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The perils of taxonomic inconsistency in quantitative palaeoecology: experiments with testate amoeba data

RICHARD J. PAYNE, MARIUSZ LAMENTOWICZ AND EDWARD A.D. MITCHELL

Payne, R.J., Lamentowicz, M. & Mitchell, E.A.D.: The perils of taxonomic inconsistency in quantitative palaeoecology: experiments with testate amoeba data

A fundamental requirement of quantitative palaeoecology is consistent taxonomy between a modern training set and palaeoecological data. In this study we assess the possible consequences of violation of this requirement by simulating taxonomic errors in testate amoeba data. Combinations of easily-confused taxa were selected and data manipulated to reflect confusion of these taxa, transfer functions based on unmodified data were then applied to these modified data sets. Initially these experiments were carried out one error at a time using four modern training sets, subsequently multiple errors were separately simulated in both four modern training sets and four palaeoecological datasets. Some plausible taxonomic confusions caused major biases in reconstructed values. In the case of two palaeoecological datasets a single consistent taxonomic error was capable of changing the pattern of environmental reconstruction beyond all recognition, totally removing any real palaeoenvironmental signal. The issue of taxonomic consistency is one which many researchers would rather ignore; our results show that the consequences of this may ultimately be severe.

Keywords: Testate amoebae; Palaeoecology; Transfer Functions; Peatlands; Palaeohydrology; Palaeoclimatology

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45 Quantitative palaeoecology generally proceeds by modelling the relationship between
46 species and an environmental variable in modern environments and then applying this
47 model to palaeoenvironmental data to produce quantitative estimates of environmental
48 changes through time. Among the basic requirements of this ‘transfer function’
49 approach is that ‘the fossil data-sets used for reconstruction purposes should be of
50 comparable taxonomy and nomenclature... as the modern training set’ (Birks 1995)
51 i.e. that individuals of the same species are identified consistently and called the same
52 name in both the modern and palaeoecological data (Belyea 2007). However, there
53 are good reasons to suppose that this assumption is sometimes violated; human error
54 is inevitable and in some microfossil groups there is considerable uncertainty
55 regarding the underlying taxonomy. Such a microfossil group is the testate amoebae, a
56 group of protists which are abundant in many aquatic to terrestrial ecosystems and
57 whose solid shells (‘tests’) may be preserved long after death (Fig. 1), allowing
58 community changes to be tracked through time. Testate amoebae are increasingly
59 used in palaeoecology, in particular as proxies for hydrological change, and therefore
60 palaeoclimate, in peatlands (Charman 2001; Mitchell et al. 2008).

61 The taxonomy of testate amoebae is not straightforward. Difficulties start with
62 the problem of applying a biological species concept to micro-organisms which, as far
63 as we know, overwhelmingly reproduce asexually and for which there are little
64 genetic data (Schlegel & Meisterfeld 2003). Testate amoeba taxonomy is built around
65 the concept of morphospecies, that consistent morphological forms represent valid
66 taxonomic units, at least in the absence of any superior approach (Finlay et al. 1996;
67 Finlay 1998). However there are no biometric data for many morphospecies, leaving
68 considerable room for personal interpretation of what degree of difference justifies the
69 erection of new morphospecies and what can simply be considered intraspecific
70 variability (Medioli et al. 1987; Odgen & Meisterfeld 1989). Delineation of species is
71 further complicated by considerable morphological variability in tests (Heal 1963;
72 Wanner 1999; Bobrov & Mazei 2004). Testate amoebae can show marked phenotypic
73 plasticity (Lüftnegger et al. 1988; Wanner & Meisterfeld 1994; Wanner 1999) and in
74 some taxa (adaptive) polymorphism (Schönborn 1992). The test morphology of taxa
75 which build their shells from particles in their environment (xenosomes) depends on
76 the available material; large particles may obscure the underlying test morphology
77 (Ogden 1983). It is probable that many described taxa may just represent extreme
78 forms of this morphological variability. A difference in taxonomies between

79 'lumpers' and 'splitters' is highly apparent in the literature. For instance the
80 *Centropyxis constricta* of Medioli & Scott (1983) would probably include 20 or more
81 species and subspecies considered separable by Chardez (1967).

82 Issues with the differentiation of morphospecies are common to other micro-
83 organisms (e.g. Mann & Droop 1996; Pawlowski et al. 2002). However in the case of
84 testate amoebae these issues are particularly acute due to the inadequacies of the
85 taxonomic literature. Unlike for instance freshwater diatom analysis, where the floras
86 of Krammer & Lange-Bertalot (1986, 1988, 1991a, b) are widely used (at least as a
87 baseline), there is no 'standard text' for testate amoeba taxonomy. The obscurity of
88 testate amoebae to many biologists, combined with the general decline in
89 morphological taxonomic research over recent decades (Lee 2000; Wheeler 2004)
90 have contributed to the poor state of testate amoeba taxonomy. Those attempting to
91 apply testate amoeba analysis in ecology and palaeoecology are forced to use a
92 fragmented body of literature, much of which dates back to the early part of the last
93 century, and much of which is mutually-contradictory. There are no clear rules for
94 separating many taxa and few taxonomic keys are available (none of which are
95 comprehensive and few of which are in English, the de facto language of modern
96 science).

