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1 **Climate-induced phenological shifts in a Batesian mimicry complex**

2

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18

19 Running title: Phenological shifts in a mimicry complex

20

21 **Summary**

22 Climate-induced changes in spatial and temporal occurrence of species, as well as species  
23 traits such as body size, each have the potential to decouple symbiotic relationships. Past  
24 work has focused primarily on direct interactions, particularly those between predators and  
25 prey and between plants and pollinators, but studies have rarely demonstrated significant  
26 fitness costs to the interacting, co-evolving organisms. Here, we demonstrate that changing  
27 phenological synchrony in the latter part of the 20th century has different fitness outcomes  
28 for the actors within a Batesian mimicry complex, where predators learn to differentiate  
29 harmful “model” organisms (stinging Hymenoptera) from harmless “mimics” (hoverflies,  
30 Diptera: Syrphidae). We define the mimetic relationships between 2,352 pairs of stinging  
31 Hymenoptera and their Syrphidae mimics based on a large-scale citizen science project and  
32 demonstrate that there is no relationship between the phenological shifts of models and their  
33 mimics. Using computer game-based experiments, we confirm that the fitness of models,  
34 mimics, and predators differs among phenological scenarios, creating a phenologically-  
35 antagonistic system. Finally, we show that climate change is increasing the proportion of  
36 mimetic interactions in which models occur first and reducing mimic-first and random  
37 patterns of occurrence, potentially leading to complex fitness costs and benefits across all  
38 three actors. Our results provide the first strong evidence for an overlooked example of  
39 fitness consequences from changing phenological synchrony.

40

41 **Keywords:** hover flies, Batesian mimicry, phenology, mismatch, climate change

42 **Significance statement**

43 Climate change can degrade ecological interactions by separating interacting species in  
44 space and time, but this is not the case in one of the best-studied examples of mimicry in  
45 which hoverflies (mimics) imitate stinging wasps and bees (models). While there is no  
46 evidence of the emergence of mimics and models tracking climate change in the same way,  
47 historical records suggest that the mimicry complex is undergoing complex shifts in  
48 evolutionary pressures under climate change through changes in the relative emergence  
49 patterns of model-mimic pairs. This finding is based on the first community-level description  
50 of mimetic relationships (comparing 2,352 pairs of species) and the most comprehensive  
51 demonstration of the importance of phenology for the fitness of mimics, models, and  
52 predators.

53 **/body**

54 **Introduction**

55 The biological consequences of climate change for individual species have been  
56 documented in detail: distributions are moving poleward (1), phenological events are  
57 changing (most often documented as an advance in spring phenological events (2)), and  
58 organism size is reducing (3). These changes also have consequences for ecological  
59 interactions, as interacting species may become separated in space (4) or time (5, 6). The  
60 challenge for the study of these eco-evolutionary processes is the linking of (i) ecological  
61 data on the nature of relationships among taxa, (ii) the shifting spatiotemporal associations  
62 among those taxa, and (iii) data on the fitness costs and benefits that result from changes in  
63 the strength of interactions. Here, we provide a comprehensive evaluation of a putative  
64 symbiosis that has the potential to provide significant insights into community-level, climate-  
65 driven ecological change: model-mimic complexes. Mimetic relationships can be cooperative  
66 (e.g. Müllerian mimicry, where multiple defended species evolve a common phenotype as a  
67 result of selection to share the burden of educating predators (7)), or parasitic (e.g. Batesian  
68 mimicry, where an undefended species exhibits the phenotype of a defended species in  
69 order to benefit from the learned aversion of predators to that phenotype (8)). Batesian  
70 mimics (8) exploit a range of sensory modalities to enhance their similarity to defended  
71 models (9). One taxon that exploits multiple sensory cues is the hoverflies (Diptera:  
72 Syrphidae), which have evolved to produce visual (10), behavioural (11) and acoustic (12)  
73 cues that resemble those of stinging Hymenoptera.

