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36 **ABSTRACT**

37

38 **Aim:** Conservation plans often struggle to account for connectivity in spatial prioritisation  
39 approaches for protecting migratory species. Protection of such species is challenging because their  
40 movements may be: uncertain and variable, span vast distances, cross international borders, and  
41 traverse land and sea habitats. Often we are faced with small samples of information from various  
42 sources and collection of additional data can be costly and timely. Therefore, it is important to  
43 evaluate what degree of spatial information provides sufficient results for directing management  
44 actions. Here we develop and evaluate an approach that incorporates habitat and movement  
45 information to advance the conservation of migratory species. We test our approach using  
46 information on threatened loggerhead sea turtles (*Caretta caretta*) in the Mediterranean.

47

48 **Location:** The Mediterranean Sea

49

50 **Methods:** We use Marxan, a spatially explicit decision support tool for selecting priority  
51 conservation areas. Four approaches with increasing amounts of information about the loggerhead  
52 sea turtle are compared, ranging from: i) the broad distribution, ii) multiple habitat types that  
53 represent foraging, nesting and inter-nesting habitats, iii) mark-recapture movement information, to  
54 iv) telemetry-derived migration tracks.

55

56 **Results:** We find that spatial priorities for sea turtle conservation are sensitive to the information  
57 used in the prioritisation process. Setting conservation targets for migration tracks altered the  
58 location of conservation priorities, indicating that conservation plans designed without such data  
59 would miss important sea turtle habitat. We discover that even a small number of tracks makes a  
60 significant contribution to a spatial conservation plan if those tracks are substantially different.

61

62 **Main Conclusions:** This study presents a novel approach for improving spatial prioritisation for  
63 conserving migratory species. We propose that future telemetry studies tailor their efforts towards  
64 conservation prioritisation needs, obtaining spatially dispersed samples over quantity. This work  
65 highlights the valuable information that telemetry research contributes to the conservation of  
66 migratory species.

67

68

69

## 70 INTRODUCTION

71

72 The increase in anthropogenic activities over the last two centuries has disrupted the movement of  
73 many organisms (Bolger et al., 2008; Harris et al., 2009). Migration and movement is essential for  
74 the persistence of many terrestrial and marine animals. Such species rely on movement between  
75 specific habitats or regions for reproduction, feeding, or thermal regulation (Alerstam et al., 2003).  
76 The destruction of movement pathways, and threats to individuals that move (e.g. bycatch), affect  
77 the fitness and survival success of migratory species (Beger et al., 2015). Protecting mobile species  
78 presents a great challenge due to the vast distances such animals often traverse, sometimes across  
79 international borders and in other cases between land and sea habitats (Martin et al., 2007). Yet,  
80 most conservation plans fail to incorporate the spatial connectivity that is needed to adequately  
81 protect migratory species (Martin et al., 2007; Runge et al., 2014).

82

83 Sea turtles are an example of an ecologically, economically and culturally important globally  
84 threatened migratory species group (IUCN, 2013). The thousands of kilometres these species travel  
85 between nesting and feeding habitats makes them highly vulnerable to an array of anthropogenic  
86 threats (Shillinger et al., 2010; Mazaris et al., 2014). These threats include, disturbance to nesting  
87 beaches from coastal development and sea level rise (Fuentes et al., 2011; Katselidis et al., 2014),  
88 turtle egg harvesting (Koch et al., 2006; Wallace et al., 2011), incidental catch in fishing gear  
89 (Lewison et al., 2004; Peckham et al., 2007), collision with boats, and the digestion of plastic  
90 material (Casale & Margaritoulis, 2010). Contributing to the vulnerability of marine turtles is their  
91 long life spans, reproductive age (e.g. loggerheads ~ 40-50 years old; Casale, 2011; Scott et al.,  
92 2012a; Avens & Snover, 2013) and different male versus female breeding patterns (Schofield et al.,  
93 2013a). Given the need for sea turtle protection and conservation, large-scale conservation plans  
94 that explicitly incorporate their complete habitat needs and migratory behaviours are lacking.

95

96 Previous sea turtle conservation efforts have primarily focused on protecting nesting sites (Casale &  
97 Margaritoulis, 2010; Mazaris et al., 2013). The central aim of these recovery efforts has been to  
98 protect female sea turtles and their nests, with little focus on males and the younger developmental  
99 stages (Schofield et al., 2013b). However, while some sea turtle populations are recovering  
100 (Tapilatu et al., 2013; Lamont et al., 2014), some continue to decline (Stewart et al., 2014; Weber et  
101 al., 2014), suggesting that there are limitations to a conservation approach that focuses on only a  
102 sub-set of the life-history stages. Population models indicate that conserving sea turtle nesting  
103 habitats alone without considering other key habitats is insufficient for species recovery (Heppell et  
104 al., 1996; Lazar et al., 2004). Currently, there are limited management actions (e.g. turtle exclusion

105 devices TEDs) to conserve sea turtles within marine waters and only recently have conservation  
106 efforts been directed towards protecting offshore sea turtle populations and their migration corridors  
107 (Pendoley et al., 2014; Seminoff et al., 2014; Baudouin et al., 2015). Successful conservation  
108 planning for sea turtles must explicitly protect all the life-stages and link their terrestrial and marine  
109 habitat requirements (Beger et al., 2015). One of the major impediments for minimising mortality in  
110 the sea is that information on the offshore distribution and movements of sea turtles is limited  
111 (Casale et al., 2007a).

112

113 Various methods have been trialled to understand sea turtle movement in offshore habitats. Since  
114 the 1950s, the most common method has been mark-recapture approaches, where tags are affixed to  
115 sea turtles at nesting sites and their location of recapture is documented (Carr & Giovannoli, 1957;  
116 Hendrickson, 1958; Caldwell et al., 1962). Mark-recapture methods have contributed to our  
117 knowledge of sea turtle migratory extent, links between release and capture sites (recaptures at sea;  
118 Casale et al., 2007b), nesting populations and growth rates (recaptures at the same nesting beaches;  
119 Monk et al., 2011). However this method is unable to provide information about entire migratory  
120 paths and remains labour-intensive (Stewart et al., 2013), characterised by low recapture rates  
121 (Avens & Snover, 2013) and slow knowledge accumulation (Godley et al., 2008). In recent  
122 decades, with the expansion of telemetry systems such as radio trackers, satellite transmitters and  
123 GPS loggers, tracking programs have proliferated (Godley et al., 2008; Hussey et al., 2015). These  
124 technologies actively improve our understanding of sea turtle migration pathways at sea (Pendoley  
125 et al., 2014; Stokes et al., 2015). While there is an increasing emphasis on telemetry to improve our  
126 understanding of sea turtles distribution, physiology and behaviour (e.g. Hochscheid et al., 2007;  
127 McCarthy et al., 2010), there is comparatively less attention paid to how this knowledge can  
128 improve management and identify conservation areas. Recent tracking studies link adult foraging  
129 grounds to existing MPAs and identifying new areas for protection (e.g. Scott et al., 2012b;  
130 Schofield et al., 2013a), however analyses that link habitat and movement information into spatial  
131 conservation prioritisations (Beger et al., 2015) remain scarce.

