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1 **Upward expansion and acceleration of forest clearance in the mountains of Southeast Asia**

2

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17

18 Southeast Asia contains about half of all tropical mountain forests, which are rich in
19 biodiversity and carbon stocks, yet there is debate as to whether regional mountain forest
20 cover has increased or decreased in recent decades. Here, our analysis of high-resolution
21 satellite datasets reveals increasing mountain forest loss across Southeast Asia. Total mean
22 annual forest loss was 3.22 Mha yr⁻¹ during 2001–2019, with 31% occurring on mountains.
23 In the 2010s, the frontier of forest loss moved to higher elevations (15.1 ±3.8 m yr⁻¹ during
24 2011–2019, $p < 0.01$) and steeper slopes (0.22 ±0.05° yr⁻¹ during 2009–2019, $p < 0.01$) that have
25 high forest carbon density relative to lowlands. These shifts led to unprecedented annual
26 forest carbon loss of 424 Tg C yr⁻¹, accelerating at a rate of 18 ±4 Tg C yr⁻² ($p < 0.01$) from
27 2001–2019. Our results underscore the immediate threat of carbon stock losses associated
28 with accelerating forest clearance in Southeast Asian mountains, which jeopardizes
29 international climate agreements and biodiversity conservation.

30
31 Tropical forests are the largest terrestrial component of the global carbon cycle¹, storing 247 Pg C
32 in above and belowground biomass². However, recent anthropogenic-influenced forest loss has
33 reshaped tropical forests profoundly³, weakening their ability to store carbon and regulate climate⁴.
34 Currently, across the tropics, the amount of carbon sequestered by intact forests and forest regrowth
35 is approximately similar to that released from forest loss, suggesting that tropical forests likely act
36 as a neutral contributor to the global carbon cycle⁵. Forest loss in the tropics, which dominates the
37 total loss worldwide in the 21st century^{6–10}, has been driven largely by agricultural intensification

38 and/or extensification to support demands for human/animal food trade, profit-driven (illegal)
39 logging, and other activities that are inherently linked to population growth¹¹⁻¹³. Of concern is that
40 acceleration of forest clearance in the future will intensify carbon stock loss, potentially
41 transforming tropical forests into significant net carbon sources^{5,14,15}, as well as disrupting
42 biodiversity patterns, human livelihoods, hydro-geomorphological processes, and ecosystem
43 functions.

44
45 The general notion is that most tropical deforestation worldwide occurs primarily in lowland areas.
46 This sentiment aligns with prior work showing substantial forest losses at low elevations, but only
47 negligible losses, and even some forest gains, in the mountains^{6,16,17}. However, in Southeast Asia
48 (hereafter SEA), where approximately half of the world's tropical mountain forests are located^{8,18}
49 and extensive forest losses in the lowlands of Indonesia have occurred^{6,9}, recent studies have
50 reported new croplands and plantations replacing mountain forests in Laos and Thailand of
51 montane mainland SEA^{19,20}. Yet the applicability of these results^{19,20} as an indicator of a regional
52 trend is debatable, as some global analyses^{7,17} indicate an increase in forest cover in this region.
53 New spatiotemporal analyses conducted at high resolution and with common vegetation definitions
54 are needed to address these inconsistencies related to topography of forest loss in the mountains
55 and lowlands of SEA.

56
57 Here, we analyze multiple high-resolution satellite datasets to provide a comprehensive assessment

58 of changes in topographical patterns of forest clearance and related carbon loss across SEA during
59 the first two decades of the 21st century. The analyses incorporate global mountain extent map¹⁸
60 with two 30-m resolution products reporting the global forest cover change⁸ and aboveground live
61 woody biomass density²¹ (for details refer to *Methods*). Owing to limitations of distinguishing tree
62 types in the satellite products used⁸, unless specifically stated, “forest losses” incorporate those
63 from primary forest, secondary forest disturbance, as well as tree-dominated plantations, including
64 oil palm and rubber. As the mountains of SEA hold more forest biomass carbon than lowlands²²
65 (Fig. S1), a better understanding of forest and related biomass carbon dynamics is crucial for
66 reducing uncertainties in the global carbon cycle, as well as guiding land management in the region.

67

68 **Results**

69 This section presents our findings of forest loss in SEA, including the patterns of forest loss and
70 related topography and carbon loss.