97 In environmental studies using testate amoebae these problems are particularly
98 serious because of the large number of tests which must be counted; typically at least
99 100 individuals per sample and 40-50 samples (Payne & Mitchell 2009). This number
100 of tests pragmatically requires that all identification and counting be carried out using
101 light microscopy under normal (200x to 400x) magnifications. Many fine taxonomic
102 distinctions rest on very subtle features which are simply not practicable under these
103 conditions (e.g. in Euglypha: Wylezich et al. 2002, Cyphoderia: Todorov et al. 2009;
104 Heger et al. in press, and Diffflugia: Ogden 1983). In palaeoecology problems are
105 compounded by the loss of diagnostic features. The division between taxa with lobose
106 and filose pseudopodia is the most fundamental in testate amoebae taxonomy but is
107 not applicable in palaeoecology. Diagnostic features of the test such as spines may be
108 lost through taphonomic processes or in sample preparation and tests may become
109 compressed (Charman et al. 2000). Taxonomic schemes used in palaeoecology are
110 therefore a compromise between practical simplicity and loss of palaeoenvironmental
111 discernment (Charman et al. 2000). Given all these problems it would be little
112 surprise if there were considerable taxonomic differences among researchers. In the

113 absence of a formal inter-comparison exercise it is impossible to know to what extent
114 different researchers apply the same name to different taxa or different names to the
115 same taxon. We can however make observations that: i) The taxonomic literature
116 lacks clarity. ii) There are considerable differences in the taxonomic resolution
117 adopted by different studies. iii) Inter-comparison exercises for other microfossil
118 groups used in Quaternary palaeoecology have shown considerable variability
119 between different analysts and research groups (Munro *et al.* 1990; Pederson &
120 Moseholm 1993; Kelly *et al.* 2002; Prygiel *et al.* 2002). For instance, in the diatom
121 inter-comparison exercise of Kelly *et al.* (2002) some taxa were identified correctly
122 less than 20% of the time. iv) When researchers are learning testate amoeba taxonomy
123 several mistakes are consistently made.

124 On the basis of these observations we feel it would be naïve to assume that
125 taxonomies are identical among all researchers. In this study we attempt to gain an
126 understanding of the possible implications of taxonomic variability for environmental
127 reconstructing by simulating possible errors in previously established modern and
128 palaeoecological datasets.

129

130 Methods

131

132 Four modern training sets and four palaeoecological datasets were used in our
133 experiments. The four modern training sets are all derived from Sphagnum-
134 dominated, mostly ombrotrophic mires and span a considerable region from North
135 America to western Asia (Table 1). They are: i) Poland, from peatlands of Poland
136 (Lamentowicz *et al.* 2005, 2007, 2008); ii) Jura, from peatlands in the Jura Mountains
137 of France and Switzerland (Mitchell *et al.* 1999, 2001); iii) Turkey, from the Sürmene
138 Ağaçaşlı Yaylası peatland in north-eastern Turkey (Payne *et al.* 2008); and iv)
139 Alaska, from peatlands in south-central Alaska (Payne *et al.* 2006). The final selected
140 transfer function models were used in our experiments to infer depth to water table
141 (DWT; Table 1). The four palaeoecological datasets are: 1. 'Site DLB', a peatland in
142 sub-Arctic Alaska (Payne *et al.* unpublished, but see Payne & Mitchell 2009); 2. Praz-
143 Rodet, a peatland in Switzerland (Mitchell *et al.* 2001); 3. Tuchola, a peatland in
144 Poland (Lamentowicz *et al.* 2008), and 4. Jelenia Wyspa, another peatland in Poland
145 (Lamentowicz *et al.* 2007). All of these palaeoecological datasets have an applicable
146 transfer function from the same area (i.e. the Alaska, Jura and Poland training sets,

147 Table 1) which was produced by the same analysts. We are as confident as possible
148 that these palaeoecological datasets and their respective transfer functions have
149 consistent taxonomic schemes.

150 A first step in our experiments was to select pairs of species which we
151 considered could be confused (Table 2). Our combinations were based on three
152 sources of evidence: i) Our assessment of the distinctiveness of the taxon based upon
153 the literature, in particular where taxa have been considered inseparable by some
154 authors. ii) Our observations of the mistakes made by undergraduate and postgraduate
155 students in learning testate amoeba taxonomy. iii) Our own experience of learning
156 testate amoeba taxonomy. We produced separate lists of taxon combinations for each
157 of our training sets, reflecting the differing communities encountered in those studies
158 and the slightly different taxonomic schemes adopted by the analysts. For simplicity
159 we refer to each of these taxon combinations as an ‘error combination’, however with
160 some of these pairings we note that the distinction between the taxa may not always
161 be clear. We would not claim that our taxon combinations reflect all possible errors or
162 that all of these errors have a high probability. However, we do feel that our taxon
163 combinations include all of the most common confusions. Three sets of experiments
164 were conducted:

165 Individual errors

166 The first group of experiments used only the modern training sets and was designed to
167 quantitatively investigate the impacts of individual errors on transfer function
168 predictions. We identified three possible ways in which each pair of species could be
169 confused: 1. All of taxon A could be recorded as taxon B. 2. All of taxon B could be
170 recorded as taxon A. 3. The taxa could be switched. The training set data were then
171 transformed to reflect each of these three types of error for each of the taxon pairs
172 identified. So for instance with the Alaska data we identified 15 taxon pairs (Table 2),
173 which could each be transformed in three different ways giving a total of 45 possible
174 individual modifications to the data. We then applied the transfer function derived
175 from the original, unmodified training set to each of these modified data-sets in turn to
176 predict depth to water table (DWT). This approach of applying a transfer function
177 based on a training set to the same training set but with simulated taxonomic errors is
178 not representative of any real-world situation but is a useful tool to investigate the
179 impact that these errors might have on transfer function results.

180 Inferred depth to water table values (termed ‘testate amoeba-inferred depth to
181 water table’: TI-DWT) were compared to predictions based on the unmodified data
182 set and residuals calculated ($\text{TI-DWT}_{\text{original}} - \text{TI-DWT}_{\text{modified}}$). Differences between
183 predictions based on the original and modified data were calculated in terms of root
184 mean square error (RMSE), R^2 and the maximum difference between predictions for
185 any one sample (Maximum Bias). All transfer function analyses were carried out
186 using C^2 (Juggins 2003).

