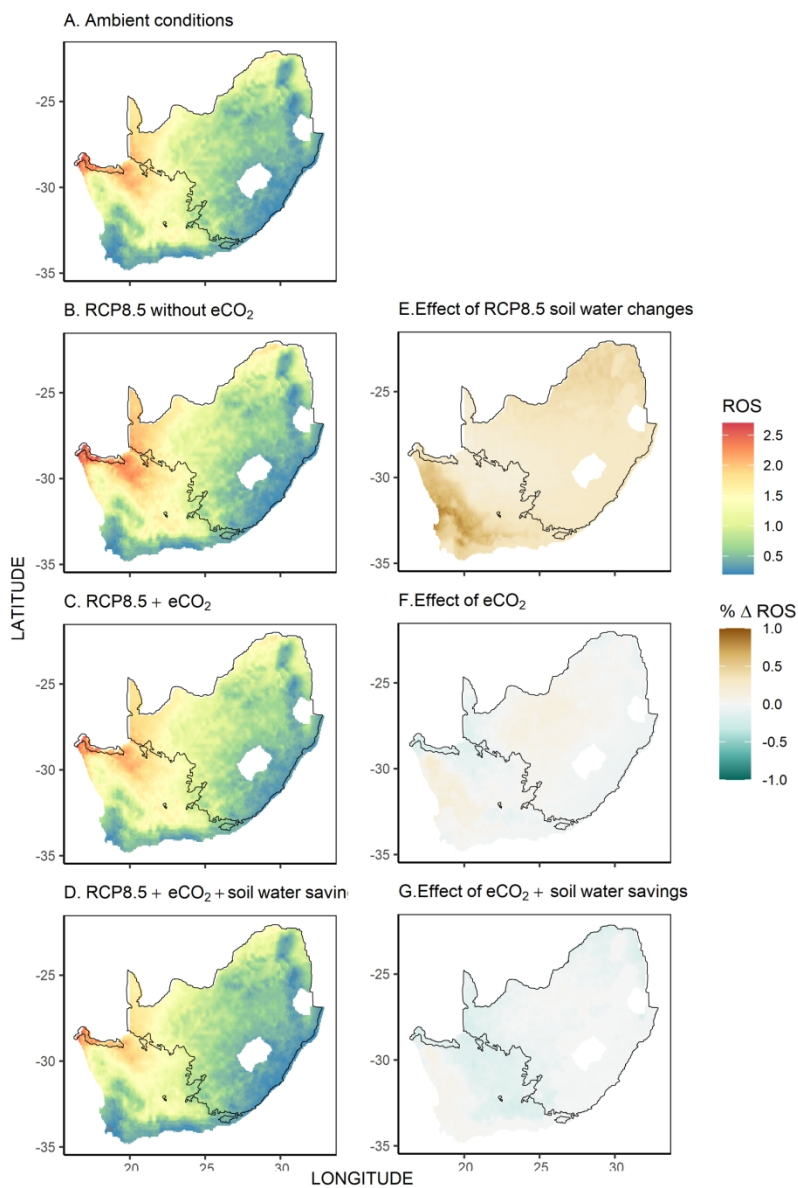
Model	Parameter	Effect of CO ₂ treatment		
		Estimate	SE	P value
A. Plant traits	Leaf curing rate	-0.009	0.003	<0.001 **
	Biomass moisture	4.63	6.18	0.5
	Biomass quantity	6.45	2.47	<0.05 *
B. Flammability	Time to ignition	0.96	0.45	0.09
	Sustainability	-12.36	17.35	0.52
	<i>MCR</i>	4.24	1.63	<0.05 *
	<i>ROS</i>	-0.43	0.23	<0.05 *
	<i>FI</i>	24.26	13.96	0.15