74

75 Previous work has suggested that the order of appearance of mimics and models is  
76 important to the success of mimicry. In a classic study, Mostler demonstrated that avian  
77 predators learn to avoid stinging Hymenoptera quickly, but that being presented with  
78 palatable Diptera can reverse that learning (13). High fidelity syrphid mimics of stinging  
79 Hymenoptera in North America emerge predominantly in spring before their models (14, 15).  
80 It has been suggested that this phenological pattern has evolved so that the mimics emerge

81 before naïve fledglings begin to feed independently and, to begin with, indiscriminately (16).  
82 Once the predators have been educated by the Hymenoptera models, the mimics show a  
83 second peak in emergence later in the summer. While consistent with theory on Batesian  
84 mimicry, other work has failed to support this idea, showing that mimics and models are  
85 largely synchronous and independent of fledging dates (17). However, no studies have  
86 attempted to test this idea comprehensively over large datasets and temporal ranges. A  
87 further source of uncertainty stems from the role of contemporary climate change in altering  
88 the spatiotemporal patterns of occurrence of mimics and models (18), with consequences  
89 for predator learning and, ultimately, the fitness of mimics, models, and the predators  
90 themselves.

91

92 Our understanding of the evolutionary consequences of global environmental change  
93 remains poorly developed, partly because of the difficulty in measuring the consequences of  
94 natural selection and partly due to interactions between phenotypic plasticity and adaptive  
95 change (19). In this paper we make use of alternative approaches to the study of  
96 evolutionary change under warming climates to quantify shifting phenological antagonism in  
97 a Batesian mimicry complex. This is achieved through (i) defining the mimetic relationships  
98 within a large pool of Batesian models and mimics, (ii) quantifying the phenological trends  
99 between those models and mimics, (iii) estimating empirically the fitness consequences of  
100 phenological asynchrony for mimics, models and predators, and finally (iv) demonstrating  
101 changes in optimal phenological patterns under recent climate change.

102

## 103 **Results and Discussion**

### 104 **Study 1: Community-level mimetic networks**

105 We selected 42 species of Syrphidae and 56 species of bees and wasps from UK species  
106 lists based on their abundance in UK biological records and their taxonomic and  
107 morphological distinctness (see Table S1 for the species list). All of the 2,352 species pairs  
108 are known to have co-occurred spatially in at least one 10 x 10 km grid square based on

109 biological records, with Jaccard overlap indices (the ratio of shared squares to the number of  
110 squares in which one or both species was found) of 0.9-33.3% (mean 12.6%). A web  
111 interface was created that randomly paired single representative images from the Syrphidae  
112 and Hymenoptera and requested users to rate the pairing on a scale from 1 (not at all  
113 similar) to 10 (extremely similar). The experiment ran from 18 March 2016 to 28 March 2017  
114 during which the 2,352 potential pairwise combinations had been rated a total of 30,300  
115 times, with a minimum of 3 ratings and a maximum of 29 ratings for individual pairs. The  
116 modal rating was 1 (indicating negligible similarity) and the mean rating  $3.0 \pm 2.2$  SD (see SI  
117 Appendix, fig. S1). We consider any mean ratings  $\geq 5$  to be indicative of a mimetic  
118 relationship, which was the case for 237 pairwise combinations (hereafter “high-fidelity  
119 pairs”). These 237 species pairs had Jaccard overlap indices of 2.6-28.4% (mean 13.5%).  
120 fig. 1 shows a matrix of those mimetic similarities, highlighting these “islands of mimicry” in  
121 the wider Batesian complex that form the basis for our subsequent analyses, and example  
122 pairs for different mean similarity scores (see SI Appendix, fig. S2 for a larger version of fig.  
123 1 with species labelled). Our results correlate significantly with data from experiments using  
124 pigeons (20) ( $r=0.757$ ,  $p=0.030$ , see SI Appendix, fig. S3), indicating that our human ratings  
125 are meaningful..

126

## 127 **Study 2: Comparative phenology of models and mimics**

128 We use the rank biserial correlation (RBC, a correlation between rank data [e.g. emergence  
129 date] and a categorical variable [e.g. species] (21)) to quantify the degree of overlap in  
130 phenology, where  $RBC=0$  is random occurrence,  $RBC=-1$  is all Hymenoptera models  
131 emerging before Syrphidae mimics and  $RBC=1$  is all Syrphidae mimics emerging before  
132 Hymenoptera models. Average RBC was significantly lower than zero (median= $-0.015$ ,  $V =$   
133  $1709500000$ ,  $p<0.001$ ) for the entire community of 56 Hymenoptera and 42 Syrphidae,  
134 showing that, on average, the Syrphidae emerge later than the Hymenoptera. When only  
135 high-fidelity model-mimic pairs were included, median RBC was slightly but significantly  
136 greater from zero (median= $0.093$ ,  $V = 12092000$ ,  $p<0.001$ ; see SI Appendix, fig. S7 for

