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Eukaryotes in soil aggregates across conservation managements: Major roles of protists, fungi and taxa linkages in soil structuring and C stock

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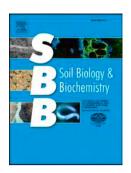
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- 1 Eukaryotes in soil aggregates across conservation managements: major roles of protists, fungi
- 2 and taxa linkages in soil structuring and C stock

4 Running title: eukaryotic diversity and roles in soil aggregates

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13 Abstract

The stabilization of soil organic carbon (SOC) promoted by conservation agriculture (CA) depends on soil aggregation. Aggregation protects SOC and creates heterogeneous microhabitats hosting diverse soil biota which in turn promote aggregation. A long-term experiment, studying the interaction of tillage with nitrogen (N) fertilization on a soybean-wheat rotation, was used to investigate eukaryotic community diversity, composition, and structure within small macroaggregates (sM) and occluded microaggregates (mM). Using high-throughput Illumina sequencing, we found (i) a different eukaryote diversity response to management intensification across soil aggregates and soil depths; (ii) a conserved core community composition of eukaryotes across CA treatments and aggregates at surface and subsurface layers; (iii) a different effect of tillage on eukaryotic community structure in sM and mM along the soil profile according to N availability; (iv) a positive association of protists, and fungi with the amount of sM and mM, and their SOC content; (v) a stronger complexity of within- and cross-domain networks (eukaryotes and eukaryotes-prokaryotes) in mM than in sM at surface layer. Overall, our findings demonstrated for

- 27 the first time that protists together with fungi play major roles in soil structuring and C cycling, and
- that Cercozoa represent hubs in soil biota aggregate networks.

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- 30 Keywords
- 31 soil aggregates; soil eukaryotes; protists; Illumina sequencing; 18S rRNA gene amplicon; soil biota
- 32 linkages.

1. Introduction

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Soil organic carbon (SOC) stability mainly depends on physical protection (Six et al., 2000; Six and Paustian, 2014), whereas molecular structure of plant residues and root exudates play a secondary role in SOC persistence (Schmidt et al., 2011; Lal et al., 2015). Organic carbon (C) is protected in soil aggregates by physically limiting the access of decomposers and enzymes and the diffusion of O_2 . According to the model of Tisdall and Oades (1982), primary particles (clay and silt particles, Ø <53 µm) are bound together by persistent bacterial, fungal, and plant debris into free microaggregates (\emptyset 53-250 µm). Free microaggregates are bound into macroaggregates (\emptyset > 250 um) by transient agents (i.e., microbial and plant polysaccharides) that are rapidly decomposed by microorganisms, and by temporary agents (i.e., roots, fungal hyphae and glomalin) that persist in the medium term. Labile SOC is mainly located in macroaggregates, while free microaggregates contain a more recalcitrant SOC pool (Elliott, 1986; Jastrow and Miller, 1998). Intensive agricultural practices, such as tillage and fertilization, shorten the life cycle of macroaggregates and diminish the formation rate of new microaggregates, worsening soil structure (Six et al., 2000). In no-tillage systems (NT), the slower turn-over of macroaggregates resulted in more sequestration of crop-derived C in microaggregates formed within macroaggregates (occluded microaggregates, mM; Ø 53-250 µm), and thus the amount of mM is crucial for the long-term Csequestration in soils (Six et al., 2000; Denef et al., 2007; Sheehy et al., 2015). In this context, the application of conservation agriculture (CA) practices (i.e., minimum tillage/NT, crop rotation and mulching) may allow the establishment of microhabitats with variable nutrient availabilities for a diverse soil biota, acting as efficient binding agent (Kong et al., 2011; Gupta and Germida, 2015; Totsche et al., 2018; Piazza et al., 2019). Moreover, CA practices may also produce yields equivalent to or even greater than conventional systems (Rusinamhodzi et al., 2011; Aune, 2012; Pittelkow et al., 2015; Himmelstein et al., 2016).

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In boreal climates, long-term NT and minimum tillage (MT) have been shown to increase the amount of macroaggregates and mM as well as their SOC content in the shallow layer (surface soil within horizon A) in comparison to conventional tillage (CT) (Franzluebbers and Arshad, 1997; Sheehy et al., 2015). This was demonstrated under various soil textures and was more evident in clay, clay-loam and silt-loam soils. Similarly, in humid tropical climates and sandy loam soils, long-term application of MT significantly increased SOC content in large soil aggregates, whereas the reverse was reported under CT (Onweremadu et al., 2007). Accordingly, Denef et al. (2007) highlighted a promotion of mM fraction and mM-associated C stocks in NT and MT compared with CT under similar climate. Moreover, nitrogen (N) fertilization was reported to increase SOC in macroaggregates and free microaggregates by decreasing the activity of cellulolytic fungi and bacteria (Ghosh et al., 2019; Duan et al., 2021). Recently, in a cold and humid Mediterranean area and in a silt-loam soil, high N fertilization rates in combination with MT not only increased mM, but also promoted a shift to low level, but more efficient C-cycling microbial enzyme activities, which were correlated to a greater accumulation of SOC (Piazza et al., 2020). Overall, in four regions across Europe the intensification of agriculture was reported to consistently reduce soil biota diversity in bulk soil, making soil food webs less diverse and composed of smaller bodied organisms (Tsiafouli et al., 2015). Although the role of bacterial and fungal communities (including arbuscular mycorrhizal fungi, AMF) in soil aggregation and SOC stabilization is widely recognized to be fundamental (Six et al., 2004; Lehmann et al., 2017; Bach et al., 2018), the diversity and potential role of other soil biota have received less attention. Soil biota diversity has indeed proved to be the major driver of C sequestration and nutrient cycling in bulk soil (De Vries et al., 2013; Wagg et al., 2019; Delgado-Baquerizo et al., 2020). Many studies demonstrated that earthworms and bacterivore nematodes are directly involved in the formation of macroaggregates by incorporating fresh organic matter inside mM and thus promoting SOC accumulation (Six et al., 2004; Pulleman et al., 2005; Bossuyt et al., 2006; Fonte et al., 2007; Zhang et al., 2013; Delgado-Baquerizo et al., 2020). Moreover, an indirect

effect on SOC accumulation by earthworms and bactrivorous nematodes was also reported and explained by the shift of soil microbial diversity through taxa regulating nutrient flow (Delgado-Baquerizo et al., 2020). Thus, in this study, we investigate the diversity and related roles of the eukaryotic component of soil biota within small macroaggregates (sM) and mM across CA managements. Moreover, since a more connected soil biota network takes up more C (Morriën et al., 2017), the study was extended to elucidate how eukaryotes are connected among each other, and to the prokaryotic community. In this context, long-term CA field experiments in the Mediterranean area, such as the one used in this study, provide a great opportunity for improving the understanding of soil eukaryotic diversity and functionality in soil aggregates and C stocks.

The following hypotheses were tested: (1) conservation tillage and N fertilization shift soil eukaryote community diversity, composition and structure, in soil aggregates along the soil profile; (2) soil aggregates differentially shape the diversity, composition and structure of soil eukaryotes; (3) some eukaryotic taxa are predictors for soil structuring and C stocks; (4) eukaryotes form structured assemblages and distinctive networks in soil aggregates (within-domain networks); (5) the traits of the eukaryotes-prokaryotes networks vary across aggregates (cross-domain networks), and some network traits can predict soil structuring and C stocks.

2. Materials and Methods

102 2.1. Field experiment

A long-term CA field experiment on a bread wheat (*Triticum aestivum* L.) - soybean (*Glycine max* L. Merr.) rotation was set up in 1993 at the Centro Interdipartimentale di Ricerche Agro-Ambientali Enrico Avanzi (Pisa, Italy; 43°40' latitude N; 10° 19' longitude E; 1 m above sea level) in an alluvial silt loam soil (131, 613 and 256 g kg⁻¹ of sand, silt and clay, respectively). The experiment was conducted comparing two tillage intensities and two N fertilization levels. The tillage intensities were: conservation tillage (minimum tillage, MT: disk harrowing at 15-cm depth)

and conventional tillage (CT: mouldboard ploughing at 25-cm depth, disking and harrowing at 15-cm depth). The N fertilization levels applied only to bread wheat were: 0 and 200 kg N ha⁻¹ (N0 and N200, respectively). The soil is classified as Typic Xerofluvent by USDA system (Soil Survey Staff, 1975) and as Fluvisol by FAO (IUSS, 2006). Climate of the site is cold, humid Mediterranean (Csa), according to the Köppen-Geiger climate classification (Kottek et al., 2006). The experiment was arranged following a split-plot design, with tillage as main-plot factor and N fertilization as subplot factor and three replicate plots (dimension: 11.5 x 14.5 m). The N fertilizer treatment was applied as urea and the rate was split into three applications, before sowing (60 kg N ha⁻¹), at the first detectable node (70 kg N ha⁻¹), and 15 days after this stage (70 kg N ha⁻¹). Under CT, almost 100% of the residues were incorporated in the 0-25 cm soil layer, whereas under MT approximately 50% of the crop residues were incorporated at 0-15 cm depth. Crops were managed applying preemergence herbicide for weed control and no disease or insect treatments (Piazza et al., 2020).

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2.2. Soil sampling and analysis of soil physical and chemical parameters

Soil sampling was carried out in Spring 2016 before soybean sowing. In each replicate plot, a homogenized sample was obtained by mixing four soil cores collected at two soil depths (surface layer: 0-15 cm; sub-surface layer: 15-30 cm). A total of twenty-four soil samples were collected (12 at the surface layer; 12 at the subsurface layer). Once in the laboratory, each sample was air-dried, gently broken apart and then passed through an 8-mm sieve. The isolation of small macroaggregates (sM; 250-2000 µm) was done from 80 g of the sieved soil samples by the wet sieving method (Six et al., 1999). Occluded microaggregates (mM; 53-250 µm) were isolated from an additional isolation of sM (i.e., starting from 80 g of the sieved soil samples) and utilising a device designed and built by Piazza et al. (2020). Once collected, the fractions were freeze-dried (FreeZone 2.5 Labconco, Kansas City, MO, USA) for 48-72 h for dry weight determination and chemical and molecular analyses. Both aggregate fractions of all samples were then analysed for SOC by CHN

134	combustion method (LECO, Italy) and SOC content was calculated and expressed in Mg ha-
135	Bremner and Mulvaney, 1982; Piazza et al., 2020).

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- 2.3. Molecular analyses
- 138 DNA was extracted from 0.25 g of sM (n=24) and mM samples (n=24) using the DNeasy 139 PowerSoil Kit (QIAGEN, Venlo, Netherlands). The DNA extracts were then quantified by a 140 spectrophotometer (NanoDrop Technology, Wilmington, DE) and stored at -20 °C. PCRs were generated from 10 ng μL⁻¹ genomic DNA in volumes of 25 μL with 0.125 U μL⁻¹ of GoTag® Hot 141 Start Polymerase (Promega Corporation, WA, USA), 0.5 µM of each primer, 0.2 mM of each 142 dNTP, 1 mM of MgCl₂ and 1x reaction buffer, using the PTC-200 96-well Peltier Thermal Cycler 143 144 (MJ Research, MA, USA). The primers were TAReuk454FWD1-ill (5'-145 TCGTCGGCAGCGTCAGATGTGTATAAGAGACAGANNNHNNNWNNNHCCAGCASCYGC 146 GGTAATTCC-3') TAReukREV3-ill (5'and 147 GTCTCGTGGGCTCGGAGATGTGTATAAGAGACAGTACTTTCGTTCTTGATYRA-3') (modified from Stoeck et al., 2010). The primer pair has attached Illumina sequencing tags, and for 148 149 the forward primer a 13 bp random sequence was included in order to improve cluster definition on 150 the MiSeq slide. Primers target the hypervariable region V4 of the small subunit ribosomal RNA 151 (SSU rRNA 18S) gene fragment. The thermal cycler was programmed as follows: 95 °C for 2 min, 152 35 cycles at 94 °C for 30 s, 50 °C for 45 sec, 72 °C for 1 min and 30 s and a final extension step at 72 °C for 10 min. PCR products were examined by electrophoresis through a 1% agarose gel in 0.5 153 154 × TBE buffer, then purified with magnetic beads (Agencourt® AMPure® XP, Beckman Coulter, 155 USA) and freshly prepared 80% ethanol, and quantified by fluorimetry with the use of Quant-iTTM 156 dsDNA HS (High-Sensitivity) Assay Kit (Invitrogen by Thermo Fisher Scientific, CA, USA), 157 following the instructions of the manufacturer. Cleaned and quantified amplicons of each library 158 were adjusted in an equimolar ratio (10 ng/µL) for the required Illumina P5 and P7 sequences

addition along with index sequences in a new PCR step. Indexing was performed using primers from the Nextera® Index kit (sets A and D; Illumina Inc., CA, USA) and the resulting metabarcoding libraries were sequenced on an Illumina MiSeq sequencer (2 * 300 bp paired-end reads) at the Genomics and Bioinformatics Laboratory (Technology Facility, Department of Biology, University of York, UK). Details are given in the Supplementary Methods 1.

