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3	The microbiome mediates the interaction between predation and heavy
4	metals
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17	

18 Abstract

19 Gut microbiota communities are fundamental ecological components in the aquatic food web. 20 Their potential to mediate how organisms respond to multiple environmental stressors remains 21 understudied. Here we explored how manipulations of the gut microbiome of *Daphnia pulex*, 22 a keystone species in aquatic communities, influenced life history (size at maturity, age at maturity, somatic growth rate and clutch size), morphology (induced defence) and body 23 24 condition (lipid status deposits) responses to combined anthropogenic (copper) and natural (predation risk) stress. Data from a factorial experiment revealed that the effect of predation 25 26 risk on traits was often mediated by copper (predation risk and copper interact). These patterns 27 align with theory linking predation risk and copper contamination via digestive physiology. 28 We also found that each stressor, and their combination, was associated with the same 29 community composition of the D. pulex microbiome. However, antibiotic manipulation of the 30 microbiome reversed 7/12 the trait responses across life history, morphology and body 31 condition. This was associated with dramatically different communities to control conditions, 32 with clear and unique patterns of microbiome community composition for each stressor and 33 their combination. Our study revealed that microbiome community composition is highly 34 correlated with the response of organisms to multiple, simultaneous stressors.

35

37 INTRODUCTION

38 In freshwater ecosystems, organisms are exposed simultaneously to a wide range of biotic and 39 abiotic stressors including predation, bacterioplankton, metals, nutrition, pH and temperature 40 (Coors and De Meester, 2008; Coors et al., 2004; Hecky and Kilham, 1988; Hunter and Pyle, 41 2004; Jackson et al., 2016; Long et al., 2004; Martins et al., 2017). Evaluating how multiple 42 simultaneous stressors impact on organisms currently focuses on understanding how life 43 history, morphological and behaviour responses reflect additive or interactive effects among 44 stressors (Jackson et al., 2016; Orr et al., 2020). Multi-stressor research has grown enormously 45 in the past decades as our understanding of modes of action of stressors have allowed 46 hypotheses and experiments to address whether stressors combine additively or interactively 47 (Altshuler et al., 2011; Folt et al., 1999; Jansen et al., 2011; Loureiro et al., 2010; Orr et al., 48 2020).

49

50 Understanding how such effects arise requires a more thorough understanding of organism 51 physiology. An emerging approach to achieve this understanding centres on examining how 52 the microbiome of target species might mediate responses to different stressors (Banerjee et 53 al., 2018; Mushegian et al., 2019; Sison-Mangus et al., 2015). The gut microbiome is a 54 potential 'interface' to mediating the response to stressors in freshwater which enter organisms 55 via contact or ingestion. Here we focus on the role of the gut microbiome of *Daphnia pulex* in 56 mediating its response to the heavy metal Copper and the biotic stress of predation risk.

57

58 These are common and co-occuring stressors in many aquatic communities around the world 59 and each form of stress is represented by a large body of historical empirical research that 60 together provides the platform for predicting how variation in the identity and relative abundance of taxa in the gut microbiome might mediate physiology, life history andmorphology under threat from both stressors.

63

64 How Copper affects life history

65 Exposure to anthropogenic sources of heavy metals such as Copper generate major responses in life history and behaviour leading often to dramatically reduced fitness (De Schamphelaere 66 67 et al., 2007; Martins et al., 2017; Sadeq and Beckerman, 2019a; Shuhaimi-Othman et al., 2010). 68 Theory and empirical data indicate that sub-lethal concentrations of Copper influence foraging 69 and assimilation of nutritional resources which can lead to as increased metabolic cost and 70 reduced energy acquisition. Copper's impact on digestive physiology is strongly predicted by 71 classic life-history theory where reduced energy intake translates into delayed maturity at a 72 smaller size, reduced reproduction and slower somatic growth rates (Barata and Baird, 2000; 73 Barata et al., 2000; Bui et al., 2016; De Schamphelaere et al., 2007; Sadeq and Beckerman, 74 2019b).

75

76 How Predation Risk affects life history and morphology

77 Several decades of research on predation risk reveal a wide array of responses in Daphnia spp. 78 Much of predation risk research, like that on sub-lethal concentrations of metals, is framed 79 around how risk of predation alters foraging, habitat use and life history and on the allocation 80 of energy to growth versus reproduction (Beckerman et al., 2007; Benard, 2004; Noonburg and 81 Nisbet, 2005; Stoks and McPeek, 2003; Taylor and Gabriel, 1992). Predation risk is well 82 known to alter foraging behaviour (Balseiro et al., 2007; Noonburg and Nisbet, 2005) with very 83 clear changes life history (Beckerman et al., 2010; Black and Dodson, 1990; Campero et al., 84 2007; Pestana et al., 2009; Pestana et al., 2010; Rose et al., 2002; Rose et al., 2001; Schulz and ⁸⁵ Dabrowski, 2001) and morphology (Carter et al., 2017; Hammill et al., 2008; Tollrian, 1995;
⁸⁶ Tollrian and Dodson, 1999) driven by the size-selectivity of the predator.

⁸⁷ Under small size selective predation, the conditions of predation risk favour somatic growth
⁸⁸ over reproduction and typically leads to later age and larger size at maturity, often along with
⁸⁹ induced morphological defence. In contrast, under large size selection by predators, prey
⁹⁰ favour reproduction over somatic growth, typically leading to early maturity at a small size
⁹¹ (Beckerman et al., 2010; Beckerman et al., 2007).

92

93 A null model for Copper-Predation Risk interactions

94 Thus, sub-lethal concentration of Copper drive changes in life history that are typically 95 associated with starvation. In contrast, predation risk drives changes in life history that are not 96 directly aligned with theory about resource limitation and are instead driven by the size-97 selectivity of the predator. Such 'univariate' patterns provide a template and null expectation 98 for how the two forms of stress might combine additively. The central role of predator size 99 and resource allocation to life history and morphological defences suggests that any abiotic 100 (e.g. Copper) stress that interferes with digestive physiology may alter the response to predation 101 risk (and vice-versa). Table 1 provides an overview of aligned and contrasting effects of metals 102 and predation that underpin the expectation of their joint effects.

103

104 The relationship between stress responses and microbiota: a case study with Daphnia pulex

In addition to providing a template for additive effects between the stressors, the null predictions also provide a reference point for evaluating how the gut microbiota mediates responses to these stressors.

108 TABLE 1 HERE

Based on this history of research centred around digestive physiology and the allocation of energy to growth and reproduction, we propose that the gut microbiome may mediate interactions between Copper and predation risk. This mechanistic hypothesis stems from assuming that digestive physiology is a shared 'mode-of-action' for the response to metals and to predation risk (*sensu* ecotoxicology). Here we evaluate this hypothesis in *Daphnia pulex* facing stress from copper and from predation risk by manipulating their microbiome with antibiotics.