137 illustration of the relationship between RBC and difference in median flight date). The rate of  
138 phenological advance per unit increase in mean annual central England temperature in the  
139 leading edge (5<sup>th</sup> percentile) of the flight period is smaller in Syrphidae ( $-14.3 \text{ days}\cdot^{\circ}\text{C}^{-1} \pm 0.9$   
140 SE) than in Hymenoptera ( $-18.9 \text{ days}\cdot^{\circ}\text{C}^{-1} \pm 1.1 \text{ SE}$ ;  $t=22.806$ ,  $p<0.001$ ), but there is no  
141 difference in the rate of shift in the median flight date (Syrphidae:  $-8.1 \text{ days}\cdot^{\circ}\text{C}^{-1} \pm 0.8 \text{ SE}$ ;  
142 Hymenoptera:  $-8.1 \text{ days}\cdot^{\circ}\text{C}^{-1} \pm 0.8 \text{ SE}$ ;  $t=0.365$ ,  $p=0.715$ ) and the Syrphidae are advancing  
143 the trailing edge (95<sup>th</sup> percentile) of their flight period faster than Hymenoptera (Syrphidae:  
144  $1.3 \text{ days}\cdot^{\circ}\text{C}^{-1} \pm 0.8 \text{ SE}$ ; Hymenoptera:  $2.6 \text{ days}\cdot^{\circ}\text{C}^{-1} \pm 1.1 \text{ SE}$ ;  $t=-6.491$ ,  $p<0.001$ ; see  
145 Supplementary Information for details). When considering only high-fidelity pairs, the results  
146 were quantitatively similar with Syrphidae ( $-14.4 \pm 0.4 \text{ days}\cdot^{\circ}\text{C}^{-1}$ ) advancing the leading edge  
147 of the flight period more slowly than Hymenoptera ( $-19.1 \pm 0.7 \text{ days}\cdot^{\circ}\text{C}^{-1}$ ;  $t=5.793$ ,  $p<0.001$ ),  
148 no difference in median flight date (Syrphidae:  $-7.0 \pm 0.4 \text{ days}\cdot^{\circ}\text{C}^{-1}$ ; Hymenoptera:  $-7.1 \pm 0.4$   
149  $\text{days}\cdot^{\circ}\text{C}^{-1}$ ;  $t=0.054$ ,  $p=0.957$ ) or the trailing edge of the flight period (Syrphidae:  $2.1 \pm 0.4$   
150  $\text{days}\cdot^{\circ}\text{C}^{-1}$ ; Hymenoptera:  $2.8 \pm 0.6 \text{ days}\cdot^{\circ}\text{C}^{-1}$ ;  $t=1.052$ ,  $p=0.294$ ). There is no significant  
151 relationship between the rates of phenological shift in high-fidelity models and mimics  
152 (leading edge: Spearman's  $\rho=-0.039$ ,  $p=0.604$ ; median:  $\rho=0.026$ ,  $p=0.728$ ; trailing edge:  $\rho=-$   
153  $0.004$ ,  $p=0.959$ ; dark points in SI Appendix, fig. S6). In other words, we find no evidence that  
154 models and mimics are advancing their phenology at the same rate.

155

### 156 **Study 3: Fitness consequences of phenological mismatch**

157 We quantified the fitness consequences of phenological change using a computer game-  
158 based behavioural experiment within which human participants could act as “predators” and  
159 make decisions concerning the profitability of three pairs of prey stimuli: (i) *Apis mellifera*  
160 and *Eristalis tenax*; (ii) *Vespula vulgaris* and *Chrysotoxum cautum*; and (iii) *Bombus*  
161 *terrestris* and *Criorhina ranunculi* (see SI Appendix, fig. S8 for stimuli and fig. S9 for an  
162 example screenshot). Participants were presented with all three model-mimic pairs in one of  
163 three phenological scenarios involving 25 models and 25 mimics: (i) mimics on average first,

164 (ii) models on average first, or (iii) random presentation with equal mean order of  
165 presentation. Mimic- and model-first scenarios were created by increasing the relative  
166 probability of the later species from 0% to 100% in increments of 2% over the 50 screens  
167 (example sequences are shown in Table S2), resulting in a mean RBC of 0.677 or -0.677  
168 (SD=0.103) depending on whether the model or mimic occurs earlier. Participants gained 5  
169 points for clicking a mimic, lost 10 points for clicking a model, and leaving the insect did not  
170 change the score.