71

72 **Accelerating forest loss and related topography.** We find a total forest loss of 61 Mha in SEA
73 during the period 2001 to 2019, which is equivalent to an annual rate of 3.22 Mha yr⁻¹ (Table 1;
74 Figs. 1a, S2c). Annual forest loss of the 2010s (4.02 Mha yr⁻¹) was nearly twice that of the 2000s
75 (2.33 Mha yr⁻¹), with the greatest loss occurring in 2016 (5.79 Mha yr⁻¹). Approximately 31% of
76 the loss during the 19-year period (19 Mha; 1.00 Mha yr⁻¹) occurred within the 61 mountainous
77 areas that occupy 1.7 million km² of the region (38% of SEA’s land surface; Fig. S2a, c). We also

78 find a significant increase in annual forest loss area across SEA since 2001, with an acceleration
79 rate of $0.17 \pm 0.03 \text{ Mha yr}^{-2}$ ($p < 0.01$). The annual rate of mountain forest loss increased 2.4-fold
80 from 0.58 Mha yr^{-1} in the first decade to 1.38 Mha yr^{-1} in the second decade (Fig. 1a).

81
82 Forest loss occurring in the lowlands of SEA significantly accelerated only during the 2000s (0.20
83 $\pm 0.04 \text{ Mha yr}^{-2}$, $p < 0.01$), with a non-significantly decreasing trend in the following decade (-0.01
84 $\pm 0.07 \text{ Mha yr}^{-2}$, $p = 0.92$). This pattern mirrors the fact that there were limited lowland forests that
85 can be converted to croplands in some regions over SEA during the 2010s, as lowland forests had
86 continued to be cleared since the 1980s (ref. 6). Regarding mountain forest loss, the near doubling
87 of acceleration rates from the first ($0.06 \pm 0.01 \text{ Mha yr}^{-2}$, $p < 0.01$) to the second ($0.11 \pm 0.03 \text{ Mha}$
88 yr^{-2} , $p < 0.01$) decade resulted from accelerated conversion of forests for crop plantation in
89 mountains¹⁹. Further, the trend in lowland forest loss was significantly different from that in
90 mountains during the 2000s ($p < 0.05$), but this difference was no longer statistically significant
91 during the 2010s (Fig. 1a). Taken together, these patterns reveal that forest loss in the mountains
92 increasingly comprised a significant portion of total forest loss in SEA from 2001 (24%) to 2019
93 (42%), which is a finding that has not been reported by prior studies^{6,9,23}.

94
95 Incorporating data on primary forest extent in 2001 (ref. 10), we further estimate that annual loss
96 of primary forest from 2001 to 2019 was 0.93 Mha yr^{-1} (Table 1; Fig. S3), with 0.26 Mha yr^{-1} (28%)
97 occurring in the mountains and 0.67 Mha yr^{-1} (72%) in the lowlands. These equate to 2.9% and

98 7.3% losses of the primary forest extent in 2001. Throughout the 19-year period, secondary forest
99 loss always exceeded primary forest loss in both lowlands and the mountains. Whereas secondary
100 forest loss accelerated significantly throughout the whole period ($0.14 \pm 0.02 \text{ Mha yr}^{-2}$, $p < 0.01$),
101 the significant acceleration in primary forest loss in the first decade ($0.11 \pm 0.02 \text{ Mha yr}^{-2}$, $p < 0.01$)
102 gave way to a non-significant decline in primary forest loss in the second decade ($-0.05 \pm 0.03 \text{ Mha}$
103 yr^{-2} , $p = 0.19$). Two trends emerged during the 2010s: (1) secondary forest loss in the mountains
104 greatly increased ($0.10 \pm 0.02 \text{ Mha yr}^{-2}$, $p < 0.05$) and (2) primary forest loss in the lowlands non-
105 significantly decreased ($-0.05 \pm 0.02 \text{ Mha yr}^{-2}$, $p = 0.06$). As the trend in secondary forest loss is
106 much larger than that of primary forest loss over the 19-year period, the ratio of primary-to-total
107 forest loss decreased from $>30\%$ to 20% . Collectively, the increase in mountain forest loss in the
108 2010s primarily originated from secondary forest loss, while the overall reduction in primary forest
109 loss resulted from reductions in the lowlands.

