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1 **TITLE:** Global impacts of energy demand on the freshwater resources of nations

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19 **CLASSIFICATION:** Environmental Sciences, Sustainability Science

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23 Wrote paper – RH, KS, FE with contributions from all authors.

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26 **ABSTRACT:** The growing geographic disconnect between consumption of goods, the
27 extraction and processing of resources, and the environmental impacts associated with
28 production activities makes it crucial to factor global trade into sustainability assessments.
29 Using an empirically validated environmentally extended global trade model we examine the
30 relationship between two key resources underpinning economies and human well-being -
31 energy and freshwater. A comparison of three energy sectors (petroleum, gas, electricity)
32 reveals that freshwater consumption associated with gas and electricity production is largely
33 confined within the territorial boundaries where demand originates. This contrasts with
34 petroleum, which exhibits a varying ratio of territorial to international freshwater
35 consumption depending on the origin of demand. For example, while the USA and China
36 have similar demand associated with the petroleum sector, international freshwater
37 consumption is three times higher for the former than the latter. Based on mapping patterns
38 of freshwater consumption associated with energy sectors at subnational scales, our analysis
39 also reveals concordance between pressure on freshwater resources associated with energy
40 production and freshwater scarcity in a number of river basins globally. These energy-driven
41 pressures on freshwater resources in areas distant from the origin of energy demand
42 complicate the design of policy to ensure security of fresh water and energy supply. While
43 much of the debate around energy is focussed on greenhouse gas emissions, our findings
44 highlight the need to consider the full range of consequences of energy production when
45 designing policy.

46

47 **SIGNIFICANCE STATEMENT:**

48 Understanding the role of international trade in driving pressures on freshwater resources is
49 key to meeting challenges at the water-energy nexus. A coupled trade and hydrological model
50 is used to examine pressures on freshwater resources associated with energy production
51 across the global economy. While the electric and gas sectors induce freshwater consumption
52 predominantly within countries where demand originates (91%, 81% respectively), the
53 petroleum sector exhibits a high international footprint (56%). Critical geographic areas and
54 economic sectors are identified providing focus for resource management actions to ensure
55 energy and freshwater security. Our analysis demonstrates the importance of broadening the
56 discourse on energy policy to include issues such as freshwater scarcity, the role of
57 international trade, and wider environmental and societal considerations.

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63 **Introduction**

64 Meeting society's demand for fresh water and energy has been identified as a major challenge
65 for society over the coming decades (1, 2). Most of the estimated 35 million km³ of fresh
66 water that exists globally is inaccessible (3). Recent estimates put renewable freshwater
67 resources in the region of between 40,000 and 66,000 km³ yr⁻¹ (4, 5), of which around 10% is
68 appropriated for human use (6–8). While this global total might be considered to fall within
69 the “safe operating space” of humanity (9), it hides substantial mismatches between
70 availability and demand in different regions (7, 8, 10) and associated pressures on renewable
71 freshwater resources (1, 11). Given that freshwater is central to maintain ecosystem function
72 (12) and biodiversity (13), pressures on freshwater resources can result in the loss of
73 ecosystem services (14–16) and associated benefits to society, ultimately impacting human
74 wellbeing both directly and indirectly (17–19).

75 Fresh water is used by the energy sector along the complete supply chain from extraction and
76 conversion of raw material through to generation of power (2, 20), such that limits on access
77 to fresh water through physical scarcity or regulatory control can have significant
78 implications for security of energy supply (21). At the same time, energy is needed for
79 extraction, treatment and distribution of fresh water (2) to meet societal demand. This
80 interdependence of fresh water and energy (2, 22–25) means that limits on one will impact
81 the other, potentially causing significant economic, environmental and social costs (23).
82 Despite growing recognition of the importance of this water-energy nexus (26), policy
83 objectives relating to fresh water and energy are often poorly integrated and concerned
84 primarily with exploitation of fresh water and/or implications of climate change on
85 freshwater resources required for energy production (27, 28). Indeed alignment of climate
86 and energy policy has led to the adoption of energy strategies that have the potential to
87 negatively affect freshwater resources (25, 27).

88 A key difference between energy and fresh water is the relative ease with which the former
89 can be transformed and transported between areas of production and demand (28). The
90 resulting geographic disconnect between sources of inputs associated with energy production
91 and final energy demand poses a significant challenge for resource management at the water-
92 energy nexus. Countries can implement policies that improve energy and freshwater resource
93 management within their own territories (23), with most developed countries exhibiting rapid

94 reform of both sectors in recent years (26). However, movement of energy resources around
95 the world coupled with increasing trade in “virtual water” (29), adds complexity to the
96 identification of policy and management options to ensure security of supply of both
97 resources along global supply chains.

98 There is an increasing understanding that international trade in natural resources driven by
99 rising national wealth and the opening up of commodity markets since the 1980s, has led to a
100 disconnect between final consumption of goods, and production activities such as the
101 extraction and processing of resources and associated environmental impacts (30). The
102 implications of this disconnect have been explored predominantly for greenhouse gas
103 emissions (31–34), but also for freshwater use (29, 35), land use change (36, 37), material use
104 (30) and biodiversity (38). Consistent across these studies is a bias in environmental transfers
105 in favour of net-importing developed nations at the expense of resource-exporting less
106 developed nations. For example, emissions saved by industrialised countries bound by
107 emissions reduction targets under the Kyoto Protocol were offset through emissions
108 associated with the import of goods from countries without such emissions targets (34). There
109 is considerable evidence to show that such carbon leakage can jeopardise climate targets (39),
110 and that carbon-importing countries gain more socio-economic benefits from international
111 trade than carbon-exporting countries (40).

112 Such studies make a compelling case to incorporate the transfer of resources through
113 international trade within national policies and sustainability assessments, so that the
114 implications of consumption of goods for environment and society can be fully considered
115 (30, 37). In the case of renewable freshwater resources, where impacts will be congruent with
116 areas of resource extraction or production of goods, understanding and locating the
117 geographic disconnect between use of fresh water and drivers of demand (29, 35, 41, 42) is
118 key for assessing sustainability. In the current study we investigate differences between
119 energy sectors in the magnitude and geographical distribution of consumption of renewable
120 freshwater resources, explore the geographical relationship between energy induced
121 freshwater consumption and the demand that drives it, and consider the implications in the
122 context of freshwater scarcity. We use a novel, empirically validated environmentally
123 extended multi-regional input-output (EE-MRIO) approach that is spatially resolved at
124 subnational scales. A spatially-resolved, comprehensive analysis is vital, as energy-driven
125 demand can be an important contributor to pressures on freshwater resources in localised
126 regions (21, 43). Our analyses focus on freshwater *consumption* as this represents loss of the

127 resource to the immediate environment (8) and so an opportunity cost in terms of ecosystem
128 benefits (44). We do not consider freshwater *withdrawal* which refers to fresh water removed
129 from a source and used for human activity before being returned to the environment (8). Our
130 analyses isolate freshwater consumption embodied in the three main energy sectors (gas,
131 electric and petroleum) globally, taking into account all processes along the supply chain
132 from material extraction, transformation to energy carriers, and distribution to final
133 consumers. Although a number of studies have examined the water-energy nexus at regional
134 and national scales using EE-MRIO techniques (24, 45) ours is the first to attempt such an
135 analysis at a global scale.

