



UNIVERSITY OF LEEDS

This is a repository copy of *Potential shift from a carbon sink to a source in Amazonian peatlands under a changing climate*.

White Rose Research Online URL for this paper:
<http://eprints.whiterose.ac.uk/146088/>

Version: Accepted Version

Article:

Wang, S, Zhuang, Q, Lähteenoja, O et al. (2 more authors) (2018) Potential shift from a carbon sink to a source in Amazonian peatlands under a changing climate. *Proceedings of the National Academy of Sciences*, 115 (49). pp. 12407-12412. ISSN 0027-8424

<https://doi.org/10.1073/pnas.1801317115>

(c) 2018. Published under the PNAS license. This is an author produced version of a paper published in the *Proceedings of the National Academy of Sciences*. Uploaded in accordance with the publisher's self-archiving policy.

Reuse

Items deposited in White Rose Research Online are protected by copyright, with all rights reserved unless indicated otherwise. They may be downloaded and/or printed for private study, or other acts as permitted by national copyright laws. The publisher or other rights holders may allow further reproduction and re-use of the full text version. This is indicated by the licence information on the White Rose Research Online record for the item.

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

1 **A potential shift from a carbon sink to a source in Amazonian peatlands under a changing**
2 **climate**

3 ¹Sirui Wang, ¹Qianlai Zhuang, ²Outi Lahteenoja, ^{3,4}Frederick C Draper, and ²Hinsby Cadillo-
4 Quiroz

5
6 ¹Department of Earth, Atmospheric & Planetary Sciences and Department of Agronomy, Purdue
7 University, West Lafayette, IN, USA

8 ²School of Life Sciences, Arizona State University, Tempe, AZ, USA

9 ³Department of Global Ecology, Carnegie Institution of Science, Stanford, CA, USA

10 ⁴International Center for Tropical Botany, Florida International University, Miami, FL, USA

11

12 Correspondence to: qzhuang@purdue.edu

13

14

15

16

17

18

19

20

21

22

23

24

25

26

27 **Abstract:** Amazonian peatlands store a large amount of soil organic carbon (SOC) and its fate
28 under a future changing climate is unknown. Here we use a process-based peatland
29 biogeochemistry model to quantify the carbon accumulation for peatland and non-peatland
30 ecosystems in the Pastaza-Marañon foreland basin (PMFB) in the Peruvian Amazon from 12,000
31 years before present to 2100 AD. Model simulations indicate that warming accelerates peat SOC
32 loss while increasing precipitation accelerates peat SOC accumulation at millennial time scales.
33 The uncertain parameters and spatial variation of climate are significant sources of uncertainty to
34 modeled peat carbon accumulation. Under warmer and presumably wetter conditions over the
35 21st century, SOC accumulation rate in the PMFB slows down to 7.9 (4.3~12.2) g C m⁻² yr⁻¹
36 from the current rate of 16.1 (9.1~23.7) g C m⁻² yr⁻¹ and the region may turn into a carbon
37 source to the atmosphere at -53.3 (-66.8~-41.2) g C m⁻² yr⁻¹ (negative indicates source),
38 depending on the level of warming. Peatland ecosystems show a higher vulnerability than non-
39 peatland ecosystems as indicated by the ratio of their soil carbon density changes (ranging from
40 3.9 to 5.8). This is primarily due to larger peatlands carbon stocks and more dramatic responses
41 of their aerobic and anaerobic decompositions in comparison with non-peatland ecosystems
42 under future climate conditions. Peatland and non-peatland soils in the PMFB may lose up to 0.4
43 (0.32~0.52) Pg C by 2100 AD with the largest loss from palm swamp. The carbon-dense
44 Amazonian peatland may switch from a current carbon sink into a source in the 21st century.

45 **Significance Statement:** We use a process-based biogeochemistry model to quantify the carbon
46 accumulation for peatland ecosystems in the Pastaza-Marañon foreland basin in the Peruvian
47 Amazon from 12,000 years before present to 2100 AD. We find that warming accelerates peat
48 carbon loss while increasing precipitation slightly enhances peat carbon accumulation at
49 millennial time scales. With these impacts, our simulations suggest that the basin might lose up

50 to 0.4 Pg C by 2100 AD with the largest loss from palm swamp. If this loss rate is true for all
51 Amazonia peatlands, we project that these carbon-dense peatlands may switch from a current
52 carbon sink into a future source in this century.

53 \body

54 Tropical peatlands cover ~441,025 km² and store a large quantity (88.6 Pg C) of soil
55 organic carbon (SOC)¹⁻⁴. These ecosystems occupy ~11% of the global peatland area and
56 account for 15-19% of the total global peat SOC stock³. Tropical peatlands are mainly
57 distributed in Southeast Asia (~56%, 247,778 km²), and South and Central America (~23%,
58 107,486 km²)³. Recently, additional 145,500 km² of tropical peatlands containing 30.6 Pg C was
59 discovered in central Congo basin, Africa⁵. Given their significant C stocks, studying their
60 responses to past climatic trends and to the future climate change is of global importance⁶⁻⁸.