187 Multiple errors

188 To investigate the cumulative impact of more than one error we also carried out
189 experiments simulating multiple errors in our modern training sets. The same taxon
190 combinations were used as in the individual errors experiments. A random numbers
191 system was used to select a taxon pair, with each pair assigned an equal probability of
192 selection. Where more than two taxa could be confused with each other only one
193 taxon pair could be selected at a time (where more than one pair were selected the
194 data were kept unchanged). Each taxon pair could be transformed in one of the three
195 ways described above with each of these three modifications given an equal
196 probability of being selected. The number of errors in the data was steadily increased
197 up to the maximum number of possible changes, with fifteen repetitions for each error
198 total. The transfer function based on the unmodified training set was then applied to
199 this modified training set and RMSE, R^2 and Maximum Bias calculated as above.

200 A related possible source of bias in inferred values is that taxonomic errors in
201 a training set lead to selection of a different transfer function model structure which
202 may, in itself, lead to differences in model output. To investigate the potential
203 implications of this issue alternative model structures (WA, WA-Tol, WA-PLS, ML)
204 were tested using the maximum number of simulated errors in each training set and 15
205 replicates. The best performing model was selected based on $\text{RMSEP}_{\text{jack}}$ with no
206 penalty for model complexity.

207 Errors in palaeoecological sequences

208 To see how the simulated errors might affect palaeoenvironmental inference we also
209 manipulated the four palaeoecological data-sets and then applied transfer functions
210 based on unmodified training sets. The same taxon combinations were used when
211 simulating errors in the palaeoecological data-sets as were used in the two
212 experiments simulating errors in training sets described above. The number of errors
213 was successively increased from one to ten. Transfer functions based on the

214 unmodified training set data were applied and TI-DWT values calculated for each
215 modified palaeoecological data-set.

216

217 Results

218

219 Individual errors

220 Results of individual error experiments are shown in Table 2. With all training sets a
221 few error combinations have a great deal more impact on predictions than most
222 others. With the Poland data much the most significant error combination is *Diffflugia*
223 *globulosa*/*Cyclopyxis arcelloides*, introducing a mean error of up to 2.5 cm (7% of the
224 total measured DWT range) depending on which of the three permutations is
225 considered, the next most important error combination is *Arcella vulgaris*/*Arcella*
226 *discoides* ($RMSE \leq 0.55$ cm, 1.5% measured range). With the Jura data the two most
227 important error combinations are *Cyclopyxis arcelloides*/*Phryganella acropodia*,
228 leading to a mean error of up to 1.95 cm (4% measured range) and *Centropyxis*
229 *aerophila*/*Centropyxis platystoma*, leading to a mean error of up to 1.1 cm (2%
230 measured range). With the Turkey data the most important error combination is
231 *Corythion dubium*/*Trinema lineare*, leading to a mean error of up to 1.7 cm (2%
232 measured range). With the Alaska data the most important error combinations are
233 *Euglypha ciliata*/*Euglypha strigosa* ($RMSE \leq 3.06$ cm, 5% measured range), *Nebela*
234 *tincta*/*Nebela penardiana* ($RMSE \leq 2.78$ cm, 4.6% measured range) and *Heleopera*
235 *petricola*/*Heleopera sphagni* ($RMSE \leq 2.13$ cm, 3.5% measured range). Maximum bias
236 data show that many of these single errors lead to the predicted TI-DWT values of
237 some samples changing by more than 10 cm, and in some cases more than 20 cm.
238 These are highly significant changes; 20 cm represents the DWT difference between a
239 lawn and a low hummock.

240 Multiple errors

241 When multiple errors are simulated there is a steady increase in the deviation of
242 predictions from those based on the unmodified data (Fig. 2). With the Alaska data
243 there is an approximately equal division between samples with TI-DWT over- and
244 under-predicted relative to the original data. However with the other three data-sets
245 there is a trend in one direction; with the Poland data this is towards under-prediction
246 of TI-DWT while with the Jura and Turkey data this is towards over-prediction of TI-
247 DWT. This directional bias is most apparent with the Jura data with the TI-DWT

248 values of the majority of samples being over-predicted relative to the unmodified data.
249 These directional biases are largely driven by just a few errors, so with the Jura data
250 the trend is mostly due to the *N. tinctoria*/*N. parvula* combination, with the Poland data
251 the trend is mostly due to the *C. arcelloides*/*D. globulosa* combination and with the
252 Turkey data the trend is mostly attributable to the *C. dubium*/*T. lineare* and *H.*
253 *petricola* /*H. rosea* combinations.

254 If alternative transfer function model structures are tested using the training sets
255 with simulated errors a different model structure is selected with 93% of replicates
256 with the Jura data, 60% of replicates with the Poland data, 40% of replicates with the
257 Turkey data and in no replicates with the Alaska data.

258

259 Errors in palaeoecological sequences

260 The consequences of these errors for palaeoecological reconstruction are shown in
261 Fig. 3A-D. With the Site DLB data (Fig. 3A) the most distinct features of the
262 reconstruction based on unmodified data are pronounced wet phases at the base of the
263 profile, from 52-56 cm and from 25-28 cm. These wet phases generally remain
264 apparent even when taxonomic errors are introduced, although with increasing
265 number of errors the phases become less distinct in some experiments. A notable
266 change with even one error is a period of higher values between 11 and 15 cm due to
267 counting *Centropyxis ecornis* as *Centropyxis laevigata*. With the Praz Rodet data (Fig.
268 3B) simulated errors make relatively little difference to reconstructed values. The
269 maximum deviation is 7.6 cm but in none of these experiments is the TI-DWT
270 reconstruction different enough to change interpretation of the record. With the
271 Tuchola data (Fig. 3C) even a single error can drastically change the pattern of the
272 reconstruction: If *Cyclopyxis arcelloides* is recorded as *Diffugia globulosa* it
273 fundamentally changes the reconstruction giving an overall reduction in predicted
274 values, introducing a period of rapidly fluctuating values between 20 and 120 cm
275 depth and adding a trough at 360 cm. Interpretation of these data with and without this
276 error would be utterly different. Increasing error load slightly increases the variability
277 of predictions, but the overall pattern is largely determined by whether or not *C.*
278 *arcelloides* and *D. globulosa* are confused.

