

This is a repository copy of Accounting global grey water footprint from both consumption and production perspectives.

White Rose Research Online URL for this paper: http://eprints.whiterose.ac.uk/144525/

Version: Accepted Version

Article:

Zhao, X, Liao, X, Chen, B et al. (3 more authors) (2019) Accounting global grey water footprint from both consumption and production perspectives. Journal of Cleaner Production, 225. pp. 963-971. ISSN 0959-6526

https://doi.org/10.1016/j.jclepro.2019.04.037

© 2019, Elsevier. This manuscript version is made available under the CC-BY-NC-ND 4.0 license http://creativecommons.org/licenses/by-nc-nd/4.0/.

Reuse

This article is distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs (CC BY-NC-ND) licence. This licence only allows you to download this work and share it with others as long as you credit the authors, but you can't change the article in any way or use it commercially. More information and the full terms of the licence here: https://creativecommons.org/licenses/

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk https://eprints.whiterose.ac.uk/

Accounting global grey water footprint from both consumption

and production perspectives

Xu Zhao^{a,*,1}, Xiawei Liao^{b,1}, Bin Chen^{c, **}, Martin R. Tillotson^d, Wei Guo^e, Yiping Li^a

^aKey Laboratory of Integrated Regulation and Resource Development on Shallow Lake s, Ministry of Education, College of Environment, Hohai University, Nanjing 210098, China

^bEnvironmental Change Institute, University of Oxford, OX1 3QY, Oxford, UK

^cState Key Joint Laboratory of Environment Simulation and Pollution Control, School

of Environment, Beijing Normal University, Beijing 100875, China

^dwater@leeds, School of Civil Engineering, University of Leeds, Leeds LS2 9JT, United Kingdom

^eSchool of Environmental Science and Engineering, Southern University of Science and Technology, Shenzhen 518055, China

*Corresponding author.

**Corresponding author.

E-mail addresses: chenb@bnu.edu.cn (B. Chen); xuzhao@hhu.edu.cn (Xu Zhao)

¹These authors contributed equally to this work

1 Abstract

2 Grey water footprint (GWF)¹ accounting has previously been conducted at the global level using a bottom-up approach but lacking detailed industrial information. Here we applied a multi-region 3 4 input-output approach based on the World Input-Output Database (WIOD) to quantify global GWF 5 of 40 countries/regions with 35 economic sectors. The GWF from both the production perspective (GWFP), and the consumption perspective (GWFC) are quantified. The results show that the global 6 7 GWFP/GWFC was 1507.9 km³ in 2009. Except for the "Agriculture, Hunting, Forestry and Fishing" 8 sector, the industrial sectors with the largest GWFC were "Food, Beverages and Tobacco", 9 "Construction", "Chemicals and Chemical Products", and "Textiles and Textile Products". The BRIC 10 countries (Brazil, Russia, India, China) had a larger GWFP than their GWFC, which accounted for 11 over half of global GWFP (53.6%), and their GWFP was mainly generated from the production of 12 domestic final demand. In contrast, the OECD29 and EU27 groups of countries i.e. the country 13 groups consisting mainly of economically advanced nations, had larger GWFC than their GWFP. 14 Overall, the OECD29 and EU27 outsourced 134.8 km³ and 64.4 km³ of their grey water respectively, 15 mostly to large newly advanced economies such as the BRIC group of countries, which, in turn, were collectively outsourcing 112 km³ of grey water. Quantitative approaches are thus suggested 16 17 for development, aimed at shared responsibility for water pollutant discharge among poor 18 exporters and wealthy consumers.

19 Keywords: Global assessment; Input-output analysis; Virtual water trade; Water pollution

¹Abbreviations: WF, water footprint; GWF, grey water footprint; GWFP, grey water footprint from the production perspective; GWFC, grey water footprint from the consumption perspective; IGWF, internal grey water footprint; GVWE, grey virtual water exports; GVWI, grey virtual water imports; WIOD, World Input-Output Database.

20 outsourcing.

21 Introduction

22 It is widely recognised that the ever growing demand for goods and services for human 23 consumption are the main drivers of water resource depletion and water pollution (Hoekstra and 24 Mekonnen, 2012; Munksgaard et al., 2005; Zhao et al., 2016a). Further, the spatially uneven 25 distribution of consumption and water resource depletion has been identified as having a 26 substantial impact on countries/regions with water intensive production (Feng et al., 2014; Liu et 27 al., 2015; Orlowsky et al., 2014; Zhao et al., 2015). Highlighting such issues has added a global 28 dimension to regional water management (Vörösmarty et al., 2015). However, despite severe 29 water pollution problems around the world, most studies seeking to understand the impact of 30 trade and consumption on water only consider water quantity and ignore water quality (van Vliet 31 et al., 2017; Zhao et al., 2016a).

32 One of the helpful concepts used to describe human impacts on water quantity and water 33 quality is that of the water footprint (WF). The WF of a product may be defined as the volume of 34 freshwater use, measured directly and indirectly, through the supply chain to support the final 35 demand of a particular product (Hoekstra et al., 2011). Three components are included in WF 36 accounting: the blue and green WF refer to the use of blue water (groundwater and surface water) 37 and green water (soil water) to support final demand. These two components represent the impact 38 of human consumption in terms of water quantity. In contrast, the grey WF (GWF) is a measure of 39 the volume of water required to assimilate the pollution load generated as a result of final demand 40 of a particular product back to ambient levels and, as such, is a water quality indicator (Chapagain 41 et al., 2006; Hoekstra et al., 2011). Hence, the GWF unifies the impact of both water quantity and quality into a homogeneous unit: freshwater volume. The GWF also allows the comparison of environmental impacts produced by pollutants discharged into waterbodies with different natural conditions (Pellicer-Martínez and Martínez-Paz, 2016). It has been quantified at different spatial levels (Cazcarro et al., 2016; Serrano et al., 2016; Zhi et al., 2015), as well as at global level (Hoekstra and Mekonnen, 2012; Liu et al., 2012; Liu et al., 2017).