61 To date, most studies on the role of tropical peatlands in the global C cycle have focused
62 on Indonesian peatlands, which have been acting during the last decades as a considerable C
63 source to the atmosphere resulting from anthropogenic activities (e.g., land exploitation and
64 fires)⁹⁻¹¹. Few studies have focused on the Amazon basin, where peatlands remain nearly intact,
65 and have been a long-term C sink^{7,8,12,13}. The 120,000 km² Pastaza-Marañón foreland basin
66 (PMFB) located in Peru is the most extensive peatland complex in the Amazon basin, with up to
67 7.5 m thick peat deposits. The basal ages vary from 0.67 to 8.9 ka (1 ka=1000 cal years before
68 present) and the peat SOC accumulation rates range from 26 to 195 g C m⁻² yr^{-18,14}. It is a
69 subsiding foreland basin, resulting from the Cenozoic uplift of the Andes Mountains¹⁵⁻¹⁸ and
70 characterized by meandering (more than 100 m in a year¹⁹) and avulsions of rivers (abrupt
71 changes in the location of river stretches^{20, 21}). Waterlogged conditions due to high precipitation

72 and low lying topography provide a favorable environment for peat accumulation^{13,14}. By
73 measuring peat characteristics at several peatland sites within the basin, and using Landsat TM
74 images, Lähteenoja¹⁴ estimated a peatland area of 21,929 km² with SOC stock of 3.116 Pg
75 (0.837-9.461 Pg) for the central parts of the PMFB. Further, by incorporating multi-sensor
76 remote sensing and adding more peat core data, Draper¹³ mapped the distribution of peatland and
77 non-peatland ecosystems in the PMFB and estimated a peatland area of 35,600±2133 km² with
78 3.14 Pg C (0.44–8.15 Pg) stored in the vegetation and peat deposits of the whole basin.

79 According to most climate models, mean air temperature of South America has been
80 projected to increase by 1.8-5.1°C for the PMFB by the end of this century²²⁻²⁵. Annual
81 precipitation is projected to increase by up to 500 mm, although a large uncertainty exists²²⁻²⁵.
82 The strong dependence of C dynamics on climate suggests that warming in the 21st century may
83 turn the peatlands in the PMFB from a long-term C sink into a source^{6,7,9,26}. However, this
84 potential change has not been quantified or modelled in any way in previous studies. Nearly all
85 models focusing on the future C dynamics of the Amazon basin have been applied to non-
86 peatland ecosystems²⁷⁻³⁵ with the exception of²⁹.

87 Process-based models offer an alternative approach to quantifying peatland C dynamics
88 and providing insights for future projection³⁶⁻⁴⁰. Recently, a peatland terrestrial ecosystem model
89 (P-TEM) was developed for both peatland and non-peatland ecosystems by combining a
90 hydrology module (HM), a soil thermal module (STM), a methane dynamics module (MDM),
91 and a C and nitrogen dynamics module (CNDM)³⁹. P-TEM has been evaluated and used for
92 estimating C stocks across the Alaskan landscape since the last deglaciation^{39,40}. Here, we
93 parameterize and evaluate the P-TEM for tropical peatlands and model the C dynamics of the
94 peatlands in the Pastaza-Marañón foreland basin, Peruvian Amazonia (Fig 1) from 9 ka to 2014

95 AD. The model parameters were optimized using published peat, vegetation and remote-sensing
96 data for the PMFB from^{13,14} as well as other published sources (SI Appendix, Tables S1 and S2).
97 The model was then used to 1) quantify past C accumulation from 12 ka to 2014 AD in
98 peatlands, and 2) predict the future trends of C accumulation under different climate scenarios in
99 the 21st century in peatland and non-peatland ecosystems within the PMFB.

100 **Past C Accumulation**

101 The annual comparison between model simulation and measurements¹⁴ at a temporal
102 resolution of a year at the thickest and largest Amazonian peatland site (Aucayacu site) reveals
103 that our model captures the historic peat SOC accumulation rates (Fig 2a) and the peat depth
104 profile (Fig 2b) for most simulation periods, but overestimates the rates between 8 and 6 ka.
105 Simulated total depth reaches 8 m (ranging 6-12 m), slightly higher than the measured 7.5 m¹⁴.
106 The correlation between simulations and measurements using 500-year bins at multiple sites with
107 different vegetation types indicates that the model well estimates SOC accumulation trajectories
108 at millennial time scales (SI Appendix, Fig S1). The model underestimates the rates between 3
109 and 2 ka at San Jorge, and between 0.5 and 0 ka at Charo. Although the model underestimates
110 the rates between 2 and 1.5 ka at Rinon (an open peatland site), the starting basal age for the
111 regional transient simulation for the open peatlands is at 1.6 ka (see SI Appendix, Table S3). As
112 indicated by the mean basal age applied in the model, pole forest (PF) has a longer SOC
113 accumulation period than palm swamp (PS) and open peatland (OP) in general (SI Appendix, Fig
114 S1), with peat initiation of PF around 2000 years ahead of PS and OP peat initiation.

115 Our simulation suggests there were strong relationships between peat C dynamics and
116 climatic change. Temporally, temperature and photosynthetically active radiation (PAR) rose
117 slightly over the whole period (SI Appendix, Figs S2a and e), whereas annual precipitation

118 decreased before 4 ka and subsequently increased (SI Appendix, Fig S2c). Under increasingly
119 warmer and drier conditions before 4 ka, the historic SOC accumulation rate declined at the
120 Aucayacu site (Fig 2a). It started to increase concurrently with the wetter conditions after 3.5 ka
121 (close to 4 ka), despite continuing warming. Overall, the historic SOC accumulation rates of the
122 Aucayacu peat core followed the historic pattern of the precipitation change (SI Appendix, Fig
123 S2c). This suggests that higher rainfall might have accelerated while warming and drought might
124 have decelerated peat SOC accumulation at millennial time scales.

125 Spatially, this relationship between the peat SOC accumulation and climate was indicated
126 by the patterns of the peat SOC densities distribution and mean historic temperature and
127 precipitation within the PMFB (Figs. 3 and S3). We find that the highest peat SOC density
128 region fell in the northeast with the highest precipitation and relatively low temperature. The
129 secondary highest peat SOC region was located in the northwest with moderate precipitation but
130 the lowest temperature. The lowest peat SOC zone fell within the southwest where the lowest
131 precipitation and highest temperature coincided. This, again, suggests that higher precipitation
132 increased whereas higher temperature reduced peat SOC accumulation at regional scales.