279 With the Jelenia Wyspa data (Fig. 3D) the difference that even a single error can
280 make is even more marked. Again the most important error is recording *C. arcelloides*
281 as *D. globulosa*. This error leads to a general under-prediction of TI-DWT by 5 cm or

282 more and an almost total difference in the pattern of change. Introducing this error
283 leads to the reconstruction of major TI-DWT peaks at 42, 95 and 110 cm, features
284 which are totally absent in the reconstruction based on unmodified data. One of the
285 most distinctive features of the TI-DWT reconstruction based on the unmodified data
286 is a period of high values between 50 and 65 cm. However in several experiments
287 with one or more errors this feature is less distinct or not apparent at all. In these
288 experiments *Centropyxis cassis* has been recorded as either *Centropyxis platystoma* or
289 *Centropyxis aerophila*. With increasing number of errors there is an increasing
290 variability in the pattern of reconstructed change, although reconstructions group
291 around two basic patterns determined by whether *C. arcelloides*/*D. globulosa* are
292 confused or not. In some experiments where both *C. arcelloides*/*D. globulosa*, and *C.*
293 *cassis* and *C. aerophila* or *C. platystoma* are confused TI-DWT values deviate from
294 the unmodified data by more than 17 cm.

295

296 Discussion

297 All of our experiments make several important assumptions: they assume that
298 mistakes are made consistently, that these are all possible errors and all have an equal
299 probability, and they do not account for tests simply over-looked or mistaken for taxa
300 not included in the transfer function and therefore excluded. While we acknowledge
301 that our experiments represent a considerable simplification of the real way in which
302 taxonomic errors may affect transfer function output the results are undeniably
303 revealing. While many possible errors make very little difference to predicted values
304 some possible errors can change predicted values drastically, giving reconstructions
305 which bear little apparent resemblance to those based on full data.

306 The specific errors which produce major impacts in our experiments seem by
307 no means improbable. For instance the confusion of *C. dubium* with *T. lineare*
308 (important in the Turkey training set) and *E. ciliata* with *E. strigosa* (important in the
309 Alaska training set) are both common mistakes among our students. The most
310 dramatic illustration of the possible impacts of taxonomic errors in our experiments is
311 provided by the experiments simulating errors in palaeoecological data sets from
312 Tuchola and Jelenia Wyspa. Major differences in reconstructions are produced by
313 confusing *D. globulosa* and *C. arcelloides*, two taxa that have a similar overall
314 morphology and would probably be grouped by Charman et al. (2000) or Medioli &
315 Scott (1983). The drastic impact that this error makes is particularly notable given the

316 relative scarcity of these taxa in the Tuchola data, constituting only 2.7% of total tests
317 and only exceeding 5% of count in 5 samples. In the Jelenia Wyspa data the taxa are
318 slightly more abundant, constituting 10.1% of total tests. The difference that this
319 single change makes to the reconstructions highlights the extent to which the pattern
320 of palaeoenvironmental reconstruction may be determined by just a few important
321 taxa. It is worryingly easy to envisage a scenario where somebody, perhaps relatively
322 new to testate amoebae palaeoecology and using one of the more agglomerative
323 taxonomies as their main guide, could make such an error to produce an
324 environmental reconstruction which is substantially biased, or in the worst case
325 entirely an artefact of taxonomic inconsistency. Taxonomic errors in a training set
326 may change the transfer function model structure selected, but it is likely that this
327 change alone would have limited impact on model output (cf. Booth 2007).

328 The large impacts of some of the simulated errors may suggest the need to
329 group these potentially problematic taxa in our transfer functions. However these taxa
330 frequently have significantly differing hydrological optima, therefore a corollary of
331 the impacts of these errors is that if these taxa are grouped considerable ecological
332 information will be lost. In the worst case grouping may considerably bias
333 reconstructions. If one of a pair of taxa is well represented in a training set and the
334 other not, the ecological optima of the group will mostly match that of the first taxon,
335 however if the second taxon is more abundant in palaeoecological samples then
336 reconstructed values will be biased.

337 In the absence of any formal taxonomic inter-comparison it is not possible to
338 make any definitive assessment of how much of a problem taxonomic inconsistency
339 may be in praxis. We would suggest that these errors are far from implausible.
340 However, whether or not these specific taxonomic errors are very likely, our results
341 suggest a wider point, that it is possible for taxonomic errors to radically distort
342 environmental reconstructions. Taxonomic errors will not necessarily make any
343 significant difference to environmental reconstruction; indeed, most errors will
344 probably make very little difference. However, there is the potential for a single
345 taxonomic mistake made consistently to so change an environmental reconstruction
346 that the real palaeoecological signal is totally masked. Although our experiments only
347 consider water table reconstruction in peatlands it is likely that similar results would
348 be found when considering reconstruction of other variables and in other
349 environments. Problems may be particularly acute in minerotrophic peatlands where

350 there may be a greater abundance of ‘difficult’ taxa (e.g. genera *Diffflugia* and
351 *Centropyxis*).

352 Taxonomic comparability is critical; what a palynomorph used in
353 palaeoecology is called matters little as long as the name is used consistently. For
354 instance, non-pollen palynomorphs are commonly referred to as simply a numbered
355 ‘type’ as the origin of the palynomorph may not be known (van Geel 2001). Given the
356 taxonomic limitations imposed by palaeoecological counting some authors have
357 considered it necessary to use a parallel naming system, for instance Joosten & de
358 Klerk (2002) have suggested the differentiation of fossil pollen from plant species
359 (and indeed modern pollen) by referring to the former in SMALL CAPITALS. While we
360 do not feel that such a system is necessarily required for testate amoebae we would
361 appeal for clarity in the description of taxonomies used in palaeoecological studies of
362 testate amoebae. Until a revised taxonomic framework with clear identification
363 criteria and keys is available and consistently used, researchers publishing training
364 sets should clearly state identification criteria and the taxa included in groupings
365 where these are not obvious.