110
111 An elevational shift in the frontier of forest loss in the region is further supported by changes in the
112 elevation and slope of mean forest loss midway through the 19-year study period (Fig.1b).
113 Piecewise regression reveals an inflection point (hereafter IP) for mean elevation of forest loss that
114 occurred in 2011 and an IP for mean slope of forest loss that occurred in 2009 (Fig.1b). Within the
115 period after the IPs, the mean elevation and slope increased significantly at rates of $15.1 \pm 3.8 \text{ m yr}^{-1}$
116 ($p < 0.01$) and $0.22 \pm 0.05^\circ \text{ yr}^{-1}$ ($p < 0.01$), respectively. Importantly, forest loss in the mountains
117 accounted for most of both the observed increases in mean elevation (64% ; $9.6 \pm 2.7 \text{ m yr}^{-1}$, $p <$

118 0.01) and slope (64%; $0.14 \pm 0.04^\circ \text{ yr}^{-1}$, $p < 0.01$) after the IPs (Fig. 2a, b).

119
120 Regional patterns of trends in the mean elevation and slope where forest loss occurred (hereafter
121 termed as forest loss topography) show that east Sumatra and Kalimantan (Indonesia), north Laos,
122 and northeast Myanmar contribute to most of the increases in forest loss topography after IPs (Figs.
123 2). In some regions, a decreasing trend in forest loss topography occurred, such as on the Malay
124 peninsula (including south Thailand and Malaysia) and in Vietnam (Fig. S5). In Indonesia, which
125 experienced the largest magnitude of forest loss (Fig. S4), a sharp increase in forest loss topography
126 occurred during the second decade (Fig. S5). These losses in Indonesia contribute to most of the
127 increase in mean elevation (44% or $6.6 \pm 1.6 \text{ m yr}^{-1}$, $p < 0.01$) and slope (41% or $0.09 \pm 0.03^\circ \text{ yr}^{-1}$,
128 $p < 0.01$) in SEA after the IPs (Fig. 2a, b). Also of regional importance were the increases in forest
129 loss topography in Laos (28% for SEA's elevation and 23% for SEA's slope) and Myanmar (26%
130 for SEA's elevation and 23% for SEA's slope). In other countries, such as Thailand and the
131 Philippines, trends in forest loss topography were comparatively small (Fig. S5).

132
133 **Carbon loss resulting from forest clearance.** The observed shift in forest loss to higher elevations
134 and steeper slopes is of concern because mountain forests in the region tend to have higher carbon
135 stocks than lowland forests²²: $141 \pm 49 \text{ Mg C ha}^{-1}$ in mountains versus $101 \pm 69 \text{ Mg C ha}^{-1}$ in
136 lowlands (Fig. S1). By incorporating the forest change calculations in the previous section with
137 forest carbon stock map²¹ (see *Methods*), we estimate the total forest carbon loss in SEA during

138 2001–2019 was 8,050 Tg, equivalent to a rate of 424 Tg C yr⁻¹ (Fig. 3a; Table 1). As with annual
139 forest loss, forest carbon stock loss increased continuously throughout the entire period,
140 accelerating significantly at a rate of 18 ±4 Tg C yr⁻² ($p < 0.01$; Fig 3a, Table 1). Nearly a third of
141 the loss in forest carbon stocks (2,584 Tg C; 136 Tg C yr⁻¹) occurred in the mountains; lowland
142 forest carbon stock losses totaled 5,466 Tg (68%; 288 Tg C yr⁻¹). Mountain forest carbon loss
143 accelerated significantly both in the first (8 ±2 Tg C yr⁻², $p < 0.01$) and second decade (10 ±4 Tg C
144 yr⁻², $p < 0.05$), whereas the significant acceleration of lowland forest carbon stock loss (27 ±5 Tg
145 C yr⁻², $p < 0.01$) in the first decade was followed by a non-significant decrease in the 2010s (-9 ±8
146 Tg C yr⁻², $p = 0.30$). These trends result in the increasing contribution of mountain forest carbon
147 loss to total forest carbon loss in the second decade of the 21st century. Moreover, increasing
148 clearance of mountain forests with dense carbon stocks results in a disproportionate loss of carbon
149 stocks relative to past times when forest loss was more prevalent at lower elevations. For example,
150 in 2019, the last year of the analysis, mountain forest carbon loss was 119 Mg C ha⁻¹ yr⁻¹, which
151 was 7% higher than that of the lowlands. If these carbon loss rate trajectories continue, annual
152 forest carbon loss in mountains will exceed that of lowlands by 2022.