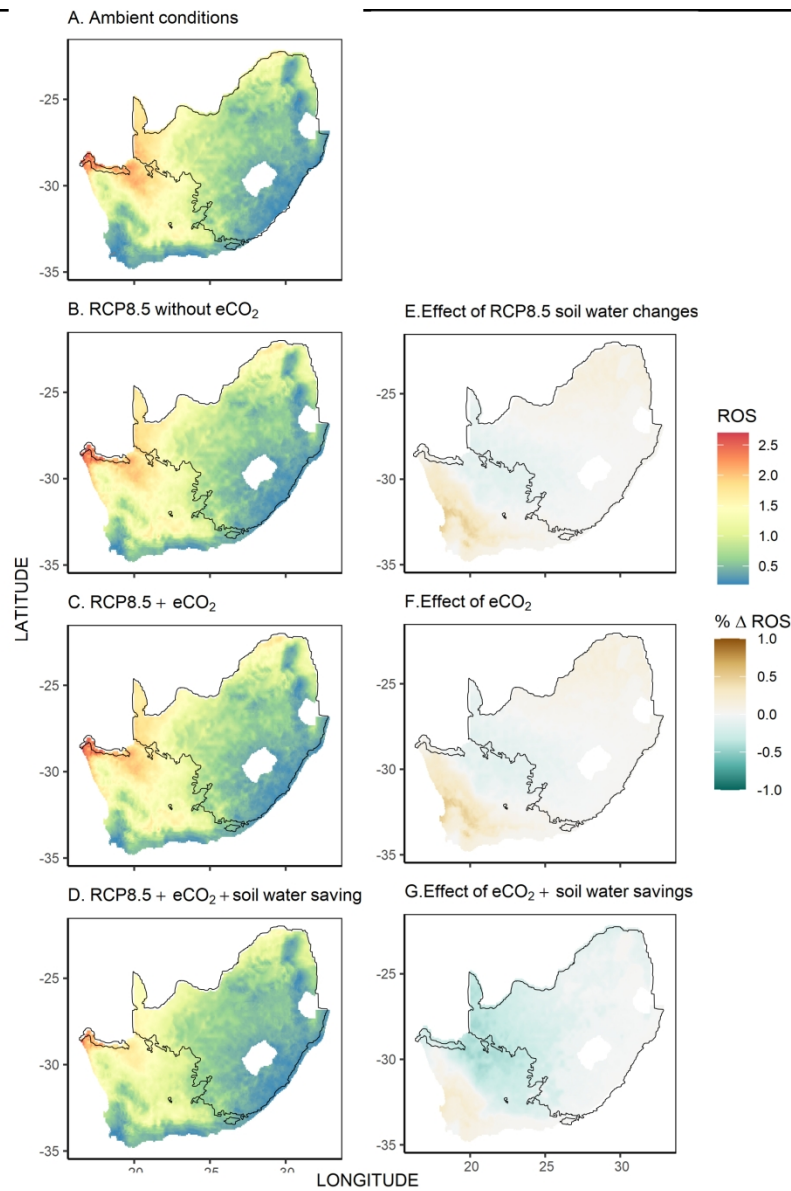
175x74mm (300 x 300 DPI)