366 Extreme caution should be used when applying transfer functions, particularly
367 when using training sets counted by different analysts. Researchers attempting to use
368 a transfer function derived by other analysts should work in close cooperation to
369 ensure the same identification criteria are consistently employed. In our experience
370 this is best done by close communication during counting, rather than trying to post-
371 hoc adjust the taxonomy of a palaeoecological data-set to fit the taxonomy of a
372 transfer function. Comparison of photographs of difficult taxa between analysts is a
373 useful approach to ensure this consistency. Where there is any doubt at all over the
374 criteria for differentiating taxa these taxa should be grouped or excluded from the
375 data-sets. The fact that extremely large reconstruction errors can be introduced by
376 relatively modest taxonomic errors adds to the case for comparing testate amoeba-
377 based records with other data in a multi-proxy approach, and ideally replicating
378 records with multiple cores. All palaeoecological techniques are imperfect, testate
379 amoeba analysis is no exception.

380 There appears to be a tendency in testate amoeba-based palaeoecological
381 reconstruction to use boot-strapping to derive estimates of standard errors and
382 consider any changes which exceed these error bars (or even do not: Hendon &
383 Charman 2004) to be a palaeoecological ‘signal’. However, these standard errors only

384 provide an estimate of the error inherent in the model, additional errors may well be
385 introduced if the transfer function does not provide an adequate fit to the
386 palaeoecological data (cf. Wilmschurst et al. 2003) or taxonomic errors are made. In
387 our experiments even quite minor taxonomic errors produced a bias that significantly
388 exceeded the boot-strapped standard errors. Boot-strapped standard errors should be
389 used with caution as other sources of error can produce biases which considerably
390 exceed these estimates.

391 To ensure taxonomic consistency there is a need for a common standard
392 taxonomy which can be applied uniformly among analysts given the constraints
393 imposed by counting large numbers of sub-fossil tests using optical microscopy. The
394 guide of Charman et al. (2000) is the best attempt at this and is widely used (79
395 citations in ‘Google Scholar’ at the time of writing). However, the taxonomic scheme
396 set out has not met with uniform acceptance with many authors either not adopting
397 this scheme or adapting it to varying extents. Major reasons for this lack of consistent
398 use may include the exclusion of some relatively common peatland taxa (e.g.
399 *Euglypha cristata*, *Tracheleuglypha dentata*) and the broad ‘types’ adopted for some
400 groups of taxa (perhaps most notably the ‘*Cyclopyxis arcelloides* type’). The guide of
401 Charman et al. (2000) provides a first attempt at a difficult task and is a very useful
402 contribution. However we would argue that now, ten years after publication, is the
403 time for a reconsideration and refinement of the scheme in an attempt to achieve a
404 broad consensus. A consistent taxonomy is essential given increasing attempts to
405 compare and combine modern data-sets while the more widespread use of testate
406 amoebae in palaeoecology means that more environmental reconstructions are being
407 produced using transfer functions derived by other researchers. Taxonomic
408 inconsistency is a neglected issue in biological sciences, but its consequences may
409 ultimately be very severe (Bortolus 2008).

410

411 Conclusions

412 • Errors of taxonomy and enumeration are inevitable in palaeoecology. Testate
413 amoeba analysis is likely to be particularly susceptible to such errors due to
414 the inadequacies of the taxonomy.

415 • Our experiments suggest that some likely confusions can produce significant
416 biases in quantitative environmental reconstructions.

417 • These results call for improvement of the taxonomic baseline. For now,

418 extreme caution should be used when applying transfer functions and especially
419 interpreting small changes.

420 • There are many possible causes of bias in environmental reconstructions.
421 Taxonomic inconsistency is but one of these.

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431 constructive comments on a previous version of the manuscript.

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433 'Truth is mighty and will prevail. There is nothing the matter with this, except that it
434 ain't so.' (Mark Twain)

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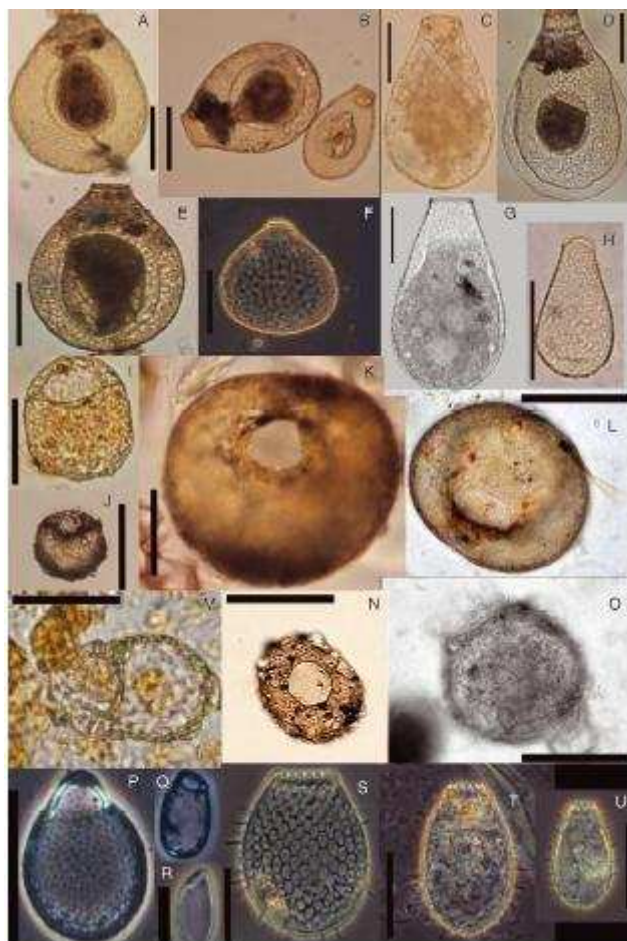
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583 FIGURES and TABLES

