

This is a repository copy of *The importance of inherent inorganics and the surface area of wood char for its gasification reactivity and catalytic activity towards toluene conversion*.

White Rose Research Online URL for this paper: https://eprints.whiterose.ac.uk/172994/

Version: Accepted Version

Article:

Korus, A., Ravenni, G., Loska, K. et al. (3 more authors) (2021) The importance of inherent inorganics and the surface area of wood char for its gasification reactivity and catalytic activity towards toluene conversion. Renewable Energy, 173. pp. 479-497. ISSN 0960-1481

https://doi.org/10.1016/j.renene.2021.03.130

© 2021 Elsevier. This is an author produced version of a paper subsequently published in Renewable Energy. Uploaded in accordance with the publisher's self-archiving policy. Article available under the terms of the CC-BY-NC-ND licence (https://creativecommons.org/licenses/by-nc-nd/4.0/).

Reuse

This article is distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs (CC BY-NC-ND) licence. This licence only allows you to download this work and share it with others as long as you credit the authors, but you can't change the article in any way or use it commercially. More information and the full terms of the licence here: https://creativecommons.org/licenses/

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk https://eprints.whiterose.ac.uk/

1	The importance of inherent inorganics and the surface area of wood char for its
2	gasification reactivity and catalytic activity towards toluene conversion
3	
4	Agnieszka Korus ^{a,1} , Giulia Ravenni ^b , Krzysztof Loska ^c , Irena Korus ^c , Abby Samson ^d , Andrzej
5	Szlęk ^a
6	
7	^a Department of Thermal Technology, Silesian University of Technology, Gliwice, Poland
8	^b Department of Chemical and Biochemical Engineering, Technical University of Denmark,
9	Lyngby, Denmark
10	^c Department of Water and Wastewater Engineering, Silesian University of Technology,
11	Gliwice, Poland
12	^d Department of Mechanical Engineering, University of Sheffield, Sheffield, UK
13	
14	Abstract
15	Gasification char is an effective catalyst for tar reforming because of the abundance of surface
16	active sites, which are available for heterogeneous conversion of hydrocarbons and interactions
17	with the reforming agents. This paper focuses on the importance of certain char properties for
18	the gasification and catalytic reforming. Specifically, the gasification reactivity of spruce char
19	is examined, along with its performance as a catalyst for toluene conversion. The material used
20	for this work was produced via gasification of spruce wood chips in the pilot TwoStage Viking
21	plant (Technical University of Denmark, Risø). To obtain a set of samples with varied surface

¹ Corresponding author, E-mail: agnieszka.korus@polsl.pl (Agnieszka Korus)

area characteristics and inorganic content, three pre-treatments were applied to samples of this char: acid washing, steam activation, and high-temperature treatment. The gasification and catalytic experiments performed with the untreated and modified materials revealed that the reactivity of the char during gasification in CO_2 depends mostly on the metal content in the sample, whereas the conversion of toluene was insensitive to the char inorganic content, but strongly correlated with the surface area available for heterogeneous reactions with toluene.

28

29 Keywords:

30 Char; gasification; inorganics; catalysis

31

32 1. Introduction

Biomass-derived char is the solid product of thermal conversion processes, such as pyrolysis 33 34 and gasification. In the past, it has been often disposed as a waste; yet, nowadays, it is 35 considered as an attractive material with versatile applications [1]. It is a heterogenous, carbon-36 rich substance, which generally has a porous structure and contains dispersed inorganics, such 37 as Ca, K and Mg, that may have been present in the parent feedstock. Because of its properties 38 and non-fossil origin, char represents a sustainable alternative to traditional materials and 39 resources employed in applications ranging from agriculture, environmental management [2] 40 and soil remediation [3], to energy storage [4] and catalysis [5].

Some types of char are particularly promising catalysts for hydrocarbon conversion, including reforming of the gasification tar for clean syngas production with a high H₂ content. Depending on the feedstock and the applied conditions, a diverse range of properties can be obtained; thus, the catalytic performance of the char in terms of tar conversion can vary from negligible to 45 nearly complete removal [6–9]. Several char properties are known to affect their catalytic
46 performance, including surface area, pore size distribution, O-containing surface
47 functionalities, and alkali and alkaline earth metal (AAEM) content, among others [8,10,11].

However, interactions between volatile hydrocarbons and the char matrix are complex, and the exact mechanism and the role of various active sites are still debated. For example, while the catalytic effect of AAEM species on char gasification is well-documented [12,13], conflicting conclusions were drawn regarding their role in the conversion of various hydrocarbons, such as methane [14–16].

Biomass-derived chars have a disordered, microcrystalline structure, which accommodates a substantial number of active sites in the form of unsaturated carbon atoms with unpaired electrons located at the crystallite edges and lattice defects [11,15,17]. Active sites catalyse reactions at the char surface; during char activation with CO₂ or H₂O, these sites can also be precursors for the formation of O-containing functional groups, such as carbonyl, carboxyl, and quinone moieties. Such O-containing functionalities can be formed under inert atmosphere as well, following the decomposition of certain polymeric constituents of biomass [10,14,18].

60 Metals incorporated in the carbonaceous lattice can also act as active sites. For example, C–O– 61 M or M₂O–C structures change the electron density at the neighbouring C atoms, thereby 62 enhancing the dissociative chemisorption of O from oxidising agents, such as H₂O or CO₂ [19– 63 22]. The O adsorption on the surface of the carbon lattice weakens the C-C bonds, allowing for 64 CO release from the char surface, which is the principle of char gasification. For this reason, AAEM species are widely-recognised catalysts for the oxidation and gasification of 65 66 carbonaceous materials. However, the role that AAEM species play in the catalytic hydrocarbon 67 conversion process has not yet been entirely elucidated.

Several studies have reported enhanced tar decomposition due to AAEM species [18,23,24]. However, these studies were carried out in the presence of gasifying agents, such as CO_2 or H_2O . Therefore, the distinct role of the metals in the heterogeneous tar conversion process was unclear because of the concurrent effect of increased gasification and activation of the catalyst by these elements.

73 Fuentes-Cano et al. [25] studied naphthalene conversion using char under inert atmosphere, and 74 observed a diminished catalytic activity for the acid-washed char. However, it is important to 75 note that the char used in their study was derived from coal, which has a dispersion of metals 76 that is different from that found in biomass [26]. Additionally, the char was acid-washed prior 77 to the steam activation, so the latter process was also influenced by the removal of inorganics. 78 Klinghoffer et al. [9] investigated the decomposition of methane over char and highlighted the 79 importance of inorganics (Ca, Na, K, Mg, P, Si, and Fe) dispersed on the carbonaceous surface 80 for catalysing the reactions. Similarly, Hervy et al. [27] studied the conversion of ethylbenzene 81 and reported the importance of Ca, P, Al, and K oxides for the catalytic performance. In contrast, Dufour et al. [15] reported that the removal of inorganics from the pine char did not 82 83 affect its methane reforming efficiency, and their observation was in agreement with some 84 previous reports on the lack of catalytic activity of metals [16,17,28].

85 Recently, Feng et al. [24] reported a decrease in tar formation during gasification of K-loaded 86 and Ca-loaded biomass at temperatures up to 700 °C. However, at all examined temperatures 87 (500 - 900 °C), the relative yield of single-ring aromatics, especially toluene, increased 88 significantly when AAEM-rich feedstock was used. Furthermore, in their studies regarding tar 89 decomposition over activated biochar, Feng et al. [29] observed that the coke formation, which 90 occurred primarily at AAEM sites, originated mainly from oxygenated hydrocarbons. These 91 results suggest that O-containing hydrocarbons are more susceptible to the catalytic effects of 92 AAEM species, whereas the decomposition of other tar components, such as light aromatics

with no heteroatoms, may be insensitive to the presence of these elements, and thus requireindependent investigations.

In this work, toluene was chosen as a representative of this class of compounds (i.e., light aromatics containing no heteroatoms). Studies involving this molecule are particularly relevant because toluene is among the most refractory and difficult tar species to remove from producer gas [7]. The objective of this work was to evaluate how the AAEM species inherently present in the char affect the heterogeneous conversion of toluene and the overall char gasification reactivity. The role and influence of inorganic species were also compared with the impacts of the surface area, porosity, and surface functional groups of each examined char.

102 Char derived from the gasification of spruce wood was selected for this study. With the goal of 103 studying the effects of selected char properties, the obtained char was modified by dedicated 104 pre-treatments (i.e., acid washing, steam activation, or high-temperature heat treatment), which 105 can remove inorganics and modify the surface area and porosity of the char surface.

106 Prior to the toluene decomposition experiments, the gasification reactivity of chars was 107 evaluated via thermogravimetric analysis in the presence of CO₂ as an oxidiser. The 108 experiments involving toluene conversion and char gasification were conducted at 800 °C, 109 which is above the minimum temperature where the catalytic effect of alkali metals is observed 110 (approx. 700 °C) [19]. Toluene decomposition was carried out under an inert atmosphere to 111 avoid the interference of the catalytic effect of inorganics on the char oxidation. Steam 112 reforming tests were performed to verify the combined effects of AAEMs in the presence of 113 both tar and a gasifying agent.

The novelty of this work lies in the parallel evaluation of the effects of selected char properties on gasification reactivity and catalytic performance of char for the decomposition of toluene. The results provide valuable insights into the relative importance of different key properties of

the char surface, which can be applied for the production of tailored carbonaceous catalyticmaterials.

119

120 **2.** Experimental (Materials and Methods)

121 **2.1. Materials**

122 The char used in this study was generated from spruce wood chips using a pilot TwoStage 123 gasification plant (the Viking gasifier) operated at Technical University of Denmark (DTU) 124 Risø, in Roskilde, Denmark. Within the TwoStage process, biomass undergoes pyrolysis during 125 the first step, and it is then gasified in a separated reactor. In this second stage, an air inlet allows 126 for partial oxidation of the tar and locally increases the temperature up to 1100 °C. Additional 127 details about the TwoStage gasification process can be found in the literature [30,31]. The 128 Viking char (VC) used in this study was collected directly from the solid residues of the 129 gasification plant. The char from the Viking gasifier has a large surface area with well-130 developed mesoporosity resulting from its participation in the reforming process. These 131 properties made it an efficient catalyst for the conversion of tar model compounds and real tar 132 mixtures [32,33]. The diversified catalyst properties investigated in this work were obtained by 133 applying pre-treatments to subsampled batches of the original char. The final VC has a stable 134 carbon structure because of the high temperatures reached during the TwoStage process. 135 Therefore, it can be assumed that the applied pre-treatments induced negligible variations in 136 the carbonaceous structure, i.e., the aromaticity or microcrystalline arrangement. These features 137 contribute significantly to the catalytic performance of chars, and they are closely related to the 138 type of feedstock and the conditions of char formation [11]. Thus, it is crucial that the 139 experimental modifications maximise the changes in the AAEM content and surface area while 140 maintaining a similar carbon structure in all studied samples. Prior to any treatment, VC was ground in an agate mortar and sieved to isolate particles in the 250-600 μm size range, selected
based on the reactor size and the reports on similar studies [8,32].

