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1 **Stream solute tracer timescales changing with discharge and reach length confound**  
2 **process interpretation**

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42  
43 **Key points:**

- 44 • Advection is the primary control on observed stream solute tracer responses.  
45 • The influence of spatial heterogeneity in morphology is muted by advection.  
46 • Interpretation of solute transport requires consideration of tracer timescales.

47 **Abstract**

48 Improved understanding of stream solute transport requires meaningful comparison of processes  
49 across a wide range of discharge conditions and spatial scales. At reach scales where solute  
50 tracer tests are commonly used to assess transport behavior, such comparison is still confounded  
51 due to the challenge of separating dispersive and transient storage processes from the influence  
52 of the advective timescale that varies with discharge and reach length. To better resolve  
53 interpretation of these processes from field-based tracer observations, we conducted recurrent  
54 conservative solute tracer tests along a 1-km study reach during a storm discharge period and  
55 further discretized the study reach into six segments of similar length but different channel  
56 morphologies. The resulting suite of data, spanning an order of magnitude in advective  
57 timescales, enabled us to (1) characterize relationships between tracer response and discharge in  
58 individual segments and (2) determine how combining the segments into longer reaches  
59 influences interpretation of dispersion and transient storage from tracer tests. We found that the  
60 advective timescale was the primary control on the shape of the observed tracer response. Most  
61 segments responded similarly to discharge, implying that the influence of morphologic  
62 heterogeneity was muted relative to advection. Comparison of tracer data across combined  
63 segments demonstrated that increased advective timescales could be misinterpreted as a change  
64 in dispersion or transient storage. Taken together, our results stress the importance of  
65 characterizing the influence of changing advective timescales on solute tracer responses before  
66 such reach scale observations can be used to infer solute transport at larger network scales.

67  
68 **Key words:** stream solute transport, transient storage, conservative tracer, storm event, statistical  
69 moments, advective timescale

## 70 **1. Introduction**

71           Meaningful comparison of stream solute transport processes across different discharge  
72 conditions and spatial scales (e.g., morphologic unit to reach to network) is necessary to  
73 accurately represent and predict solute transport through stream networks [e.g., Covino et al.,  
74 2011; Kelleher et al., 2013]. Such comparison remains a persistent methodological and  
75 conceptual problem due to the uncertain distinction between spatial variability of solute transport  
76 processes, such as dispersion and transient storage, and changes caused by varying discharge  
77 conditions [e.g., González-Pinzón et al., 2013]. At the reach scale, on the order of tens to  
78 thousands of meters, solute transport processes are commonly interpreted through stream solute  
79 tracer tests [e.g., Stream Solute Workshop, 1990]. Direct comparison of observed tracer  
80 responses, however, is impossible without separating the spatially variable solute transport  
81 process from the unique timescale (commonly transit time distribution) of each test [Harvey et  
82 al., 1996]. Unfortunately, this separation is not straightforward because the tracer response  
83 changes primarily as a function of the down-stream advective transport time (commonly modal  
84 transport time, hereafter advective time), which shifts with discharge and reach length selection  
85 [e.g., Ward et al., 2013a].

86           Traditionally, comparison of solute tracer responses is performed by standardizing reach  
87 lengths through dimensionless numbers that relate physical processes to advective times [Runkel,  
88 2002; Wagner and Harvey, 1997]. This standardization is thought to yield the appropriate  
89 “window of detection” inherent in tracer studies [i.e., the time from tracer first arrival to last  
90 detection; Harvey and Wagner, 2000] and allow for assessment of rapid exchanges between the  
91 stream and connected subsurface—the solute transport process often deemed most important to  
92 stream ecosystem functions [e.g., Boulton et al., 2010; Hester and Gooseff, 2010]. Subsurface

93 and surface tracer exchange flowpaths that return to the stream within the window of detection,  
94 but have residence times slower than the advective time, are defined as short-term storage  
95 (commonly “transient storage”) [e.g., Harvey et al., 1996]. Conversely, tracer flowpaths that  
96 return outside of the window of detection (i.e., not recovered) are considered long-term storage  
97 and reflected in a tracer-based channel water balance [Payn et al., 2009; Schmadel et al., 2014a].  
98 However, the window of detection changes with discharge and, consequently, defines an  
99 arbitrary boundary between short- and long-term storage regardless of the actual physical  
100 flowpaths [e.g., Ward et al., 2013a; Wondzell, 2006]. The window of detection may also be  
101 influenced by the tracer type and associated resolution of observations. For example, some  
102 tracers can provide more late-time tailing information than others due to differences in detection  
103 limit sensitivity (e.g., fluorescent tracers in comparison to salt tracers) [Drummond et al., 2012].  
104 While the tracer selection and interaction between the advective time and window of detection  
105 can influence the interpretation of short-term storage, the storage flowpaths themselves can  
106 change with discharge [Wondzell, 2011; Zarnetske et al., 2007], further complicating meaningful  
107 comparison of tracer responses and thus impeding accurate conceptualization of solute transport  
108 processes.

109         Recent studies have shown, with the availability of high-frequency discharge  
110 observations, that analyzing solute transport processes across different discharge conditions is  
111 essential to improve conceptualization of how transient storage and dispersion change or  
112 compete with advection [Dudley-Southern and Binley, 2015; Ward et al., 2013a; Zimmer and  
113 Lautz, 2014]. A more refined conceptual understanding of stream solute transport across a range  
114 of discharges can better facilitate the upscaling of reach scale observations to infer solute  
115 transport at larger network scales. Because solute transport processes are dynamic and spatially

116 variable along a stream network, current upscaling strategies use relationships between replicate  
117 reach scale processes (e.g., long-term storage assessed from tracer observations) and discharge  
118 [e.g., Covino et al., 2011; Stewart et al., 2011]. While there are reported relationships between  
119 solute transport processes and discharge—both for long-term [Covino et al., 2011; Ward et al.,  
120 2013a; Ward et al., 2013b] and short-term storage [Schmid et al., 2010; Ward et al., 2013a; Ward  
121 et al., 2013b]—varying advective times reflected in reach scale observations can lead to  
122 misinterpreting a change in discharge as an apparent change in transient storage or dispersion  
123 [e.g., Gooseff et al., 2007]. For example, when fixing study reach lengths throughout a watershed  
124 (e.g., standard 100-m or 200-m reaches [Payn et al., 2009; Ward et al., 2013b]), headwater  
125 reaches with lower discharge typically have larger windows of detection and advective times  
126 compared to higher discharge, downstream reaches. An apparent process relationship with  
127 discharge may be incorrectly inferred because a common assumption is that tracer studies across  
128 fixed reach lengths are directly comparable [Covino et al., 2011; Payn et al., 2009; Ward et al.,  
129 2013b], despite preexisting knowledge that different discharges in those fixed reaches will yield  
130 different windows of detection. The applicability of this assumption is uncertain as few studies  
131 have compared stream solute tracer studies across different discharges to characterize the  
132 influence of changing advective times and windows of detection on interpreting the processes of  
133 interest.

134         In this study, we directly examine the extent to which solute transport process  
135 information assessed from observed conservative solute tracer responses is comparable between  
136 different advective times and windows of detection. The objectives of this study are (1) to  
137 characterize relationships between physical solute transport processes and discharge in reach  
138 segments with distinct channel morphologies, and (2) to determine how study reach length

139 influences the interpretation of these processes. To achieve these objectives, we conducted seven  
140 conservative solute tracer injections during a storm discharge period with in-stream responses  
141 recorded at seven downstream locations—enabling the comparison of up to 21 reach segments of  
142 different spatial scales, windows of detection, and advective times for each injection. We co-  
143 injected salt and fluorescein dye tracers for each injection, which also allowed us to test how  
144 tracer type and the associated data resolution influences interpreting solute transport.

145

## 146 **2. Background on the Challenge of Interpreting Physical Solute Transport Processes from** 147 **Stream Solute Tracers**

148         The stream solute transport timescale, defined here as the advective time and window of  
149 detection, can be directly estimated from an observed in-stream conservative solute tracer time  
150 series (hereafter breakthrough curve, or BTC, Figure 1a). Following an instantaneous tracer  
151 injection, the elapsed time from injection to peak concentration is commonly interpreted as the  
152 down-stream advective time ( $t_{ad}$ ) [e.g., Haggerty et al., 2002]. The window of detection ( $t_w$ ) of  
153 an observed BTC is often quantified as the elapsed time from tracer first arrival ( $t_1$ ) to 99% of  
154 recovered signal above background ( $t_{99}$ ) [e.g., Mason et al., 2012; Ward et al., 2013a]. The  
155 elapsed time from  $t_{ad}$  to  $t_{99}$  describes the persistence of tailing in an observed BTC, which is an  
156 indicator of transient storage (hereafter the transient storage index, or TSI) [Mason et al., 2012].  
157 The unique shape of a BTC is controlled by complex interactions of solute transport processes.  
158 For example, dispersive processes that contribute to spreading of the BTC include molecular  
159 diffusion and turbulent mixing [e.g., Fischer, 1975]. However, a BTC typically provides a reach-  
160 average representation of solute transport processes. Therefore, the processes most commonly

161 interpreted from the shape of a BTC are down-stream advection, longitudinal dispersion, and  
162 transient storage [e.g., Ward et al., 2013a].

163         The extent to which down-stream advection or processes of interest (i.e., dispersion and  
164 transient storage) control the shape of the BTC remains unclear as different combinations can  
165 result in similar shapes. At one extreme, if dispersion and transient storage are constant between  
166 high and low discharges in the same reach, differences between the BTCs are controlled solely  
167 by differences in advective times (Figure 1b). For example, during lower discharges that result in  
168 smaller advective velocities, there will be more time for dispersion and transient storage to act on  
169 the tracer. At the other extreme, if advective times are constant (e.g., discharges are equivalent)  
170 between two different reaches, variation in the shape of the BTC and, therefore, the transport  
171 timescale is controlled by differences in dispersion and transient storage (Figure 1d). In practice,  
172 in-stream tracer observations reflect interacting advective times with dispersive and transient  
173 storage processes, which can yield BTCs with similar shapes for different reasons (Figure 1c).  
174 Therefore, we anticipate that an understanding of the proportional controls of advective  
175 timescales relative to dispersion and transient storage is necessary to compare stream solute  
176 tracer studies across discharges, reaches, and spatial scales.

177

### 178 **3. Methods**

#### 179 **3.1. Site Description and Discharge Measurements**

180         The Selke River originates in the Harz Mountains in central Germany, flows through  
181 steep gradient, deeply incised valleys, and transitions to lower gradients underlain by alluvial  
182 deposits up to 10 m thick [Trauth et al., 2015] (Figure 2a). The study reach is a ~1-km section  
183 situated immediately downstream of this transition in largely agricultural lowlands

184 (51°43'21.4"N, 11°18'17.6"E). Continuous stream discharge was generated from stage—  
185 measured at 10-minute intervals (LTC Levellogger Junior M10, Solinst, Georgetown, Canada,  
186 accuracy +/- 0.1% of reading) about 15 m upstream of site E (Figure 2b)—applied to a site-  
187 specific power function rating curve developed by researchers at Helmholtz Centre for  
188 Environmental Research. The confidence bounds of this rating curve were produced following  
189 Schmadel et al. [2010] to provide the 95% confidence intervals of the discharge estimates (see  
190 Figure S1 in supporting information S1 for details). Discharge within the study reach increased  
191 from 0.27 to 2.35 m<sup>3</sup> s<sup>-1</sup> with stage remaining below bankfull during the July 2014 storm event  
192 used in this study (Figure 2c). The reach-specific rating curve was corroborated with continuous  
193 stream discharge recorded 4.6 km upstream at the Meisdorf federal gaging station. This station  
194 reports that discharge can range from 0.2 to 16 m<sup>3</sup> s<sup>-1</sup> during storm events and seasonal  
195 snowmelt, and the long-term (1921 to 2014) annual mean discharge is 1.5 m<sup>3</sup> s<sup>-1</sup>.

196         The study reach has general channel morphologic characteristics of meandering glides  
197 and shallow riffles, pool-riffle sequences, point bars, and in-stream gravel bars (Figure 2b). The  
198 streambed primarily consists of medium sand to coarse gravel. Visually different morphologic  
199 characteristics are expected to manifest as distinct solute transport controls across changing  
200 discharge conditions [e.g., Gostner et al., 2013]. Therefore, seven in-stream monitoring sites  
201 were selected within the study reach to bracket sections of visually distinct morphologies while  
202 preserving similar lengths. Labeled moving downstream from A to G (Figure 2b), these sites  
203 delineate the six individual reach segments (100 to 152 m). The three upstream most segments,  
204 AB, BC, and CD, consist primarily of meandering glides connected by shallow riffles visible  
205 under low-flow conditions. Moving downstream, the morphology becomes more complex:  
206 segment DE contains one in-stream bar, two point bars, and a pronounced pool-riffle sequence;

207 segment EF contains a pool-riffle sequence at the downstream end; and segment FG contains two  
208 pool-riffle sequences and a large in-stream and point bar. The segment-wise average streambed  
209 slopes are about 0.7, 0.4, 0.1, 0.4, 0.4 and 0.8% for segments AB, BC, CD, DE, EF, and FG,  
210 respectively, estimated from a 1-m digital elevation model acquired from the Land Survey  
211 Administration Saxony-Anhalt dated April 2009. The reach average streambed slope is 0.5% and  
212 active channel width is 9 m.