153

154 In agreement with the forest loss trends, the frontier of forest carbon loss also climbed to higher
155 elevations and steeper slopes during 2001–2019 (Fig. 3b). However, there are stark regional
156 differences in forest carbon loss patterns with respect to topography (Fig. 4). In maritime SEA
157 during the 2000s, most forest carbon losses took place in the lowlands (Fig. 4a), particularly on

158 some Indonesian islands (e.g., Sumatra, Kalimantan) and the Malay peninsula (Fig. 4c). Forest
159 carbon loss in the lowlands of maritime SEA accounted for 65% of SEA's total carbon loss in the
160 2000s. In the 2010s, lowland forest carbon loss decreased, particularly in Sumatra and Kalimantan
161 (Fig. 4d). However, positive trends in annual forest carbon loss occurred throughout many
162 mountainous areas of mainland SEA, pushing upwards and accelerating in the mountains of Laos,
163 and Myanmar. Although forest and related carbon loss in Vietnam and the Malay peninsula
164 increased (Figs. 4b, S4), the topography of forest loss in those regions decreased (Figs. 2, S5),
165 indicating that forest (carbon) loss accelerated in regions with lower elevations, a pattern that is
166 opposite to those observed in Myanmar and Laos. Overall, we conclude that the hotspots of forest
167 carbon loss, while mirroring those of forest loss in general, were found predominantly in lowland
168 maritime SEA in the 2000s. They were then located disproportionately in the mountains of
169 mainland SEA in the 2010s, particularly in northern Laos and northeast Myanmar, locations
170 strongly associated with increased forest loss at higher elevations and on steeper slopes (Fig. 2c,
171 d).

172

173 **Discussion**

174 In this section, we discuss the net changes in forest loss, implications, and potential limitations that
175 need to further address in future studies. Finally, we summarize our findings.

176

177 **Net changes.** In the dynamic environments of SEA, forest losses were also counteracted to some

178 degree by forest gains during the study in both lowland and mountain areas. Using the data
179 developed by Hansen et al.⁸, we determine that forest gains during the period of 2001–2012 were
180 10.3 Mha (0.86 Mha yr⁻¹) in the lowlands and 2.7 Mha (0.23 Mha yr⁻¹) in the mountains (Fig. S6).
181 These gains result in the net-to-gross loss proportion of 56% and 66% in lowlands and mountains,
182 respectively, during this abbreviated period. The lower net-to-gross loss rate in lowlands may be
183 related to extensive oil palm and timber plantation establishment following the removal of forest
184 or older plantations²⁴, as maturing plantations would be counted as forest gain once plants exceed
185 the threshold 5-m tree height definition of Hansen et al.^{8,25}. By assuming that the net-to-gross loss
186 ratios during 2013–2019 are the same as that in the earlier period, we estimate a 23.6 Mha (1.24
187 Mha yr⁻¹) net forest loss in the lowlands and 12.5 Mha (0.66 Mha yr⁻¹) net forest loss in the
188 highlands during 2001–2019 (Fig. S6). These estimates of net loss are likely conservative, given
189 that forest loss accelerated at a rate of 0.17 ± 0.03 Mha yr⁻² ($p < 0.01$) throughout the entire period
190 (Table 1).

191
192 Overall, our net estimates also reveal a clear fingerprint of mountain forest loss that is accelerating
193 in some countries of SEA (e.g., Indonesia, Myanmar, and Laos) during the early 21st century,
194 primarily owing to expansion of agriculture for crop plantation^{19,20}. The accelerating mountain
195 forest loss in the 2010s originated from secondary forest loss also mirrors the replacement of
196 swidden fields with other agriculture systems. For example, a notable shift from swidden fields,
197 where secondary forests regenerate during fallow period, to permanent agriculture systems is

198 reported in mountains of Laos²⁶, indicating that these forest losses in the mountains of SEA are
199 partly a result of agriculture intensification. This pattern, however, is different from agricultural
200 expansion in the Midwestern United States, which made the farms in the northeastern United States
201 not profitable and hence resulted in forest regeneration in that region²⁷.

202

203 **Implications.** Our results demonstrate not only a continuation of forest loss in SEA as reported in
204 sub-regions during prior periods^{6,9}, but an acceleration in loss that includes encroachment into
205 forests at higher elevations with higher carbon density. These trends influence the roles tropical
206 forests play within the context of global climate mitigation, biodiversity conservation, and global
207 carbon cycling. For example, the observed acceleration in forest carbon loss counters efforts to
208 limit global warming to below 2 °C by the end of this century²⁸. The climb in the forest loss frontier
209 also represents a challenge for climate change assessments, as current earth system models do not
210 differentiate mountain from lowland forest loss because of their coarse spatial resolutions¹⁹,
211 potentially resulting in the misrepresentation of climate feedbacks. In addition to the warming
212 triggered by forest carbon loss to the atmosphere through biochemical feedbacks, tree replacement
213 also increases surface temperature at a variety of scales through biophysical feedbacks^{28,29}. In the
214 mountains of SEA, where most deforested lands are converted to croplands¹⁹, warming effects
215 related to forest loss tend to be amplified due to suppressed evapotranspiration, raising local
216 temperatures by up to 2 °C²⁹⁻³¹. The acceleration of mountain forest loss in the region has likely
217 already enhanced these warming effects and influenced the carbon budget.