117

118 The Daphnia gut is colonized by a wide range of bacteria. They obtain these microbes via the 119 transmission from host parents to offspring or acquisition from sediments and food in ponds 120 and lakes that enter organism's gut via filter feeding (Gillis et al., 2005; Grossart et al., 2009; 121 Mushegian et al., 2019). As a result, the Daphnia gut offers a niche for selective microbes that 122 may provide benefits and services to their hosts (Sison-Mangus et al., 2015). Gut microbiota 123 offer a variety of functions and physiological processes to their hosts associated with 124 metabolism, development, fecundity immunity, and behaviour (Dattagupta et al. 2009, 125 Nicholson et al. 2012, Sommer et al. 2013, Gorokhova et al. 2015, McKenney and Pamer 2015, 126 Sampson and Mazmanian 2015, Sison-Mangus et al. 2015). Bacterial communities may thus 127 influence the potential interaction between metals and predation (Gorokhova et al. 2015) at the 128 nexus of energy acquisition, assimilation and allocation to growth, defence and reproduction. 129

To evaluate whether the gut microbiome mediates response to multiple stressors and offer insight into how this might happen, we manipulated the microbiome of *Daphnia pulex* using antibiotics as part of a fully factorial experiment evaluating how body condition, life history and morphological defences respond to exposure to predation risk and copper. Our experimental design is motivated by asking the following two questions: First, does the effect

of predation risk on body condition, life history, and induced defences vary by the presence of Copper? This question centres around evaluating the predictions shown in Table 1 and are a formal test for synergistic or antagonistic effects of the two stressors. Second, does antibiotic treatment which will alter directly the gut microbiota, disrupt the interaction(s) between copper and predation risk leading potentially to augmentation or reversals of the patterns, defined in Table 1. These questions are motivated by the shared importance of digestive physiology on the response to predation risk and metals.

142

143 We first report on whether the effects of predation risk vary by copper in the absence of antibiotic exposure (e.g. Table 1). We then document the reversal of numerous interactions 144 between predation risk and copper under antibiotic treatment suggesting a central role of the 145 146 digestive physiology and the gut microbiota in mediating response to copper and predation 147 risk. We then associate these responses, and the experimental treatments causing them, with 148 clear changes in taxonomic and functional diversity of the microbiota. This final step 149 formalises emergent hypotheses about functional groups of bacteria that appear to mediate the 150 response to combined metal (anthropogenic) and predation risk (natural) stressors.

151

152 MATERIAL AND METHODS

153 Daphnia culturing

The clone *D. pulex* (LD33) was collected from field populations (in 2010) in the UK and maintained in long-term stock culture in the Department of Animal and Plant Sciences, University of Sheffield. Stock cultures were acclimated in ASTM hard water under controlled conditions at a temperature of 20 ± 2 °C, photoperiod 16h light: 8h dark and light intensity 26 μ E M ⁻²s⁻¹. Prior to experiments, animals were acclimated to test media over three weeks as recommended in the OECD guideline. The cultures were maintained in 2L tanks with approximately 25 individuals and fed every day with the green algae *C. vulgaris* fo. *Viridis* (strain number: CAAP 211/12). The algal cultures were grown in Ebert medium (Ebert group, Zoologisches Institut Evolutionsbiologie, Switzerland) and kept on a table shaker in controlled room at 20 ± 2 °C under an 8 h dark16 h light photoperiod with 40 μ E M ⁻²s⁻¹.

164

165 Experimental media

166 Daphnids were exposed in a factorial experiment to copper, predator cues (*Chaoborus flavicans* 167 extract) and antibiotics. The control medium and other treatment groups consisted of 168 autoclaved ASTM water, 300μ l Marinure (nutritional seaweed extract), and the green algae *C*. 169 *vulgaris* at 2 x 10^5 cells/ml.

170

171 We used a concentration of 5µg/l aqueous copper (II) chloride dihydrate (Fisher Scientific UK 172 C/7920/48) for the copper (Cu) treatment (see Sadeq and Beckerman, 2019a). For predation exposure, concentrated chemical cues were extracted using the procedure of (see Beckerman 173 174 et al., 2010; Carter et al., 2017; Dennis et al., 2011; Hammill et al., 2008; Lind et al., 2015; 175 Tollrian, 1995) and added to treatment media at a concentration of 1µl / ml. Nominal and realised concentrations of Cu were strongly correlated (ICP-MS; r2 = 0.99, F = 1.91, p < 0.002; 176 177 performed in a separate specialist chemistry laboratory at the University of Sheffield; see Sadeq 178 and Beckerman, 2019b).

179

The antibiotic treatment was made using ampicillin (Sigma-Aldrich Company Ltd A9393-25G) and kanamycin sulphate (Sigma Aldrich Company Ltd 60615-5G) delivered together at 9.5µg and 4µg per 150ml (sensu Sison-Mangus et al., 2015). Antibiotic treatments were delivered only on Day 1 of the experiment to 'clear' the microbiome; this allowed re-colonisation of the gut microbiota, in-situ, for the remainder of the experiments (see below). 185

186 *Experiment set up*

We performed two factorial experiments to acquire data. First, we performed a classic life table experiment with n=15 replicates per treatment to collect data on the life history and morphology of daphnia exposed to predation risk and copper, with and without antibiotic exposure. Second, we repeated this experiment, but collected n=30 individuals at maturity for microbiome analysis. The two experiments were required because analysis of the microbiome requires destructive sampling before the life table assays are completed.

193

Both experimental groups were exposed to the same conditions and following treatments:
Control, Cu, Predation, Cu-Predation, Antibiotics (AB), AB-Cu, AB-Predation and AB-CuPredation.

197

198 1. Life History Experiment

Experiments were initiated by transferring <24-hour old neonates individually into six-well plates (10ml). The daphnids were fed daily, their media changed and each individual photographed every day using a Cannon camera (EOS 350D DSLR) placed onto a Leica MZ6 modular stereomicroscope (Leica Microsystems GmbH, Wetzler, Germany).

203

Using these photos and observations, we captured data on six response variables: size at maturity, age at maturity, clutch size, induction of morphological defences, somatic growth rate and body condition represented by lipid status. Age and size at maturity were estimated as size and age of adults on the day neonates first appear in their brood pouch (max 12 days). Size was estimated with image analysis (linear measurement from head to the base of the carapace spine) using ImageJ(Rasband, 1997-2018). Clutch size was recorded by counting the number of eggs in the first clutch. Somatic growth rate was calculated as ln(size at maturity/initial size)
/ (age at maturity (days)). Lipids were counted daily and calculated as the sum of
droplets/exposure period (Gilbert, 2004; Wacker and Martin-Creuzburg, 2007). The induction
score was calculated based on a composite of pedestal size and spike number (Carter et al.,
2017; Dennis et al., 2011; Hammill et al., 2008; Lind et al., 2015).

215

216 2. Microbiome Methods (Bacterial communities' identification)

As in the life history experiment, all treatments were initiated on embryos/neonates and guts were collected from adults who had just released their first brood (e.g. Age at Maturity above). As the antibiotic treatment was on day 1 only, the guts were expected to have a microbiome acquired from living and feeding naturally after antibiotic exposure but under the control or experimental treatments.