213 The individual segments can be combined into a total of 21 unique segments of different  
214 lengths (Figure 2d). These 21 combinations were used to test if combining shorter segments into  
215 longer reaches integrates differences in processes or only generates apparent differences due to  
216 changing advective timescales and windows of detection. Similar to the varying reach length  
217 analysis of Gooseff et al. [2013], we recombined individual segments into every possible unique  
218 combination based on available tracer data.

219

### 220 **3.2. In-Stream Solute Tracer Observations**

221 Seven instantaneous conservative solute tracer injections were performed over the storm  
222 discharge period (July 7-11, 2014). For all injections, dissolved salt (NaCl) and fluorescein dye  
223 tracers were co-injected at the same location and measured at sites A through G, resulting in 49  
224 observed BTCs each. In-stream responses of the salt tracer were observed at 30-second intervals  
225 (Schlumberger CTD diver 10m, Schlumberger Water Services, Delft, Netherlands, accuracy 1%  
226 of reading) using fluid specific conductance as a surrogate for salt concentration. Background  
227 specific conductance was corrected to zero. In some cases, the background specific conductance  
228 drifted during the measurement period, which was observed by an additional sensor placed about  
229 10 m upstream of the injection site. This background signal was shifted downstream based on the

230 advective time observed from the corresponding BTC and subtracted to correct for background  
 231 (hereafter salt BTC). Because instantaneous and constant rate tracer injection techniques are  
 232 expected to provide similar transit time distributions [Payn et al., 2008], instantaneous tracer  
 233 injections were chosen over constant rate injections to minimize the influence of dynamic  
 234 discharge conditions on measurements, similar to Ward et al. [2013a]. In-stream responses of the  
 235 fluorescein dye tracer (hereafter fluorescein BTC) were measured at 10-second intervals using  
 236 flow-through fluorimeters (Albilgia Sarl GGUN-FL30, Switzerland) at the same monitoring sites.  
 237 No background correction was performed for the fluorescein BTCs because the instruments were  
 238 calibrated in the field with stream water to reduce background interference and no dye was  
 239 present prior to injection.

240 The observed tracer BTC from these types of experiments is an incomplete temporal  
 241 signal due to detection limits and signal to noise interferences at late times [e.g., Drummond et  
 242 al., 2012], but can provide meaningful dispersion and short-term storage information [e.g., Bellin  
 243 et al., 2015; González-Pinzón et al., 2013]. In this study, we truncated each observed BTC to  $t_{99}$   
 244 to reduce the subjective selection of a response due to the injected tracer above background from  
 245 noise at late times, limiting interpretation to observed short-term storage [after Mason et al.,  
 246 2012; Ward et al., 2013a; Ward et al., 2013b]. An approximation to the window of detection ( $t_w$   
 247 =  $t_{99} - t_1$ ) was quantified by solving equation (1) for  $t_{99}$ ,

$$248 \quad 0.99 = \frac{\int_{t=t_1}^{t=t_{99}} C(t)dt}{\int_{t=t_1}^{t=t_{CLIP}} C(t)dt}, \quad (1)$$

249 where  $C$  is the observed, background corrected BTC ( $\text{g m}^{-3}$ ),  $t$  is the time after injection (s),  $t_1$  is  
 250 the time from injection to tracer first arrival (s), and  $t_{CLIP}$  is the time at which the observed BTC

251 was initially clipped based on visual inspection. All BTCs (i.e., salt and fluorescein) used  
 252 hereafter were truncated to  $t_{99}$ .

253 We compared the salt and fluorescein BTCs to examine how much temporal information,  
 254 such as  $t_w$ , varies given differing tracer sensitivities and resolutions. In this comparison, we  
 255 computed the associated temporal metrics (including those in Figure 1a and statistical moments  
 256 as described in section 3.4.1) and examined the slope of the trend-line constructed through linear  
 257 regression between those metrics corresponding to the fluorescein and salt BTCs. If a slope of  
 258 unity and intercept of zero were within the associated 95% t-based confidence intervals (i.e., the  
 259 ratio of fluorescein metrics to salt metrics is 1:1), the two BTCs were considered to provide  
 260 similar temporal information. For this trend-line and those presented hereafter, we also tabulated  
 261 the coefficient of determination ( $R^2$ ).

262

### 263 **3.3. Net Change in Discharge and Unrecovered Tracer Mass**

264 Sections of the study reach have been reported as losing during baseflow conditions, but  
 265 the hydraulic gradients between the stream and adjacent alluvial aquifer may alternate seasonally  
 266 [Schmidt et al., 2012]. We used the salt BTCs and rating curve estimates to examine the general  
 267 pattern of net changes to discharge and mass losses through the study reach during the storm  
 268 discharge period. Specifically, we estimated discharge at site A via dilution gaging [after  
 269 Kilpatrick and Cobb, 1985],

$$270 \quad Q = \frac{M}{\int_{t=t_1}^{t=t_{99}} C(t) dt}, \quad (2)$$

271 where  $Q$  is the stream discharge ( $m^3 s^{-1}$ ) and  $M$  is the tracer mass injected (g) which ranged from  
 272 9 to 16 kg of NaCl in this study. The section of the study reach between the injection and site A  
 273 was treated as a mixing length where we assume no tracer mass was lost. If some mass was lost

274 over this section,  $Q$  will be slightly overestimated. We quantified the net change in discharge  
 275 (downstream minus upstream) using estimates at sites A (dilution gaging) and E (rating curve),  
 276 normalized by the upstream discharge to express as percent change. Assuming that the error in  
 277 dilution gaging is roughly 8% [after Schmadel et al., 2010] and using the 95% confidence  
 278 intervals of the site E discharge estimates (Figure 2c), the 95% confidence intervals of these net  
 279 changes were approximated. A net change in discharge is considered significant if zero is outside  
 280 of these intervals. Likewise, tracer mass recoveries were estimated from the discharge estimates  
 281 near site E,

$$282 \quad M_R = Q_D \int_{t=t_1}^{t=t_{99}} C_D(t) dt, \quad (3)$$

283 where  $M_R$  is the tracer mass recovered (g),  $Q_D$  is the downstream stream discharge (i.e., site E)  
 284 ( $\text{m}^3 \text{s}^{-1}$ ), and  $C_D$  is the observed tracer concentration at site E from the tracer released at the  
 285 injection site ( $\text{g m}^{-3}$ ). The percent unrecovered mass was estimated as the difference  $M_R - M$   
 286 normalized by  $M$ . Again, because 95% confidence intervals were approximated for discharge  
 287 estimates, we estimated the confidence intervals of unrecovered tracer mass. These confidence  
 288 interval estimates do not include uncertainty due to  $\int C_D(t) dt$  in equation (3). Although this  
 289 uncertainty is expected to be small relative to the uncertainty in  $Q_D$ , a potential artifact of  
 290 objectively truncating the BTC to  $t_{99}$  is artificially low  $M_R$  estimates.

291

### 292 **3.4. Temporal Metrics**

293 We estimated temporal metrics (including those in Figure 1a and statistical moments as  
 294 described below) of each fluorescein and salt BTC to compare field-based tracer observations  
 295 and interpret relationships between solute transport processes and discharge. Following this

296 reach scale analysis, we used the temporal metrics for each individual reach segment to examine  
 297 how segment-wise processes changed with discharge. Lastly, we used the same metrics of each  
 298 segment combination (Figure 2d) to test whether combining shorter segments into longer reaches  
 299 influenced the interpretation of how solute transport processes change with discharge.

300

### 301 **3.4.1. Relationships between Reach Scale Temporal Metrics and Discharge**

302 Using the  $t_{ad}$  and  $t_{99}$  estimates, TSI was estimated (Figure 1a). To compare metrics across  
 303 discharges,  $t_w$  and TSI were normalized by  $t_{ad}$  to remove the variation in the BTC caused by  
 304 advection [after Gooseff et al., 2007; Ward et al., 2013a],

$$305 \quad t_{w,norm} = t_w/t_{ad}, \quad (4)$$

$$306 \quad TSI_{norm} = TSI/t_{ad}, \quad (5)$$

307 where  $t_{w,norm}$  defines the window of detection relative to advective time and  $TSI_{norm}$  reflects the  
 308 persistence of tailing relative to advective time. These metrics provide the overall influences of  
 309 dispersion and short-term storage on the shape of the BTC independent of advective time  
 310 assuming that  $t_w$  and TSI are linearly related to  $t_{ad}$ . While this assumption is appropriate in  
 311 advective-dominated stream systems [Gooseff et al., 2007; Ward et al., 2013a], other types of  
 312 normalization may be necessary to better account for potential nonlinearities [e.g., Gelhar et al.,  
 313 1992]. Each observed BTC was normalized to express only the available temporal signature,

$$314 \quad c(t) = \frac{C(t)}{\int_{t=t_1}^{t=t_{99}} C(t)dt}, \quad (6)$$

315 where  $c$  reflects the recovered transit time distribution and the integral of  $C(t)$  with respect to  
 316 time represents the zeroth temporal moment. We estimated the statistical moments of the transit

317 time distributions by empirically calculating temporal moments of the recovered BTCs.  
 318 Specifically, we estimated the first temporal moment ( $M_1$ ),

$$319 \quad M_1 = \int_{t=t_1}^{t=t_{99}} tc(t)dt, \quad (7)$$

320 which provides an estimate of mean arrival time. The  $n$ th-order temporal moment centered about  
 321  $M_1$  (hereafter central moment) was estimated as

$$322 \quad \mu_n = \int_{t=t_1}^{t=t_{99}} (t - M_1)^n c(t)dt \quad \text{for } n > 1. \quad (8)$$

323 The second central moment ( $\mu_2$ ) provides the temporal variance, a metric of symmetrical  
 324 spreading; the third central moment ( $\mu_3$ ) reflects the temporal extent of late-time tailing. We  
 325 computed the normalized metrics of the coefficient of variation (CV) and skewness ( $\gamma$ ) to further  
 326 compare BTCs across discharges,

$$327 \quad CV = \frac{\mu_2^{1/2}}{M_1}, \quad (9)$$

$$328 \quad \gamma = \frac{\mu_3}{\mu_2^{3/2}}, \quad (10)$$

329 where CV expresses the rate of symmetrical spreading relative to mean arrival time and  $\gamma$  reflects  
 330 the extent of late-time tailing relative to symmetrical spreading (see Table S1 in supporting  
 331 information S1 for a summary of all temporal metrics used). Therefore, these metrics should  
 332 provide an indication of how dispersion and short-term storage processes change with discharge,  
 333 respectively.

334 Relationships between discharge and the normalized reach scale metrics ( $t_{w,norm}$ ,  $TSI_{norm}$ ,  
 335 CV, and  $\gamma$ ) for each monitoring site were constructed through linear regression. The discharges  
 336 used for this relationship were those estimated from the rating curve near site E. Significant  
 337 slopes were those whose 95% t-based confidence interval did not include zero.

### 338 3.4.2. Relationships between Segment-Wise Temporal Metrics, Discharge, and Length

339 The transfer function,  $g(t)$ , reflects the unique temporal signature of a solute traveling  
 340 through a given reach segment independent of the input signal. When a single tracer injection is  
 341 observed at different locations within a reach, the transfer function can be related to the upstream  
 342 and downstream transit time distributions by convolution, assuming linear, time-invariant  
 343 transport from one segment to the next,

$$344 \quad c_{ds}(t) = \int_{t=t_1}^{t=t_{99}} g(\tau) c_{us}(t - \tau) d\tau, \quad (11)$$

345 where  $c_{ds}$  is the observed transit time distribution at a downstream monitoring site and  $c_{us}$  is the  
 346 observed transit time distribution at the upstream monitoring site (input signal). To isolate  
 347 segment-wise responses to discharge, we quantified metrics of the temporal characteristics of  
 348 segment-specific transfer functions [after Ward et al., 2016] rather than apply a more  
 349 complicated deconvolution scheme [e.g., Cirpka et al., 2007].

