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1 **Local and regional controls of phylogenetic structure at the high-latitude range limits**  
2 **of corals**

3

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15

16 Key words: range limit, phylogenetic structure, niche conservatism, environmental filtering, coral,  
17 trait

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21

22 **Abstract**

23

24 Understanding how range edge populations will respond to climate change is an urgent research  
25 priority. Here we used a phylogenetic community ecology approach to examine how ecological and  
26 evolutionary processes shape biodiversity patterns of scleractinian corals at their high-latitude range  
27 limits in eastern Australia. We estimated phylogenetic signal in seven ecologically important  
28 functional traits and conducted tests of phylogenetic structure at local and regional scales using the  
29 net relatedness (NRI) and nearest taxon indices (NTI) for presence/absence and abundance data.  
30 Regional tests showed light phylogenetic clustering, indicating that coral species found in this  
31 subtropical-to-temperate transition zone are more closely related to each other than are species on  
32 the nearby, more northerly Great Barrier Reef. Local tests revealed variable patterns of  
33 phylogenetic clustering and overdispersion and higher than expected phylogenetic turnover among  
34 sites. In combination, these results are broadly consistent with the hierarchical filtering model,  
35 whereby species pass through a regional climatic filter based on their tolerances for marginal  
36 conditions and subsequently segregate into local assemblages according to the relative strength of  
37 habitat filtering and species interactions. Conservatism of tested traits suggests that corals will  
38 likely track their niches with climate change. Nevertheless, high turnover of lineages among sites  
39 indicates that range shifts will probably vary among species and highlights the vulnerability and  
40 conservation significance of high-latitude reefs.

41

42

## 43 INTRODUCTION

44

45 Understanding local and regional controls on biodiversity is fundamental to understanding species  
46 range dynamics and how they may be affected by global change. Ecological community structure  
47 arises from a combination of processes that operate at different spatial and temporal scales. Species  
48 may occur in the same communities by occupying different niches (i.e. limiting similarity) and / or  
49 because they have specialised traits that allow them to persist locally (i.e. environmental filtering)  
50 [1]. Classical ecological theory predicts that the relative importance of these processes in driving  
51 biodiversity patterns varies systematically along stress-gradients, with abiotic factors predominant  
52 at high elevation, high latitude, and other abiotically stressful regions, and biotic interactions  
53 relatively more important in abiotically more benign regions [2, 3]. Moreover, a growing body of  
54 literature suggests that biotic and abiotic filters operate at different spatial scales, with  
55 environmental conditions often found to be the dominant process at regional scales and biotic  
56 interactions relatively more important at local scales [4, 5]. Quantifying the relative strength of  
57 these processes at multiple spatial scales is necessary to accurately predict species range shifts in  
58 response to climate change [3]. For instance, when biotic interactions are important in setting  
59 species range limits, models solely based on climatic influences could overestimate species range  
60 shifts in response to climate change [3].

61

62 Phylogenies are increasingly used to investigate ecological and evolutionary drivers of biodiversity  
63 patterns. The central idea is that patterns of phylogenetic community structure differ depending on  
64 whether environmental filtering or competition is the dominant assembly process [6]. This is based  
65 on the premise that closely related species have similar traits [i.e. trait conservatism; 7] and  
66 consequently require similar environmental conditions and resources to persist. Close relatives  
67 therefore compete more intensely for the same resources, which is thought to limit coexistence and  
68 to drive trait and niche difference [2]. Environmental filtering, on the other hand, can prevent

69 species with unsuitable traits from occurring locally. Under classical ecological theory, competition  
70 is therefore generally expected to limit the similarity of co-occurring species, leading to  
71 phylogenetic overdispersion, while environmental filtering is thought to drive the co-occurrence of  
72 closely related species, leading to phylogenetic clustering [6]. For example, seminal work by  
73 Cavender-Bares et al. [8] found that co-occurring oak trees in Florida were phylogenetically  
74 overdispersed, suggesting that competition restricts close relatives from inhabiting the same areas.  
75 Recent advances in coexistence theory challenge the universality of this pattern, highlighting that  
76 competition can also lead to the co-occurrence of similar species, if the traits in question are  
77 associated with competitive dominance [9].

78  
79 Gradients in phylogenetic alpha and beta diversity and tests for phylogenetic structure have been  
80 used to infer the dominant assembly processes operating in local and regional assemblages [6]. This  
81 work indicates that biotic and abiotic filters often operate at different spatial scales, such that local  
82 assemblages are embedded in regional assemblages and species sequentially pass through historical,  
83 abiotic and biotic filters [4, 10]. For example, regional climate might initially filter species with  
84 certain traits (e.g. cold tolerance) into large patches of homogeneous habitat (e.g. alpine mountains)  
85 that are further partitioned at finer scales (e.g. soil type). Species then segregate into local  
86 assemblages based on the relative strength of habitat filtering versus competition [4, 6]. These  
87 processes can be detected by spatially nesting analyses [6] and by combining metrics of  
88 phylogenetic alpha and beta diversity [4]. For example, environmental filtering at local scales  
89 combined with high turnover of habitat types in the region is expected to manifest in local  
90 phylogenetic clustering and high phylogenetic beta diversity among sites [4].

91  
92 While studies of phylogenetic community ecology are well established in terrestrial ecosystems,  
93 this approach is underrepresented in marine ecosystems [reviewed in 11, but see 12, 13]. Due to  
94 their ecological diversity, eco-morphological plasticity and habitat sensitivity, scleractinian corals

95 are especially suited to examining the role of evolutionary history in shaping extant communities. In  
96 eastern Australia, diverse coral assemblages extend into high latitudes ( $32^{\circ}48'S$ ) south of the Great  
97 Barrier Reef [14-16], where various coral species occur at the edges of their ranges and experience  
98 cooler, darker and more seasonal and variable conditions than their tropical counterparts [17]. These  
99 high-latitude reefs represent ideal systems in which to investigate how evolutionary stability of  
100 species traits and phylogenetic relatedness may influence biodiversity patterns in marginal and  
101 changing environments. A recent study found that these range-edge coral assemblages are shaped  
102 by trait-mediated environmental filtering, with similar species co-occurring in the region [15] along  
103 strong environmental gradients [17]. As phylogenetic relatedness does not predict species'  
104 ecological similarity per se [18], investigating the evolutionary stability of ecological traits in these  
105 physiologically challenging environments is a critical next step in predicting species responses to  
106 climate change. For instance, if species' climatic tolerances are stable over time (i.e. conserved),  
107 they will probably only expand their ranges to environments that are similar to their core ranges,  
108 and be constrained in their evolutionary responses to novel environmental conditions [7].

109

110 In this study, we take a hierarchical approach (Fig. 1) to disentangle the relative importance of  
111 regional to local processes in generating biodiversity patterns of corals in the subtropical-to-  
112 temperate transition zone ( $26^{\circ}36'S$  to  $32^{\circ}48'S$ ) along the eastern Australian coast. To investigate  
113 evolutionary stability of species traits, we initially estimate phylogenetic signal in ecologically  
114 important functional traits. Under niche conservatism we expect closely related species to be more  
115 similar in their functional traits than distantly related species, implying that there has been limited  
116 evolutionary change since species have diverged from a common ancestor [7]. We subsequently  
117 combine analyses of phylogenetic alpha and beta diversity and of phylogenetic structure at multiple  
118 spatial scales to evaluate how evolutionary and ecological processes shape the distribution of corals  
119 in this biogeographic transition zone (Fig.1). To examine the shared history of high-latitude coral  
120 assemblages with the Great Barrier Reef (GBR; situated to the north of the study region), we test

121 whether species found on eastern Australian high-latitude reefs are a random subset of the GBR  
122 species pool in terms of their phylogenetic relatedness (Fig. 1A, B). Consistent with stronger  
123 influence of environmental filtering at the high-latitude margins of species ranges [19], we expect  
124 that the species within the high-latitude species pool are more closely related to one another than  
125 are the species within the Great Barrier Reef species pool. We then perform site-level tests of  
126 phylogenetic structure to investigate whether local assemblages are phylogenetically clustered or  
127 overdispersed (Fig. 1C) and whether patterns vary with latitude. Due to increasingly marginal  
128 environmental conditions for corals towards higher latitudes, we expect coral species to be more  
129 closely related to each other at higher latitude sites and high turnover of lineages (i.e. high  
130 phylogenetic beta diversity; Fig. 1C) among sites in this biogeographic transition zone.