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2.4. Bioinformatic analyses

Raw data generated from the Illumina MiSeq sequencing run were processed and analyzed following the pipelines of QIIME 2 (2018.4) and USEARCH (v10.0.240) (Edgar, 2010; Caporaso et al., 2012). Forward and reverse paired-end sequences were assembled independently for each sample using -fastq_mergepairs USEARCH command. Primer sequences were then trimmed off by employing cutadapt plugin (2018.4) with default settings. To avoid potential errors in sequencing data, quality of sequence reads was checked by -fastq_eestats2 USEARCH command, using the expected number of errors in a read as a measure of quality for filtering (Edgar and Flyvbjerg, 2015). Reads were then trimmed at the length where the "drop-off" point for the maximum expected error value occurred (250 bp). Quality filtered reads were de-replicated by -fastx_uniques USEARCH command, then Operational Taxonomic Units (OTUs) were generated using USEARCH by clustering sequence reads at the 97% similarity threshold. During the process, chimeric sequences and singletons were removed from the dataset. For the curation, the sequences were aligned using ClustalW and then Neighbor Joining (NJ) phylogenetic tree was built in MEGA7 (Kumar et al., 2016) (https://www.megasoftware.net). The most abundant sequence of the eukaryotic OTU in each cluster was selected, and used as representative sequence for that OTU after branch collapsing. For the curation, the sequences were aligned using ClustalW and then Neighbor Joining (NJ) phylogenetic tree was built in MEGA7 (Kumar et al., 2016) (https://www.megasoftware.net). The most abundant sequence of the eukaryotic OTU in each

cluster was selected, and used as representative sequence for that OTU after branch collapsing. The OTUs were phylogenetically assigned using the 18S SSU SILVA database (version 132, release date 13.12.2017) (Quast et al., 2012; Yilmaz et al., 2013) by clustering sequence reads at the 97% similarity threshold. After curation, the representative sequences were re-aligned using ClustalW and the phylogenetic tree was built in MEGA7 using the Neighbor Joining (NJ) analysis with 1 000 boostrap replicates and the Kimura 2-parameter model (uniform rates).

The suitability of the eukaryotic community sampling was verified by rarefaction curves plotting the number of eukaryotic classes/phyla *versus* the number of sequence reads, while accumulation curves were calculated plotting the number of classes/phyla *versus* the number of soil samples using the package Vegan in R (Oksanen et al., 2013). Since there was a high variability in the number of reads per sample, sequencing depth per sample was standardized to the median number of reads across the samples in each data matrix using the same package in R (standardized datasets). All representative sequences were deposited in the NCBI GenBank database (SUB5948379 submission: MN178662-MN178794 accession numbers).

Prokaryotic data were obtained from the same soil matrices and depths (Piazza, 2019) and were based on the V4 region of the 16S rRNA gene sequenced using a MiSeq Illumina approach (SUB5941754 submission: MN171543-MN172157 accession numbers).

2.5. Statistical analyses

To test hypothesis 1 - Conservation tillage and N fertilization shift soil eukaryote community diversity, composition and structure in soil aggregates along soil profile - analysis of variance (ANOVA), Venn diagrams and permutational analysis of variance (PERMANOVA) were applied. Concerning diversity, richness, Shannon index (H') and Simpson index ($\lambda = 1 - \lambda'$) were calculated at class level and analysed by two-way ANOVA, according to the experimental design. These analyses were done in Vegan package in R and plotted by ggplot2 (Wickham and Chang, 2008).

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Data were In--transformed when needed to fulfil the assumptions of ANOVA. Post-hoc Tukey-B significant difference test was used for comparison among treatments. Concerning composition, Venn diagrams were drawn to visualize the OTUs unique to the treatments as well as the shared ones. The standardized datasets were used to generate the Venn diagrams by the online tool InteractiVenn (http://www.interactivenn.net; Heberle et al., 2015). Concerning structure, the relative abundances of eukaryotes were calculated at class level and the permutational analysis of variance (PERMANOVA) (Anderson and Braak, 2003) and the analysis of homogeneity of multivariate dispersion (PERMDISP) were used to test the effect of treatments (Togerson, 1958; Clarke and Gorley, 2006). Response data were square-root transformed to down-weight the dominant taxa and the Bray-Curtis index of dissimilarity was used to measure ecological distance. When PERMANOVA indicated a significant effect, the principal coordinate analysis (PCO) was carried out (Anderson et al., 2008) to visualize the most relevant patterns in the response data. In each PCO biplot, only the taxa with a strong correlation (r = 0.50-0.80) with the ordination scores on each PCO axis were displayed. P-values were calculated using the Monte-Carlo test and residuals were permuted according to the experimental model (Oksanen et al., 2013). Multivariate analyses were performed using PRIMER 6 and PERMANOVA+ software (Clarke and Gorley, 2006; Anderson et al., 2008). To test hypothesis 2 - Soil aggregates differentially shape the diversity, composition and structure of soil eukaryotes - eukaryotic richness, H' and λ at phylum level and at both soil depths were analysed by one-way ANOVA, using soil matrix (sM vs mM) as fixed factor, and tillage and N fertilization as covariates. Analyses were performed in Vegan package in R and data were plotted by ggplot2. Data were ln-transformed when needed to fulfil the assumptions of ANOVA, and the post-hoc Tukey-B significant difference test was used for comparison among treatments. Moreover, the effect of matrix on composition and structure were analysed at phylum level using the Venn diagrams and PERMANOVAs, as described above.

234 To test hypothesis 3 - Some eukaryotic taxa are potential predictors for soil structuring and C 235 stocks - multiple regression analysis was applied using as independent variables the standardized 236 relative abundances (calculated as described above) of eukaryotic taxa at class level. To account for 237 the compositional nature of the data, an additive log-ratio transformation was applied (Gloor et al., 238 2017). The dependent variables were sM and mM weights, and SOC content per unit of surface in 239 sM and mM. The assumptions of the linear regression model were verified (Supplementary Method 240 2) and the multiple linear regression analysis was applied using a stepwise method with the 241 following probability criteria: P < 0.05 to accept and of P > 0.05 to remove a phylum or a 242 within/cross-domain network traits. Multiple regressions were performed using the SPSS software 243 package version 25.0 (SPSS Inc., Chicago, IL, United States). Details about regression analysis are 244 reported in Supplementary Methods 2. To test hypothesis 4 - Eukaryotes form structured assemblages and distinctive networks in soil 245 246 aggregates - we built networks using the SParse InversE Covariance estimation for Ecological 247 **ASsociation** Inference (SPIEC-EASI) 0.1 R package version in 248 (https://github.com/zdk123/SpiecEasi/). SPIEC-EASI is a pipeline for inferring sparse inverse 249 covariance matrix within and between multiple compositional datasets, under joint sparsity penalty 250 (Kurtz et al., 2015). The within-domain analyses were performed on the standardized eukaryotic 251 dataset at class level for each soil matrix and depth. The neighborhood selection (MB method) was 252 applied as graphical inference model (Meinshausen and Bühlmann, 2010), since it has been shown 253 to better perform than other available methods (e.g., CCREPE, SPARCC, SPIEC-EASI glasso) 254 (Kurtz et al., 2015). The Stability Approach to Regularization Selection (StARS) was applied to 255 select the optimal sparsity parameter (Liu et al., 2010), and the StARS variability threshold was set 256 to 0.05 and $n\lambda$ to 100 for all networks. We evaluated the weights of the edges in the networks using 257 SPIEC-EASI (frequency *versus* edge weights = modularity), and we plotted the degree distributions 258 of frequencies of the edges using adj2igraph (Kurtz et al., 2015). In the networks a node represents 259 a connected taxon, an edge the connection between taxa, a singleton an unconnected taxon and a

dyad two connected taxa. For the eukaryotic networks (within-domain network) we calculated the following parameters: number of nodes excluding singletons, number of edges, number of singletons and dyads, number of subnetworks (a subnetwork is a network composed by at least three nodes), mean nodes per subnetwork, linkage density (complexity: the average number of edges per node), percentage of positive interactions and modularity. Moreover, in each network, we calculated the frequency of the phyla within the subnetworks, the mean of edges and nodes, and the percentage of positive edges for each phylum. Details about trait calculations are given in Supplementary Methods 3.

To test hypothesis 5 - The traits of the eukaryotes-prokaryotes networks vary across aggregates we inferred the associations between eukaryotes and prokaryotes domains by the cross-domain extension of SPIEC-EASI (Kurtz et al. 2015; Tripton et al., 2018). The same traits calculated for within-domain networks, except for the positive edges, were assessed (Supplementary Methods 3). In addition, the percentage of eukaryotes/prokaryotes per subnetwork was calculated and the number of subnetworks with only eukaryotes, only prokaryotes and with both domains were counted. The significance of the cross-domain relationships was tested by the Mantel test (Mantel and Valand, 1970) on the standardized read data that were centered and normalized and using the function Jaccard in PC-ORD 5 to build the resemblance matrix (Grandin, 2006) and also by the co-Correspondence Analysis (CoCA) (ter Braak and Schaffers, 2004) in CANOCO 5 (ter Braak and Smilauer, 2012). To test if some network traits can predict soil structuring and C stocks - a multiple linear regression was performed after verification of the assumptions (Supplementary Methods 3). The analysis was performed using as independent variables the log(1+x)-transformed and normalized within- and cross-domain network traits in sM and mM (i.e., traits of eukaryotic networks and of eukaryotic-prokaryotic networks), and using as dependent variables sM and mM weights, and SOC content per unit of surface in sM and mM. Scripts for within and cross-domain network construction and analysis are available in Supplementary Methods 4.

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3. Results and discussion

3.1. Illumina sequencing information

MiSeq sequencing yielded a total number of 2 940 322 reads from the 48 soil samples, and following quality-filtering a total of 2 884 052 sequence reads having a length of 392 bp were obtained. After BLAST against the 18S SSU SILVA database (Quast et al., 2012; Yilmaz et al., 2013), we found 2 036 277 reads, ranging from 3 to 106 540 reads per sample that were assigned to a total of 4 211 OTUs. After sequence curation and the removal of Plantae sequences, 863 809 reads, ranging from 6 421 and 46 429 reads per sample, were retrieved and assigned to 133 OTUs, 56 classes and 27 phyla (Fig. S1, Fig. S2). The rarefaction and accumulation curves demonstrated that sampling effort was sufficient as the curves reached the asymptote (Fig. S3, Fig. S4).