350 We estimated the change in window of detection between consecutive downstream and  
 351 upstream BTCs ( $\Delta t_w$ ) as an estimate of the transfer function window of detection,

$$352 \quad \Delta t_w = t_{w,ds} - t_{w,us}. \quad (12)$$

353 To allow for comparison between segments, segment lengths, and discharges, we normalized this  
 354 metric by the change in the advective time,

$$355 \quad \Delta t_{w,norm} = \Delta t_w / \Delta t_{ad}, \quad (13)$$

356 where  $\Delta t_{ad} = t_{peak,ds} - t_{peak,us}$ , and  $t_{peak,ds}$  and  $t_{peak,us}$  are the times to peak concentrations of the  
 357 downstream and upstream BTCs, respectively. Next, we estimated the change in TSI and  
 358 normalized by the corresponding change in advective time,

$$359 \quad \Delta TSI_{norm} = (TSI_{ds} - TSI_{us}) / \Delta t_{ad}. \quad (14)$$

360 We used the change in  $t_{ad}$ ,  $t_w$ , and TSI between downstream and upstream BTCs as surrogates to  
 361 the corresponding metrics of the actual transfer function. Because the segment-wise temporal  
 362 moments are theoretically linearly additive to produce the overall reach estimate [e.g., Riml and  
 363 Wörman, 2011], and have been interpreted as such in existing field studies [Ward et al., 2014],  
 364 we isolated statistical moments of the transfer function while eliminating the need to solve  
 365 equation (11),

$$366 \quad \Delta M_1 = M_{1,ds} - M_{1,us}, \quad (15)$$

$$367 \quad \Delta \mu_n = \mu_{n,ds} - \mu_{n,us} \text{ for } n > 1. \quad (16)$$

368 Again, to compare between segments, segment lengths, and discharges, we computed the  
 369 coefficient of variation and skewness of the transfer function,

$$370 \quad \Delta CV = \frac{\Delta \mu_2^{1/2}}{\Delta M_1}, \quad (17)$$

$$371 \quad \Delta \gamma = \frac{\Delta \mu_3}{\Delta \mu_2^{3/2}}. \quad (18)$$

372 We used the normalized segment metrics ( $\Delta t_{w,norm}$ ,  $\Delta TSI_{norm}$ ,  $\Delta CV$ , and  $\Delta \gamma$ ) to investigate (1)  
 373 relationships with discharge, (2) the spatial variability of segment response to discharge, and (3)  
 374 whether estimates of these metrics were dependent on segment length.

375 Relationships with discharge were investigated by constructing linear trend-lines between  
 376 the normalized metrics calculated for each segment (and combination of segments) and discharge  
 377 near site E. Discharge was anticipated to vary slightly between segments (i.e., potentially a net  
 378 losing stream), but if a relationship is present, discharge near site E will suffice because the large  
 379 magnitude of change in discharge due to the storm event overwhelms the potential spatial change  
 380 over the study reach. The metrics estimated for each of the seven discharges that reflect a single  
 381 segment provided the sample population for the associated trend-line. Additionally, we pooled

382 all normalized segment-wise metrics and examined the corresponding slope with discharge to  
383 provide an indication of the overall stream system response to changing discharge.

384 Spatial variability between segment responses to discharge was examined by comparing  
385 each metric ( $\Delta t_{w,norm}$ ,  $\Delta TSI_{norm}$ ,  $\Delta CV$ , and  $\Delta \gamma$ ) sample population that reflected the individual  
386 segments (i.e., not the combined segments). This comparison was completed through a  
387 parametric test of comparing means in a one-way analysis of variance (ANOVA). We  
388 additionally completed a nonparametric test of comparing medians in a Kruskal-Wallis one-way  
389 ANOVA. The nonparametric test does not require the assumption of normally distributed  
390 residuals like the parametric version, so it provides additional information to prevent over-  
391 interpreting results from small sample sizes where the parent distribution is unknown. A 95%  
392 confidence level where  $p < 0.05$  indicates means (denoted by  $p_{ANOVA}$ ) or medians (denoted by  
393  $p_{KW}$ ) are different and segments are a significant source of variability. While these p-values  
394 indicate whether means and medians of at least one segment are different from all of the others,  
395 these tests were also repeated to compare each segment distribution to another on a paired basis.  
396 We selected these statistical tests because they are among the simplest to apply, acknowledging  
397 that the small sample size limits the ability to identify underlying distributions or satisfy the  
398 assumptions of additional statistical tests.

399 To determine how the normalized transfer function metrics ( $\Delta t_{w,norm}$ ,  $\Delta TSI_{norm}$ ,  $\Delta CV$ , and  
400  $\Delta \gamma$ ) depended on segment length, we grouped corresponding estimates according to the number  
401 of combined segments. Each group provided the sample populations that were compared using  
402 the parametric and nonparametric tests outlined above. We only considered the sample  
403 populations for one, two, and three combined reach segments (Figure 2d) that contained more  
404 than 20 values representing different spatial scales and discharges to reduce over-interpretation

405 of results. In this case, a  $p_{ANOVA}$  and  $p_{KW}$  of less than 0.05 indicates that the normalized transfer  
406 function metrics are dependent on the transport timescale (i.e.,  $\Delta t_{ad}$  and  $\Delta t_w$ ) due to reach length  
407 selection. Hence, if this condition is recognized, interpreting how solute transport processes  
408 change with discharge is influenced by reach length selection.

409

## 410 **4. Results**

### 411 **4.1. In-Stream Solute Tracer Observations**

412 The influence of time-varying discharge conditions on the observed BTCs was assumed  
413 negligible due to the relatively short measurement periods. For the seven tracer injections  
414 measured at sites A through G, the duration of each corresponding measurement period from  
415 tracer injection to last detection at G ranged from 40-150 minutes (see Figure S2 in supporting  
416 information S1 for all observed BTCs). The possible change in discharge was between 1% and  
417 6% over these measurement periods (red lines in Figure 2c).

418 Through the storm discharge period, some of the salt and fluorescein BTCs (49 each  
419 possible) were deemed erroneous based on initial quality control (e.g., debris potentially  
420 blocking the sensor, power failure, or data with visually high noise compared to other  
421 observations). This quality control resulted in a total of 29 co-located salt and fluorescein BTCs.  
422 In addition to these co-located BTCs, there were 12 instances where only salt BTCs were  
423 obtained and 4 instances where only fluorescein BTCs were obtained. A total of 5 fluorescein  
424 BTCs were omitted based on visually high noise. Data availability is summarized in Figure S2  
425 and Table S2 in supporting information S1.

426 The co-located salt and fluorescein BTCs provided similar reach scale temporal metrics  
427 ( $t_w$ , TSI,  $M_1$ ,  $\mu_2$ , and  $\mu_3$ ) (Figure 3, left column). All the slopes between these metrics estimated

428 from salt BTCs and those estimated from fluorescein BTCs were statistically the same as 1 (i.e.,  
429 1 falls within the associated 95% confidence interval). The intercepts were also statistically the  
430 same as zero for all these metrics. The fluorescein tracer produced longer detectable tailing,  
431 resulting, mostly, in higher  $t_w$  and TSI estimates (above the 1:1 line in Figure 3, left column). The  
432 higher order central moment estimates began to deviate further from the 1:1 line at higher values  
433 (i.e., those associated lower discharges). All slopes and intercepts of the normalized metrics  
434 ( $t_{w,norm}$ ,  $TSI_{norm}$ , CV, and  $\gamma$ ) were statistically less than 1 and greater than zero, respectively  
435 (Figure 3, right column). However, the largest average difference between these metrics  
436 estimated from salt and fluorescein BTCs was ~10%, indicating that both tracers provide similar  
437 temporal metric estimates following objective truncation to  $t_{99}$ .

438         As both tracers provided similar temporal metric estimates yet incomplete datasets, we  
439 used fluorescein BTCs to fill in gaps in the salt BTC dataset, resulting in a total of 45 BTCs (41  
440 salt and 4 fluorescein) to allow for greater coverage of advective times and windows of  
441 detection, and thus a more complete analysis. Fluorescein BTCs were selected to fill in gaps in  
442 the salt BTC dataset, and not vice versa, because there were more salt BTCs available and salt is  
443 more readily available and commonly used in stream tracer studies.

444

#### 445 **4.2. Net Change in Discharge and Unrecovered Tracer Mass**

446         The study reach was generally net losing through the storm discharge period (Figure 4a).  
447 The net change in discharge from site A to E ranged from -15 to -20% and was significant for  
448 every injection with the exception of injection 3 (where the discharge estimate at site E was  
449 highest) and injection 1 (where there was no salt BTC available at site A). A similar pattern  
450 occurred for the unrecovered tracer mass estimates (Figure 4b). For injections 4 to 7,

451 unrecovered mass was significant and ranged from -15 to -25%, providing further evidence that  
452 this study reach was a net losing stream.

453

### 454 **4.3. Relationships between Reach Scale Temporal Metrics and Discharge**

455       Regardless of the tracer type, the normalized metrics were mostly insensitive to changing  
456 discharge where statistically significant relationships were the exception rather than the norm.  
457 The reach scale metrics of the transit time distributions ( $t_w$ , TSI,  $M_1$ ,  $\mu_2$ , and  $\mu_3$ ) all decreased  
458 with increasing discharge (Figure 5, left column, estimated from 41 salt and 4 fluorescein BTCs).  
459 There was an increasing trend between these metrics and downstream location (bottom-to-top  
460 order of lines in the left column of Figure 5, representing increased length). The normalized  
461 metrics ( $t_{w,norm}$ ,  $TSI_{norm}$ , CV, and  $\gamma$ ) along the study reach were not significantly related to  
462 discharge in most cases (Figure 5, right column). The exceptions are that CV was significantly  
463 negatively related to discharge only at sites C and D and  $\gamma$  was significantly positively related to  
464 discharge at site A (Table 1). Repeated analyses using only the salt BTCs and using only the  
465 fluorescein BTCs produced similar results that most relationships with discharge were not  
466 significant (see Figures S3 and S4 and Tables S3 and S4 in supporting information S1 for the  
467 results of these analyses).

468

### 469 **4.4. Relationships between Segment-Wise Temporal Metrics, Discharge, and Length**

#### 470 **4.4.1. Segment-Wise Responses to Changing Discharge**

471       The normalized transfer function metrics ( $\Delta t_{w,norm}$ ,  $\Delta TSI_{norm}$ ,  $\Delta CV$ , and  $\Delta \gamma$ ) indicated that  
472 some reach segments were more sensitive to changing discharge than others where more  
473 variability in the distribution is interpreted as a higher sensitivity (Figure 6). Segment EF was

474 consistently the most sensitive to discharge based on the interquartile range. For  $\Delta t_{w,norm}$ , there  
475 was no significant difference between the segment-wise means and medians ( $p_{ANOVA} = 0.30$  and  
476  $p_{KW} = 0.12$ ). For  $\Delta TSI_{norm}$ , the means were statistically the same ( $p_{ANOVA} = 0.30$ ), but there was  
477 evidence that the medians were different ( $p_{KW} = 0.04$ ). The segment-wise means and medians of  
478  $\Delta\gamma$  were statistically the same ( $p_{ANOVA} = 0.11$  and  $p_{KW} = 0.06$ ). Note that some negative  $\Delta\gamma$   
479 estimates resulted from isolating the central moments of the transfer functions (equation (16)).  
480 The variation of  $\Delta CV$  means and medians between segments was significant ( $p_{ANOVA} < 0.001$   
481 and  $p_{KW} = 0.01$ ). When each segment distribution was compared to another on a paired basis,  
482  $\Delta t_{w,norm}$  was different between segments AB and BC, BC and CD, and BC and DE (also  
483 visualized by non-overlapping notches in Figure 6);  $\Delta TSI_{norm}$  of segment AB was different from  
484 segments BC, CD, and DE. The mean and median  $\Delta CV$  of segment AB was different from all  
485 other segments. The mean and median  $\Delta CV$  between all segments other than AB were the same.  
486 The means and medians of  $\Delta\gamma$  were statistically the same between each segment with the  
487 exception of segments AB and FG. See Table S5 in supporting information S1 for p-values of all  
488 these pairwise comparisons.

489 An apparent relationship between the normalized transfer function metrics and discharge  
490 was not clearly recognized. The slopes of the relationship between  $\Delta t_{w,norm}$  and discharge for  
491 each segment and combination were generally negative (15 out of 21 segments; see Table S6 in  
492 supporting information S1 for slopes and  $R^2$  values). However, the slope between this metric and  
493 discharge was only significant for segments AC and AD. The slopes of the relationship between  
494  $\Delta TSI_{norm}$  and discharge were generally negative (15 out of 21 segments), but none were  
495 significant. The slopes of the relationship between  $\Delta CV$  and discharge were also generally  
496 negative (19 of 21 segments) with only two significant within segments AD and BD (negative

497 slopes). A relationship between  $\Delta\gamma$  and discharge was unclear, with variable slopes (9 out of 21  
498 segments with positive slopes, 12 out of 21 with negative slopes) that were not significant.  
499 Pooling all 21 segments indicated that  $\Delta t_{w,norm}$ ,  $\Delta TSI_{norm}$ , and  $\Delta CV$  overall had a slight downward  
500 trend with discharge while the relationship between  $\Delta\gamma$  and discharge was not significant (see  
501 Table S6 in supporting information S1 for slopes).