47 The accounting of WF (including GWF) may be approached from two perspectives. Accounting 48 from the production perspective includes the production of goods/services for domestic 49 consumption, and those for export (Hoekstra and Mekonnen, 2012; Peters, 2008). Regional water 50 management has traditionally focused on local water supply i.e. managing water to support the 51 water demand of local production (Lenzen et al., 2013). However, WF accounting from the 52 production perspective takes this a step further to divide local water supply into domestic use 53 (internal WF) and export use (Hoekstra and Mekonnen, 2012). As a result, the internal and external 54 driving force of local water supply due to final demand can be shown (Zhao et al., 2016b). In 55 contrast, accounting from the consumption perspective includes water used in the production of 56 all goods/services intended for local consumption (internal WF), including imports (external WF). 57 From the consumption perspective, all water use, no matter whether it is within or outside the 58 regional boundary, occurring along the chains of production and distribution are allocated to the 59 final product consumer (Wiedmann, 2009). Hence, the consumption perspective can show the 60 impacts on water resources of consumers in one region from the same or another region, thus 61 informing consumer responsibility. Overall, it is important to address both perspectives to 62 understand how the environmental costs of water pollution are redistributed from the 63 countries/regions of consumers to that of producers.

64 In terms of the quantification of WF (including GWF), two approaches are commonly taken 65 i.e. the bottom-up and top-down approach. Global GWF accounting has previously been conducted 66 using the bottom-up approach for agricultural products from both production and consumption 67 perspectives (Hoekstra and Mekonnen, 2012). Compared to the top-down approach, the bottom-68 up approach is less sensitive to trade data and enables a detailed inventory of the agricultural 69 sector (Hoekstra and Mekonnen, 20102), but is more difficult to quantify for the industrial sector 70 owing to the complexity of the supply chain (Feng et al., 2011). However, the breakdown of the 71 industrial sector is important in GWF accounting because of the substantial amount of waterborne 72 pollutants discharged by the sector, such as from the textiles, chemicals, and papermaking 73 industries. A top-down approach based on input-output analysis allows the quantification of the 74 GWF with details of industrial sectors. In addition, the top-down approach allows the quantification 75 of the GWF in a unified framework from both consumption and production perspectives. Recent 76 studies have shown that a multi-region input-output (MRIO) analysis is an appropriate 77 methodology for quantifying the WF (including the GWF) from the consumption perspective 78 (Lenzen et al., 2013; Serrano et al., 2016). However, as far as we know, studies which have 79 quantified the global GWF from both production and consumption perspectives under an MRIO 80 framework are rare.

In this study we account for the global GWF from the production perspective (GWFP for short) and the consumption perspective (GWFC for short) using the top-down approach. A global MRIO table, the World Input-Output Database (WIOD), was used to study the global GWF. The WIOD covers 40 countries/regions and 35 economic sectors, including 1 agricultural sector, 17 industrial sectors, and 17 service sectors (Timmer et al., 2015). GWFP may be further divided into the internal GWF (IGWF) and grey virtual water exports (GVWE), and GWFC is subdivided into IGWF and grey
 virtual water imports (GVWI). The implications from the results of GWFC and GWFP, as well as the
 comparison between the two perspectives, are also discussed.

89 **Data**

90 The WIOD (www.wiod.org) provides the MRIO table from 1995 onwards and (blue, green, and 91 grey) water use data between 1995 and 2009 (Genty, 2012; Timmer et al., 2015). It is currently the 92 only database providing grey water use for a range of sectors (Serrano et al., 2016). There are 35 93 sectors for each country/region including 1 agricultural sector, 17 industrial sectors, and 17 service 94 sectors (See Fig. 2 for details). Hence, this study made use of the data from the WIOD for 2009 in 95 order to acquire the most up-to-date grey water use data (http://www.wiod.org/database/eas13). The MRIO table in the WIOD lists 40 countries/regions (see Fig. 2 for details), including 27 EU 96 97 countries, 29 OECD countries, and the BRIC countries (Brazil, Russia, India and Mainland China); 98 see Table 1 for details. The economic activities of Chinese Taipei was separated from Mainland 99 China (hereafter China for short), according to the data provided. Other countries/regions not in 100 the list were categorized as Rest of World (ROW).

In the WIOD, grey water use for different sectors was estimated using different methods depending on data availability (Genty, 2012). The grey water use of crop and livestock production was quantified based on non-point source pollution of nitrogen, according to Mekonnen and Hoekstra (2010a; 2010b). The sum of grey water use for crop and livestock production was aggregated and assigned to the "Agriculture, Hunting, Forestry and Fishing" sector in the WIOD (Genty, 2012). The grey water use of hydropower was estimated based on Mekonnen and Hoekstra (2011b), and assigned to the "Electricity, Gas and Water Supply" sector. The grey water use for industrial production (excluding hydropower) was a measure of the part of the return flow which
is discharged into the environment without prior treatment with the assumption that the dilution
factor is 1 (Mekonnen and Hoekstra, 2011a). The total industrial grey water use was then
distributed to WIOD sectors based on the share of water use in the database generated by the
EXIOPOL project² (Genty, 2012; EXIOPOL, 2011).

113 Methodology

In a fundamental input-output framework, the total output of different sectors in region r in column vector form \mathbf{x}^r is the sum of intermediate input in matrix form \mathbf{Z}^{rr} and final demand in column vector form \mathbf{f}^r , shown as follows:

117
$$\mathbf{x}^r = \mathbf{Z}^r \mathbf{i} + \mathbf{f}^r$$
(1)

118 where i is a column vector of 1.