143 Three different pre-treatments were applied to three different subsamples of the raw material 144 (VC), specifically, (i) acid (HCl) washing to remove inherent inorganics, (ii) steam activation 145 to increase the surface area, and (iii) thermal treatment at 1000 °C under N₂ to reduce char 146 porosity. In the first pre-treatment, aimed at the removal of inorganics, two cycles of acid wash 147 were performed by shaking the subsample of VC in 2M HCl solution for 24 h. After the second 148 cycle, the char was washed with distilled water to remove Cl⁻ ions (verified by addition of 149 AgNO₃ to the filtrate). The char was then vacuum filtered and dried at 60 °C under N₂ flow 150 until a constant mass was reached. This modified VC char was referred to as VC_HCl. To 151 increase the surface area, another VC subsample was subjected to steam activation by heating 152 a fixed bed of char up to 800 °C in a quartz tube reactor under N₂ flow, followed by 120 min of 153 isotherm under a flow of 15/85 vol.% of H₂O/N₂. The char was then cooled to ambient 154 temperature under N₂ flow and referred to as VC_SA. This pre-treatment also increased the 155 AAEM concentration in the char due to the carbon burn-off. The last modification involved the 156 heat treatment of the VC. A fixed bed of the third VC subsample was heated at a rate of 10 157 K/min up to 1000 °C, maintained at that isothermal condition for 180 min, and then cooled to 158 ambient temperature, all under N₂ flow. The goal of this procedure was to modify the pore 159 distribution of the sample by inducing pore collapse and/or sintering. The obtained material was 160 referred to as VC_HT.

161 These pre-treatments produced a set of samples with similar carbonaceous structure yet varying 162 surface areas and concentrations of inorganics, thus enabling investigations into the 163 individualised effects of these two parameters on the activity of char.

165 **2.2.** Char analysis

Different techniques were applied to characterise the fresh chars and verify the outcomes of the pre-treatments. In addition, the spent chars recovered after 50 min of toluene conversion tests were analysed to determine the effect of the heterogeneous reaction on the char properties.

The C, H, N, and S (CHNS) contents were determined with an elemental analyser (EA3000, EuroVector, Italy) and assessed on a dry basis. Since negligible amounts of S and N were detected in the samples, only the H/C ratios are reported in this work. At least three repetitions of each measurement were carried out, and the averaged values with standard deviations are reported.

The Na, K, Ca, Mg, and Fe contents in the bulk of the char were analysed with a SpectrAA 880 Varian Atomic Absorption Spectrometer (AAS) after sample digestion with HNO₃ in a Milestone MLS1200 MEGA microwave. Because of the strong heterogeneity of the chars, four repetitions of each sample measurement were performed, and the averaged values with standard deviations are reported.

The surface functional groups on the chars were examined via attenuated total reflectance Fourier transform infrared (ATR-FTIR) spectroscopy using a Perkin Elmer Spectrum 100 spectrometer with a Universal Attenuated Total Reflectance module equipped with a germanium crystal. Each spectrum was constructed from 32 scans, with a resolution of 4 cm⁻¹ within the range of 4000-700 cm⁻¹ wavenumbers. The averaged spectra of five measurements are reported herein for each material.

The surface area and pore size distribution (PSD) of each sample were analysed with a Micromeritics TriStar II 3020 analyser. The samples were outgassed under N₂ flow at 200 °C for 24 h prior to the analysis; the adsorption-desorption isotherm of N₂ at 77 K was performed for the relative pressures (p/p_0) in the range of 0-0.95, and the CO₂ adsorption at 273 K was

189 carried out for p/p_0 up to 0.03. Two separate calculations regarding the surface area of each char 190 were performed using the Brunauer-Emmett-Teller (BET) model, specifically, one from the N₂ 191 adsorption isotherm that accounted for the mesopores and some of the micropores, and the other 192 for the CO₂ adsorption data, which accommodated the whole micropore range. Additionally, 193 for comparison, the surface area of the micropores was estimated using the t-plot method 194 applied to the N₂ adsorption data and using a statistical thickness curve (Carbon Black STSA). 195 All measurements were carried out in duplicate, and the averaged values with standard 196 deviations are reported. The PSD was calculated using the two-dimensional version of non-197 local density functional theory (2D-NLDFT) [33] and the kernels for slit pore carbon that were 198 included in the Micromeritics software. Combining data from N₂ and CO₂ adsorption allows 199 for a more detailed PSD analysis of the materials wherein a significant fraction of pores are narrower than 7 Å [33]. 200

The structure and chemistry of the char surface were also evaluated with a scanning electron microscope (Prisma E, Thermo Scientific) coupled to an energy dispersive spectroscopy detector (SEM-EDS). For each sample, several transversal and longitudinal sections were examined to account for the strongly heterogenous structure of the chars.

205

206 **2.3. Reactivity and kinetic parameters calculation**

Char reactivity during gasification with CO_2 was examined by thermogravimetric analysis (TGA) performed with a Netzsch STA 409 LUXX instrument, according to the following procedure: 5 mg of the sample was heated up to 1000 °C at a 10 K/min rate under 100 mL/min N_2 flow, and then cooled to the desired gasification temperature (800, 850, or 900 °C); after 15 min of N_2 purging, the flow was changed to 50/50 vol.% CO_2/N_2 , and the char was gasified under isothermal conditions until a constant residual mass was reached. For all fresh and spent chars, the gasification was carried out at 800 °C, and the registered mass loss was used to calculate the gasification reactivity. Additionally, for VC and VC_HCl, the tests at 850 and 900 °C were performed to determine the kinetic parameters of the gasification reaction. All tests were carried out in duplicate; the averaged values with corresponding standard deviations are reported for the char reactivities. For the kinetic parameters, to account for a kinetic compensation effect, raw data from both repetitions were averaged prior to calculating the activation energy and the pre-exponential coefficient.

220 The reactivity at 50 % conversion (R_{50}) can be determined using Eq. 1 [34],

221
$$R_{50} = k_0 \exp(-E/RT) p_{CO_2}^n$$
(1)

where k_0 is the pre-exponential factor (bar⁻ⁿ·s⁻¹), *E* is the activation energy (kJ·mol⁻¹), *R* is the gas constant (kJ·mol⁻¹·K⁻¹), *T* is the absolute temperature (K), p_{CO_2} is the partial pressure of CO₂ (bar), and *n* is the reaction order. This analysis was applied to examine the influence of the pre-treatments on the gasification of the VC char.

The representative R_{50} value for the 20-80 % conversion of each sample was calculated using data obtained from the TGA tests of char gasification at 800 °C, according to the procedure described by Ollero et al. [34]. The conversion of the sample was determined from Eq. 2

229
$$X_i = (m_0 - m_i)/(m_i - m_a)$$
(2)

where m_0 is the initial mass of the sample, m_i is the mass at each time, and m_a is the mass of the ash that remained after the conversion.

232 The reactivity R_i at each time was calculated for each datapoint using Eq. 3

233
$$R_i = (-dm_i/dt)/(m_i - m_a)$$
(3)

The first estimate of normalised structural profile function for each datapoint was determinedfrom Eq. 4

236
$$F_i = R_i / R_{50}$$
 (4)

where R_{50} is the first estimate of the reactivity at 50 wt.% conversion. Then, the 5th order polynomial function was fitted to the $F_i - X_i$ data to obtain the normalised structural profile F(X).

Eq. 5, Finally, the weighted mean value of R_{50} was calculated using Eq. 5,

241
$$R_{50} = \frac{\sum_{i} \omega_{i} R_{i} / F(X)}{\sum_{i} \omega_{i}}$$
(5)

242 where ω_i is the statistical weight determined as $\omega_i = R_i(1 - X_i)$.

The kinetic parameters for VC and VC_HCl were calculated based on the reactivities measured at 800, 850, and 900 °C, under the assumption of an Arrhenius-type temperature dependence, using the linear equation presented in Eq. 6:

246
$$\ln R_{50} = \ln k_0 - \left(\frac{E}{RT}\right) + n \ln(p_{CO_2})$$
(6)

Only one partial pressure of CO_2 was examined; thus, the reaction order, *n*, was assumed *a priori* to be 0.36, based on the data available in the literature [34].

To correlate the reactivity of the chars with their specific properties, the R₅₀ values were used as a parameter to link the specific surface area of the samples with their inorganic content, represented by K, which is one of the most reactive AAEM.

252

253

255 **2.4.** Toluene conversion tests

The catalytic performances of the untreated and modified chars during the conversion of toluene were assessed by examining its conversion over a fixed bed of char in a laboratory quartz tube reactor. Specifically, the heterogenous conversion of toluene over the different chars was assessed in pyrolytic conditions (N₂ atmosphere) at 800 °C, for either 30 or 50 min. The VC and VC_HCl samples were also tested under oxidising conditions, with 15 vol.% of steam in N₂, during 50 min of steam reforming of toluene at 800 °C.