171

172 The phenological scenario had a significant effect on predation rates on models ( $\chi^2=49.218$ ,  
173  $df=2$ ,  $p<0.001$ ) and mimics ( $\chi^2=34.544$ ,  $df=2$ ,  $p<0.001$ ), and on the score achieved by  
174 human predators ( $\chi^2=51.282$ ,  $df=2$ ,  $p<0.001$ ; see Table S3 for full model results). Random  
175 presentation produced the highest fitness (survival rate) in mimics, and these were  
176 significantly higher than in model-first ( $z=3.073$ ,  $p=0.006$ ) or mimic-first scenarios ( $z=5.773$ ,  
177  $p<0.001$ ; fig. 2A). The model-first sequence of prey items produced the highest fitness  
178 (survival rate) in models, and those outcomes were significantly higher than random  
179 ( $z=3.050$ ,  $p=0.006$ ) or mimic-first ( $z=6.983$ ,  $p<0.001$ ; fig. 2B), in agreement with the theory  
180 underpinning Batesian mimicry. However, random presentation produced significantly lower  
181 predator scores than mimic-first ( $z=5.390$ ,  $p<0.001$ ) or model-first ( $z=6.849$ ,  $p<0.001$ ; fig.  
182 2C). These results highlight the phenological antagonism among the three actors within the  
183 mimicry system: models benefit from educating predators (model-first), mimics benefit from  
184 unpredictability (random), and predators benefit from consistent education on either prey  
185 item (model-first or mimic-first).

186

### 187 **Temporal trends in fitness**

188 Finally, we can infer temporal trends in fitness outcomes for model-mimic pairs under  
189 contemporary climate change based on how their RBCs change over time. From Study 3 we  
190 know that there are differences in fitness outcomes where the order of appearance of

191 models and mimics corresponds to mean RBCs of  $<-0.677$  (“model-first”), -  
192  $0.677 < \text{RBC} < 0.677$  (“random”) or  $>0.677$  (“mimic-first”), and so we can apply those  
193 thresholds to the biological recording data to infer fitness consequences of real world  
194 sequences. Therefore, for each year, we classify each of the 237 high-fidelity pairs based on  
195 their RBC into a “model-first”, “random” or “mimic-first” pattern. There was a significant  
196 increase between 1960 and 2005 in the proportion of interactions in which the RBC  
197 corresponded to a model-first pattern ( $\rho=0.454$ ,  $p=0.001$ , fig. 3A), a significant decline in the  
198 proportion of mimic-first sequences ( $\rho=-0.427$ ,  $p=0.003$ , fig. 3B), and a weakly significant  
199 decline in the proportion of pairs in which the species occurred randomly ( $\rho=-0.295$ ,  
200  $p=0.044$ , fig. 3C; see SI Appendix, fig. S10 for sensitivity analysis around thresholds).  
201 Hence, we can infer a positive fitness trend for models from increasing model-first pairings  
202 (where models perform best) and decreasing mimic-first pairings (where models perform  
203 worst), a mixed fitness trend for mimics due to the decreasing proportion of random pairs  
204 (where mimics perform best) and a decrease in mimic-first pairs (where mimics perform  
205 worst), and positive fitness benefits for predators from the increasing proportions of model-  
206 first pairs (where predators perform best) and decreasing proportion of random pairs (where  
207 predators perform worst).

208

209 The evidence is building for a significant impact of phenological decoupling in a wide variety  
210 of systems. Snowshoe hares that moult after snowmelt show significant increases in  
211 mortality (22), many studies have shown that avian fitness is compromised if peak  
212 abundance of food does not coincide with chick rearing (e.g. (23)), and tritrophic studies  
213 suggest that oak-caterpillar-bird systems may have little room for buffering from warming  
214 springs (24). These exemplar studies are being carried out against a back drop of dynamic  
215 shifts in the degree of phenological synchrony over the past few decades (5, 25). However,  
216 while previous studies have tended to find the negative aspects of phenological shifts, our