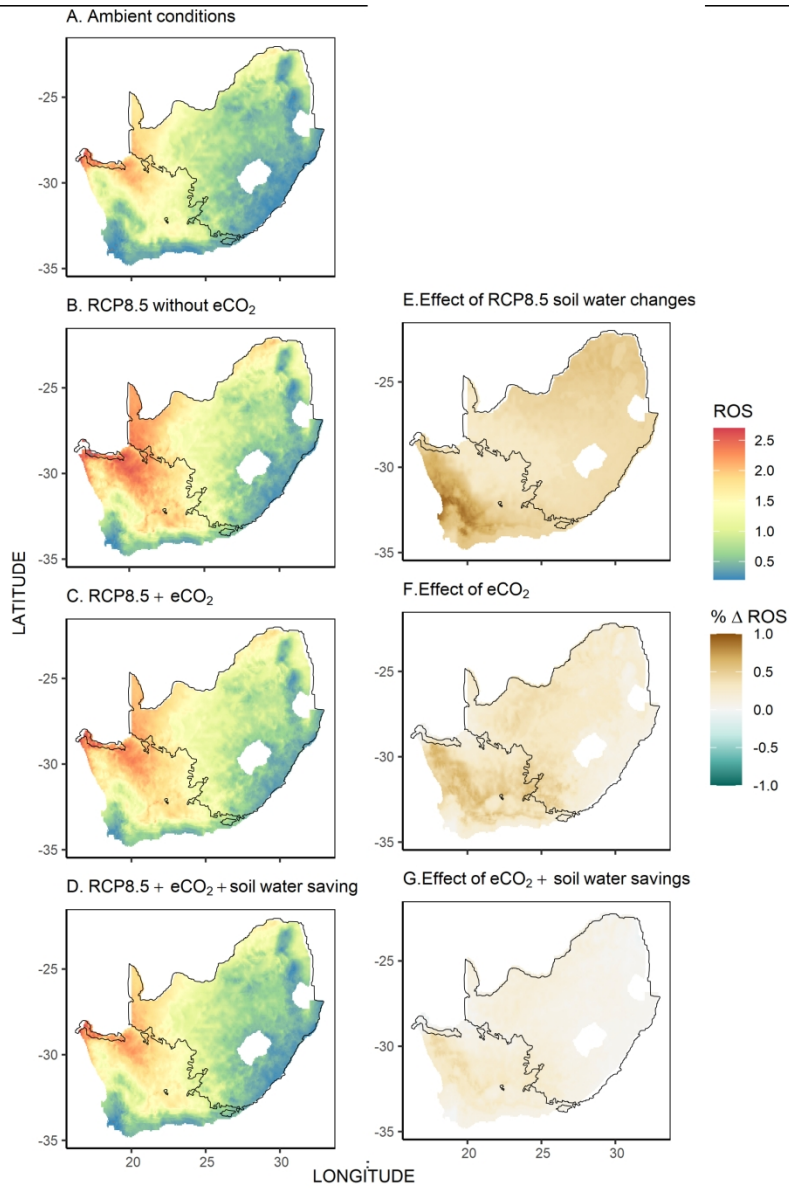
175x89mm (300 x 300 DPI)