584

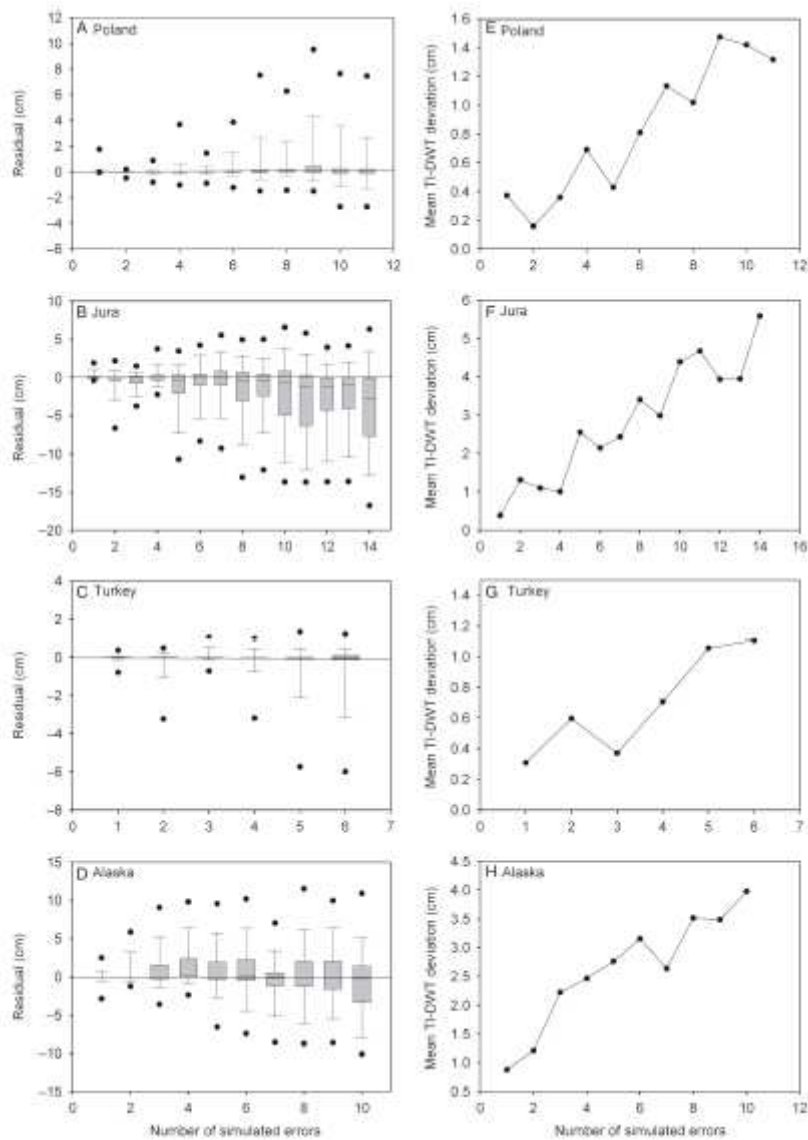
585 Figure 1. Illustrations of selected testate amoeba taxa discussed in this paper. A.
 586 *Nebela tinctoria* var. *major*. B. *N. tinctoria* var. *major* and *N. tinctoria*. C. *N. marginata*. D. *N.*
 587 *carinata*. E. *N. tinctoria* var. *major*. F. *N. flabellulum*. G. *N. penardiana*. H. *N. militaris*.
 588 I. *Centropyxis aerophila*. J. *C. aerophila* var. *sphagnicola*. K. *C. ecornis*. L. *C.*
 589 *laevigata*. M. *C. platystoma*. N. *Phryganella acropodia*. O. *Diffugia globulosa*. P.
 590 *Corythion dubium*. Q & R. *Trinema lineare*. S. *Euglypha ciliata*. T. *E. compressa*. U.
 591 *E. strigosa*. Scale bar is 20µm for P,Q and R, 50µm for others.



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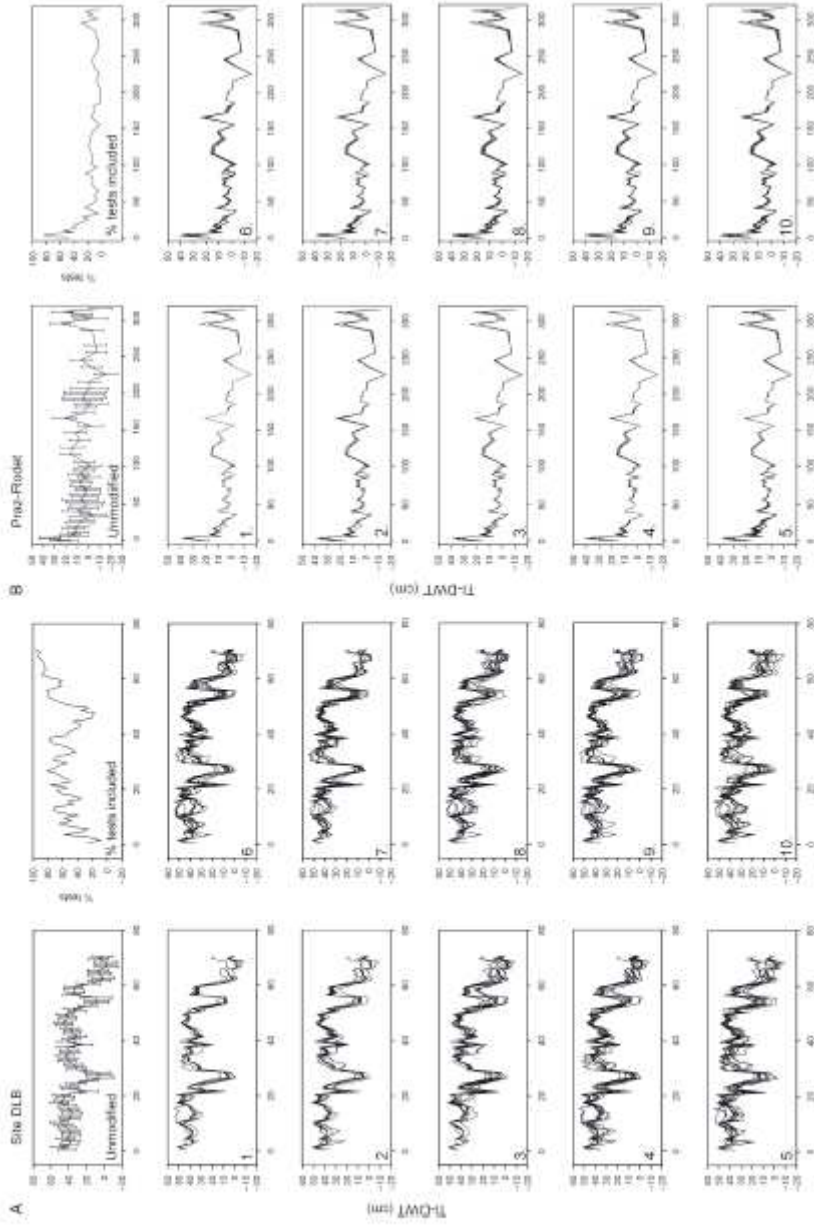
594 Figure 2. Results of multiple error experiments (see Methods) with four modern
 595 training sets. Plots A-D show residuals ($TI-DWT_{original} - TI-DWT_{modified}$), plots E-H
 596 show the same data presented as an overall mean TI-DWT deviation. Box plots show
 597 median (central line), first and third quartiles (grey box), tenth and ninetieth
 598 percentiles ('whiskers') and fifth and ninety-fifth percentiles (dots).