222

The guts were dissected under a stereomicroscope with sterilised needles and transferred into phosphate buffered saline buffer (PBS) and then to micro-centrifuge tubes (1.5ml). The guts from each treatment were pooled to ensure sufficient material for microbiome sequencing and as such represent an average microbiome community among 30 individuals in each treatment. Samples were frozen in liquid nitrogen and then stored in the freezer at -20 °C for bacteria abundance and diversity analysis.

229

Samples of *D. pulex* guts were analysed using 16S rDNA sequencing (RTL Genomics, Texas,
USA). This technique is a well-established method for identifying taxonomy and phylogeny of
bacteria. The analysis yielded numerous OTU (operational taxonomic unit) data (Woo et al.
2008). OTU were taxonomically classified by identifying sequences to the highest similarity
among bacterial taxa using the SINA method (Pruesse et al., 2012).

235

236 Statistical analysis – microbiome data

Genus level OTU data for the microbiome data were analysed using the phyloseq package for R (McMurdie and Holmes, 2013). We subset the data to exclude counts of less than 200 and assessed the changes in bacterial diversity and community composition among treatments using hierarchical clustering and non-metric multidimensional scaling (NMDS). Results were qualitatively similar with exclusion criteria of 1000 counts.

242

243 Statistical analysis – phenotype data

All phenotype data were analysed using R 4.0.2 (R Core Team, 2020). We first analysed the 244 245 non-antibiotic treatment data using MANOVA to evaluate the baseline question: does the effect 246 of predation risk on body condition, life history and induced defences vary by copper exposure. 247 We then analysed the full trait data with MANOVA to evaluate whether the effect of predation risk on all phenotypic traits varied by copper and then whether this interaction (or not) varied 248 249 by the antibiotic treatment manipulating the gut microbiome. In both cases, MANOVA was 250 followed by univariate ANOVA for each individual trait. In each analysis, we used Type II 251 sums of squares implemented in the Anova() function of the car package (Fox and Weisberg, 252 2019) for R to assess significance in the MANOVA and ANOVA models.

253

254 Data availability

All data and analysis scripts are available at www.github.com/andbeck/microbiome_lifehistory

257 **RESULTS**

We first report on whether and how the microbiome was affected by the two stressors under control and antibiotic conditions. These data provide two fundamental results: whether the 260 microbial community changed – which is necessary to understand the effects reported from the 261 life table experiment – and how the microbial community changed – which is necessary to 262 generate our functional hypotheses linked to the association between microbiome and stress 263 response. Against these microbiome data, we then report on the effects of stressors and 264 antibiotic treatment on life history, morphology and condition (lipids). Specifically, we report 265 on a set of interactive and additive effects of Cu and predation that were often reversed by the 266 antibiotic treatment.

267

268 1. Microbiome

269 1.1. The diversity and composition of the bacterial communities in the digestive tract in270 response to stressors

271 We detected ~10,000 unique OTU_S among our samples. Specifically, we detected OTUs in 272 Proteobacteria, Bacteroidetes, Actinobacteria and Firmicutes phyla, Betaproteobacteria, Flavobacteriia, Gammaproteobacteria, Actinobacteria, Bacilli and Alphaproteobacteria 273 274 Flavobacteriales. classes. Burkholderiales. Xanthomonadales, Propionibacteriales. 275 Micrococcales, Bacillales, Rhizobiales, Sphingomonadales and Pseudomonadales orders, 276 Comamonadaceae, Flavobacteriaceae, Xanthomonadaceae, Propionibacteriaceae, 277 Micrococcaceae, Sanguibacteraceae, Staphylococcaceae, Phyllobacteriaceae, Pseudomonadaceae, Rhizobiaceae, Moraxellaceae 278 Sphingomonadaceae, and 279 Oxalobacteraceae families and Limnohabitans, Delftia, Chryseobacterium, Flavobacterium, 280 Stenotrophomonas. Propionibacterium, Arthrobacter, Sanguibacter. Staphylococcus, 281 Mesorhizobium, Sphingomonas, Pseudomonas, Rhizobium, Psychrobacter and 282 Janthinobacterium genera.

283

- 284 Hierarchical clustering of bacterial abundances (Fig. 2) revealed substantive changes caused
- 285 by antibiotic, Cu and predation treatments in the diversity of the bacterial community in the
- 286 gut of D. pulex. Across treatments, the dominant classes were Betaproteobacteria,
- 287 Gammaproteobacteria,

Figure 1. Heatmap based on hierarchical clustering of the bacterial community composition in different taxa (row labels; phylum, class, order, family and genus) associated with the *Daphnia pulex* gut in six treatments (column names; chronic exposure to different stress(ors). OTUs are classified using 16S rDNA gene sequences and are plotted for values >200.





Alpaproteobacteria and Flavobacteria. The data suggest that antibiotics and our stressors alter
 the relative abundance of four key genera: *Limnohabitans, Delftia, Chryseobacterium* and
 Flavobacterium. Specifically, *Limnohabitans* and *Flavobacterium* dominated the control, no-

antibiotic conditions, but these were replaced by *Delftia, Chryseobacterium* and *Stenotrophomonas* under antibiotic treatments (Fig 1).

294

295 We gained further insight into the microbiome community changes linked to predation, copper 296 and antibiotic treatments via nonmetric-multidimensional scaling analysis (Fig. 3). First, under 297 conditions with no antibiotics, the copper treatment, the predation treatment and their 298 combination each shifted the microbiome to the same community structure (Fig 2, point A). 299 Second, the addition of antibiotics shifted the control microbiome (Fig 2, point B). Finally, 300 under antibiotic treatment, each stressor also shifts the community, but now uniquely (Fig 2, 301 points C 1,2,3). In contrast to their effects under no-antibiotic conditions, each stressor is here 302 associated with a microbiome community with a distinct relative abundance





Figure 2. The bacterial diversity in *D. pulex* gut in response to two stressors, Cu and predation and to antibiotic treatment. The data are analysed via NMDS using bray-curtis dissimilarity. The red dots represent the control treatments (without antibiotics), while the blue dots are the treatments under antibiotics. Point A designates the community under copper, predation or combined stress, but in the absence of antibiotics. Point B represents the shift in the control community in the absence of antibiotics to their presence (grey dotted arrow). Points C-1, -2, and -3 indicate the community under copper, predation or mixed stressors, but after the antibiotic treatments.