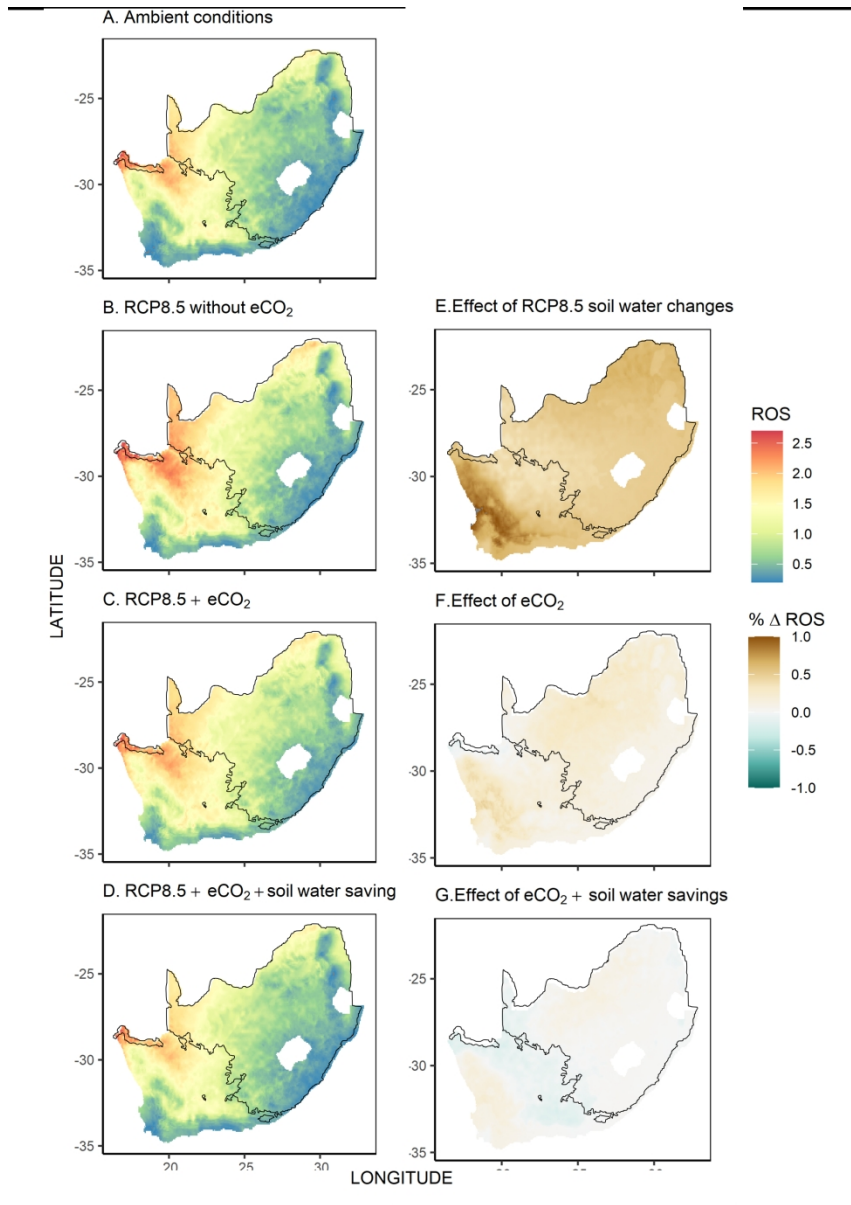
446x629mm (130 x 130 DPI)



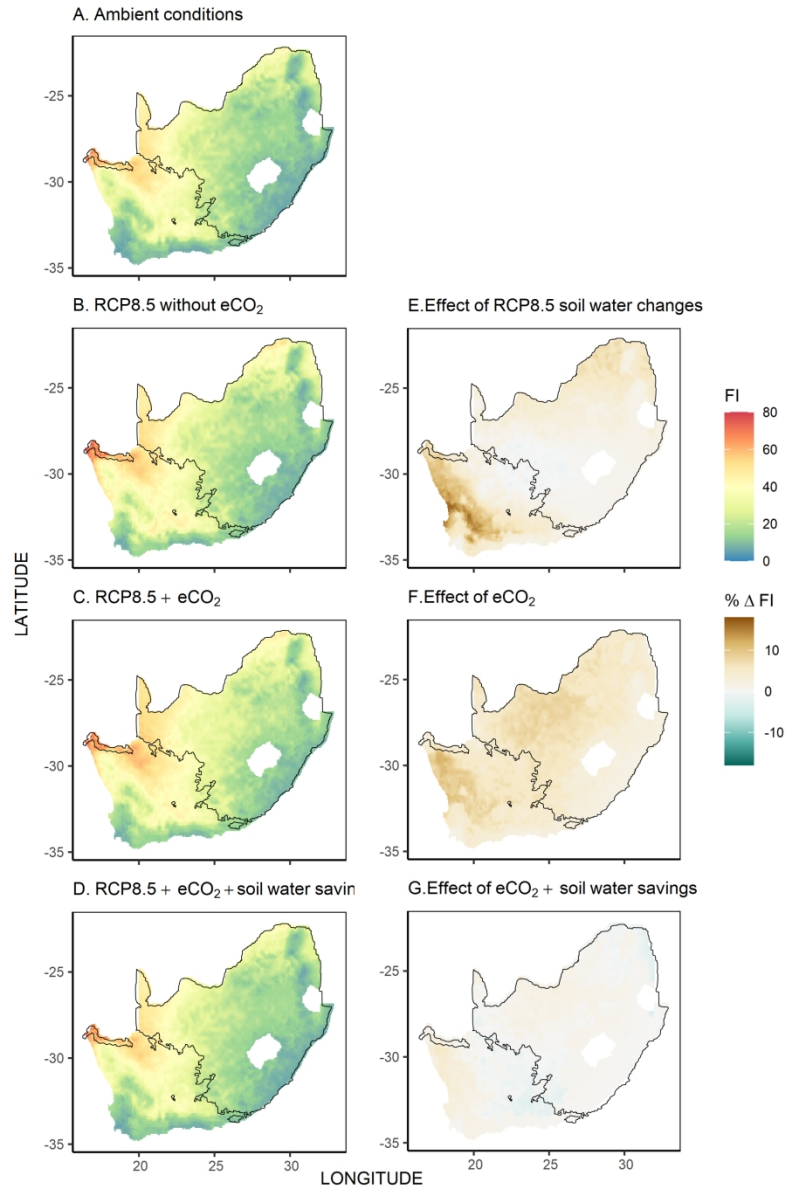
446x628mm (130 x 130 DPI)