119 In order to reflect the functional relationship between final demand and total output, the technical coefficient \mathbf{A}^{rr} is introduced. The relationship between \mathbf{Z}^{rr} , \mathbf{A}^{rr} , and \mathbf{x}^{r} can be denoted as 120 $\mathbf{Z}^{rr}\mathbf{i} = \mathbf{A}^{rr}\mathbf{x}^{r}$, where \mathbf{A}^{rr} is the technical coefficient of region r in matrix form, acquired through 121 122 dividing the total output of each sector by their intermediate inputs. Hence, equation (1) becomes 123 $\mathbf{x}^r = \mathbf{A}^{rr} \mathbf{x}^r + \mathbf{f}^r$. The above equation can be further transformed to the fundamental inputoutput formula $\mathbf{x}^r = (\mathbf{I} - \mathbf{A}^{rr})^{-1} \mathbf{f}^r$, where $(\mathbf{I} - \mathbf{A}^{rr})^{-1}$ is the Leontief Inverse Matrix showing the 124 125 total output of different sectors necessary for one unit of final demand. 126 In a generalized environmental input-output model for a single region, the environmental

127 accounts (in our case, the sectoral grey water use accounts) are added as an additional row to the

² EXIOPOL is an integrated project funded by the European Commission to build an input-output framework linking other socio-economic models to estimate environmental impacts of economic activities.

input-output table to link the economic sectors to grey water use. In order to acquire the GWF of region *r*, we first introduce the direct grey water use intensity of region *r*, \mathbf{d}^r , which represents the direct grey water use per unit of output in each sector, i.e. $\mathbf{gwu}^r = \hat{\mathbf{d}}^r \mathbf{x}^r$. Here, \mathbf{gwu}^r is the sectoral grey water use of region *r* in column vector form, and $\hat{\mathbf{d}}^r$ is the direct grey water use intensity of region *r* in diagonal form. Combining the above equation with the fundamental inputoutput formula, we can obtain the GWF of region *r* derived from its final demand:

134
$$\mathbf{gwf}^r = \hat{\mathbf{d}}^r (\mathbf{I} - \mathbf{A}^{rr})^{-1} \hat{\mathbf{f}}^r$$
(2)

To quantify the GWF of many countries/regions from both consumption and production perspectives, we applied the "Water Embodied in Trade" (WET) approach based on the generalized environmental input-output model. Full details of this approach may be found in Feng et al. (2011) and Zhao et al. (2015), and herein we provide a simplified illustration of the methodology. The basic framework of WET for *p* regions in matrix form can be shown as:

$$\begin{bmatrix} \mathbf{x}^{1} \\ M \\ \mathbf{x}^{r} \\ M \\ \mathbf{x}^{r} \end{bmatrix} = \begin{bmatrix} \mathbf{A}^{11} & \mathbf{L} & \mathbf{0} & \mathbf{L} & \mathbf{0} \\ M & \mathbf{O} & \mathbf{M} & \mathbf{M} \\ \mathbf{0} & \mathbf{L} & \mathbf{A}^{rr} & \mathbf{L} & \mathbf{0} \\ M & \mathbf{M} & \mathbf{O} & \mathbf{M} \\ \mathbf{0} & \mathbf{L} & \mathbf{0} & \mathbf{L} & \mathbf{A}^{pp} \end{bmatrix} \begin{bmatrix} \mathbf{x}^{1} \\ M \\ \mathbf{x}^{r} \\ \mathbf{x}^{p} \end{bmatrix} + \begin{bmatrix} \mathbf{y}^{1} + \sum_{s \neq 1} \mathbf{e}^{1s} \\ M \\ \mathbf{y}^{r} + \sum_{s \neq p} \mathbf{e}^{rs} \\ M \\ \mathbf{y}^{p} + \sum_{s \neq p} \mathbf{e}^{ps} \end{bmatrix}$$
(3)

where \mathbf{y}^{r} is the domestic final demand of region r in vector form, and $\sum_{s\neq r} \mathbf{e}^{rs}$ is the exports of region r to fulfill the final demand of different regions s.

The IGWF of a region may be defined as the grey water use derived from domestic production to support the final demand of the region. Under the WET framework, the IGWF of region r may be denoted as $\mathbf{igwf}^r = \sum \hat{\mathbf{d}}^r (\mathbf{I} \cdot \mathbf{A}^{rr})^{-1} \hat{\mathbf{y}}^r$. The GVWE of a region is defined as the grey water use derived from the production of exported products from that region, and the GVWE of region r may thus be denoted as $\mathbf{gvwe}^r = \sum \hat{\mathbf{d}}^r (\mathbf{I} \cdot \mathbf{A}^{rr})^{-1} \sum_{s \neq r} \hat{\mathbf{e}}^{rs}$. The GVWI is defined as the grey water use derived from the production of imported products for local final demand, and the GVWI 149 of region *r* can therefore be shown as $\mathbf{gvwi}^r = \sum \sum_{s \neq r} \hat{\mathbf{d}}^s (\mathbf{I} - \mathbf{A}^{ss})^{-1} \hat{\mathbf{e}}^{sr}$

where $\hat{\mathbf{d}}^{s}$ is the direct grey water intensity of region *s* in diagonal matrix form, $\hat{\mathbf{e}}^{sr}$ is the export from region *s* to fulfill the final demand of different regions *r* in diagonal matrix form, and \mathbf{A}^{ss} is the technical coefficient of region *s* in matrix form.

By combining the IGWF and GVWE we can derive the GWFP of region r, $gwfp^r = igwf^r + gvwe^r$, and the GWFC of region r is the combination of IGWF and GVWI, $gwfc^r = igwf^r + gvwi^r$.