3.2. Effect of conservation management on eukaryotic diversity in soil aggregates

To test if conservation tillage and N fertilization shift the diversity of soil eukaryotes in soil aggregates (hypothesis 1), richness and diversity indices were determined along the soil profile in small macroaggregates (sM) and occluded microaggregates (mM). A greater eukaryotic diversity in sM was consistently found at both soil layers under CT compared to MT, as shown by the significant increase of richness (+26%) and H' and λ (+9%) (Fig. S5). This higher eukaryotic diversity might be due to larger root development and higher availability of root exudates, organic matter (e.g., nutrients, organic acids), water and oxygen, reported under deep ploughing systems, and which have been shown to promote microbial growth and soil biota diversity/functionality (Guan et al., 2014; Edwards et al., 2015; Ercoli et al., 2017; Piazza et al., 2020), according to the response of individual taxonomic units to habitat and trophic conditions (van Capelle et al., 2012). At surface layer, N fertilization significantly increased λ under MT, suggesting an increase in number of relative abundances of taxa regulated by N availability. Conversely, under CT, λ was

high and not modified by N fertilization, suggesting non-limiting N availability due to improved
plant growth and higher mineralization rate of residues. A low eukaryotic diversity was previously
found in bulk soil under N fertilization (Lentendu et al., 2014), and within soil aggregates a higher
microbial diversity was found at low nutrient availability and NT compared with high nutrient
availability and ploughing (Lagomarsino et al., 2012; Zhang et al., 2013; Bach et al., 2018).

In mM an opposite pattern was found at surface layer, and no effect of N fertilization alone or in interaction with tillage was observed at both soil depths (Fig. S5). At surface layer, the eukaryotic diversity indices increased by 6% under MT compared to CT. Under tillage intensification, macroaggregates are indeed disrupted and occluded microaggregates became free in the soil (Six et al., 2000), potentially reducing diversity and C sequestration. This is supported by the highest SOC accumulation observed in mM under MT (Piazza et al., 2020), and by the high soil biota diversity found in the present study within mM.

To test if soil aggregates differentially shape the diversity of soil eukaryotes (hypothesis 2), richness, H' and λ were determined in sM and mM. Overall eukaryotic diversity was significantly higher in mM than in sM (Fig. S5) (i.e., at surface layer: richness +16%, H' +6%, and λ +2%; at subsurface layer: richness +20 and H' +5%). These results are in accordance with the higher richness and H' of bacteria and fungi found in free microaggregates compared to large macroaggregates (Bach et al., 2018). Our findings support the fact that soil aggregates are distinct habitat spaces with eukaryotes adapted to SOM resources, pore-space network, and water and oxygen availability characteristic of sM and mM.

3.3. Effect of conservation management on eukaryotic composition in soil aggregates

To test if conservation tillage and N fertilization shift the composition of soil eukaryotes in soil aggregates (hypothesis 1) and how soil aggregates shape their composition (hypothesis 2), this parameter was evaluated along the soil profile in sM and mM. Across management practices and soil depths, the eukaryotic phyla Cercozoa (21%), Ciliophora (13%), Chlorophyta (11%), Nematoda

(11%) and Glomeromycota (9%) were predominant in sM, whereas in mM the predominant phyla were Ciliophora (19%), Cercozoa (18%) Chlorophyta (15%), and Ascomycota (11%) (Fig. 1a). The other phyla showed an abundance ≤ 8%. Sun et al. (2021) found that protists were the most dominant eukaryote (33.9% of the total eukaryotic sequences) in bulk soil. By contrast, Treonis et al. (2018) analysing the whole eukaryotic structure in bulk soil found a high abundance of fungi, Arthropoda, Nematoda and Anellida (40, 20, 20 and 11%, respectively), and a low abundance of protists (0.63%). Among protists, Rhizaria was the group with the highest relative abundance in arable soil, comprising as dominant taxa Cercozoa and Amoebozoa (Bates et al., 2013; Degrune et al., 2019a; Santos et al., 2020). Similarly, in another study, fungi were reported to be the most abundant (i.e., Ascomycota, Basidiomycota, fungi Incertae sedis and Glomeromycota), followed by Alveolata, Metazoa, Rhizaria, Stramenopiles, and Viridiplantae (Chen et al., 2012). Therefore, we can assume that the differences between bulk soil and soil aggregates depend on the variability of pH, moisture and organic nutrient availability that shift soil biota at multiple trophic levels.

In both soil aggregates, the majority of taxa were common to all managements across soil

In both soil aggregates, the majority of taxa were common to all managements across soil depths, whereas some were unique to certain managements, as shown by the Venn diagrams at class resolution (Fig. 2b-c,e-f). This is also shown by the pie charts representing the proportion of the 56 classes retrieved in each management (Fig. 2a,d). Moreover, focusing on the shared taxa between sM and mM, ca. 80% of eukaryotes were common to both soil aggregates, averaging soil depths (Fig. 3a,b). Accordingly, a large conserved core community of soil prokaryotes and fungi was found across managements at the same site in bulk soil (Piazza et al., 2019). This is also consistent with the findings obtained in other studies in different soil types and managements in bulk soil as well as in specific rhizocompartments (Lentendu et al., 2014; Edwards et al., 2015; Pershina et al., 2018).

However, the exclusive presence of some eukaryotic taxa in the different systems and soil aggregates (Fig. 2) suggests that long-term tillage and N fertilization may drive the development of communities of specialized taxa putatively having specific functions (e.g., soil aggregate and/or

SOC accumulation and nutrient cycling). As example, at the surface layer, in sM the classes *Nassophorea* and *Perkinsea* were exclusively found in MTN0, whereas *Pezizomycetes* and *Gastrotricha* in CTN0 (Fig. 2a,b), while many taxa were exclusively found in mM, such as Chlorophyta, *Eutardigrada*, *Eurotiomycetes*, *Nassophorea* and *Thecofilosea* in MTN0; Alveolata, *Chilopoda* and *Rhabditophora* in MTN200 and *Dictyostelia* and *Pezizomycetes* in CTN200 (Fig. 2d,e). A more in-depth description of the eukaryotic composition and exclusiveness across managements and aggregates is reported in the Supplementary Results and Discussion 1.

3.4. Effect of conservation management on eukaryotic community structures in soil aggregates

To test if conservation tillage and N fertilization shift the structure of soil eukaryotes in soil aggregates (hypothesis 1), the relative abundance pattern of taxa was determined in soil aggregates. Despite the high degree of similarity among treatments in term of community composition, we highlighted a strong effect of the interaction between tillage and N fertilization on the eukaryotic community structures in sM at both soil layers, and in mM only at the surface layer (Table 1, Fig. 1). Similarly, soil fungal community structure in bulk soil was strongly shaped by the interaction between tillage and N fertilization at surface and subsurface layers (Piazza et al., 2019). However, to our knowledge no studies have focused on the effect of the interaction of these practices on the eukaryotic communities in soil aggregates, whereas a huge number of studies was performed to assess the effect of tillage or N fertilization on the diversity/abundance and functionality of single eukaryotic group in bulk soil (e.g., fungi: Jansa et al., 2003; Wang et al., 2019; Zhao et al., 2019; micro-arthropods, nematodes and protozoa: Adl et al., 2006; Zhang et al., 2012; Briones and Schmidt, 2017; Cai et al., 2020; protists: Zhao et al., 2019; Sun et al., 2021).

In the PCO plots, CTN0 and CTN200 showed similar community structures within sM at surface layer (Fig. 1b) that were characterised by *Colpodea* and OTU1*Spirotrichea* (Ciliophora, Alveolata) and *Conoidasida* (Apicomplexa, Alveolata). This supports that under ploughing N

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availability is not limiting and the community structures are not affected by N fertilization. The class *Colpodea* is a well-known dominant clade of mainly bacterivorous protists (Foissner, 1998) and, according to our results, they were shown highly abundant in disturbed soils. By contrast, Spirotrichea were found in more stable environments (Lüftenegger et al., 1985), not supporting the large abundance found within sM under ploughed soils. This result not supported by literature may suggest a high resistance of Spirotrichea to natural and anthropogenic stresses. Moreover, Apicomplexa that are the third most abundant protistan group in soil, after Cercozoa and Ciliophora (Fierer, 2017) are described as putative parasites of invertebrates (Del Campo et al., 2019). This is in agreement with the lower occurrence of the invertebrates Chromadorea and Enoplea found in our study under CT compared with MT. In MT, N fertilization determined a strong shift at the surface layer (Fig. 1b), with community structure within sM under MTN200 characterized by a large abundance of taxa belonging to Cercozoa, *Tubulinea* (Lobosa, Amoebozoa) and *Chromadorea* (Nematoda) together with fungi, (i.e., OTU1Microbotryo: Microbotryomicetes, Ascomycota and Mortierellomycota), and under MTN0 characterized by a large abundance of taxa belonging to Imbricatea (Cercozoa), Enoplea (Nematoda) and OTU1Nassophorea (Ciliophora, Alveolata). The dominance of Ascomycota in sM under MTN200 is in accordance with their higher abundance in macroaggregates under mineral fertilization compared with no fertilization (Liao et al., 2018; Wang et al., 2021). This confirms the importance of fungi as binding agents in soil aggregates (Six et al., 2000). Cercozoa were reported to be affected by several abiotic factors, as soil moisture, clay content and N availability (Lentendu et al., 2014; Fiore-Donno et al., 2019). However, although our analyses did not allow discrimination of which cercozoan classes were favoured under MTN200 (Fig. 1b), the detection of *Imbricatea* within sM under MTN0 supports that N availability is a major driver of cercozoan communities. This class had unexpectedly high abundance in sM under MTN0 (Fig. 1b), although it was shown to be highly favoured by organic fertilizers (Lentendu et al., 2014). According to our results (Fig. 1b), the heterotroph lineage Tubulinea are dominant in highly Nfertilizer soils (Sun et al., 2021), and the ciliate Nassophorea, characterizing the community

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structures of sM under MTN0, supports their role in energy transfer between trophic levels under low N availability (Gao et al., 2016). Finally, the dominance in sM under MT of *Enoplea* and *Chromadorea* known to be plant and animal nematode parasites is consistent with their general trend in soil aggregates (Jiang et al., 2017). Their abundance in N0 and N200, respectively, might be explained by specific predator-prey interactions occurring within intra-aggregate pores at differential N availabilities.

At the subsurface layer (Fig. 1c), N fertilization drove a stronger shift of the eukaryotic community structure in sM under CT compared with MT that showed similar structures irrespective to N fertilization, highlighting an opposite pattern as compared with the one observed at surface layer (Fig. 1b). Under CTN0, sM was characterised by high abundance of *Chilopoda* (Arthropoda, Animalia) and Sarcomonadea (Cercozoa, Rhizaria), and under CTN200 by high abundance of Colpodea and Conoidasida (Alveolata) (Fig. 1c). Indeed, under ploughing, the subsurface layer is less compacted than under MT and shows a lower bulk density, resulting in an increase of pore size and aeration (Berisso et al., 2012). This may allow a larger root development under N fertilization and a higher variation in soil moisture and temperature compared with no fertilization (Piazza et al., 2020). Small macroaggregates are inaccessible to living centipedes (Chilopoda), thus their abundance in this fraction under CTN0 can be only related to a role as binding agents or as dead biomass consumed by decomposers. By contrast, Sarcomonadea, previously found in soil as the dominant class within the phylum Cercozoa (Degrune et al., 2019b), are likely to play an active role also in sM at low N availabilities. The abundance of Colpodea in CTN200 at the subsurface layer (Fig. 1c) is consistent with their dominance at the surface layer (Fig. 1b) and can be explained by bacterial pray changes following N application, while no information is available on Conoidasida trophic functional role.