131

## 132 **MATERIALS AND METHODS**

133

### 134 **Ecological data**

135

#### 136 Coral species abundances

137 We determined species abundance distributions of scleractinian corals at 17 locations  
138 (Supplementary Fig. S1) along a latitudinal gradient (26°36'S to 32°48'S) in the subtropical-to-  
139 temperate transition zone in coastal eastern Australia [15]. We quantified percent cover of 85 coral  
140 species from photographic belt transects using the random point count method [600 points per  
141 transect, total of 51,000 points; see 15].

142

#### 143 Coral functional traits and phylogeny

144 To examine functional strategies of corals, we characterised each species based on seven traits  
145 known to influence the ecology of corals [20]. We included the three structural traits colony  
146 morphology, corallite formation, and corallite size, as they undergo strong environmental filtering

147 at high-latitudes [15], as well as reproductive mode, sexual pattern, larval development rate and  
148 symbiont transmission mode to characterise reproductive strategies [21-23]. Following Keith et al.  
149 [24] egg size was used as a proxy for larval development rate, because small eggs develop more  
150 rapidly [25]. The character states of all categorical traits were expressed as separate binary traits,  
151 where '1' indicates that the species displays the trait and '0' that it does not. For example,  
152 reproductive mode was classified as two separate binary traits: brooding and spawning (see  
153 supplementary material for more information on traits). Due to the known variation with latitude in  
154 other traits [e.g. growth rate; 26], and our inability to quantify this variation due to data deficiency,  
155 we did not include more traits in our analyses. We used the coral 'supertree' of Huang and Roy [13]  
156 to construct phylogenies of the 85 coral species recorded in our surveys and of the 412 species from  
157 the GBR species pool obtained from the Atlas of Living Australia [27].

158

## 159 **Statistical analyses**

160

### 161 Tests for phylogenetic signal in traits

162 We tested for phylogenetic signal in traits using the D statistic [28] for binary traits (i.e. colony  
163 morphology, corallite formation, reproductive mode, sexual pattern and symbiont transmission),  
164 and Pagel's  $\lambda$  [29] for the traits corallite size and larval development rate. Lower D values indicate  
165 stronger phylogenetic signal (i.e. more conserved trait evolution), whereby a value of 0 equates to a  
166 Brownian model (i.e. conserved trait evolution) and a value of 1 to random phylogenetic structure  
167 (i.e. no signal). Negative D values indicate that traits are extremely clumped (i.e. more conserved  
168 than under Brownian motion) and D values greater than 1 indicate phylogenetic overdispersion [28].  
169 Conversely, higher values of Pagel's  $\lambda$  indicate stronger phylogenetic signal, where  $\lambda$  equal to zero  
170 means no phylogenetic signal and  $\lambda$  equal to 1 indicates trait evolution consistent with Brownian  
171 motion. P values indicate departure from random phylogenetic structure and from Brownian  
172 structure (i.e. non-significant P values for Brownian motion indicate conserved trait evolution).

173

174 Patterns in phylogenetic diversity and tests for phylogenetic structure

175 We quantified phylogenetic diversity patterns of corals using two metrics that are sensitive to  
176 patterns at different depths of the phylogeny and identify processes operating at different  
177 evolutionary time-scales [6], mean pairwise distance (MPD) and mean nearest taxon distance  
178 (MNTD). MPD represents the average phylogenetic distance between all species in an assemblage  
179 and is sensitive to tree-wide (i.e. more ancestral) patterns. MNTD represents the average distance  
180 between each species and its closest relative and is thus sensitive to patterns at the tips of the  
181 phylogeny (i.e. more recent events). In comparative analyses, MPD and MNTD values can be  
182 affected by differences in species richness among sites. At our study sites, coral species richness  
183 ranged between 52 and 2 species; we therefore standardised the observed MPD and MNTD values  
184 with respect to expected values calculated under a null model that controls for coincidental  
185 gradients in species richness. More specifically, we calculated the net relatedness index (NRI) and  
186 the nearest taxon index (NTI) to test whether species in local assemblages were phylogenetically  
187 clustered or overdispersed i.e. relatively more closely or distantly related than randomly expected,  
188 respectively [6, 30]. We compared the observed MPD and MNTD values to null distributions  
189 generated by shuffling the species names across the tips of the phylogeny (1000 times). This null  
190 model randomises the relatedness of species to one another, but leaves the abundance distributions  
191 and species richness of local assemblages unchanged. In addition to these site level tests, we also  
192 conducted a regional test to determine whether coral species found in the subtropical-to-temperate  
193 transition zone (i.e. all species that we recorded at the 17 study sites; Fig. 1B) are more closely or  
194 distantly related than the species in the GBR phylogeny (Fig. 1A). For both local and regional tests,  
195 positive NRI/NTI values indicate phylogenetic clustering while negative values indicate  
196 phylogenetic overdispersion [30] (Fig. 1C). We performed nonparametric multidimensional scaling  
197 ordination to visualise the relationships among patterns in phylogenetic structure, community-  
198 weighted trait means, and coral community structure at the study sites.

199

200 Phylogenetic beta diversity was calculated in a way analogous to the NRI and NTI metrics, by  
201 comparing the observed phylogenetic beta diversities to the patterns expected under a null model  
202 that randomly shuffles the species names across the GBR phylogeny. Positive values of beta\_NRI  
203 and beta\_NTI indicate higher than randomly expected phylogenetic turnover of species in pairs of  
204 local assemblages; negative values indicate lower than expected phylogenetic turnover. As the coral  
205 supertree contains a number of polytomies, we conducted all analyses for 1000 fully resolved trees  
206 [13] and report results as the means over the 1000 trees. We calculated all metrics using both  
207 presence/absence and abundance data, with respective results identified by the letters ‘pa’ and ‘ab’  
208 (e.g. beta\_NRIpa, beta\_NRIab). When used with abundance data the interpretation of metrics  
209 changes from phylogenetic distances among taxa to distances among individuals. We used  
210 generalised additive models to evaluate the relationships between phylogenetic alpha diversity  
211 metrics and latitude. We selected smoothness using restricted maximum likelihood and specified  
212 smoothness with up to three effective degrees of freedom. We used Procrustes analysis to test the  
213 association of phylogenetic beta diversity patterns and geographic distance (in degrees latitude) and  
214 visualised site-level beta diversities using boxplots. All analyses were conducted using the packages  
215 ‘picante’ [31], ‘caper’ [32], ‘vegan’ [33], and ‘mgcv’ [34] in R [35].

216

## 217 **RESULTS**

218

### 219 **Tests for phylogenetic signal in traits**

220

221 We found a phylogenetic signal for all tested functional traits, as their phylogenetic structures were  
222 significantly different from random and not significantly different from Brownian motion (Table 1).  
223 Indeed, six of the binary traits (sexual pattern, symbiont transmission, laminar growth, meandroid  
224 corallites, encrusting-to-massive and branching-to-tabular growth) had negative D values,

225 indicating that they were more phylogenetically conserved than under a Brownian model [28].  
226 Arborescent growth had the weakest signal and showed modest clumping (Table 1).

227

## 228 **Patterns in phylogenetic diversity and tests for phylogenetic structure**

229

230 Overall, the regional test for phylogenetic structure indicates that the coral species present on high-  
231 latitude reefs (the high-latitude species pool) are more closely related to one another than are the  
232 species in the GBR species pool. Phylogenetic structure for NRI (NRI = 0.842) and NTI (NTI =  
233 1.289; Table S1 in supplementary material) was related to a clustering pattern, although values  
234 calculated for 1000 phylogenetic trees were only significantly clustered for approximately 25% of  
235 trees (third quartile: NRI = 1.193, NTI = 1.516). Site level tests of phylogenetic structure revealed  
236 contrasting patterns for NRI and NTI, except for northern- and southernmost sites where both  
237 metrics were positive (Fig. 2). Phylogenetic structure for most sites was clustered at the tips of the  
238 phylogeny (NTI > 0) and overdispersed or close to random when measured at deep levels in the  
239 phylogeny (NRI < 0), with NRI and NTI metrics generally showing both low positive and negative  
240 values. NRI<sub>pa</sub> showed a concave relationship with latitude ( $R^2 = 0.549$ , deviance explained = 60.5,  
241  $p = 0.002$ ), but NRI<sub>ab</sub> ( $R^2 = 0.162$ , DE = 25.5,  $p = 0.192$ ), NTI<sub>pa</sub> ( $R^2 = 0.006$ , DE = 6.12,  $p =$   
242 0.356), and NTI<sub>ab</sub> ( $R^2 = 0.028$ , DE = 9.240,  $p = 0.252$ ) were independent of latitude (Fig. 2).