262 The test rig, described in detail in previous work [35], allowed for the pre-mixing of gaseous 263 reactants in the evaporator. Toluene (and water if steam was required) were fed through a 264 syringe pump into the individually-controlled heating section filled with quartz wool and 265 purged with N₂. Vapours were then introduced into the main reactor containing the fixed char 266 bed maintained at the desired reaction temperature. The condensable reaction products were 267 captured downstream of the reactor in the cooled impinger bottles filled with dichloromethane, 268 and analysed post-run with GC-FID; the permanent gases were analysed online with GC-TCD. 269 The blank tests revealed that the efficiency of the sampling train was above 98 % and the toluene 270 thermal decomposition in the absence of a catalytic bed was negligible. The procedures for the 271 pyrolytic and reforming tests, detailed in previous work, were followed closely [35,36]; the 272 only difference was that the gas analysis was performed online with a 10 min interval required to separate the released gases. Toluene was fed with a constant concentration of 12 g/Nm³ for 273 274 either 30 or 50 min. Two types of reaction were examined: pyrolytic conversion under pure N_2 , 275 and steam reforming under 15/85 vol.% H₂O/N₂ flow. Temperature, and toluene and steam 276 concentrations that were applied for the conversion tests were based on the reports on similar 277 studies [7,9,35,36] to enable the comparison with other catalytic materials. After each measurement, the char bed was cooled to ambient temperature under N₂ flow; the 278 279 physicochemical properties of the spent materials were analysed along with the fresh chars to evaluate the changes that occurred during the reactions with toluene (and steam for the reforming tests). All runs were carried out in duplicate, and the averaged values with standard deviations are reported. The results of the toluene conversion tests were used to calculate the toluene conversion, η_{T} , from Eq. 7,

284
$$\eta_T = (m_f - m_r)/m_f \cdot 100\%$$
 (7)

where m_f and m_r represent the mass of toluene (mg) fed into the reactor and recovered in the impingers, respectively. The η_T values were then used to estimate the catalytic performance of each material and to determine the correlation between specific properties of the tested chars, as reported in Sections 3.3 and 3.4, respectively.

289 An overview of the char samples produced from pre-treatments of VC and the corresponding 290 spent samples after toluene conversion is presented in Fig. 1. The pyrolytic conversion tests 291 (N₂) allowed for evaluation of the direct roles of inorganics and surface area of char on the 292 heterogenous reactions with toluene. The reforming tests (Steam) were aimed at the evaluation 293 of the indirect role of inorganics, i.e. their catalytic effect on char oxidation that occurred 294 concurrently, yet independently, to the toluene conversion, when steam was present. Thus, the 295 reforming tests were limited to the original char (VC) and the sample with the removed 296 inorganics (VC_HCl).



Fig. 1. Schematic describing the toluene decomposition experiments under nitrogen or steamcatalysed by untreated and modified VC.

300

301 **3. Results and discussion**

302 3.1. Physicochemical properties of chars

The applied pre-treatments provided a set of chars with various properties, and the fresh chars were analysed to relate these properties to the observed gasification reactivity and activity towards toluene removal. The spent chars were also examined to better understand the nature of the heterogenous conversion of toluene. The analysis of the CHNS content in the fresh and spent samples revealed either non-detectable or negligible amounts of N and S in all chars; the atomic H/C ratios are presented in Fig. 2. The untreated Viking char (VC) had the highest share of H atoms in the structure, and all applied pre-treatments resulted in a decrease of the relative amount of H. This was likely because of the dehydrogenation of the char under high temperatures (VC_HT), the formation of larger aromatic structures due to interactions with steam (VC_SA), and the removal of lighter H-rich labile compounds during the acid wash (VC_HCl).

The coke formed during tar conversion is usually low in H because of the dehydrogenation of aromatic tar compounds that occurs during their polymerisation on the catalyst surface. Thus, the chars recovered after toluene pyrolysis had lower H/C ratios, with the exception of VC_SA. As revealed by the comparison of VC and VC_SA, the steam activation drastically diminished H/C ratio of the char. Therefore, it is possible that the toluene-derived coke deposited on the spent char had higher H content than the fresh VC_SA, resulting in the increased H/C ratio in VC_SA_50N.



Fig. 2. The H/C atomic ratios in the untreated (VC) and modified (VC_HCl, VC_SA, VC_HT)
Viking chars, both fresh and recovered after 50 min of toluene pyrolysis (50N) or steam
reforming (50S)

325

326 **3.1.1. Metal content in chars**

327 The AAS results illustrating the content of the most common AAEMs and Fe in biochars are 328 presented in Fig. 3. Despite the strong heterogeneity of the samples, which leads to high 329 uncertainties in the measured values, some statistically significant differences were observed. 330 The acid washing removed the AAEM species almost completely; K content was reduced by 331 99 wt.% and Ca by 98 wt.% in the VC_HCl sample. Fe was more resilient to the treatment; 332 only 76 wt.% of this element was removed, making Fe the most abundant metal in the analysed 333 VC_HCl. Steam activation of VC caused carbon burn-off, resulting in a 2-3 times higher 334 concentration of metals in the VC_SA, compared with the untreated char. A slight reduction in 335 the concentration of inorganics in the bulk of the char was observed after the high-temperature 336 thermal treatment (VC_HT), likely due to partial vaporisation of the ash fraction.

Toluene conversion under an inert atmosphere led to intense coke formation, resulting in an increased carbon content in the spent catalyst. However, no significant decrease in the concentration of inorganics was observed, except in the case of the VC_SA char.

During toluene decomposition in the presence of steam, the steam reforming of carbon caused an increased concentration of metals in the VC_50S. The same effect was not observed in the VC_HC1_50S sample, which underwent less intense gasification reactions, as discussed further in Section 3.3.



Fig. 3. Content of selected metals in the fresh Viking char (VC), acid-washed VC (VC_HCl),
steam-activated VC (VC_SA), and high-temperature annealed VC (VC_HT), as well as in spent
chars recovered after 50 min of toluene conversion under pure nitrogen (50N) or 15 vol.% of
steam (50S)

354 **3.1.2. SEM-EDS analysis**

The dispersion of the reactive elements on the surface of each char was assessed using SEM-EDS. As shown in Fig. 4, the surfaces of all char samples were strongly heterogeneous, and the measured concentrations varied in different areas of the same material. Therefore, the mean concentration values for the analysed surfaces were used to evaluate the differences between the materials.

Unsurprisingly, C (not shown) and O were the most abundant elements detected on the surface of all samples. The C concentration on the surface of the char samples was in agreement with the values obtained from elemental analysis. In comparison with VC, the chars obtained following all applied pre-treatments and the spent chars all contained reduced amounts of O on their surface. The diminished oxygen presence in the EDS spectra is consistent with the destruction of some O-containing surface functionalities, as revealed by ATR-FTIR analysis.

366 The most abundant inorganic species detected on the surface of fresh char samples were Ca and 367 K, followed by Fe and Mg, in good agreement with the AAS results. In addition, Si and traces 368 of S, P, and Na were detected. The surface concentrations of some AAEM species, namely K, 369 Ca, and Mg, were higher than the values measured with the AAS technique, which examined 370 the bulk volume of the sample, suggesting that these elements accumulated on the char surface. 371 The VC is a gasification residue, meaning it is the product of numerous volatile-solid 372 interactions, which are known to induce the migration of AAEM species towards the char 373 surface [10,37].

374 Compared with VC, higher concentrations of K and Ca were detected on the surface of VC_SA 375 and VC_HT samples. Since the heat treatment did not significantly change the bulk 376 concentration of these metals, the presence of these inorganic regions at the surface suggest that 377 the applied thermal treatment may have caused the migration of metal atoms towards the surface of the catalyst, as previously observed by Klinghoffer et al. [14]. The accumulation of metals on the surface of the VC_SA char can be explained by (i) the migration of these elements towards the surface [38] and (ii) the increased concentration of inorganics in the bulk of the char due to carbon burn-off, as confirmed by the AAS analysis detailed in Section 3.1.1.

382 The EDS analysis confirmed the successful removal of inorganic species from the VC_HCl 383 sample, which contained significantly lower concentrations of Ca, K, and Mg compared with 384 the other samples. Only Si and Cl appeared to persist on the surface of VC HCl. The presence 385 of Si can be explained by its poor solubility in acid, while the presence of Cl was most likely a 386 consequence of the acid wash treatment, during which Cl⁻ ions bound to the carbon surface and 387 were not removed by rinsing the char with deionised water. Cl was also present in a small 388 concentration (<0.5 wt.%) in the VC and VC_SA samples, but among the spent chars, only 389 VC HCl 50N had traces of Cl on the surface. The lack of Cl on the spent samples suggests that 390 the high temperature caused the desorption of this element.

391 The SEM micrographs correlated well with previously published images of wood char 392 [14,37,39], as illustrated by the selected SEM images presented in Fig. 5. The VC exhibited a 393 fibrous, anisotropic structure typical of softwood, which was evidently preserved in the charred 394 material (Fig. 5a). Most of the EDS scans revealed that AAEM were dispersed across the 395 surface of the chars. In most cases, these elements could not be recognised visually in the SEM 396 micrograph due to their small size and good dispersion. Only a few round-shaped irregular 397 clusters of inorganics, as described by Fuentes-Cano et al. [39], were visible (Fig. 5b). On the 398 surface of VC_HT, agglomeration and sintering of inorganics were observed, as shown in Fig. 399 5c. This phenomenon was similar to that reported by Klinghoffer et al. [14,37] for chars treated 400 with high temperature (1000 °C).



- 406 Fig. 4. The concentration of selected elements determined from SEM-EDS analysis
- 407 (horizontal lines indicates the average values for each sample)
- 408

The structural changes resulting from the char pre-treatments occurred mainly in the nanoscale, below the resolution of the SEM micrographs. Similarly, no visible modification could be observed following the heterogeneous decomposition of toluene. This can be explained by the relatively short reaction time (50 min). Structural changes caused by cracking reactions on the surface of char have been reported after several hours of char exposure to aromatic compounds. However, the char that was used in those experiments was produced at temperatures below 900 °C, thus it may have been less thermally-stable than VC [6,39].









421

10 µm

10 µm

c)



424 Fig. 5. SEM micrographs of the VC chars: (a) an anisotropic char grain of untreated VC; (b)

- 425 cluster of inorganics (Ca and Si) on the surface of VC_SA_50N; (d) inorganics' migration
- 426 towards the surface and sintering of the VC_HT char

427

423

422

429 **3.1.3. Surface functional groups**

430 The chemistry of the char surface was examined using ATR-FTIR to verify the effect of the 431 pre-treatments on the surface functional groups and evaluate potential correlations between the 432 functional group content and the catalytic performance of the different materials. The spectra 433 of fresh chars are presented in Fig. 6, and the results for the spent chars are provided in the 434 Supplementary Information. The relatively weak absorption at high wavenumbers suggested 435 insignificant amounts of hydroxyl groups or aliphatic structures in the examined samples 436 [40,41]; O-H and C-H bending motions would be associated with strong peaks around 3700-3000 and 2900-2800 cm⁻¹, respectively. Instead, the observed bands originate mainly from the 437 438 vibrations of aromatic structures and various carbon-oxygen bonds.