502

#### 503 **4.4.2. Segment-Wise Metrics Dependence on Length**

504 We found a general pattern of decreasing variability in the transfer function metric  
505 estimates for longer segments (Figure 7). This decrease was due to both increased averaging of  
506 the heterogeneity in longer segments and an overall decrease in the number of samples available  
507 (i.e., where fewer extremes in the underlying distribution are reflected in longer segments). We  
508 statistically compared only the distributions associated with one to three combined segments  
509 where the numbers of individual observations in the sample populations were greater than 20.  
510 The means and medians of the  $\Delta t_{w,norm}$  and  $\Delta TSI_{norm}$  distributions did not significantly change  
511 when reach segments were combined (i.e.,  $p_{ANOVA} = 0.53$  and  $0.44$  and  $p_{KW} = 0.76$  and  $0.96$ ,  
512 respectively). The means and medians of the  $\Delta CV$  distributions were significantly different (i.e.,  
513  $p_{ANOVA}$  and  $p_{KW} < 0.001$ ). By combining segments in this study reach under the seven discharge  
514 conditions, the mean and median of  $\Delta CV$  significantly increased over longer lengths. While the  
515 medians of  $\Delta\gamma$  distributions decreased slightly over longer lengths ( $p_{KW} = 0.04$ ), the means were  
516 statistically the same ( $p_{ANOVA} = 0.81$ ).

517

518

519

## 520 5. Discussion

### 521 5.1. A Changing Transport Timescale Complicates Comparison of Stream Solute Tracer 522 Studies

523 The extent to which dispersive and transient storage processes change with discharge can  
524 be easily misinterpreted from in-stream solute tracer tests based on different transport timescales  
525 (defined as the advective time and window of detection). For example, similar to findings from  
526 other tracer studies [e.g., Schmid et al., 2010], magnitudes of several BTC metrics decreased  
527 with increasing discharge (Figure 5, left column). A possible interpretation of this result is that  
528 physical processes have changed (e.g., activation or deactivation of storage flowpaths), which is  
529 to be expected as varying discharge alters the turbulent energy, wetted geometry, hydraulic  
530 gradients, and connections to storage flowpaths [Leopold and Maddock, 1953]. However,  
531 through the suite of transport timescales of this study, the normalized temporal metrics ( $t_{w,norm}$ ,  
532  $TSI_{norm}$ , CV and,  $\gamma$ ) were mostly insensitive to changes in discharge (Figure 5, right column and  
533 Table 1). This lack of a strong relationship indicates that determining how differences in  
534 advective times and windows of detection control the shape of the BTC is needed when  
535 interpreting changes in physical solute transport processes. However, we recommend that similar  
536 tracer tests should be repeated across more reaches of different stream systems and discharge  
537 conditions to test the robustness of this conclusion.

538 We anticipate general changes to some transport processes due to varying discharge  
539 based on previous studies. The relative influence of dispersive processes can rise with discharge  
540 due to increased turbulent mixing [D'Angelo et al., 1993]. Likewise, we expect CV to increase  
541 with discharge. Subsurface short-term storage volume can increase with discharge in stream  
542 systems with relatively coarse (e.g., sand to gravel) streambeds [Dudley-Southern and Binley,

543 2015; Schmid et al., 2010; Zimmer and Lautz, 2014], which would likely cause an increase in  
544 prolonged tracer tailing or  $\gamma$ . Conversely, streams with low subsurface short-term storage  
545 potential (i.e., low streambed hydraulic conductivity and valley slope) are less likely to undergo  
546 substantial storage flowpath changes with discharge [Wondzell, 2011]. In this latter case, an  
547 increase in stream discharge should result in reduced tailing and corresponding decrease in  $\gamma$   
548 because transport of the tracer is expected to be more sensitive to advective timescales than  
549 changing subsurface short-term storage processes. Similarly, based on a fluid-mechanics  
550 classification scheme after Jackson et al. [2013], we expect surface short-term storage to  
551 generally increase with discharge. The Selke study reach is not constrained by its valley and is  
552 set in highly permeable gravels [Schmidt et al., 2012]; therefore, we expected the reach-scale  
553 extent of short-term storage to increase with discharge, which would likely cause an increase in  
554  $\gamma$ . Contrary to expectation, our estimates of CV were mostly insensitive to changes in discharge  
555 with some evidence of a negative relationship with discharge (sites C and D, Table 1). Although  
556  $\gamma$  was mostly insensitive to changes in discharge, there was some indication of a positive  
557 relationship at site A, suggesting that the extent of storage increased with discharge. The lack of  
558 a clear relationship with discharge indicates that these metrics either scaled directly with the  
559 transport timescale or short-term storage and dispersion did not change appreciably despite the  
560 order of magnitude change in observed discharge. Direct scaling with the transport timescale is  
561 the more likely conclusion because a negative response in CV suggests that this metric is still  
562 influenced by changing transport timescales rather than changes in dispersive processes.

563 As expected, the advective time of the transfer function was inversely proportional to  
564 discharge and directly proportional to reach length—the longest advective time occurred at the  
565 lowest discharge and longest segment length (Figure 8a). Similarly, the window of detection

566 expanded with an increase in advective time (Figure 8b). Theoretical temporal moments (i.e.,  
567 those derived from commonly used transient storage models) suggest that both the coefficient of  
568 variation and skewness decrease nonlinearly with increasing reach length (and advective time) as  
569 the predicted BTC becomes more symmetrical [Schmid, 2002]. Contrary to theory, we observed  
570 an increasing trend in  $\Delta CV$  with advective time (Figure 8c), which corresponds to an increase in  
571 reach length (Figure 7). Consistent with theory,  $\Delta \gamma$  did have a slight decreasing trend with  
572 advective time (Figure 8d), but did not show a significant change with increased reach length  
573 (Figure 7). While Ward et al. [2013a] found that comparison of tracer observations across  
574 different discharges and reaches requires normalizing by associated advective times, the  
575 increasing trend in the coefficient of variation indicates that an expanding window of detection  
576 associated with longer advective times still limits direct comparison. A larger window of  
577 detection allows more time for processes like dispersion and short-term storage to act on the  
578 tracer. Consequently, tracer observations made at short advective times are more sensitive to the  
579 influence of truncation, or lack of an ability to measure the entire tail, than those made at longer  
580 advective times. An estimate of the coefficient of variation is clearly more sensitive to truncation  
581 than an estimate of skewness. This issue of truncation may also partially explain why dispersion  
582 in aquifers assessed from tracers typically increases with the spatial scale of observations [e.g.,  
583 Gelhar et al., 1992]. Our results provide evidence that when tracer observations are compared,  
584 the influence of differences in the advective time and window of detection must be explicitly  
585 considered to prevent misinterpreting changes in solute transport processes. However, any  
586 change in the coefficient of variation or skewness at a given advective time is likely due to the  
587 influence of heterogeneity in channel morphology and process change associated with discharge.

588           In addition, other sources of uncertainty may restrict comparisons of tracer observations  
589 between different discharge conditions. A potential source of uncertainty is that short-term  
590 storage flowpaths may change differently on the rising and falling limbs of the hydrograph [e.g.,  
591 Ward et al., 2013a; Zimmer and Lautz, 2014]. For example, storage flowpaths within the Selke  
592 study reach might be activated during the rising limb and subsequently deactivated during the  
593 falling limb. To better understand this influence, more tracer tests along the rising limb would be  
594 needed (Figure 2c). Furthermore, segment-wise tracer mass loss, as the study reach is anticipated  
595 to be net losing (Figure 4b), might change with discharge. Fortunately, if the stream only loses  
596 water without having significant gains, mass losses will not have substantially influenced the  
597 temporal information contained in the BTC because in-stream concentrations remain unchanged.  
598 Testing this mass loss influence would require tracer injections at the limits of each segment  
599 (such as those by Payn et al. [2009]). Considerations of dynamic mixing length and data  
600 resolution would be necessary to perform finer scale dilution gaging and mass recovery  
601 techniques during high discharges.

602

## 603 **5.2. Sensitivity of Tracer Detection is Less Important at High Discharges**

604           Detection of fluorescein was slightly more sensitive than detection of salt in late-time  
605 tailing based on generally larger windows of detection ( $t_w$ ). Accordingly, the higher order  
606 moments ( $\mu_2$  and  $\mu_3$ ) were generally larger for the fluorescein tracer (Figure 3, left column).  
607 These differences also caused the normalized metrics ( $t_{w,norm}$ ,  $TSI_{norm}$ , CV and  $\gamma$ ) to be generally  
608 larger than those estimated from the salt tracer. Despite these differences, however, similar  
609 conclusions were drawn from both tracers: changes to the transport timescales confound the  
610 ability to observe changes in other processes because deviations between the metrics of the two

611 different tracer types were small relative to changes in advective time. The metrics provided by  
612 the fluorescein and salt tracers were in close agreement at high discharges, indicating that the  
613 ability to detect short-term storage processes through tracer tests is limited at high discharges  
614 regardless of the tracer type selected.

615         An artifact of incomplete tracer response due to detection limits or signal to noise  
616 interferences in late-time tailing is artificially low temporal moment estimates [e.g., Drummond  
617 et al., 2012]. Consequently, objective truncation limits this study to observable short-term  
618 storage and may partially explain the lack of a robust trend between  $\Delta CV$  and  $\Delta \gamma$  and discharge.  
619 Although we expect dispersion and short-term storage to change with discharge, this truncation  
620 may have reduced the influence of these processes relative to advection in the observations. One  
621 potential way to address this late-time detection issue is to apply statistical techniques to better  
622 isolate the tracer signal from noise [e.g., Aquino et al., 2015; Drummond et al., 2012]. Other  
623 sensing techniques and tracers may also help address this truncation issue. For example, highly  
624 sensitive tracers (e.g., synthetic DNA (deoxyribonucleic acid) molecules) have been shown to  
625 reflect much longer residence time storage flowpaths [Foppen et al., 2013; Foppen et al., 2011].  
626 The use of such tracers thereby allows for a better determination of which portion of mass loss  
627 may be long-term storage and, in turn, which portion is groundwater recharge. Otherwise,  
628 modeling techniques to approximate the late-time tail may be necessary to provide a more  
629 complete representation of the actual transit time distribution and support a better understanding  
630 of relationships between short- and long-term storage processes and discharge [e.g., Drummond  
631 et al., 2012; Stonedahl et al., 2012].

632

633

### 634 **5.3. Temporal Metrics Do Not Reflect Observed Morphologic Patterns**

635 Heterogeneity in stream morphologic characteristics is expected to cause differences in  
636 short-term storage [e.g., Gostner et al., 2013; Wondzell and Gooseff, 2013]. Likewise, this  
637 heterogeneity can cause dispersive processes to respond differently to changes in discharge. For  
638 example, increased turbulence due to higher discharge over features like riffles can increase  
639 dispersion, but higher discharge through slower moving sections like pools can reduce dispersion  
640 [Dyer and Thoms, 2006]. Based on this understanding, dispersion in sections with riffles (e.g.,  
641 segment BC, Figure 2b) should increase with discharge compared to sections with pool-riffle  
642 sequences (e.g., segment DE, Figure 2b). The response of the window of detection ( $\Delta t_{w,norm}$ ) to  
643 changing discharge was different between segments BC and DE where DE was consistently  
644 higher (Figure 6 and Table S5 in supporting information S1), which corresponds to the more  
645 expected spreading of the tracer over riffle features. However, the symmetrical spreading relative  
646 to mean arrival time ( $\Delta CV$ ) between segments BC and DE did not respond significantly  
647 differently to changing discharge, suggesting that the influence of spatially variable morphology  
648 was muted relative to changing advective timescales. The estimates of  $\Delta CV$  within segment AB  
649 were consistently the largest, the least sensitive to changing discharge, and different from all  
650 other reach segments (Figure 6 and Table S5 in supporting information S1). Segment AB is a  
651 glide section with the steepest streambed slope (0.7%). We believe that the relatively uniform  
652 straight character and steep gradient of this segment (Figure 2b) caused the highest symmetrical  
653 spreading relative to mean arrival time while remaining the least sensitive to changing turbulent  
654 energy. Still, the means and medians of the normalized segment metrics were statistically  
655 equivalent between most segments, suggesting that the influence of segment spatial  
656 heterogeneity was muted or obscured by changes to transport timescales inherent in tracer

657 observations. Only the most drastic differences in morphology appear to be reflected by tracer  
658 observations in our study.