136 In the first stage of the analysis a MRIO table derived from the Global Trade Analysis Project
137 (GTAP) (46) (SI Appendix section 1) that quantifies economic transactions between 57
138 sectors across 129 countries/regions, is linked to data from the hydrological model
139 WaterGAP that provides freshwater consumption data associated with agricultural, energy,
140 domestic, and industrial activity (47–49) (SI Appendix section 2). The environmental
141 extension to the MRIO that this link provides allows us to reattribute direct sectorial
142 freshwater consumption following the trade transactions to the final consumer of a finished
143 commodity, a process known as footprinting (SI Appendix section 3). The approach to this
144 country/region-scale analysis is comparable to other studies that have examined international
145 trade as a driver of pressures on freshwater resources (29) but which have not specifically
146 addressed issues around the water-energy nexus. The second stage of analysis refines
147 country/region values for freshwater consumption calculated in the EE-MRIO to sub-
148 country/region scales (0.5×0.5 degree grid cell resolution) to describe spatial heterogeneity
149 in freshwater consumption (35) (SI Appendix section 4). This is a vital step, as locality is
150 critical to determining the implications of freshwater consumption given the uneven
151 distribution of renewable freshwater resources (7, 42). Based on this 0.5×0.5 degree grid
152 cell resolution data, patterns of freshwater consumption associated with energy demand are
153 considered within the context of available renewable freshwater resources in the world's river
154 basins (4) to identify areas of critical importance for security of fresh water and energy
155 supply (SI Appendix section 6).

156 **Results**

157 **Overview of freshwater consumption.** Before presenting the results of the EE-MRIO
158 analysis and considering freshwater consumption induced by the global energy sector from a

159 consumption based perspective, we provide a brief overview of the underpinning data to
160 place our analysis within the wider context of freshwater consumption associated with human
161 activity. Data from the WaterGAP model indicates that the crop sector dominates freshwater
162 consumption accounting for 91.85% (1237 km³ yr⁻¹) of the 1314 km³ yr⁻¹ of global annual
163 freshwater consumption. This figure correspond to findings in previous studies (35) that have
164 emphasised agricultural production as the principal driver of pressures on freshwater
165 resources globally. Industrial and domestic demand accounts for 5.88% (77 km³ yr⁻¹) of the
166 remaining freshwater consumption, again corresponding to findings stated in (35).

167 Of this industrial and domestic freshwater consumption 23.78% (or 1.40% of global total
168 freshwater consumption) is directly associated with energy sectors considered in this analysis.
169 Although this figure is comparatively small, the importance of considering freshwater
170 consumption associated with energy sectors arises for two reasons. Firstly, freshwater
171 consumption associated with energy extraction and refining may be highly locally
172 concentrated and so contribute to social, environmental and economic problems in specific
173 regions (21), a question we examine through our spatially explicit impact analysis. Secondly,
174 our assessment employs EE-MRIO analysis to calculate the sum of embodied freshwater
175 within all the products required to meet final demand in isolated energy sectors. Thus we
176 identify not only freshwater consumption associated with specific energy sectors (e.g. oil
177 extraction, oil refining, etc.) but also freshwater consumption associated with inputs required
178 by these sectors (e.g. steel production for infrastructure; crops for biofuel) that could
179 contribute to pressures on freshwater resources through higher intensities or in different
180 geographic areas than the directly energy related activities.

181 **Country/region energy-driven freshwater consumption footprints.** Our analysis finds that
182 when measuring total freshwater consumption along global supply chains, the electricity
183 sector consumes 6.48 km³ of freshwater per year, with the petroleum sector consuming 1.60
184 km³ yr⁻¹ and the gas sector 0.30 km³ yr⁻¹. For each of the 129 countries/regions within the
185 EE-MRIO, total freshwater consumption is disaggregated to describe the amount that occurs
186 within the country/region where demand originates (i.e. territorial consumption), and the
187 amount that is sourced internationally along energy supply chains (Fig. 1). The proportion of
188 internationally sourced freshwater consumption is highest for activity induced by the
189 petroleum sector (Fig 1A) at 56% of total consumption for this sector. For the electricity (Fig
190 1B) and gas (Fig 1C) sectors respectively, 9% and 19% of total sector-induced freshwater
191 consumption is sourced internationally. For the petroleum sector as a whole, the largest

192 consumers of fresh water are the USA ($0.34 \text{ km}^3 \text{ yr}^{-1}$), China ($0.29 \text{ km}^3 \text{ yr}^{-1}$) and India (0.19
193 $\text{ km}^3 \text{ yr}^{-1}$). Together, these three countries account for 50% of total freshwater consumption
194 within this sector. These countries exhibit markedly different patterns of territorial and
195 international consumption (Fig. 1). For the USA 73% of total freshwater consumption
196 associated with the petroleum sector occurs internationally, this contrasts with China where
197 22% occurs internationally, and India where there is an almost even division (52% territorial,
198 48% international).

199 Given that the USA and China have comparable total freshwater consumption associated with
200 their energy sectors (Fig. 1) we focus on the geographic and sectorial patterns of freshwater
201 consumption of these two in further detail, while noting that the technique can be extended to
202 all countries/regions (see SI Appendix, Fig. S1). Countries and sectors have been aggregated
203 for illustration purposes (Fig. 2), with the underlying analysis based on 129 countries/regions
204 and 57 sectors (see SI Appendix section S3). Consistent with the patterns shown in Fig. 1,
205 freshwater consumption by the petroleum sector in the USA is geographically diverse (Fig.
206 2A) occurring in northern America (27%), western Asia (29%), southern Asia (13%), eastern
207 Asia (7%) and northern Africa (6%). This contrasts with the Chinese petroleum sector (Fig.
208 2B) where 78% of freshwater consumption occurs within China, with the remainder
209 occurring mainly in other Asian countries/regions (13%) and in eastern Africa (4%). The
210 majority of freshwater consumption associated with the electricity (Fig. 2C & 2D) and gas
211 (Fig. 2E & 2F) sectors for the USA and China is located within the territory where demand
212 originates.

213 Utilisation of goods or services along the supply chain of energy provision is reflected in the
214 breakdown of freshwater consumption by sector of activity. For both the USA and China, the
215 EE-MRIO demonstrates that the majority of freshwater consumed to produce petroleum (Fig.
216 2A & 2B) is by the crop sector (76% and 44% respectively), the electric sector (12% and
217 10% respectively), the oil sector - relating to extraction of raw materials (2% and 16%
218 respectively), direct use in the petroleum sector itself (2% and 8% respectively), and to a
219 lesser extent, sectors relating to industry (e.g. metal and machinery production) and services
220 (e.g. insurance, banking, other support services). A similar pattern is found for the gas sector
221 (Fig. 2E & 2F), with crops (71% and 37% respectively for the USA and China) dominating.
222 In contrast, the majority of freshwater consumption by the electricity sector (Fig. 2C & 2D) is
223 associated with the sector itself (91% and 64% respectively for the USA and China),
224 followed by crops (8% and 19% respectively).