439 Four characteristic absorption regions can be distinguished in the spectrum of VC, specifically, sharp bands at 1566, 1427, and 874 cm⁻¹ and increased intensity in the fingerprint region 440 441 between 1300-900 cm⁻¹. The sharp bands are most likely related to the aromatic rings in the 442 char. The carbon-oxygen bonds in structures with pronounced resonance effects (i.e., where the 443 energy of the bond is between that of a single and a double bond) may be also responsible for the absorption at 1566 cm⁻¹ and 1427 cm⁻¹, e.g. the asymmetric $v_{as}(CO_2)$ and the symmetric 444 $v_s(CO_2)$ stretching of carboxylates, respectively [42,43]. The band at 874 cm⁻¹ is indicative of 445 446 one predominant substitution pattern in the aromatic structures, which involves only one or two 447 neighbouring H atoms bound to the ring [41]. The increased absorption in the fingerprint region between 1300-900 cm⁻¹ can be attributed to the overlapping frequency ranges for the vibrations 448 449 of single C–O bonds in various conformations, such as ethers and esters, as well as C–C bonds 450 in the carbonaceous matrix of the char. Moreover, a weak, broad band around 3700-3000 cm⁻¹ 451 suggests a small amount of lattice water or a trace amount of hydrogen-bonded hydroxyl groups. The increased absorption around 2600-1700 cm⁻¹ may indicate the presence of some 452

453 C=O groups, such as lactones or quinones, or it may result from the overtones of aromatic ring454 stretching.

455 All the treatments applied to VC decreased the char absorption in certain regions, indicating 456 that the pre-treatments caused a significant loss of surface functionalities. Acid washing 457 (VC_HCl) led to almost complete disappearance of the 1427 and 874 cm⁻¹ bands, indicating the 458 destruction or rearrangement of carboxylate ions and changes in aromatic rings substitution, 459 respectively. Steam activation (VC_SA) had a similar effect, with an additional decrease of the 460 1560 cm⁻¹ band and slightly diminished absorption in the fingerprint area. The spectrum of the 461 thermally-treated char (VC_HT) was drastically flattened, suggesting major decomposition of 462 the surface functionalities. Acid washing most likely protonated carboxylates and changed the substitution pattern in aromatic structures, thus shifting the sharp 874 cm⁻¹ band towards 800 463 464 cm⁻¹. Carboxylates thermally decompose around 440 °C [44], so it is possible that they were 465 released from the char surface during heating under an inert atmosphere, or even in the presence 466 of steam, despite its oxidising effect, thus removing the 1427 cm⁻¹ band from the VC_HT and 467 VC_SA char spectra. The depleted absorption of the VC_SA sample, compared with VC, 468 suggests that the atmosphere in the second stage of the Viking gasifier where the untreated char 469 was created, was more reactive than the steam pre-treatment conditions. Therefore, although 470 the applied steam activation enhanced the surface area development of the VC SA char (discussed further in Section 3.1.4), it was not sufficient to maintain the O-containing 471 472 functionalities of the VC.





475 Fig. 6. ATR-FTIR spectra of the fresh chars: untreated VC, acid-washed VC_HCl, steam476 activated VC_SA, and thermally-annealed VC_HT

478 **3.1.4. Surface area and pore size distribution**

479 The structures of the fresh and spent char samples were examined by measuring the N₂ 480 adsorption at 77 K and the CO₂ adsorption at 273 K. The adsorption branches of both isotherms 481 were then combined to obtain the pore size distribution (PSD), presented in Fig. 7, using the 482 2D-NLDFT model. All samples contained significant quantities of micropores, among which, two characteristic groups could be distinguished, namely, ultramicropores (<7 Å [45]; here, 3-483 484 5 Å) and supermicropores (10-20 Å). Additionally, some mesopores (>20 Å) with a wide 485 variety of sizes were observed in all PSD plots. The incremental area of the ultramicropores in the VC reached 70 m²/g, and these structures were preserved following acid wash (VC_HCl) 486 487 and thermal pre-treatments (VC_HT), although steam activation decreased the area to $60 \text{ m}^2/\text{g}$ 488 (VC_SA). The main peak in the supermicropores' distribution curve (at 11 Å) increased by 10 m^2/g due to acid washing and by 20 m^2/g due to steam activation, whereas thermal annealing 489

490 resulted in a 10 m²/g decrease. The area of larger supermicropores and mesopores was similar 491 for all samples, except the VC SA char. The consumption of carbon during steam gasification 492 has a pore-widening effect [46], which was likely responsible for the decrease in 493 ultramicroporous area and the increase in the supermicroporous and mesoporous area of the 494 steam-activated material. High-temperature annealing resulted in pore collapse and/or sintering, 495 thereby diminishing the area of supermicropores in VC_HT. In contrast, the removal of some 496 labile carbon species during the acid wash may explain the increased microporosity of the 497 VC HCl char.

498 Heterogeneous toluene conversion over the catalyst surface results in coke deposition inside 499 the pores, thus decreasing the surface area of the material. In general, micropores are relatively 500 more prone to blockage and deactivation [6,29,47]; therefore, more mesoporous structures 501 typically have higher longevity. Indeed, the PSD in the spent chars recovered after 50 min of 502 catalytic toluene conversion in an inert atmosphere revealed a significant decrease in their 503 supermicroporous areas. In particular, VC_50N and VC_HT_50N suffered the most severe 504 deactivation and their pores between 10-15 Å became completely filled with coke. In the 505 slightly more resilient VC_HCl_50N and VC_SA_50N chars, residual amounts of pores with 506 sizes of ca. 14 Å were preserved. This observation suggests that supermicropores were 507 progressively filled from the smallest to the largest, during coke deposition. Interestingly, the 508 surface area contained in ultramicropores did not diminish in the spent samples. Since these 509 pores are smaller than the kinetic diameter of toluene (5.85 Å [48]), they likely did not 510 participate in toluene conversion. Their presence in the PSD of spent chars suggests that the 511 entrance to those pores was not blocked by deposited coke, so it is plausible that they were not 512 embedded in the supermicropores, but rather, were directly accessible from the meso and 513 macropores, which were not completely filled with coke. The observed changes in the pore 514 distribution following catalysis suggest that (i) the coke deposition occurred primarily in the 515 micropores and (ii) the ultramicroporous structures did not participate in the catalytic 516 conversion of toluene, so they were resistant to deactivation. The general exclusion of 517 micropores from catalysing tar conversion because of diffusional limitations was previously 518 reported by Fuentes-Cano et al. [39]. In this study, the inclusion of CO₂ adsorption 519 measurements in the PSD calculation provided additional information on the ultramicropores, 520 allowing the conclusion that only the smallest micropores did not contribute to toluene 521 conversion, and the critical size of the pores participating in the conversion correlate well with 522 the kinetic diameter of the compound. The results of the 2D-NLDFT model also suggested that 523 micropores that are too narrow for toluene did not get blocked by the coke deposited at their 524 entrance, but remained unaffected by both the toluene conversion and the deactivation 525 processes. Verification of this proposal requires further investigation, e.g., using additional 526 samples to determine if discussions of biochar microporosity should distinguish between 527 ultramicropores and supermicropores.

528 During toluene conversion in the presence of steam, the gasification of the char and the 529 deposited coke occurred, thus the pore structure was preserved. This phenomenon was more 530 pronounced in the case of VC, which demonstrated greater reactivity than VC_HCl during the 531 gasification, as discussed further in Section 3.2. Most of the original pores were preserved in 532 the spent VC 50S char, and the slight decrease in the microporous area was most likely the 533 result of the pore-widening effect of carbon consumption during reactions with steam. In 534 contrast, since the VC_HCl char was more resistant to gasification, the beneficial role of steam 535 was limited in this case. The VC_HCl_50S char had fewer micro and mesopores, although the 536 pore blocking was less intense compared with the tests under an inert atmosphere 537 (VC_HCl_50N). Toluene reforming with VC_HCl char confirmed that the coke deposition 538 occurred preferentially in the smaller supermicropores (~11 Å), as the area corresponding to 539 these structures decreased the most following catalysis.



Fig. 7. NLDFT pore size distributions calculated from the combined results of N₂ and CO₂
adsorption

545 The total surface area of the samples was calculated using a BET model based on the N₂ adsorption isotherm (Fig. 8). These data were also used to estimate the area of the micropores 546 547 with the t-Plot method. The BET model was also applied to the CO₂ adsorption data; the 548 calculated surface area accounts for micropores up to approximately 10 Å. It was determined 549 that VC had a very high surface area, comparable with that of activated carbons. Both acid wash 550 and steam activation effectively increased the surface area, as a result of the removal of 551 inorganics and carbon burn-off, respectively. In contrast, the heat treatment decreased the BET-552 determined surface area due to pore collapsing.

553 Coke deposition during the pyrolytic toluene conversion significantly diminished the total 554 surface area of all four examined spent chars. Conversely, the area of micropores determined 555 by the CO₂ measurements was only slightly diminished; most of these pores were the 556 ultramicropores (<5 Å), which were not affected by the coke deposition according to the PSD 557 results. It should be noted that the microporous area calculated with t-Plot method was similar 558 to that derived from CO₂ adsorption for the fresh chars. However, there was a large discrepancy 559 between the t-Plot and CO₂ adsorption results from the spent char analysis. It is plausible that 560 the coke deposit effectively limited N₂ diffusion into the smallest pores, while the CO₂ was still 561 able to enter those structures. Therefore, the assessment of the microporous carbons with CO₂ 562 adsorption proved to be crucial for understanding the structural changes in the char during 563 deactivation with coke deposition. Furthermore, these results confirmed that the incorporation 564 of data from the CO₂ measurement into the PSD calculation is necessary to obtain a realistic 565 distribution of the ultramicropores. The microporous areas determined from t-Plot and CO₂ data 566 were consistent for the material recovered after steam reforming of toluene over the VC because 567 the structure of the VC_50S sample was well-preserved. However, for the partially coke-568 covered VC HCl 50S char, the t-Plot method again underestimated the area of the micropores 569 by not accounting for the ultramicropores.