217 data suggest that climate change will result in an increase in phenologically-optimal  
218 emergence patterns that benefit (at least in part) all three actors within the mimicry system.  
219  
220 Building on past work (e.g. 13), we have now developed comprehensive evidence to support  
221 the theory (26) that the evolutionary costs and benefits of mimicry to models, mimics, and  
222 predators depend upon the relative phenological patterns of models and mimics. We have  
223 used these findings to help understand the implications of changes in the temporal overlap  
224 of models and mimics in a classical Batesian mimicry system. The different actors within the  
225 mimicry system each experience costs and benefits from different phenological patterns:  
226 hymenopteran models benefit in all cases from increased model-first, decreased mimic-first  
227 and decreased random patterns because these simplify – and, hence, accelerate – predator  
228 learning of aposematic signals. Mimics benefit from the decrease in mimic-first patterns and  
229 increase in model-first patterns because these accelerate predator learning of aposematic  
230 cues. However, mimics may suffer from a decline in randomness (their optimal phenological  
231 scenario according to Study 2), if predators are able to shift prey preferences as mimics  
232 increase in relative abundance later in the season. Finally, predators benefit from the  
233 reduction of randomness if they are able to respond by exploiting mimics when they become  
234 numerically dominant, but also show more subtle responses to declines in mimic-first and  
235 increased in model-first patterns. The results illustrate the benefits of integrating mechanistic  
236 and observational data to study large-scale eco-evolutionary processes within a  
237 phenologically antagonistic Batesian mimicry complex.

238

## 239 **Methods**

### 240 **Study 1: Community-level mimetic networks**

241 Rather than relying upon subjective, isolated descriptions of mimetic relationships between  
242 particular species from published work, we derived a full matrix of mimetic associations  
243 using an extensive, online citizen science project. Human ratings of visual similarity correlate  
244 with data from experiments with birds and with ratings based on morphometric analysis (10)

245 and provide a Gestalt perspective on similarity that avoids issues with the definition of  
246 particular traits and with variations in both size and shape that can complicate computational  
247 image analysis (e.g. (27)). The online mimicry experiment, which can still be found at  
248 [www.mimicryexperiment.net](http://www.mimicryexperiment.net), was based around a simple PHP script that randomly selected  
249 two images – one from a pool of 42 Hymenoptera images and one from a pool of 56  
250 Syrphidae images (one image for each species). The landing page of the online study  
251 contained brief details about the project and contact details for the lead author if participants  
252 required any further information. Participants were instructed to click on a button if they  
253 consented to take part. Study 1 was approved by the University of Leeds Faculty of  
254 Biological Sciences Research Ethics Committee (ref BIOSCI 16-006). Species were  
255 selected for inclusion based on a hierarchical process: first, the most common species were  
256 selected from the highest ranked abundance in the Hoverfly Recording Scheme (HRS) and  
257 the Bees, Wasps, and Ants Recording Scheme (BWARS) datasets (see below for details of  
258 those schemes). The rationale behind this criterion was that more common species are  
259 more likely to interact and, therefore, to have a true mimetic relationship if species with a  
260 similar morphology were present. We calculated the proportion of these putative model-  
261 mimic pairs that are known to have co-occurred in the same 10km grid square in the HRS  
262 and BWARS dataset. Secondly, congeners with close morphological similarity were  
263 excluded. There are many genera of UK Syrphidae and Hymenoptera with similar  
264 morphology and we refrained from using physically similar congeners to reduce phylogenetic  
265 autocorrelation and redundant morphological variation in the dataset (e.g. *Melanostoma*  
266 *mellinum* was included but *M. scalare* was excluded, while *Lasioglossum leucozonium* was  
267 included but *L. villosulum* was excluded). Thirdly, additional species were added to  
268 incorporate further morphological diversity where particularly distinct morphologies were  
269 known to the authors (particularly the rarer *Chrysotoxum cautum*, *Criorhina ranunculi* and  
270 *Arctophila superbiens* from among the Syrphidae and *Vespula rufa* from among the  
271 Hymenoptera, which are all thought to be involved in mimetic relationships (28)). Since the  
272 pairing of images was done at random, the number of comparisons between pairs of images