304

305 under antibiotic treatments. These data show that our antibiotic treatments had substantial

- 306 effects on the gut microbiome and that the response of the microbiome to stressors also varies
- 307 by antibiotic treatments.
- 308

309 2.1 Life Table Experiment – Copper–predation interactions (No antibiotics)

- 310 The left-hand block of each Fig. 4 panel provides substantial insight into whether and how the
- 311 natural predation risk stress and the anthropogenic copper metal stress interact among six traits.
- 312 Overall, the effect of predation on the phenotype under control conditions varied by copper
- 313 (Table 1A; MANOVA, Pillai's Trace = 0.568, approximate F = 7.9, p < 0.001). Underpinning

314 this multivariate response are several significant univariate responses (Table 2A; all patterns 315 described below are associated with p.values < 0.05): Predation risk and Cu both reduced size at maturity, but their effects were additive; The effect of Cu on Age at Maturity varied by 316 317 predation rsk - Cu increased age at maturity in the absence of predation, but because age was 318 later under predation risk without copper, there was no Cu effect under predation risk. The 319 effect of Cu on somatic growth rate varied by predation risk - Cu reduced somatic growth from 320 a much higher point in the absence of predation risk; Cu reduced reproductive output (clutch 321 size), but there was no effect of predation. The effects of Cu and predation risk on mprhological 322 defences were additive - Cu reduced the morphological defence induced by predation risk; 323 Predation risk reduced body condition, but there was no effect of Cu.

324



Figure 3. Interaction plot of the effect of copper and predation risk (multi-stressors) on life history, morphological defence to predation and body condition, under control and antibiotic treatment conditions (3-way ANOVA): size at maturity (A), age at maturity (B), somatic growth rate (C), clutch size (D), induced defence (E) and body condition (F). All values are mean \pm SE, n=15.

327

328 2.2. Whole phenotype (MANOVA)

Here, and in the next section, we first report first whether the three-way interaction among stressors (predation risk, Cu and antibiotics) was significant. If this was not significant, we then report on significant two-way interactions (all p.value < 0.05).

332

The effect of predation risk on the phenotype comprised of the six traits we measured varied by the presence of copper, and this interaction varied by the antibiotic treatment (Fig 3; Table 3A; 3-way interaction; Pillai's trace = 0.156, df = 6,92, approximate F = 2.846, p = 0.0137). Against the background that the microbiome community shifted dramatically with treatments (Fig 1,2), this result suggests a strong association between how organisms respond to multiple stressors and their gut bacteria community.

339

340 2.2 Univariate Trait Responses

341 Table 2 documents the 7/12 reversals of treatment effects caused by the exposure to antibiotics 342 at Day 1 (neonates) of the experiment. These reversals underpin the statistical interaction 343 among copper, predation and antibiotic treatments revealed in the following univariate 344 ANOVAs of each trait (Table 4A; see text below).

345 TABLE 2 HERE

346 *2.2.1 Size at maturity*

347 There was no evidence of a three-way interaction (Cu: Predation:Antibiotics interaction; F =

348 1.4, df = 1,97, p = 0.25; Fig. 4a; Table 4A). We found that the effect of Predation on size at

maturity varied by antibiotics (Predation:Antibiotics interaction; F = 14.3, df = 1,97, p<0.0002) where Predation reduced body size for the bacteria-free daphnids, but increased body size under antibiotics. Furthermore, Cu had significant effect on body size which varied by antibiotic treatments (Cu:Antibiotics interaction; F = 56, df = 1,97, p<0.0003). Cu reduced body size in control treatment, but increased the size under antibiotic treatments.

354

355 2.2.2. Age at maturity

We found no evidence of a three-way interaction (Cu:Predation:Antibiotics interaction; F = 2.7, df = 1,97, p = 0.1; Fig. 4b; Table 4A). Cu had significant effect on age and this varied by antibiotics (Cu:Antibiotics interaction; F = 49, df = 1,97, p<0.0003) where maturation happened later as Cu increased under control, but earlier under antibiotic conditions. The effect of Cu on age varied by predation (Cu:Predation interaction; F = 40, df = 1,97, p<0.0008) where maturation happened later as Cu increased under control, but earlier under antibiotic reaction treatments.

363

364 *2.2.3. Growth*

The results indicated that the impact of predation on somatic growth rate did vary by Cu and this interaction varied by antibiotic exposures (Cu:Predation:Antibiotics interaction; F = 8.5, df = 1,97, p<0.004; Fig. 4c; Table 4A). Cu reduced somatic growth rate under control, but increased growth under antibiotics. The reduction without antibiotics was much weaker under predator conditions and the increase under antibiotics was much stronger under predator conditions.

371

372 *2.2.4. Clutch size*

There was no evidence of a three-way interaction (Cu: Predation:Antibiotics interaction; F = 2.4, df = 1,97, p = 0.12; Fig. 4d; Table 4A). However, we found that the effect Cu on clutch size varied by antibiotics (Cu:Antibiotics interaction; F = 12.2, df = 1,97, p<0.0007) where Cu reduced number of eggs in the absence of antibiotics and increased them in their presence.

377

378 2.2.5. Induced defence

We found that the effect of predation on neck-teeth production varied by Cu and this interaction varied by antibiotics (Cu: Predation:Antibiotics interaction; F = 7.4, df = 1,97, p<0.008; Fig. 4e; Table 4A). Cu reduced neck-teeth production under control, but increased spike production under antibiotics.

383

384 2.2.6. Body Condition

There was no evidence for a three-way interaction (Cu: Predation:Antibiotics interaction; F = 0.6, df = 1,97, p = 0.5; Fig. 4f; Table 4A). We found that the effect of predation on lipid droplets varied by antibiotic exposures (Predation:Antibiotics interaction; F = 9.2, df = 1,97, p<0.003). Predation risk reduced the number of lipid droplets under control conditions, but had no effect under antibiotics where condition was on average lower than control-control conditions.

390

391 **DISCUSSION**

Freshwater communities including lakes, ponds and streams harbour valuable biodiversity and provide important ecosystems services like freshwater and recreation. However, the organisms in them experience multiple simultaneous threats which impacts on their capacity to deliver these services. Here we evaluate the response of a keystone species in freshwater lakes and ponds – daphnia – to the combined effects of an anthropogenic stress – the heavy metal copper – and a natural stress – predation risk. 399 As detailed at the end of the introduction, we aimed to answer two major questions. The first 400 was whether the two stressors combine to affect life history, induced morphological defences 401 and body condition of daphnia in an additive versus interactive manner. The second was 402 whether the gut microbiome mediated these responses. Our experiments were motivated by 403 theory and empirical data suggesting the responses of *D. pulex* to predation and copper are both 404 mediated, in part, by digestive physiology. We found strong evidence of interactions among 405 the stressors, strong associations between gut microbiota composition and experimental 406 treatment responses and detected the antibiotic treatment driven reversal of 7/12 interactions 407 between Cu and predation risk. The manipulation of the gut microbiota is associated with 408 major changes in life history responses to multiple stressors.

409

In the following sections, we first review the array of responses to copper and predation risk under no-antibiotic conditions (no manipulation of the microbiome) against research on this topic from other laboratories. Then, we discuss in more detail how these patterns changed and were often reversed under manipulation of the gut microbiome. Finally, we discuss potential functional links between the substantial change in the abundance of the four genera of bacteria, feeding/digestive physiology and our life history, morphology and condition traits.