659

#### 660 **5.4. Reach Scale Tracer Study Limitations May Complicate Upscaling**

661       Appropriately representing the heterogeneity and dynamics of solute transport processes  
662 across a network is critical to predict solute transport to receiving water bodies [e.g., Kiel and  
663 Cardenas, 2014; Wollheim et al., 2008]. The approach of many previous studies to examine  
664 solute transport processes throughout a network has been to conduct solute tracer tests over fixed  
665 reaches of equal length regardless of differences in transport timescales or morphologic  
666 characteristics [e.g., Covino et al., 2011; Gooseff et al., 2013; Stewart et al., 2011], with some  
667 calling for a length-normalized metric to compare processes between reaches or discharges  
668 [Runkel, 2002]. This study indicates that a change in the transport timescale (i.e., set by the  
669 interacting advective time and window of detection) resulting from a change in discharge or  
670 reach length selection can artificially manifest as spatial heterogeneity or process change.  
671 Therefore, we argue that fixing reaches of equal length or using length-normalized metrics to  
672 compare stream solute tracer observations may not be the best approach. Rather, since we  
673 observed variability between segment morphologies and combining segments influences the  
674 interpretation of some transport processes, setting reach lengths and monitoring sites for tracer  
675 experiments should focus more on an understanding of transport timescales. One possible  
676 approach would be to fix the advective timescale between reaches, adjusting reach length  
677 accordingly. Ultimately, design of comparable tracer experiments may require preliminary  
678 studies or modeling efforts to initially quantify the range of transport timescales. Only  
679 comparable tracer observations will clarify relationships between solute transport processes and

680 discharge. Improving the ability to compare tracer observations will likely lead to more accurate  
681 process-discharge relationships.

682         Spatial patterns of stream morphology (e.g., streambed slope and channel width)  
683 throughout a network may also be an important consideration when representing solute transport  
684 processes within a network model [e.g., Zarnetske et al., 2007]. Fortunately, it may be possible  
685 to represent the heterogeneity of advective velocity at long reach scales by identifying the spatial  
686 pattern of the stream channel from imagery [Schmadel et al., 2014b]. Advances in upscaling  
687 from reach-based studies to networks will require consideration of within-reach spatial  
688 heterogeneity, spatial patterns in discharge and advective velocity, more sophisticated techniques  
689 to map the inter-connected surface and subsurface waters, and a comprehensive analysis of tracer  
690 tests conducted over a range of discharges, spatial scales, and geologic settings. For example,  
691 tracer studies could be paired with independent and complementary measures of transport, such  
692 as through geophysics [e.g., Toran et al., 2013; Ward et al., 2012] or well networks in the  
693 adjacent aquifer [e.g., Voltz et al., 2013; Zarnetske et al., 2011], to better develop relationships  
694 between transient storage processes and discharge. However, from this study, one clear way  
695 forward is to better acknowledge and quantify the influence of changing advective timescales  
696 and windows of detection on the interpretation of processes such as dispersion and transient  
697 storage in our future reach scale solute tracer studies.

698

## 699 **6. Conclusions**

700         Using a suite of transport timescales (advective times and windows of detection) reflected  
701 in conservative stream solute tracer responses observed during a storm discharge period, we  
702 demonstrate that changes in advective times dominate and, consequently, mask other transport

703 processes like dispersion and transient storage. While a possible interpretation of the tracer data  
704 could be that solute transport processes changed with discharge, we found through further  
705 analysis that the differences in the tracer data were generated primarily by variation in advective  
706 times. Furthermore, most individual segments within the original study reach did not respond  
707 differently to changes in discharge, suggesting that the influence of distinct channel  
708 morphologies was muted by the differences in advective times between tracer tests. Only the  
709 largest differences in morphologies were reflected in the tracer observations. We also found that  
710 through combining segments into longer reaches, differences of transport timescales could  
711 manifest as an incorrect interpretation of how processes change with discharge.

712         Based on these findings, this study provides general recommendations for future tracer  
713 studies. First, while a high-sensitivity tracer (fluorescein) provided more late-time tailing  
714 information than that obtained from a lower-sensitivity tracer (salt), both independently provided  
715 the same conclusion that the impact of changing advective times obscured other solute transport  
716 process changes. Furthermore, differences between these tracer types were smaller at higher  
717 discharges, indicating that tracer selection is less important than considering the influence of  
718 changing timescales on the interpretation of changing processes. Second, the influence of  
719 changing advective times and windows of detection must be established before tracer tests can be  
720 compared. Otherwise, a change in the transport timescale could be misinterpreted as a change in  
721 dispersion or transient storage. We show that this influence can be approximated by normalizing  
722 temporal metrics directly assessed from the observations by characteristics of the transport  
723 timescale including advective time and mean arrival time. However, the normalized metrics did  
724 not provide a sufficient correction to completely isolate dispersion and transient storage.  
725 Therefore, development of more complete relationships between processes and discharge may

726 require better approximations of late-time tailing through other methods. Lastly, we recommend  
727 that selection of study reaches for design of tracer studies should be less influenced by length  
728 scales, such as fixed reaches of equal length proposed in previous studies, and more so by the  
729 transport timescales, which can be approximated from preliminary tests or modeling efforts. This  
730 recommendation is based on evidence that differences between windows of detection still limit  
731 the comparison of tracer responses despite normalization by characteristics such as the advective  
732 time. Improving the ability to compare tracer observations will clarify relationships between  
733 solute transport processes and discharge necessary to use reach scale observations to infer solute  
734 transport at larger network scales.

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749 **References**

- 750
- 751 Aquino, T., A. Aubeneau, and D. Bolster (2015), Peak and tail scaling of breakthrough curves in  
752 hydrologic tracer tests, *Advances in Water Resources*, 78, 1-8.
- 753 Bellin, A., D. Tonina, and A. Marzadri (2015), Breakthrough curve moments scaling in  
754 hyporheic exchange, *Water Resources Research*, 51(2), 1353-1358.
- 755 Boulton, A. J., T. Datry, T. Kasahara, M. Mutz, and J. A. Stanford (2010), Ecology and  
756 management of the hyporheic zone: stream-groundwater interactions of running waters and  
757 their floodplains, *Journal of the North American Benthological Society*, 29(1), 26-40.
- 758 Cirpka, O. A., M. N. Fienen, M. Hofer, E. Hoehn, A. Tessarini, R. Kipfer, and P. K. Kitanidis  
759 (2007), Analyzing Bank Filtration by Deconvoluting Time Series of Electric Conductivity,  
760 *Ground Water*, 45(3), 318-328.
- 761 Covino, T., B. McGlynn, and J. Mallard (2011), Stream-groundwater exchange and hydrologic  
762 turnover at the network scale, *Water Resources Research*, 47(12).
- 763 D'Angelo, D. J., J. R. Webster, S. V. Gregory, and J. L. Meyer (1993), Transient Storage in  
764 Appalachian and Cascade Mountain Streams as Related to Hydraulic Characteristics, *Journal*  
765 *of the North American Benthological Society*, 12(3), 223-235.
- 766 Drummond, J. D., T. P. Covino, A. F. Aubeneau, D. Leong, S. Patil, R. Schumer, and A. I.  
767 Packman (2012), Effects of solute breakthrough curve tail truncation on residence time  
768 estimates: A synthesis of solute tracer injection studies, *Journal of Geophysical Research:*  
769 *Biogeosciences*, 117(G3).
- 770 Dudley-Southern, M., and A. Binley (2015), Temporal responses of groundwater-surface water  
771 exchange to successive storm events, *Water Resources Research*, 51(2), 1112-1126.
- 772 Dyer, F. J., and M. C. Thoms (2006), Managing river flows for hydraulic diversity: an example  
773 of an upland regulated gravel-bed river, *River Research and Applications*, 22(2), 257-267.
- 774 Fischer, H. B. (1975), Discussion of "Simple method for predicting dispersion in streams" by  
775 R.S. McQuivey and T.N. Keefer, *Journal of Environmental Engineering Division American*  
776 *Society of Civil Engineers*, 101(3), 453-455.
- 777 Foppen, J. W., J. Seopa, N. Bakobie, and T. Bogaard (2013), Development of a methodology for  
778 the application of synthetic DNA in stream tracer injection experiments, *Water Resources*  
779 *Research*, 49(9), 5369-5380.
- 780 Foppen, J. W., C. Orup, R. Adell, V. Poulalion, and S. Uhlenbrook (2011), Using multiple  
781 artificial DNA tracers in hydrology, *Hydrological Processes*, 25(19), 3101-3106.
- 782 Gelhar, L. W., C. Welty, and K. R. Rehfeldt (1992), A critical review of data on field-scale  
783 dispersion in aquifers, *Water Resources Research*, 28(7), 1955-1974.
- 784 González-Pinzón, R., R. Haggerty, and M. Dentz (2013), Scaling and predicting solute transport  
785 processes in streams, *Water Resources Research*, 49(7), 4071-4088.
- 786 Gooseff, M. N., R. O. Hall, and J. L. Tank (2007), Relating transient storage to channel  
787 complexity in streams of varying land use in Jackson Hole, Wyoming, *Water Resources*  
788 *Research*, 43(1), W01417.
- 789 Gooseff, M. N., M. A. Briggs, K. E. Bencala, B. L. McGlynn, and D. T. Scott (2013), Do  
790 transient storage parameters directly scale in longer, combined stream reaches? Reach length  
791 dependence of transient storage interpretations, *Journal of Hydrology*, 483(0), 16-25.
- 792 Gostner, W., P. Parasiewicz, and A. J. Schleiss (2013), A case study on spatial and temporal  
793 hydraulic variability in an alpine gravel-bed stream based on the hydromorphological index  
794 of diversity, *Ecology*, 6(4), 652-667.

- 795 Haggerty, R., S. M. Wondzell, and M. A. Johnson (2002), Power-law residence time distribution  
796 in the hyporheic zone of a 2nd-order mountain stream, *Geophysical Research Letters*, 29(13),  
797 18-11-18-14.
- 798 Harvey, J. W., and B. J. Wagner (2000), Quantifying hydrologic interactions between streams  
799 and their subsurface zones, in *Streams and Ground Waters*, edited by J. B. Jones and P. J.  
800 Mulholland, pp. 3-44, Academic Press, San Diego, CA.
- 801 Harvey, J. W., B. J. Wagner, and K. E. Bencala (1996), Evaluating the Reliability of the Stream  
802 Tracer Approach to Characterize Stream-Subsurface Water Exchange, *Water Resources*  
803 *Research*, 32(8), 2441-2451.
- 804 Hester, E. T., and M. N. Gooseff (2010), Moving Beyond the Banks: Hyporheic Restoration Is  
805 Fundamental to Restoring Ecological Services and Functions of Streams, *Environmental*  
806 *Science & Technology*, 44(5), 1521-1525.
- 807 Jackson, T. R., R. Haggerty, and S. V. Apte (2013), A fluid-mechanics based classification  
808 scheme for surface transient storage in riverine environments: quantitatively separating  
809 surface from hyporheic transient storage, *Hydrol. Earth Syst. Sci.*, 17(7), 2747-2779.
- 810 Kelleher, C., T. Wagener, B. McGlynn, A. S. Ward, M. N. Gooseff, and R. A. Payn (2013),  
811 Identifiability of transient storage model parameters along a mountain stream, *Water*  
812 *Resources Research*, 49(9), 5290-5306.
- 813 Kiel, B. A., and M. B. Cardenas (2014), Lateral hyporheic exchange throughout the Mississippi  
814 River network, *Nature Geosci*, 7(6), 413-417.
- 815 Kilpatrick, F. A., and E. D. Cobb (1985), Measurement of discharge using tracers, U.S.  
816 Geological Survey. *Techniques of Water-Resources Investigations*, Book 3 (Chapter A16,  
817 52pp.).
- 818 Leopold, L. B., and T. Maddock (1953), *The Hydraulic Geometry of Stream Channels and Some*  
819 *Physiographic Implications*, Geological Survey Professional Paper 252, U.S. Department of  
820 the Interior, Washington, D.C.
- 821 Mason, S. J. K., B. L. McGlynn, and G. C. Poole (2012), Hydrologic response to channel  
822 reconfiguration on Silver Bow Creek, Montana, *Journal of Hydrology*, 438-439(0), 125-136.
- 823 Payn, R. A., M. N. Gooseff, B. L. McGlynn, K. E. Bencala, and S. M. Wondzell (2009), Channel  
824 water balance and exchange with subsurface flow along a mountain headwater stream in  
825 Montana, United States, *Water Resources Research*, 45(11).
- 826 Payn, R. A., M. N. Gooseff, D. A. Benson, O. A. Cirpka, J. P. Zarnetske, W. B. Bowden, J. P.  
827 McNamara, and J. H. Bradford (2008), Comparison of instantaneous and constant-rate  
828 stream tracer experiments through non-parametric analysis of residence time distributions,  
829 *Water Resources Research*, 44(6).
- 830 Riml, J., and A. Wörman (2011), Response functions for in-stream solute transport in river  
831 networks, *Water Resources Research*, 47(6), W06502.
- 832 Runkel, R. L. (2002), A new metric for determining the importance of transient storage, *Journal*  
833 *of the North American Benthological Society*, 21(4), 529-543.
- 834 Schmadel, N. M., B. T. Neilson, and D. K. Stevens (2010), Approaches to estimate uncertainty  
835 in longitudinal channel water balances, *Journal of Hydrology*, 394(3-4), 357-369.
- 836 Schmadel, N. M., B. T. Neilson, and T. Kasahara (2014a), Deducing the spatial variability of  
837 exchange within a longitudinal channel water balance, *Hydrological Processes*, 28(7), 3088-  
838 3103.