273 was not equal across the dataset. The ratings gathered during the experiment exhibited a  
274 highly positively skewed distribution, with 53.2% of raw ratings and 56.1% of mean ratings  
275 being <3, while only 5.1% of raw ratings and 0.3% of mean pair ratings were >7 (see SI  
276 Appendix, fig. S1). Representative pairs of images are shown in fig. 1 for mean pair  
277 similarities of 7.5, 4.9, 3.1 and 1.0, along with a matrix of similarity values for all 2,352  
278 comparisons. To validate the online experiment, we compared the ratings given to  
279 comparisons between eight of the hoverflies in our study that had previously been compared  
280 to *Vespula vulgaris* in pigeon experiments (20). In that previous experiment, pigeons were  
281 trained to peck at an image of *Vespula vulgaris* in return for a food reward, and then shown  
282 different Syrphidae images. The peck rate in response to the Syrphidae images was  
283 assumed to be proportional to the pigeon's perception of the similarity of the Syrphidae to  
284 the original training stimulus. Despite the low number of species in the pigeon data, as in  
285 previous studies that have used human ratings there was a significant correlation between  
286 our human ratings experiment and the pigeon peck rate data ( $r=0.757$ ,  $p=0.030$ , see SI  
287 Appendix, fig. S3), indicating that our human ratings reflect morphological similarity in much  
288 the same way that birds might assess it.

289

## 290 **Study 2: Comparative phenology of models and mimics**

291 We extracted biological records from two extensive, long-term biological recording schemes,  
292 the HRS and BWARS datasets, to recreate past trends in phenology in model and mimic  
293 communities. Both HRS and BWARS are citizen science projects that rely on ad hoc  
294 recording by a community of recorders, and require compulsory fields relating to the species  
295 identity, location of sighting, and name of recorder, with desirable fields that describe  
296 ecological variables (e.g. flowers visited, pollen collected, prey/host). While abundance data  
297 is very occasionally recorded for each submitted record on a broad categorical scale, these  
298 data are rarely available and so have not been incorporated into this analysis. Validation of  
299 records for both schemes involves checking the validity of species names and geographical  
300 coordinates, while subsequent verification is based on identification difficulty, known spatial

301 distributions, known seasonal phenology, and scarcity (29, 30). Prior to analysis, we cleaned  
302 the biological records to remove records from before 1960, records without ordinal dates,  
303 and records without valid species names. After this processing, the HRS contained 620,460  
304 records of 288 hoverfly species from between 1960 and 2014 at time of analysis (27 January  
305 2015). The BWARS dataset contained 451,624 records of 547 species from between 1960  
306 and 2013 at time of analysis (27 January 2015). Biological recording data from the HRS and  
307 BWARS datasets show spatiotemporal patterns that are consistent with other UK biological  
308 recording data: a strong increase in the number of records through the latter part of the 20<sup>th</sup>  
309 Century (see SI Appendix, fig. S4) and a strong concentration of records in the south of the  
310 country and around centres of population density (see SI Appendix, fig. S5). Further  
311 descriptive statistics for the Syrphidae can be found in (31).

312

313 For each year in which a species was recorded, we calculated the 5<sup>th</sup>, 50<sup>th</sup> and 95<sup>th</sup>  
314 percentile flight dates (representing the leading edge, median, and trailing edge of the flight  
315 period, respectively) and then conducted linear regressions of these dates against a general  
316 measure of mean annual UK temperature (central England temperature [CET] (32)). CET  
317 uses an average of values from a set of three long-term meteorological stations in central  
318 England to create an averaged trend over the country. We calculated Pearson correlation  
319 coefficients to provide a measure of the strength of the trends, and regression coefficients to  
320 provide a measure of the rate of change in phenology. A comparison of model and mimic  
321 shifts across the three flight dates can be seen in SI Appendix, fig. S6.

322

323 To quantify the relative phenology of the model and mimic communities, we assumed that  
324 the sequence of occurrence is the relevant phenological metric for learning within a Batesian  
325 mimicry system. We calculated rank biserial correlations (RBC) of the flight dates of all  
326 potential model-mimic pairs and for the subset of pairs for which mean similarity ratings were  
327  $\geq 5$ . The RBC approach gives a correlation coefficient ranging from -1 (no overlap, first  
328 sample entirely before second sample) to +1 (no overlap, first sample entirely after second

329 sample). An RBC of 0 indicates random occurrence. To illustrate the relationship between  
330 RBCs and differences in flight dates within the BWARS and HRS datasets, the RBCs  
331 calculated for the real-world species pairs from the BWARS and HRS datasets are plotted  
332 against the difference in median flight dates in SI Appendix, fig. S7.