225 To illustrate the mechanism that drives the dominance of freshwater consumption associated
226 with crops within energy sectors (Fig. 2), the EE-MRIO was used to describe how an increase
227 in one unit (i.e. US\$1) of output of the USA petroleum sector induces production activities
228 and corresponding freshwater consumption to support them (SI Appendix section 3.2.). For
229 an increase in US\$1 of output from the USA petroleum sector, US\$2.52 of economic activity
230 is induced upstream in the global economy. This is associated with an additional $2,500 \text{ m}^3 \text{ yr}^{-1}$
231 of freshwater consumption. In economic terms, of the US\$2.52 of induced activity, 31% is
232 in the oil sector (extraction of materials), 45% in the petroleum sector itself (refining,
233 distribution etc.) and 1% in crop production. Expressed in terms of freshwater consumption
234 ($\text{m}^3 \text{ yr}^{-1}$), the one per cent of additional economic activity in the crop sector accounts for 76%
235 of the additional fresh water consumed. This contrasts with induced activity in the oil and
236 petroleum sectors which drive only four per cent of additional freshwater consumption but
237 account for three quarters of additional economic activity.

238 **Sub-country/region energy-driven freshwater consumption footprints for USA and**
239 **China.** Using the approach of (35), the global distribution of freshwater consumption
240 associated with the individual energy sectors in the USA and China was mapped to 0.5×0.5
241 degree grid cells (Fig. 3; SI Appendix, Fig. S3 and S4). Data at the country/region scale was
242 disaggregated based on intensity of freshwater consumption and location of economic
243 activity within each 0.5×0.5 degree grid cell corresponding to the economic sectors within
244 the EE-MRIO (see SI Appendix section 4) to reveal spatial heterogeneity within
245 countries/regions. Using the petroleum sector as an exemplar (Fig. 3), reveals a statistically
246 strong correlation between geographic patterns of freshwater consumption for the USA (Fig.
247 3A) and China (Fig. 3B) ($r = 0.98$, $F = 2776.78$, $df = 110$, $p < 0.001$). This correlative
248 relationship is likely driven by areas of common global resource extraction, manufacturing
249 and agricultural production across Asia, North Africa, Europe and the Americas. However,
250 there exist significant differences (Table S6) between the USA (Fig. 3A) and China (Fig. 3B)
251 in patterns of freshwater consumption in absolute terms driven by the higher international
252 demand on freshwater resources associated with the USA petroleum sector, as demonstrated
253 at the country/region level (Fig. 1 and 2).

254 **Implications of freshwater consumption.** The implications of freshwater demand induced
255 by energy sectors are dependent on the geographic overlap between location of activities
256 required to meet demand (Fig. 3), and available freshwater resources (4). However, analyses

257 of such relationships are complicated by the lack of a single universally accepted indicator
258 with which to examine availability of freshwater resources (18), and the fact that impacts can
259 arise through two mechanisms, first and second order water scarcity (50).

260 First order scarcity represents a physical shortage of freshwater. Here we employ two
261 common metrics of first order scarcity; (i) freshwater availability per person and; (ii) the
262 ratio of freshwater withdrawals to availability (18). We examine geographic concordance
263 between these indices and aggregated freshwater consumption for the three energy sectors
264 (petroleum, electric, gas) for the USA and China. Bivariate mapping (see methods and SI
265 Appendix section 6) identifies common areas of spatial overlap between high freshwater
266 consumption induced by the energy sector and river basins that can be considered to
267 experience high first order water scarcity based on thresholds proposed in the literature (18)
268 (SI Appendix, Fig. S5-S8). An ensemble measure identifies major river basins in India,
269 Pakistan, China and the USA (Fig. 4) as being areas where energy induced freshwater
270 consumption is occurring within a context of high first order water scarcity, irrespective of
271 the metric used.

272 Second order water scarcity arises through a lack of social adaptive capacity and reflects the
273 economic and social context in which pressures on freshwater resources are occurring (50–
274 52). The socio-economic context can be as important as physical scarcity in determining
275 implications for society of pressures on freshwater resources (18, 52). Various approaches to
276 calculate a “Water Poverty Index” reflecting second order scarcity have been suggested (50,
277 51); however varying availability of socioeconomic data at sub-country/region scale limits
278 their application in the current study.

279 We examine second order water scarcity using two indices (see SI Appendix section 6) that
280 provide socio-economic indicators at differing spatial scales. The Human Development
281 Index (HDI) is a multidimensional measure that captures a range of social and economic
282 factors that could influence second order water scarcity and has been employed in previous
283 studies that considered social adaptive capacity and freshwater resources (52, 53). Using this
284 national scale measure we find no correlation between freshwater consumption associated
285 with the energy sector for the USA ($\rho = -0.01$, $df = 119$, $P > 0.05$) or China ($\rho = 0.03$, df
286 $= 119$, $P > 0.05$) globally. However, spatial mapping suggests overlap between countries where
287 high energy induced freshwater consumption is occurring within the context of low and

288 medium values for the Human Development Index (Fig. S10 and S11) in India, Pakistan,
289 China and parts of the middle-east.

290 Our second indicator provides data on the prevalence of child malnutrition at a 0.5×0.5
291 degree grid resolution and has been used in a previous study (54) as a measure of social
292 adaptive capacity. Indicators of human health such as malnutrition have been used in a
293 number of studies examining pressures on freshwater resources (53, 55) as, together with
294 economic and social factors, they represent facets relevant to understanding social adaptive
295 capacity (51, 52, 56) and therefore second order scarcity. As with national scale analysis, the
296 lack of correlation between energy induced freshwater consumption and our indicator of
297 social adaptive capacity (prevalence of child malnutrition) for both the USA ($r = 0.01$, $F 0.01$,
298 $df 1,43.70$, $p > 0.05$) and China ($r = -0.01$, $F 0.0045$, $df 1, 40.47$, $p > 0.05$) results from the
299 complex spatial relationship between the two. This relationship is revealed using bivariate
300 mapping at sub-national scales, where areas of high energy induced freshwater consumption
301 are demonstrated to be occurring within the context of low social adaptive capacity within
302 India, Pakistan, south-east Asia, north east Africa and parts of the middle-east (SI Appendix
303 Fig. S12 and S13). The two independent metrics (i.e. HDI, prevalence of child malnutrition)
304 are therefore consistent in identifying a number of geographic regions where energy induced
305 freshwater consumption is occurring within a context of low social adaptive capacity,
306 potentially contributing to second order water scarcity.