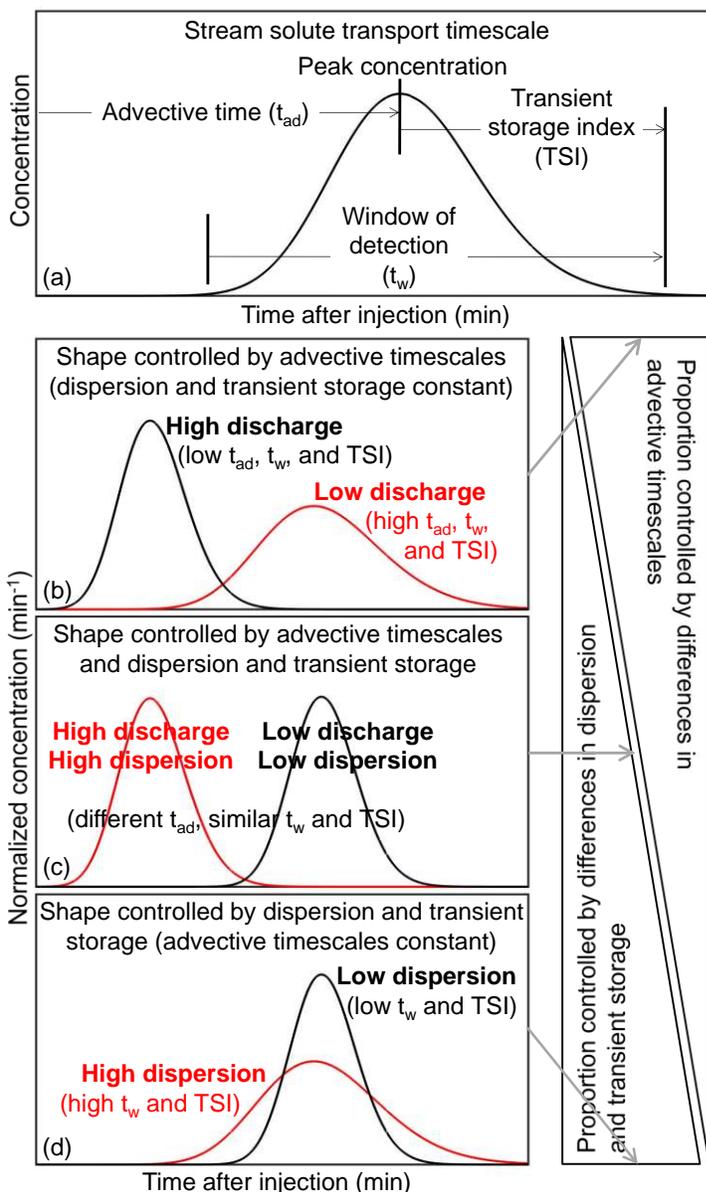
- 839 Schmadel, N. M., B. T. Neilson, J. E. Heavilin, D. K. Stevens, and A. Wörman (2014b), The  
840 influence of spatially variable stream hydraulics on reach scale solute transport modeling,  
841 *Water Resources Research*, 10.1002/2014WR015440.
- 842 Schmid, B. (2002), Persistence of Skewness in Longitudinal Dispersion Data: Can the Dead  
843 Zone Model Explain It After All?, *Journal of Hydraulic Engineering*, 128(9), 848-854.
- 844 Schmid, B. H., I. Innocenti, and U. Sanfilippo (2010), Characterizing solute transport with  
845 transient storage across a range of flow rates: The evidence of repeated tracer experiments in  
846 Austrian and Italian streams, *Advances in Water Resources*, 33(11), 1340-1346.
- 847 Schmidt, C., A. Musolff, N. Trauth, M. Vieweg, and J. H. Fleckenstein (2012), Transient  
848 analysis of fluctuations of electrical conductivity as tracer in the stream bed, *Hydrol. Earth  
849 Syst. Sci.*, 16(10), 3689-3697.
- 850 Stewart, R. J., W. M. Wollheim, M. N. Gooseff, M. A. Briggs, J. M. Jacobs, B. J. Peterson, and  
851 C. S. Hopkinson (2011), Separation of river network-scale nitrogen removal among the main  
852 channel and two transient storage compartments, *Water Resources Research*, 47(10),  
853 W00J10.
- 854 Stonedahl, S. H., J. W. Harvey, J. Detty, A. Aubeneau, and A. I. Packman (2012), Physical  
855 controls and predictability of stream hyporheic flow evaluated with a multiscale model,  
856 *Water Resources Research*, 48(10), W10513.
- 857 Stream Solute Workshop (1990), Concepts and Methods for Assessing Solute Dynamics in  
858 Stream Ecosystems, *Journal of the North American Benthological Society*, 9(2), 95-119.
- 859 Toran, L., J. E. Nyquist, A. C. Fang, R. J. Ryan, and D. O. Rosenberry (2013), Observing  
860 lingering hyporheic storage using electrical resistivity: variations around stream restoration  
861 structures, Crabby Creek, PA, *Hydrological Processes*, 27(10), 1411-1425.
- 862 Trauth, N., C. Schmidt, M. Vieweg, S. E. Oswald, and J. H. Fleckenstein (2015), Hydraulic  
863 controls of in-stream gravel bar hyporheic exchange and reactions, *Water Resources  
864 Research*, 51(4), 2243-2263.
- 865 Voltz, T., M. Gooseff, A. S. Ward, K. Singha, M. Fitzgerald, and T. Wagener (2013), Riparian  
866 hydraulic gradient and stream-groundwater exchange dynamics in steep headwater valleys,  
867 *Journal of Geophysical Research: Earth Surface*, 118(2), 953-969.
- 868 Wagner, B. J., and J. W. Harvey (1997), Experimental design for estimating parameters of rate-  
869 limited mass transfer: Analysis of stream tracer studies, *Water Resources Research*, 33(7),  
870 1731-1741.
- 871 Ward, A. S., M. N. Gooseff, M. Fitzgerald, T. J. Voltz, and K. Singha (2014), Spatially  
872 distributed characterization of hyporheic solute transport during baseflow recession in a  
873 headwater mountain stream using electrical geophysical imaging, *Journal of Hydrology*,  
874 517(0), 362-377.
- 875 Ward, A. S., M. Fitzgerald, M. N. Gooseff, T. J. Voltz, A. M. Binley, and K. Singha (2012),  
876 Hydrologic and geomorphic controls on hyporheic exchange during base flow recession in a  
877 headwater mountain stream, *Water Resources Research*, 48(4), W04513.
- 878 Ward, A. S., M. N. Gooseff, T. J. Voltz, M. Fitzgerald, K. Singha, and J. P. Zarnetske (2013a),  
879 How does rapidly changing discharge during storm events affect transient storage and  
880 channel water balance in a headwater mountain stream?, *Water Resources Research*, 49(9),  
881 5473-5486.
- 882 Ward, A. S., N. M. Schmadel, S. M. Wondzell, C. Harman, M. N. Gooseff, and K. Singha  
883 (2016), Hydrogeomorphic controls on hyporheic and riparian transport in two headwater

- 884 mountain streams during base flow recession, *Water Resources Research*,  
885 10.1002/2015WR018225.
- 886 Ward, A. S., R. A. Payn, M. N. Gooseff, B. L. McGlynn, K. E. Bencala, C. A. Kelleher, S. M.  
887 Wondzell, and T. Wagener (2013b), Variations in surface water-ground water interactions  
888 along a headwater mountain stream: Comparisons between transient storage and water  
889 balance analyses, *Water Resources Research*, 49(6), 3359-3374.
- 890 Wollheim, W. M., B. J. Peterson, S. M. Thomas, C. H. Hopkinson, and C. J. Vörösmarty (2008),  
891 Dynamics of N removal over annual time periods in a suburban river network, *Journal of*  
892 *Geophysical Research: Biogeosciences*, 113(G3).
- 893 Wondzell, S. M. (2006), Effect of morphology and discharge on hyporheic exchange flows in  
894 two small streams in the Cascade Mountains of Oregon, USA, *Hydrological Processes*,  
895 20(2), 267-287.
- 896 Wondzell, S. M. (2011), The role of the hyporheic zone across stream networks, *Hydrological*  
897 *Processes*, 25(22), 3525-3532.
- 898 Wondzell, S. M., and M. N. Gooseff (2013), 9.13 Geomorphic Controls on Hyporheic Exchange  
899 Across Scales: Watersheds to Particles, in *Treatise on Geomorphology*, edited by J. F.  
900 Shroder, pp. 203-218, Academic Press, San Diego.
- 901 Zarnetske, J. P., R. Haggerty, S. M. Wondzell, and M. A. Baker (2011), Dynamics of nitrate  
902 production and removal as a function of residence time in the hyporheic zone, *Journal of*  
903 *Geophysical Research: Biogeosciences*, 116(G1).
- 904 Zarnetske, J. P., M. N. Gooseff, T. R. Brosten, J. H. Bradford, J. P. McNamara, and W. B.  
905 Bowden (2007), Transient storage as a function of geomorphology, discharge, and  
906 permafrost active layer conditions in Arctic tundra streams, *Water Resources Research*,  
907 43(7), W07410.
- 908 Zimmer, M. A., and L. K. Lautz (2014), Temporal and spatial response of hyporheic zone  
909 geochemistry to a storm event, *Hydrological Processes*, 28(4), 2324-2337.
- 910

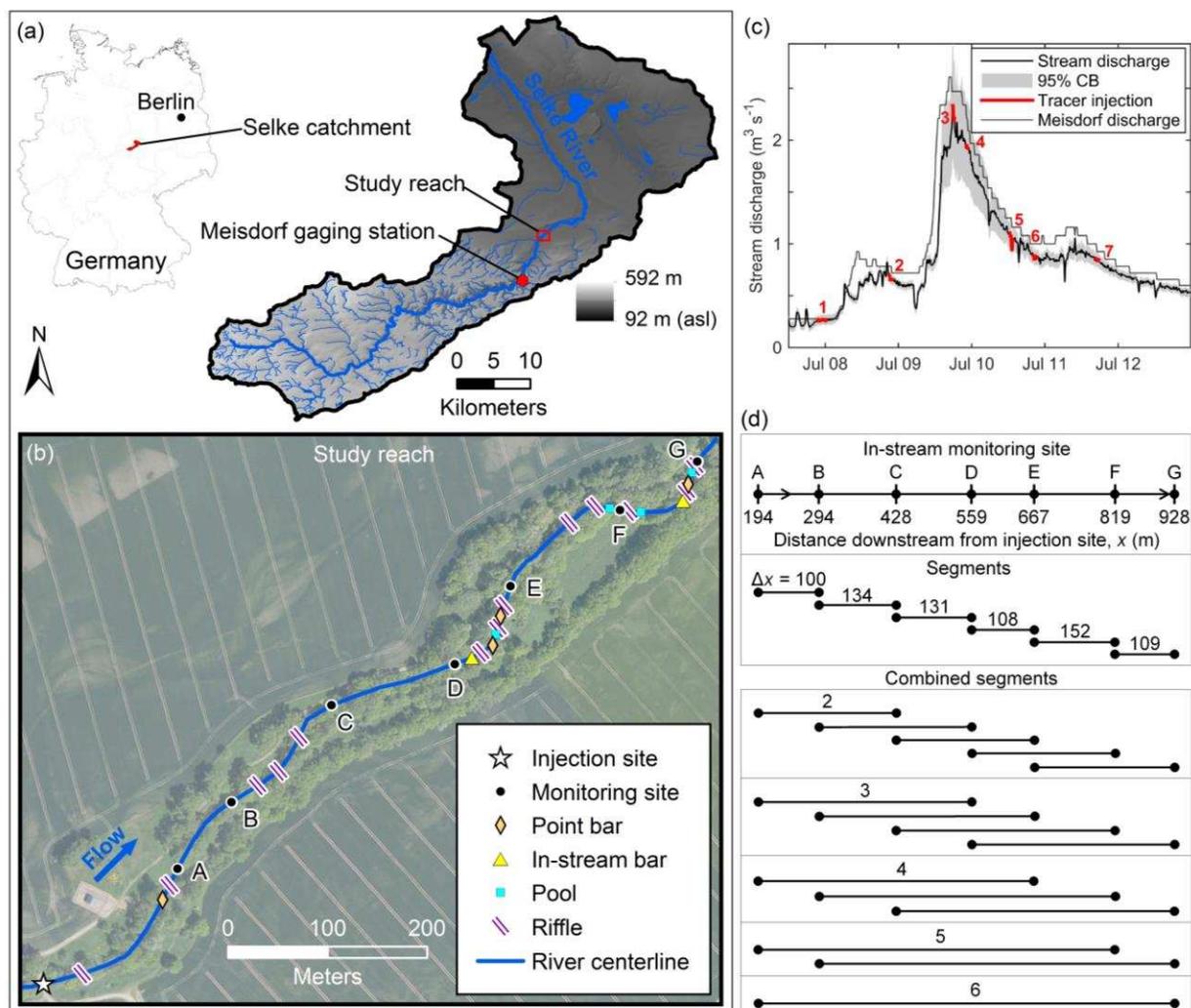
911 **Table 1.** The reach scale temporal metrics, window of detection ( $t_{w, \text{norm}}$ ) and transient storage  
 912 index ( $\text{TSI}_{\text{norm}}$ ) normalized by the advective time, coefficient of variation (CV), and skewness ( $\gamma$ ),  
 913 of the transit time distributions measured at in-stream monitoring sites A through G and their  
 914 relationships (slope) with changing stream discharge (Q) (also see Figure 5, right column). Note  
 915 that 41 salt and 4 fluorescein breakthrough curves were used to estimate these metrics and  
 916 relationships. A positive or negative slope indicates the metrics increase or decrease with an  
 917 increasing Q, respectively. A slope is considered significant (bold text) if zero is outside of the  
 918 associated 95% confidence interval.