243

244 Assemblages that showed basal and terminal clustering (red circles in Fig. 3) were characterised by  
245 species with branching-to-tabular morphology, large corallites, fast development rates,  
246 hermaphroditic sexual pattern, brooding reproduction, and vertical symbiont transmission (e.g.  
247 Pocillopora, Stylophora, Acanthastrea). This pattern was seen across sites except at Black Rock,  
248 which had high abundance of laminar species that were gonochoric broadcast spawners and  
249 transmit their symbionts horizontally (e.g. Turbinaria spp.). Assemblages that showed contrasting  
250 patterns for NRI and NTI (green diamonds and blue triangles in Fig. 3) generally showed large

251 variation in species traits, except for Flinders Reef (orange square in Fig. 3), which associated  
252 strongly with arborescent morphology, spawning reproduction and horizontal symbiont  
253 transmission (e.g. *Acropora* spp.). Note that vectors represent associations of community-weighted  
254 trait means with the nMDS axes and that individual species don't necessarily possess the trait  
255 combinations listed above.

256

257 Analyses of phylogenetic betadiversity showed that most sites showed higher than expected  
258 phylogenetic turnover at both basal and terminal levels ( $\beta\text{-NRI}$  and  $\beta\text{-NTI} > 0$ ), and that this  
259 pattern was stronger when abundance information was used (Fig. 4). With the exception of  
260  $\beta\text{-NRIpa}$  (Procrustes correlation  $t = 0.48$ ;  $p = 0.015$ ), phylogenetic turnover patterns were  
261 independent of geographic distance (Fig. 4). Indeed, some geographically close site pairs (within 0-  
262 3 degrees latitude of each other) had high terminal phylogenetic beta diversity ( $\beta\text{-NTIab}$ ),  
263 indicating high turnover of abundant species from site to site (Fig. 4).  $\beta\text{-NRIpa}$  and  $\beta\text{-NTIpa}$   
264 were positively correlated with each other (Procrustes correlation  $t = 0.67$ ,  $p = 0.001$ ), indicating  
265 that site pairs with high basal turnover tended to also have high terminal turnover ( $\beta\text{-NTI} > 0$ ).  
266 This relationship was not significant ( $t = 0.21$ ,  $p = 0.929$ ) when species abundances were considered.

267

## 268 **DISCUSSION**

269

### 270 **Phylogenetic niche conservatism of corals**

271

272 Understanding species range dynamics and how they may be affected by global change requires the  
273 consideration of multiple ecological and evolutionary processes [36]. Our results suggest that coral  
274 species attenuate at higher latitudes in response to habitat filtering that is mediated by conserved  
275 ecological niches. Niche conservatism has been invoked to explain large-scale biogeographic  
276 gradients, such as the latitudinal diversity gradient [reviewed in 7]. This is based on the premise that

277 many groups originated in the tropics and are limited from colonising temperate regions by niche  
278 conservatism (such as lack of adaptations to survive cold winter temperatures). Conversely, recent  
279 studies have found multimodality of latitudinal gradients in species distributions and thermal  
280 affinities of marine species [e.g. 37] and linked these patterns to adaptations of ‘bridge species’ at  
281 the edges of the tropics and subsequent colonisation of temperate regions [i.e. the out of the tropics  
282 hypothesis’; 38]. While rigorous tests of these hypotheses require data for entire species ranges, our  
283 results indicate trait conservatism for corals at their high-latitude range limits and we find  
284 phylogenetic diversity patterns that are consistent with species sorting in these marginal  
285 environments. These findings are consistent with a recent study of habitat affinities of marine  
286 invertebrates over geological time - niche conservatism was common in the oceans, but variable  
287 among clades, with coral and sponges among the most conservative [39].

288

289 We found a phylogenetic signal in all tested species traits (Table 1). Larval development rate,  
290 corallite size, sexual patterns and symbiont transmission showed the strongest conservatism among  
291 the tested traits. Indeed, six of the binary traits (sexual pattern, symbiont transmission, laminar  
292 growth, meandroid corallites, encrusting-to-massive and branching-to-tabular growth) were more  
293 phylogenetically conserved than under a Brownian model [28]. Structural traits such as corallite  
294 size and colony morphology undergo strong filtering in this region, whereby coral species with  
295 similar trait expressions frequently co-occur at the same sites [15]. Corallite size has been linked  
296 with energy storage and reproductive investment [40], and colony shape with light interception [41]  
297 and it is plausible that these traits enhance energy acquisition [15] in these colder and light-limited  
298 environments [17]. This is also consistent with the observation that corals on high-latitude reefs in  
299 Australia and Japan share traits relevant to environmental tolerance [42].

300

301 Faster larval development rates and larger depth ranges were recently found to increase the  
302 likelihood that coral species traverse faunal breaks [24]. Indeed, in our study species with relatively

303 fast development rates, including *Paragoniastrea australensis*, *Turbinaria* spp., *Acanthastrea* spp.  
304 and *Pocillopora* spp., were abundant and widely distributed across the high latitude study region  
305 [15]. Larval developmental time has important ecological implications for marine organisms  
306 because it affects dispersal distance and connectivity [22, 25]. For instance, high relative abundance  
307 of brooding corals, whose larvae are ready to settle upon release [25], is thought to be important for  
308 population establishment and persistence on Lord Howe Island, the world's southernmost coral reef  
309 [43, 44]. In our study, the strong association of brooding reproduction and fast development rates  
310 with coral assemblages that were phylogenetically clustered (red circles in Fig. 3) also support the  
311 hypothesis that the capacity to derive new recruits from local populations is important in these  
312 marginal habitats [43].

### 313 314 **Phylogenetic community ecology of high-latitude coral assemblages**

315  
316 Overall, regional tests revealed phylogenetic clustering of coral assemblages, indicating that coral  
317 species on high-latitude reefs are on average more closely related with one another than are species  
318 in the GBR species pool. While this is generally consistent with regional filtering, NRI and NTI  
319 values only showed significant clustering for approximately 25% of the 1000 phylogenetic trees.  
320 Site level results (Fig. 2) support the role of local filtering, indicating that patterns of phylogenetic  
321 clustering and overdispersion vary among sites and at ancestral and more recent temporal scales.  
322 This is consistent with other studies that found closely related or functionally similar species co-  
323 occurring at larger spatial scales and more distantly related or dissimilar species co-occurring  
324 locally. For example, local tree frog assemblages in the Americas were generally overdispersed,  
325 while regional assemblages were frequently clustered in their body sizes [10]. Moreover, the  
326 authors found that regional processes tend to be stronger in temperate than in tropical zones, such  
327 that antagonistic regional and local filtering can substantially mask variation in trait dispersion [10].