In MTN0 and MTN200 at subsurface layer sM were characterized by *Chromadorea* (Nematoda), *Vampyrellidea* (Cercozoa), OTU1*Nassophorea* (Ciliophora, Alveolata) and many fungi (e.g., *Glomeromycetes*, *Sordariomycetes*) (Fig. 1c). The abundance of *Choromadorea* and

440 Nassophorea is consistent with surface layer observations. Moreover, Vampyrellidea, observed for 441 the first time in sM, are fungivores that may control the parasitic rust fungus of wheat under MT 442 (Adl and Gupta, 2006). Finally, the abundance of Glomeromycetes and Sordariomycetes confirms 443 their crucial role in driving soil aggregates under undisturbed conditions (Rillig et al., 2015; Wang 444 et al., 2021). 445 In mM, a strong interaction effect between tillage and N fertilization was found on the 446 eukaryotic community structures only at surface layer, consisting in a strong shift of the structure 447 under CTN0 compared with the other treatments (Table 1, Fig. 1d). This effect is in line with the 448 aggregate distribution of mM found by Piazza et al. (2020). The shifts of aggregate distribution and 449 eukaryotic community structure toward more mM and distinct soil biota communities under CT at 450 low N availability can be related to a lower sequestration of C within mM and thus in differences of 451 the related functional soil biota. By contrast, the lack of effect at subsurface layer is unexpected 452 since the percentage of mM was significantly decreased by tillage intensification (CT < MT; - 21%) 453 (Piazza et al., 2020). However, this inconsistency could be due to the coverage of the V4 region 454 primer set, its taxonomic resolution or limitation in amplifying rare taxa or taxa with lower 455 proportions of template DNA in DNA extracts (Choi and Park, 2020). Moreover, considering 456 aggregate pore size and animal body size, the presence of traces of animal DNA (i.e., nematodes, 457 Arthropoda) within aggregates is likely not attributable to the occurrence of living animals, but to 458 the process of aggregate formation which utilises organic decaying material as binding agent. 459 Nitrogen fertilization determined a strong shift at the surface layer in the eukaryotic community 460 structure of mM under CT (Fig. 1d) from Glomeromycetes (Glomeromycota), Imbricatea, 461 Sarcomonadea and Vampyrellidea (Cercozoa) in CTN0 to OTU1Xantho (Ochrophyta), 462 Oligohymenophorea (Ciliophora), Stramenopiles (Chromista) and the fungus Tremellomycetes in 463 CTN200. This is the first time that Glomeromycota have been detected within mM fraction. 464 Previously, using a cloning approach targeting the long-fragment SSU-ITS-LSU (Krüger et al., 465 2009) we could not detect AMF within mM (data not shown), and this was also supported by

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several works reporting their major roles only in macroaggregates (e.g., Miller and Jastrow, 2000, Rillig et al., 2002). However, in this study, the observed large proportion of Glomeromycota (14%) in mM under CTN0 supports the fact that tillage under unfertilized conditions may not negatively affect the development of the extraradical mycelium, potentially improving the production of glomalin and enhancing soil aggregate stability (Bedini et al., 2002). Similarly, the high abundance of Cercozoa in mM under CTN0 suggests for the first time that this phylum plays a major role within mM under ploughed and no fertilized conditions at surface layer. This is consistent with the findings of Degrune et al. (2019a) that highlighted under ploughing and at topsoil distinct cercozoan communities in microhabitats (i.e., drilosphere and rhizosphere) compared with bulk soil. Moreover, the distinct eukaryotic community found at surface layer in mM under CTN200 additionally supports that, under ploughing, nutrient availabilities in microhabitats allow the dominance of functional protists (Alveolata: Oligohymenophorea; Chromista: OTU1Xantho and Stramenopiles), potentially contributing to OM decomposition and mineralization through several functional groups. In addition, scarce information is available on the functional roles in agricultural soils of Tremellomycetes, a heterogeneous group comprising saprotrophs, animal parasites, and fungicolous species. Similar community structures were observed at surface layer within mM in MTN0 and MTN200 [e.g., Sarcomonadea1 (Cercozoa), and fungi such as Dothideomycetes (Ascomycota) and OTU1Cystobasidio (Basidiomycota)] (Fig. 1d). These results support the hypothesis of a major role played by Cercozoa together with distinct classes of fungi also within mM under MT. However, it is well known that the 18S barcoding utilised in this work is less efficient compared with the ITS for detecting many groups of fungi (Schoch et al., 2012). To test if soil aggregates differentially shape the structure of soil eukaryotes (hypothesis 2), the relative abundance pattern of taxa was determined in sM and mM. Significant differences among matrices (sM vs mM) were found and supported by PERMANOVAs (Table 1). PCO biplots showed that at both soil layers more phyla were linked to mM as compared with sM (Fig. 3c,d).

Recently, Liao et al. (2018) and Wang et al. (2021) used an Illumina sequencing approach for studying at phylum and class level the bacterial and fungal community structures within soil aggregates across different fertilization treatments. Although differences in community structure were detected for both bacteria and fungi, fungal community in sM and free microaggregates differed more than bacteria (Liao et al., 2018; Wang et al., 2021). Our findings support that some unresolved taxa belonging to Glomeromycota and Ascomycota are positively associated with mM, as previously reported in free microaggregates for unclassified Ascomycota (Wang et al., 2021) and for a group of unclassified *Glomerales* (Lu et al., 2018). Similarly to the results of Jiang et al. (2017), the total abundance of nematodes increased with increasing aggregate size. Finally, the alveolate Apicomplexa, Ciliophora, and Dinoflagellata were preferentially found in mM, whereas the amoebozoan Conosa and Lobosa in sM. This result additionally confirms the functional role played by protists within microenvironments. Moreover, Mollusca in mM and Arthropoda and Anellida in sM at both soil layers can be considered as preferential binding agents for aggregate fractions.

Eukaryotic taxa predictors for soil structuring and C stocks

To test if some eukaryotic taxa are predictors for soil structuring and C stock (hypothesis 3) we utilised a multiple regression analysis that allowed to identify the eukaryotic taxa that were good predictors for the amount of sM and mM and their SOC content, irrespective of management and soil depth (Table S1). Specifically, *Microbotryomycetes* and Alveolata were moderately strongly related to the amount of sM, with *Microbotryomycetes* identified as best predictor. Similarly, Cercozoa and *Chytridiomycetes* were related to the amount of mM, with Cercozoa playing the major role. Moreover, *Microbotryomycetes*, Cercozoa and Alevolata were moderately related to SOC in sM, with *Microbotryomycetes* consistently found to be the best predictor. Finally, *Chytridiomycetes* and Cercozoa were moderately related to SOC in mM, with *Chytridiomycetes* the best predictor. Previously, Bach et al. (2018) identified bacterial and fungal indicators in free and

large macroaggregates. However, it is the first time that *Microbotryomycetes* and *Chytridiomycetes* have been shown to be correlated with the amount of sM and mM and their SOC content, respectively. Both fungal classes correlated with the pattern of C-cycling enzymes and SOC content in bulk soil (Piazza et al., 2019), and their abundance was high in sM and mM, respectively (Degrune et al., 2019b). *Microbotryomycetes* were also recently identified as good predictors of slow and passive SOC decomposition parameters (Hale et al., 2019). Our findings on the positive association of protists, Alveolata and Cercozoa, with the amount of sM and mM and their SOC content, support the multiple agroecological roles of protists found in bulk soil (Cavalier-Smith and Chao, 2003; Delgado-Baquerizo et al., 2020). Moreover, our results confirm previous works reporting that protists are shaped by pore size reduction and soil aeration (Berisso et al., 2012; Degrune et al., 2019a), features related to soil aggregates. Thus, we can assume that Alveolata are playing major role in soil, promoting sM formation and slowing down the decay of SOM within sM, while Cercozoa are crucial microorganisms in mM taking part to long-term sequestration and storing of SOC.

3.6. How eukaryotes are interlinked among each other and to prokaryotes in soil aggregates, and network traits predictor for soil structuring and C stocks

To test if eukaryotes form structured assemblages and distinctive networks in soil aggregates (hypothesis 4), and how eukaryotes are linked to prokaryotes (hypothesis 5), within- and cross-domain networks were built for sM and mM. At the surface layer within- and cross-domain networks were more complex in mM than in sM, whereas at the subsurface layer they did not vary (Table 2, Fig. 4, Fig. S6). In the cross-domain networks, both sM and mM showed a general trend toward a higher percentage of eukaryotes per subnetwork compared to prokaryotes at the surface layer compared with the subsurface layer (Table 3, Fig. 4b,d). Moreover, at both soil depth, in the sM cross-domain networks the majority of subnetworks were composed of both eukaryotes and

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prokaryotes, whereas in the mM cross-domain networks the subnetworks were half composed of eukaryotes and half of prokaryotes, and few subnetworks were mixed (Table 3). This is the first study that demonstrated that eukaryotes, components of soil biota communities usually studied separately, formed structured associations within each other and with prokaryotes in soil aggregates. Moreover, mM consistently had tighter connectivity compared with sM in both withinand cross-domain networks (Fig. 4, Fig. S6). This might be related to microhabitat conditions (i.e., wetter and more nutrient rich microhabitats) in mM that shift biotic interactions from facilitation to competition, leading to higher correlations between eukaryotic taxa or eukaryotic and prokaryotic taxa, as previously reported for fungi and bacteria in bulk soil and roots across land uses and agricultural managements with a gradient of nutrient availabilities (e.g., SOC, P levels) and pH (de Menezes et al., 2015; Banerjee et al., 2016, 2018; Wang et al., 2021). Other explanations could be a higher proportion of viable cells and spores and a lower niche heterogeneity (i.e., nutrients) in mM respect to sM, leading to tighter within- and cross-domain networks. Finally, a higher plant residue diversity in mM could also explain the within- and cross-domain mM network traits, as previously shown for plant community composition or host selectivity against microbial network complexity (Xiong et al., 2021). In the within-domain networks of sM and mM and at both soil layers, Cercozoa were highly co-occurring in the subnetworks respect to the other eukaryotic phyla, as shown by the network

on the within-domain networks of sM and mM and at both soil layers, Cercozoa were nightly co-occurring in the subnetworks respect to the other eukaryotic phyla, as shown by the network traits (Table 3, Supplementary Results and Discussions 2). It is noteworthy the high percentage of taxa belonging to Cercozoa (21%) involved in the largest subnetworks, composed of 33, occurring at surface layer in the within-domain mM networks. In addition, while fungi, mainly Ascomycota and Basidiomycota, were highly co-occurring in the within-domain sM subnetworks at both soil layer, protists, as Ochrophyta and Stramenopiles together with Chlorophyta, were co-occurring in the within-domain mM subnetworks at surface and subsurface layer, respectively (Table 2). Following the definition of keystone taxa by Banerjee et al. (2018), we can support the agroecological theory that some Cercozoa, as *Sarcomondea* and *Vampyrillidea* (Fig. 4c), can be

suggested as "hubs" (keystone taxa) within mM. Consistently, in the sM and mM cross-domain networks at the surface layer, Cercozoa can be considered as "hubs" as they are directly connected to prokaryotes in the largest subnetworks (Fig. 4b,d). Moreover, protists were involved in the largest subnetworks and showed variable direct connections (Fig. 4b,d, Fig. S6b,d). Indeed, the linkages of protists, according to their feeding versatility (Geisen, 2016), varied from direct connections to many prokaryotes, mainly belonging to Chloroflexi in the cross-domain sM networks (Fig. 4b, Fig. S6b), to direct linkages to other eukaryotes (e.g., fungi) or other protists in mM networks (Fig. 4d, Fig. S6d). So far, studies on the functional roles of soil biota in the formation of soil aggregates have mainly focused on the role played by a single functional group, e.g. AMF, earthworms, nematodes, termites and microarthropods (mites and collembolans) (e.g., Lee and Foster, 1991; Pulleman et al., 2005; Rillig and Mummey, 2006; Siddiky et al., 2012a,b; Zhang et al., 2016). However, only recently, Cercozoa and other protists, as Lobosa and Ciliophora, were shown to be positively related with ecosystem services, i.e. nutrient cycling and OM decomposition (Delgado-Baquerizo et al., 2020). Details about traits and taxa co-occurrences in within- and cross-domain networks and the significance of cross-domain relationships (Mantel test and CoCA) are given in Supplementary Results and Discussion 2, Fig. S7 and Fig. S8.