133 The climatic effects on the long-term peat SOC accumulation in the PMFB can be
134 explained by our simulated C fluxes and hydrological factors. Peat accumulated SOC where the
135 rate of soil C input was higher than the decomposition⁴¹. Soil C input from litters was largely
136 controlled by and was proportional to plant net primary productivity (NPP). Soil decomposition
137 was modeled as heterotrophic respiration (R_H)⁴² (see SI Appendix). Increasing temperature and
138 PAR stimulated the plant C uptake by increasing NPP. However, warming might have created
139 favorable conditions for microbial decomposition⁴³. Warming also increased the
140 evapotranspiration, decreasing water table, thereby reducing anaerobic respiration and increasing

141 aerobic respiration⁴⁴. Increasing precipitation had a positive effect on NPP. It also lifted the
142 water table and decreased R_H . This, in turn, enhanced peat C accumulation. In our previous study
143 for the northern (Alaskan) peatlands⁴⁰, under the warmer conditions, the stimulation of NPP
144 exceeded the stimulation of R_H , thereby increasing SOC accumulation in northern peatlands
145 during the Holocene Thermal Maximum (HTM). Similarly, we find that R_H within 1 m depth
146 followed the increasing trend of temperature with a decrease at 4 ka when precipitation increased
147 (SI Appendix, Fig S4b). This suggests that warmer condition in the PMFB enhances R_H while
148 wetter condition decreases R_H . The volumetric soil moisture (VSM) (SI Appendix, Fig S4c) and
149 water table (SI Appendix, Fig S4d) started decreasing at 8 ka as precipitation became lower. At
150 the same time, R_H kept increasing under such drier condition. Interestingly, when climate
151 became wetter at ~4 ka (SI Appendix, Fig S2c), the VSM abruptly increased. The water table
152 also stopped dropping and showed an increase pattern. Meanwhile, R_H started decreasing. This
153 again suggests that higher precipitation may decrease R_H and thus slows the peat SOC
154 decomposition by increasing the soil moisture and raising the water table. As warming
155 continued, the increase of VSM and water table were slight, presumably due to the enhanced
156 evapotranspiration.

157 To examine how temperature and precipitation have impacted NPP in this region. The
158 attributions of these two key drivers and soil water content to NPP for both historical periods and
159 the 21st century were analyzed with the Analysis of Variance table (ANOVA) and the F-test of
160 the multi-variate linear regression between annual mean NPP and climate variables. For the
161 historical simulation at Aucayacu site, we find that, although higher precipitation and higher
162 temperature increase the NPP (SI Appendix, Table S4 and S5), those two factors have limited
163 effects presumably because the temperature exceeds the optimum temperature for photosynthesis

164 and soil water content is already suitable for plant growth. The variable with the highest
165 importance is VSM, indicating the hydrological condition plays the most important role in
166 determining the NPP. Such hydrological condition is modeled by various factors including the
167 temperature, precipitation, solar radiation and others such as soil porosity, soil layers
168 characteristics that are described in our previous hydrological modeling studies^{39,60}.

169 Our historical simulations at Aucayacu and in the PMFB suggest that NPP was consistent
170 with the temporal patterns of precipitation and VSM (SI Appendix, Fig S2c). The spatial
171 correlations between NPP, vegetation C density, and mean historic precipitation were detected
172 (SI Appendix, Figs S3b and S5) when observing each peatland vegetation type separately. Pixels
173 with higher vegetation C density and NPP fell within the northeastern wetter region while lower
174 vegetation C density and NPP pixels were in the southwestern drier region. Our explanation is
175 that during the historical period, the point where NPP will no longer positively respond to the
176 increasing precipitation and VSM has not been reached. Still, the wetter condition stimulates the
177 SOC accumulation by increasing NPP and decreasing R_H . However, for the simulations under
178 three future climate scenarios (SI Appendix, Table S6), we find that the F values of the
179 precipitation started decreasing as the precipitation continued to be higher from RCP 2.6 to 8.5.
180 This suggests that precipitation becomes less and less important in the future for NPP as it
181 increases, based on the existing suitable hydrological condition.

182 In our model, GPP is a function of atmospheric CO₂ concentrations in addition to
183 physical variables (see SI Appendix). The CO₂ effects is modeled with a Michaelis-Menten
184 equation considering CO₂ concentrations inside leaves which is assumed to be directly
185 proportional to atmospheric CO₂ concentrations when stomata are fully open. When moisture is a
186 limiting factor, the limitation on CO₂ assimilation is modeled by the modifying the conductance

187 of leaves to CO₂ diffusion. The moisture availability is expressed as the ratio of actual
188 evapotranspiration (EET) to potential evapotranspiration (PET). The relationship between CO₂
189 concentration inside stomatal cavities (C_i) and in the atmosphere (C_a) is proportional to relative
190 moisture availability:

$$191 \quad G_V = 0.1 + \left(\frac{0.9EET}{PET}\right)$$
$$192 \quad C_i = G_V C_a$$

193 where G_V is a unitless multiplier that accounts for changes in leaf conductivity to CO₂
194 resulting from changes in moisture availability. When there is sufficient water in soils, EET will
195 not be limited by water, which will reach its maximum value, G_V is close to 1. This suggests that
196 inside of leaves CO₂ will be close to ambient CO₂. When the ecosystem has sufficient
197 precipitation, GPP and NPP will not respond to increasing precipitation.