328

329 We found that most assemblages contained species that were distributed randomly or slightly  
330 overdispersed with respect to the entire tree ( $NRI < 0$ ), but were clustered towards the tips of the  
331 tree ( $NTI > 0$ ). Thus, although local assemblages support numerous lineages that are widely spread  
332 across the entire phylogeny, they harbour many close relatives within those lineages. Such evenness  
333 in tree-wide phylogenetic structure ( $NRI$ ) of local assemblages indicates that a diverse range of  
334 lineages has been able to disperse to and establish on many high-latitude reefs, including many  
335 species in the Acroporidae family that are the predominant reef builders on tropical coral reefs [45].  
336 Nevertheless, the presence of close relatives within those lineages indicates the co-occurrence of  
337 species with similar traits and tolerances for marginal conditions. Branch tips represent more recent  
338 divergences, where new species have formed. Species that are closely related at the tips of the  
339 phylogeny are therefore generally expected to occupy more similar niches and to have similar  
340 resource needs when species traits are conserved [6]. The traits branching-to-tabular morphology,  
341 large corallite sizes, hermaphroditic sexual pattern, brooding reproductive mode, fast development  
342 rates, and vertical symbiont transmission were most strongly associated with phylogenetically  
343 clustered assemblages (red circles in Fig. 3) that included Pocillopora, Stylophora, Acanthastrea  
344 and Micromussa species. Moreover, gonochoric sexual pattern, broadcast spawning reproduction,  
345 horizontal symbiont transmission and laminar growth also appear to be advantageous at sites that  
346 were phylogenetically clustered (green diamonds in Fig. 3). In particular, these sites had high  
347 abundance of several Turbinaria species that tend to do well in marginal environments [15, 46].

348

349 Contrary to our expectation, species at higher latitudes were not phylogenetically more related to  
350 each other than those at lower latitudes.  $NRI_{pa}$  even showed a convex relationship with latitude,  
351 indicating that phylogenetic clustering was strongest at opposite ends of the study region (Fig. 2).  
352 Indeed, Mudjimba Island, the northernmost site in our study, contained species that were  
353 phylogenetically clustered at deep and shallow levels of the phylogeny. In particular, Mudjimba  
354 comprised many lineages from the robust clade (e.g. genera Acanthastrea, Paragoniastrea,

355 Goniastrea, Dipsastrea, Favites, Pocillopora and Stylophora) and contained several species within  
356 most genera. Moreover, several abundant species (e.g. Acanthastrea) were clustered in those  
357 lineages, leading to higher NRI<sub>ab</sub> than NRI<sub>pa</sub> values (Fig. 2). Our study was restricted to  
358 subtropical and temperate latitudes and future work that examines these patterns across the entire  
359 tropical-to-temperate gradient is needed to assess the importance of climatic niche conservatism in  
360 shaping phylogenetic structure of coral assemblages throughout core and marginal species ranges.

361

362 Species interactions and tolerances for local environmental conditions both influence species  
363 abundances. Abundance information can therefore considerably improve inferences from  
364 community phylogenetic studies and improve the power to detect limiting-similarity competition  
365 [47]. In our study, phylogenetic structure for several sites (i.e. Flinders Reef, Flat Rock and Cook  
366 Island) changed from clustered to overdispersed when abundance information was used in the  
367 analysis, highlighting the influence of biotic interactions at these sites and the importance of using  
368 abundance data. For example, Flinders Reef, the most speciose of our survey sites, had the highest  
369 occurrence of species from the Acroporidae family of all sites, leading to phylogenetic clustering  
370 for both NRI and NTI when presence/absence data were used. However, high dominance of the  
371 arborescent species *Acropora intermedia*, which occupied a large patch of the reef in a mono-  
372 specific stand (site F and photo in Fig. 3), resulted in weak overdispersion at the tips of the tree  
373 when species abundances were considered (NTI<sub>ab</sub> < 0), pointing to the influence of biotic  
374 interactions. This example highlights that abundance-weighted metrics more accurately reflect  
375 assembly processes, particularly where there is high dominance of particular species with unusual  
376 traits (arborescent growth was absent or rare at other sites; Fig. 3).

377

378 Phylogenetic beta diversity measures the phylogenetic distance among local assemblages and  
379 allows us to connect local and regional processes [48]. Our phylogenetic betadiversity analyses  
380 show that most site pairs had higher than expected phylogenetic turnover at both basal and terminal

381 levels (Fig. 4). These patterns were stronger when species abundances were considered, consistent  
382 with species present at many sites but only abundant at certain sites [48]. Site pairs with high basal  
383 turnover (beta\_NRIpa) generally also had high terminal turnover (beta\_NTIpa), but this relationship  
384 was not significant for abundance-weighted data. Except for beta\_NRIpa, which was greater for  
385 geographically distant sites, we found no relationship between patterns in phylogenetic turnover and  
386 geographic distance. Indeed, beta\_NTIab was particularly high among site pairs that were  
387 geographically close (within 0-3 degrees latitude or 0-300 km of each other), indicating high  
388 turnover of abundant species among these sites (Fig. 4 e and f). In particular, site pairs in the  
389 Solitary Islands region (30° southern latitude; Fig. S1 in supplementary material) were  
390 phylogenetically most dissimilar from each other at the branch tips (Fig. 4f), even though closely  
391 related species co-occurred locally (Fig. 2). These patterns of local phylogenetic clustering (Fig. 2)  
392 and higher than randomly expected phylogenetic turnover (Fig. 4f) are consistent with high spatial  
393 and temporal variability in environmental conditions and biological communities in this region,  
394 where assemblages range from kelp- to coral-dominated [14, 16, 17].

395

## 396 **Conclusions and caveats**

397

398 In combination, patterns in phylogenetic alpha and beta diversity are broadly consistent with the  
399 hierarchical filtering model of community assembly operating on high-latitude reefs (Fig. 1); that is  
400 regional filtering of the high-latitude species pool relative to the GBR species pool (Fig. 1A-B),  
401 segregation of species into local assemblages depending on the relative strength of biotic and  
402 abiotic processes, and high phylogenetic turnover in this biogeographic transition zone (Fig. 1C).  
403 Recent advances in coexistence theory suggest that patterns of phylogenetic clustering in  
404 observational data alone can be insufficient to clearly establish the role of environmental filtering [9,  
405 49, 50]. While experiments are therefore needed to tease apart the relative strength of niche and  
406 competitive fitness differences among coral species, our results are consistent with findings that

407 environmental tolerance is important in structuring high-latitude coral assemblages [15, 17, 42, 43].  
408 More specifically, we hypothesise that the colder, light limited and more variable environmental  
409 conditions that govern coral biodiversity patterns in this region [17] impose abiotic filters at the  
410 high-latitude range limits for corals. Nevertheless, our local scale analyses revealed the signature of  
411 competitive processes at some sites, particularly for abundance-weighted data, highlighting the  
412 importance of using abundance data when investigating assembly processes. Conservatism of tested  
413 traits suggests that coral distributions will likely shift with climate change. While this is consistent  
414 with recent observations of coral range expansion on high-latitude reefs in Japan [51] and Australia  
415 [52], the higher than expected phylogenetic beta diversity between most site pairs recorded in this  
416 study and species-specific relationships with environmental parameters [17, 53], suggest that coral  
417 range shifts will likely vary among species. High phylogenetic turnover among sites highlights the  
418 vulnerability and conservation significance of high-latitude reefs [14] and the need to protect a  
419 substantial proportion of reefs to preserve the breadth of phylogenetic diversity in the region.

420

421 **Ethics:** Methods were approved under the University of Queensland ethics guidelines (193/10).

422 **Data accessibility:** Supporting data are freely available as electronic supplementary material.

423 **Authors' contributions:** BS, ES, MB and JMP designed the study; BS and MB collected the field  
424 data; BS identified corals, performed data analysis and led the writing; all authors contributed to  
425 writing the manuscript.

426 **Competing interests:** We declare we have no competing interests.

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436

## 437 **References**

438 [1] Chase, J.M. & Leibold, M.A. 2003 *Ecological niches: Linking classical and contemporary*  
439 *approaches*, University of Chicago Press; 212 p.

440 [2] Darwin, C. 1859 *The origin of species by means of natural selection*. London, Murray.

441 [3] Louthan, A.M., Doak, D.F. & Angert, A.L. 2015 Where and when do species interactions set  
442 range limits? *Trends Ecol. Evol.* **30**, 780-792.

443 [4] Swenson, N.G., Erickson, D.L., Mi, X., Bourg, N.A., Forero-Montaña, J., Ge, X., Howe, R., Lake,  
444 J.K., Liu, X., Ma, K., et al. 2012 Phylogenetic and functional alpha and beta diversity in  
445 temperate and tropical tree communities. *Ecology* **93**, S112-S125.

446 [5] Weiher, E. & Keddy, P.A. 1995 Assembly rules, null models, and trait dispersion - new  
447 questions front old patterns. *Oikos* **74**, 159-164.