570 Since the ultramicropores did not participate in toluene conversion, the term "useful surface 571 area" of the char was introduced, and this value was determined as the integrated area under the 572 PSD plot for the pores wider than the toluene kinetic diameter (5.85 Å). This surface parameter 573 was further utilised to correlate the catalyst surface development with the toluene conversion, 574 as reported in Section 3.4.

575



Fig. 8. The chars' surface areas calculated with a BET model based on N₂ and CO₂ adsorption
data, and the microporous area calculated with a t-Plot model applied to the N₂ measurements

579

580 **3.2.** Gasification reactivity of the chars

The reactivities, R_{50} , of the untreated and modified chars were analysed to evaluate the influence of the changes in the surface area and inorganic content on the rate of char gasification with CO₂. In addition to the TGA measurements at the temperature corresponding to the toluene 584 reforming tests (800 °C), TGA runs were performed at 850 and 900 °C to determine the kinetic 585 parameters for the VC and VC HCl samples. These parameters were employed to compare the 586 gasification of the untreated char catalysed with inherent inorganics, with the non-catalytic 587 gasification of the acid-washed material. The kinetic analysis was carried out as described in 588 Section 2.3, and the obtained R_{50} values are presented in Fig. 9. The R_{50} values from the VC and VC_HCl gasification at 800 °C differed by an order of magnitude, and at higher 589 590 temperatures the discrepancy between the chars become even more pronounced. While 591 increasing the temperature from 800 °C to 900 °C, the reactivity increased from 0.56×10^{-3} s⁻¹ to 3.75×10^{-3} s⁻¹ and from 0.07×10^{-3} s⁻¹ to 0.22×10^{-3} s⁻¹ for the VC and VC_HCl chars, 592 593 respectively. It can be expected that removing char inorganics during the acid wash and the 594 consequent lack of AAEM catalytic effects resulted in the VC_HCl char's poor sensitivity to 595 temperature changes. The 100 °C rise in the reaction temperature increased the reactivity of the 596 VC char over six times and increased the VC_HCl char's reactivity only three times, indicating 597 the higher activation energy of the VC char within the examined temperature range (Table 1).

598

599

601



600 Fig. 9. Untreated (VC) and acid-washed (VC_HCl) char reactivity as a function of the

temperature of the CO₂ gasification

Table 1. Kinetic parameters of untreated (VC) and acid-washed (VC_HCl) char gasification calculated based on the reactivities determined at 800, 850 and 900 °C

	E, kJ·mol ⁻¹	k_0 , bar ⁻¹ ·s ⁻¹	R^2	
VC	199	3.61E+06	0.997	
VC_HCl	119	5.42E+01	0.964	

In addition to the fresh chars, the spent materials recovered after 50 min of toluene conversion were also analysed in terms of their reactivity during CO₂ gasification at 800 °C in order to examine the influence of the coke formation and steam activation that occurred during these tests. The R_{50} values of fresh and spent chars are compiled in Table 2.

609

610 Table 2. Overview of the reactivity (R_{50}) of fresh and spent char samples

	Encel	Spent	Spent
	Fresh	50 min in N ₂	50 min in steam/N ₂
	$R_{50} \times 10^3 (1/s)$	$R_{50} \times 10^3 (1/s)$	$R_{50} \times 10^3 (1/s)$
VC	0.56 ± 0.02	0.46 ± 0.04	0.55 ± 0.03
VC_HCl	0.07 ± 0.01	0.06 ± 0.01	0.06 ± 0.01
VC_SA	0.94 ± 0.12	0.68 ± 0.01	
VC_HT	0.69 ± 0.04	0.45 ± 0.01	

611

612 The pre-treatments applied to VC significantly affected its gasification reactivity. In particular, 613 acid washing drastically decreased R_{50} . This result confirmed the important role that inorganics 614 play in determining the gasification reactivity of chars. In contrast, the steam activation and heat treatment effectively increased R_{50} . In the case of VC_SA, this was because of the increased surface area and AAEM content. In the case of VC_HT, this may be the result of metals rising to the surface following high-temperature treatment [37].

The pyrolytic conversion decreased the R_{50} values of VC, VC_SA, and VC_HT. The toluenederived coke was deposited on the char surface, thereby hindering the access to smaller pores (decreasing the surface area) and covering the AAEM species. This also significantly reduced the number of accessible active sites on the char surface, thus inhibiting the reactivity of spent chars. Steam reforming of toluene over the VC char maintained the reactivity of the char due to its continuous re-activation with steam.

624 No significant differences between the R_{50} values of fresh and spent VC_HCl chars were 625 observed. Since most of the AAEM species were already removed from the sample during the 626 pre-treatment, the coke layer did not have a deactivating effect, while the steam-activating 627 effect was negligible due the generally poor reactivity of the VC_HCl sample. The comparison 628 of the mass loss curves of the fresh and spent VC and VC_HCl chars, presented in Fig. 10, 629 confirmed that the gasification of the VC char that was spent during toluene pyrolysis was 630 slower than that of fresh VC and the sample recovered after steam reforming. Interestingly, the 631 mass loss curves for both spent VC_HCl chars were slightly less rapid than the fresh material. 632 It is possible that although the toluene conversion did not affect the reactivity of the char, the 633 char became more stable due to the high-temperature exposure during the conversion 634 experiments. The difference between the shapes of the mass loss curves of VC and VC_HCl 635 chars should be also noted. It is generally acknowledged [49] that the rate of biomass char 636 gasification increases with the char burn-off due to the increasing concentration of inorganics. 637 This behaviour resulted in the curved mass loss plots obtained for fresh and spent VC chars. 638 Because of the lack of inorganics in the VC_HCl samples, the mass loss functions of these acid-639 washed materials were linear, as no AAEM densification effect occurred.



Fig. 10. Mass loss curves during CO₂ gasification of the untreated (VC) and acid-washed
(VC_HCl) chars, and the chars recovered after 50 min of toluene conversion in pure nitrogen
(50N) or 15.5 vol.% of steam (50S)

645 **3.3.** Catalytic effect of the chars on toluene decomposition

The toluene conversions (η_T) obtained from the experiments summarised in Fig. 1 are reported in Table 3. The relatively high values of η_T confirm the good performance of this char for the heterogeneous conversion of toluene. The values of η_T are similar to values reported in literature for toluene conversion over char under similar conditions [35,50]. In addition, in N₂ atmosphere, all the samples exhibited reduced conversion over time due to coke deposition and blocking of active sites. This mechanism is the same as that which reduced the gasification reactivity of spent chars.

653

Atmosphere:	N_2	N_2	Steam/N ₂
Feeding time:	30 min	50 min	50 min
VC	75 ± 2	67 ± 1	74 ± 2
VC_HCl	79 ± 2	72 ± 1	75 ± 1
VC_SA	88 ± 4	78 ± 1	
VC_HT	70 ± 1	60 ± 1	

Table 3. Toluene conversion $(\eta_T, \%)$ over the tested chars under an inert N₂ atmosphere and in the presence of steam

657

The different chars varied significantly in their catalytic performances. It is especially interesting to note that the performance of the VC_HCl char was not impaired by the removal of inorganics. On the contrary, its efficiency for toluene pyrolytic conversion was slightly higher than the untreated VC char, especially for the 50 min runs, where the toluene conversion increased from 67 % to 72 % as a result of this pre-treatment. The steam activation increased the surface area of VC_SA and significantly enhanced its catalytic affinity, while the degraded surface structure resulting from the thermal treatment of VC_HT inhibited its performance.

665 These results suggested that the metals inherently present in VC did not play any major role in 666 the heterogenous conversion of toluene. The enhanced performance of the VC HCl char may 667 be related to the observed increase in its surface area or the protonation of the active sites. The 668 role of AAEM during volatile-char interactions was further investigated with steam reforming 669 tests. Under oxidising conditions, the char surface is involved in two types of reactions: (1) the 670 heterogeneous toluene conversion to coke and (2) the solid carbon gasification. As suggested 671 by the previous studies [9,36], these two reactions occur independently, and the ratio of their 672 individual rates determine if the catalyst loses or maintains its activity throughout the reforming 673 process. Thus, steam introduction should result in the solid carbon gasification, which was not 674 the case in the pyrolytic runs. Therefore, in this mode, despite no effect of AAEM on 675 heterogenous toluene conversion, their catalytic effect on carbon gasification was expected to 676 influence the reforming process in an indirect way, i.e. by enhancing char/coke gasification. 677 The drastic difference in the R_{50} values of the VC and VC_HCl chars was expected to offset the 678 balance between the rates of the toluene-derived coke formation and char consumption. Indeed, 679 the introduction of steam resulted in an increase in toluene conversion from 67 % to 74 % for 680 VC and only from 72 to 75 % for VC_HCl sample. This observation suggests that, although 681 heterogeneous conversion of toluene was not directly affected by the presence of inorganics, 682 they indirectly contributed to the increased efficiency of the conversion over VC by enhancing 683 gasification of this char and thus prolonging its activity.

The surface functionalities identified by FTIR did not appear to play an important role in the decomposition of toluene. All applied pre-treatments decreased the quantity of O-containing groups on the char surface, although different trends in the η_T values were observed for different modified materials.