307 Considered in the context of first order scarcity (Fig. 4; SI Appendix section 6) there is
308 spatial concordance between geographic areas experiencing high levels of first order
309 (physical driven) (Fig. 4; SI Appendix, Fig. S5-S8) and second order (socio-economic driven)
310 (SI Appendix Fig. S12 and S13) water scarcity and highest energy induced freshwater
311 consumption in a number of river basins, notably in India and Pakistan (see SI Appendix,
312 section 6).

313 **Discussion**

314 Differences between countries in terms of the degree to which energy induced freshwater
315 consumption (Fig. 1) is derived from international sources has important implications for
316 management of renewable freshwater resources. For countries such as China, where energy
317 induced freshwater consumption is largely sourced internally, there is a direct incentive to
318 manage pressures on freshwater resources to ensure security of energy and freshwater supply.
319 Pressures on freshwater resources, of which energy production represents one facet, are

320 increasingly recognised by the Chinese government as a critical issue affecting human
321 wellbeing, economic development and national security within the country (57–59). Country
322 focused analysis using EE-MRIO techniques has demonstrated the physical and virtual
323 transfer of freshwater resources between Chinese provinces to support economic activity (60,
324 61). In demonstrating that globally driven demand for freshwater resources, in this instance
325 associated with energy sectors, contributes to pressures on freshwater resources within
326 countries/regions far removed from where final demand lies our analysis compliments these
327 findings (60, 61). Patterns of freshwater stress across China detailed by (60) correspond to
328 areas identified in our sub-national scale analysis as being where demand induced by energy
329 sectors is occurring within the context of high first order scarcity (Fig 4).

330 In contrast to China, for certain countries/regions and energy sectors (e.g. US petroleum
331 sector), consumption of fresh water along complex international supply chains (35, 62, 63)
332 complicates the development of policy responses and management options at the water-
333 energy nexus. Territorial pressure on freshwater resources has been identified by the US
334 government as a threat to energy security (64) , a result supported by regional US analysis
335 (25). However, our analyses demonstrate that the US petroleum sector is reliant on economic
336 activity in countries/regions of the world that are exposed to significant pressures on
337 renewable freshwater resources (e.g. India, Pakistan; Fig. 4), and where it may be difficult to
338 implement the necessary market reforms (29) to safeguard freshwater resources. This is of
339 particular relevance for activity in transboundary river basins such as the Indus, identified as
340 an area of India and Pakistan associated with high energy-induced freshwater consumption
341 occurring in the context of both first (Fig. 4) and second (Fig. S12 and S13) order water
342 scarcity. Consideration of the water-energy nexus must be in terms of both the territorial and
343 international demand for freshwater resources to enhance both our understanding of the
344 security of energy supply, and broader issues of sustainability through the link between
345 freshwater resources, human wellbeing and economic development.

346 Findings in the current study can be placed within an emerging literature that suggests an
347 imbalance in the use of natural resources (29, 30, 65–67) with exchanges between developed
348 and less developed countries having become increasingly ecologically unequal. The analysis
349 of virtual freshwater transfers to affluent eastern provinces of China from other provinces in
350 the study of (60), highlights that such an imbalance in resource use can also occur within
351 countries. To address such transfers (60) suggests a number of policy mechanisms based on
352 shared producer and consumer responsibility (68) that could be implemented and used to

353 fund agricultural and industrial freshwater efficiency programmes. In the context of findings
354 in the current study we would suggest such mechanisms could also be used at the global level
355 to ensure both security of energy supply in areas where final demand lies, and to address
356 social, economic and environmental issues where freshwater consumption to meet this
357 demand originates. Ultimately, as argued by (25, 69), the analysis presented here provides
358 information that can be used by policy makers to identify critical sectors and geographic
359 regions at the water-energy nexus. When developing energy policy, decisions can then be
360 made to invest in protecting these critical points to reduce social, environmental and
361 economic burdens. For example in the 1970s the government of Saudi Arabia identified
362 threats to territorial freshwater resources as a major issue for the oil industry such that the
363 industry is now based almost entirely on the use of desalination technology and brackish
364 water (70), a fact reflected in our analysis which finds comparatively low freshwater
365 consumption in this region. Our analysis provides information which could enable transfer of
366 resources between countries to enable similar sectorial changes to protect freshwater
367 resources and ensure security of energy supply.

368 Demand associated with each energy sector generates a long chain of interactions in its
369 production processes as all of the resources - the material feedstock and energy inputs, the
370 infrastructure requirements (factories, machinery, processing equipment, transportation,
371 worker canteens etc.), the financial services utilised and so on - need to be “produced” and in
372 turn themselves require numerous inputs. The use of EE-MRIO therefore provides a different
373 perspective on freshwater consumption that moves beyond considering a single aspect of
374 energy production (e.g. petroleum refining; electric generation) to incorporate understanding
375 of the inputs required to undertake such activity. Generation of each input consumes
376 freshwater in the process, with the amount of consumption varying dependent on how
377 freshwater intensive the sector is, such that there can be large disparities between economic
378 activity within a sector and the associated freshwater consumption. In breaking down energy
379 sectors using EE-MRIO (Fig. 2) it is possible to identify in which inputs most freshwater
380 consumption is embodied, and thus consider strategies to reduce overall freshwater
381 consumption by targeting specific sectors.

382 Across energy sectors our analysis demonstrates that agricultural production represents a
383 major contributor to total freshwater consumption (Fig. 2). The dominance of agriculture
384 within our analysis (Fig. 2) is a reflection of high levels of freshwater consumption associated
385 with crop production (35) that subsequently flows to energy sectors, as opposed to a high

386 input of crop materials themselves. This was demonstrated in the analysis of the USA
387 petroleum sector in terms of both induced economic activity (US\$) and freshwater
388 consumption ($\text{m}^3 \text{ yr}^{-1}$). This result is also consistent with analysis that compares sectorial
389 water footprint results across bottom-up (process based) and top-down (EE-MRIO) methods
390 (71), finding substantial differences in water footprints in agricultural and industry sectors
391 depending on the method employed. These differences arise as EE-MRIO calculates the full
392 supply chain water demands of final energy consumption, and hence it does not just sum the
393 direct water consumption associated with only those supply chain components deemed
394 important as is the case in bottom-up approaches. As a result (71) demonstrates that by using
395 EE-MRIO a higher proportion of a nations water footprint will be attributed to industry rather
396 than crops and livestock, as a large proportion of agricultural water use is consumed by
397 industrial sectors as production inputs (e.g. biofuel feedstock). SI section S5 provides an
398 overview of the different approaches to water footprinting.