448 [6] Webb, C.O., Ackerly, D.D., McPeck, M.A. & Donoghue, M.J. 2002 Phylogenies and community  
449 ecology. *Annu. Rev. Ecol. Syst.* **33**, 475-505.

450 [7] Wiens, J.J., Ackerly, D.D., Allen, A.P., Anacker, B.L., Buckley, L.B., Cornell, H.V., Damschen, E.I.,  
451 Davies, T.J., Grytnes, J.A., Harrison, S.P., et al. 2010 Niche conservatism as an emerging  
452 principle in ecology and conservation biology. *Ecol. Lett.* **13**, 1310-1324.

453 [8] Cavender-Bares, J., Ackerly, D.D., Baum, D.A. & Bazzaz, F.A. 2004 Phylogenetic  
454 overdispersion in Floridian oak communities. *Am. Nat.* **163**, 823-843.

455 [9] Mayfield, M.M. & Levine, J.M. 2010 Opposing effects of competitive exclusion on the  
456 phylogenetic structure of communities. *Ecol. Lett.* **13**, 1085-1093.

457 [10] Algar, A.C., Kerr, J.T. & Currie, D.J. 2011 Quantifying the importance of regional and local  
458 filters for community trait structure in tropical and temperate zones. *Ecology* **92**, 903-914.

- 459 [11] Vamosi, S.M., Heard, S.B., Vamosi, J.C. & Webb, C.O. 2009 Emerging patterns in the  
460 comparative analysis of phylogenetic community structure. *Mol. Ecol.* **18**, 572-592.
- 461 [12] Leprieur, F., Colosio, S., Descombes, P., Parravicini, V., Kulbicki, M., Cowman, P.F.,  
462 Bellwood, D.R., Mouillot, D. & Pellissier, L. 2015 Historical and contemporary determinants of  
463 global phylogenetic structure in tropical reef fish faunas. *Ecography* **39**, 825-835.
- 464 [13] Huang, D. & Roy, K. 2015 The future of evolutionary diversity in reef corals. *Philosophical  
465 Transactions of the Royal Society of London B: Biological Sciences* **370**, 20140010.
- 466 [14] Beger, M., Sommer, B., Harrison, P.L., Smith, S.D.A. & Pandolfi, J.M. 2014 Conserving  
467 potential coral reef refuges at high latitudes. *Divers. Distrib.* **20**, 245-257.
- 468 [15] Sommer, B., Harrison, P.L., Beger, M. & Pandolfi, J.M. 2014 Trait-mediated environmental  
469 filtering drives assembly at biogeographic transition zones. *Ecology* **95**, 1000-1009.
- 470 [16] Harriott, V.J., Smith, S.D.A. & Harrison, P.L. 1994 Patterns of coral community structure of  
471 subtropical reefs in the Solitary-Islands Marine Reserve, Eastern Australia. *Marine Ecology-  
472 Progress Series* **109**, 67-76.
- 473 [17] Sommer, B., Beger, M., Harrison, P.L., Babcock, R.C. & Pandolfi, J.M. 2017 Differential  
474 response to abiotic stress drives species distributions at biogeographic transition zones.  
475 *Ecography*.
- 476 [18] Losos, J.B. 2008 Phylogenetic niche conservatism, phylogenetic signal and the  
477 relationship between phylogenetic relatedness and ecological similarity among species. *Ecol.  
478 Lett.* **11**, 995-1003.
- 479 [19] Fischer, A.G. 1960 Latitudinal Variations in Organic Diversity. *Evolution* **14**, 64-81.
- 480 [20] Madin, J.S., Hoogenboom, M.O., Connolly, S.R., Darling, E.S., Falster, D.S., Huang, D., Keith,  
481 S.A., Mizerek, T., Pandolfi, J.M., Putnam, H.M., et al. 2016 A Trait-based approach to advance  
482 coral reef science. *Trends Ecol. Evol.* **31**, 419-428.
- 483 [21] Baird, A.H., Guest, J.R. & Willis, B.L. 2009 Systematic and biogeographical patterns in the  
484 reproductive biology of scleractinian corals. In *Annu. Rev. Ecol. Evol. Syst.* (pp. 551-571).

- 485 [22] Harrison, P.L. 2011 Sexual reproduction of scleractinian corals. In *Coral Reefs: An*  
486 *Ecosystem in Transition* (eds. Z. Dubinsky & N. Stambler), pp. 59-85, Springer.
- 487 [23] Madin, J.S., Anderson, K.D., Andreasen, M.H., Bridge, T.C.L., Cairns, S.D., Connolly, S.R.,  
488 Darling, E.S., Diaz, M., Falster, D.S., Franklin, E.C., et al. 2016 The coral trait database, a curated  
489 database of trait information for coral species from the global oceans. *Scientific Data* **3**,  
490 160017.
- 491 [24] Keith, S.A., Baird, A.H., Hughes, T.P., Madin, J.S. & Connolly, S.R. 2013 Faunal breaks and  
492 species composition of Indo-Pacific corals: the role of plate tectonics, environment and habitat  
493 distribution. *Proceedings of the Royal Society B: Biological Sciences* **280**.
- 494 [25] Figueiredo, J., Baird, A.H. & Connolly, S.R. 2013 Synthesizing larval competence dynamics  
495 and reef-scale retention reveals a high potential for self-recruitment in corals. *Ecology* **94**,  
496 650-659.
- 497 [26] Harriott, V.J. 1999 Coral growth in subtropical eastern Australia. *Coral Reefs* **18**, 281-291.
- 498 [27] The Atlas of Living Australia. Accessed 3 December 2014 at [www.ala.org.au](http://www.ala.org.au)
- 499 [28] Fritz, S.A. & Purvis, A. 2010 selectivity in mammalian extinction risk and threat types: a  
500 new measure of phylogenetic signal strength in binary traits. *Conserv. Biol.* **24**, 1042-1051.
- 501 [29] Pagel, M. 1999 Inferring the historical patterns of biological evolution. *Nature* **401**, 877-  
502 884.
- 503 [30] Webb, C.O. 2000 Exploring the phylogenetic structure of ecological communities: An  
504 example for rain forest trees. *Am. Nat.* **156**, 145-155.
- 505 [31] Kembel, S.W., Cowan, P.D., Helmus, M.R., Cornwell, W.K., Morlon, H., Ackerly, D.D.,  
506 Blomberg, S.P. & Webb, C.O. 2010 Picante: R tools for integrating phylogenies and ecology.  
507 *Bioinformatics* **26**, 1463-1464.
- 508 [32] Orme, D., Freckleton, R., Thomas, G., Petzoldt, T., Fritz, S., Isaac, N. & Pearse, W. caper:  
509 Comparative analyses of phylogenetics and evolution in R. R package version 0.5.2.

510 [33] Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'Hara, R.B., Simpson, G.,  
511 Solymos, P., Stevens, M.H.H. & Wagner, H. 2012 vegan: Community ecology package. R  
512 package version 2.0-5.

513 [34] Wood, S.N. 2011 Fast stable restricted maximum likelihood and marginal likelihood  
514 estimation of semiparametric generalized linear models. *Journal of the Royal Statistical*  
515 *Society: Series B (Statistical Methodology)* **73**, 3-36.

516 [35] R Core Team. 2012 R: A language and environment for statistical computing. R  
517 Foundation for Statistical Computing. Vienna, Austria.

518 [36] Sexton, J.P., McIntyre, P.J., Angert, A.L. & Rice, K.J. 2009 Evolution and ecology of species  
519 range limits. In *Annu. Rev. Ecol. Evol. Syst.* (pp. 415-436).

520 [37] Stuart-Smith, R.D., Edgar, G.J., Barrett, N.S., Kininmonth, S.J. & Bates, A.E. 2015 Thermal  
521 biases and vulnerability to warming in the world's marine fauna. *Nature* **528**, 88-92.

522 [38] Jablonski, D., Belanger, C.L., Berke, S.K., Huang, S., Krug, A.Z., Roy, K., Tomasovych, A. &  
523 Valentine, J.W. 2013 Out of the tropics, but how? Fossils, bridge species, and thermal ranges in  
524 the dynamics of the marine latitudinal diversity gradient. *Proceedings of the National Academy*  
525 *of Sciences* **110**, 10487-10494.