132

133 Sea turtle tagging and telemetry programs are rarely explicitly shaped by conservation planning  
134 objectives, and their execution is logistically difficult and expensive (satellite transmitters range  
135 from US\$2000-5000 each; Godley et al., 2008; seaturtle.org, 2013). Such information often remains  
136 in the sea turtle behaviour and ecology literature without any attempt to use it for conservation  
137 (Godley et al., 2008). Recent studies that have used telemetry to inform and improve conservation  
138 have been restricted to examining species movements (Stokes et al., 2015) and building distribution  
139 models (Schofield et al., 2013a). Presently, attempts to use sea turtle migration information to

140 enhance systematic conservation planning remain scarce (Beger et al., 2015), and the sensitivity of  
141 conservation outcomes to the number and quality of tracks used has never been assessed.  
142 Furthermore, conservation plans are being made for mobile species such as sea turtles often without  
143 considering the potential input that migration information could contribute (Martin et al., 2007;  
144 Runge et al., 2014).

145  
146 Here, we aim to develop and test approaches for incorporating information on habitat use and  
147 migration into conservation prioritisation for migratory species. The Mediterranean Sea and its  
148 endangered loggerhead sea turtle *Caretta caretta* (Linnaeus, 1758; IUCN, 2013) population provide  
149 an excellent case study for tackling this issue. We assess the potential impact of data limitations on  
150 conservation prioritisation outcomes by examining the value of different kinds of spatial  
151 information for identifying the location of areas that are a priority for sea turtle conservation.

152

153

## 154 **METHODS**

155

### 156 **Study area and database**

157

158 The study area was the entire Mediterranean Sea to a seafloor depth of 1,000 m<sup>1</sup>. We divided the  
159 resulting shallow Mediterranean Sea including coastal land areas with nesting beaches into planning  
160 units of 10 x 10 km, consistent with EU guidelines (Directive 2007/2/EC) and other large-scale  
161 regional planning studies (e.g. Mazor et al., 2014).

162

163 We assembled available sea turtle data (for data sources see Appendix 1) to create maps of three sea  
164 turtle habitat types (Fig. 1a).

165

166 Nesting habitat: First, the locations of 131 loggerhead nesting beaches were collated from over  
167 thirty published resources (Table S1 in Supporting Information). We did not aim to predict potential  
168 additional (unreported) locations of beaches using species distribution modelling methods because  
169 female sea turtles display natal homing and factors that affect their site selection within this homing  
170 range are not well known (Garcon et al., 2009). Planning units along the beach within a 10 km  
171 radius from each known nesting site were designating as nesting beach habitat. We note here that  
172 we did not aim to differentiate between major and minor nesting sites, but rather map the majority

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<sup>1</sup> Areas below 1,000 m were excluded because: a) most important foraging habitats for sea turtles in the Mediterranean Sea are generally classified in shallow waters along the continental shelf, b) anthropogenic threats are mainly concentrated along the coast and c) the General Fisheries Commission for the Mediterranean (GFCM) recommended the prohibition of towed dredges and trawl nets fisheries at depths beyond 1000 m (Recommendation GFCM/2005/1 on the ‘‘management of certain fisheries exploiting demersal and deep-water species’’) which has been adopted by the EU (Regulation 1967/2006).

173 of nesting sites (defined as sites averaging  $\geq 20$  nests per year to capture smaller nesting beaches) to  
174 represent the distribution of sea turtles.

175

176 Inter-nesting habitat: We created inter-nesting habitat data using a 10 km buffer from nesting  
177 beaches (Tucker et al., 1995; Waayers et al., 2011). These neritic areas are important habitat for  
178 female sea turtles during the time between laying clutches (Schofield et al., 2010) and for juvenile  
179 turtles making their way to the ocean post-hatching (Bolten, 2003).

180

181 Foraging habitat: Given that sea turtle foraging habitat is not yet fully mapped in the  
182 Mediterranean, we modelled foraging habitats using MaxEnt (Version 3.3.3k;  
183 <http://www.cs.princeton.edu/~schapire/maxent/> Phillips et al., 2004, 2006; Appendix S1 in  
184 Supporting Information). This model is intended as a simplified baseline representation of foraging  
185 grounds in the Mediterranean Sea as it incorporates location data from both adult and juvenile sea  
186 turtles. The MaxEnt species distribution modelling software models occupancy across space using  
187 presence-only species data. We collated sea turtle sighting locations from EurOBIS (2014), several  
188 scientific papers and location and telemetry data contributed by seaturtle.org (2013; Table S2).  
189 Telemetry data points that were spatially aggregated exhibiting high sinuosity on the continental  
190 shelf (defined by the 200 m isobaths; Kallianiotis et al., 2000; Sardà et al., 2004) were included,  
191 because such patterns indicate foraging (McCarthy et al. 2010; Dodge et al. 2014). Thus, transiting  
192 movements (and those off the continental shelf) were excluded, resulting in a total of 9,058 data  
193 points (see Fig. S1). These point data were combined with 22 environmental variables (for a list of  
194 variables see Table S3). The resulting model was validated by a random sub-sampling method that  
195 was repeated 15 times and used 25% of the data (Phillips et al., 2004, 2006). To create a  
196 distribution map of suitable foraging habitat we used the tenth percentile training presence logistic  
197 threshold ( $>0.36$ ). By using this threshold, we defined suitable habitat to include 90% of the data we  
198 used to develop the model. Our resulting map of foraging habitat was consistent with findings by  
199 localised studies that identified foraging grounds in the region (Broderick et al., 2007; Casale et al.,  
200 2013; Stokes et al., 2015).

201

202 Migration information: For our analyses of loggerhead turtle migration movements we compiled  
203 available satellite tracking data from EurOBIS (<http://www.eurobis.org/> 2014) and seaturtle.org  
204 (<http://seaturtle.org/>; Table S4). A total of 34 individual tracks were collected from a variety of  
205 sources across the Mediterranean Sea and were used in this study (Fig. 1b – individual tracks cannot  
206 be shown due to data protection; Appendix S3). More tracking data should be obtained if this

207 methods is to be used to robustly assign priority conservation areas for the regions sea turtle  
208 population.