156 **Results**

157 Global grey water footprint from the production perspective

Total and sectoral GWFP of different countries/regions were collected from the WIOD (Genty, 2012). The total global GWFP was 1507.9 km³, with China identified as having the largest GWFP, with a total of 536.6 km³, followed by the USA and India with 193.2 km³ and 174.6 km³ respectively. Other countries with large GWFP's were Russia (65.0 km³), Canada (33.6 km³), Brazil (31.4 km³) and Indonesia (31.0 km³). In terms of country groups, the GWFP of the BRIC countries accounted for over half of global GWFP (53.6%), and the GWFP of the OECD29 and EU27 bloc amounted to 24.0% and 6.8%, respectively.

165 In terms of sectoral GWFP, almost all GWFP was allocated to seven sectors including 166 "Agriculture, Hunting, Forestry and Fishing", "Basic Metals and Fabricated Metal", "Chemicals and 167 Chemical Products", "Pulp, Paper, Printing and Publishing", "Food, Beverages and Tobacco", 168 "Textiles and Textile Products", and "Other Non-metallic Mineral". The "Agriculture, Hunting, 169 Forestry and Fishing" sector had the largest GWFP, accounting for 60% of total GWFP, followed by 170 the "Basic Metals and Fabricated Metal" and "Chemicals and Chemical Products" sectors accounting for 13% and 11% of GWFP, respectively.

| 172 | Environmental externality and responsibility cannot be discussed without addressing |
|-----|---|
| 173 | interpersonal equity. Taking into account the per capita value, Canada had the largest GWFP, |
| 174 | generating 997.8 m ³ /person of grey water. This is attributed to the relatively low population of |
| 175 | Canada. With the exception of the USA (629.8 m ³ /person), Chinese Taipei (507.3 m ³ /person), and |
| 176 | China (403.8 m ³ /person), the top 10 countries/regions with the largest per capita GWFP were all |
| 177 | from Eastern and Southern Europe, including Bulgaria (685.7 m^3 /person), Hungary (535.5 |
| 178 | m ³ /person), Slovenia (466.0 m ³ /person), Russia (455.0 m ³ /person), Poland (413.9 m ³ /person), and |
| 179 | Romania (370.3 m ³ /person). |
| 180 | GWFP can be divided into IGWF and GVWE to identify internal and external drivers to local |
| 181 | grey water production; in total GVWE constitutes 24.6% of total GWFP. The share of GVWE varies |
| 182 | among countries/regions from 12.3% (Greece) to 87.9% (Netherlands). The countries/regions with |
| 183 | the largest GWFP had a relatively lower share of GVWE, for example China (22.8%), USA (20.6%), |
| 184 | India (12.4%), and Russia (17.6%). Obviously, these countries are large economies and produce |
| 185 | mainly to fulfill domestic consumption. |
| 186 | Global grey water footprint from the consumption perspective |
| 187 | Insert Fig. 1 |
| 188 | When examined from a consumption perspective, the GWFC embodied in a country/region's |
| 189 | final demand indicates how much grey water would be generated in order to meet its final |
| 190 | consumed goods and services. Since the total GWFC is equal to the total GWFP globally, GWFC also |
| 191 | refers to the redistribution of grey water use generated from the production side to the |

192 consumption side through the supply chain. The national/regional GWFC's are presented in Fig. 1a.

193 For specific countries/regions, China tops the list with a GWFC of 440.0 km³, accounting for nearly 30% of GWFC. The USA is ranked second with a GWFC of 214.8 km³, and other countries ranked 194 195 sequentially among the top 5 are India, Russia and Germany. The GWFC of the top five countries 196 accounted for 60% of global GWFC. In terms of country groups, we found the EU27 accounted for 197 11% of global GWFC, and the results were 33% and 46% for OECD29 and BRIC, respectively. These 198 findings indicate a large amount of GWFC was generated by the so-called "emerging economies". 199 When taking into account the per capita value (Fig. 1b), North America had the highest GWFC per capita, with Canada and the USA ranking first and second respectively. In 2009, the 200 201 consumption of an average Canadian and American citizen generated water pollution requiring 202 914.8 m³/person and 700.2 m³/person freshwater respectively. Excepting Chinese Taipei, Russia, 203 and Bulgaria, the top 10 countries/regions with the largest GWFC per capita were all from the 204 OECD29, including Canada, USA, Australia (551.9 m³), Slovenia (515.4 m³), Netherlands (514.6 m³), 205 Luxembourg (506.3 m³), and Belgium (477.9 m³). Some countries had a large total GWFC but small 206 per capita amount, for instance India ranked third in terms of its national GWFC but had the lowest 207 per capita amount of only 133.9 m³. Similarly, the inverse relationship can be seen in some 208 countries, for example in 2009 Luxembourg generated only 0.25 billion m³ GWFC in total, equating 209 to 506.3 m³ per person.

210

Insert Fig. 2

In terms of sectoral distribution, the top five sectors contributing the most to global GWFC
were "Agriculture, Hunting, Forestry and Fishing", "Food, Beverages and Tobacco", "Construction",
"Chemicals and Chemical Products", and "Textiles and Textile Products". These five sectors were
responsible for 31.0%, 21.0%, 9.8%, 5.3%, and 4.4% respectively of GWFC throughout the world in

2009, amounting to 71.5% of GWFC. The national/regional level of sectoral distribution are shown
in Fig. 2. At national/regional level the pattern varies; the contribution of "Agriculture, Hunting,
Forestry and Fishing" to national/regional GWFC ranged from 15% (Bulgaria) to 52% (Latvia). For
"Food, Beverages and Tobacco", the contribution ranged from 10% (Romania) to 38% (Finland), for
"construction" from 0.1% (Luxembourg) to 21% (China), for "Chemicals and Chemical Products"
between 2% (China) and 24% (Malta), and for "Textiles and Textile Products" between 1% (Hungary)
and 20% (Turkey).