526 [39] Hopkins, M.J., Simpson, C. & Kiessling, W. 2014 Differential niche dynamics among major  
527 marine invertebrate clades. *Ecol. Lett.* **17**, 314-323.

528 [40] Leuzinger, S., Anthony, K.R.N. & Willis, B.L. 2003 Reproductive energy investment in  
529 corals: scaling with module size. *Oecologia* **136**, 524-531.

530 [41] Hoogenboom, M.O., Connolly, S.R. & Anthony, K.R.N. 2008 Interactions between  
531 morphological and physiological plasticity optimize energy acquisition in corals. *Ecology* **89**,  
532 1144-1154.

533 [42] Mizerek, T.L., Baird, A.H., Beaumont, L.J. & Madin, J.S. 2016 Environmental tolerance  
534 governs the presence of reef corals at latitudes beyond reef growth. *Glob. Ecol. Biogeogr.* **25**,  
535 979-987.

- 536 [43] Keith, S.A., Woolsey, E.S., Madin, J.S., Byrne, M. & Baird, A.H. 2015 Differential  
537 establishment potential of species predicts a shift in coral assemblage structure across a  
538 biogeographic barrier. *Ecography* **38**, 1225-1234.
- 539 [44] Noreen, A.M.E., Schmidt-Roach, S., Harrison, P.L. & van Oppen, M.J.H. 2015 Diverse  
540 associations among coral host haplotypes and algal endosymbionts may drive adaptation at  
541 geographically peripheral and ecologically marginal locations. *J. Biogeogr.* **42**, 1639-1650.
- 542 [45] Wallace, C.C. 1999 *Staghorn corals of the world*. Collingwood, Australia, CSIRO Publishing.
- 543 [46] Butler, I.R., Sommer, B., Zann, M., Zhao, J.x. & Pandolfi, J.M. 2013 The impacts of flooding  
544 on the high-latitude, terrigenoclastic influenced coral reefs of Hervey Bay, Queensland,  
545 Australia. *Coral Reefs* **32**, 1149-1163.
- 546 [47] Freilich, M.A. & Connolly, S.R. 2015 Phylogenetic community structure when competition  
547 and environmental filtering determine abundances. *Glob. Ecol. Biogeogr.* **24**, 1390-1400.
- 548 [48] Fine, P.V.A. & Kembel, S.W. 2011 Phylogenetic community structure and phylogenetic  
549 turnover across space and edaphic gradients in western Amazonian tree communities.  
550 *Ecography* **34**, 552-565.
- 551 [49] Godoy, O., Kraft, N.J.B. & Levine, J.M. 2014 Phylogenetic relatedness and the determinants  
552 of competitive outcomes. *Ecol. Lett.* **17**, 836-844.
- 553 [50] Kraft, N.J.B., Adler, P.B., Godoy, O., James, E.C., Fuller, S. & Levine, J.M. 2015 Community  
554 assembly, coexistence and the environmental filtering metaphor. *Funct. Ecol.* **29**, 592-599.
- 555 [51] Yamano, H., Sugihara, K. & Nomura, K. 2011 Rapid poleward range expansion of tropical  
556 reef corals in response to rising sea surface temperatures. *Geophysical Research Letters* **38**.
- 557 [52] Baird, A.H., Sommer, B. & Madin, J.S. 2012 Pole-ward range expansion of *Acropora* spp.  
558 along the east coast of Australia. *Coral Reefs* **31**, 1063-1063.
- 559 [53] Madin, J., Allen, A., Baird, A., Pandolfi, J. & Sommer, B. 2016 Scope for latitudinal extension  
560 of reef corals is species specific. *Frontiers of Biogeography* **8**.

561

562 Table 1. Phylogenetic signal in ecological traits of corals in the subtropical-to-temperate transition  
 563 zone in eastern Australia using Pagel's  $\lambda$  or D as the test statistic. P (random) and P (Brownian)  
 564 indicate significance levels of departure from random and from Brownian phylogenetic structure,  
 565 respectively ( $P < 0.05$ ). Not significant departure from Brownian structure indicates conserved trait  
 566 evolution.

567

<b>Trait</b>	<b>Test statistic</b>	<b>P (random)</b>	<b>P (Brownian)</b>
Larval development rate	$\lambda = 0.997$	$< 0.001$	n.s.
Corallite size	$\lambda = 0.980$	$< 0.001$	n.s.
Sexual pattern (hermaphroditic / gonochoric)	$D = -0.522$	$< 0.001$	n.s.
Symbiont transmission	$D = -0.513$	$< 0.001$	n.s.
Laminar morphology	$D = -0.340$	$< 0.001$	n.s.
Meandroid corallite formation	$D = -0.128$	$< 0.001$	n.s.
Encrusting-to-massive morphology	$D = -0.160$	$< 0.001$	n.s.
Branching-to-tabular morphology	$D = -0.093$	$< 0.001$	n.s.
Reproductive mode (brood / spawn)	$D = 0.009$	$< 0.001$	n.s.
Ceriod corallite formation	$D = 0.033$	$< 0.001$	n.s.
Plocoid corallite formation	$D = 0.058$	$< 0.001$	n.s.
Tabular morphology	$D = 0.146$	$< 0.001$	n.s.
Arborescent morphology	$D = 0.554$	0.04	n.s.

568

569 **Figure captions:**

570

571 Figure 1. Conceptual representation of the hierarchical filtering assembly model tested for high-  
572 latitude reefs of eastern Australia. Coral species from the Great Barrier Reef species pool (A) pass  
573 through a regional filter e.g. based on species tolerances for marginal reef conditions (i.e. regional  
574 high-latitude species pool) (B), and subsequently segregate into local assemblages according to  
575 local biotic and/or abiotic filters that lead to clustered (more closely related) or overdispersed (more  
576 distantly related than expected) phylogenetic structure (C). There is substantial turnover of lineages  
577 from site to site (phylogenetic beta diversity) due to environmental gradients in the region (C).

578

579 Figure 2. Local tests for phylogenetic structure of coral assemblages arrayed along the subtropical-  
580 to-temperate transition in eastern Australia, showing net relatedness index (NRI, a and c) and  
581 nearest related taxon index (NTI, b and d) values using presence/absence (a and b) and abundance  
582 data (c and d). Positive values indicate phylogenetic clustering and negative values reflect  
583 phylogenetic overdispersion. Abbreviations for site names, with latitude increasing left to right: M  
584 = Mudjimba; IG = Inner Gneerings; F = Flinders; FR = Flat Rock; CI = Cook Island; JN = Julian  
585 Rocks Nursery; JF = Julian Rocks False Trench; NS = North Solitary Island; NR = North Rock;  
586 NWS = North West Solitary Island; WH = Woolgoolga; SWS = South West Solitary Island; SS =  
587 South Solitary Island; MB = Muttonbird Island; BR = Black Rock; CT = Cabbage Tree Island.

588

589 Figure 3. Non-parametric multidimensional scaling ordination of phylogenetic diversity patterns of  
590 corals on high-latitude reefs in eastern Australia. The ordination is based on the abundance-  
591 weighted net relatedness index (NRIab) and symbols represent patterns in phylogenetic structure at  
592 the study sites as indicated in the legend; see Fig. 2 for abbreviations of site names. Vectors identify  
593 traits that primarily discriminate assemblages: enc-mass = encrusting-to-massive morphology;  
594 branch-tab = branching-to-tabular morphology; rate = larval development rate; size = corallite size.