### 209 **The value of sea turtle information for conservation**

210  
211 We examined the value of sea turtle information for conservation using scenario exploration with  
212 Marxan, a commonly used decision-support tool, and its derivative algorithm, Marxan with  
213 Connectivity (Beger et al., 2010a; 2010b). For each scenario (approach), we developed a set of  
214 spatial plans that met our conservation targets and connectivity objectives for the least possible cost  
215 (Ball et al., 2009). Below, we describe each planning approach highlighting the incorporation of  
216 additional data layers. To focus on the effects that different kinds of information have on spatial  
217 priorities, we kept the number of iterations (1000 runs) and the associated cost (equal cost per  
218 planning unit) consistent in all planning approaches.

219

220 The changes in spatial priorities signify the potential knowledge gained from investing in additional  
221 and more complex information. For new information to be useful for planning, it must improve our  
222 ability to make a decision or modify a plan (Maxwell et al., 2015). In the context of this analysis,  
223 we want to explore what information helps us better identify conservation priority sites that protect  
224 the entire turtle life cycle. First, we prioritise using the extant distribution range of sea turtles  
225 (Approach 1 - Range), then by multiple habitat types (nesting, inter-nesting and foraging,) (Approach 2 - Habitats), followed by movement information extracted from mark-recapture data  
226 (Approach 3 - Mark Recapture) and finally, the incorporation of satellite tracking data (Approach 4  
227 - Tracks). Within Approach 4, we tested the influence of the number of tracks used on resulting  
228 conservation priorities. Our conservation objectives to protect a given percentage of sea turtle  
229 spatial distribution (targets) varied according to approach (Table 1; Appendix S2).

231

232 We parameterised Marxan both without representing any connections between planning units  
233 (Approach 1 - Range, and Approach 2 - Habitats; Ball et al., 2009; Table 1) and by incorporating  
234 ecological connectivity into the objective function (Approach 3 - Mark-Recapture and Approach 4 -  
235 Tracks; Beger et al., 2010a; 2010b; Table 1). When including connectivity, we calibrated the  
236 Connectivity Strength Modifier (CSM - for methods see Beger et al., 2010b) to 50 (Fig. S2).

237

#### 238 Approach 1 - Range

239 In this approach we represented the overall distribution of loggerhead sea turtles by a single broad  
240 distribution map in the Mediterranean Sea, combining nesting, inter-nesting and foraging habitat  
241 data into one single distribution range (target was 20% of the species distribution) This is a basic

242 approach that is commonly used in conservation planning given the normal paucity of fine-scale  
243 spatial habitat data (e.g. IUCN distribution ranges).

244

#### 245 Approach 2 - Habitats

246 For this approach we set specific conservation targets for nesting (target 60%), inter-nesting (target  
247 40%) and foraging habitat (target 20%), simulating a situation where the three main habitats used  
248 by turtles are known. Dividing the broad distribution range into specific habitats with set targets  
249 ensures that priority conservation areas will be selected for each habitat type.

250

#### 251 Approach 3 - Mark-recapture

252 Mark-recapture studies define at least two points on a turtle's travel, its start (tagging location) and  
253 end points (recapture location). To represent this type of information in conservation planning, we  
254 targeted the three habitats used by turtles while also ensuring connectivity between nesting and  
255 foraging sites. Here, we simulated mark-recapture data using tracking routes (34 tracks) to select  
256 planning units associated with nesting beaches and foraging habitat. For this purpose, we  
257 considered foraging and nesting habitat to be planning units where tracks demonstrated sinuosity  
258 (obvious foraging behaviour; McCarthy et al., 2010) and overlapped with our modelled foraging  
259 grounds and our mapped nesting beaches (Fig. 1a). Tracks that did not move across more than 50  
260 planning units were discarded from the analysis as based on typical distances that Mediterranean  
261 loggerhead sea turtles move between nesting and foraging grounds (Zbinden et al., 2008; Schofield  
262 et al., 2013a). This analysis enabled us to allocate connectivity links between the identified foraging  
263 and nesting planning units at either end of the track, assuming non-directional connectivity in  
264 Marxan and ignoring the remaining tracked pathways (Beger et al., 2010b).

265

#### 266 Approach 4 - Tracks

267 To capture information about the pathways turtles take to cross vast distances and incorporate links  
268 between habitats along the entire journey, we applied a method that incorporates telemetry-derived  
269 movement information into Marxan with Connectivity (Beger et al., 2015). This approach allows  
270 for connectivity strength values to be assigned between and across sites by deriving a connectivity  
271 matrix that connects all planning units along each satellite track (Fig. 2). By symmetrically linking  
272 all planning units along an individual turtle's pathway, this method allows for spatial dependencies  
273 to exist between places that are not adjacent to each other (Beger et al., 2010b). Planning units that  
274 are travelled through by more than one individual turtle are deemed increasingly important for  
275 migration and contribute more to the connectivity of the solutions. Applying this method, we

276 targeted the three habitats (i.e. nesting, inter-nesting, foraging) used by turtles and the connectivity  
277 information provided from our 34 telemetry tracks (see Migration information).

278

279 Comparing planning approaches

280 We compared the four approaches by calculating Spearman Rank Correlation between the selection  
281 frequency outputs from Marxan, and mapping the resulting spatial conservation priorities. Selection  
282 frequency is the number of times that a planning unit is selected as part of a near-optimal solution in  
283 Marxan. This frequency can be seen as a measure of relative importance, where units selected a  
284 high percentage of times could be considered more valuable than those appearing less frequently in  
285 solutions.

286

287 We then tested how the number of telemetry tracks altered the resulting conservation plan. To  
288 investigate the value of new spatial information for identifying conservation priorities, we randomly  
289 selected an increasing number of tracks from the pool of known tracks; 0 (no tracks), 5, 10, 15, 20,  
290 25, 30, 34 (max). The Marxan analysis was repeated ten times for each group of tracks to account  
291 for variability in the selected tracks. From these solutions we calculated the Spearman rank  
292 correlation of the selection frequency outputs and compared it with that of a solution that includes  
293 all 34 tracks. To further examine the increased inclusion of telemetry tracks, we used a Bray-Curtis  
294 dissimilarity matrix method as described in Linke et al., (2012) and displayed our results in a  
295 dendrogram. This method compared the Marxan best solution outputs (solution with the lowest  
296 objective function score) when run with different numbers of tracks.

297

## 298 **RESULTS**

299

300 Conservation priorities that were evident in Approach 4 (Tracks) were not well represented in the  
301 other three approaches. For example, Approach 3 (Mark-Recapture), which had the highest  
302 Spearman rank correlation coefficient of the three approaches when compared with a plan that  
303 incorporates tracking data (Approach 4 – Tracks), indicated that the spatial priority areas from the  
304 plans do not significantly overlap ( $\rho = 0.08$ ). Thus, results show that links between habitats are  
305 not protected by chance when protecting sea turtle habitat, but need to be separately represented.