399 Analysis based on MRIO therefore provides a complimentary perspective on freshwater
400 consumption to bottom-up approaches that has a number of implications relevant for policy.
401 Firstly, although transfer of technology and expertise between countries relating to the
402 industrial side of energy production has a role to play in relieving pressures on renewable
403 freshwater resources, particularly at point localities (e.g. industrial plants, power station),
404 large gains could also be achieved in relation to agricultural production. Adoption of
405 precision irrigation techniques and new crop varieties could represent a “soft-path” to
406 addressing pressures on renewable freshwater resources focussed on improvements in
407 efficiency (72) that would complement those already adopted on the industrial side of energy
408 production. For example (73) demonstrate that reducing freshwater consumption of global
409 crop production to a level that represents the top 25th percentile of current production values
410 could deliver 39% freshwater savings compared to current levels of consumption. In the
411 context of the current analysis such savings would cascade through the global economy,
412 reducing pressures on renewable freshwater resources associated with demand for crops
413 driven by the energy sector (Fig. 2) and delivering benefits to the environment and society. It
414 is not our purpose to propose the most effective form of governance, but rather to inform the
415 debate encompassing those promoting market-based mechanisms and the monetary valuation
416 of ecosystem services, to those advocating more collective and deliberative forms of local
417 level governance (74, 75).

418 Secondly, the importance of agriculture as a driver of freshwater consumption has
419 implications associated with production of energy from biofuel feedstocks suggesting that
420 even modest increases in biofuel production, driven by recent USA and European mandates,
421 could displace freshwater consumption associated with food production to that associated
422 with the energy sector. This finding is consistent with scenarios produced by the IEA that
423 project an 85% increase in freshwater consumption associated with energy between 2010 and
424 2035 driven primarily by expanding biofuel production (21), and results presented in (25) that
425 demonstrate the impact on freshwater resources of increased reliance on bioethanol in
426 California as a result of changes in energy policy since 1990. Such findings emphasise the
427 importance of the spatial aspect of EE-MRIO (29, 35) as this will allow policy to target
428 feedstock production towards countries/regions based on availability of renewable freshwater
429 resources and local socio-economic conditions (42), so contributing to sustainable
430 production.

431 While our analysis advances our understanding of the relationship between energy production
432 and freshwater resources, there are nonetheless a number of limitations, and improvements
433 that require future research. Many of these limitations are common to EE-MRIO analysis;
434 Daniels et al. (42) provide a detailed discussion specific to freshwater resources. Of these
435 limitations, aggregation error, which refers to a lack of product specificity within sectors and
436 to the grouping of countries into regional blocks (29, 42) will most significantly affect our
437 findings in relation to sub-country/region scale mapping of industrial activity. Our estimates
438 of freshwater consumption within a specific sector assume homogeneity in levels of
439 freshwater use efficiency that may mask distinct differences in spatial patterns associated
440 with different industrial processes. A second limitation of our analysis is that total freshwater
441 consumption at the country/region level is assigned to individual 0.5×0.5 degree grid cells in
442 proportion to the location of industry and intensity of freshwater consumption within the grid
443 cell without taking account of distinct sub-country/region patterns that may be associated
444 with individual supply chains. For example although freshwater consumption in the
445 electricity sector is defined spatially based on the location and type of power stations (48),
446 our analysis treats electricity as a pooled resource. In reality, within a specific country/region
447 co-location of electric production and industry may mean that a higher proportion of
448 generated electricity is being used for industrial process in some areas, and a higher
449 proportion for domestic use in others. A third limitation is that for any future analysis using
450 our methodology, the expected rapid expansion of second generation bioenergy feedstocks

451 will need to be incorporated both with the MRIO table through disaggregation of agricultural
452 sectors, and within the crop models contained within WaterGAP.

453 In addition to the EE-MRIO specific limitations discussed above, an additional limitation to
454 our analysis relates to understanding the relationships between pressures placed on renewable
455 freshwater resources, and the implications such pressures have for individuals and
456 communities. Difficulties in the construction of indicators that reflect pressures on renewable
457 freshwater resources arise through the wide range of environmental, economic and social
458 factors that interact to contribute to freshwater scarcity (18, 51). Our analysis addresses this
459 challenge by using a range of possible indicators relevant to both first and second order water
460 scarcity (see Fig. 4; SI Appendix, section 6 and Fig. S10 - S13) to identify concordance
461 between regions with high freshwater scarcity and consumption associated with energy
462 sectors. However, the relative coarse scale of our analysis (0.5×0.5 degree grid; river basin;
463 country/region) and difficulty in obtaining data of relevance for understanding second order
464 water scarcity limits our ability to understand this relationship. Nevertheless, we identify
465 coincident locations of demand for freshwater resources associated with energy sectors and
466 areas subjected to high first and second order scarcity, notably in India and Pakistan. In such
467 areas analysis indicates demand induced by energy sectors is occurring within a context of
468 both physical freshwater scarcity and low social adaptive capacity to address the challenges
469 that freshwater scarcity poses for human wellbeing and economic development. This provides
470 the information necessary to conduct targeted studies along critical supply chains and channel
471 investment and expertise to address pressures at local scales.

472 Our analysis lies at the interface of global efforts to meet societal energy and freshwater
473 needs while addressing climate change. By demonstrating the global connectedness of the
474 energy system and demands on freshwater resources that can be far removed from where
475 final energy demand resides, we provide decision makers with a key piece of knowledge to
476 address future energy security whilst at the same time considering social, environmental and
477 economic consequences of decisions. Given rising populations and the critical
478 interdependence of freshwater, food and energy demand, our work examines an important
479 threat for global freshwater resources that has previously not been considered in detail. The
480 fossil-based sector represents a major contributor to increasing atmospheric CO₂ (76) and as
481 such strategies to reduce greenhouse gas emissions form the dominant discourse within
482 energy policy. We argue that energy policy should increasingly be designed to incorporate
483 not only implications for greenhouse gas emissions, but also consideration of other

484 consequences that will affect global ecosystems and the goods and services that flow from
485 them to society. Failure to do so may mean that we address climate change at the expense of
486 existing natural resources on which human wellbeing and economies depend.

487 **Materials and methods**

488 **Country/region freshwater consumption footprints.** The freshwater resources embodied in
489 a country/regions consumption are calculated using EE-MRIO analysis (see SI Appendix
490 section 3.1). EE-MRIO analysis is well suited to calculating consumption-based
491 environmental accounts at the national and supra-national level (42, 63, 77) as it enables
492 trade flows across the full supply-chain of product categories traded globally to be linked to
493 non-economic measures such as freshwater consumption.

494 The MRIO is based on data from GTAP (78) which is constructed from 2007 global
495 economic data, and contains domestic and international monetary transactions among 57
496 industry sectors across 129 countries/regions (see SI Appendix section 1). Our analysis
497 focuses on three of these sectors -electricity, gas and petroleum - as these represent major
498 sources of energy for the global economy. These three represent the sectors in GTAP where
499 raw materials are transformed into energy carriers that then flow to end users. For example
500 the GTAP petroleum sector (as used in this analysis) receives inputs from the GTAP oil
501 sector, with the latter relating to activity associated with extraction of raw materials (e.g.
502 crude oil). Refined products from the petroleum sector are then sold to industry and final
503 consumers (e.g. goods manufacturers, services and households).