Site	x (m)	Number of Qs	$t_{w, \text{norm}}$		$\text{TSI}_{\text{norm}}$		CV		$\gamma$	
			Slope with Q	$R^2$	Slope with Q	$R^2$	Slope with Q	$R^2$	Slope with Q	$R^2$
A	194	6	2.0E-02	0.07	-1.4E-02	0.03	-1.8E-02	0.60	<b>4.6E-01</b>	0.97
B	294	7	-5.4E-02	0.40	-1.5E-02	2.2E-02	-6.8E-03	0.37	-5.5E-04	6.0E-06
C	428	7	-8.4E-02	0.53	-6.2E-02	0.48	<b>-1.1E-02</b>	0.76	-8.2E-03	4.7E-03
D	559	7	-8.6E-02	0.42	-7.2E-02	0.30	<b>-1.2E-02</b>	0.60	-2.0E-02	2.6E-02
E	667	5	-1.2E-01	0.62	-9.5E-02	0.54	-1.6E-02	0.72	-1.1E-01	0.43
F	819	6	3.9E-02	0.09	5.1E-02	0.11	-1.2E-03	3.5E-03	1.8E-01	0.31
G	928	7	-7.6E-02	0.24	-5.3E-02	0.10	-8.6E-03	0.18	4.4E-03	3.5E-04

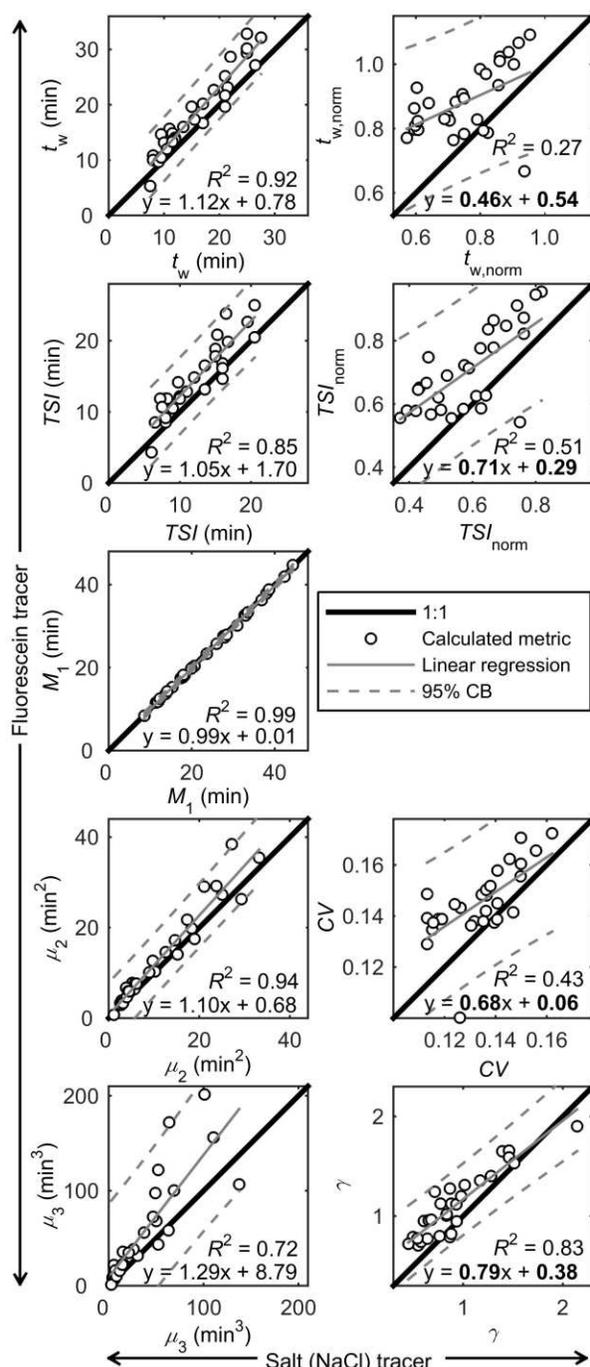
**Bold** indicates significant value



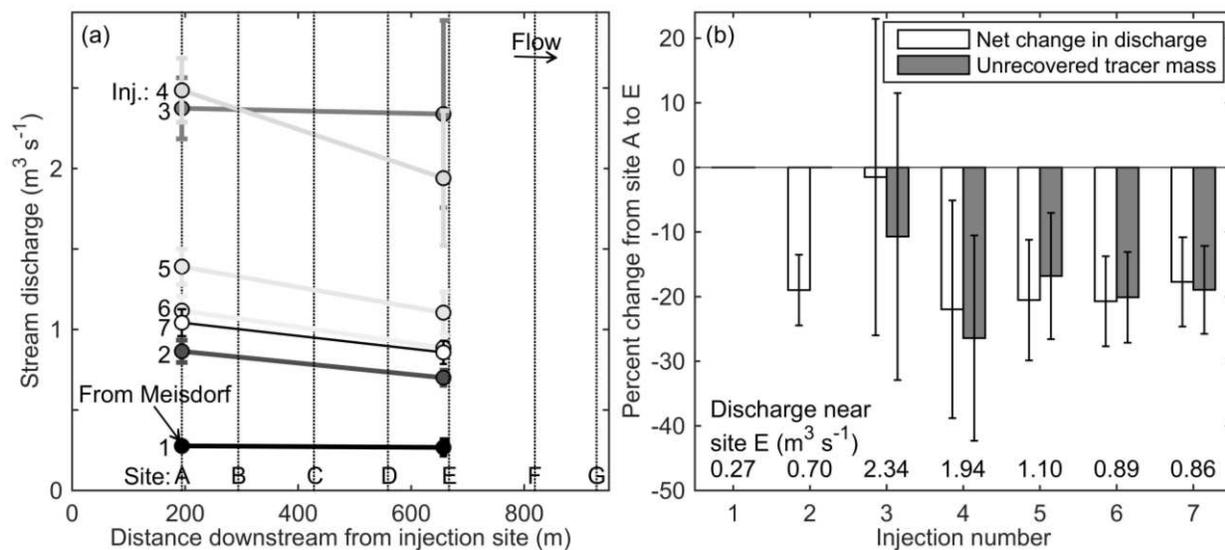
919  
 920 **Figure 1.** (a) The stream solute transport timescale, defined here as the advective time ( $t_{ad}$ ) and  
 921 window of detection ( $t_w$ ), can be directly estimated from an observed in-stream conservative  
 922 solute tracer time series (breakthrough curve, or BTC). Following an instantaneous injection, the  
 923 elapsed time from tracer injection to peak concentration describes  $t_{ad}$ . The elapsed time from  
 924 tracer first arrival to last detection describes  $t_w$ . The elapsed time from  $t_{ad}$  to tracer last detection  
 925 provides an indicator of transient storage (transient storage index, or TSI). Below is an  
 926 illustration of the challenge of interpreting solute transport processes from stream solute tracers.  
 927 (b) At one extreme, if underlying processes (e.g., dispersion and transient storage) are constant  
 928 but  $t_{ad}$  is different between high and low discharge conditions, the shape of the BTC is controlled  
 929 by differences in advective timescales. (d) At the other extreme, if  $t_{ad}$  is constant between two  
 930 reaches, the shape of the BTC is controlled by differences in the underlying processes. (c) In  
 931 practice, shapes of observed BTCs are controlled by different proportions of advective  
 932 timescales and these underlying processes.



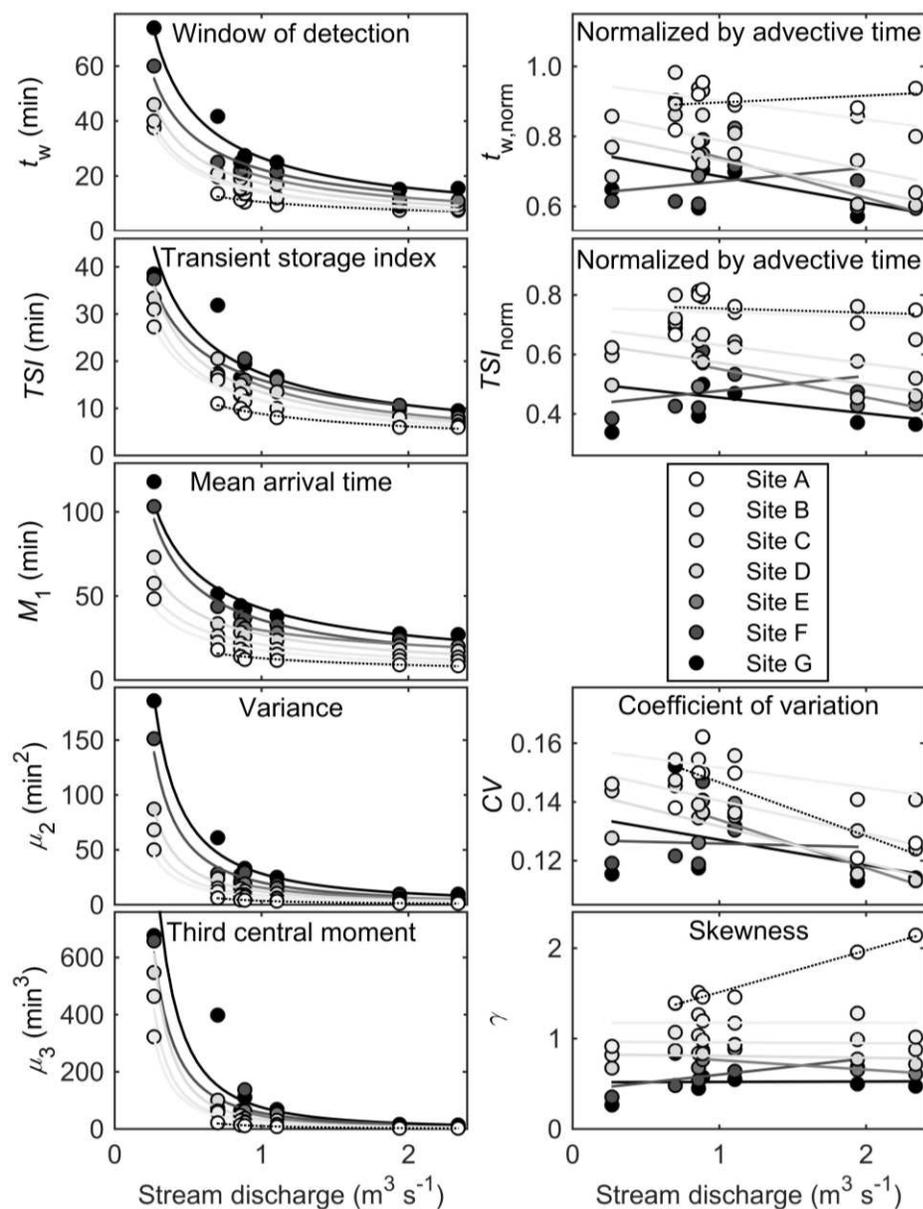
933  
 934 **Figure 2.** (a) The Selke River catchment located in central Germany and the location of the  
 935 study reach and Meisdorf federal gaging station. (b) In-stream monitoring sites and general  
 936 channel morphologic characteristics of the study reach. (c) Stream discharge estimated within the  
 937 study reach from an established rating curve (15 m upstream of site E) and the associated 95%  
 938 confidence bounds (CB), timing of tracer test durations (elapsed time from injection to last  
 939 detection at site G), and discharge at the Meisdorf gaging station located ~4.6 km upstream of  
 940 the injection site. (d) The monitoring sites delineate six reach segments, which were recombined  
 941 into every possible unique combination (grouped by the number of individual segments) to  
 942 produce segments of various lengths and transport timescales.