306

307 We found that conservation priorities substantially changed as we added different aspects of turtle  
308 information (Fig. 3a; Fig. 4). Despite the weak correlations, approaches that incorporated more  
309 habitat and movement information (e.g. Approach 2 - Habitats  $\rho = -0.12$  and Approach 3 - Mark-  
310 Recapture  $\rho = -0.23$ ) than a broad species distribution range (Approach 1 - Range  $\rho = -0.08$ ),

311 were more successful at capturing migration pathways (comparison with Approach 4 - Tracks) in  
312 the resulting spatial plans. Including movement data can also increase the cost of conservation plans  
313 as movement corridors may mean more area or costly planning units are needed to reach  
314 conservation targets (see Table S5).

315  
316 We found that when sample sizes are low, which is often the case with tracking sea turtle and other  
317 large marine animals, even a small number of tracks (~5) can substantially increase the correlation  
318 ( $\rho = 0.6$ ) with plans that include all thirty-four tracks (Fig. 3b). We discovered that the largest  
319 Bray-Curtis dissimilarity was between conservation plans that did include sea turtle tracks and those  
320 that did not (see Group A vs. Group C in Fig. 5). The second largest dissimilarity was between  
321 plans that had a low number of tracks (Group B and Group D in Fig. 5) and a corresponding low  
322 Spearman rank correlation ( $\sim \rho < 0.7$  Table S6) when compared with solutions that included  $\geq 20$   
323 tracks and resulted in a higher Spearman rank correlation ( $\sim \rho > 0.7$ ; Group C in Fig. 5). This  
324 dissimilarity was due to the low number of tracks (5-15 tracks) included in the plans and because  
325 the spatial variability captured was insufficient for the entire region. Given these results it seems  
326 that plans with  $> 20$  tracks were needed to capture the spatial heterogeneity of turtle movement  
327 across the Mediterranean Sea from our given sample size (34 tracks). Thus, plans with over twenty  
328 tracks did not vary considerably to those with 34 tracks.

329

## 330 **DISCUSSION**

331

332

333 We demonstrated that migratory pathways provide critical information for identifying habitats for  
334 inclusion in spatial planning. We discovered that the inclusion of satellite tracking data makes a  
335 substantial difference to spatial priorities. Moreover, prioritisation without the use of such tracks is  
336 sub-optimal for wide ranging species that move between multiple habitats.

337

338 This study highlights the value of incorporating critical habitat and migration information for  
339 conservation planning of migratory species. Our example system of loggerhead sea turtles in the  
340 Mediterranean Sea showed significant changes in spatial priorities when increasing the amount of  
341 sea turtle information (see four approaches; Fig. 3; Fig. 4). Sea turtle migration was best captured  
342 by incorporating the entire movement track rather than critical habitat information (Approach 2 -  
343 Habitats), species range (Approach 1 - Range), or mark-recapture data (start and end points of  
344 movements; Approach 3 - Mark-Recapture; Fig. 3; Fig. 4). We managed to collate data from 34 sea  
345 turtle tracks in this study and discovered that even a small number of very different tracks (e.g. five)

346 can substantially alter conservation priority sites and help capture the known spatial extent of the  
347 migratory life cycle of sea turtles (Fig. 3b; Fig. 5). As new methods emerge, we suggest that future  
348 conservation plans for sea turtles and other migratory species should attempt to incorporate  
349 available habitat and telemetry data where possible.

350

351 Our results suggest that in order to capture sea turtle habitat connectivity in conservation plans, a  
352 good quantity of heterogeneous tracks across the study area is needed (Fig. 5). Our case study  
353 example in the Mediterranean with a limited sample size (34 tracks; Fig. S3), found that >20 sea  
354 turtle tracks that were widely sampled across the study region were able to capture sea turtle  
355 movement. While we stress that more data is always better and higher sample sizes are preferable,  
356 such information is not always readily available and conservation decisions are often made with  
357 scarce data (Bottrill et al., 2008). This study suggests that limited data that is well dispersed across  
358 the study region can actually contribute valuable information to begin conservation planning. Given  
359 our findings that more heterogeneously placed tracks provide the best value of information, future  
360 data collection efforts could be made more useful for conservation by taking a complimentary  
361 sampling approach, and targeting regions that currently have fewer or no tracking studies (e.g. the  
362 eastern Mediterranean; Fig. 1b; Stokes et al., 2015).

363

364 Telemetry studies provide a wealth of connectivity information that is not often applied to  
365 conservation planning. We found that a limited but heterogeneous assemblage of tracks makes a  
366 substantial contribution to improve a spatial conservation plan towards better representing turtles'  
367 life cycles. This result could perhaps provide better direction for the timely and costly collection of  
368 telemetry data. We recommend that currently available telemetry data be extracted where possible,  
369 perhaps using monetary incentives or intellectual safeguards, and compiled into databases for the  
370 incorporation of species migration information into conservation plans. Established collaborative  
371 frameworks such as the EU, or the IUCN, could be potential starting points. Future work should  
372 aim to carry out value-of-information analyses (e.g. Maxwell et al., 2015; Canessa et al., 2015) in  
373 order to assess the trade-off between investing in the collection of more tracking data, or gaining  
374 new information for improved conservation outcomes. This type of analysis can help inform cost-  
375 effective conservation decisions.

376

377 Another challenge in addressing species movements is determining how much connectivity  
378 information is needed. Relying on too few tracks means there is also a risk of over-fitting to a  
379 limited number of data tracks. As an attempt to overcome these challenges, this study used a  
380 calibration method where planning units that contained a track were selected over 50% of the time

381 (Fig. S2). The method ensures that connectivity is represented, but it does not necessarily mean that  
382 50% of all migration links are captured in the solution. Determining the level of connectivity that is  
383 needed will largely depend on the species of interest as well as the conservation budget and  
384 conservation objectives. For example, connectivity is especially important for sea turtles that exhibit  
385 high mortality rates within movement pathways (Lewison et al., 2004; Casale, 2011). However,  
386 connectivity may not be particularly useful for species that are less threatened during the  
387 movement/migration phase or those that have large dispersal patterns without clear migration  
388 trajectories. Importantly, the area and cost of a conservation plan are likely to increase as the  
389 importance of connectivity is increased (Table S5). Hence, we suggest that the level of connectivity  
390 required could be pre-determined and a measure of minimum connectivity should be set per species.

391

392 This study demonstrates and tests a method for prioritising the conservation of migratory species.  
393 However, such an approach could be built upon to provide priority areas for sea turtle conservation  
394 in the region. A suitable conservation plan should aim to incorporate all available telemetry studies  
395 (e.g. the 195 tracks identified by Luschi & Casale (2014)), comparable and consistent data for sea  
396 turtle habitat across the Mediterranean region, robust species distribution modelling, as well as the  
397 associated cost of conservation actions (Carwardine et al., 2008). This study has touched on several  
398 of these requirements however a comprehensive data pooling from organisations and scientific  
399 literature is required if priority for the region are to be robustly and transparently determined. Our  
400 method here explored connectivity between nesting and foraging grounds however other  
401 connectivity should be included such as links between breeding sites, wintering habitats and  
402 developmental grounds (Casale et al., 2013; Schofield et al., 2013a). Similarly, migration tracks  
403 should be evaluated by different age classes, sexes and weighted by direction of usage and the  
404 number of individuals that it represents as a proportion of the entire region.