416

417 Phenotype Results - No Antibiotics

418 Our results demonstrate that the effects of predation risk on several life history traits varies by 419 the presence of copper (Fig 3, all left panels; Table 1A, 2A). This analysis is framed by 420 ecological and ecotoxicological theory associated with resource limitation and predation risk 421 by small size-selective predators experienced by *D. pulex*. (Table 1) and is one of the most 422 comprehensive analyses of these two stressors on multiple phenotypic traits. The theory that 423 our results align under is defined by resource limitation associated with Cu where life history 424 theory predicts later maturity, decreased somatic growth rates, smaller size at maturity and 425 smaller clutches and theory about size selective predation risk which predicts that under threats 426 from small size-selective predators, individuals mature later, but at the same or larger sizes, 427 experience faster juvenile somatic growth and have the same or larger clutches (Beckerman et 428 al., 2010).

Some of our 'no antiobiotic' results are in line with the findings of several other groups. For 430 431 example, Pyle and colleagues (DeMille et al., 2016; Hunter and Pyle, 2004; Mirza and Pyle, 432 2009) have investigated the interaction between predation risk and Cu in several Daphnia 433 species and for several traits. In their 2004 work, they found that concentrations of Cu and Ni 434 had inhibitory impacts on neck tooth induction in D. pulex (more clearly in Cu). In their 2009 435 work, they showed that Cu and predation had morphological changes in D. pulex neonates 436 leading to fewer and shorter neckteeth compared to kairomone treatments alone. DeMille et al. 437 (2016) further revealed that clones of *D. pulicaria* collected from lakes representing a gradient 438 of Cu concentrations exhibited different responses to Chaoborus kairomones in the presence 439 of Cu. This work suggests that acclimation or adaptation to local metal concentrations is linked 440 to changing capacity to respond to predation risk.

441

Our results, however, are more comprehensive than these bodies of research which are typically at the scale of a single trait response and do not evaluate how metals (or salts for example) might interfere with predator induced morphological and life history (but see the work of the Pyle group). We, and Pyle et al., suggest that because Cu interferes with digestive physiology, it may impact on induced defences by altering the pathways for allocating energy to defences.

⁴²⁹

We extended this conjecture to invoke the microbiome. We now review in more detail theresults from our manipulation of the microbiome.

449

450 Microbiome mediation

451 Our experimental manipulation of the gut microbiome with antibiotics generated four distinct 452 patterns in the bacterial microbiome community (Fig 2, Point A-C) and clear changes in the 453 relative abundance of bacteria taxa among the treatments (Fig 1). Our data show that 454 Limnohabitans and Flavobacterium dominated the control conditions, but these were replaced 455 by Delftia, Chryseobacterium and Stenotrophomonas under antibiotic treatments. Given our 456 experimental design, the new gut communities under antibiotic treatment are acquired throughout development via feeding from the existing bacterial community in the water 457 458 column. The most striking pattern in the data are that under no antibiotic conditions, where 459 parents and offspring experience the same bacterial community, exposure to either treatment 460 or their combination leads to the same microbiome community composition (Fig 2, Point A). 461 This suggests, that against the background, natural microbial community, these natural and 462 anthropogenic stressors act generically to alter gut biota in a similar way.

463

464 Our data on taxanomic diversity and relative abundance complement evidence focused on life 465 history and the role of specific microbiome taxa but in the absence of other stressors. Our data 466 suggest a functional relationship between Limnohabitans sp (Betaproteobacteria), who's 467 relative abundance was dramatically altered in our experiments and multiple traits (body 468 condition, life history and induced morphological defences) that respond to Copper and 469 predation risk. Qi et al. (2009) foundthat re-infection of aposymbiotic D. pulex to 470 Limnohabitans sp., the dominant bacterial species in D. pulex, led to elevated reproduction. 471 Peerakietkhajorn et al. (2016) found that Limnohabitans sp. play an essential role in conferring 472 fecundity by showing that bacteria-free *Daphnia* recovers fecundity when inoculated with 473 *Limnohabitans* sp. However, host-microbiota interactions within *D. magna* are known to be 474 highly specific, e.g., only certain strains of the genus *Limnohabitans* are able to recover the 475 fitness of germ-free individuals after re-inoculation (Peerakietkhajorn et al., 2015). Indeed, the 476 work of Akbar et al (2020) on associations between the microbiome, diet quality, and life 477 history revealed substantial variation in the genus Comamonadaceae, the genus in which 478 Limnohabitans is classified.

479

480 We are, however, unable to assess whether these changes are caused by stress induced 481 alteration of the gut microbiome, stress induced limited uptake of the bacteria from the 482 community or stress induced change in the external bacterial community. The latter is unlikely, 483 however, as under antibiotic treatment, the subsequent adults contain a vastly different 484 community depending on the stressors. This suggests that stressors either alter the gut 485 community or the uptake, but only when the community starting point is first altered by 486 antibiotics in the first place. Distinguishing among these mechanisms requires observation and 487 experiments that evaluate and manipulate environmental and host filtering of the microbiome 488 through ontogeny or among environmental conditions (e.g. Skelton et al., 2017).

489

A striking result of our antibiotic treatment and associated change in gut bacterial community are the 7/12 instances where the pattern under no antibiotics is reversed under antibiotic treatments (Fig 3; Table 2). For example, not only did antibiotics increase age at maturity but the effect of copper on age shifted from positive to flat under no-predator conditions and from flat to negative under predation conditions (Fig. 4). Antibiotic treatment also shifted the negative effect of copper on somatic growth and reproduction to a positive one. Finally, the negative effect of copper on induced morphological defences under predation risk with un497 manipulated microbiota was reversed when the gut microbiome had been manipulated. In the 498 absence of information on the chemical, functional and nutritional properties of all of bacteria 499 genera that comprise the change in microbiota community structure, we are unable to ascribe 500 causal inference to these patterns. However, it is clear that the community composition of 501 daphnia guts is deeply connected to their life history, to morphological defences and to body 502 condition.

503

504 Overall, our data, albeit on a single clone, support continued work to test the compelling 505 hypothesis that digestive physiology, mediated by the microbiome, has a substantial influence 506 on *D. pulex* response to multiple simultaneous stressors. This study is the first to combine a 507 study of interactions between natural and anthropogenic stress with manipulation of the gut 508 microbiota. Our data highlight stressor interactions, stress mediated changes in microbiota and 509 microbiota linked reversals of stressor interactions.

511 Table 1. Copper and Predation both affect the energy budget of organisms leading to various aligned and contrasting predicted responses in life

512 history, inducible defences and body condition. Theory and empirical data about copper is linked strongly to life history theory about starvation

513 where reduced energy leads to later maturity at a smaller size, reduced growth rates and reduced reproduction (de Shelampeare 2007). This also

514 parallels theory on non-size selective predation (Abrams and Rowe 1996). In contrast, theory about small size selective predation, as

515 implemented in this study, predicts investment into growth over reproduction, a decoupling of growth and development leading to increase in

516 size and age at maturity, induced morphological defences (Beckerman et al 2010) but equivocal data on reproduction. Copper does not generate

517 induced defences, hence the null symbol.

