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1 Supplementary Materials

2	Title: Extraction of extracellular polymeric substances (EPS) from Red Soils (Ultisols)
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1. Materials and methods

24 **1.1 Site description and soil sampling**

25 Two soils of contrasting ecosystems and Fe content were selected for the study. Forest soil was sampled from a cedar forest (114°21'E, 30°1'N) located in the He 26 27 Shengqiao town, Xianning city, Hubei province, China. The region is characterized by a subtropical climate, with an average annual temperature of 16.8 °C and precipitation 28 of 1300 mm. Paddy soil was collected from a paddy field under a rice-rice cropping 29 system from National Agro-Ecosystem Observation and Research Station (116°55'E, 30 31 28°15'N) in Yingtan city, Jiangxi province, China. This region is characterized by a typical subtropical monsoonal climate, with an average annual temperature of 17.6 °C 32 and precipitation of 1785 mm. Both soils are derived from Quaternary red clay. The 33 34 soils were sampled randomly from 0-20 cm depth with three replicates per site, then soil sample were homogeneous mixing to produce composite soils. Then the soil 35 samples (forest soils and paddy soils) were sieved (2 mm) and divided into divided 36 37 into two subsamples. One subsample was air dried and used to analyze soil properties, 38 while the other subsample was incubated for extraction of EPS as described below.

39

1.2 Determination of soil characteristics

Soil pH was measured in triplicate in soil-water prepared at a ratio of 1: 2.5. Soil
particle size distributions were determined using the laser diffraction method
(Mastersizer 3000, Malvern, UK). The contents of organic carbon and total nitrogen
were analyzed in triplicate with an elemental analyzer (Vario MICRO cube, Elementar,
Germany). The free iron oxides (Fe_d) and amorphous iron oxides (Fe_o) were extracted

using dithionite-citrate-bicarbonate (DCB) solution and oxalic acid
ammonium-oxalate solution (McKeague and Day 1966), respectively. The extracts
were diluted and iron concentration was measured with an atomic absorption
spectrophotometer (AA240FS, Varian, USA).

49 **1.3 Soil incubation**

Inevitably, artefacts can be caused by soil sampling and sieving, which may 50 affect subsequent analysis. For example, soil sampling may cause cell lysis. 51 52 Consequently, we may overestimate cell lysis if we do not include the incubation 53 treatment. Thus, when we compared the extraction methods, the soils sample were incubated. Briefly, the subsample (100 g dry weight equivalent) were pre-incubated at 54 25 °C for 2 weeks to ensure microbial activity had stabilized caused by sampling, 55 56 sieving and adjustment of soil moisture (e.g. Redmile-Gordon et al., 2014; Brookes et al., 2017). During the incubation period, the water content of the soils was kept at 57 60% of maximum water holding capacity (WHC). In order to maintain constant soil 58 59 water content, deionized water was added gravimetrically twice weekly. After pre-incubation, microcosms were incubated under the same conditions but with an 60 added substrate for a further 10 days as did Redmile-Gordon et al. (2014), since EPS 61 62 production is understood to be the greatest sometime shortly after the exponential 63 phase of growth. The exponential phase of growth in soils given additional substrate at these temperatures tends to last between 1 and 7 days. Hence, the time-point of 10 64 65 days was selected to ensure this phase had passed. The extra substrate was used to stimulate the microbial biomass to produce EPS (Nunan et al. 2003). Glycerol was 66

selected since it produced no sugars or proteins on decomposition (Redmile-Gordon 67 et al. 2014), which could otherwise affect the quantification of components in EPS. 68 The glycerol was added to soils at a concentration of 10 mg C g⁻¹ soil. In addition, to 69 ensure that the growth of soil microbes was not limited by the lack of nutrients, 70 71 ammonium nitrate and mono-ammonium phosphate were added to soils at concentrations of 1.16 mg N g⁻¹ soil and 0.166 mg P g⁻¹ soil, respectively. We added 72 C:N:P according to average global stoichiometric ratios of C:N:P in soil microbial 73 biomass (60:7:1) (Cleveland and Liptzin, 2007). 74