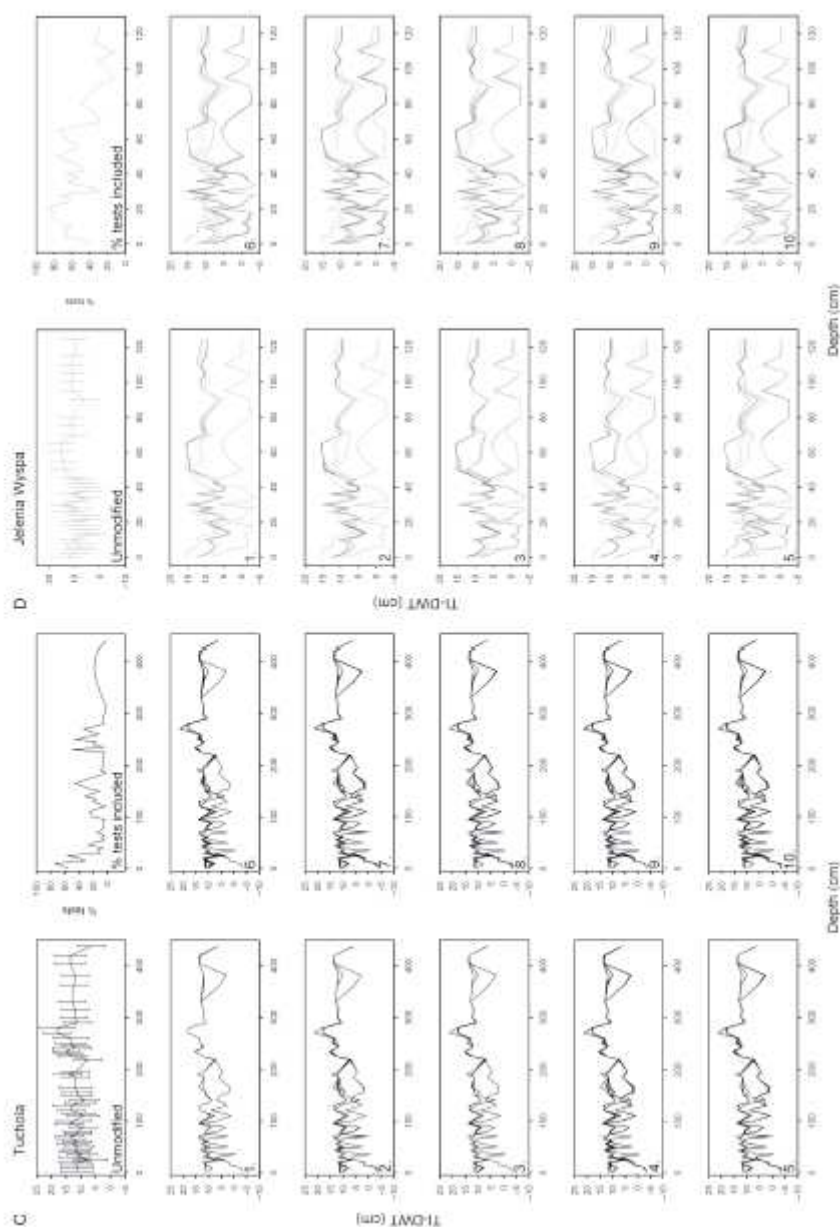


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600 Figure 3. Results of errors in palaeoecological sequences experiments (see Methods)
 601 with palaeoecological data from A) 'Site DLB', Alaska, B) Praz-Rodet, Swiss Jura,
 602 C) Tuchola, Poland, and D) Jelenia Wyspa, Poland. For each dataset the plot on the
 603 upper left shows reconstruction based on unmodified data and the adjacent plot shows
 604 percentage of tests contributed by the taxa which could be confused. Other plots show
 605 reconstructions for increasing number of errors from 1-10 with fifteen cycles of
 606 random re-selection for each error total.

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Table 1.

Attributes of the datasets used in this study showing number of samples (n), and for modern training sets: transfer function model structure, jack-knifed root mean square error of prediction (RMSEP), Maximum Bias and R^2 . Location given in parentheses after palaeoecological data set name indicates applicable transfer function.