To test if some network traits can predict soil structuring and C stocks, we utilized a multiple regressions analysis that identified the number of edges and mean nodes per network as predictors for the amount of sM and mM and their SOC content, irrespective of management and soil depth (Table S2). Although network analysis is now largely applied to study soil biota co-occurrence and plant-microbe associations across different treatments (e.g., de Menezes et al., 2015; Farrer et al., 2019; Feng et al., 2019), little is known about soil biota networks within aggregates (Jiang et al., 2015, 2017) and few studies have dissected the predictable power of the network traits (topological properties) on ecosystem services. Jiang et al. (2015, 2017) indicated that aggregate fractions (large and small macroaggregates, and free microaggregates) showed a strong effect on the association networks of nematodes and bacteria and using the topological properties they could identify large

macroaggregates network as organized soil food web, showing functional interrelationships between bacterivorous nematodes and bacteria. In accordance with our results, the topological properties of soil biota networks should be taken into consideration for dissecting soil structuring as well as C cycling.

4. Conclusions

We have shown that soil aggregation is essential for a complete 'multifunctional' perspective of soil biota. A full understanding of relationships between soil biota and soil functions requires analyses emphasizing the feedbacks between soil structure and soil biota, rather than a unidirectional approach simply addressing the roles of single key functional groups. Next generation sequencing tools have been confirmed in this study to be crucial in the understanding of eukaryotic structures and soil biota networks and have the potential to further reveal their contributions to soil functions. Indeed, our findings demonstrate for the first time that protists together with fungi play major roles in soil structuring and C cycling, and that Cercozoa represent hubs in the soil biota aggregate networks. This supports the fact that their conservation is fundamental to prevent soil degradation and to enhance SOC accumulation in agroecosystems.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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References

- 628 Adl, M.S., Gupta, V.S., 2006. Protists in soil ecology and forest nutrient cycling. Canadian Journal
- 629 of Forest Research 36, 1805–1817.
- Adl, M.S., Coleman, D.C., Read, F., 2006. Slow recovery of soil biodiversity in sandy loam soils of
- Georgia after 25 years of no-tillage management. Agriculture, Ecosystem and Environment
- 632 114, 323–334.
- Anderson, M., Braak, C.T., 2003. Permutation tests for multi-factorial analysis of variance. Journal
- of Statistical Computation and Simulation 73, 85–113.
- Anderson, M.J., Gorley, R.N., Clarke, R.K., 2008. Permanova+ for Primer: Guide to Software and
- 636 Statistical Methods. Primer-E Limited.
- 637 Aune, J. B., 2012. Conventional, organic and conservation agriculture: production and
- environmental impact. In: Lichtfouse E. (Eds), Agroecology and strategies for climate change.
- Sustainable Agriculture Reviews, vol 8. Springer: Dordrecht, NL, pp. 149–165.
- Bach, E.M., Williams, R.J., Hargreaves, S.K., Yang, F., Hofmockel, K.S., 2018. Greatest soil
- microbial diversity found in micro-habitats. Soil Biology and Biochemistry 118, 217–226.
- Banerjee, S., Baah-Acheamfour, M., Carlyle, C.N., Bissett, A., Richardson, A.E., Siddique, T.,
- Bork, E.W., Chang, S.X., 2016. Co-occurrence patterns of soil bacterial communities.
- Environmental Microbiology 18, 1805–1816.

- Banerjee, S., Schlaeppi, K., van der Heijden, M.G., 2018. Keystone taxa as drivers of microbiome
- structure and functioning. Nature Reviews Microbiology 16, 567–576.
- Bates, S.T., Clemente, J.C., Flores, G.E., Walters, W.A., Parfrey, L.W., Knight, R., et al., 2013.
- Global biogeography of highly diverse protistan communities in soil. The ISME Journal 7, 652.
- Bedini, S., Pellegrino, E., Avio, L., Pellegrini, S., Bazzoffi P, Argese E, Giovannetti, M., 2009.
- Changes in soil aggregation and glomalin-related soil protein content as affected by the
- arbuscular mycorrhizal fungal species *Glomus mosseae* and *Glomus intraradices*. Soil Biology
- and Biochemistry 41, 1491–1496.
- Berisso, F. E., Schjønning, P., Keller, T., Lamandé, M., Etana, A., de Jonge, L.W., Iversen, B.V.,
- Arvidsson, J., Forkmann, J., 2012. Persistent effects of subsoil compaction on pore size
- distribution and gas transport in a loamy soil. Soil and Tillage Research 122, 42–51.
- Bossuyt, H., Six, J., Hendrix, P.F, 2006. Interactive effects of functionally different earthworm
- species on aggregation and incorporation and decomposition of newly added residue carbon.
- 658 Geoderma 130, 14–25.
- Bremner, J.M., Mulvaney, C.S., 1982. Nitrogen total 1. Methods of soil analysis. Part 2. Chem.
- Microbiol. Properties, pp. 595-624.
- Briones, M.J.I., Schmidt, O., 2017. Conventional tillage decreases the abundance and biomass of
- earthworms and alters their community structure in a global meta-analysis. Global Change
- 663 Biology 23, 4396–4419.
- 664 Cai, S., Wang, J., Lv, W., Xu, S., Zhu, H., 2020. Nitrogen fertilization alters the effects of
- earthworms on soil physicochemical properties and bacterial community structure. Applied
- 666 Soil Ecology 150, 103478.
- 667 Caporaso, J.G., Lauber, C.L., Walters, W.A., Berg-Lyons, D., Huntley, J., Fierer, N., et al, 2012.
- Ultra-high-throughput microbial community analysis on the Illumina HiSeq and MiSeq
- platforms. The ISME Journal 6, 1621.

- 670 Cavalier-Smith, T., Chao, E.E.Y., 2003. Phylogeny and classification of phylum Cercozoa
- 671 (Protozoa). Protist 154, 341–358.
- 672 Chen, M., Li, X., Yang, Q., Chi, X., Pan, L., Chen, N., et al., 2012. Soil eukaryotic microorganism
- succession as affected by continuous cropping of peanut-pathogenic and beneficial fungi were
- 674 selected. Plos One 7, e40659.
- 675 Choi, J., Park, J.S., 2020. Comparative analyses of the V4 and V9 regions of 18S rDNA for the
- extant eukaryotic community using the Illumina platform. Scientific Report 10, 1–11.
- 677 Clarke, K.R., Gorley, R.N., 2006. PRIMER v6: User Manual/Tutorial. PRIMER-E: Plymouth, UK.
- de Menezes, A.B., Prendergast- Miller, M.T., Richardson, A.E., Toscas, P., Farrell, M.,
- Macdonald, L.M., Baker, G., Wark, T. and Thrall, P.H., 2015. Network analysis reveals that
- bacteria and fungi form modules that correlate independently with soil
- parameters. Environmental Microbiology 17, 2677–2689.
- De Vries, F.T., Thébault, E., Liiri, M., Birkhofer, K., Tsiafouli, M.A., Bjørnlund, L., Bracht
- Jørgensen, H., Brady, M.V., Christensen, S., de Ruiter, P.C., d'Hertefeldt, T., Frouz, J.,
- Hedlund, K., Hemerik, L., Hol, W.H.G., Hotes, S., Mortimer, S.R., Setälä, H., Sgardelis, S.P.,
- Uteseny, K., W.H., Wolters, V., Bardge, R.D., 2013. Soil food web properties explain
- ecosystem services across European land use systems. Proceedings of the National Academy of
- 687 Sciences 110, 14296–14301.
- Degrune, F., Boeraeve, F., Dufrêne, M., Cornélis, J.T., Frey, B., Hartmann, M., 2019a. The
- pedological context modulates the response of soil microbial communities to agroecological
- management. Frontiers in Ecology and Evolution 7, 261.
- Degrune, F., Dumack, K., Fiore-Donno, A.M., Bonkowski, M., Sosa-Hernández, M.A., Schloter,
- M., Kautz, T., Fischer, D., Rillig, M.C., 2019b. Distinct communities of Cercozoa at different
- soil depths in a temperate agricultural field. FEMS Microbiology Ecology 95, fiz041.
- 694 Del Campo, J., Heger, T.J., Rodríguez-Martínez, R., Worden, A.Z., Richards, T.A., Massana, R.,
- Keeling, P.J., 2020. Assessing the diversity and distribution of apicomplexans in host and free-

- living environments using high-throughput amplicon data and a phylogenetically informed
- reference framework. Frontiers in Microbiology 10, 2373.
- 698 Delgado-Baquerizo, M., Reich, P.B., Trivedi, C., Eldridge, D.J., Abades, S., Alfaro, F.D., Bastida,
- F., Berhe, A.A., Cutler, N.A., Gallardo, A., García-Velázquez, L., Hart, S.C., Hayes, P.E., He,
- J., Hseu, Z., Hu, H., Kirchmair, M., Neuhauser, S., Pérez, C.A., Reed, S.C., Santos, F., Sullivan,
- B.W., Trivedi, P., Wang, J., Weber-Grullon, L., Williams M.A., Singh, B.K., 2020. Multiple
- elements of soil biodiversity drive ecosystem functions across biomes. Nature Ecology and
- 703 Evolution 4, 210–220.
- Denef, K., Zotarelli, L., Boddey, R.M., Six, J., 2007. Microaggregate-associated carbon as a
- diagnostic fraction for management-induced changes in soil organic carbon in two Oxisols. Soil
- Biology and Biochemistry 39, 1165–1172.
- 707 Duan, Y., Chen, L., Zhang, J., Li, D., Han, X., Zhu, B., Li, Y., Zhao, B., Huang, P., 2021. Long-
- term fertilisation reveals close associations between soil organic carbon composition and
- microbial traits at aggregate scales. Agriculture, Ecosystem and Environment 306, 107169.
- 710 Edgar, R.C., 2010. Search and clustering orders of magnitude faster than BLAST. Bioinformatics
- 711 26, 2460-2461.
- 712 Edgar, R.C., Flyvbjerg, H., 2015. Error filtering, pair assembly and error correction for next
- generation sequencing reads. Bioinformatics 31, 3476-3482.
- 714 Edwards, J., Johnson, C., Santos-Medellín, C., Lurie, E., Podishetty, N.K., Bhatnagar, S., Eisen,
- J.A., Sundaresan, V., 2015. Structure, variation, and assembly of the root-associated
- microbiomes of rice. Proceedings of the National Academy of Sciences 112, E911–E920.
- 717 Elliott, E.T., 1986. Aggregate structure and carbon, nitrogen, and phosphorus in native and
- 718 cultivated soils. Soil Science Society of America Journal 50, 627–633.
- 719 Ercoli, L., Schüßler, A., Arduini, I., Pellegrino, E., 2017. Strong increase of durum wheat iron and
- 720 zinc content by field-inoculation with arbuscular mycorrhizal fungi at different soil nitrogen
- availabilities. Plant and Soil 419, 153–167.