405

406 In summary, this study highlights the value of habitat and movement information to advance the  
407 conservation of migratory species. Our findings on loggerhead sea turtles of the Mediterranean Sea  
408 are expected to provide one example of a broader application for the protection of migratory  
409 species. We recommend future research aims to incorporate and evaluate the value of telemetry  
410 information into conservation plans for migratory species (Runge et al., 2014), especially those that  
411 are threatened, to ensure that mortality is reduced across their whole life cycle. Determining the  
412 value of investing in the collection of more spatial data for species or extracting information from  
413 existing resources can help inform spatial planning more immediately. When there is only a short  
414 window of time to act for threatened species it is critical that decision makers invest and act in areas  
415 which will be most effective at ensuring species persistence (Bottrill et al., 2008).

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727 References for ‘Supporting Information’ for Mazor et al. are found at the end of Appendix S3.  
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## BRIEF TITLES OF SUPPORTING INFORMATION

**Table S1.** Nesting habitat: A total of 131 loggerhead (*Caretta caretta*) nesting beaches were recorded from the following literature.

**Table S2.** Foraging habitat: References for data extracted from EurOBIS (2014), scientific literature and seaturtle.org (2013) to collect point data (9058 point locations) on sea turtles when foraging.

**Table S3.** Environmental Variables (Variables included in final model marked with \*)

**Table S4.** Migration information: A total of 34 sea turtle tracks were obtained via EurOBIS (2014) and seaturtle.org (2013). All data extracted from these sources is reference below.

**Table S5.** The opportunity cost of each scenario (cost is assumed equal for each planning unit). The Connectivity Strength Modifier (CSM; Beger et al., 2010b) was calibrated to 50 (Fig. S1). All values in the table represent the average value when run in Marxan 1000 times. The “number of planning units” indicates the number of 10 x 10 km units needed for reservation to meet biodiversity targets.

**Table S6.** Spearman rank correlation coefficient when running conservation plans in Marxan with different numbers of sea turtle tracks (0, 5, 10, 15, 20, 25, 30, 34). The selection frequency outputs from Marxan were compared against a solution with all 34 tracks included. These values indicate the similarity between spatial priorities in the solutions. We tested the number of tracks with 10 repetitions to test for variation between selected tracks in our random samples (indicated by a letter).

**Figure S1.** Map of 9058 data points (data supplied by reference Table S2) used to construct the foraging habitat model as described in full detail in Appendix S1.

**Figure S2.** Graphs showing the trade-off curve of the connectivity strength modifier (CSM) with the number of connected planning units (those containing a sea turtle track). By assessing a trade-off curve with the number of planning units that overlap with tracking data we could determine the appropriate Connectivity Strength Modifier (CSM - Beger et al., 2010b). We aimed for planning units containing tracks to be selected >50% of the time when run 1000 times in Marxan. We used a CSM of 50 (equal cost per planning unit).

**Figure S3.** Graphs showing the length (km) of each of the 34 tracks used in this study. See Table S4 for the sources of the 34 tracks.

**Appendix S1.** Sea turtle foraging distribution model created using MaxEnt.

**Appendix S2.** Setting conservation targets

**Appendix S3.** Information for each sea turtle track. The start and end country that the tracks were found, starting positions were usually nesting sites. Further information is unable to be given due to data privacy.

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**BIOSKETCH**

Tessa Mazor is a research fellow at The Commonwealth Scientific and Industrial Research Organisation (CSIRO). This work was carried out during her PhD at the University of Queensland, and the Centre of Excellence for Environmental Decisions (CEED, <http://ceed.edu.au/>). Her research interests include conservation planning for threatened marine species, the application of systematic planning tools in the marine realm and the development of sustainable management practises for marine ecosystems.

Author contributions: All authors conceived the ideas and contributed to the writing; T.M. conducted the analysis and led the writing.

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## APPENDIX 1 – DATA SOURCES

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1121 **TABLES**

1122

1123 **Table 1.** Summary of the planning approaches, including increasing amounts of data and  
 1124 information on the distribution and movement of sea turtles. Each plan aims to derive conservation  
 1125 priorities for loggerhead sea turtles (*Caretta caretta*) in the Mediterranean Sea, and uses systematic  
 1126 conservation decision tool Marxan.  
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<b>Approach for sea turtles conservation planning</b>	<b>Targets</b>	<b>How connectivity was incorporated</b>
<b>1.</b> Range	The distribution of sea turtles as a whole (not per habitat type) overall target = 20%	Not at all
<b>2.</b> Habitats	Nesting = 60% Inter-nesting habitat = 40% Foraging habitat = 20%	Targets for habitats used in different life-stages
<b>3.</b> Mark-Recapture	Nesting = 60% Inter-nesting habitat = 40% Foraging habitat = 20%	Connections between the priority habitats
<b>4.</b> Tracks	Nesting = 60% Inter-nesting habitat = 40% Foraging habitat = 20%	Connections between each track is prioritized

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1150 **FIGURE LEGEND**

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1153 **Figure 1**

1154 a) Three types of loggerhead sea turtle (*Caretta caretta*) habitat: nesting habitat, inter-nesting  
1155 habitat and foraging habitat. b) Map of the Mediterranean Sea divided by geographical sub-areas as  
1156 determined by the General Fisheries Commission of the Mediterranean Sea (GSCM). The total  
1157 number of sea turtles tracks that cross each sub area were calculated and represented in this map.  
1158 Individual tracks were unable to be displayed due to data confidentiality reasons, see Appendix S2  
1159 for further information on data sources.

1160

1161 **Figure 2.** Assignment of connectivity values derived from sea turtle telemetry paths. The squares  
1162 correspond to planning units of this study (10 x 10 km; consistent with EU guidelines (Directive  
1163 2007/2/EC) and other large-scale regional planning studies (Levin et al., 2013; Mazor et al., 2013;  
1164 Mazor et al., 2014) and result in a connectivity matrix.

1165

1166 **Figure 3.** a) Spearman rank correlation of selection frequency outputs, comparing four conservation  
1167 plans with increasing data complexity on sea turtle movement and habitat: Approach 1 - single  
1168 species distribution range, Approach 2 - habitat differentiation (nesting, inter-nesting, foraging),  
1169 Approach 3 – three habitat types and movement information from mark-recapture data, and  
1170 Approach 4 – three habitat types and movement information from 34 sea turtle tracks. b) Graph of  
1171 the average Spearman rank correlation of selection frequency outputs, comparing scenarios with a  
1172 subset of tracks vs. scenarios with all 34 tracks. The standard deviation is shown for each scenario  
1173 (calculated from ten repeated Marxan runs). This analysis used an equal cost for each planning unit.