333

### 334 **Study 3: Fitness costs of phenological mismatch**

335 We designed a computer game within which human participants could act as “predators”  
336 and make decisions concerning the profitability of different “prey”. A psychological approach  
337 was selected above computational algorithms for a number of reasons. First, as we note  
338 elsewhere in the manuscript, we have demonstrated that human ratings of similarity are  
339 correlated with those of avian model systems. Second, the available algorithms for  
340 quantifying model-mimic similarity such as the distance transform method (27), and the  
341 neural net approach of Bain et al. (33) are highly data hungry, requiring extensive data to  
342 parameterize them compared to our “Gestalt” system based on human assessments. There  
343 is no reason to expect that these computational methods would produce a more ecologically  
344 relevant result than the use of human scores. Indeed, the neural network classifier of Bain  
345 et al. (33) ranked the similarity of hoverfly species to wasps in a very similar manner to  
346 humans ( $R^2 = 0.74$ ,  $P < 0.001$ ; 34). Third, we were interested in generating a fitness  
347 measure that was more than just similarity but also included speed-accuracy trade-offs that  
348 are important to prey survival (35). The game was built in the Vizard (WorldViz) virtual reality  
349 environment, programmed in Python, and displayed using an Oculus Rift DK2 immersive  
350 virtual reality headset (full field of view horizontal visual angle =  $100^\circ$ ) while participants were  
351 seated. Participants were given an opportunity to read an information sheet describing the  
352 study and ask any questions before giving written informed consent to take part. Study 3  
353 was approved by the University of Leeds Faculty of Biological Sciences Research Ethics  
354 Committee (ref BIOSCI 15-021). The experiment began with a short training phase during  
355 which participants were asked to “eat” (click) red triangles and leave blue circles, and a  
356 scoreboard kept track of their points. All participants performed well during this phase.

357 Following the training phase, participants were presented with the experimental stimuli.  
358 These stimuli comprised three pairs of insect images chosen to represent the three broad  
359 groups of model-mimic relationships within the UK Syrphidae-Hymenoptera community: (i)  
360 *Apis mellifera* and *Eristalis tenax*; (ii) *Vespula vulgaris* and *Chrysotoxum cautum*; and (iii)  
361 *Bombus terrestris* and *Criorhina ranunculi* (see SI Appendix, fig. S8 for stimuli). Participants  
362 were presented with a series of sequences of insect images against a grass background to  
363 enhance the noise in the image and reduce the use of image artefacts (e.g. cropping of  
364 images) as cues to the identity of the insects (see SI Appendix, fig. S9). Each screen  
365 contained only one insect at any time. The insects were displayed in random positions and  
366 orientations on each trial, all within a 30° central visual angle (horizontally and vertically).  
367 The insect images covered approximately 5° visual angle. The three pairs of insects were  
368 presented in one of three phenological scenarios involving 25 models and 25 mimics: (i)  
369 mimics more likely to be first, (ii) models more likely to be first, or (iii) random presentation  
370 (example sequences are shown in Table S2). Sequences were generated by increasing the  
371 probability of occurrence of one image from 0% to 100% in increments of 2% over the 50  
372 presentations, producing a series of unique model-first sequences with a mean RBC of  
373 0.678 (SD=0.103). Sequences were reversed to give the mimic-first sequences in order to  
374 retain the same RBC. Each participant was asked to decide whether or not to “eat” an insect  
375 by pressing a key on a keyboard, after which a score counter would change to reflect  
376 whether the decision was “correct”: consuming harmless Syrphidae increased the score by 5  
377 points, consuming stinging Hymenoptera reduced the score by 10 points, and leaving the  
378 insect did not change the score. Participants were told that “eating” some types of insects  
379 would reduce their score while eating others would increase it, and that the aim was to score  
380 as many points as possible.

381

382 We analysed the data from 45 participants to evaluate the consequences of phenological  
383 scenario on (i) mimic survival, (ii) model survival, and (iii) final participant scores (as a proxy

384 for predator fitness). Each participant experienced all three of the phenological scenarios,  
385 where each scenario was presented using a different model-mimic pair so that the  
386 participants also saw all three model pairs (e.g. a participant may have seen *B. pratorum*  
387 and *C. ranunculi* in the model-first scenario, *A. mellifera* and *E. tenax* in the mimic-first  
388 scenario, and *V. vulgaris* and *C. cautum* in the random scenario). For tests (i), and (ii) we  
389 used generalised linear mixed effects models with binomial errors in the lme4 package in R  
390 to analyse the survival or predation of each target as a binary variable. The trial number  
391 (from 1 to 150) was a covariate and phenological scenario was a fixed effect, while the  
392 model-mimic pair and the participant ID were entered as random effects. For analysis (iii) the  
393 same models were run but with general linear mixed effects models with the participant  
394 score as the response and normal error distribution. The residuals of the model for analysis  
395 (iii) were checked to ensure that the data met the assumptions of normality and  
396 homogeneous variance. Full model results can be found in Table S3.