688 The role of steam in the toluene reforming process was confirmed by analysing the released 689 gases presented in Fig. 11. The pyrolytic conversion of toluene yielded only H₂ and CH₄ (see 690 Supplementary Information); however, during the steam reforming process, CO₂ and CO 691 originating from the solid carbon gasification were also detected. Greater amounts of carbon 692 oxides and H₂ were released during the tests with VC, relative to VC_HCl, thus confirming that 693 the rate of the char/coke reaction with steam was faster when inorganics were present in the 694 catalyst. Some studies [51] have described the significant influence of the inherent AAEM 695 species on the water-gas shift reaction (i.e., $CO + H_2O \rightarrow CO_2 + H_2$), leading to increased 696 selectivity towards CO₂ during catalytic gasification [19], which might explain the particularly 697 high CO₂ yield during the test with the VC char. The linear mass loss of the VC_HCl char 698 registered during TGA tests revealed a constant gasification rate for this material. Additionally, 699 the similar reactivities of the spent and fresh samples suggested that the coke layer formed 700 during toluene conversion did not inhibit the gasification of the catalyst bed. Thus, the CO 701 release directly from the heterogenous water-steam reaction (i.e., via $C + H_2O \rightarrow CO + H_2$) was 702 constant throughout the steam reforming experiment with VC_HCl. In contrast, the toluene-703 derived coke was less reactive than the AAEM-rich VC char, as revealed by the lower R_{50} value for the spent, coke-covered VC char $(0.46 \times 10^{-3} \text{ s}^{-1})$ recovered after pyrolytic toluene conversion 704 relative to the R_{50} of the fresh VC char (0.56×10⁻³ s⁻¹). Thus, as the reaction proceeded, the coke 705 706 contributed more and more to the total amount of solid carbon interacting with steam, so the 707 rate of gasification and the CO yield both decreased. The continuous deposition and steam 708 consumption of the coke inhibited the gasification of the char bed, but also helped preserve the reactivity of the VC char recovered after the reforming test $(0.55 \times 10^{-3} \text{ s}^{-1})$. The non-catalysed 709 710 gasification has also been correlated with increased CH₄ formation [19], which was observed 711 during toluene reforming over the VC_HCl char in this study.

712

713





Fig. 11. Gas evolution during the steam reforming of toluene over the untreated (VC) and acidwashed (VC_HCl) chars

719 In addition to the permanent gases, some liquid by-products of toluene conversion were 720 detected (Fig. 12 and Supplementary Fig. S2). The demethylation of toluene resulted in the 721 formation of a substantial amount of benzene. Moreover, some secondary reactions led to 722 methyl substitution either directly on the aromatic ring or on the existing methyl group of 723 toluene molecules, yielding small amounts of xylenes and ethylbenzene. Some of the ethyl 724 groups underwent dehydrogenation to produce styrene. It was previously reported [52] that 725 toluene decomposition to benzene intensifies as the catalyst becomes saturated with coke. 726 Therefore, the selectivity towards benzene formation was lower during (i) pyrolytic toluene 727 conversion over the VC_SA char, which had a higher capacity for coke due to its larger and 728 more mesoporous surface area, and during (ii) steam reforming over the VC char, where the 729 coke was continuously removed by steam gasification, compared with the other chars. The secondary reactions leading to the formation of substituted benzenes formation are also 730

intensified when the catalyst is deactivated. Therefore, larger amounts of these compounds weregenerated during toluene pyrolysis, compared with steam reforming.

733 The construction of the test rig did not allow for gravimetric assessment of the mass gain/loss 734 of the catalyst bed; therefore, carbon balance was performed to estimate carbon accumulation 735 in the reactor during toluene conversion tests. The amounts of carbon in the liquid and gaseous 736 products, as well as in the unreacted share of toluene are presented in Fig. 13 as a percentage 737 of the amount of carbon fed into the reactor with the toluene. For the pyrolytic tests, where no 738 char/coke gasification occurred, the amount of the fed toluene that was converted into coke can 739 be therefore estimated by difference, as indicated by the grey sections of the VC, VC_HCl, 740 VC_SA, and VC_HT bars. These sections were relatively similar for all examined chars, with 741 the highest coke deposition on the most efficient and most mesoporous sample (i.e., VC_SA), 742 and the lowest coke yield on the least effective char (i.e., VC HT).

743 Different information can be obtained from Fig. 13 for the toluene reforming tests. In this case 744 a part of the char bed and toluene-derived coke underwent steam gasification. Hence, an 745 unknown share of released gases comprised carbon that was introduced into the reactor with the char bed and not with the fed toluene. Thus, for these experiments, the amount of created 746 747 coke cannot be calculated from the performed carbon balance. However, results presented as 748 the bars VC_steam and VC_HCl_steam represent the overall change in C content in the reactor 749 that occurred as a result of three concurrent reactions involving the solid carbon: coke 750 formation, coke gasification, and char gasification. The total amount of C released during the 751 steam test with the VC char exceeds the amount fed with the toluene, which suggests that the 752 total amount of coke and char consumed during the reforming process was higher than the 753 amount of the deposited coke. However, it is impossible to determine the proportion of the 754 gasified coke versus the char. Compared with the pyrolytic test, the presence of steam also 755 decreased the amount of coke/char in the bed of the acid-washed VC_HCl char, yet the lower gasification rate observed for this catalyst did not prevent the accumulation of carbon on its surface. The deactivation of the spent chars was also confirmed by the reduction in their FTIR spectral intensities (see the Supplementary Information); some functional groups remained in the VC char after steam reforming, whereas the spectra of both spent VC_HCl chars and the VC char recovered after pyrolysis were flattened.

Fig. 12. Relative molecular yields of liquid by-products following 50 min of toluene conversion over the chars under inert (N_2) and oxidising (steam) conditions

Fig. 13. Carbon distribution in the reaction products based on the amount of C fed with the toluene during conversion under inert (N_2) and oxidising (steam) atmosphere

3.4. The gasification reactivity and catalytic properties of biochar based on AAEM content and surface area

773 Four char samples (the untreated VC and three modifications) were evaluated as catalysts for 774 toluene decomposition, while the CO₂ gasification reactivities of the fresh and spent materials 775 were determined using TGA. The main physicochemical properties of the chars were analysed 776 to identify the differences between the samples, focusing primarily on their surface area and 777 inherent inorganic content. This approach allowed for comparing the performance of the chars 778 in terms of the parameters generally attributed to high reactivity and good catalytic properties 779 of carbonaceous materials. Interestingly, it was observed that the char reactivity and catalytic 780 efficiency were not equally affected by the applied pre-treatments. This result suggests that 781 different features of the char play crucial roles in determining its gasification reactivity and catalytic properties. Therefore, an attempt was made to understand how the most important parameters of the char (surface area, surface functionalities, and inorganic content) influence the char performance (reactivity, R_{50} , and toluene conversion, η_T).

785 In Fig. 14a, the R_{50} values of the chars are compared with the K content, which is one of the 786 main and most reactive AAEM species. A general trend of increasing gasification reactivity 787 with increasing concentration of K is clear. The most drastic difference in the chars' K content, 788 introduced by acid washing and steam activation, resulted in R_{50} values that were 10 times lower 789 (VC_HCl) and two times higher (VC_SA) than the VC. The char reactivity was most likely 790 influenced by other parameters as well, e.g., a decrease in R_{50} was observed for coke-covered 791 materials. Both the pre-treatment and application for toluene conversion also affected the 792 structural parameters of the chars, including the functional groups, pore size distribution, and 793 likely the organisation of the carbon lattice within the chars, i.e., the presence of the unsaturated 794 carbons. As a result, it is impossible to isolate the effect of the inorganics on the reactivity; 795 however, a relatively good linear fit of the reactivity with respect to the potassium content was 796 observed in this study, as described by Eq. 8,

$$R_{50} = 0.00004C_K + 0.058 \tag{8}$$

798 where C_K is the potassium content (ppm) in the char.

The plot of the R_{50} values as a function of the BET surface area of the chars, which varied drastically among the analysed samples, revealed no clear correlation (Fig. 14b). Despite the small decrease in the reactivity of each fresh char with the reduction in surface area, resulting from the exposure to toluene, the main differences in the gasification reactivity was observed between the main types of chars with the most diverse AAEM content (i.e., between the VC, VC_HCl, and VC_SA groups), regardless of the significant changes in their surface areas caused by their interactions with toluene. There was no clear relationship between the reactivityof chars and the development of their surface functional groups observed by ATR-FTIR.

Fig. 14. The reactivity, R_{50} , of the fresh (blue) chars and the recovered chars after 50 min of toluene pyrolysis (orange) and steam reforming (green) plotted against (a) the K content in the char and (b) the BET surface area determined from the N₂ adsorption isotherm

812

813 The reactivity of the char was affected substantially by its inherent inorganic content, whereas 814 the wide variations in the surface area did not contribute significantly to the gasification rate of 815 these materials. In contrast, the catalytic activity of the char towards pyrolytic toluene 816 conversion was largely unaffected by the presence of inherent inorganics. The acid-washed char 817 (VC_HCl) performed better than the untreated sample, despite the removal of AAEM species, 818 while the steam-activated (VC SA) char contained an increased metal content and an improved 819 catalytic efficiency. These results indicated that there was no clear correlation between the 820 catalyst's inorganic content and its toluene conversion capabilities. However, the changes in 821 the surface area and pore size distribution, which are also commonly attributed to the catalytic 822 properties of the materials, correspond relatively well with the observed differences in toluene 823 pyrolysis, as presented in Fig. 15. The useful surface area employed in this comparison was 824 calculated from the 2D-NLDFT pore size distribution and determined as the cumulative surface 825 area of the pores with sizes larger than the kinetic diameter of toluene (5.85 Å), because only 826 those pores are expected to participate in the toluene conversion. The four fresh chars differed 827 in terms of their inorganic content and O-functional groups, which could be expected to obscure 828 the correlation between the surface area and toluene conversion. However, the linear 829 relationship between η_T (%) and the useful surface area was observed for both toluene feeding 830 times, as described by Eq. 9 and 10,

831
$$\eta_T(t=50\min) = 0.069A_u + 23.7$$
 (9)

$$\eta_T(t=30\min) = 0.068A_u + 33.4 \tag{10}$$

833 where A_u is the useful surface area (m²/g). The similarity between the slopes of both fitted 834 functions suggests a uniform conversion mechanism that depends mainly on the char's capacity 835 for coke accumulation.

Therefore, it can be expected that under an inert atmosphere, the inherent inorganic content in the char has a negligible effect on the conversion of toluene, and that the catalytic properties are determined mainly by the char's structural parameters. For chars with a similar nature (i.e., 839 obtained from the same feedstock and only slightly modified), the useful surface area seemed 840 to be the main feature affecting the toluene conversion. However, further investigations are 841 required to confirm this hypothesis. The results of this study appear to contradict Fuentes-Cano 842 et al.'s observation [25] of decreased naphthalene removal over acid-washed coal char, 843 suggesting that some important process parameters must be accounted for, while examining the 844 role of inorganics in the tar conversion process. Different procedures for samples preparation 845 were applied in the two studies. Steam activation performed after the removal of inorganics 846 may have affected the char activation process (e.g., inhibited the formation of O-functionalities 847 on the surface). Moreover, the lack of the methyl group in naphthalene could have impacted its 848 sensitivity to catalysis promoted by AAEM species. Indeed, our previous work regarding the 849 conversion of benzene suggested that non-methylated aromatics have different sensitivity to the 850 catalyst compared with more reactive, methylated species [36].