- Farrer, E.C., Porazinska, D.L., Spasojevic, M.J., King, A.J., Bueno de Mesquita, C.P., Sartwell,
- S.A., Smith, J.G., White, C.T., Schmidt, S.K., Suding, K.N., 2019. Soil microbial networks
- shift across a high-elevation successional gradient. Frontiers in Microbiology 10, 2887.
- Feng, K., Zhang, Y., He, Z., Ning, D., Deng, Y., 2019. Interdomain ecological networks between
- plants and microbes. Molecular Ecology Resources 19, 1565–1577.
- Fierer, N., 2017. Embracing the unknown: disentangling the complexities of the soil microbiome.
- 728 Nature Reviews Microbiology 15, 579–590.
- 729 Fiore-Donno, A.M., Richter-Heitmann, T., Degrune, F., Dumack, K., Regan, K.M., Marhan, S.,
- Boeddinghaus, R.S., Rillig, M.C., Friedrich, M.W., Kandeler, E., Bonkowski, M., 2019.
- Functional traits and spatio-temporal structure of a major group of soil protists (Rhizaria:
- 732 Cercozoa) in a temperate grassland. Frontiers in Microbiology 10, 1332.
- 733 Foissner, W., 1998. An updated compilation of world soil ciliates (protozoa, Ciliophora), with
- ecological notes, new records, and descriptions of new species. European Journal of
- 735 Protistology 34, 195–235.
- Fonte, S.J., Kong, A.Y., van Kessel, C., Hendrix, P.F., Six, J., 2007. Influence of earthworm
- activity on aggregate-associated carbon and nitrogen dynamics differs with agroecosystem
- management. Soil Biology and Biochemistry 39, 1014–1022.
- 739 Franzluebbers, A.J., Arshad, M.A., 1997. Particulate organic carbon content and potential
- mineralization as affected by tillage and texture. Soil Science Society of America Journal 61,
- 741 1382–1386.
- 742 Gao, F., Warren, A., Zhang, Q., Gong, J., Miao, M., Sun, P., Xu, D., Huang, J., Yi, Z., Song, W.,
- 743 2016. The all-data-based evolutionary hypothesis of ciliated protists with a revised
- classification of the phylum Ciliophora (Eukaryota, Alveolata), Scientific Report 6, 24874.
- Geisen, S., 2016. The bacterial-fungal energy channel concept challenged by enormous functional
- versatility of soil protists. Soil Biology and Biochemistry 102, 22–25.

- Ghosh, A., Bhattacharyya, R., Dey, A., Dwivedi, B. S., Meena, M. C., Manna, M. C., Agnihortri, R.
- 748 2019. Long-term fertilisation impact on temperature sensitivity of aggregate associated soil
- organic carbon in a sub-tropical inceptisol. Soil and Tillage Research 195, 104369.
- Gloor, G. B., Macklaim, J. M., Pawlowsky-Glahn, V., Egozcue, J. J. 2017. Microbiome datasets are
- compositional: and this is not optional. Frontiers in Microbiology 8, 2224.
- Grandin, U., 2006. PC-ORD version 5: a user-friendly toolbox for ecologists. Journal of Vegetation
- 753 Science 17, 843–844.
- Guan, D., Al-Kaisi, M. M., Zhang, Y., Duan, L., Tan, W., Zhang, M., Li, Z., 2014. Tillage practices
- affect biomass and grain yield through regulating root growth, root-bleeding sap and nutrients
- uptake in summer maize. Field Crops Research 157, 89-97.
- 757 Gupta, V.V., Germida, J.J., 2015. Soil aggregation: Influence on microbial biomass and
- 758 implications for biological processes. Soil Biology and Biochemistry 80, A3–A9.
- Hale, L., Feng, W., Yin, H., Guo, X., Zhou, X., Bracho, R., Pegoraro, E., Penton, C.R., Wu, L.,
- Cole, J., Konstantinidis, K.T., Luo, Y., Tiedje, J.M., Schuur, E.A.G., Zhou, J., 2019. Tundra
- microbial community taxa and traits predict decomposition parameters of stable, old soil
- organic carbon. The ISME Journal 2019 13, 2901–2915.
- Heberle, H., Meirelles, G.V., da Silva, F.R., Telles, G.P., Minghim, R., 2015. InteractiVenn: a web-
- based tool for the analysis of sets through Venn diagrams. BMC Bioinformatics 16, 169.
- Himmelstein, J., Ares, A., Gallagher, D., Myers, J., 2016. A meta-analysis of intercropping in
- Africa: impacts on crop yield, farmer income, and integrated pest management effects.
- 767 International Journal of Agricultural Sustainability 15, 1–10.
- 768 IUSS working group WRB., 2006. World reference base for soil resources 2006 A framework for
- international classification, correlation and communication. Food and Agriculture Organization
- of the United Nations: Rome, Italy.

- Jansa, J., Mozafar, A., Kuhn, G., Anken, T., Ruh, R., Sanders, I.R., Frossard, E.J.E.A., 2003. Soil
- tillage affects the community structure of mycorrhizal fungi in maize roots. Ecological
- 773 Applications 13, 1164–1176.
- Jastrow, J.D., Miller, R.M., 1998. Soil aggregate stabilization and carbon sequestration: feedbacks
- through organomineral associations. In: Lal, R., Kimble, J.M., Follett, R.F., Stewart, B.A.
- (Eds). Soil processes and the carbon cycle. CRC Press, New York, USA, pp. 207–223.
- 777 Jiang, Y., Liu, M., Zhang, J., Chen, Y., Chen, X., Chen, L., Li, H., Zhang X., Sun, B., 2017.
- Nematode grazing promotes bacterial community dynamics in soil at the aggregate level. The
- 779 ISME Journal 11, 2705–2717.
- Jiang, Y., Sun, B., Li, H., Liu, M., Chen, L., Zhou, S., 2015. Aggregate-related changes in network
- patterns of nematodes and ammonia oxidizers in an acidic soil. Soil Biology and Biochemistry
- 782 88, 101–109.
- 783 Kong, A.Y., Scow, K.M., Córdova-Kreylos, A.L., Holmes, W.E., Six, J., 2011. Microbial
- community composition and carbon cycling within soil microenvironments of conventional,
- low-input, and organic cropping systems. Soil Biology and Biochemistry 43, 20–30.
- Kottek, M., Grieser, J., Beck, C., Rudolf, B., Rubel, F., 2006. World map of the Köppen-Geiger
- 787 climate classification updated. Meteorologische Zeitschrift 15, 259–263.
- Krüger, M., Stockinger, H., Krüger, C., Schüßler, A., 2009. DNA- based species level detection of
- Glomeromycota: one PCR primer set for all arbuscular mycorrhizal fungi. New Phytologist
- 790 183, 212–223.
- 791 Kumar, S., Stecher, G., Tamura, K., 2016. MEGA7: molecular evolutionary genetics analysis
- version 7.0 for bigger datasets. Molecular Biology and Evolution 33, 1870-1874.
- Kurtz, Z.D., Müller, C.L., Miraldi, E.R., Littman, D.R., Blaser, M.J., Bonneau, R.A., 2015. Sparse
- and compositionally robust inference of microbial ecological networks. PLoS Computational
- 795 Biology 11, e1004226.

- 796 Lagomarsino, A., Grego, S., Kandeler, E., 2012. Soil organic carbon distribution drives microbial
- activity and functional diversity in particle and aggregate-size fractions. Pedobiologia 55, 101–
- 798 110.
- 799 Lal, R., Negassa, W., Lorenz, K., 2015. Carbon sequestration in soil. Current Opinion in
- 800 Environmental Sustainability 15, 79–86.
- Lee, K.E., Foster, R.C., 1991. Soil fauna and soil structure. Soil Research 29, 745–775.
- 802 Lehmann, A., Zheng, W., Rillig, M.C., 2017. Soil biota contributions to soil aggregation. Nature
- Ecology and Evolution 1, 1828.
- Lentendu, G., Wubet, T., Chatzinotas, A., Wilhelm, C., Buscot, F., Schlegel, M., 2014. Effects of
- long-term differential fertilization on eukaryotic microbial communities in an arable soil: a
- multiple barcoding approach. Molecular Ecology 23, 3341–3355.
- Liao, H., Zhang, Y., Zuo, Q., Du, B., Chen, W., Wei, D., Huang, Q., 2018. Contrasting responses of
- bacterial and fungal communities to aggregate-size fractions and long-term fertilizations in
- soils of northeastern China. Science of the Total Environment 635, 784–792.
- Liu, H., Roeder, K., Wasserman, L., 2010. Stability Approach to Regularization Selection (StARS)
- for High Dimensional Graphical Models. Advances in Neural Information Processing Systems
- 812 24, 1432–40.
- 813 Lu, X., Lu, X., Liao, Y., 2018. Effect of tillage treatment on the diversity of soil arbuscular
- mycorrhizal fungal and soil aggregate-associated carbon content. Frontiers in Microbiology 9,
- 815 2986.
- 816 Lüftenegger, G., Foissner, W., Adam, H., 1985. R- and K-selection in soil ciliates: a field and
- experimental approach. Oecologia 66, 574–579.
- Mantel, N., Valand, R.S., 1970. A technique of nonparametric multi-variate analysis. Biometrics
- 819 26, 547–558.
- Meinshausen, N., Bühlmann, P., 2010. Stability selection. Journal of the Royal Statistical Society
- Series B (Statistical Methodology) 72, 417–473.

- 822 Miller, R.M., Jastrow, J.D., 2000. Mycorrhizal fungi influence soil structure. In: Kapulnik, Y.,
- Douds, D.D. (Eds). Arbuscular mycorrhizas: physiology and function. Springer, Dordrecht,
- 824 ND, pp. 3–18.
- Morriën, E., Hannula, S.E., Snoek, L.B., Helmsing, N.R., Zweers, H., De Hollander, M., Soto, R.L.,
- Bouffaud, M., Buée, M., Dimmers, W., Duyts, H., Geisen, S., Girlanda, M., Griffiths, R.I.,
- Jørgensen, H., Jensen, J., Plassart, P., Redecker, D., Schmelz, R.M., Schmidt, O., Thomson,
- B.C., Tisserant, E., Uroz, S., Winding, A., Bailey, M.J., Bonkowski, M., Faber, J. H., Martin,
- F., Lemanceau, P., de Boer, W., van Veen J.A., van der Putten, W.H., 2017. Soil networks
- become more connected and take up more carbon as nature restoration progresses. Nature
- 831 Communication 8, 1–10.
- Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'hara, R.B., et al., 2013.
- Package 'vegan'. Community Ecology Package, version 2.9.
- Onweremadu, E.U., Onyia, V.N., Anikwe, M.A.N., 2007. Carbon and nitrogen distribution in
- water-stable aggregates under two tillage techniques in Fluvisols of Owerri area, southeastern
- Nigeria. Soil and Tillage Research 97, 195–206.
- 837 Pershina, E.V., Ivanova, E.A., Korvigo, I.O., Chirak, E.L., Sergaliev, N.H., Abakumov, E.V.,
- Provorov, N.A., Andronov, E.A., 2018. Investigation of the core microbiome in main soil types
- from the East European plain. Science of the Total Environment 631, 1421–1430.
- Piazza, G., 2019. Soil carbon and microbial diversity under conservation management practices.
- Doctoral dissertation, Scuola Superiore Sant'Anna, Pisa, Italy.
- Piazza, G., Ercoli, L., Nuti, M., Pellegrino, E., 2019. Interaction between conservation tillage and
- nitrogen fertilization shapes prokaryotic and fungal diversity at different soil depths: evidence
- from a 23-year field experiment in the Mediterranean area. Frontiers in Microbiology 10, 2047.
- Piazza, G., Pellegrino, E., Moscatelli, M.C., Ercoli, E., 2020. Long-term conservation tillage and
- 846 nitrogen fertilization effects on soil aggregate distribution, nutrient stocks and enzymatic
- activities in bulk soil and occluded microaggregates. Soil and Tillage Research 196, 104482.