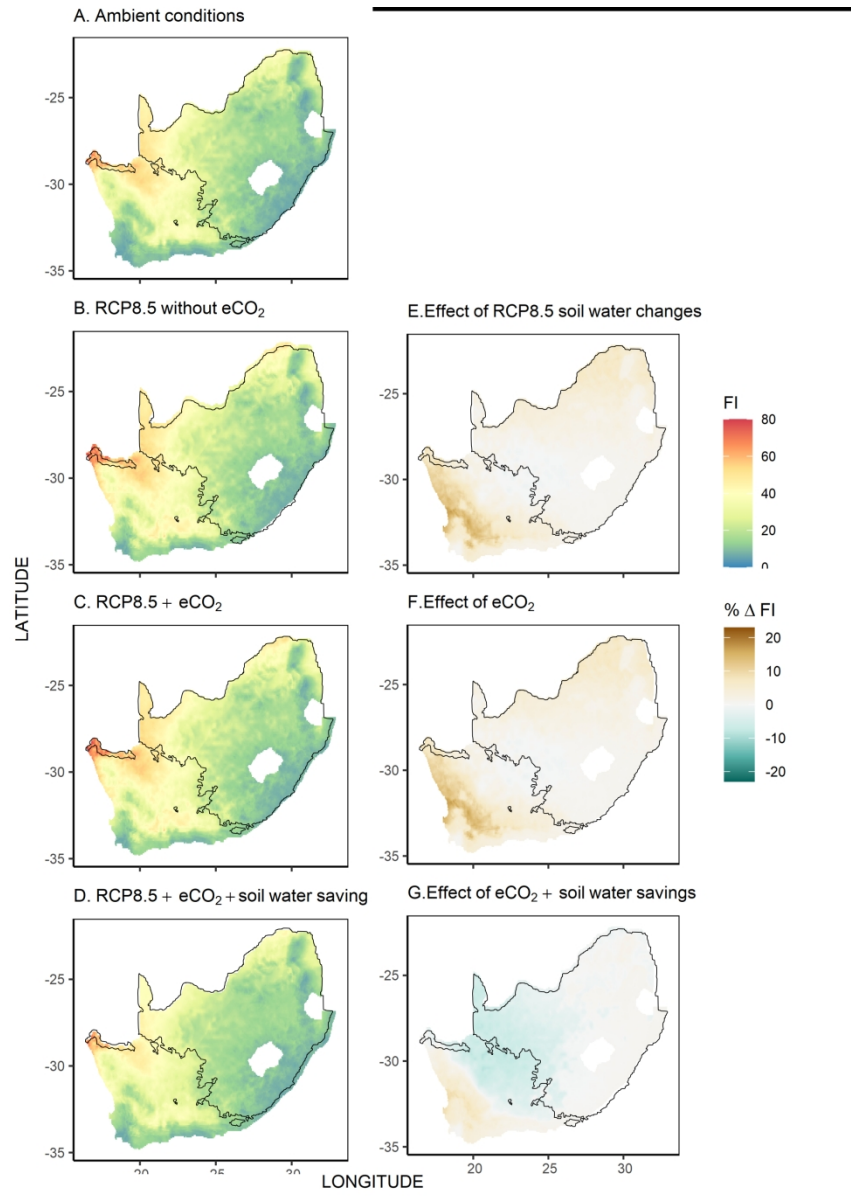
446x628mm (130 x 130 DPI)