218

219 The acceleration in forest loss also affects biodiversity conservation in the region because a great
220 number of endemic species are found in the mountains of SEA³². While widespread conversion of
221 forests to croplands significantly reduces species richness and alters community composition in
222 general, loss of mountain forest habitat is particularly detrimental^{33,34}. Tropical montane species
223 typically live within specific hydro-thermal environments, which are dramatically altered during
224 forest conversion, increasing extinction risk^{35,36}. Deforestation also interacts with climate changes,
225 forcing species to redistribute³⁷, often to higher and cooler locations. Mountain forest loss threatens
226 to reduce the area of suitable habitat to accommodate these types of relocations³⁸.

227

228 Beyond the direct loss of carbon associated with vegetation biomass removal and habitat loss, forest
229 loss also affects the carbon cycle through diminishing photosynthesis and altering soil carbon
230 stocks. For example, forest loss directly lowers landscape-wide photosynthesis due to decreases in
231 leaf area and alteration of vegetation functioning. Forest conversion also alters basic water balance
232 processes including evapotranspiration, infiltration, and water storage³⁹⁻⁴¹, thereby modulating
233 vegetation growth and associated carbon assimilation. Soil erosion accelerated by forest conversion,
234 particularly on sloping lands, exhumes soil carbon that may be quickly released to the atmosphere,
235 or transported into downslope flood plain locations, water bodies, or the ocean, where it is
236 stored/lost at variable time scales^{42,43}. Unfortunately, because of the absence of regional data on
237 soil carbon stocks, we were not able to account for losses of this component, which for some forest

238 conversion outcomes are substantial^{3,44}.

239

240 **Uncertainties and caveats.** With regard to uncertainties in our analysis, fragmentation and edge
241 effects of forest losses can alter microclimates, and thus regulates the growth and structure of nearby
242 trees, causing additional long-term carbon losses on the landscape that we could not quantify⁴⁵.
243 Additional uncertainty relates to our inability to detect forest conversions at scales smaller than a
244 Landsat pixel, for example, those related to small-scale, fallow-based swidden agriculture, which
245 is still practiced in some areas of SEA^{20,46}. Again, our estimates also represent absolute forest
246 carbon losses, not net losses that incorporate biomass carbon gains which could not be calculated
247 from available data with confidence. Even with these uncertainties in mind, the acceleration of loss
248 in mountain forests with high carbon density that we find based on immediate vegetative biomass
249 changes alone portends additional redistributions and losses of carbon in the near future, potentially
250 nudging SEA's forests to be a net carbon source in the global carbon budget^{15,47}, rather than a
251 neutral actor⁵. To reduce the above uncertainties, future studies could integrate higher resolution
252 satellite and lidar datasets to map primary and secondary forests and related biomass carbon loss
253 more accurately. More studies on above and belowground carbon recovery associated with forest
254 regrowth are also needed.

255

256 In summary, our results reveal changing topographical patterns associated with forest loss in
257 Southeast Asia during the first two decades of the 21st century. The shift is characterized by an

258 upward expansion in the frontier of forest exploitation, from predominantly occurring in the
259 lowlands to increasingly encroaching forests at higher elevations with comparatively higher carbon
260 stocks and more sensitive species. The acceleration of this trend throughout the two decades
261 provides new insight regarding forest and carbon dynamics in the region that has not been
262 recognized in prior climate change assessments, nor parameterized in current model configurations
263 simulating impacts. Such exclusion misrepresents regional biophysical and biochemical feedbacks
264 of deforestation. Collectively, knowledge of the ascent of the frontier of forest loss across Southeast
265 Asia is needed to develop effective policies to manage concomitant negative impacts on
266 biodiversity, water resources, land degradation, and the carbon cycle. This knowledge is valuable
267 for developing strategies to reduce future losses of remaining forests that still have great ability to
268 preserve valuable ecosystem services, including atmospheric carbon dioxide capture and
269 biodiversity conservation.

270 **Methods**

271 This section provides details on the datasets and methods used for quantifying changes in
272 topographical patterns of forest clearance and related carbon loss across SEA.