- Pittelkow, C.M., Liang, X., Linquist, B.A., Van Groenigen, K.J., Lee, J., Lundy, M.E., van Gestel,
- N., Six, J., Venterea R.T., van Kessel, C., 2015. Productivity limits and potentials of the
- principles of conservation agriculture. Nature 517, 365–368.
- Pulleman, M.M., Six, J., Uyl, A., Marinissen. J.C.Y., Jongmans, A.G., 2005. Earthworms and
- management affect organic matter incorporation and microaggregate formation in agricultural
- soils. Applied Soil Ecology 29, 1–15.
- Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., et al., 2012. The SILVA
- ribosomal RNA gene database project: improved data processing and web-based tools. Nucleic
- 856 Acids Research 41, D590–D596.
- 857 Rillig, M.C., Aguilar-Trigueros, C.A., Bergmann, J., Verbruggen, E., Veresoglou, S.D., Lehmann,
- A., 2015. Plant root and mycorrhizal fungal traits for understanding soil aggregation. New
- Phytologist 205, 1385–1388.
- Rillig, M.C., Mummey, D.L., 2006. Mycorrhizas and soil structure. New Phytologist 171, 41–53.
- 861 Rillig, M.C., Wright, S.F., Eviner, V.T., 2002. The role of arbuscular mycorrhizal fungi and
- glomalin in soil aggregation: comparing effects of five plant species. Plant and Soil 238, 325–
- 863 333.
- Rusinamhodzi, L., Corbeels, M., Van Wijk, M.T., Rufino, M.C., Nyamangara, J., Giller, K.E.,
- 2011. A meta-analysis of long-term effects of conservation agriculture on maize grain yield
- under rain-fed conditions. Agronomy for Sustainable Development 31, 657.
- Santos, S.S., Schöler, A., Nielsen, T.K., Hansen, L.H., Schloter, M., Winding, A., 2020. Land use
- as a driver for protist community structure in soils under agricultural use across Europe.
- Science of the Total Environment 717, 137228.
- 870 Schmidt, M.W., Torn, M.S., Abiven, S., Dittmar, T., Guggenberger, G., Janssens, I.A., Kleber, M.,
- Kögel-Knabner, I., Lehmann, J., Manning, D.A.C., Nannipieri, P., Rasse, D.P., Steve Weiner,
- S., Susan E. Trumbore, S.E., 2011. Persistence of soil organic matter as an ecosystem property.
- 873 Nature 478, 49–56.

- 874 Schoch, C. L., Seifert, K. A., Huhndorf, S., Robert, V., Spouge, J. L., Levesque, C. A., Chen, W.,
- Fungal Barcoding Consortium, 2012. Nuclear ribosomal internal transcribed spacer (ITS)
- region as a universal DNA barcode marker for Fungi. Proceedings of the National Academy of
- 877 Sciences 109, 6241-6246.
- 878 Sheehy, J., Regina, K., Alakukku, L., Six, J., 2015. Impact of no-till and reduced tillage on
- aggregation and aggregate-associated carbon in Northern European agroecosystems. Soil and
- 880 Tillage Research 150, 107–113.
- 881 Siddiky, M.R.K., Kohler, J., Cosme, M., Rillig, M.C., 2012a. Soil biota effects on soil structure:
- Interactions between arbuscular mycorrhizal fungal mycelium and collembola. Soil Biology
- and Biochemistry 50, 33–39.
- 884 Siddiky, M.R.K., Schaller, J., Caruso, T., Rillig, M.C., 2012b. Arbuscular mycorrhizal fungi and
- Collembola non-additively increase soil aggregation. Soil Biology and Biochemistry 47, 93–99.
- 886 Six, J., Bossuyt, H., Degryze, S., Denef, K., 2004. A history of research on the link between (micro)
- aggregates, soil biota, and soil organic matter dynamics. Soil and Tillage Research 79, 7–31.
- 888 Six, J., Elliott, E.T., Paustian, K., 1999. Aggregate and soil organic matter dynamics under
- conventional and no-tillage systems. Soil Science Society of America Journal 63, 1350–1358.
- 890 Six, J., Elliott, E.T., Paustian, K., 2000. Soil macroaggregate turnover and microaggregate
- formation: a mechanism for C sequestration under no-tillage agriculture. Soil Biology and
- 892 Biochemistry 32, 2099–2103.
- 893 Six, J., Paustian, K., 2014. Aggregate-associated soil organic matter as an ecosystem property and a
- measurement tool. Soil Biology and Biochemistry 68, A4–A9.
- 895 Soil Survey Staff, 1975. Soil taxonomy. A Basic System of Soil Classification for Making and
- Interpreting Soil Surveys. USDA-SCS Agric. Handb.: Washington, DC, USA.
- 897 Stoeck, T., Bass, D., Nebel, M., Christen, R., Jones, M.D., Breiner, H.W., Richards, T.A., 2010.
- Multiple marker parallel tag environmental DNA sequencing reveals a highly complex
- 899 eukaryotic community in marine anoxic water. Molecular Ecology 19, 21–31.

- 900 Sun, A., Jiao, X.Y., Chen, Q., Trivedi, P., Li, Z., Li, F., Zheng, Y., Lin, Y., Hu, H.W. He, J.-Z.,
- 901 2021. Fertilization alters protistan consumers and parasites in crop-associated microbiomes.
- 902 Environmental Microbiology 23, 2169–2183.
- 903 ter Braak, C.J.F., Schaffers, A.P., 2004. Co-correspondence analysis: a new ordination method to
- relate two community compositions. Ecology 85, 834–846.
- 905 ter Braak, C.J.F., Smilauer, P., 2012. Canoco reference manual and user's guide: software for
- ordination (version 5.0). Microcomputer Power: Ithaca, USA.
- 907 Tipton, L., Müller, C.L., Kurtz, Z.D., Huang, L., Kleerup, E., Morris, A., Bonneau, R., Ghedin., E.,
- 908 2018. Fungi stabilize connectivity in the lung and skin microbial ecosystems. Microbiome 6,
- 909 12.
- 910 Tisdall, J.M., Oades, J., 1982. Organic matter and water-stable aggregates in soils. European
- 911 Journal of Soil Science 33, 141–163.
- 912 Totsche, K.U., Amelung, W., Gerzabek, M.H., Guggenberger, G., Klumpp, E., Knief, C.,
- Lehndorff, E., Mikutta, R., Peth, S., Prechtel, A., Ray, N., Kögel-Knabner, I., 2018.
- Microaggregates in soils. Journal of Plant Nutrition and Soil Science 181, 104–136.
- 916

915

- 917 van Capelle, C., Schrader, S., & Brunotte, J., 2012. Tillage-induced changes in the functional
- 918 diversity of soil biota-A review with a focus on German data. European Journal of Soil
- 919 Biology 50, 165-181.
- 920 Wagg, C., Schlaeppi, K., Banerjee, S., Kuramae, E.E., van der Heijden, M.G., 2019. Fungal-
- bacterial diversity and microbiome complexity predict ecosystem functioning. Nature
- 922 Communication 10, 1–10.
- 923 Wang, Q., Ma, M., Jiang, X., Guan, D., Wei, D., Zhao, B., Chen, S., Cao, F., Li, L., Yang, X., Li,
- J., 2019. Impact of 36 years of nitrogen fertilization on microbial community composition and

- soil carbon cycling-related enzyme activities in rhizospheres and bulk soils in northeast China.
- 926 Applied Soil Ecology 136, 148–157.
- Wang, X., Bian, Q., Jiang, Y., Zhu, L., Chen, Y., Liang, Y., Sun, B., 2021. Organic amendments
- drive shifts in microbial community structure and keystone taxa which increase C
- mineralization across aggregate size classes. Soil Biology and Biochemistry 153, 108062.
- 930 Wickham, H., Chang, W., 2008. ggplot2: An implementation of the Grammar of Graphics. R
- package version 0.7.
- 932 Xiong, C., Zhu, Y.G., Wang, J.T., Singh, B., Han, L.L., Shen, J.P., Li, P. Wang, G., Wu, C., Ge,
- A., Zhang, L., He, J., 2021. Host selection shapes crop microbiome assembly and network
- complexity. New Phytologist 229, 1091–1104.
- 935 Yilmaz, P., Parfrey, L.W., Yarza, P., Gerken, J., Pruesse, E., Quast, C., et al., 2013. The SILVA and
- 936 "all-species living tree project (LTP)" taxonomic frameworks. Nucleic Acids Research 42,
- 937 D643–D648.
- P38 Zhang, S., Li, Q., Lü, Y., Zhang, X., Liang, W., 2013. Contributions of soil biota to C sequestration
- varied with aggregate fractions under different tillage systems. Soil Biology and Biochemistry
- 940 62, 147–156.
- P41 Zhang, X., Li, Q., Zhu, A., Liang, W., Zhang, J., Steinberger, Y., 2012. Effects of tillage and
- residue management on soil nematode communities in North China. Ecological Indicators 13,
- 943 75–81.
- 244 Zhang, Z., Zhang, X., Mahamood, M., Zhang, S., Huang, S., Liang, W., 2016. Effect of long-term
- combined application of organic and inorganic fertilizers on soil nematode communities within
- aggregates. Scientific Report 6, 31118.
- 247 Zhao, Z.B., He, J.Z., Geisen, S., Han, L.L., Wang, J.T., Shen, J.P., Wei, W., Fang, Y., Li, P., Zhang,
- L., 2019. Protist communities are more sensitive to nitrogen fertilization than other
- microorganisms in diverse agricultural soils. Microbiome 7, 1–16.

Figure captions

Fig. 1. Long-term effect of conservation management on community composition of eukaryotes in two soil matrices: small macroaggregates (sM) and occluded microaggregates (mM). Managements were: MTN0 (minimum tillage and 0 kg N ha⁻¹), MTN200 (MT and 200 kg N ha⁻¹), CTN0 (conventional tillage and 0 kg N ha⁻¹) and CTN200 (CT and 200 kg N ha⁻¹). Neighbor-joining (NJ) tree of 56 eukaryotic taxon representative sequences (classes) found in (a) sM and (d) mM. NJ trees are based on the sequences obtained from the amplification of the V4 region (18 SSU rRNA gene). The eukaryotic taxa were assigned to Operational Taxonomic Unit (OTU) (at class phylogenetic resolution) by BLAST against the 18S SSU SILVA database by clustering sequence reads at the 97% similarity threshold. For each OTU, the proportion of sequences retrieved in each management (MTN0, light green; MTN200, dark green; CTN0, light red; CTN200, dark red) and soil depth (0–15 cm: light grey; 15–30 cm: dark grey) are shown in the pie charts. Venn diagrams of eukaryotic classes uniquely retrieved and shared across managements in sM at (b) 0-15 cm and (c) at 15-30 cm, and in mM at (e) 0-15 cm and (f) at 15-30 cm.

Fig. 2. Long-term effect of conservation management on relative abundances and community structures of eukaryotes in two soil matrices: small macroaggregates (sM) and occluded microaggregates (mM). Managements were: MTN0 (minimum tillage and 0 kg N ha⁻¹), MTN200 (MT and 200 kg N ha⁻¹), CTN0 (conventional tillage and 0 kg N ha⁻¹) and CTN200 (CT and 200 kg N ha⁻¹). (a) Relative abundances of eukaryotes at phylum level across treatments, matrices and soil depths. (b) Principal Coordinates Analysis (PCO) biplots on the interaction of tillage and N fertilization on the eukaryotic community structure at class level in sM at 0-15 cm and (c) at 15-30 cm, and (d) in mM at 0-15 cm. The output of the PCO biplots is based on the significant effect of

treatments following the permutational analysis of variance (PERMANOVA). We displayed only the taxa with a strong correlation (r = 0.50-0.70) with the ordination scores on each PCO axis.