222 External and internal grey water footprint from the consumption perspective

223 The GWFC may be divided into IGWF and GVWI. Globally, the GVWI was 371.4 km³ accounting 224 for 24.6% of the total GWFC. The top five countries/regions with the largest GVWI were the USA 225 (61.4 km³), Japan (26.0 km³), China (25.8 km³), Germany (25.0 km³), and Canada (12.7 km³). The 226 share of GVWI varied significantly at national/regional scale (Fig. 3), ranging from 5.9% (China) to 227 98.8% (Netherlands). The EU27 bloc had the largest share of GVWI; globally the top five countries 228 with the largest GVWI were the Netherlands, Malta, Luxembourg, Cyprus, and Ireland. As a result, 229 the GVWI for the EU27 bloc constituted 63.5% of the total GWFC. In contrast the BRIC countries 230 were among the main countries having the largest share of IGWF. Over 90% of GWFC in China (94.1%) and India (94.1%) were generated within their own territories, as well as 83.5% in 231 232 Indonesia, 82.9% in Russia, and 82.2% in Brazil. The share of IGWF varied greatly between the EU27 233 and BRIC countries, probably because the BRIC countries have larger land areas and relatively self-234 dependent economies.

235

Insert Fig. 3

236 Global outsourcing of grey water

| 237 | If one country imports more grey water than it exports to other countries through its trade |
|-----|---|
| 238 | balance i.e. is a net importer of grey water, we can say that this country is outsourcing its water |
| 239 | pollution to other countries. Conversely, if a country exports more grey water than it imports, i.e. |
| 240 | is a net exporter of grey water, we can say this country is being outsourced by other countries. In |
| 241 | 2009, there were 10 countries/regions being outsourced compared to 30 outsourcing |
| 242 | countries/regions (Fig. 4). As can be seen in Fig. 4, the top 5 outsourcing countries were Japan, USA, |
| 243 | Germany, UK and South Korea, while the top 5 countries/regions being outsourced were China, |
| 244 | India, Poland, Brazil, and Canada. |
| 245 | Insert Fig. 4 and Fig. 5 |
| 246 | Grey water outsourcing shows different patterns per country/region (Fig. 5). For example, as |
| 247 | the largest outsourcing country, Japan outsourced 24.6 km ³ of grey water in 2009. Breaking down |
| 248 | the aggregate result, Japan outsourced 24.7km ³ of grey water to 34 out of 40 countries/regions, |
| 249 | but had only 0.1 km ³ of grey water outsourced to it by 6 countries/regions. The two countries Japan |
| 250 | outsourced most to were China and USA, which accounted for 46.1% and 18.3% of total grey water |
| 251 | outsourcing from this country. It is also worth noting that the volume of grey water outsourced by |
| 252 | Japan to China (11.4km ³) was even higher than grey water consumption within Japan for producing |
| 253 | goods and services (6.4km ³). As the second largest outsourcing country, the USA outsourced 34.7 |
| 254 | km ³ of grey water to 12 countries/regions, but was also being outsourced by 28 countries/regions |
| 255 | to the tune of 13.1 km ³ , resulting in total outsourcing of 21.6 km ³ grey water. The USA mainly |
| 256 | outsourced grey water to China, India and Canada, accounting for 77.4% of its total grey water |
| 257 | outsourcing to a total of 12 countries/regions. However, USA was being outsourced a large amount |

of grey water mainly from Japan, Mexico, and Korea, accounting for 66.2% of its total grey water
being outsourced. The top three countries outsourcing the largest share of grey water were Malta,
the Netherlands, and Cyprus where net exports represented 96%, 90%, and 88% of consumptionbased GWF respectively.

262 The largest net grey water exporting country, i.e. country being outsourced, was China with 263 96.7 km³ from the other 41 countries/regions contained in our study (Fig. 5). This volume 264 corresponds to 18% of GWFP for China, and was close to the GWFP of the EU27 (102.8 km³). The 265 top three countries/regions that outsourced to China were the USA (19.9 km³), Japan (11.4 km³), 266 and Germany (6.6 km³). These three countries were also the top three outsourcing countries 267 examined by our study. Beyond China, the volume being outsourced fell to 12 km³ for the second 268 largest country India, and to 3.7km³ for the third largest, Poland. The top three countries that being 269 outsourced the largest share of their GWFP were Bulgaria, Hungary, and Poland. Net exports of 270 grey water for these three countries represented 38%, 24%, and 23% of GWFP, respectively.

271 **Discussion**

272 The quantification of GWFC showed that the top 10 countries/regions with the greatest GWFC 273 were also the most populous (Fig. 6). This is simply because these countries require tremendous 274 amounts of products to satisfy final demand of their populations. Final demand may be met 275 through either domestic production or imports of externally produced goods and services. Similarly, 276 the GWFC may either be mostly generated within the country/region with larger IGWF, or outside 277 the country/region with larger GVWI. The GWFC of the newly advanced economies of China, India, 278 Russia, Brazil (i.e. the BRIC countries), as well as Indonesia, was mostly generated within the 279 national boundary, with the share of IGWF in GWFC making up 94.1%, 94.1%, 82.9%, 82.2%, and

280 83.5%, respectively. Similarly, the USA and Canada were found to depend mainly on domestic 281 production, but also generated a substantial share of GVWI accounting for 28.6% and 41.2% of 282 their GWFC, respectively. Therefore, the BRIC countries, the USA, and Canada could substantially 283 reduce their GWFC by reducing their IGWF i.e. regulating production related pollutant discharges 284 associated with domestic final demand within their national boundaries. This is important because 285 the total IGWF of these 7 nations (842.4 km³) made up over half the global GWFC, i.e. 55.9%.