198 At northern high latitudes, in addition to CO₂ fertilization effects, warming also enhances
199 photosynthesis, stimulating plant productivity (NPP) and thus increasing SOC accumulation^{42,5-}
200 ⁴⁷. In contrast, warming in the tropical regions generally led to temperatures above the optimum
201 level for photosynthesis^{48,49}, which is also suggested by the ANOVA analysis (SI Appendix,
202 Tables S4 and S6) as increasing temperature in the future has less and less positive effects on
203 NPP. Increasing temperature accelerates R_H , however, at the same time. The less sensitivity of
204 NPP versus R_H to warming might ultimately result in the SOC loss in the PMFB under warmer
205 conditions.

206 However, hydrology, NPP, and SOC accumulation can also be controlled by autogenic
207 processes of peatlands such as transition from minerotrophic to ombrotrophic conditions⁸. This
208 transition is largely induced by the form and thickness of the peat deposit and less affected by
209 prevailing climatic conditions – as long as the rainfall is sufficient to sustain an rain-fed bog.
210 Interestingly, in the Aucayacu peatland, the transition from minerotrophic to ombrotrophic

211 conditions occurred around 4 ka (3.5 ka) – exactly when precipitation started to increase. It
212 might have been a coincidence, but it is also possible that the increased precipitation enabled the
213 appearance of ombrotrophic bogs. If this is the case, a change in the precipitation did not affect
214 the NPP directly but indirectly by inducing a change in the peatland type. Since our model
215 cannot simulate the paleo-ecological change including the shifts between different peatland
216 ecosystem types through time, our results may only partly explain the observed patterns, with
217 much information still relying on paleo-ecological studies^{14,50-52}. The relationship between NPP
218 and precipitation for peatland ecosystems in the region should be further studied in the future.

219 Another key control of the current distribution of peat depths and SOC densities within
220 the PMFB is the active lateral migration of rivers¹⁴. The current distribution of peat SOC
221 densities can be explained by both climatic and geological factors. The Amazon river networks
222 can be affected under future climate conditions, which will affect peatland dynamics (e.g.,
223 formation and area change). Further, our model did not differentiate the minerotrophic vs.
224 ombrotrophic conditions for the peatland ecosystems, which will introduce biases. Incorporating
225 these dynamics into future analysis shall improve our predictions of SOC for this region.

226 Our uncertainty analysis suggests that the uncertainty of the simulated past C
227 accumulation rates was mainly due to parameters, spatial variations of climate variables (SI
228 Appendix, Fig S3), and the uncertain peat basal ages (SI Appendix, Table S3). Specifically,
229 using the mean peat basal age by averaging the basal ages of peat samples for each peatland type
230 is a top uncertainty source. The variation of peat characteristics (e.g., bulk density, C content,
231 peat depth) and limited number of samples are also sources of the uncertainty.

232

233 **Current C stocks**

234 Overall, model simulations of current peatland C stocks are comparable to the field
235 measurements of¹³. Specifically, PF has the SOC density of 1900 Mg C ha⁻¹, consistent with the
236 field measurements (800-2200 Mg C ha⁻¹, SI Appendix, Fig S6a)¹³. PS has the next highest SOC
237 density (1100 Mg C ha⁻¹), which is sufficiently within the measured range of 300-1390
238 Mg C ha⁻¹. The SOC density of OP is 535 Mg C ha⁻¹, also within the measured 392-1492
239 Mg C ha⁻¹. The high SOC density of PF corresponds to the longer SOC accumulation period
240 compared to the other types (SI Appendix, Figs S1a, b and S6a). Our simulations are even closer
241 to the field measurements¹³ when vegetation C density was examined (SI Appendix, Fig S6b).
242 The simulated lowest vegetation C density was in PF (86 Mg C ha⁻¹) and within the measured
243 range of 80-100 Mg C ha⁻¹.

244 The model estimates a total SOC of 3.922 (2.208~5.777) Pg in the PMFB including
245 3.519 (1.833~5.344) Pg in the peatland soils, which is higher than the measured total peat SOC,
246 2.844 Pg (SI Appendix, Table S7)¹³. The simulated vegetation C stock of 1.104 (1.097~1.137)
247 Pg with 0.34 (0.338~0.369) Pg on the PMFB peatlands is also higher than the measured value
248 (0.293 Pg C)¹³. Our model may overestimate the soil and vegetation C stocks. The uncertainty of
249 the simulated C stocks is mainly due to the spatial variations of the interpolated mean
250 temperature (25-29°C) and precipitation (2200-2900 mm) (SI Appendix, Fig S3).

251

252

253 **Future Projection**

254 Under the RCP 2.6 scenario (see below for climate description), the SOC accumulation
255 rate in all ecosystem types within the PMFB decreases from 16 (9~24) to 7.9 (4.3~12.2)
256 g C m⁻² yr⁻¹ and the SOC accumulation rate in the peatlands dramatically decreases from 56

257 (29~85) to 23 (15~32) g C m⁻² yr⁻¹ (Table 1). The PS exhibits the biggest drop from 65 to 26
258 g C m⁻² yr⁻¹. Spatially, the majority of pixels within the PMFB have positive SOC
259 accumulation and vegetation C change, but some areas with PS have SOC loss (Figs 4 and S7).
260 Overall, 0.067 (0.037~0.108) Pg SOC, including 0.06 (0.03~0.1) Pg SOC in the peatlands, will
261 be accumulated in the PMFB by the end of the 21st century under moderately warmer and wetter
262 conditions of this climate scenario (Table 1). There will be 0.0148 Pg C accumulated in
263 vegetation, including 0.0048 Pg C in peatland vegetation.