851 However, the most important difference can be traced back to the feedstock used for the char 852 preparation; the metals in coal are present in the form of clusters, contrary to the well-dispersed 853 ions and salts in biomass. Therefore, it is plausible that the inorganics in the coal char examined 854 by Fuentes-Cano et al. [25] remained in the form of agglomerates, as indicated by the inorganic 855 clusters observed in the SEM-EDS scans of the coal char [25], whereas the K and Ca elements 856 were well-dispersed on the biochar surface (Fig. 5e). Since the metal particles and clusters have 857 stronger catalytic effects than the structures integrated in the carbon matrix (e.g., C-O-M 858 phenolates [20,38]), the inherent inorganics in coal and biomass likely have different affinities 859 for catalysing tar conversion reactions.

860

862

Fig. 15. Correlation between the pyrolytic toluene conversion (η_T) during the 30- and 50-min tests and the useful surface area (pores >5.85 Å) of the chars

866 **4.** Conclusions

This work examined the effect of the surface area and inherent inorganic contents of woodderived char on its gasification reactivity and catalytic performance. The untreated and modified chars were gasified with CO₂, and they were also used as catalysts for toluene conversion. The goals of the modifications were to change the surface area and inorganic content of the untreated char: acid wash removed the inorganics, steam activation increased the surface area and inorganic concentration, and high-temperature annealing decreased the porosity of the char.

The results indicated that the reactivity of the char was strongly correlated with the quantity of AAEM species, yet there was no clear effect of the surface area on the char's gasification rate. Conversely, toluene pyrolytic conversion over the char was unaffected by changes in the 877 concentration of inorganics, yet it correlated well with the surface area available for the reaction 878 with toluene. Therefore, it was concluded that enhanced oxygen adsorption on the char surface 879 due to the AAEM species incorporated in the carbonaceous matrix significantly increased char 880 gasification rate, but it did not contribute to toluene reactions on the surface. The 881 microcrystalline structure of the char is rich in unsaturated carbons that can attract volatile 882 compounds, so the determining factor for the toluene removal efficiency was the area of the 883 char surface that was available for toluene adsorption (i.e., the area comprised of pores with a 884 size greater than the kinetic diameter of toluene). However, the inorganics in the char could 885 indirectly enhance toluene conversion during the reforming process, by catalysing the 886 gasification, and thus activation, of the char.

887 It is important to note that the observations presented herein were made for one type of catalyst 888 and one tar-representing compound (toluene). Therefore, further research into the catalytic 889 effects of different char properties is required to unambiguously determine their importance. 890 For example, studies regarding conversion of O-containing, polyaromatic, or non-substituted 891 aromatics, and mechanistic investigations of their heterogeneous decomposition will lead to a 892 more comprehensive insight on how different metal species might contribute uniquely to their 893 conversion. Moreover, because the nature of the metal-char bonding is known to have a strong 894 impact on the reactivity of these active sites [20], it is advisable to examine the role of 895 inorganics introduced in a form different to that of the inherent char elements, e.g., using a 896 catalyst impregnated with metal salts.

897

898 Acknowledgements

899 This work was supported by the Statutory Fund of the Faculty of Energy and Environmental900 Engineering of the Silesian University of Technology.

901 Appendix A. Supplementary data

902 Supplementary data to this article can be found online at

903

905 [1] F. Patuzzi, D. Prando, S. Vakalis, A.M. Rizzo, D. Chiaramonti, W. Tirler, T. Mimmo,
906 A. Gasparella, M. Baratieri, Small-scale biomass gasification CHP systems:
907 Comparative performance assessment and monitoring experiences in South Tyrol (Italy),

908 Energy. 112 (2016) 285–293. https://doi.org/10.1016/j.energy.2016.06.077.

- 909 [2] J. Lehmann, S. Joseph, Biochar for environmental management: science, technology and
 910 implementation, Routledge, 2015.
- 911 [3] E.F. Zama, B.J. Reid, H.P.H. Arp, G.X. Sun, H.Y. Yuan, Y.G. Zhu, Advances in research
 912 on the use of biochar in soil for remediation: a review, J. Soils Sediments. 18 (2018)
 913 2433–2450. https://doi.org/10.1007/s11368-018-2000-9.
- 914 [4] V. Augustyn, P. Simon, B. Dunn, Pseudocapacitive oxide materials for high-rate
 915 electrochemical energy storage, Energy Environ. Sci. 7 (2014) 1597–1614.
 916 https://doi.org/10.1039/c3ee44164d.
- 917 [5] J. Lee, K.H. Kim, E.E. Kwon, Biochar as a Catalyst, Renew. Sustain. Energy Rev. 77
 918 (2017) 70–79. https://doi.org/10.1016/j.rser.2017.04.002.
- 919 [6] F. Nestler, L. Burhenne, M.J. Amtenbrink, T. Aicher, Catalytic decomposition of
 920 biomass tars: The impact of wood char surface characteristics on the catalytic
 921 performance for naphthalene removal, Fuel Process. Technol. 145 (2016) 31–41.
 922 https://doi.org/10.1016/j.fuproc.2016.01.020.

- 923 [7] D. Fuentes-Cano, A. Gómez-Barea, S. Nilsson, P. Ollero, Decomposition kinetics of
 924 model tar compounds over chars with different internal structure to model hot tar
 925 removal in biomass gasification, Chem. Eng. J. 228 (2013) 1223–1233.
 926 https://doi.org/10.1016/j.cej.2013.03.130.
- 927 [8] G. Ravenni, Z. Sárossy, J. Ahrenfeldt, U.B. Henriksen, Activity of chars and activated
 928 carbons for removal and decomposition of tar model compounds A review, Renew.
 929 Sustain. Energy Rev. 94 (2018) 1044–1056.
 930 https://doi.org/10.1016/J.RSER.2018.07.001.
- 931 [9] S. Hosokai, K. Kumabe, M. Ohshita, K. Norinaga, C.Z. Li, J. Hayashi, Mechanism of
 932 decomposition of aromatics over charcoal and necessary condition for maintaining its
 933 activity, Fuel. 87 (2008) 2914–2922. https://doi.org/10.1016/j.fuel.2008.04.019.
- 934 [10] Y. Shen, Chars as carbonaceous adsorbents/catalysts for tar elimination during biomass
 935 pyrolysis or gasification, Renew. Sustain. Energy Rev. 43 (2015) 281–295.
 936 https://doi.org/10.1016/J.RSER.2014.11.061.
- [11] D. Xu, L. Yang, K. Ding, Y. Zhang, W. Gao, Y. Huang, H. Sun, X. Hu, S.S.A. SyedHassan, S. Zhang, H. Zhang, Mini-Review on Char Catalysts for Tar Reforming during
 Biomass Gasification: The Importance of Char Structure, Energy and Fuels. 34 (2020)
 1219–1229. https://doi.org/10.1021/acs.energyfuels.9b03725.
- [12] Y. Huang, X. Yin, C. Wu, C. Wang, J. Xie, Z. Zhou, L. Ma, H. Li, Effects of metal
 catalysts on CO2 gasification reactivity of biomass char, Biotechnol. Adv. 27 (2009)
 568–572. https://doi.org/10.1016/j.biotechadv.2009.04.013.
- [13] Z. Zhang, S. Pang, T. Levi, Influence of AAEM species in coal and biomass on steam
 co-gasification of chars of blended coal and biomass, Renew. Energy. 101 (2017) 356–
 363. https://doi.org/10.1016/J.RENENE.2016.08.070.

- 947 [14] N.B. Klinghoffer, M.J. Castaldi, A. Nzihou, Influence of char composition and
 948 inorganics on catalytic activity of char from biomass gasification, Fuel. 157 (2015) 37–
 949 47. https://doi.org/10.1016/j.fuel.2015.04.036.
- 950 [15] A. Dufour, A. Celzard, V. Fierro, E. Martin, F. Broust, A. Zoulalian, Catalytic
 951 decomposition of methane over a wood char concurrently activated by a pyrolysis gas,
 952 Appl. Catal. A Gen. 346 (2008) 164–173. https://doi.org/10.1016/j.apcata.2008.05.023.
- 953 M.H. Kim, E.K. Lee, J.H. Jun, S.J. Kong, G.Y. Han, B.K. Lee, T.-J. Lee, K.J. Yoon, [16] 954 Hydrogen production by catalytic decomposition of methane over activated carbons: 955 kinetic study, Int. J. Hydrogen Energy. 29 (2004)187–193. 956 https://doi.org/10.1016/S0360-3199(03)00111-3.
- 957 [17] N. Muradov, F. Smith, A. T-Raissi, Catalytic activity of carbons for methane
 958 decomposition reaction, in: Catal. Today, Elsevier, 2005: pp. 225–233.
 959 https://doi.org/10.1016/j.cattod.2005.02.018.
- D. Feng, Y. Zhao, Y. Zhang, S. Sun, S. Meng, Y. Guo, Y. Huang, Effects of K and Ca
 on reforming of model tar compounds with pyrolysis biochars under H2O or CO2, Chem.
 Eng. J. 306 (2016) 422–432. https://doi.org/10.1016/J.CEJ.2016.07.065.
- 963 [19] Y. Wu, J. Wang, S. Wu, S. Huang, J. Gao, Potassium-catalyzed steam gasification of
 964 petroleum coke for H2 production: Reactivity, selectivity and gas release, Fuel Process.
 965 Technol. 92 (2011) 523–530. https://doi.org/10.1016/j.fuproc.2010.11.007.
- 966 [20] S.G. Chen, R.T. Yang, The active surface species in alkali-catalyzed carbon gasification:
 967 Phenolate (C-O-M) groups vs clusters (particles), J. Catal. 141 (1993) 102–113.
 968 https://doi.org/10.1006/jcat.1993.1122.
- 969 [21] S.G. Chen, R.T. Yang, Mechanism of alkali and alkaline earth catalyzed gasification of