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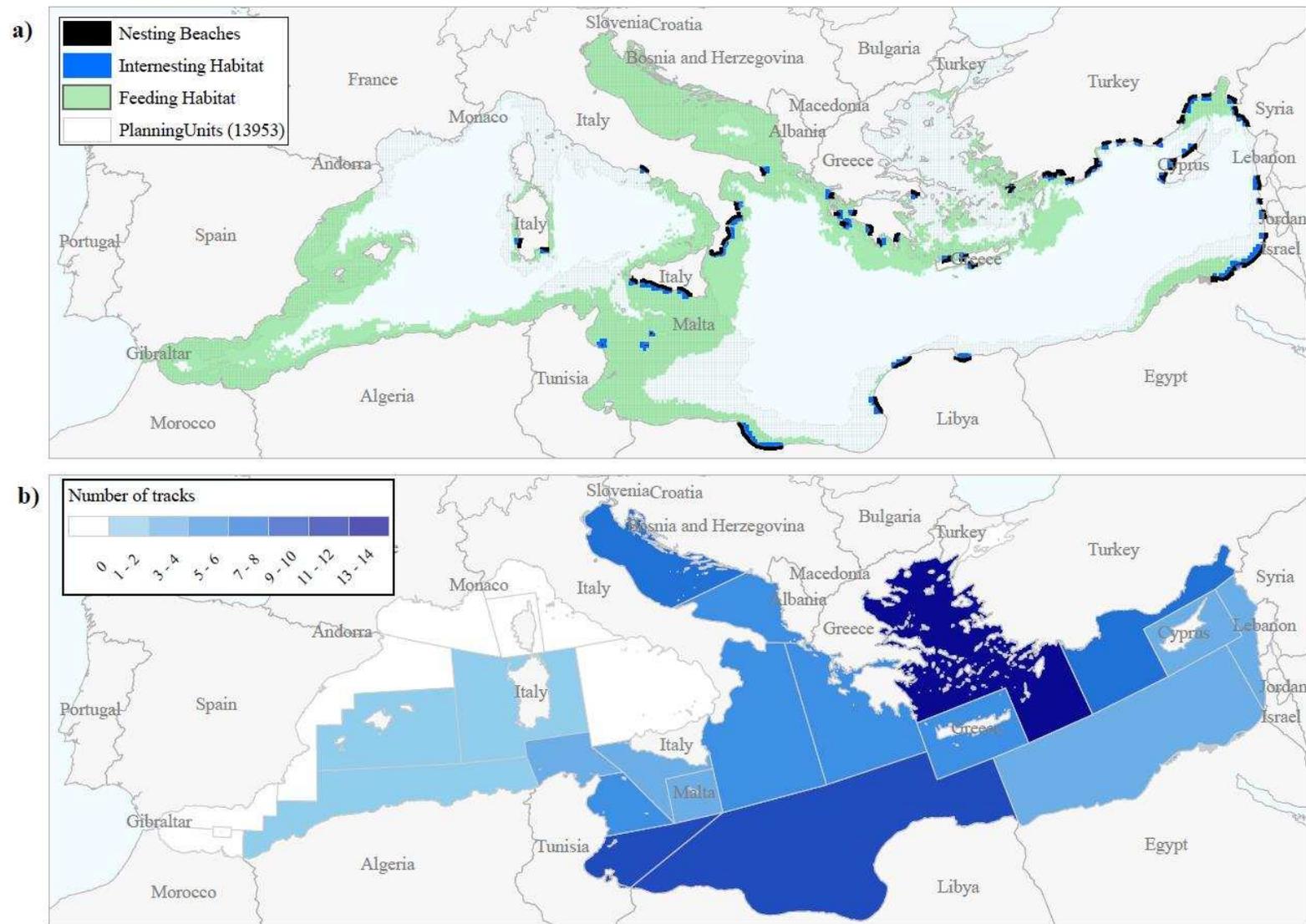
1175 **Figure 4.** Maps of four conservation plans in the Mediterranean Sea with increasing data  
1176 complexity for sea turtle movement: Approach 1 - Range, Approach 2 - Habitats (nesting, inter-  
1177 nesting, foraging), Approach 3 – Mark-Recapture data, and Approach 4 – Tracks (34 telemetry  
1178 tracks). Priority areas are those planning units that have a high percentage of selection (selection  
1179 frequency).

1180

1181 **Figure 5.** Dendrogram comparing the dissimilarity of solutions (Bray-Curtis dissimilarity matrix  
1182 method; Linke et al., 2012) with increasing numbers of tracks. Each node on the dendrogram  
1183 represents the number of tracks (0, 5, 10, 15, 20, 25, 30, and 34 tracks) used in the analysis and the  
1184 repetition letter (each number of tracks was run 10 times each as represented by letters a – j). These

1185 letters and numbers link to Supporting Information Table S6. Four groups were identified as  
1186 denoted by cycles and letters A, B, C, D. The main split between solutions is between analyses  
1187 without tracks and those that include tracks (Group A and B).