264 Under the RCP 8.5 scenario, the SOC accumulation rate declines from 16 to -53 (-67~-
265 41) C m⁻² yr⁻¹ and the rate in peatlands declines from 56 to -123 (-152~-91) C m⁻² yr⁻¹ (Table
266 1). Again, the highest decline of the rate is for PS, from 65 to -135 g C m⁻² yr⁻¹. The pixels with
267 SOC and vegetation C loss dominate the region (Figs 4 and S7). Under this climate scenario, the
268 PMFB will act as a C source of 0.413 (0.319~0.518) Pg C by 2100 AD (Table 1). Peatlands will
269 lose 0.31 (0.23~0.38) Pg C compared with 0.1 Pg C loss from non-peatland ecosystems.
270 Vegetation will lose 0.07 Pg C, including 0.02 Pg C from peatland vegetation. Among all
271 peatland ecosystem types, PS could be severely affected by the climate due to its large area
272 within the PMFB and within the whole Amazon Basin⁵³. It must be taken into account that the
273 tendency of the model to overestimate the current soil and vegetation C stocks in the PMFB (SI
274 Appendix, Table S7) might affect these values to some extent.

275 Under the intermediate RCP 4.5 scenario, the SOC accumulation rate declines from 16 to
276 -19 C m⁻² yr⁻¹ and the SOC accumulation rate in peatlands declines from 56 to -45 C m⁻² yr⁻¹.
277 Peatlands will lose 0.12 Pg C compared with 0.034 Pg C from non-peatlands.

278 Three extra simulations were conducted as sensitivity tests to examine the effects of
279 potential drier climate in the PFMB on SOC accumulation rates. We assume that (1) The future

280 precipitation will decrease 5% in our study region over the century, but holding air temperature
281 change as in the original RCP 2.6; (2) The future precipitation will decrease 10% but holding air
282 temperature change as in RCP 4.5; (3) The future precipitation will decrease 15% but holding air
283 temperature change as in RCP 8.5. The precipitation was manually decreased at monthly step for
284 each grid cell from 2014 to 2100 AD to achieve the certain percentage total reduction at the end
285 of 2100 AD. Our simulations show that the C accumulations are +0.027 (0.02~0.068), -0.203 (-
286 0.349~-0.167), and -0.594 (-0.731~-0.51) Pg C under the three sensitivity simulations (Table 1).
287 These extra simulations suggest that the slightly drier condition will decrease but will not have
288 significant effects on the C accumulation in this region.

289 The modeled current C stocks agree with the field observations at the Aucayacu site,
290 which is a PF site. However, instead of PF, PS is the dominant peatland type in the study area
291 and its SOC accumulation rates at Charo site is underestimated. Thus, using PS as representative
292 peatland types for regional simulations under future climate scenarios may evolve uncertainty
293 due to the underestimation during the parameterization.

294 In conclusion, the warming in the 21st century may weaken the C sink function of the
295 Amazonian peatlands in the PMFB or may entirely switch them from a long-term carbon sink
296 into a source, depending on the severity of the warming. The same has also been predicted for
297 the Amazonian rainforest in general²⁶. The vegetation and SOC density changes (future total C
298 stock changes (Table 1) divided by the corresponding areas (SI Appendix, Table S7) of peatlands
299 and non-peatland) were calculated to compare with other studies. Our model estimation of
300 vegetation C change for the non-peatland (mainly flooded forest) ecosystem in the 21st century
301 (+0.23~-1.17 kg C m⁻²) is well within the range of other studies (+0.6~-1.2 kg C m⁻²) on the
302 future vegetation C change from Amazonian rainforest dieback^{26,54} (SI Appendix, Table S8). Our

303 estimation of SOC change for the non-peatland ecosystem in the 21st century (+0.18~-3.35 kg C
304 m⁻²) is also comparable to -3.88 kg C m⁻² from other studies^{26,54}. Further, we find that the ratio of
305 SOC density changes for peatlands and non-peatland ecosystems in the next 100 years ranges
306 from 3.9 to 5.8 (SI Appendix Table S8). This indicates that future warming is likely to affect the
307 Amazonian peatlands more dramatically than non-peatland ecosystems, although the total area of
308 peatlands is much smaller than that of non-peatland ecosystems within the PMFB (31000 km² vs.
309 47000 km², see SI Appendix for areas). The high vulnerability of peatland ecosystems to future
310 climate is presumably due to its large amount of existing SOC stock (3.5 Pg C) compared with
311 non-peatland ecosystems (0.4 Pg C). Another possible reason is that, in addition to the non-linear
312 function defining the volumetric soil moisture (VSM) effect on heterotrophic respiration within
313 the unsaturated zone, there is also a linear relation between water-table depth and aerobic
314 respiration in the model (see SI Appendix for decomposition calculation). Future warming
315 increases evapotranspiration that subsequently decreases VSM and increases aerobic respiration
316 for both peatland and non-peatland ecosystems. For peatlands, it further lowers the water table
317 which, in turn, increases aerobic respiration. This suggests that peatland ecosystems may suffer
318 larger SOC decomposition under the changing climate and may help explain its vulnerability. In
319 addition, increased land use change, expansion of commercial agriculture, transport
320 infrastructure, and hydropower development form a threat to the persistence of the considerable
321 C stock⁵⁵. The most carbon dense ecosystems of the whole Amazon basin may turn into C
322 sources.

323

324

325 **Methods**

326 **The peatland biogeochemistry model and its parameterization**

327 In P-TEM, peat SOC accumulation is determined by the difference between NPP and
328 aerobic and anaerobic respiration R_H at a monthly step (see^{1,2} and SI Appendix for model
329 details). Parameters in P-TEM were first optimized with data of annual C fluxes and stocks in the
330 Amazon basin taken from literature (SI Appendix, Table S1) in order to obtain the prior
331 distribution of the parameter space for peatland ecosystems (see SI Appendix for details)
332 Specifically, site-level measurements of tree biomass from Amazonian peatlands¹³ were used to
333 compare with model simulations to optimize parameters. Due to the lack of NPP measurements,
334 NPP values used in the model are field measurements from neighboring white-sand forests (for
335 pole forest peatlands) and seasonally flooded forests (for palm swamp peatlands and flooded
336 forests) (SI Appendix, Table S1). Second, a Bayesian approach was used to optimize parameters
337 (SI Appendix, Table S2) with Monte Carlo ensemble simulations driven by the extracted paleo
338 climate data (SI Appendix, Fig S2) at five peatland sites (SI Appendix, Fig S1).