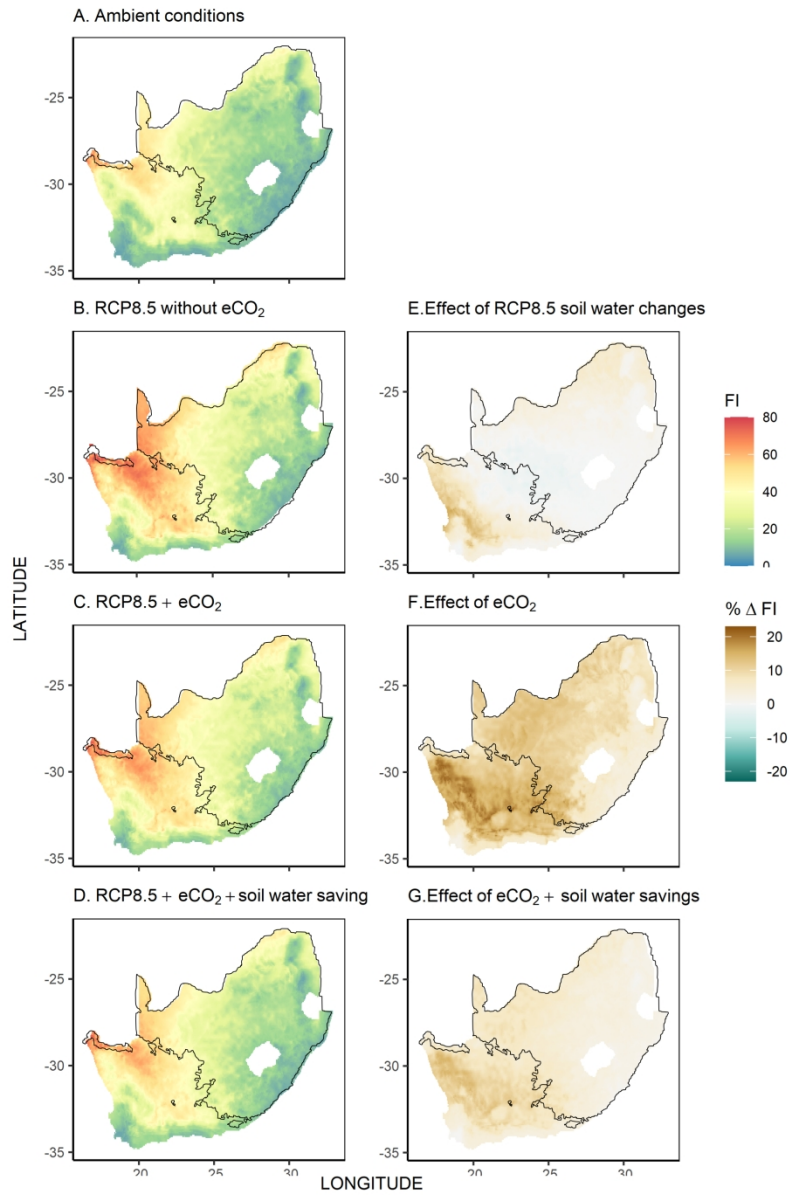
446x629mm (130 x 130 DPI)