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1.4 Extraction of EPS

Before EPS extraction, soluble microbial products (SMP) were extracted from 76 soils (Redmile-Gordon et al., 2014). Soluble microbial products (SMP) are freely 77 soluble extracellular polymers not actually bound to the cells or EPS matrix (see 78 Comte et al., 2006). The transient nature of SMP means that it is not a defining 79 component of the extracelluar matrix or 'biofilm'. Therefore, any extraction method 80 81 should include SMP. To extract SMP, 25 mL pre-cooled CaCl₂ solution (pH 7.0, 10 mM) was added to the moist soil (2.5 g dry weight equivalent), shaken at 120 rpm 82 (4 °C) for 30 min, then centrifuged at 4000 g for 30 min (4 °C). EPS was 83 subsequently extracted from the residues with the followin five method. 84

85

1.4.1 Hot water extractable polysaccharide (HWEP)

The HWEP extraction methods were slightly modified from the method described by Ghani et al. (2003). This method were frequently used to extract EPS EPS-like fraction (e.g. polysaccharides) from soils. Briefly, 25 mL of ultra-pure water was added to the residues (after extraction of SMP), incubated in a water bath (80 °C)
for 7 h, then centrifuged at 4000 g for 30 min (4 °C). The supernatant was passed
through 0.45 µm cellulose nitrate membrane filters and then was used for EPS
determination (within 4 days). The pellet was washed using phosphate buffer saline
(PBS), centrifuged at 4000 g for 30 min and the supernatant of PBS was discarded,
then the pellet was stored at -80 °C pending ATP analysis.

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5 **1.4.2 Hot dilute acid extractable polysaccharide (HDAEP) extraction**

The HWEP extraction methods were slightly modified from the method 96 97 described by Spohn and Giani (2010). This method were frequently used to extract EPS-like fraction (e.g. polysaccharides) from soils. Briefly, 25 mL of dilute sulfuric 98 acid (0.125 M H₂SO₄) was added to the residues (after extraction of SMP), incubated 99 100 in a water bath (80 °C) for 7 h, then centrifuged at 4000 g for 30 min (4 °C). The supernatant was passed through 0.45 µm cellulose nitrate membrane filters and then 101 was used for EPS determination (within 4 days). The pellet was washed using 102 103 phosphate buffer saline (PBS), centrifuged at 4000 g for 30 min and the supernatant of PBS was discarded, then the pellet was stored at -80 °C pending ATP analysis. 104

105 **1.4.3 Easily extractable glomalin (EEG) extraction**

The EEG protocol was first developed by Wright and Upadhyaya (1996) to extract a protein (i.e. glomalin) deposited by arbuscular mycorrhizal fungi (AMF). The protocol described by Wright and Upadhyaya (1996) was followed. Briefly, 20 mL of 20 mM citrate buffer (pH 7.0) was added to the residues (after extraction of SMP), autoclaved at 121 °C for 30 min, then centrifuged at 4000 g for 30 min (4 °C). The supernatant was passed through 0.45 μm cellulose nitrate membrane filters and then was used for EPS determination (within 4 days). The pellet was washed using phosphate buffer saline (PBS), centrifuged at 4000 g for 30 min and the supernatant of PBS was discarded, then the pellet was stored at -80 °C pending ATP analysis.

115 **1.4.4 Sodium sulfide extraction**

The theory of EPS extraction with SS is that SS can reduce the Fe³⁺ to insoluble 116 FeS, thus resulting in disintegration of Fe³⁺ bound EPS matrix Nielsen and Keiding 117 (1998). The extraction procedure was slightly modified from the method described by 118 119 Zhu et al. (2015). Briefly, 25 mL of 20 mM sodium sulfide (Na₂S•9H₂O) was added to the residues (after extraction of SMP), shaken at 180 rpm (4 °C) for 2 h, then 120 centrifuged at 4000 g for 30 min. The supernatant was passed through 0.45 µm 121 122 cellulose nitrate membrane filters and then was used for EPS determination (within 4 days). The pellet was washed using phosphate buffer saline (PBS), centrifuged at 123 4000 g for 30 min and the supernatant of PBS was discarded, then the pellet was 124 125 stored at -80 °C pending ATP analysis.