339 The distribution of vegetation types was taken from¹³ at a resolution of 90 m×90 m and
340 was resized to 1.69 km×1.69 km. Vegetation types in the region include three peat-forming
341 vegetation types (pole forest (PF), palm swamp (PS), and open peatland (OP, peatland lacking
342 closed canopy)) and a non-peat forming type (flooded forest (FF)) (Fig 1). OP was assumed to
343 have minimal NPP and vegetation biomass during the simulation¹³.

344

345 **Climate Data**

346 The climate forcing data for historic simulations include temperature, precipitation,
347 photosynthetically active radiation (PAR), vapor pressure at a monthly step and CO₂ at an annual

348 step from 12 ka to 1990 AD, simulated by CCSM3 (TraCE-21ka) at a spatial resolution of
349 $3.75^\circ \times 3.75^\circ$. Climate forcing data for modern simulations is from Climate Research Unit
350 (CRU2.0) at a monthly step from 1990 to 2014 AD at a resolution of $0.5^\circ \times 0.5^\circ$. For future
351 simulations, we applied the Representative Concentration Pathway (RCP) 2.6 (mean annual
352 temperature in the PMFB has the smallest increase (by $\sim 0.5^\circ\text{C}$), mean annual precipitation
353 increases by ~ 260 mm, and CO_2 increases by ~ 80 ppm at 2050 AD and decreases by ~ 30 ppm at
354 2100 AD), RCP 4.5 (by $\sim 1.5^\circ\text{C}$, ~ 290 mm, and CO_2 increases by ~ 150 ppm at 2100 AD), and
355 RCP 8.5 (by $\sim 2.7^\circ\text{C}$, ~ 350 mm, and ~ 600 ppm at 2100 AD) at a monthly step from 2014 to 2100
356 AD at $0.5^\circ \times 0.5^\circ$ as possible future climate scenarios. The CRU data together with the modern
357 digital elevation data at $1.69\text{ km} \times 1.69\text{ km}$ were input into interpolation software ANUSPLIN4.4.
358 We then downscaled the paleo-climate data (TraCE-21ka, $3.75^\circ \times 3.75^\circ$) and the RCP data
359 ($0.5^\circ \times 0.5^\circ$) based on the spatial variations of the interpolated CRU data ($1.69\text{ km} \times 1.69\text{ km}$) by
360 assuming that the spatial variations of CRU to be the same as that of paleo and RCP data.

361

362 **Model Application and Uncertainty Analysis**

363 A 500-year run was conducted for each peatland ecosystem type ahead of the basal age
364 using parameters of non peat-forming FF to determine the initial SOC within the upper 1 m
365 mineral soil underlying the peat deposit. The model was first run from 12 ka to 2014 AD for
366 validation at five peatland sites (SI Appendix, Fig S1). The simulated SOC accumulation rates of
367 PS, OP, and PF were firstly compared with measured SOC accumulation rates^{7,14} annually in
368 500-year bins. Second, we applied the model to a regional simulation with interpolated monthly
369 paleo-climate data for the PMFB. We averaged all the measured basal ages of each peatland type
370 to determine the mean basal age of each peatland ecosystem type (SI Appendix, Table S3). Basal

371 ages at sites where they have not been measured were calculated using mean SOC accumulation
372 rates, bulk density, peat depth and C content of each peatland ecosystem type derived from^{7,13,14},
373 following the equations in⁷. We conducted the simulation from 1 ka to 2014 AD for FF. Finally,
374 we conducted the simulations for future projection using the interpolated RCP 2.6, RCP 4.5 and
375 RCP 8.5. A series of simulations were conducted to examine the effects of the spatial variation of
376 climate and the posterior distribution of the parameter on the estimated C accumulation rates and
377 stocks (see SI Appendix for details). Twenty sets of parameters were randomly drawn from the
378 posterior parameter space. All pixels in the study area were then assigned with the same climate
379 forcing data which were random combinations between temperature and precipitation.