397

### 398 **Temporal Trends in Fitness**

399 To explore the sensitivity of the temporal trends (fig. 3, main text) to the choice of thresholds,  
400 we conducted a sensitivity analysis by recalculating the RBC trends for thresholds that were  
401 1 standard deviation above (-0.781 and 0.781, for model-first and mimic-first, respectively)  
402 and below (-0.574 and 0.574) the mean of the distribution of RBCs (-0.677 and 0.677). The  
403 results were qualitatively robust to the variation in threshold, with the number of model-first  
404 scenarios always increasing (-1SD:  $\rho=0.428$ ,  $p=0.002$ ; mean:  $\rho=0.455$ ,  $p=0.001$ ; +1SD:  
405  $\rho=0.489$ ,  $p=0.001$ ), mimic-first scenarios always decreasing but not always significantly so  
406 (-1SD:  $\rho=-0.607$ ,  $p<0.001$ ; mean:  $\rho=-0.427$ ,  $p=0.003$ ; +1SD:  $\rho=-0.120$ ,  $p=0.431$ ), and  
407 random scenarios always decreasing but not always significantly so (-1SD:  $\rho=-0.158$ ,  
408  $p=0.285$ ; mean:  $\rho=-0.295$ ,  $p=0.044$ ; +1SD:  $\rho=-0.414$ ,  $p=0.005$ ). SI Appendix fig. S10  
409 shows the comparison of time series.

410

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417

418 **Author contributions**

419 C.H. and T.N.S. conceived the project; C.H. conducted the citizen science experiment,  
420 conducted the behavioural experiment, analysed the biological records, and analysed the  
421 data; J.B. coded the behavioural experiment and advised on behavioural methods. All  
422 authors discussed the results and provided comments on the manuscript.

423

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- 509

510 **Figure legends**

511 **Fig. 1.** Above: Heat map of mimetic ratings between 56 Syrphidae and 42 Hymenoptera.  
512 Colours indicate the mean similarity rating for each pair. The heat map has been ordered to  
513 place the Syrphidae with the highest mean similarity on the left of the plot and the  
514 Hymenoptera with the highest mean similarity at the bottom of the plot. See SI Appendix, fig.  
515 S2 for an enlarged version with species names. Below: Representative pairs of  
516 Hymenoptera (upper row) and Syrphidae mimics (lower row) that were rated as high  
517 similarity (A=Anthophora plumipes, E=Arctophila superbiens, rating=7.5), medium-high  
518 similarity (B=Vespula vulgaris, F=Chrysotoxum festivum, rating=4.9), low-medium similarity  
519 (C=Apis mellifera, G=Cheilosia impressa, rating=3.1), and low similarity (D=Bombus  
520 pasuorum, H=Baccha elongata, rating=1.0). Photographs copyright Steven Falk and used  
521 with permission.

522

523 **Fig. 2.** Fitness consequences of phenological asynchrony in Batesian mimics, models, and  
524 predators. Dotted horizontal line at  $y=1$  in (A) and (B) shows model-first odds ratio, against  
525 which the other two phenological patterns are compared: (A) Batesian mimics are predated  
526 more under mimic-first scenarios, benefitting most from model-first scenarios. (B) Models  
527 are predated more often under mimic-first scenarios, benefitting more from random  
528 scenarios. (C) Predator fitness is greatest under model-first scenarios, then mimic-first, and  
529 lowest under random scenarios. Error bars are 95% confidence intervals.

530

531 **Fig. 3.** Trends in the number of high-fidelity model-mimic pairs in which (A) the model  
532 emerges first, (B) the mimic emerges first, and (C) emergence is random. Mimetic fidelity is  
533 derived from a large citizen science study, phenological trends are derived from >1 million  
534 biological records, and the three categories of emergence are defined using known fitness  
535 consequences from novel behavioural experiments. Shaded areas are 95% confidence  
536 intervals.