126 **1.4.5 Cation exchange resin (CER) extraction**

Extraction buffer was prepared as 4 mM NaH₂PO₄, 2 mM Na₃PO₄, 9 mM NaCl and 1 mM KCl at pH 7.0 (Frolund et al. 1996). CER (91973, Sigma-Aldrich) was pretreated for a pH of 7.0 and was washed three times with the EPS extraction buffer before use. The CER was added at the amount of 70 g CER g⁻¹ volatile solids (VS), and the quantity of VS in the soil was determined according to Redmile-Gordon et al. (2014). 70 g CER g⁻¹ volatile solids (VS) was widely used in EPS extraction from 133 sludge (e.g. Frolund et al., 1996), since EPS can be extracted with maximum 134 efficiency with this amount CER. The theory of EPS extraction with CER is that a 135 combination of shear forces and resin-Na⁺ cause ion exchange with multivalent 136 cations (mainly Ca²⁺ and Mg²⁺) that link EPS resulting in dissolution of EPS 137 macrostructure (Wilén et al., 2003).

To extract EPS, 25 mL of the pre-cooled extraction buffer and pre-weighed CER were added to the residues (after extraction of SMP), shaken at 180 rpm (4 °C) for 2 h, then centrifuged at 4000 g for 30 min. The supernatant was passed through 0.45 μm cellulose nitrate membrane filters and then was used for EPS determination (within 4 days). The pellet was washed using phosphate buffer saline (PBS), centrifuged at 4000 g for 30 min and the supernatant of PBS was discarded, then the pellet was stored at -80 °C pending ATP analysis.

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1.5 Characterization of EPS

The polysaccharide content was determined by the anthrone-sulfuric acid method (Brink Jr et al., 1960) using glucose (G116307, Aladdin) as the standard. The protein content was estimated using the Bradford method (Bradford, 1976) with bovine serum albumin (A104912, Aladdin) as the standard. The HAE content was measured using the method described by Wang and Fujii (2011) using humic acid as the standard (H16752, Sigma-Aldrich). All of these components of EPS were quantified by UV-visible spectroscopy (A580, Ao Yi Instrument CO. LED., China).

153 **1.6 Measurement of ATP content of soils before and after EPS extraction**

154 ATP was extracted from soils (without EPS extraction) and the pellet (after EPS

extraction) using the method of Redmile-Gordon et al. (2011) with Extractant A and B. 155 Extractant A contained 1.1 M trichloroacetic acid, 0.6 M imidazole, 0.25 M sodium 156 hydrogen phosphate. Extractant B was similar to Extractant A, except that it contained 157 5 mL added 0.1 mM ATP. During extraction, some of the ATP may be sorbed on soil 158 colloids, or denatured. This was corrected for by determining the proportion of added 159 ATP recovered in soil Extract B and using this percentage recovery to correct for the 160 same loss processes in soil during extraction with soil Extractant A. Briefly, moist 161 soils or the pellet with 25 ml Extractant A or B were ultrasonified for 2 min, cooled on 162 163 ice for 5 min, then filtered (Whatman 42).

The analysis of ATP as described by Qiu et al. (2016). Briefly, 10 μ L extract was added to 150 μ L arsenate buffer, 13 μ L 1M NaOH, and 50 μ L luciferin-luciferase (GN202-01, Beijing yuanpinghao biotechnology co. LTD, China) and the bioluminescence of the mixture was measured using a Multimode Plate Reader (EnVision, PerkinElmer, USA) in dark 96-well microplates.

169 **1.7 Statistical analyses**

The statistical analysis was conducted with SPSS 23.0. We check normality and homoscedasticity of variances by ShapiroeWilks test and Levene's test, respectively. In some cases, data was log transformed to meet the normal distribution. An analysis of variance (ANOVA) followed by Duncan's multiple-comparison test at p < 0.05 was used to evaluate the differences between EPS content extracted by different methods in the same soil. Student's t-tests were used to evaluate the statistical significance of differences between means, of 1) the content of C, N and iron oxides in the two soils, and 2) soil microbial ATP content in the same soil before and after EPS extraction.