518

Trait	Copper	Predation Risk
Size at Maturity		
Age at Maturity		
Growth Rate	Ļ	
Reproduction		
Induced Defence		
Body Condition		-

522 Table 2. Summary of the impact of antibiotics treatment on life history, induced defence and body condition. Five of six trait responses to

523 copper are reversed by the antibiotic treatment and two of six trait responses to copper are reversed. These underpin the statistical interactions

524 detected (see Fig. 1) from the 2 x 2 x 2 factorial phenotype assay. The terms in brackets represent the direction of effect of the stress (Copper or

525 Predation) under control (normal text) or antibiotic (italics). For example, copper and predation Decreased size at maturity average under 526 control conditions. However, copper and predation *Increased* size at maturity under artification (acc Fig. 2a)

526 control conditions. However, copper and predation *Increased* size at maturity under antibiotics (see Fig 3a).

527

Trait	Copper	Predation Risk
Size at Maturity	Reversal (Decrease - Increase)	Reversal (Decrease - Increase)
Age at Maturity	Reversal (Increase - Decrease)	Consistent (Increase - Increase)
Growth Rate	Reversal (Decrease - Increase)	Consistent (Decrease - <i>Decrease)</i>
Reproduction	Reversal (Decrease - Increase)	Consistent (Decrease - <i>Decrease)</i>
Induced Defence	Reversal (Decrease - Increase)	Consistent (Increase - Increase)
Body Condition	Consistent (Decrease - Decrease)	Reversal (Decrease - No Change)

528

530 Appendix

Table 1A. MANOVA Results – No Antibiotics. This table reports multivariate Pillia's trace,
 approximate F-values, numerator and denominator degrees of freedom and p-values testing

533 whether the effect of Cu on six traits varies by predation risk (Cu:Predation).

	Df	Pillai's	approx F	n_Df	d_Df	p.value	
		Trace					
Cu	1	0.768	19.895	6	36	4.29E-10	***
Predation	1	0.684	12.994	6	36	9.15E-08	***
Cu:Predation	1	0.568	7.9003	6	36	1.80E-05	***

- 538 Table 2A. Univariate ANOVA Results No Antibiotics. This table reports Sums of Squares,
- 539 F-values, degrees of freedom and P-values testing whether the effect of Cu on each of the six
- 540 traits varies by predation risk (Cu:Predation). 'size' = size at maturity; 'age' = age at

541 maturity; 'growth' = somatic growth rate; 'repro' = reproduction/clutch size; 'morph' =

- 542 induced morphological defence; 'lipid' = body condition.
- 543

	p.value	F-values	df	SS	term	model
***	4.05E-08	45.2	1	0.141	Cu	size
**	0.00195	11	1	0.0343	Predation	size
	0.169	1.96	1	0.00614	Cu:Predation	size
**	0.0016	11.4	1	2.16	Cu	age
***	5.54E-05	20.2	1	3.82	Predation	age
***	7.56E-07	34	1	6.41	Cu:Predation	age
***	1.63E-08	49	1	5.97E-04	Cu	growth
***	2.31E-05	22.8	1	2.78E-04	Predation	growth
***	5.31E-06	27.4	1	3.33E-04	Cu:Predation	growth
**	0.00667	8.17	1	8.04	Cu	repro
	0.107	2.71	1	2.67	Predation	repro
	0.338	0.94	1	0.926	Cu:Predation	repro
**	0.00629	8.3	1	1500	Cu	morph
***	3.57E-08	45.7	1	8270	Predation	morph
	0.0618	3.69	1	667	Cu:Predation	morph
	0.551	0.361	1	7.76	Cu	lipid
**	0.00157	11.5	1	246	Predation	lipid
	0.271	1.24	1	26.7	Cu:Predation	lipid

545

546 547

548

- 550 Table 3A MANOVA Results Full Model with Antibiotics. This table reports multivariate
- 551 Pillia's trace, approximate F-values, numerator and denominator degrees of freedom and P-
- values testing whether the effect of Cu on six traits varies by predation risk and whether these
- 553 interactions vary by Antibiotic treatment (Cu:Predation:Antibiotic).
- 554

	Df	Pillai's	approx F	n_Df	d_Df	p.value	
		Trace					
Cu	1	0.140	2.507	6	92	0.0271754	*
Predation	1	0.652	28.746	6	92	< 2.2e-16	***
Antibiotic	1	0.517	16.470	6	92	8.38E-13	***
Cu:Predation	1	0.330	7.571	6	92	1.29E-06	***
Cu:Antibiotic	1	0.634	26.667	6	92	< 2.2e-16	***
Predation:Antibiotic	1	0.233	4.663	6	92	0.0003499	***
Cu:Predation:Antibiotic	1	0.156	2.846	6	92	0.0137695	*

555

556

558 Table 4A – Univariate ANOVA Results – Full Model with Antibiotics. This table reports

559 Sums of Squares, F-values, degrees of freedom and P-values testing whether the effect of Cu

560 on six traits varies by predation risk and whether these interactions vary by Antibiotic

treatment (Cu:Predation:Antibiotic). 'size' = size at maturity; 'age' = age at maturity;
'growth' = somatic growth rate; 'repro' = reproduction/clutch size; 'morph' = induced

- 563 morphological defence; 'lipid' = body condition.
- 564

model	term	sumsq	df	f-statistic	p.value	sig
size	Cu	0.0125	1	4.35	0.0397	*
size	Predation	0.0012	1	0.418	0.519	
size	Antibiotic	0.0682	1	23.7	4.38E-06	***
size	Cu:Predation	0.00225	1	0.783	0.378	
size	Predation:Antibiotic	0.16	1	55.7	3.69E-11	***
size	Cu:Antibiotic	0.0413	1	14.3	2.65E-04	***
size	Cu:Predation:Antibiotic	0.00389	1	1.35	0.248	
age	Cu	1.24	1	5.2	0.0248	*
age	Predation	9.83	1	41.1	5.23E-09	***
age	Antibiotic	12.2	1	51.1	1.62E-10	***
age	Cu:Predation	9.53	1	39.8	8.31E-09	***
age	Predation:Antibiotic	11.7	1	48.7	3.64E-10	***
age	Cu:Antibiotic	0.135	1	0.566	0.454	
age	Cu:Predation:Antibiotic	0.635	1	2.65	0.106	
growth	Cu	8.31E-06	1	0.691	0.408	
growth	Predation	2.75E-04	1	22.8	6.25E-06	***
growth	Antibiotic	7.34E-04	1	61.1	6.65E-12	***
growth	Cu:Predation	2.77E-04	1	23.1	5.73E-06	***
growth	Predation:Antibiotic	0.00104	1	86.6	4.28E-15	***
growth	Cu:Antibiotic	3.72E-05	1	3.09	0.0819	
growth	Cu:Predation:Antibiotic	1.03E-04	1	8.55	0.00431	**
repro	Cu	1.1	1	1.75	0.189	
repro	Predation	6.93	1	11	0.00127	**
repro	Antibiotic	0.00495	1	0.00787	0.93	
repro	Cu:Predation	2.44E-04	1	3.88E-04	0.984	
repro	Predation:Antibiotic	7.69	1	12.2	7.12E-04	***
repro	Cu:Antibiotic	5.29E-04	1	8.41E-04	0.977	
repro	Cu:Predation:Antibiotic	1.53	1	2.42	0.123	
morph	Cu	69.9	1	0.376	0.541	
morph	Predation	23300	1	125	3.62E-19	***
morph	Antibiotic	927	1	4.99	0.0277	*
morph	Cu:Predation	29.1	1	0.156	0.693	
morph	Predation:Antibiotic	1930	1	10.4	0.00172	**
morph	Cu:Antibiotic	30.5	1	0.164	0.686	
morph	Cu:Predation:Antibiotic	1370	1	7.39	0.00777	**