273

274 **High-resolution forest cover change and primary forest extent products.** To quantify forest
275 cover change over SEA from 2001 to 2019, we used a high-resolution remote sensing product that
276 maps tree cover change at a spatial resolution of 30 m (version 1.7; ref. 8). The dataset has user's
277 and producer's accuracies of > 83% over the tropics⁸. A previous independent assessment indicated
278 that, in SEA, the data have user's and producer's accuracies of 93.2% and 81.2%, respectively¹⁹.
279 This dataset defines trees as "all vegetation taller than 5 m in height", and forest loss (including via
280 deforestation and forest degradation) as "the mortality or removal of all tree cover within a 30 m
281 pixel"^{8,25}. This operational definition results in the case that planted vegetation, such as rubber and
282 oil palm plantations, is mapped as trees when taller than 5 m. Removal of such vegetation is counted
283 as tree cover loss. Following these definitions, the data provide maps of forest cover loss and the
284 year of loss during 2001–2019 and forest cover gain during 2001–2012. Forest loss across SEA
285 exhibits a continuous increase trend from 2001–2019, confirming that changes in loss detection
286 method do not dominate the long-term trend. To separate forest loss type, we further used a dataset
287 on the extent of primary forests at 30 m spatial resolution for the year 2001 in SEA¹⁰.

288

289 **Topography data.** We used both mountain extent maps and a digital elevation model to quantify

290 the topographic pattern of forest loss. Mountain extent in SEA was mapped by a series of mountain
291 polygons developed by the Global Mountain Biodiversity Assessment (GMBA) inventory (version
292 1.2; ref. 18). The GMBA inventory defines a 2.5' pixel as mountainous if the geometrical amplitude
293 between the highest and lowest elevation exceeds 200 m. Following this definition, there are 61
294 mountain regions in SEA (Fig. S2a), occupying 1.7 million km² (38%) of SEA's land surface. The
295 remaining 62% of SEA's land surface is all treated as lowland. The associated elevation
296 information in lowlands and mountains, at a spatial resolution of 30 m, is collected from the
297 Advanced Spaceborne Thermal Emission and Reflection Radiometer Global Digital Elevation
298 Model (version 3; ref. 48). Slope information is estimated from elevation data using the average
299 maximum method⁴⁹.

300

301 **Forest carbon stocks.** We calculated forest carbon losses by incorporating high-resolution
302 aboveground live woody biomass (AGB) density map of Zarin et al.²¹ into our analyses of forest
303 loss. The map represents AGB density (in a unit of Mg biomass per hectare) at a spatial resolution
304 of 30 m circa 2000. The AGB map was generated using a random forest model and a statistical
305 model from measured forest biomass, GLAS lidar data, and gridded variables such as Landsat 7
306 ETM+ reflectance and biophysical variables, such as precipitation²¹. Owing to lack of data, we
307 estimate belowground biomass (BGB) at the pixel level with the empirical allometric model of
308 Mokany et al.⁵⁰ that has been widely used for BGB estimations^{2,51}: $BGB = 0.489 \times AGB^{0.89}$. Total
309 forest vegetation biomass, calculated as the sum of AGB and BGB, was converted to total forest

310 biomass carbon stocks using a conversion factor of 0.5 (refs. 2, 21).

311

312 **Forest and carbon loss calculations and analysis.** We estimated forest loss area by summing the
313 area of forest loss pixels that is dependent on its geographical location⁴⁴. The area of forest carbon
314 loss was calculated by overlapping the forest loss data with forest carbon stock density map
315 (including aboveground and belowground). We used committed emissions of carbon from forests
316 to the atmosphere upon forest loss, even though some of the carbon associated with tree removal
317 degrades on site or over time or is embedded within wood products¹⁵.

318

319 As both forest loss area and forest carbon loss showed near-uniform increases over time, we applied
320 a simple least-squares linear regression model to quantify the rate of change (Figs. 1a, 3a, S4, S5).
321 In contrast, because trends in mean elevation and slope of lands incurring forest loss in the 2000s
322 and 2010s were nonlinear (Fig. 1b), we used a piecewise linear regression model⁵²⁻⁵⁴ to (1)
323 determine where the trends in the time-series of mean elevation and slope change (i.e., inflection
324 points), and (2) quantify the trends before and after the inflection points. We also used a statistical
325 model in Real Statistics Resource Pack to test if the difference in trends between mountain forest
326 (carbon) loss and lowland forest (carbon) loss was statistically significant⁵⁵.