286

Insert Fig. 6 and Table 1

287 Conversely, the GWFC of Japan and Germany, also at the top of the global GWFC list, was 288 found to depend mainly on external production, with shares of GVWI accounting for 75.5% and 289 83.5% of their respective GWFC. These two countries, along with other developed nations such as 290 USA, UK, Italy, the Netherlands, and France, are at the top of the list of grey water outsourcing 291 countries/regions. The OECD29 and EU27 i.e. the country groups consisting mainly of economically 292 developed nations, outsourced 134.8 km³ and 64.4 km³ of their grey water mostly to large 293 developing countries such as BRIC, which imported 112 km³ of grey water. Overall, countries which 294 rely mainly on external production in order to meet domestic final demand raises the question of 295 who is responsible for discharge of waterborne pollution, and how this burden shifting 296 could/should be shared(Davis and Caldeira, 2010; Zhao et al., 2016a). Similar questions have been 297 raised in relation to CO_2 emissions, but rarely for water quality issues (Peters, 2008; Wiedmann, 298 2009). One of the few exceptions is from Zhao et al. (2016a), who studied the megacity Shanghai 299 as a case to show how wealthy consumers have largely transferred water quality stress to other 300 Chinese provinces. As a result, they suggested taking measures at national/regional, industrial, and 301 consumer levels to obtain shared responsibility between wealthy consumers and often poorer

exporters, as well as promoting greater demand-side management (Zhao et al., 2016a). Based on
the results of the present study, we recommend similar measures could be adopted at the global
level for mitigating water quality stress induced by consumption in developed countries/regions.
As a result, quantitative approaches could be developed aimed at shared responsibility for water
pollution among poorer exporters and wealthy consumers.

307 Overall, GWF is an appropriate tool for global assessment of water pollution because it is a 308 homogenous indicator enabling comparison of water pollution impacts spatially and across 309 regions/countries (Pellicer-Martínez and Martínez-Paz, 2016). However, the concept itself has a 310 number of limitations which currently constrain its accuracy for global assessment. First, the results 311 of GWF will vary substantially based on the selection of different water quality standards and data 312 sources of natural pollutant concentrations (Mekonnen and Hoekstra, 2015). Second, GWF is 313 generally assessed using individual classes of pollutant. Assessments of GWF have overwhelmingly 314 been focused on nitrogen-related pollutants, despite the cocktails of pollutants known to exist in 315 polluted waterbodies, and which may therefore significantly underestimate the GWF (Liu et al., 316 2017). For example, Liu et al. (2017) found that the volume of freshwater needed to dilute 317 phosphorus inputs is much higher than for nitrogen loads. Third, it is difficult to validate the results 318 of GWF since the indicator cannot be measured directly (Liu et al., 2012). However, despite all 319 these limitations, GWF still has great potential as a sustainability indicator for water pollution at different spatial levels (Pellicer-Martínez and Martínez-Paz, 2016). The present study also contains 320 321 the well documented limitations typical of input-output analyses and MRIO approaches. In 322 particular, regional and sectoral aggregation of MRIO tables exists, and a higher degree of 323 aggregation tends to more severe issues around quantification accuracy (Lenzen et al., 2013). Last,

but not least, both the data of grey water use by sector and MRIO table need updating to include more recent years. MRIO data is often outdated on release because of the problems of dealing with incomplete, conflicting, and misaligned data (Lenzen et al., 2012). Discharges of water pollutants by sector and ambient water quality standards to quantify grey water use by sector are often not available, or cannot be accessed, for many countries (Liu et al., 2017). Overcoming the limitations of the GWF concept and the relative models is therefore key to developing the acceptability of GWF in the policy domain.

331 **Conclusions**

332 We have quantified the global GWFC/GWFP for 40 countries/regions from both the 333 production and consumption perspectives. Although the GWFC and GWFP have the same value at 334 global level, they were found to vary greatly among different countries/regions. Generally, the 335 results from the production perspective (GWFP) helped to identify country/regional hotspots with 336 potential water quality problems, and further manifest the internal (IGWF) and external (GVWE) 337 driving forces to local water pollution discharge. We found that most of the GWFP was 338 concentrated in a small number of geographically large countries/regions; with about 59% of the 339 global GWFP found in North America and BRIC countries. Despite their large IGWF, these countries 340 also topped the list for the largest GVWE, sharing 59% of the global GVWE, meaning that the 341 pollutant discharge of these countries was largely driven by external market forces. The results 342 from the consumption perspective (GWFC), allocating water pollution discharge to the final 343 consumer of products and services, confirms previous research that developed countries (e.g. Japan, Germany) and country groups (OECD29, EU27) have outsourced their pollution to 344 345 developing countries (e.g. the BRIC countries) to a large extent.

346 The MRIO approach provides a unified framework to account for the GWF from both 347 production and consumption perspectives enabling detailed information on industrial sectors. 348 Globally, industrial sectors accounted for 40.0% of the GWFP, but 56.1% of GWFC. Most industrial 349 sectors (13 out of 17) had smaller GWFP but higher GWFC due to their demand for intermediate 350 products (raw materials) which discharge large amounts of water pollution. Hence, it is necessary 351 to take a life-cycle perspective for key industrial sectors to make sure the supply of raw materials 352 in undertaken in an environmentally conscientious way (Lenzen et al., 2007). Our results identified 353 these key industrial sectors from the GWFC point of view: sectors such as "Food, Beverages and 354 Tobacco", "Construction", and "Textiles and Textile Products" are recommended to take additional 355 responsibility for their upstream grey water use. The top-down approach from the consumer 356 perspective thus provides an appropriate tool for such analysis. However, it is essential to reduce 357 the high levels of uncertainty derived from estimating complex production chains from the 358 consumption perspective using the MRIO approach (Peters, 2008). To reduce such uncertainty, 359 further efforts need to be made towards developing the global MRIO table with more detailed 360 sectoral information, as well as developing corresponding and updated data for sectoral grey water 361 use at the national level. In addition, improvements in the scientific robustness of GWF is important 362 in its use as a tool for waterborne pollutant mitigation (Liu et al., 2017).