380 Reference

- 381 1. Page, S. E. et al. The amount of carbon released from peat and forest fires in Indonesia during 1997. *Nature*, 420(6911), 61-65 (2002).
- 382 2. Page, S. E. et al. A record of Late Pleistocene and Holocene carbon accumulation and climate change from an equatorial peat bog (Kalimantan,
383 Indonesia): implications for past, present and future carbon dynamics. *Journal of Quaternary Science*, 19(7), 625-635 (2004).
- 384 3. Page, S. E. et al. Global and regional importance of the tropical peatland carbon pool. *Global Change Biology*, 17(2), 798-818 (2011).
- 385 4. Rieley, J. O. et al. Tropical peatlands: carbon stores, carbon gas emissions and contribution to climate change processes. *Peatlands and Climate Change*.
386 International Peat Society, Vapaudenkatu, 12(40100), 148-182 (2008).
- 387 5. Dargie, G. C. et al. Age, extent and carbon storage of the central Congo Basin peatland complex. *Nature*, 542, 86-90 (2017).
- 388 6. Guzmán Castillo, W. Valor económico del manejo sostenible de los ecosistemas de aguaje (*Mauritia flexuosa*). In *International Congress on*
389 *Development, Environment and Natural Resources: Multi-level and Multi-scale Sustainability* (Vol. 3, pp. 1513-1521) (2007).
- 390 7. Låhteenoja, O. et al. Amazonian peatlands: an ignored C sink and potential source. *Global Change Biology*, 15(9), 2311-2320 (2009a).
- 391 8. Låhteenoja, O., & Page, S. High diversity of tropical peatland ecosystem types in the Pastaza-Marañón basin, Peruvian Amazonia. *Journal of*
392 *Geophysical Research: Biogeosciences*, 116(G2) (2011).
- 393 9. Rieley, J. O., & Page, S. E. Wise use of tropical peatlands: focus on Southeast Asia: synthesis of results and conclusions of the UK Darwin Initiative and
394 the EU INCO EUTROP, STRAPEAT and RESTORPEAT Partnerships together with proposals for implementing wise use of tropical peatlands.
395 ALTERRA–Wageningen University and Research Centre and the EU INCO–STRAPEAT and RESTORPEAT Partnerships, 266 (2005).
- 396 10. Maltby, E. et al. Carbon dynamics in peatlands and other wetland soils regional and global perspectives. *Chemosphere*, 27(6), 999-1023 (1993).
- 397 11. Miettinen, J. et al. Two decades of destruction in Southeast Asia's peat swamp forests. *Frontiers in Ecology and the Environment*, 10(3), 124-128
398 (2011).
- 399 12. Låhteenoja, O. et al. Amazonian floodplains harbour minerotrophic and ombrotrophic peatlands. *Catena*, 79(2), 140-145 (2009b).
- 400 13. Draper, F. C. et al. The distribution and amount of carbon in the largest peatland complex in Amazonia. *Environmental Research Letters*, 9(12), 124017
401 (2014).
- 402 14. Låhteenoja, O. et al. The large Amazonian peatland carbon sink in the subsiding Pastaza-Marañón foreland basin, Peru. *Global Change Biology*, 18(1),
403 164-178 (2012).
- 404 15. Dumont, J. F. et al. Wetland and upland forest ecosystems in Peruvian Amazonia: plant species diversity in the light of some geological and botanical
405 evidence. *Forest Ecology and Management*, 33, 125-139 (1990).
- 406 16. Dumont, J. F. et al. Morphostructural provinces and neotectonics in the Amazonian lowlands of Peru. *Journal of South American Earth Sciences*, 4(4),
407 373-381 (1991).
- 408 17. Räsänen, M. E. et al. Evolution of the western Amazon lowland relief: impact of Andean foreland dynamics. *Terra Nova*, 2(4), 320-332 (1990).
- 409 18. Räsänen, M. E. et al. Recent and ancient fluvial deposition systems in the Amazonian foreland basin, Peru. *Geological Magazine*, 129(03), 293-306
410 (1992).

- 411 19. Kalliola et al. Upper Amazon channel migration: implications for vegetation perturbation and succession using bitemporal Landsat images.
412 *Naturwissenschaften*, 79, 75–79 (1992).
- 413 20. Smith ND et al. Anatomy of an avulsion. *Sedimentology*, 36,1–23 (1989).
- 414 21. Neller RJ et al. On the formation of blocked valley lakes by channel avulsion in Upper Amazon foreland basins. *Zeitschrift für Geomorphologie*, 36,
415 401–411 (1992)
- 416 22. Guimberteau, M. et al. Future changes in precipitation and impacts on extreme streamflow over Amazonian sub-basins. *Environmental Research*
417 *Letters*, 8(1), 014035 (2013).
- 418 23. Marengo, J. A. et al. Development of regional future climate change scenarios in South America using the Eta CPTec/HadCM3 climate change
419 projections: Climatology and regional analyses for the Amazon, São Francisco and the Parana River Basins. *Clim Dyn* 38:1829–1848.
420 doi:10.1007/s00382-011-1155-5 (2012).
- 421 24. Sánchez, E. et al. Regional climate modelling in CLARIS-LPB: a concerted approach towards twenty-first century projections of regional temperature
422 and precipitation over South America. *Climate Dynamics*, 45(7-8), 2193-2212 (2015).
- 423 25. Zulkafli, Z. et al. Projected increases in the annual flood pulse of the Western Amazon. *Environmental Research Letters*, 11(1), 014013 (2016).
- 424 26. Cox, P. M. et al. Amazonian forest dieback under climate-carbon cycle projections for the 21st century. *Theoretical and applied climatology*, 78(1),
425 137-156 (2004).
- 426 27. Tian, H. et al. Effect of interannual climate variability on carbon storage in Amazonian ecosystems. *Nature*, 396(6712), 664-667 (1998).
- 427 28. Tian, H. et al. Climatic and biotic controls on annual carbon storage in Amazonian ecosystems. *Global Ecology and Biogeography*, 9(4), 315-335
428 (2000).
- 429 29. Li, W. et al. Future precipitation changes and their implications for tropical peatlands, *Geophys. Res. Lett.*, 34, L01403, doi:10.1029/2006GL028364
430 (2007).
- 431 30. Cleveland, C. C. et al. A comparison of plot-based satellite and Earth system model estimates of tropical forest net primary production. *Global*
432 *Biogeochem. Cycles*, 29, 626–644. doi: 10.1002/2014GB005022 (2015).
- 433 31. Rowland, L. et al. Modelling climate change responses in tropical forests: similar productivity estimates across five models, but different mechanisms
434 and responses, *Geosci. Model Dev.*, 8, 1097-1110, doi:10.5194/gmd-8-1097-2015 (2015).
- 435 32. Restrepo-Coupe, N. et al. Do dynamic global vegetation models capture the seasonality of carbon fluxes in the Amazon basin? A data-model
436 intercomparison. *Glob Change Biol*, 23: 191–208. doi:10.1111/gcb.13442 (2017).
- 437 33. Powell, T. L. et al. Confronting model predictions of carbon fluxes with measurements of Amazon forests subjected to experimental drought. *New*
438 *Phytologist*, 200: 350–365. doi: 10.1111/nph.12390 (2013).
- 439 34. Delbart, N. et al. Mortality as a key driver of the spatial distribution of aboveground biomass in Amazonian forest: results from a vegetation
440 model. *BG*, 7(10), 3027-3039 (2010).
- 441 35. Schulman et al. Parameters for global ecosystem models. *Nature*, 399(6736), 535-536 (1999).
- 442 36. Frolking, S. et al. A new model of Holocene peatland net primary production, decomposition, water balance, and peat accumulation. *Earth System*
443 *Dynamics*, 1(1), 1-21 (2010).
- 444 37. Spahni, R. et al. Transient simulations of the carbon and nitrogen dynamics in northern peatlands: from the Last Glacial Maximum to the 21st
445 century. *Climate of the Past*, 9(3), 1287-1308 (2013).
- 446 38. Kleinen, T. et al. A dynamic model of wetland extent and peat accumulation: results for the Holocene. *Biogeosciences*, 9(1), 235-248 (2012).
- 447 39. Wang, S. et al. Quantifying peat carbon accumulation in Alaska using a process-based biogeochemistry model. *JGR: Biogeosciences*, 121(8), 2172-
448 2185 (2016a).
- 449 40. Wang, S. et al. Quantifying soil carbon accumulation in Alaskan terrestrial ecosystems during the last 15000 years. *Biogeosciences*, 13(22), 6305
450 (2016b).
- 451 41. Loisel, J. et al. Global-scale pattern of peatland Sphagnum growth driven by photosynthetically active radiation and growing season
452 length. *Biogeosciences*, 9(7), 2737-2746, (2012).
- 453 42. Yu, Z. et al. Sensitivity of northern peatland carbon dynamics to Holocene climate change. *Carbon cycling in northern peatlands*, 55-69, (2009).
- 454 43. Nobrega, S. et al. Deeper snow enhances winter respiration from both plant-associated and bulk soil carbon pools in birch hummock
455 tundra. *Ecosystems*, 10(3), 419-431 (2007).
- 456 44. Hobbie, S. E. et al. Controls over carbon storage and turnover in high latitude soils. *Global Change Biology*, 6(51), 196-210 (2000).
- 457 45. Davidson, E. A., & Janssens, I. A. Temperature sensitivity of soil carbon decomposition and feedbacks to climate change. *Nature*, 440(7081), 165-173
458 (2006).
- 459 46. Jones, M. C., & Yu, Z. Rapid deglacial and early Holocene expansion of peatlands in Alaska. *Proceedings of the National Academy of*
460 *Sciences*, 107(16), 7347-7352 (2010).
- 461 47. Yu, Z. et al. Global peatland dynamics since the Last Glacial Maximum. *Geophysical Research Letters*, 37(13) (2010).
- 462 48. Jenkinson, D. S. et al. Model Estimates of CO₂ Emissions from Soil in Response to Global Warming. *Nature*, 351(6324), 304 (1991).
- 463 49. Cox, P. M. et al. Modelling vegetation and the carbon cycle as interactive elements of the climate system. *International Geophysics*, 83, 259-279
464 (2002).