595 Photos portray assemblage types at the study sites. All photos by B. Sommer.

596

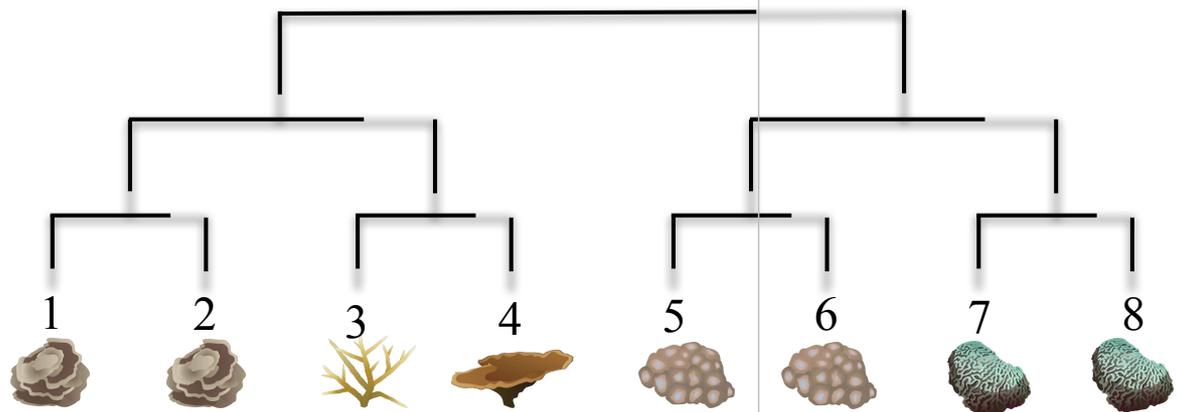
597 Figure 4. Phylogenetic betadiversity patterns of corals in the subtropical-to-temperate transition  
598 zone in eastern Australia. Standardised effect sizes for beta\_NRI (a-c) and beta\_NTI (d-f) using  
599 presence/absence (a, d) and abundance (b-c, e-f) data; pairwise turnover values plotted versus  
600 geographic distance (in degrees latitude; a-b, d-e) and boxplots for site level turnover patterns (c, f).  
601 The dashed line represents the expectation under a null model of random shuffling of taxa across  
602 the tips of the phylogeny. Positive values indicate higher than expected turnover of species in pairs  
603 of local assemblages, negative values indicate lower turnover than expected. See Fig. 2 for  
604 abbreviations of site names.

**A**

Great Barrier Reef species pool (phylogeny)

Species (1-8)

Trait (e.g. morphology)



REGIONAL FILTER e.g. high-latitude climate

**B**

High-latitude species pool

Species

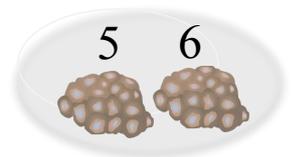
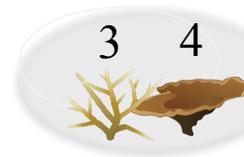
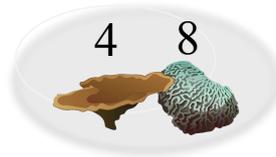
Trait



LOCAL FILTERS e.g. abiotic filtering, biotic interactions

**C**

Local communities (1-3)