Acknowledgements: This work was supported by the National Key R&D Program of China (2017YFC0405203, 2016YFC0401703), the Fundamental Research Funds for the Central Universities (No. 2016B13814), and the National Natural Science Foundation of China (No. 5177090079, 51579071, 51779075).

17

367 **References**

Cazcarro, I., Duarte, R., Sánchez-Chóliz, J., 2016. Downscaling the grey water footprints of
production and consumption. J Clean Prod 132, 171-183.
https://doi.org/10.1016/j.jclepro.2015.07.113

Chapagain, A.K., Hoekstra, A.Y., Savenije, H.H.G., Gautam, R., 2006. The water footprint of cotton
consumption: An assessment of the impact of worldwide consumption of cotton products on the
water resources in the cotton producing countries. Ecol Econ 60, 186-203.
<u>https://doi.org/10.1016/j.ecolecon.2005.11.027</u>

- Davis, S.J., Caldeira, K., 2010. Consumption-based accounting of CO2 emissions. Proc Natl Acad Sci
 USA 107, 5687-5692. <u>https://doi.org/10.1073/pnas.0906974107</u>
- EXIOPOL (2011): A new environmental accounting framework using externality data and inputoutput tools for policy analysis, FP6 Research Project: 2007-2011. http://www.feemproject.net/exiopol/
- Feng, K., Chapagain, A., Suh, S., Pfister, S., Hubacek, K., 2011. Comparison of Bottom-up and Top down Approaches to Calculating the Water Footprints of Nations. Econ Syst Res 23, 371-385.
 https://doi.org/10.1080/09535314.2011.638276
- Feng, K., Hubacek, K., Pfister, S., Yu, Y., Sun, L., 2014. Virtual Scarce Water in China. Environ Sci
 Technol 48, 7704-7713. https://doi.org/10.1021/es500502q
- 385 Genty, A. (Ed.). 2012. Final database of environmental satellite accounts: technical report on their
- compilation. WIOD Deliverable 4.6, Documentation. www.wiod.org/publications/source_docs/Environmental_Sources.pdf. Accessed 17 April 2016.
- Hoekstra, A.Y., Chapagain, A.K., Aldaya, M.M., Mekonnen, M.M., 2011. The water footprint
 assessment manual: Setting the global standard. Earthscan.
- Hoekstra, A.Y., Mekonnen, M.M., 2012. The water footprint of humanity. Proc Natl Acad Sci USA
 109, 3232-3237. <u>https://doi.org/10.1073/pnas.1109936109</u>
- Lenzen, M., Kanemoto, K., Moran, D., Geschke, A., 2012. Mapping the structure of the world
- 393 economy. Environ Sci Technol 46, 8374-8381. https://doi.org/10.1021/es300171xLenzen, M.,
- 394 Moran, D., Bhaduri, A., Kanemoto, K., Bekchanov, M., Geschke, A., Foran, B., 2013. International
- trade of scarce water. Ecol Econ 94, 78-85. https://doi.org/10.1016/j.ecolecon.2013.06.018
- Lenzen, M., Murray, J., Sack, F., Wiedmann, T., 2007. Shared producer and consumer responsibility —
- 397 Theory and practice. Ecol Econ 61, 27-42. <u>https://doi.org/10.1016/j.ecolecon.2006.05.018</u>

398 Liu, C., Kroeze, C., Hoekstra, A.Y., Gerbens-Leenes, W., 2012. Past and future trends in grey water

- footprints of anthropogenic nitrogen and phosphorus inputs to major world rivers. Ecol Indic 18,
 42-49. https://doi.org/10.1016/j.ecolind.2011.10.005
- Liu, J., Mooney, H., Hull, V., Davis, S.J., Gaskell, J., Hertel, T., Lubchenco, J., Seto, K.C., Gleick, P.,
 Kremen, C., Li, S., 2015. Systems integration for global sustainability. Science 347.
 https://doi.org/10.1126/science.1258832
- Liu, J., Savenije, H.H.G., 2008. Time to break the silence around virtual-water imports. Nature 453,
 587. <u>https://doi.org/10.1038/453587c</u>
- Liu, W., Antonelli, M., Liu, X., Yang, H., 2017. Towards improvement of grey water footprint
 assessment: With an illustration for global maize cultivation. J Clean Prod 147, 1-9.
 https://doi.org/10.1016/j.jclepro.2017.01.072
- 409 Mekonnen, M. and Hoekstra, A. (2010a). The green, blue and grey water footprint of farm