## FIGURES



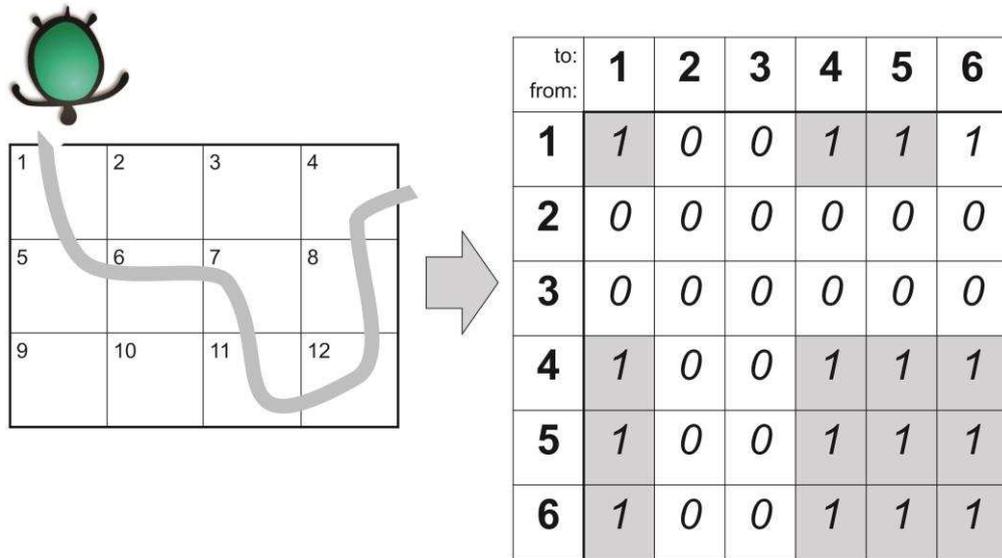
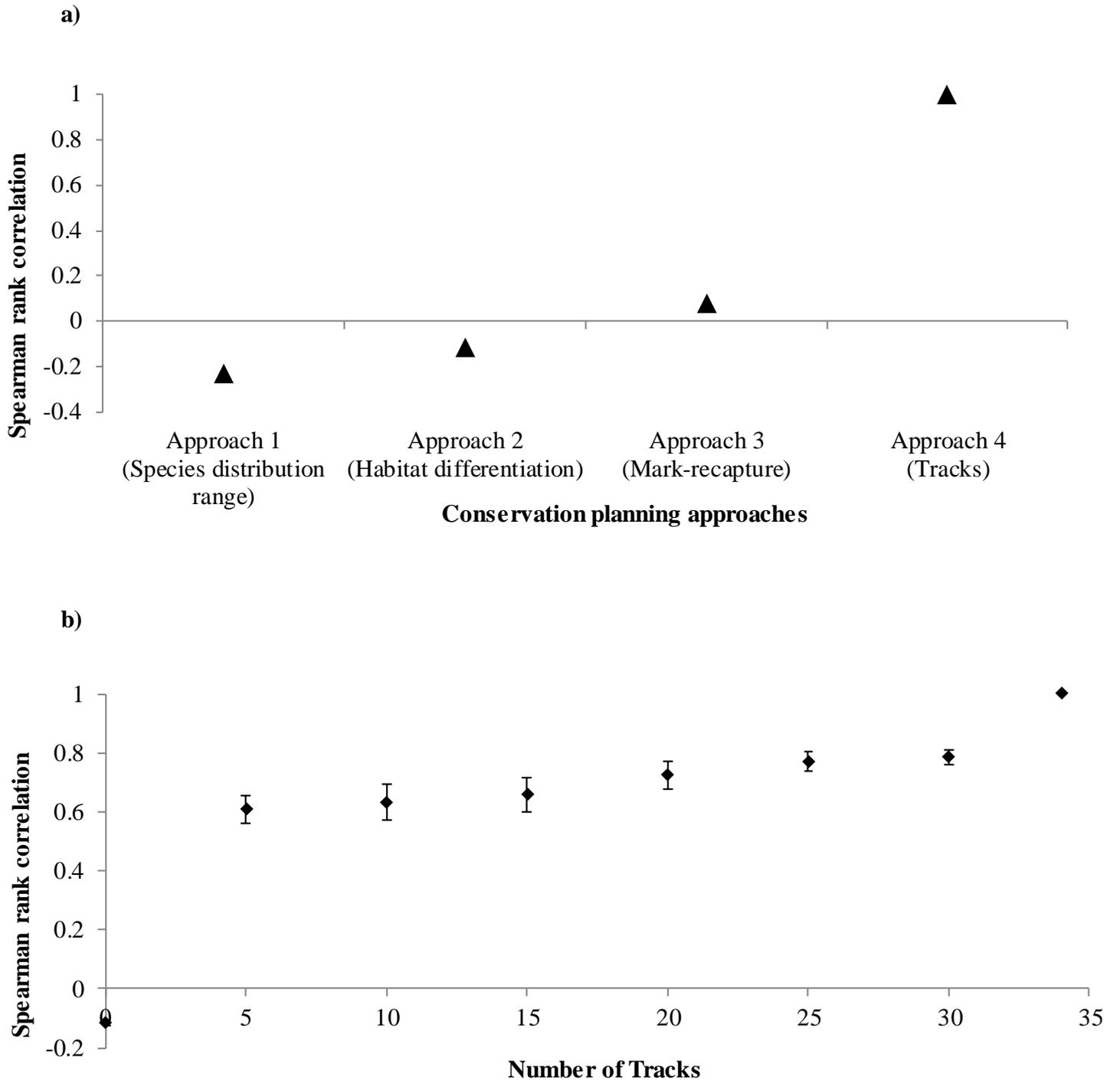
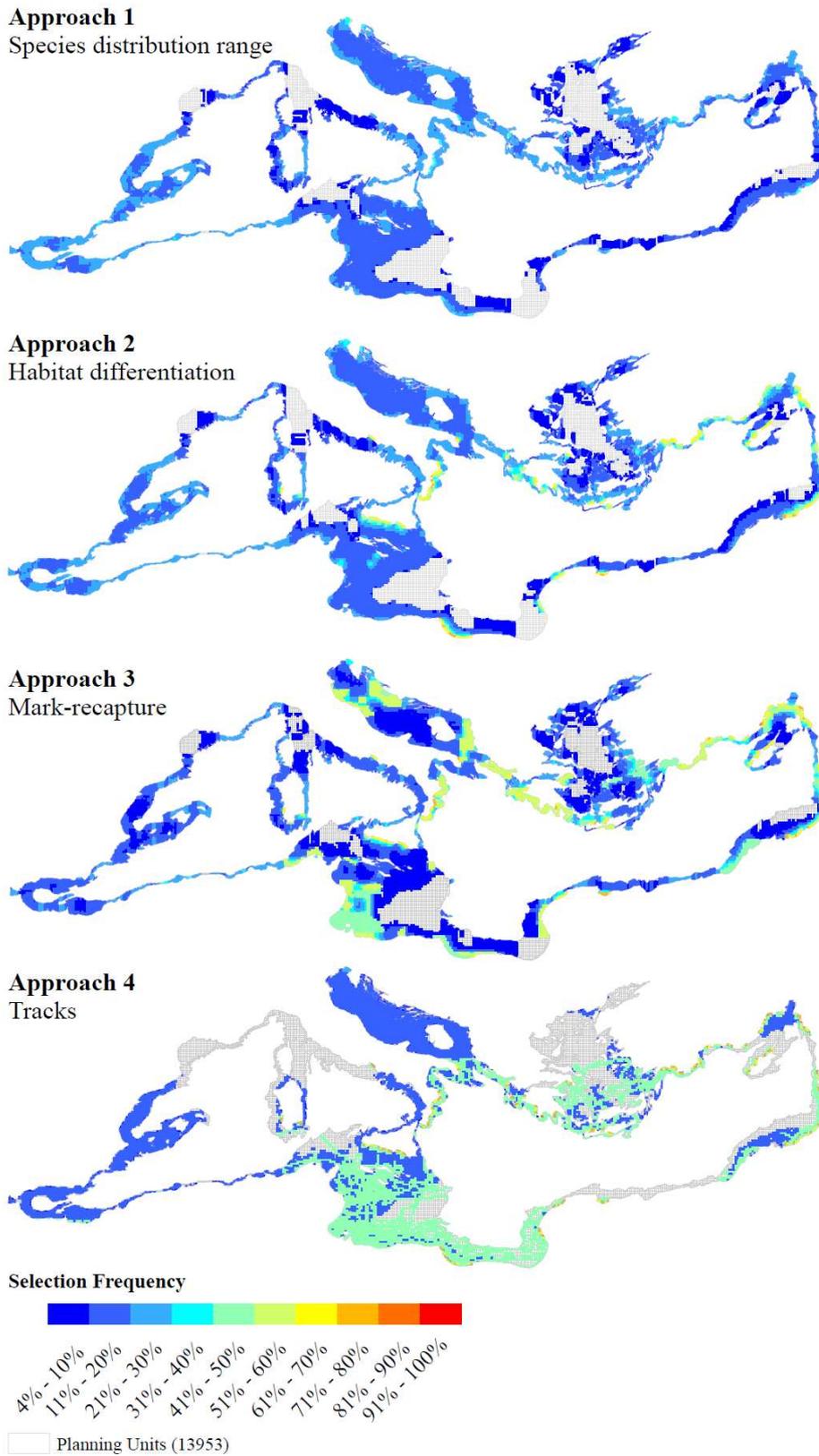