- 970 graphite by CO2 and H2O studied by electron microscopy, J. Catal. 138 (1992) 12–23.
 971 https://doi.org/10.1016/0021-9517(92)90003-Z.
- 972 [22] H. Zhu, X. Wang, J. Zhang, K. Yao, G. Yu, X. Wang, Investigation of K2CO3-Catalyzed
 973 Pyrolysis and Steam Gasification of Coal Char, Energy Technol. 3 (2015) 961–967.
 974 https://doi.org/10.1002/ente.201500094.
- 975 [23] T. Sueyasu, T. Oike, A. Mori, S. Kudo, K. Norinaga, J. Hayashi, Simultaneous Steam 976 Reforming of Tar and Steam Gasification of Char from the Pyrolysis of Potassium-977 199–208. Loaded Woody Biomass, Energy & Fuels. 26 (2012)978 https://doi.org/10.1021/ef201166a.
- 979 [24] D. Feng, H. Sun, Y. Ma, S. Sun, Y. Zhao, D. Guo, G. Chang, X. Lai, J. Wu, H. Tan,
 980 Catalytic Mechanism of K and Ca on the Volatile–Biochar Interaction for Rapid
 981 Pyrolysis of Biomass: Experimental and Simulation Studies, Energy & Fuels. (2020)
 982 acs.energyfuels.0c01946. https://doi.org/10.1021/acs.energyfuels.0c01946.
- D. Fuentes-Cano, F. Parrillo, G. Ruoppolo, A. Gómez-Barea, U. Arena, The influence
 of the char internal structure and composition on heterogeneous conversion of
 naphthalene, Fuel Process. Technol. 172 (2018) 125–132.
 https://doi.org/https://doi.org/10.1016/j.fuproc.2017.12.015.
- [26] C.Z. Li, Importance of volatile-char interactions during the pyrolysis and gasification of
 low-rank fuels A review, Fuel. 112 (2013) 609–623.
 https://doi.org/10.1016/j.fuel.2013.01.031.
- 990 M. Hervy, S. Berhanu, E. Weiss-Hortala, A. Chesnaud, C. Gérente, A. Villot, D. Pham [27] 991 Minh, A. Thorel, L. Le Coq, A. Nzihou, Multi-scale characterisation of chars mineral 992 189 (2017)88-97. species for cracking, Fuel. tar 993 https://doi.org/10.1016/j.fuel.2016.10.089.

- [28] Z. Bai, H. Chen, W. Li, B. Li, Hydrogen production by methane decomposition over coal
 char, Int. J. Hydrogen Energy. 31 (2006) 899–905.
 https://doi.org/10.1016/J.IJHYDENE.2005.08.001.
- 997 [29] D. Feng, Y. Zhang, Y. Zhao, S. Sun, J. Wu, H. Tan, Mechanism of in-situ dynamic
 998 catalysis and selective deactivation of H2O-activated biochar for biomass tar reforming,
 999 Fuel. 279 (2020) 118450. https://doi.org/10.1016/j.fuel.2020.118450.
- 1000 J. Ahrenfeldt, U. Henriksen, T.K. Jensen, B. Gøbel, L. Wiese, A. Kather, H. Egsgaard, [30] 1001 Validation of a continuous combined heat and power (CHP) operation of a two-stage 1002 biomass gasifier, Energy and Fuels. 20 (2006)2672-2680. 1003 https://doi.org/10.1021/ef0503616.
- 1004 [31] P. Brandt, E. Larsen, U. Henriksen, High tar reduction in a two-stage gasifier, Energy
 1005 and Fuels. 14 (2000) 816–819. https://doi.org/10.1021/ef990182m.
- S. Hosokai, J.-I. Hayashi, T. Shimada, Y. Kobayashi, K. Kuramoto, C.-Z. Li, T. Chiba,
 Spontaneous Generation of Tar Decomposition Promoter in a Biomass Steam Reformer,
 Chem. Eng. Res. Des. 83 (2005) 1093–1102. https://doi.org/10.1205/cherd.04101.
- IO09 [33] J. Jagiello, J. Kenvin, A. Celzard, V. Fierro, Enhanced resolution of ultra micropore size
 determination of biochars and activated carbons by dual gas analysis using N2 and CO2
 with 2D-NLDFT adsorption models, Carbon N. Y. 144 (2019) 206–215.
 https://doi.org/10.1016/j.carbon.2018.12.028.
- 1013 [34] P. Ollero, A. Serrera, R. Arjona, S. Alcantarilla, The CO2 gasification kinetics of olive
 1014 residue, Biomass and Bioenergy. 24 (2003) 151–161. https://doi.org/10.1016/S09611015 9534(02)00091-0.
- 1016 [35] A. Korus, A. Samson, A. Szlęk, Catalytic conversion of toluene over a biochar bed under

- 1017an inert atmosphere The comparison of chars from different types of wood and the role1018of selected metals, Fuel. 279 (2020) 118468. https://doi.org/10.1016/j.fuel.2020.118468.
- 1019 [36] A. Korus, A. Szlęk, I. Korus, A. Samson, Catalytic conversion of methylated aromatics
 1020 over wood-derived chars The role of reforming agents and the effect of methyl groups,
- 1021 Fuel. 280 (2020) 118521. https://doi.org/10.1016/j.fuel.2020.118521.
- 1022 [37] N.B. Klinghoffer, M.J. Castaldi, A. Nzihou, Catalyst properties and catalytic
 1023 performance of char from biomass gasification, Ind. Eng. Chem. Res. 51 (2012) 13113–
 1024 13122. https://doi.org/10.1021/ie3014082.
- 1025 [38] D. Feng, Y. Zhao, Y. Zhang, J. Gao, S. Sun, Changes of biochar physiochemical
 1026 structures during tar H2O and CO2 heterogeneous reforming with biochar, Fuel Process.
 1027 Technol. 165 (2017) 72–79. https://doi.org/10.1016/j.fuproc.2017.05.011.
- [39] D. Fuentes-Cano, L. von Berg, A. Diéguez-Alonso, R. Scharler, A. Gómez-Barea, A.
 Anca-Couce, Tar conversion of biomass syngas in a downstream char bed, Fuel Process.
 Technol. 199 (2020) 106271. https://doi.org/10.1016/J.FUPROC.2019.106271.
- 1031 [40] R.M. Silverstein, F.X. Webster, D.J. Kiemle, D.L. Bryce, Spectrometric Identification
 1032 of Organic Compounds, 8th ed., Wiley, 2014.
- 1033 [41] W. Zieliński, A. Rajca, Metody spektroskopowe i ich zastosowanie do identyfikacji
 1034 związków organicznych, second ed., Wydawnictwa Naukowo-Techniczne, Warsaw,
 1035 2000.
- I036 [42] S.C. Manna, S. Mistri, A.D. Jana, A rare supramolecular assembly involving ion pairs
 of coordination complexes with a host–guest relationship: synthesis, crystal structure,
 photoluminescence and thermal study, CrystEngComm. 14 (2012) 7415–7422.
 https://doi.org/10.1039/C2CE25916H.

- IO40 [43] G.E. Dunn, R.S. McDonald, Infrared spectra of aqueous sodium benzoates and
 salicylates in the carboxyl-stretching region: chelation in aqueous sodium salicylates,
 Can. J. Chem. 47 (1969) 4577–4588. https://doi.org/10.1139/v69-758.
- 1043 [44] N. Montoya Sánchez, A. de Klerk, Oxidative ring-opening of aromatics:
 1044 Thermochemistry of sodium, potassium and magnesium biphenyl carboxylates,
 1045 Thermochim. Acta. 645 (2016) 31–42. https://doi.org/10.1016/j.tca.2016.11.002.
- 1046 [45] Thommes Matthias, K. Katsumi, N.A. V, O.J. P, R.-R. Francisco, R. Jean, S.K.S. W,
 1047 Physisorption of gases, with special reference to the evaluation of surface area and pore
 1048 size distribution (IUPAC Technical Report), Pure Appl. Chem. 87 (2015) 1051.
 1049 https://doi.org/10.1515/pac-2014-1117.
- 1050 [46] M. Molina-Sabio, M.T. Gonzalez, F. Rodriguez-Reinoso, A. Sepúlveda-Escribano,
 1051 Effect of steam and carbon dioxide activation in the micropore size distribution of
 1052 activated carbon, Carbon N. Y. 34 (1996) 505–509. https://doi.org/10.1016/00081053 6223(96)00006-1.
- M. Hervy, E. Weiss-Hortala, D. Pham Minh, H. Dib, A. Villot, C. Gérente, S. Berhanu,
 A. Chesnaud, A. Thorel, L. Le Coq, A. Nzihou, Reactivity and deactivation mechanisms
 of pyrolysis chars from bio-waste during catalytic cracking of tar, Appl. Energy. 237
 (2019) 487–499. https://doi.org/10.1016/j.apenergy.2019.01.021.
- I. Jae, G.A. Tompsett, A.J. Foster, K.D. Hammond, S.M. Auerbach, R.F. Lobo, G.W.
 Huber, Investigation into the shape selectivity of zeolite catalysts for biomass
 conversion, J. Catal. 279 (2011) 257–268. https://doi.org/10.1016/j.jcat.2011.01.019.
- 1061 [49] C. Fushimi, T. Wada, A. Tsutsumi, Inhibition of steam gasification of biomass char by
 1062 hydrogen and tar, Biomass and Bioenergy. 35 (2011) 179–185.
 1063 https://doi.org/10.1016/J.BIOMBIOE.2010.08.017.

1064 [50] P.N. Bhandari, A. Kumar, D.D. Bellmer, R.L. Huhnke, Synthesis and evaluation of
1065 biochar-derived catalysts for removal of toluene (model tar) from biomass-generated
1066 producer gas, Renew. Energy. 66 (2014) 346–353.
1067 https://doi.org/10.1016/J.RENENE.2013.12.017.

- [51] Z. Ma, R. Xiao, H. Zhang, Catalytic steam reforming of bio-oil model compounds for
 hydrogen-rich gas production using bio-char as catalyst, Int. J. Hydrogen Energy. 42
 (2017) 3579–3585. https://doi.org/10.1016/j.ijhydene.2016.11.107.
- 1071 [52] A. Korus, A. Samson, A. Szlęk, A. Katelbach-Woźniak, S. Sładek, Pyrolytic toluene
 1072 conversion to benzene and coke over activated carbon in a fixed-bed reactor, Fuel. 207
 1073 (2017) 283–292. https://doi.org/10.1016/j.fuel.2017.06.088.