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Fig. 3. Venn diagrams of eukaryotic phyla uniquely retrieved and shared in sM and mM at 0-15 cm (a) and at 15-30 cm (b). Principal Coordinates Analysis (PCO) biplots on effect of soil matrix (small macroaggregates sM *vs* occluded microaggregates mM) on eukaryotic community structure at phylum resolution at (a) 0-15 cm and at (b) 15-30 cm soil depth. The PCO biplots are based on the significant effect of soil matrix according to the permutational analysis of variance (PERMANOVA). In the biplots, only the taxa with a strong correlation (r = 0.50-0.70) with the ordination scores on each PCO axis were displayed.

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Fig. 4. Eukaryotic networks (within-domain networks) and eukaryotic-prokaryotic networks (crossdomain Associations networks) (at class phylogenetic resolution) in small macroaggregates (sM) (a,b) and occluded microaggregates (mM) (c,d) at surface layer (0-15 cm soil depth). Withindomain networks were built using the SParse InversE Covariance estimation for Ecological **ASsociation** Inference (SPIEC-EASI) package 0.1 R version in (https://github.com/zdk123/SpiecEasi/), while cross-domain networks were built by the crossdomain extension of SPIEC-EASI (Kurtz et al., 2015; Tipton et al., 2018). Details about network construction are given in Material and Methods and Supplementary Methods 3, and R scripts are provided in Supplementary Methods 4.

Table 1Results of PERMANOVA and variation partitioning of the long-term effect of conservation management (tillage and N fertilization) within small macroaggregates (sM) and occluded microaggregates (mM) and of the effect of matrix (sM *vs* mM) on eukaryotic community structure at 0-15 and 15-30 cm soil depths.

	df	Pseudo-F	P (perm)	Variance explained	df	Pseudo-F	P (perm)	Variance explained
		0-15 cm			15-30 cm			
Eukaryotes at ci	lass leve	el - sM		_				
TIL^\dagger	1	2.24	0.025^{\ddagger}	11.96	1	4.02	0.003	22.69
N fert	1	2.11	0.017	10.76	1	2.21	0.012	9.05
TIL x N fert	1	1.99	0.024	19.13	1	2.55	0.019	23.20
PERMDISP								
TIL			0.693				0.071	
N fert			0.078				0.359	
Eukaryotes at ci	lass leve	el - mM				_(()		
TIL	1	4.60	0.003	19.94	1	1.05	0.411	-
N fert	1	3.90	0.002	16.02	1	1.64	0.127	=
TIL x N fert	1	3.79	0.006	30.85	1	0.90	0.569	-
PERMDISP								
TIL			0.237				-	
N fert			0.362				-	
Eukaryotes at p	hylum le	evel - sM vs mi	M					
Matrix [§]	1	1.07	0.001	38.31	1	11.18	0.001	42.51
TIL x N fert	1	2.90	0.004	3.74	1	1.22	0.287	4.51
N fert	1	1.54	0.140	10.69	1	2.36	0.014	2.84
PERMDISP								
Matrix			0.025				0.069	

[†]PERMANOVA was performed following a split-plot design with tillage (TIL) as main-plot factor and N fertilization (N fert) as subplot factor and with three replicate plots per treatment: TIL (minimum tillage and conventional tillage) and N fert (0 kg N ha⁻¹ and 200 kg N ha⁻¹).

[‡]In bold statistically significant values ($P \le 0.05$).

[§]PERMANOVA was performed using the matrix as fixed factor, TIL and N fert as covaribales and 12 replicate plots per matrix.

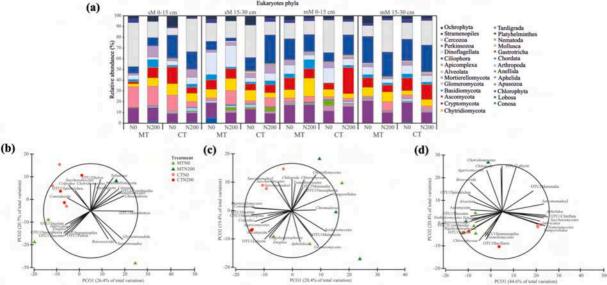
Table 2Traits of within- (eukaryotes) and cross-domain (eukaryotes - prokaryotes) network in macroaggregates (sM) and occluded microaggregates (mM) at 0-15 and 15-30 cm soil depth (for the network diagrams see Fig. 4 and S10).

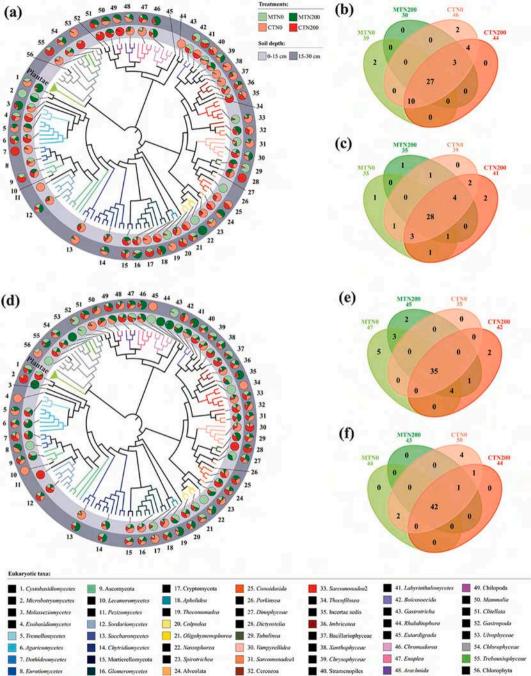
Traits	sM 0-15	sM 15-30	mM 0-15	mM 15-30
Eukaryotes Number of nodes excluding singletons	37	33	46	39
Number of edges	29	25	44	31
Number of singletons	12	16	8	15
Number of dyads	3	6	1	6
Number of subnetworks	6	5	4	5
Mean nodes per subnetwork	5.17 ± 0.60	4.20 ± 0.49	11.0 ± 7.34	5.40 ± 1.03
Linkage density (complexity)	1.57 ± 0.11 a	1.51 ± 0.13	$1.89 \pm 0.14 \text{ b}$	1.56 ± 0.13
% Positive interactions	78.4	75.8	95.7	79.5
Modularity	3	5	4	5
Identity of phyla with a frequency within the subnetworks $\geq 10\%$	Basidiomycota 15.9% Cercozoa 16.8%	Ascomycota 20% Basidiomycota 10% Cercozoa 11.7%	Cercozoa 11.6% Ochrophyta 16.7%	Chlorophyta 12% Cercozoa 21.7% Stramenopiles 16.7%
Eukaryotes - Prokaryotes				
Total number of nodes excluding singletons	79	83	96	67
Number of eukaryotic nodes excluding singletons	37	38	49	30
Number of prokaryotic nodes excluding singletons	42	45	47	37
Number of edges	74	72	109	54
Number of singletons	22	18	6	36
Number of dyads	3	8	2	5
Number of subnetworks	7	4	8	11
Mean nodes per subnetwork	10.43 ± 4.43	8.38 ± 2.74	23.50 ± 19.50	5.18 ± 0.70
Linkage density (complexity)	1.87 ± 0.11 a	1.76 ± 0.11	2.26 ± 0.12 b	1.61 ± 0.11
Modularity	9	9	11	9
Percentage of eukaryotes per subnetwork	68.39 ± 12.26	50.79 ± 12.03	74.70 ± 14.61	39.01 ± 14.69
Percentage of prokaryotes per subnetwork	31.61 ± 12.26	49.21 ± 12.03	25.30 ± 14.61	60.99 ± 14.69
Number of subnetworks with only eukaryotes	2	1	2	4
Number of subnetworks with only prokaryotes	0	1	0	5
Number of mixed subnetworks	5	6	2	2

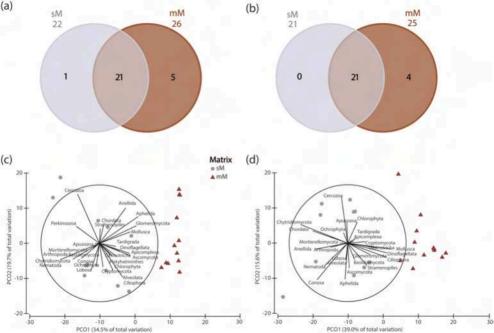
Table 3Traits of within- (eukaryotes) and cross-domain (eukaryotes - prokayotes) network in macroaggregates (sM) and occluded microaggregates (mM) at 0-15 and 15-30 cm soil depth (for the network diagrams see Fig. 4 and S10).

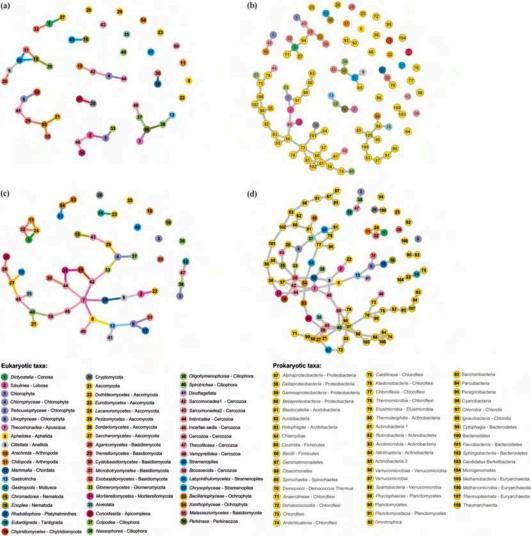
Traits	sM 0-15	sM 15-30	mM 0-15	mM 15-30
Eukaryotes				
Number of nodes excluding singletons	37	33	46	39
Number of edges	29	25	44	31
Number of singletons	12	16	8	15
Number of dyads	3	6	1	6
Number of subnetworks	6	5	4	5
Mean nodes per subnetwork	5.17 ± 0.60	4.20 ± 0.49	11.0 ± 7.34	5.40 ± 1.03
Linkage density (complexity)	1.57 ± 0.11 a	1.51 ± 0.13	$1.89 \pm 0.14 \text{ b}$	1.56 ± 0.13
% Positive interactions	78.4	75.8	95.7	79.5
Modularity	3	5	4	5
Identity of phyla with a frequency within the subnetworks $\geq 10\%$ †	Basidiomycota 15.9% Cercozoa 16.8%	Ascomycota 20% Basidiomycota 10% Cercozoa 11.7%	Cercozoa 11.6% Ochrophyta 16.7%	Chlorophyta 12% Cercozoa 21.7% Stramenopiles 16.7%
Eukaryotes - Prokaryotes		, (7)		
Total number of nodes excluding singletons	79	83	96	67
Number of eukaryotic nodes excluding singletons	37	38	49	30
Number of prokaryotic nodes excluding singletons	42	45	47	37
Number of edges	74	72	109	54
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Linkage density (complexity)	1.87 ± 0.11 a	1.76 ± 0.11	$2.26 \pm 0.12 \text{ b}$	1.61 ± 0.11
Modularity	9	9	11	9
Percentage of eukaryotes per subnetwork	68.39 ± 12.26	50.79 ± 12.03	74.70 ± 14.61	39.01 ± 14.69
Percentage of prokaryotes per subnetwork	31.61 ± 12.26	49.21 ± 12.03	25.30 ± 14.61	60.99 ± 14.69
Number of subnetworks with only eukaryotes	2	1	2	4
Number of subnetworks with only prokaryotes	0	1	0	5
Number of mixed subnetworks	5	6	2	2

[†] For details about the traits of within-domain networks for each eukaryotic phylum occurring in sM and mM at both soil depths see Table S5.









Highlights

- Eukaryote diversity responded differently to managements across soil aggregates
- A core community of eukaryotes was found across managements, aggregates and depths
- Tillage shifted eukaryotic structure in sM and mM according to N availability
- Protists and fungi positively correlated with the amount of sM, mM and SOC content
- Within- and cross-domain networks of mM at 0-15 cm showed the highest complexity

Declaration of interests

☑ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.						
□The authors declare the following financial interests/persons potential competing interests:	onal relationships which may be considered					
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