504 Sectorial freshwater consumption by country/region derived from the hydrological model
505 WaterGAP (4, 47–49) (see SI Appendix section 2) provides an environmental extension to
506 the MRIO model following the method given in (79). Freshwater consumption data for 19
507 crop and 12 livestock sectors were derived from WaterGAP, with details of the development
508 of the WaterGAP irrigation and livestock models and assumptions provided in (49, 80). This
509 data was aggregated into the eight crop and two livestock sectors in the MRIO model for each
510 country/region by allocating these to the corresponding sectors (see SI Appendix section 3).
511 Freshwater consumption associated with electricity production in WaterGAP (48) was
512 allocated to the corresponding country/region electricity sectors in the MRIO. WaterGAP
513 allocates all other (i.e. excluding crops, livestock, electricity, domestic) freshwater
514 consumption into a single ‘industry’ sector, which represents 4.18% of total freshwater
515 consumption within the EE-MRIO (47, 48). To disaggregate this among sectors not yet

516 assigned a freshwater consumption value, country/region totals for industry in the WaterGAP
517 model are apportioned among the industry sectors in the MRIO based on their expenditure on
518 the water sector. Here the strength of the interaction with the GTAP water sector is taken as
519 indicative of differences in freshwater consumption between the GTAP sectors (6). Water
520 prices between countries are considered, however the price of water within a country is
521 assumed to be constant, as no within-country price data was available. Data validation for key
522 industrial sectors was also performed against industry and modelling figures from the
523 literature (see SI Appendix section 3.3).

524 Freshwater directly consumed by industry sectors is reallocated through supply chains to the
525 finished products in which it becomes embodied using the standard input-output equation
526 originating from Leontief (81) (see SI Appendix section 3.1.), and used by many in footprint
527 analysis (for example see (29, 30, 42, 61)). Total freshwater consumption for an individual
528 country/region is the sum of embodied freshwater along these supply chains to meet absolute
529 demand for finished products. Sectorial consumption is determined by the country/region's
530 demand for a specific product, such as electricity or petroleum. The embodied freshwater can
531 be traced back to the sector and country/region in which it was originally extracted from the
532 environment to determine the location of appropriation for the consumption activity.

533 **Sub- country/region energy-driven freshwater consumption.** Country/regional patterns of
534 freshwater consumption were mapped to 0.5 x 0.5 degree grid cells using the approach of (35,
535 48). Country/region totals for freshwater consumption in each sector were derived from the
536 EE-MRIO. Values for intensity of freshwater consumption associated with crops and
537 livestock (49, 80), electricity (48) and dwellings (47) were derived from WaterGAP at the 0.5
538 x 0.5 degree grid cell resolution. Country/region totals from the EE-MRIO were then
539 assigned to each 0.5 x 0.5 degree in proportion to intensity of freshwater consumption for the
540 corresponding sector within that 0.5 x 0.5 degree grid cell derived from WaterGAP. Due to
541 aggregation of the industry sector within WaterGAP (48) outlined above, this approach was
542 modified by initially using data from a range of sources (see SI Appendix section 4) to
543 identify 0.5 x 0.5 degree grid cells in which activity associated with key industrial processes
544 (e.g. mineral extraction and refining, oil extraction) was located. Freshwater consumption at
545 the country/region level for the corresponding sector was assigned to each of this subset of
546 0.5 x 0.5 degree grid cells in proportion to intensity of freshwater consumption associated
547 with industry derived from WaterGAP (47, 48). Finally, the remainder of freshwater

548 consumption associated with industrial processes was assigned to each 0.5×0.5 degree grid
549 cells based on aggregate industrial freshwater consumption derived from WaterGAP (47, 48),
550 after accounting for that already assigned in the previous step. Correlations between patterns
551 of freshwater consumption between the USA and China were assessed using a modified *t* test
552 to account for spatial autocorrelation (see SI Appendix section 4.1).

553 **Implications of freshwater consumption.** Freshwater consumption associated with the USA
554 and China energy sectors mapped to a 0.5×0.5 degree grid resolution (Fig. 3; Fig. S3) was
555 aggregated to river basins as defined by the WaterGAP model. Patterns of first order water
556 scarcity within each river basin were assessed using two common measures: (i) the
557 Falkenmark water stress indicator (18) which measures freshwater availability per person
558 (Fig. S5 and S7); (ii) and the percentage ratio of total freshwater withdrawals to availability
559 (18) (Fig. S6 and S8). In both cases freshwater availability was defined as the total renewable
560 freshwater resources derived from the WaterGAP model (4). To create an ensemble measure
561 based on these two indices, firstly total freshwater consumption associated with the USA and
562 China energy sectors was categorised from low (one) to high (five) using a logarithmic scale
563 (Fig. S5-S8). Secondly, each basin was assigned to a first order water scarcity category from
564 low (one) to high (five) based on proposed thresholds for each of the indices (see SI
565 Appendix section 5 and (18); SI Appendix, Fig. S5-S8). For the Falkenmark water stress
566 indicator thresholds for freshwater scarcity were taken from (18) such that (i) river basins with less
567 than $1700 \text{ m}^3 \text{ yr}^{-1}$ per person are considered to experience water stress; (ii) river basins with less than
568 $1000 \text{ m}^3 \text{ yr}^{-1}$ per person are considered to experience water scarcity; and (iii) river basins with less
569 than $500 \text{ m}^3 \text{ yr}^{-1}$ per person are considered to experience absolute scarcity. For the water resources
570 vulnerability index using thresholds taken from (4) a river basin can be considered as; (i) water scarce
571 if the percentage ratio of withdrawals to availability is between 20% and 40%; and (ii) severely water
572 scarce if the percentage ratio of withdrawals to availability exceeds 40%. Thirdly, the score for total
573 freshwater consumption associated with the energy sector (category one-to-five; Fig. S9) was
574 combined with each of the first order water scarcity indicators (category one-to-five; Fig. S9)
575 independently to calculate an index of coincident energy induced freshwater consumption
576 and first order water scarcity. A river basin with high energy induced freshwater consumption
577 (category five) and high first order scarcity (category five) would score the maximum of ten
578 on this coincident index (Fig. S5-S8). Finally, an ensemble measure was calculated by taking
579 an average score of the index of coincident energy induced freshwater consumption and first
580 order water scarcity calculated from the two indices (Fig. 4; SI Appendix section 6).

581 Second order scarcity was examined using two proxy indices for social adaptive capacity, the
582 Human Development Index (HDI) at country/region scale and prevalence of child
583 malnutrition at 0.5 x 0.5 degree grid resolution (SI Appendix section 6). Correlation between
584 these two indices at country/region level ($r = -0.75$, $df = 118$, $p < 0.001$) suggests that data on
585 the prevalence of child malnutrition, which capture within country/region heterogeneity, is
586 indicative of patterns revealed by the HDI which represents a more complex view of social
587 adaptive capacity based on social, economic and health factors. Bivariate mapping was used
588 to identify areas of coincident low adaptive capacity and high energy induced freshwater
589 consumption associated with the USA (SI Appendix, Fig. S10 and S12) and China (SI
590 Appendix, Fig. S11 and S13), for both HDI and prevalence of child malnutrition. Spatial
591 overlap between river basins identified in the context of high first and second order stress and
592 high energy induced freshwater consumption were assessed visually due to difference in
593 spatial scale of data (see SI Appendix, section 6).