Phylogenetic structure

overdispersed

clustered

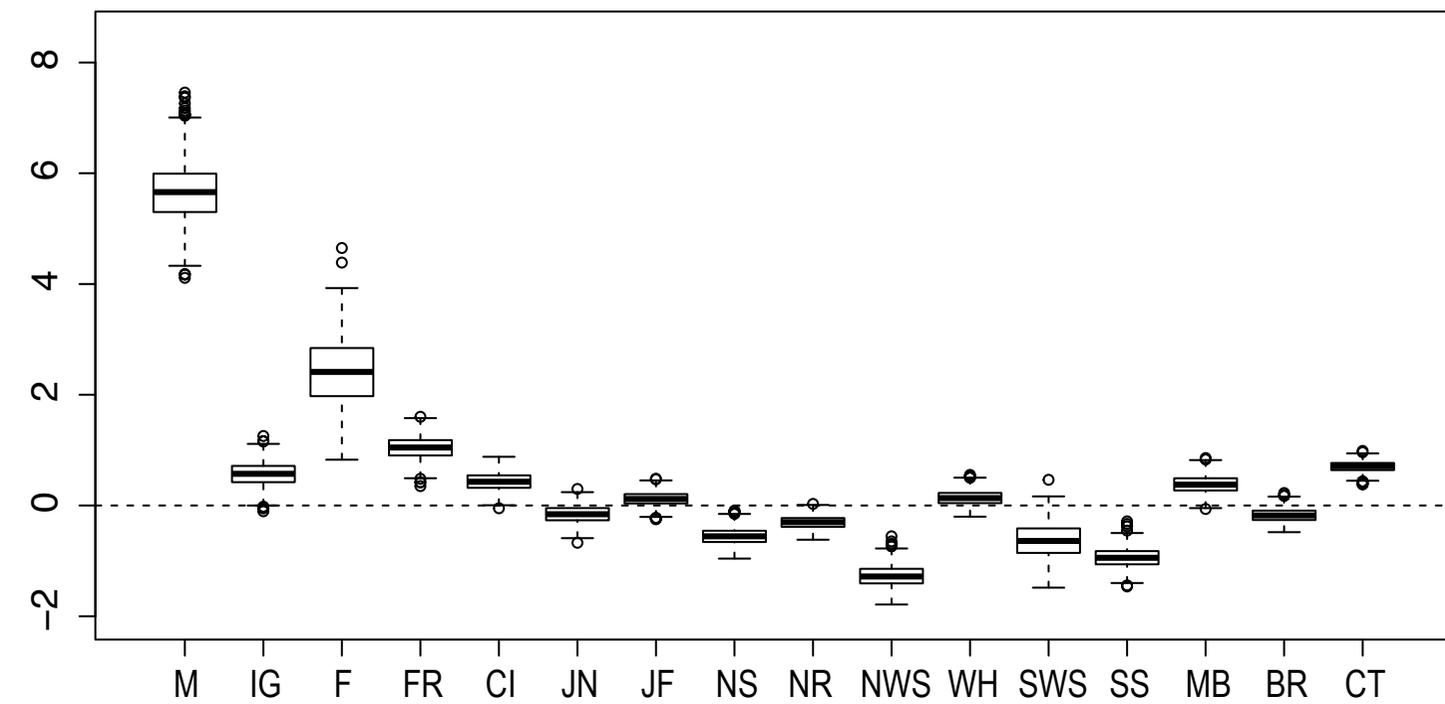
clustered

Phylogenetic beta diversity

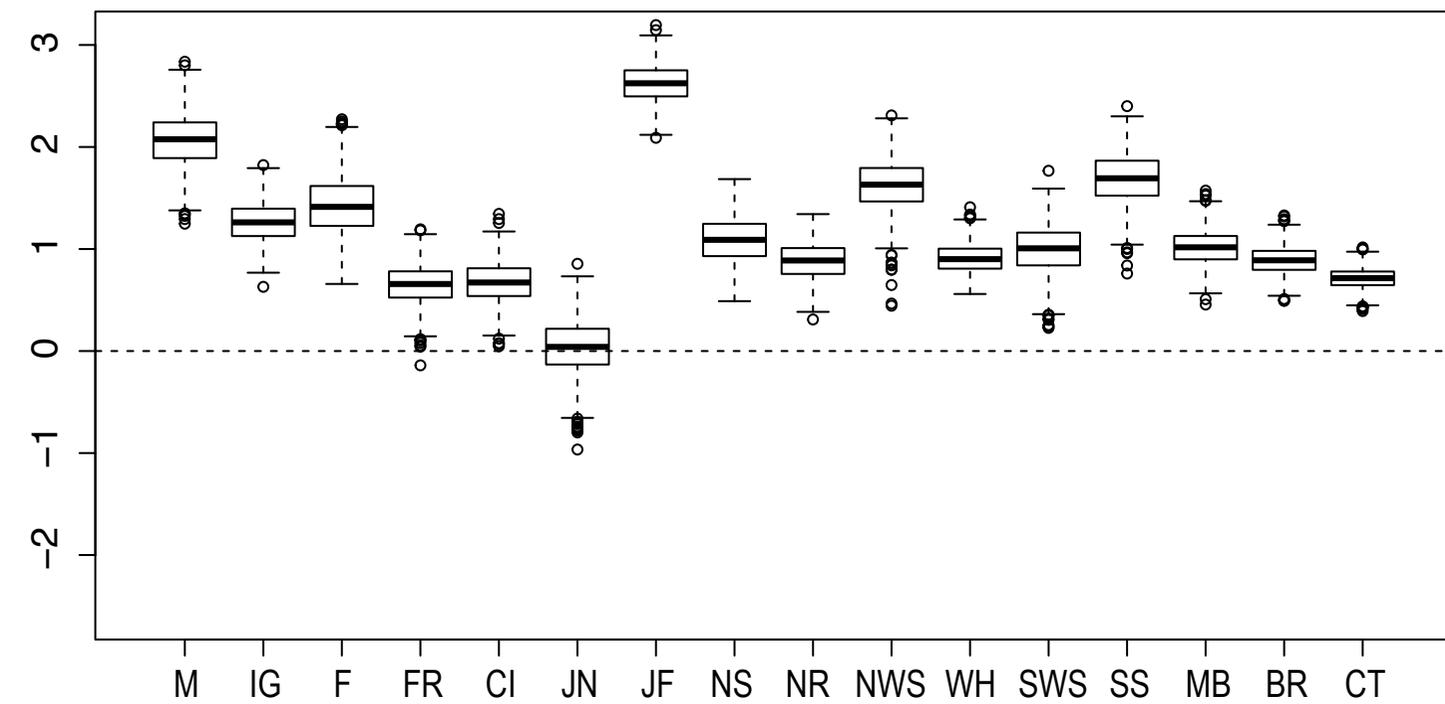
Turnover of lineages among site pairs



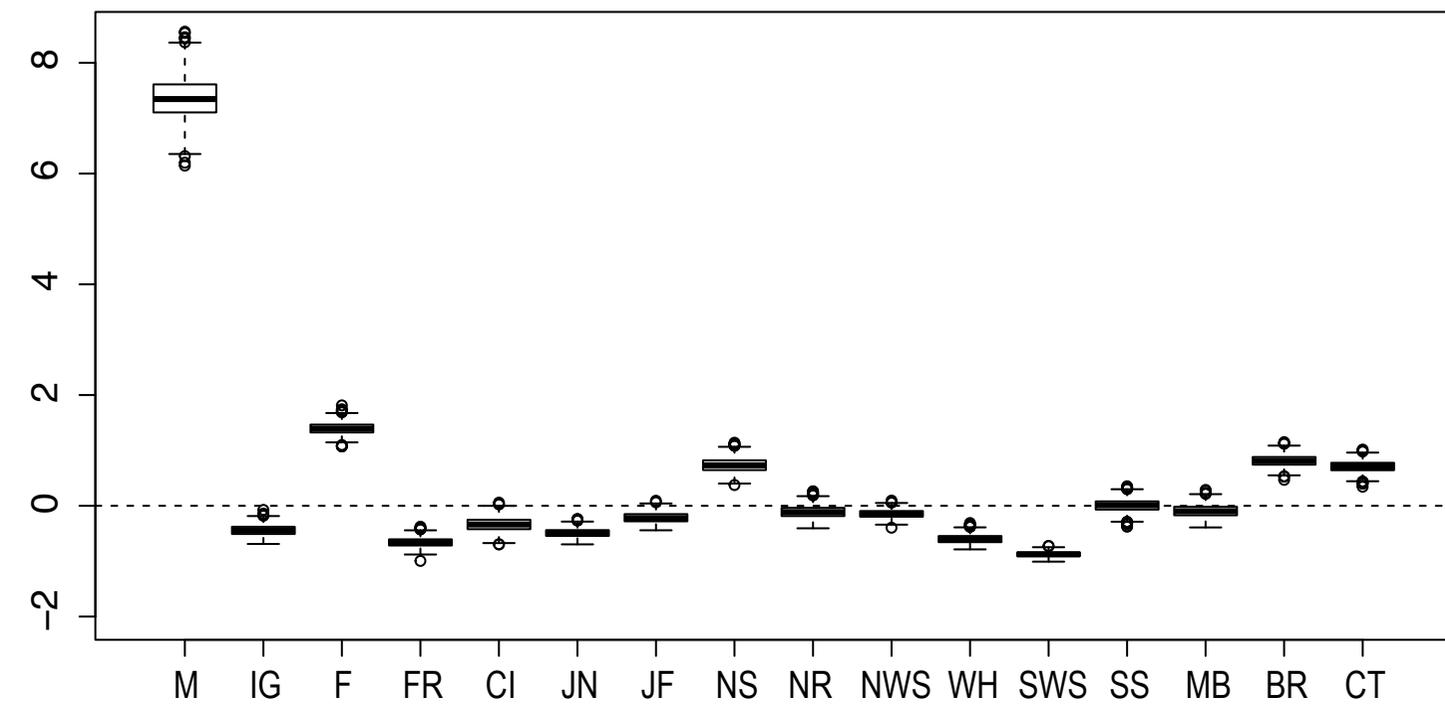
(a)



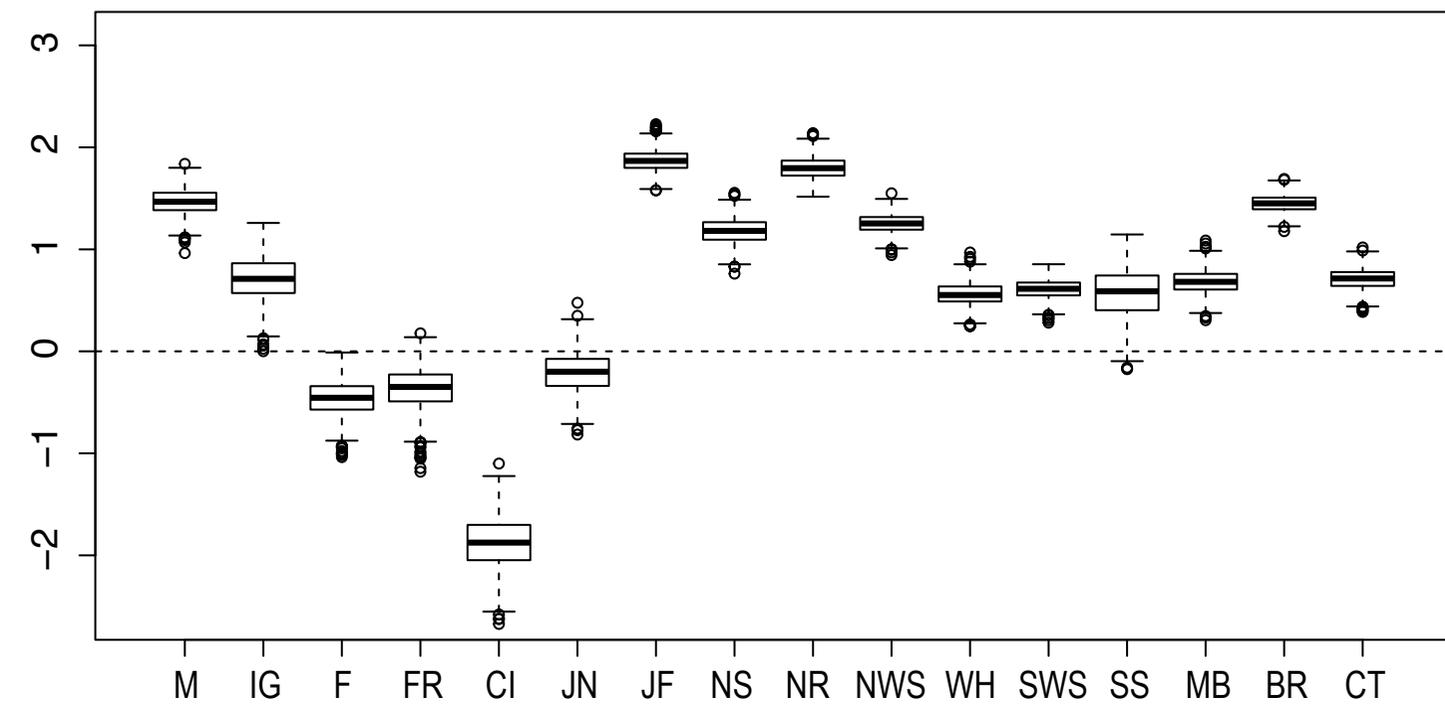
(b)



(c)



(d)



- NRI & NTI clustered
- ▲ NRI & NTI overdispersed
- ◆ NRI overdispersed, NTI clustered
- NRI clustered, NTI overdispersed