lipid	Cu	12.8	1	1.01	0.318	
lipid	Predation	118	1	9.31	0.00295	**
lipid	Antibiotic	0.0582	1	0.0046	0.946	
lipid	Cu:Predation	25.2	1	1.99	0.162	
lipid	Predation:Antibiotic	2.47	1	0.195	0.66	
lipid	Cu:Antibiotic	116	1	9.15	0.00318	**
lipid	Cu:Predation:Antibiotic	6.91	1	0.546	0.462	

565

566

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572

573 **Conflict of interest**

574 The authors have no conflicts of interest associated with this publication.

575

576 **REFERENCES**

577

578	Akbar S, Gu L, Sun Y, Zhou Q, Zhang L, Lyu K, et al. Changes in the life history traits of
579	Daphnia magna are associated with the gut microbiota composition shaped by diet
580	and antibiotics. Science of The Total Environment 2020; 705: 135827.

- Altshuler I, Demiri B, Xu S, Constantin A, Yan ND, Cristescu ME. An integrated multi disciplinary approach for studying multiple stressors in freshwater ecosystems:
 Daphnia as a model organism. Oxford University Press, 2011.
- Balseiro E, Modenutti B, Queimalinos C, Reissig M. Daphnia distribution in Andean
 Patagonian lakes: effect of low food quality and fish predation. Aquatic Ecology
 2007; 41: 599-609.
- Banerjee S, Schlaeppi K, van der Heijden MGA. Keystone taxa as drivers of microbiome
 structure and functioning. Nature Reviews Microbiology 2018; 16: 567-576.
- Barata C, Baird DJ. Determining the ecotoxicological mode of action of chemicals from
 measurements made on individuals: results from instar- based tests with Daphnia
 magna Straus. Aquatic Toxicology 2000; 48: 195-209.

Barata C, Baird DJ, Minarro A, Soares A. Do genotype responses always converge from lethal to nonlethal toxicant exposure levels? Hypothesis tested using clones of Daphnia magna straus. Environmental Toxicology and Chemistry 2000; 19: 23142322.

596 Beckerman AP, Rodgers GM, Dennis SR. The reaction norm of size and age at maturity 597 under multiple predator risk. Journal Of Animal Ecology 2010; 79: 1069-1076. 598 Beckerman AP, Wieski K, Baird DJ. Behavioural versus physiological mediation of life 599 history under predation risk. Oecologia 2007; 152: 335-343. 600 Benard MF. Predator-induced phenotypic plasticity in organisms with complex life histories. Annual Review of Ecology Evolution and Systematics 2004; 35: 651-673. 601 602 Black AR, Dodson SI. Demographic Costs of Chaoborus-Induced Phenotypic Plasticity in 603 Daphnia pulex. Oecologia 1990; 83: 117-122. 604 Bui T-KL, Do-Hong LC, Dao T-S, Hoang TC. Copper toxicity and the influence of water 605 quality of Dongnai River and Mekong River waters on copper bioavailability and 606 toxicity to three tropical species. Chemosphere 2016; 144: 872-878. 607 Campero M, Slos S, Ollevier F, Stoks R. Sublethal pesticide concentrations and predation 608 jointly shape life history: behavioral and physiological mechanisms. Ecological 609 Applications 2007; 17: 2111-2122. 610 Carter MJ, Lind MI, Dennis SR, Hentley W, Beckerman AP. Evolution of a predatorinduced, nonlinear reaction norm. Proceedings of the Royal Society B-Biological 611 612 Sciences 2017; 284. 613 Coors A, De Meester L. Synergistic, antagonistic and additive effects of multiple stressors: 614 predation threat, parasitism and pesticide exposure in Daphnia magna. Journal of 615 Applied Ecology 2008; 45: 1820-1828. Coors A, Hammers-Wirtz M, Ratte HT. Adaptation to environmental stress inDaphnia 616 617 magnasimultaneously exposed to a xenobiotic. Chemosphere 2004; 56. 618 De Schamphelaere KA, Forrez I, Dierckens K, Sorgeloos P, Janssen CR. Chronic toxicity of 619 dietary copper to Daphnia magna. Aquatic Toxicology (Amsterdam, Netherlands) 2007; 81: 409-418. 620 621 DeMille C, Arnott S, Pyle G. Variation in copper effects on kairomone-mediated responses in 622 Daphnia pulicaria. Ecotoxicology and environmental safety 2016; 126: 264-272. 623 Dennis SR, Carter MJ, Hentley WT, Beckerman AP. Phenotypic convergence along a 624 gradient of predation risk. Proceedings of the Royal Society B-Biological Sciences 625 2011; 278: 1687-1969. 626 Folt CL, Chen CY, Moore MV, Burnaford J. Synergism and antagonism among multiple 627 stressors. Limnology and Oceanography 1999; 44: 864-877. 628 Fox J, Weisberg S. An R companion to applied regression: Sage publications, 2019. 629 Gilbert JJ. Females from resting eggs and parthenogenetic eggs in the rotifer Brachionus 630 calyciflorus: lipid droplets, starvation resistance and reproduction. Freshwater 631 Biology 2004; 49: 1505-1515. 632 Gillis P, Chow-Fraser P, Ranville J, Ross P, Wood C. Daphnia need to be gut-cleared too: the 633 effect of exposure to and ingestion of metal-contaminated sediment on the gut-634 clearance patterns of D. magna. Aquatic toxicology 2005; 71: 143-154. Grossart HP, Dziallas C, Tang KW. Bacterial diversity associated with freshwater 635 636 zooplankton. Environmental Microbiology Reports 2009; 1: 50-55. 637 Hammill E, Rogers A, Beckerman AP. Costs, benefits and the evolution of inducible defences: a case study with Daphnia pulex. Journal of Evolutionary Biology 2008; 21: 638 639 705-715. 640 Hecky R, Kilham P. Nutrient limitation of phytoplankton in freshwater and marine environments: a review of recent evidence on the effects of enrichment 1. Limnology 641 and oceanography 1988; 33: 796-822. 642 643 Hunter K, Pyle G. Morphological responses of Daphnia pulex to Chaoborus americanus 644 kairomone in the presence and absence of metals. Environmental Toxicology and 645 Chemistry: An International Journal 2004; 23: 1311-1316.