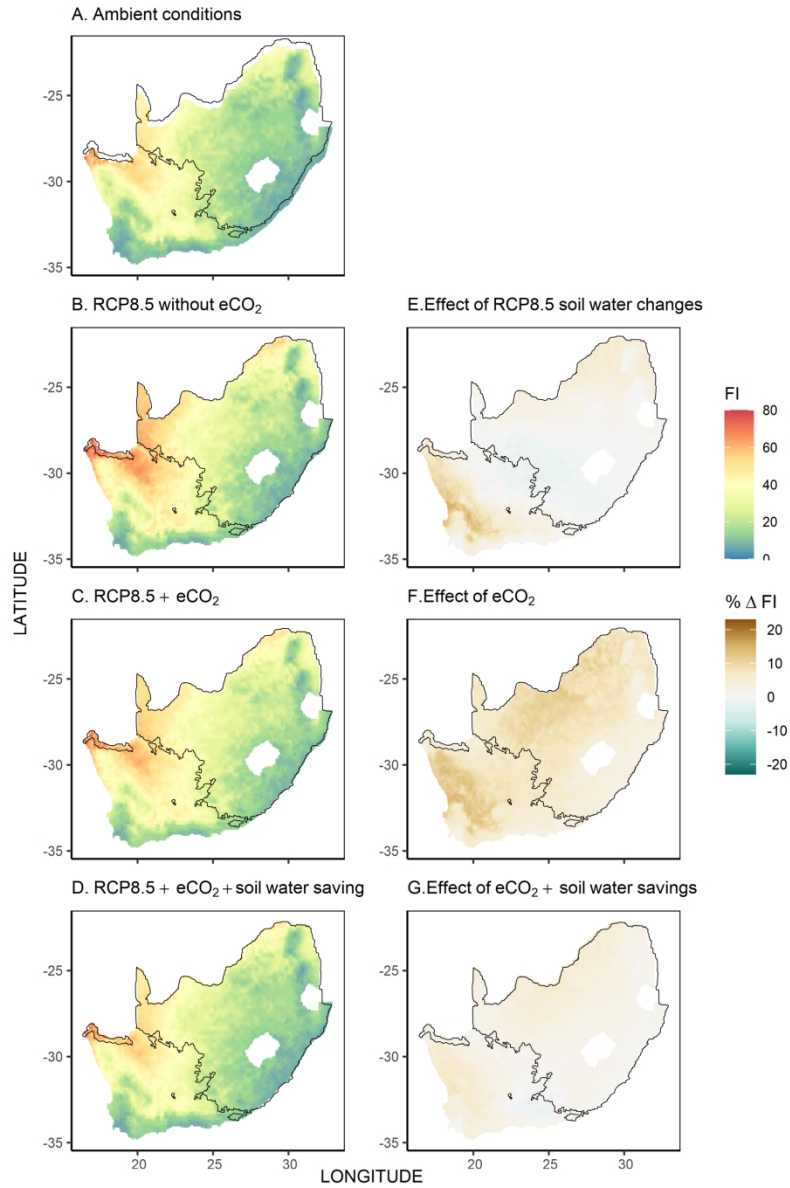
446x627mm (130 x 130 DPI)



447x630mm (130 x 130 DPI)



446x627mm (130 x 130 DPI)



446x627mm (130 x 130 DPI)

Target CO₂ (ppm)	% within S.D.	% within 50 ppm	% within 100 ppm
400	72.82	92.85	99.66
800	70.89	38.42	66.09

For Peer Review Only

Environmental variable	Mean	Minimum	Maximum
Temperature (°C)	19.25	0	42.92
Relative Humidity (%)	77.93	5.12	100.00
Daily precipitation (mm)	0.87		
Monthly precipitation (mm)	18.53		
Total accumulated precipitation (mm)	111.2		
Solar inclination (W/m ²)	663.6		
Solar azimuth (°)	179.96		
Solar elevation (°)	4.94		
Solar hour angle (°)	-0.0002		
Solar declination (°)	-0.12		
Solar air mass	5.42		