178 **References**

- Bradford, M.M., 1976. A rapid and sensitive method for the quantitation of
 microgram quantities of protein utilizing the principle of protein-dye binding.
 Analytical Biochemistry 72, 248–254.
- Brink Jr, R.H., Dubach, P., Lynch, D.L., 1960. Measurement of carbohydrates in soil
 hydrolyzates with anthrone. Soil Science 89, 157–166.
- Cleveland, C.C., Liptzin, D., 2007. C: N: P stoichiometry in soil: is there a "Redfield
 ratio" for the microbial biomass?. Biogeochemistry 85, 235–252.
- Comte, S., Guibaud, G., Baudu, M., 2006. Biosorption properties of extracellular
 polymeric substances (EPS) resulting from activated sludge according to their
 type: soluble or bound. Process Biochemistry 41, 815–823.
- 189 Filisetti-Cozzi, T.M., Carpita, N.C., 1991. Measurement of uronic acids without

interference from neutral sugars. Analytical Biochemistry 197, 157–162.

- Ghani, A., Dexter, M., Perrott, K.W., 2003. Hot-water extractable carbon in soils: a
 sensitive measurement for determining impacts of fertilisation, grazing and
 cultivation. Soil Biology and Biochemistry 35, 1231–1243.
- McKeague, J.A., Day, J., 1966. Dithionite-and oxalate-extractable Fe and Al as aids in
 differentiating various classes of soils. Canadian Journal of Soil Science 46,
 13–22.
- 197 Nunan, N., Wu, K.J., Young, I.M., Crawford, J.W., Ritz, K., 2003. Spatial distribution

- of bacterial communities and their relationships with the micro-architecture of
 soil. FEMS Microbiology Ecology 44, 203–215.
- Qiu, G.Y., Chen, Y.F., Luo, Y., Xu, J.M., Brookes, P.C., 2016. The microbial ATP
 concentration in aerobic and anaerobic Chinese soils. Soil Biology and
 Biochemistry 92, 38-40.
- Redmile-Gordon, M.A., White, R.P., Brookes, P.C., 2011. Evaluation of substitutes
 for paraquat in soil microbial ATP determinations using the trichloroacetic acid
 based reagent of Jenkinson and Oades (1979). Soil Biology and Biochemistry 43,
 1098–1100.
- Redmile-Gordon, M.A., Brookes, P.C., Evershed, R.P., Goulding, K.W.T., Hirsch, P.R.,
 208 2014. Measuring the soil-microbial interface: Extraction of extracellular
 209 polymeric substances (EPS) from soil biofilms. Soil Biology and Biochemistry
 210 72, 163–171.
- Spohn, M., Giani, L., 2010. Water-stable aggregates, glomalin-related soil protein,
 and carbohydrates in a chronosequence of sandy hydromorphic soils. Soil
 Biology and Biochemistry 42, 1505–1511.
- Wright, S.F., Upadhyaya, A., 1996. Extraction of an abundant and unusual protein
 from soil and comparison with hyphal protein of arbuscular mycorrhizal fungi.
 Soil Science 161, 575–586.
- Wang, Y., Fujii, T., 2011. Evaluation of methods of determining humic acids in
 nucleic acid samples for molecular biological analysis. Bioscience

- Biotechnology and Biochemistry 75, 355–357.
- 220 Zhu, L., Yu, H.T., Liu, Y.M., Qi, H.Y., Xu, X.Y., 2015. Optimization for extracellular
- 221 polymeric substances extraction of microbial aggregates. Water Science and
- 222 Technology 71, 1106–1112.

- **Table S1** Physical and chemical properties of the two soils. Data (means \pm SE, n = 3) followed by the different letters within a column indicate
- significant differences (p < 0.05).

Soil	Clay (%)	Silt (%)	Sand (%)	pН	SOC (g kg ⁻¹)	TN (g kg ⁻¹)	C/N ratio	$\operatorname{Fe}_{d}(\operatorname{g}\operatorname{kg}^{-1})$	$\operatorname{Fe}_{o}\left(g\ \mathrm{kg}^{-1}\right)$
Forest soil	18.5±0.5	79.6±0.6	1.9±0.2	5.6	8.75±0.15b	0.86±0.33b	9.9±0.3b	29.4±0.5a	2.53±0.07a
Paddy soil	15.6±0.2	61.7±0.7	22.7±0.9	4.7	21.66±0.58a	1.84±0.05a	11.8±0.1a	6.8±0.8b	1.64±0.03b

Abbreviations: SOC, soil organic carbon; TN, total nitrogen; Fe_d, dithionite-citrate-bicarbonate extractable iron oxides; Fe_o, oxalate extractable

iron oxides.