327

328 To demonstrate the spatial pattern of increases following inflection points, we separated them into
329 each 0.25° cell and used the equations:

330
$$H_{t,k} = \frac{\sum \bar{h}s_t + h_{t,k}s_{t,k}}{\sum s_t + s_{t,k}} \quad (1)$$

331
$$I_{t,k} = \frac{\sum \bar{i}s_t + i_{t,k}s_{t,k}}{\sum s_t + s_{t,k}} \quad (2)$$

332 where $H_{t,k}$ and $I_{t,k}$ are the mean elevation and slope in year t for the k th 0.25° cell. \bar{h} (245.5 m) and
 333 \bar{i} (9.3°) are the mean elevation and slope of forest loss across SEA after inflection points,
 334 respectively. $s_{t,k}$ and s_t are forest loss area in year t for the k th 0.25° cell and other cells, respectively.
 335 While the elevation and slope data for other cells are assumed to be the means of SEA (\bar{h} and \bar{i}),
 336 the elevation and slope data for the k th 0.25° cell are realistic. Thus, trends in the time-series after
 337 inflection points are caused by the changes only in the k th 0.25° cell. We then used a piecewise
 338 linear regression model to calculate trends in mean elevation and slope before and after identified
 339 inflection points. Following this method, we calculated the trends caused by each cell for countries
 340 (by summing all cells in each country), mountains (by summing all cells in mountains) and
 341 lowlands (by summing all cells in lowlands).

342

343

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351 inventory; Zarin/WHRC for providing the aboveground biomass density maps.

352

353 **Author contributions**

354 Z.Z. designed the research; Y.F. performed the analysis; Y.F. and A.D.Z. wrote the draft. All authors
355 contributed to the interpretation of the results and the writing of the paper.

356

357 **Competing interests**

358 The authors declare that they have no competing interests.

359

360 **Data availability**

361 The global map of forest cover loss and gain are available at
362 https://earthenginepartners.appspot.com/science-2013-global-forest/download_v1.7.html. The
363 ASTER elevation data are available at <https://earthdata.nasa.gov/>. The GMBA inventory is

364 available at https://ilias.unibe.ch/goto_ilias3_unibe_cat_1000515.html. The aboveground biomass
365 maps are available at <https://www.globalforestwatch.org/map/global/>. The primary extent data are
366 available at <https://glad.umd.edu/dataset/primary-forest-humid-tropics>. All datasets are also
367 available upon request from Z. Zeng.

368

369 **Code availability**

370 The scripts used to generate all the results are MATLAB (R2020a). Analysis scripts are available
371 at <https://doi.org/10.6084/m9.figshare.14586528>.

372

373 **Additional information**

374 Correspondence and requests for materials should be addressed to Z. Zeng.

375

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497 **Table legend**

498 **Table 1. Forest and related carbon loss in the mountains and lowlands of Southeast Asia**

499 **(SEA).** One asterisk (*) indicate statistically significant trends at level of $p < 0.05$.

500

501

502 **Table 1. Forest and related carbon loss in the mountains and lowlands of Southeast Asia (SEA).** One asterisk (*) indicate statistically
 503 significant trends at level of $p < 0.05$.

Variables	Year range	All forests			Primary forests			Secondary forests		
		SEA	Mountains	Lowlands	SEA	Mountains	Lowlands	SEA	Mountains	Lowlands
Gross forest loss (Mha yr ⁻¹)	2001–2019	3.22	1.00	2.22	0.93	0.26	0.67	2.29	0.74	1.55
	2001–2009	2.33	0.58	1.76	0.72	0.18	0.54	1.61	0.40	1.21
	2010–2019	4.02	1.38	2.64	1.11	0.33	0.78	2.91	1.05	1.86
Gross forest gain (Mha yr ⁻¹)	2001–2019	1.32	0.34	0.98	N.A.	N.A.	N.A.	1.32	0.34	0.98
Gross forest carbon loss (Tg C yr ⁻¹)	2001–2019	424	136	288	167	48	119	257	88	169
	2001–2009	330	88	242	128	33	95	202	55	147
	2010–2019	508	179	329	202	62	140	306	117	189
Forest loss acceleration (10 ⁻² Mha yr ⁻²)	2001–2019	17±3*	8±1*	9±2*	4±1*	2±0*	2±1	14±2*	7±1*	7±1*
	2001–2009	26±5*	6±1*	20±4*	11±2*	2±0*	9±1*	15±4*	3±1*	12±3*
	2010–2019	10±9	11±3*	-1±7	-5±3	1±1	-5±2	15±6*	10±2*	5±4
Forest carbon loss acceleration (Tg C yr ⁻²)	2001–2019	18±4*	10±1*	8±3*	7±2*	3±0*	4±2	11±2*	7±1*	5±2*
	2001–2009	35±7*	8±2*	27±5*	19±3*	4±1*	15±2*	16±5*	4±1*	12±4*
	2010–2019	1±12	10±4*	-9±8	-7±6	1±2	-9±4	8±6	9±2*	0±4
Trend in mean elevation (10 ⁻¹ m yr ⁻¹)	2001–2019	64±13*	46±15*	16±3*	50±17*	16±16	27±7*	52±11*	38±11*	7±2*
	2001–2011	1±19	11±28	0±5	-56±15*	-66±23*	-16±9	8±18	20±23	-3±4
	2011–2019	151±38*	95±57*	37±10*	195±30*	127±46*	85±16*	113±37*	62±46*	21±7*
Trend in mean slope (10 ⁻² ° yr ⁻¹)	2001–2019	11±2*	12±2*	3±1*	11±3*	6±3*	8±2*	9±2*	11±2*	0±0
	2001–2009	-4±3	1±5	-2±0	-17±0*	-14±0*	-9±0*	-4±0	2±4	-4±1*
	2009–2019	22±5*	20±8*	7±2	31±7*	21±8*	19±6*	19±5*	17±7*	4±2*