465 50. Swindles, G. et al. Ecology of testate amoebae in an Amazonian peatland and development of a transfer function for palaeohydrological reconstruction.
466 *Mic. Ecol.* 68, 284–298 (2014).

467 51. Kelly, T. J. et al. The vegetation history of an Amazonian domed peatland. *Palaeogeography, Palaeoclimatology, Palaeoecology*, 468, 129-141 (2017).

468 52. Roucoux, K. H. et al. Vegetation development in an Amazonian peatland. *Palaeogeography, Palaeoclimatology, Palaeoecology*, 374, 242-255 (2013).

469 53. Ruokolainen, K. et al. On Amazonian peatlands. *International Mire Conservation Group Newsletter*, 4, 8-10 (2001).

470 54. Rammig, A. et al. Estimating the risk of Amazonian forest dieback. *New Phytologist*, 187: 694-706. doi:10.1111/j.1469-8137.2010.03318.x (2010).

471 55. Roucoux, K.H. et al. Threats to intact tropical peatlands and opportunities for their conservation. *Conservation Biology*. Accepted Author Manuscript.
472 doi:10.1111/cobi.12925 (2017).

473 Figure Legends

474 Fig 1. Distribution of peat- and non peat-forming vegetation in the PMFB at the resolution of 90 m×90 m¹³. The map was resized
475 to 1.69 km×1.69 km. Colors represent vegetation types: open peatland (pink), palm swamp (red), pole forest (green), and flooded
476 forest (dark blue). Yellow represents open water and light blue represents other. See the Figs 1 and 4 of¹³ for original map.

477 Fig 2. Comparison between simulation and measurement¹⁴ at the Aucayacu site. (a) SOC accumulation rates and (b) peat depth.
478 Shaded areas represent the range due to uncertainties from the posterior distributions of the parameters after the parameterization.
479 0 cm at ~9 ka indicates no peat accumulation.

480 Fig 3. Current (2014 AD) SOC density of flooded forest, palm swamp, open peatland, pole forest and their combination in the
481 PMFB.

482 Fig 4. Changes of SOC density from 2014 to 2100 AD under RCP 2.6 and RCP 8.5 future climate scenarios in flooded forest,
483 palm swamp, pole forest, open peatland, and their combination in the PMFB. Blue and green represent the SOC accumulation.
484 Yellow and red represent the SOC loss.

485 Acknowledgements: Author Contributions: Q.Z. and S.W. conceived and designed the study; S.W., O.L., F.D. processed the
486 data; S.W. and Q.Z. drafted the paper; All authors contributed to the revision of the paper.

487 Additional Information: Competing financial interest: The authors declare no competing financial interests.