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767 **Figure Legends**

768

769 **Fig. 1:** Territorial and international freshwater consumption associated with (A) petroleum,
770 (B) electricity and (C) gas sectors for major economies (The G20, BRICS (Brazil, Russia,
771 India, China and South Africa) and MINTs (Mexico, Indonesia, Nigeria and Turkey)). An
772 expanded version showing all countries/regions can be found in the supplementary material
773 (SI Appendix, Fig. S1).

774 **Fig. 2:** Freshwater consumption by country/region and sector across three energy sectors.
775 Sankey diagrams capture the relationship between the regional and sectorial consumption of
776 freshwater driven by demand for petroleum products (A, B), for electricity (C, D), and for gas
777 (E, F) in the USA and China respectively. Grey bars indicate percentage of total freshwater
778 consumption by geographic region and sector. Coloured lines describe the relationship
779 between the region where demand originates and the sector within the region where
780 freshwater consumption is occurring. See SI Appendix Table S2 for details of country/region
781 and sector aggregation.

782 **Fig. 3:** Spatial pattern of global freshwater consumption driven by freshwater demand from
783 the petroleum sector in the (A) USA and (B) China. Numbers represent total freshwater
784 consumption within each 0.5×0.5 degree grid cell standardised per unit area ($\text{m}^3 \text{yr}^{-1}$ per
785 km^2).

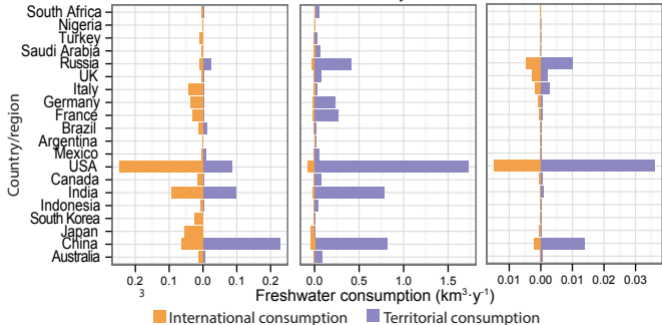
786 **Fig. 4:** Spatial relationship between freshwater consumption driven by demand for the (A)
787 USA and (B) Chinese energy sectors and pressures on freshwater resources. River basins
788 were assigned to a category (one to five) based on freshwater consumption. This was
789 combined independently with two measures of first order scarcity assigned to categories (one

790 to five) to produce two independent measures of overlap between energy induced freshwater
791 consumption and first order scarcity. The mean of these independent measures represents a
792 composite index value of coincident energy induced freshwater consumption and first order
793 water scarcity (see Methods and SI section 6). High values (orange and red) indicate spatial
794 overlap between river basins where high energy induced freshwater consumption is occurring
795 within a context of high first order water scarcity.