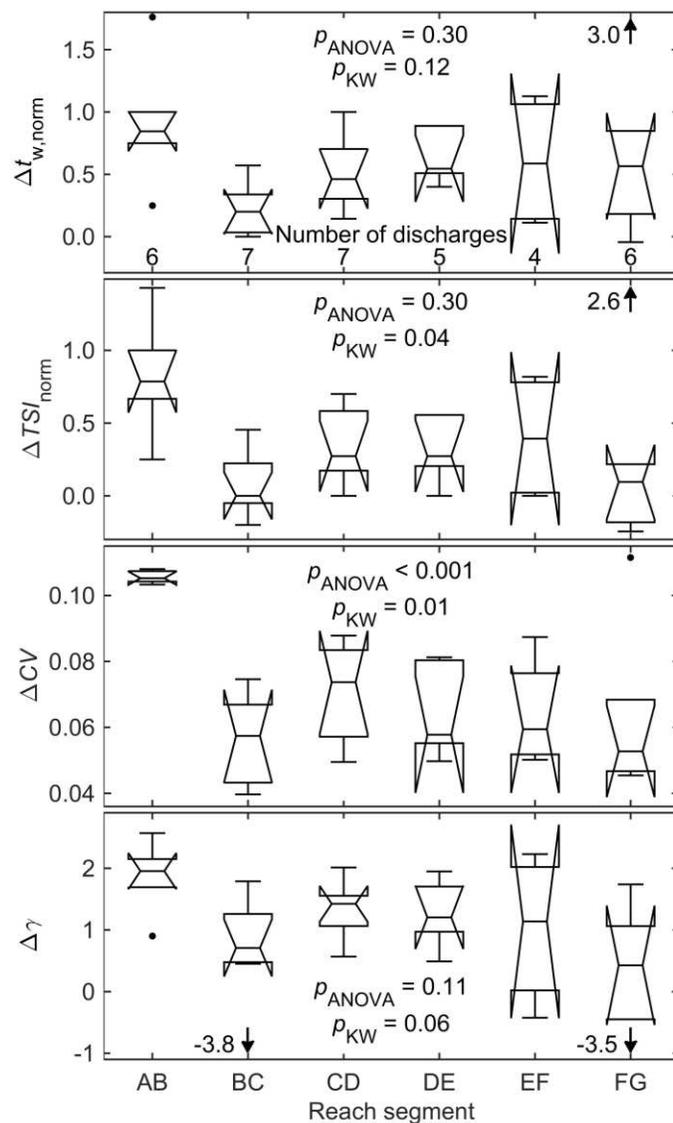
943  
 944 **Figure 3.** Comparison of the co-located salt (NaCl) and fluorescein breakthrough curves  
 945 measured at in-stream monitoring sites A through G (i.e. 29 each salt and fluorescein). In the left  
 946 column, the metrics compared are the window of detection ( $t_w$ ), transient storage index (TSI),  
 947 mean arrival time ( $M_1$ ), variance ( $\mu_2$ ), and third central moment ( $\mu_3$ ). In the right column, the  
 948 normalized metrics are compared, which include  $t_w$  and TSI normalized by the associated  
 949 advective time ( $t_{w, \text{norm}}$  and  $\text{TSI}_{\text{norm}}$ ), the coefficient of variation (CV), and skewness ( $\gamma$ ).  
 950 Corresponding linear trend-lines are shown along with their slopes, intercepts, coefficients of  
 951 determination ( $R^2$ ), and 95% confidence bounds (CB). Bold text indicates that either the slope is  
 952 statistically different than 1 or the intercept is statistically different than zero.



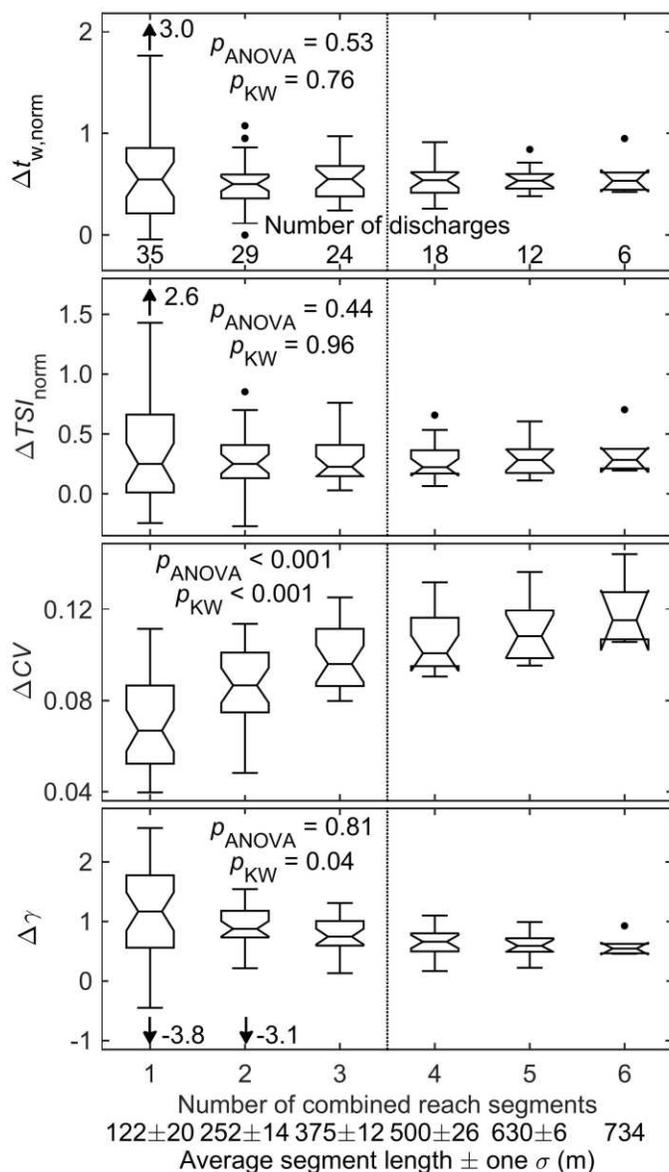
953  
 954 **Figure 4.** (a) Discharge estimates at site A (estimated from dilution gaging) and near site E  
 955 (estimated from established rating curve). (b) Percent net change in discharge from site A to E  
 956 and percent unrecovered salt tracer mass from the injection to site E. Error bars are the estimated  
 957 95% confidence intervals.



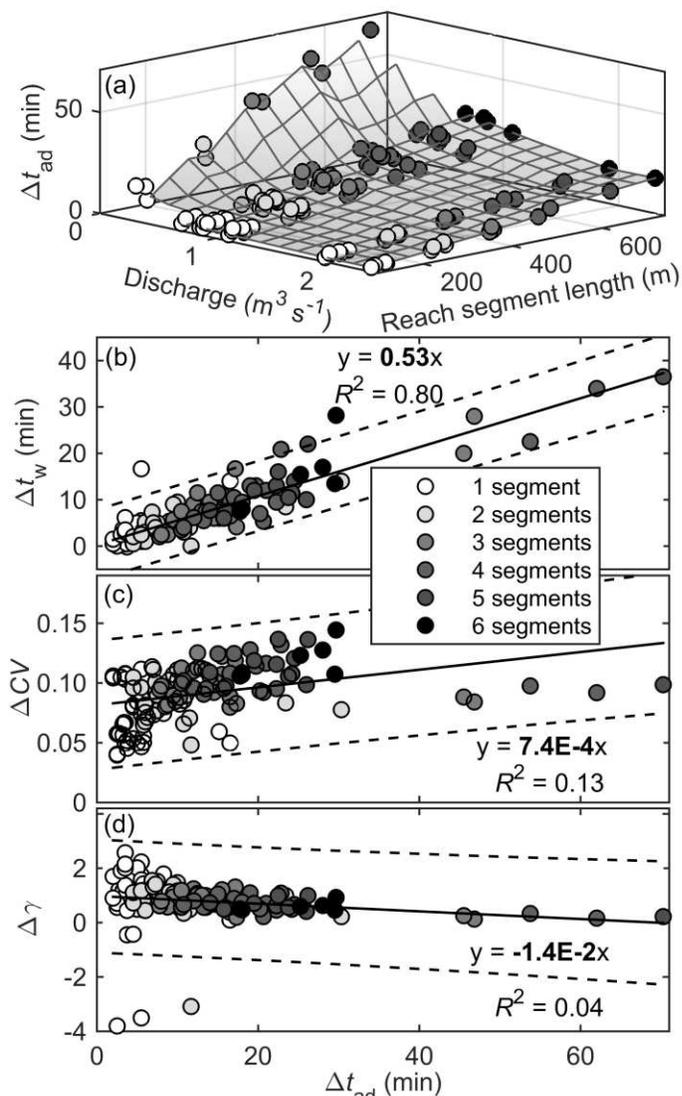
958  
 959 **Figure 5.** In the left column, the reach scale temporal metrics of the transit time distributions  
 960 measured at in-stream monitoring sites A through G. Specifically, these metrics are the window  
 961 of detection ( $t_w$ ), transient storage index (TSI), mean arrival time ( $M_1$ ), variance ( $\mu_2$ ), and  
 962 third central moment ( $\mu_3$ ). Note that 41 salt and 4 fluorescein breakthrough curves were used to  
 963 estimate these metrics. Each metric decreases with increasing discharge and increases with  
 964 increasing travel distance (shown by a general site-specific exponential function and white to  
 965 black color, respectively). To better compare across discharges, shown in the right column are  
 966  $t_w$  and TSI normalized by the advective time ( $t_{w, \text{norm}}$  and  $\text{TSI}_{\text{norm}}$ ), the coefficient of variation (CV),  
 967 and skewness ( $\gamma$ ). Linear trend-lines were estimated for each site and a slope is considered  
 968 significant if zero is outside of the associated 95% confidence interval (see Table 1).



969  
 970 **Figure 6.** Segment-wise distributions (box-and-whisker plots of the quartiles where the notches  
 971 are the 95% confidence intervals about the medians) of the transfer function metrics that express  
 972 the window of detection ( $\Delta t_{w, \text{norm}}$ ) and transient storage index ( $\Delta TSI_{\text{norm}}$ ) normalized by the  
 973 change in advective time, coefficient of variation ( $\Delta CV$ ), and skewness ( $\Delta \gamma$ ) over changing  
 974 discharge. Note that 41 salt and 4 fluorescein breakthrough curves were used to estimate these  
 975 metrics, resulting in 35 unique values (i.e., segment-wise estimates corresponding to number of  
 976 discharges). Distribution means were compared through a one-way analysis of variance  
 977 (ANOVA) and medians were compared through a Kruskal-Wallis one-way ANOVA. For mean  
 978 and median comparisons, a  $p_{\text{ANOVA}} < 0.05$  or  $p_{\text{KW}} < 0.05$  indicates the mean or median of at least  
 979 one segment is statistically different from all others, respectively. See Table S5 in supporting  
 980 information S1 for segment-to-segment pairwise comparisons.



981  
 982 **Figure 7.** The temporal metrics of the transfer function, the window of detection ( $\Delta t_{w, \text{norm}}$ ) and  
 983 transient storage index ( $\Delta \text{TSI}_{\text{norm}}$ ) normalized by the advective time, coefficient of variation  
 984 ( $\Delta \text{CV}$ ), and skewness ( $\Delta \gamma$ ), lumped together for each number of individual segments combined  
 985 together (see Figure 2d). The corresponding distributions are represented as box-and-whisker  
 986 plots of the quartiles where the notches are the 95% confidence intervals about the medians.  
 987 Each distribution contains unique estimates representing different segment lengths and  
 988 discharges (denoted as the number of discharges). Distribution means were compared through a  
 989 one-way analysis of variance (ANOVA) and medians through a Kruskal-Wallis one-way  
 990 ANOVA. A  $p_{\text{ANOVA}} < 0.05$  or  $p_{\text{KW}} < 0.05$  indicates the mean or median of at least one  
 991 distribution is statistically different from all others, respectively. Only the distributions  
 992 representing the original segments, two combined segments, and three combined segments (see  
 993 Figure 2d) were compared (vertical dashed line).



994  
 995 **Figure 8.** (a) The advective time of the transfer function ( $\Delta t_{ad}$ ) relative to discharge and reach  
 996 segment length. A surface was linearly interpolated between these estimates to illustrate the  
 997 general trend. (b) The window of detection of the transfer function ( $\Delta t_w$ ) relative to  $\Delta t_{ad}$  for each  
 998 segment combination. A linear trend-line (solid line) and associated 95% confidence bounds  
 999 (dashed lines) were approximated. The slope of this trend-line and coefficient of determination  
 1000 ( $R^2$ ) are shown. (c) The coefficient of variation ( $\Delta CV$ ) and (d) skewness ( $\Delta \gamma$ ) for each segment  
 1001 combination relative to the corresponding  $\Delta t_{ad}$  estimates and corresponding trend-line and 95%  
 1002 CBs. Bold text indicates a significant value (i.e., zero is outside of the associated 95%  
 1003 confidence interval).