Figure 2.



**Figure 3.**



**Figure 4.**

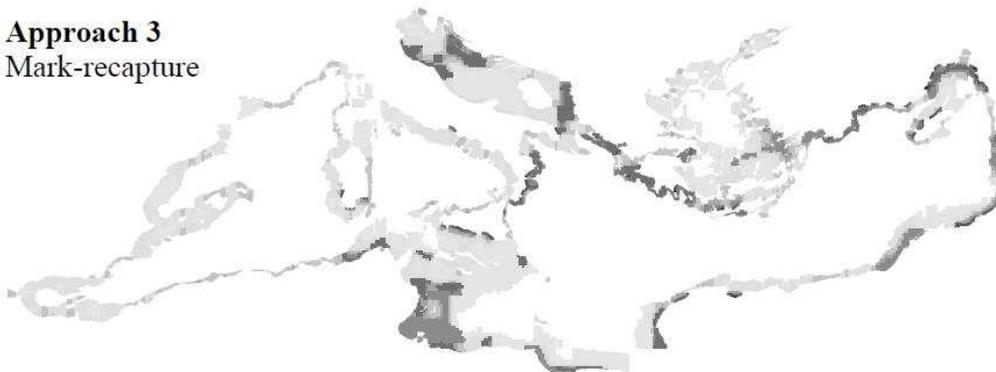
**Approach 1**  
Species distribution range



**Approach 2**  
Habitat differentiation



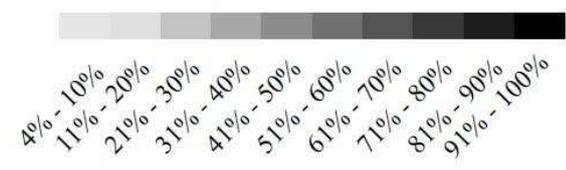
**Approach 3**  
Mark-recapture



**Approach 4**  
Tracks



**Selection Frequency**



**Figure 4.**

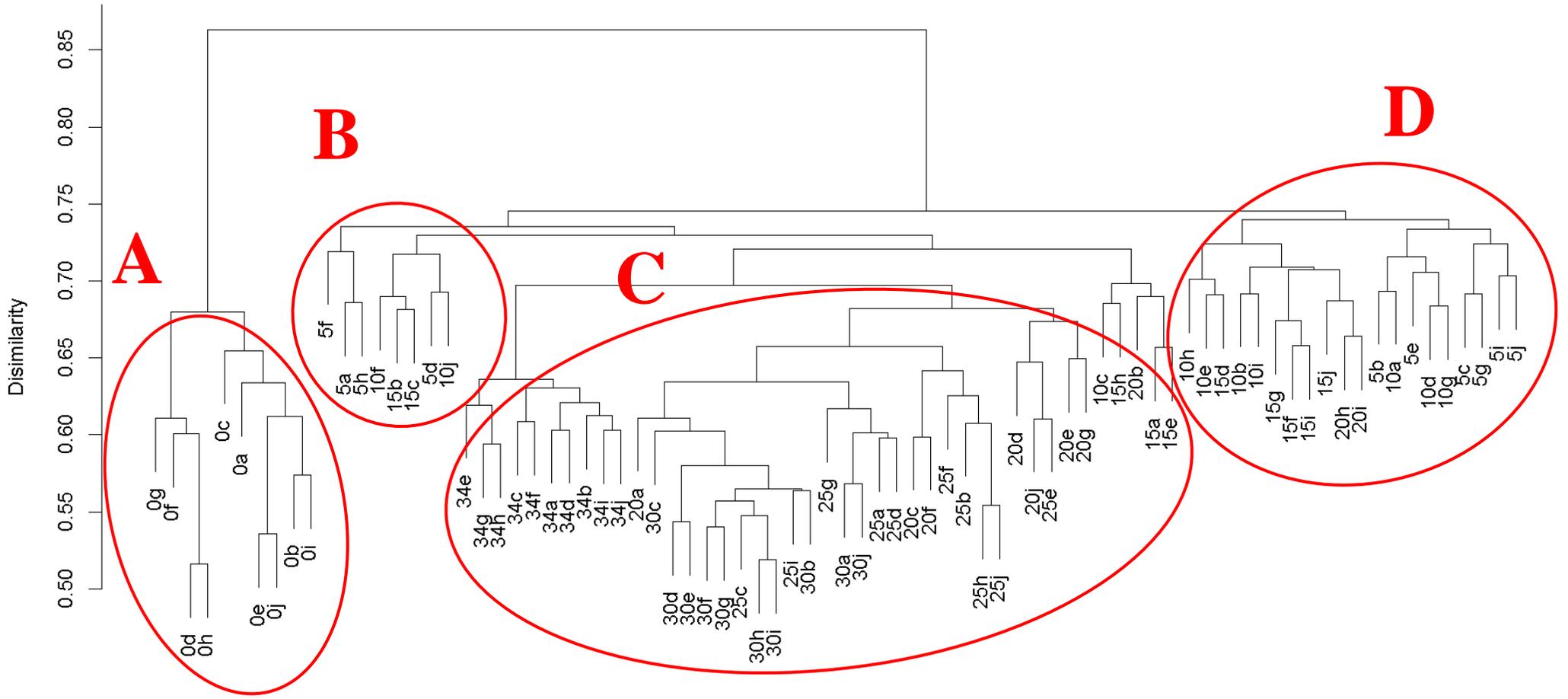


Figure 5.