505 **Figure legends**

506 **Figure 1. Time-series of forest loss area and associated topography across Southeast Asia**
507 **during the period 2001–2019. a**, Annual forest loss area in lowlands (light pink bars) and
508 mountains (light blue bars) and the ratio of mountain forest loss area to total forest loss area (orange
509 line). Inset bars show trends in lowland and mountain forest loss area in the 2000s and 2010s.
510 Different letters above the bars indicate statistically significant differences ($p < 0.05$) between
511 trends for lowlands and mountains during the 2000s (black letters) and 2010s (red letters). **b**, Mean
512 elevation (solid black lines) and slope (solid red lines) of lands incurring forest loss. Dashed lines
513 are trend lines for mean elevation (black) and slope (red) before and after inflection points, which
514 were estimated by piecewise regression. Inset bars show trends in mean elevation (black) and slope
515 (red) before and after inflection points. Error bars indicate the standard error of linear trends. One
516 and two asterisks (*, **) indicate statistically significant trends at levels of $p < 0.05$ and $p < 0.01$,
517 respectively.

518
519 **Figure 2. Trends in mean elevation and slope of lands incurring forest loss following the**
520 **inflection points (IPs). a–b**, Trend in mean elevation (**a**) and slope (**b**) following the IPs in eight
521 countries of Southeast Asia (SEA), for all of SEA, lowlands, and mountains. Three countries in
522 SEA (Brunei, East Timor, and Singapore) are not presented here because their combined forest loss
523 is only 0.2% of the SEA total. The error bars indicate the standard error of the linear trend. One
524 and two asterisks (*, **) indicate statistically significant trends at levels of $p < 0.05$ and $p < 0.01$,

525 respectively. **c–d**, Spatial patterns of trends in mean elevation (**c**) and slope (**d**) of lands incurring
526 forest loss in 0.25° cells across SEA. Black dots indicate mountain regions. The IPs for mean
527 elevation and slope are around 2011 and 2009, respectively (see Fig. 1b). Trends in mean elevation
528 and slope of lands incurring forest loss in each 0.25° cell or each country (or in lowlands and
529 mountains) were calculated considering the weight of forest loss using Eqs. (1) and (2), respectively
530 (See *Methods*).

531
532 **Figure 3. Time-series of forest carbon loss and associated topography across Southeast Asia**
533 **during the period 2001–2019. a**, Annual forest carbon loss in lowlands (light pink bars) and
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538 letters above the bars indicate statistically significant differences ($p < 0.05$) between trends for
539 lowlands and mountains during the 2000s (black letters) and 2010s (red letters). **b**, Carbon loss in
540 elevation-year space.

541
542 **Figure 4. Spatial patterns of forest carbon loss across Southeast Asia during the period 2001–**
543 **2019. a**, Mean annual forest carbon loss in the 2000s. **b**, Change in mean annual forest carbon loss
544 in 2010s relative to 2000s. **c–d**, Trend in mean annual forest carbon loss in the 2000s (**c**) and 2010s

545 (d). Black dashed lines show mainland SEA (inside the box) and maritime SEA (outside the box).

546 Black dots indicate mountain regions.

547

548

549

550 **Figure 1. Time-series of forest loss area and associated topography across Southeast Asia**

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593