- 410 animals and animal products, Value of Water Research Report Series No.48, Volume I and II.
- 411 UNESCO-IHE, Delft, the Netherlands.
- 412 Mekonnen, M. and Hoekstra, A. (2010b). The green, blue and grey water footprint of crops and
- derived crop products. Value of Water Research Report Series No.47, Volume I and II. UNESCO-
- 414 IHE, Delft, the Netherlands.
- 415 Mekonnen, M., Hoekstra, A. (2011a). National water footprint accounts: the green, blue and grey
- 416 water footprint of production and consumption. Value of Water Research Report Series No.50,
- 417 Volume I and II. UNESCO-IHE, Delft, the Netherlands.
- 418 Mekonnen, M., Hoekstra, A. (2011b). The water footprint of electricity from hydropower. Value of
- 419 Water Research Report Series No.51. UNESCO-IHE, Delft, the Netherlands.
- 420 Mekonnen, M.M., Hoekstra, A.Y., 2015. Global Gray Water Footprint and Water Pollution Levels
- 421 Related to Anthropogenic Nitrogen Loads to Fresh Water. Environ Sci Technol 49, 12860-12868.
 422 https://doi.org/10.1021/acs.est.5b03191
- 423 Munksgaard, J., Wier, M., Lenzen, M., Dey, C., 2005. Using Input Output Analysis to Measure
- 424 the Environmental Pressure of Consumption at Different Spatial Levels. J Ind Ecol 9, 169-185.
- 425 <u>https://doi.org/10.1162/1088198054084699</u>
- 426 Orlowsky, B., Hoekstra, A.Y., Gudmundsson, L., Seneviratne, S.I., 2014. Today's virtual water
 427 consumption and trade under future water scarcity. Environ Res Lett 9, 074007.
 428 https://doi.org/10.1088/1748-9326/9/7/074007
- 429 Pellicer-Martínez, F., Martínez-Paz, J.M., 2016. Grey water footprint assessment at the river basin
- 430 level: Accounting method and case study in the Segura River Basin, Spain. Ecol Indic 60, 1173-1183.
 431 https://doi.org/10.1016/j.ecolind.2015.08.032
- 432 Peters, G.P., 2008. From production-based to consumption-based national emission inventories.
- 433 Ecol Econ 65, 13-23. <u>https://doi.org/10.1111/jiec.12454</u>
- 434 Serrano, A., Guan, D., Duarte, R., Paavola, J., 2016. Virtual Water Flows in the EU27: A
- 435 Consumption-based Approach. J Ind Ecol 20, 547-558. <u>https://doi.org/10.1111/jiec.12454</u>
- 436 Timmer, M.P., Dietzenbacher, E., Los, B., Stehrer, R., de Vries, G.J., 2015. An Illustrated User Guide
- to the World Input–Output Database: the Case of Global Automotive Production. Review of
 International Economics 23, 575-605. <u>https://doi.org/10.1111/roie.12178</u>
- Vörösmarty, C.J., Hoekstra, A.Y., Bunn, S.E., Conway, D., Gupta, J., 2015. Fresh water goes global.
 Science 349, 478. http://doi.org/10.1126/science.aac6009
- van Vliet, M.T.H., Flörke, M., Wada, Y., 2017. Quality matters for water scarcity. Nature Geosci 10,
 800-802. http://doi.org/10.1038/ngeo3047
- Wiedmann, T., 2009. A review of recent multi-region input–output models used for consumptionbased emission and resource accounting. Ecol Econ 69, 211-222.
 https://doi.org/10.1016/j.ecolecon.2009.08.026
- 446 Zhao, X., Liu, J., Liu, Q., Tillotson, M.R., Guan, D., Hubacek, K., 2015. Physical and virtual water 447 transfers for regional water stress alleviation in China. Proceedings of the National Academy of 448 of Sciences the United States of America 112, 1031-1035. 449 https://doi.org/10.1073/pnas.1404130112
- 450 Zhao, X., Liu, J., Yang, H., Duarte, R., Tillotson, M.R., Hubacek, K., 2016a. Burden shifting of water
- 451 quantity and quality stress from megacity Shanghai. Water Resour Res, 52, WR018595
- 452 <u>https://doi.org/10.1002/2016WR018595</u>
- 453 Zhao, X., Tillotson, M., Yang, Z., Yang, H., Liu, J., 2016b. Reduction and reallocation of water use of

- 454 products in Beijing. Ecol Indic 61, Part 2, 893-898. <u>https://doi.org/10.1016/j.ecolind.2015.10.043</u>
- 455 Zhi, Y., Yang, Z., Yin, X., Hamilton, P.B., Zhang, L., 2015. Using gray water footprint to verify economic
- 456 sectors' consumption of assimilative capacity in a river basin: model and a case study in the Haihe
- 457 River Basin, China. J Clean Prod 92, 267-273. <u>https://doi.org/10.1016/j.jclepro.2014.12.058</u>





Fig .1 Greywater footprint from the consumption perspective (GWFC) of 40 countries/regions,

(a) national/regional total, (b) per capita





Fig. 3 The share of internal grey water footprint (IGWF) and grey virtual water import (GVWI)



Fig. 4 Net import of grey water footprint (km³)



Fig. 5 Grey water outsourcing among 40 countries/regions (the ribbons and links in the same color as the country/region mean the country/region was a net exporter of grey virtual water, otherwise it was net importer of grey virtual water. For example, the red ribbon and links from China shows China was a net exporter of grey virtual water to other countries/regions).



Fig. 6 comparison of the top 10 GWFC and GWFP countries

| Groups | Countries | GWFC | GWFP | IGWF | GVWI | GVWE | GWFC |
|--------|--|-------|-------|-------|-----------------|-------|----------------|
| | | Unit: | Unit: | Unit: | Unit: | Unit: | per |
| | | km³ | km³ | km³ | km ³ | km³ | capita |
| | | | | | | | Unit: |
| | | | | | | | m ³ |
| EU27 | Austria, Belgium, Bulgaria, Cyprus, | 167.2 | 102.8 | 61.1 | 106.1 | 41.7 | 335 |
| | Czech, Germany, Denmark, Spain, | | | | | | |
| | Estonia, Finland, France, UK, Greece, | | | | | | |
| | Hungary, Ireland, Italy, Lithuania, | | | | | | |
| | Luxembourg, Latvia, Malta, | | | | | | |
| | Netherlands, Poland, Portugal, | | | | | | |
| | Romania, Slovak, Slovenia, Sweden | | | | | | |
| OECD29 | Australia, Austria, Belgium, Canada, | 496.4 | 361.6 | 258.7 | 237.7 | 102.9 | 416 |
| | Czech, Germany, Denmark, Spain, | | | | | | |
| | Estonia, Finland, France, UK, Greece, | | | | | | |
| | Hungary, Ireland, Italy, Japan, Korea, | | | | | | |
| | Luxembourg, Latvia, Mexico, | | | | | | |
| | Netherlands, Poland, Portugal, | | | | | | |
| | Slovak, Slovenia, Sweden, Turkey, | | | | | | |
| | USA | | | | | | |
| BRIC | Brazil, Russia, India, China | 695.4 | 807.5 | 643.8 | 51.6 | 163.7 | 241 |

 Table 1 Comparison of grey water footprint among EU27, OECD29, and BRIC countries