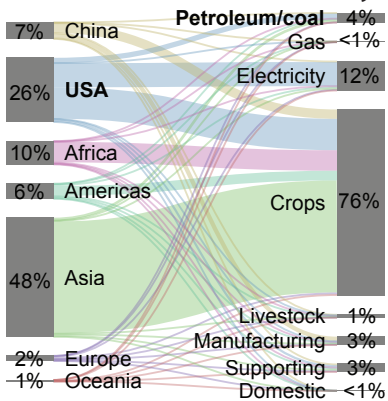
A. Petroleum sector

B. Electricity sector

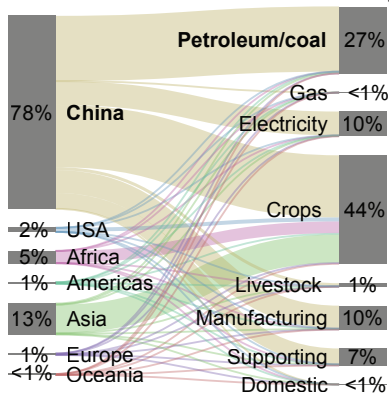
C. Gas sector



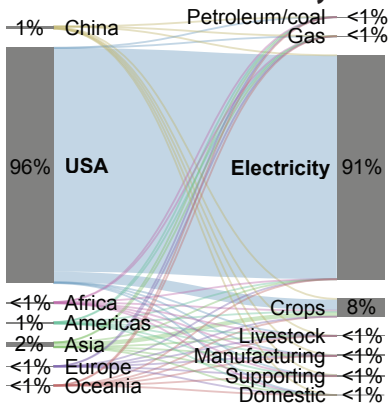
A. USA PETROLEUM 0.33 km³.y⁻¹



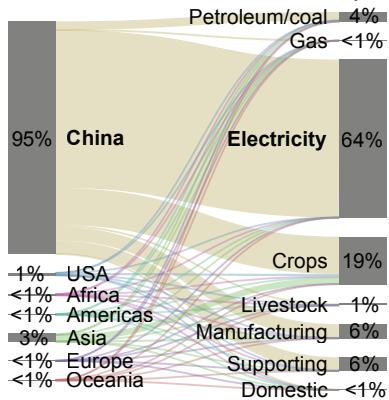
B. CHINA PETROLEUM 0.29 km³.y⁻¹



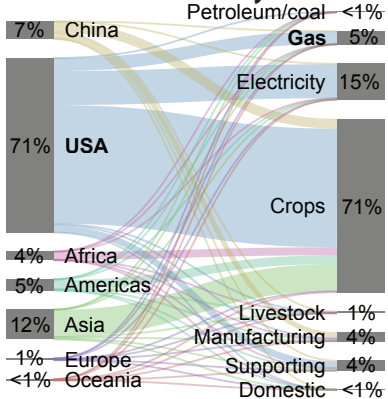
C. USA ELECTRIC 1.80 km³.y⁻¹



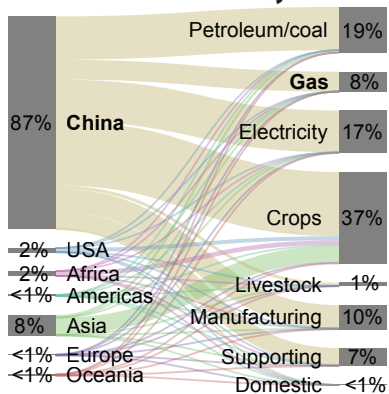
D. CHINA ELECTRIC 0.86 km³.y⁻¹



E. USA GAS 0.05 km³.y⁻¹



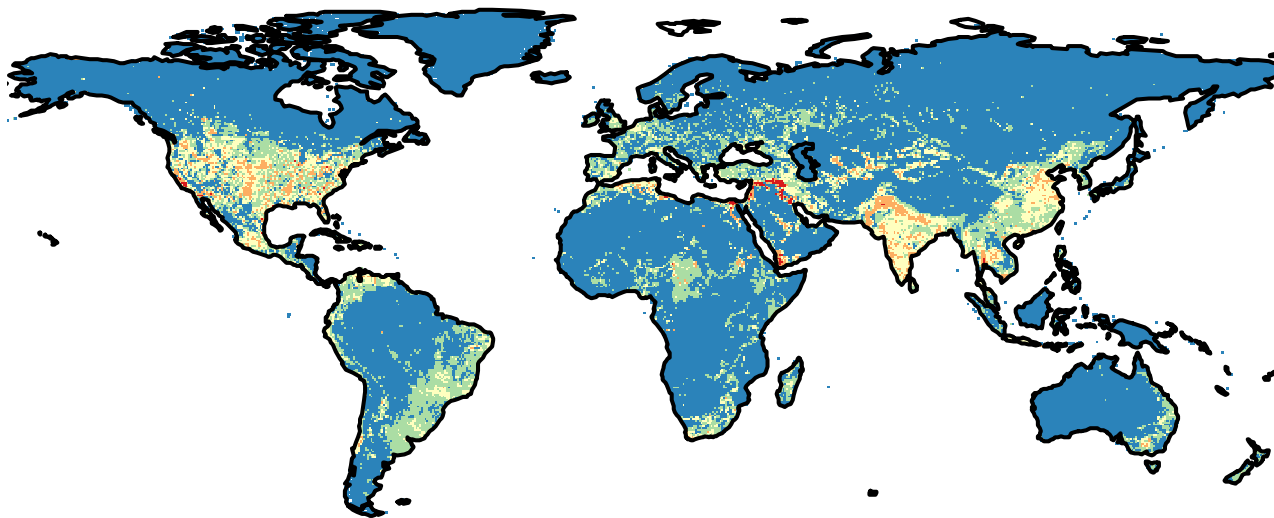
F. CHINA GAS 0.02 km³.y⁻¹



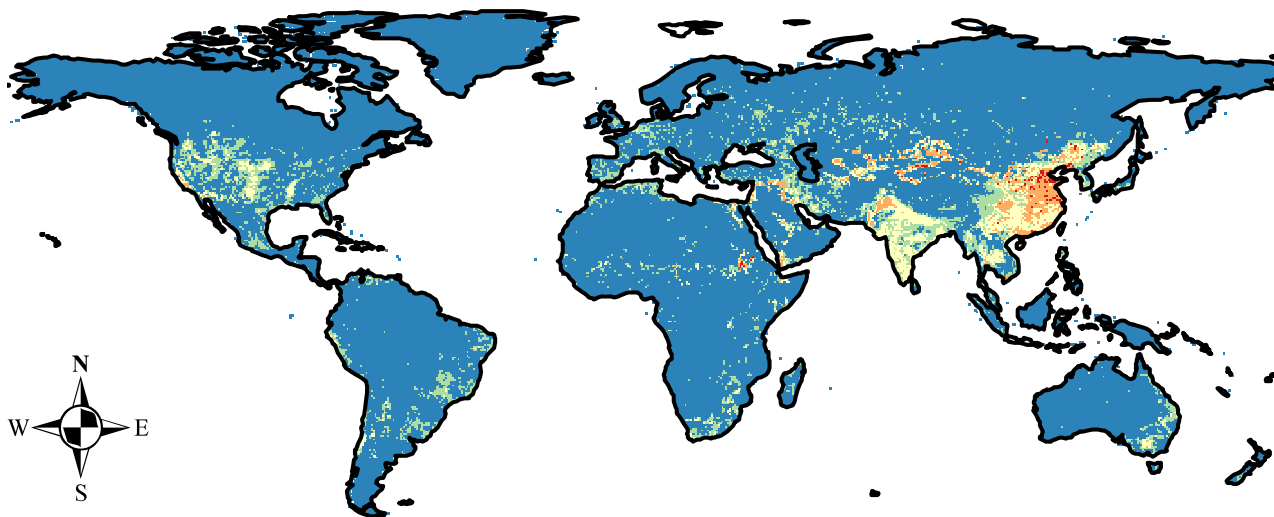
Country/region



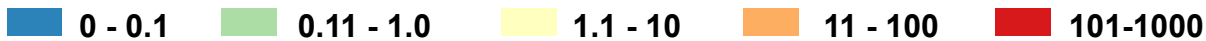
A. USA Petroleum sector



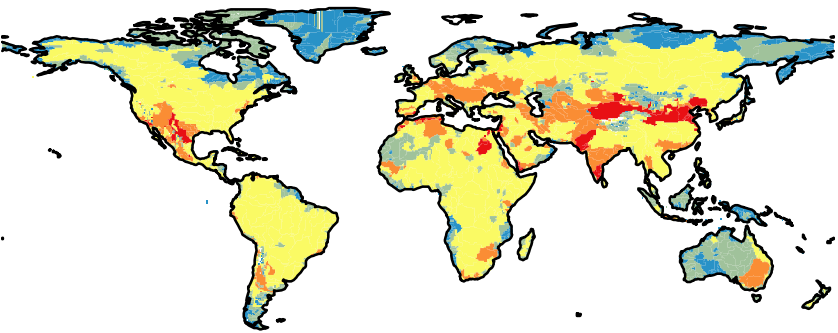
B. China Petroleum sector



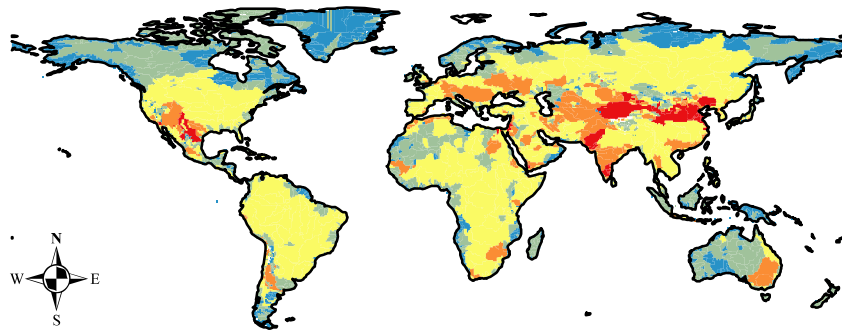
Freshwater consumption per 0.5 x 0.5 degree grid cell (m³·y⁻¹ per km²)



A. USA energy sectors



B. China energy sectors



Overlap between energy induced freshwater consumption and 1st order water scarcity

Composite index value	1,2	3,4	5,6	7,8	9,10
Category	1	2	3	4	5
Freshwater consumption $m^3 \cdot y^{-1}$	<100	<1000	<10000	<100000	>100000
Falkenmark index $m^3 \cdot y^{-1}$ person	>3400	<3400	<1700	<1000	<500
% ratio withdrawal to availability	<1	<10	<20	<40	>40