647 in freshwater ecosystems: a meta-analysis. Global Change Biology 2016; 22: 180-648 189. 649 Jansen M, De Meester L, Cielen A, Buser CC, Stoks R. The interplay of past and current 650 stress exposure on the water flea Daphnia. Functional Ecology 2011; 25: 974-982. Lind MI, Yarlett K, Reger J, Carter MJ, Beckerman AP. The alignment between phenotypic 651 652 plasticity, the major axis of genetic variation and the response to selection. 653 Proceedings of the Royal Society of London B: Biological Sciences 2015; 282. 654 Long KE, Van Genderen EJ, Klaine SJ. The effects of low hardness and pH on copper 655 toxicity to Daphnia magna. Environmental Toxicology and Chemistry: An 656 International Journal 2004; 23: 72-75. Loureiro S, Svendsen C, Ferreira AL, Pinheiro C, Ribeiro F, Soares AM. Toxicity of three 657 658 binary mixtures to Daphnia magna: comparing chemical modes of action and 659 deviations from conceptual models. Environmental Toxicology and Chemistry 2010; 660 29: 1716-1726. 661 Martins C, Jesus FT, Nogueira AJ. The effects of Copper and Zinc on survival, growth and 662 reproduction of the cladoceran Daphnia longispina: introducing new data in an "old" 663 issue. Ecotoxicology 2017; 26: 1157-1169. McMurdie PJ, Holmes S. phyloseq: An R Package for Reproducible Interactive Analysis and 664 665 Graphics of Microbiome Census Data. PLOS ONE 2013; 8: e61217. Mirza R, Pyle G. Waterborne metals impair inducible defences in Daphnia pulex: 666 667 morphology, life-history traits and encounters with predators. Freshwater Biology 668 2009; 54: 1016-1027. 669 Mushegian AA, Arbore R, Walser J-C, Ebert D. Environmental sources of bacteria and 670 genetic variation in behavior influence host-associated microbiota. Applied and 671 environmental microbiology 2019; 85. Noonburg EG, Nisbet RM. Behavioural and physiological responses to food availability and 672 predation risk. Evolutionary Ecology Research 2005; 7: 89-104. 673 Orr JA, Vinebrooke RD, Jackson MC, Kroeker KJ, Kordas RL, Mantyka-Pringle C, et al. 674 675 Towards a unified study of multiple stressors: divisions and common goals across 676 research disciplines. Proceedings of the Royal Society B: Biological Sciences 2020; 677 287: 20200421. 678 Peerakietkhajorn S, Kato Y, Kasalický V, Matsuura T, Watanabe H. Betaproteobacteria L imnohabitans strains increase fecundity in the crustacean D aphnia magna: symbiotic 679 680 relationship between major bacterioplankton and zooplankton in freshwater 681 ecosystem. Environmental microbiology 2016; 18: 2366-2374. 682 Peerakietkhajorn S, Tsukada K, Kato Y, Matsuura T, Watanabe H. Symbiotic bacteria contribute to increasing the population size of a freshwater crustacean, D aphnia 683 684 magna. Environmental Microbiology Reports 2015; 7: 364-372. 685 Pestana JLT, Loureiro S, Baird DJ, Soares A. Fear and loathing in the benthos: Responses of 686 aquatic insect larvae to the pesticide imidacloprid in the presence of chemical signals 687 of predation risk. Aquatic Toxicology 2009; 93: 138-149. Pestana JLT, Loureiro S, Baird DJ, Soares AMM. Pesticide exposure and inducible 688 689 antipredator responses in the zooplankton grazer, Daphnia magna Straus. 690 Chemosphere 2010; 78: 241-248. 691 Pruesse E, Peplies J, Glöckner FO. SINA: accurate high-throughput multiple sequence 692 alignment of ribosomal RNA genes. Bioinformatics 2012; 28: 1823-1829. 693 Qi W, Nong G, Preston JF, Ben-Ami F, Ebert D. Comparative metagenomics of Daphnia 694 symbionts. BMC genomics 2009; 10: 172. Rasband WS. Image J. U.S. National Institutes of Health, Bethesda, 695

Jackson MC, Loewen CJG, Vinebrooke RD, Chimimba CT. Net effects of multiple stressors

- 696 Maryland, USA, 1997-2018.
- Rose RM, Warne MS, Lim RP. Some life history responses of the cladoceran Ceriodaphnia
 cf. dubia to variations in population density at two different food concentrations.
 Hydrobiologia 2002; 481: 157-164.

Rose RM, Warne MSJ, Lim RP. Factors associated with fish modify life history traits of the
 cladoceran Ceriodaphnia cf. dubia. Journal of Plankton Research 2001; 23: 11-17.

- Sadeq SA, Beckerman AP. The chronic effects of copper and cadmium on life history traits
 across Cladocera species: a meta-analysis. Archives of environmental contamination
 and toxicology 2019a; 76: 1-16.
- Sadeq SA, Beckerman AP. Evaluating additive versus interactive effects of copper and
 cadmium on Daphnia pulex life history. Environmental Science and Pollution
 Research 2019b.
- Schulz R, Dabrowski JM. Combined effects of predatory fish and sublethal pesticide
 contamination on the behavior and mortality of mayfly nymphs. Environmental
 Toxicology and Chemistry: An International Journal 2001; 20: 2537-2543.
- Shuhaimi-Othman M, Nadzifah Y, Ahmad A. Toxicity of copper and cadmium to freshwater
 fishes. World Acad Scie Engin Tech 2010; 65: 869-871.
- Sison-Mangus MP, Mushegian AA, Ebert D. Water fleas require microbiota for survival,
 growth and reproduction. ISME J 2015; 9: 59-67.
- Skelton J, Geyer KM, Lennon JT, Creed RP, Brown BL. Multi-scale ecological filters shape
 the crayfish microbiome. Symbiosis 2017; 72: 159-170.
- Stoks R, McPeek MA. Antipredator behavior and physiology determine Lestes species
 turnover along the pond-permanence gradient. Ecology 2003; 84: 3327-3338.
- Taylor BE, Gabriel W. To Grow or Not to Grow Optimal Resource-Allocation for Daphnia.
 American Naturalist 1992; 139: 248-266.
- Team RC. R: A Language and Environment for Statistical Computing Vienna, Austria: R
 Foundation for Statistical Computing, 2020.
- Tollrian R. Chaoborus Crystallinus Predation on *Daphnia pulex* Can Induced
 Morphological-Changes Balance Effects of Body-Size on Vulnerability. Oecologia
 1995; 101: 151-155.
- Tollrian R, Dodson SI. Predator induced defenses in cladocerans. In: Tollrian R, Harvell CD,
 editors. The Ecology and Evolution of Inducible Defenses. Princeton University
 Press, Princeton, NJ, 1999.
- Wacker A, Martin-Creuzburg D. Allocation of essential lipids in Daphnia magna during
 exposure to poor food quality. Functional Ecology 2007; 21: 738-747.