Location	n	Model structure	RMSEP _{jack} (cm)	Max Bias _{jack} (cm)	R^2_{jack}	Reference
Modern training sets:						
Poland	84	WA-Tol, Inverse deshrinking	4.6	9.0	0.71	Lamentowicz <i>et al.</i> (2007)*
Jura	37	WA-PLS (2 component)	8.0	21	0.62	Mitchell <i>et al.</i> (1999, 2001) [†]
Turkey	42	ML	7.1	21	0.81	Payne <i>et al.</i> (2008)

Alaska	91	WA-PLS (2 component)	9.7	14	0.55	Payne <i>et al.</i> (2006)
Palaeoecological data sets:						
Site DLB (Alaska)	71					Payne <i>et al.</i> (unpublished)
Praz Rodet (Jura)	57					Mitchell <i>et al.</i> (2001)
Tuchola (Poland)	50					Lamentowicz <i>et al.</i> (2008)
Jelenia Wyspa (Poland)	38					Lamentowicz <i>et al.</i> (2007)

*Values slightly different from published due to re-calculation of percentages. †Re-calculated using WA-PLS, see Payne and Mitchell (2009).

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Table 2. Results of individual error experiments (Methods section 1) for A) Poland (Lamentowicz *et al.* 2007), B) Jura (Mitchell *et al.* 1999, 2001), C) Turkey (Payne *et al.* 2008), D) Alaska (Payne *et al.* 2006). Showing, taxon pair (A and B), percentage of total tests these taxa represent, number of occurrences of each taxon (N), DWT optima estimated by weighted averaging ('WA Optima') and impact of simulated errors in terms of RMSE, maximum bias and R^2 between TI-DWT based on original and modified datasets. Each taxon pair could be changed in three ways: all of taxon A could be counted as taxon B (A→B), all of taxon B could be counted as taxon A (A←B), and the two taxa could be switched (A↔B).

A) Poland

Taxon A	Taxon B	% total		N		WA optima		RMSE		
		A	B	A	B	A	B	A→B	A←B	A↔B
<i>Corythion dubium</i>	<i>Corythion-Trinema</i> type	0.80	0.03	13	4	23.08	20.90	0.03	0.00	0.03
<i>Cyclopyxis arcelloides</i>	<i>Diffugia globulosa</i>	3.63	1.74	33	6	4.36	-0.18	2.33	0.28	2.49
<i>Nebela parvula</i>	<i>Nebela tincta</i>	1.37	2.40	32	33	19.04	21.59	0.04	0.08	0.08
<i>Nebela bohémica</i>	<i>Nebela collaris</i>	2.49	0.12	24	6	11.60	19.72	0.19	0.02	0.20
<i>Nebela militaris</i>	<i>Nebela collaris</i>	1.21	0.12	15	6	25.11	19.72	0.12	0.01	0.11
<i>Heleopera sphagni</i>	<i>Heleopera petricola</i>	0.42	1.56	15	31	13.29	13.02	0.00	0.01	0.01
<i>Heleopera sylvatica</i>	<i>Heleopera petricola</i>	0.16	1.56	5	31	20.10	13.02	0.01	0.05	0.06
<i>Euglypha strigosa</i>	<i>Euglypha compressa</i>	0.25	0.43	10	11	19.75	6.92	0.11	0.06	0.17
<i>Euglypha compressa</i>	<i>Euglypha ciliata</i>	0.43	0.41	11	8	6.92	6.51	0.02	0.02	0.02
<i>Euglypha ciliata</i>	<i>Euglypha strigosa</i>	0.41	0.25	8	10	6.51	19.75	0.40	0.05	0.07
<i>Centropyxis cassis</i>	<i>Centropyxis aerophila</i>	0.27	0.07	5	3	13.98	7.41	0.03	0.03	0.07
<i>Centropyxis aerophila</i>	<i>Centropyxis platystoma</i>	0.07	0.03	3	2	7.41	8.68	0.00	0.00	0.00
<i>Centropyxis cassis</i>	<i>Centropyxis platystoma</i>	0.27	0.03	5	2	13.98	8.68	0.05	0.01	0.05
<i>Amphitrema stenostoma</i>	<i>Amphitrema wrightianum</i>	0.11	0.65	5	5	0.08	0.06	0.01	0.06	0.06
<i>Arcella artocrea</i>	<i>Arcella catinus</i>	0.03	3.05	4	35	11.64	15.08	0.00	0.15	0.15
<i>Arcella discoides</i>	<i>Arcella vulgaris</i>	7.58	2.20	33	17	1.36	3.15	0.43	0.16	0.55
<i>Arcella gibbosa</i>	<i>Arcella hemispherica</i>	0.59	0.59	6	5	0.77	-0.23	0.02	0.02	0.05

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B) Jura

Taxon A	Taxon B	% total		N		WA optima		RMSE		
		A	B	A	B	A	B	A→B	A←B	A↔B
<i>Arcella artocrea</i>	<i>Arcella catinus</i>	0.10	1.64	7	19	13.16	26.33	0.06	0.88	0.92
<i>Centropyxis aerophila</i>	<i>Centropyxis platystoma</i>	2.10	0.95	17	8	17.17	23.31	1.10	0.50	1.07
<i>Corythion dubium</i>	<i>Trinema</i> type	5.31	3.70	33	20	24.97	26.38	0.49	0.34	0.36
<i>Cyclopyxis arcelloides</i>	<i>Diffugia globulosa</i>	0.55	0.24	7	1	11.12	3.00	0.02	0.01	0.03
<i>Cyclopyxis arcelloides</i>	<i>Phryganella acropodia</i>	0.55	2.99	7	28	11.12	28.25	0.32	1.76	1.95
<i>Diffugia longicollis</i>	<i>Diffugia oblonga</i>	0.37	0.02	3	1	27.35	16.00	0.26	0.01	0.27
<i>Euglypha alveolata</i>	<i>Euglypha tuberculata</i>	0.01	0.01	1	1	41.00	8.00	0.02	0.01	0.03
<i>Euglypha ciliata</i>	<i>Euglypha compressa</i>	2.08	0.29	31	8	21.66	26.25	0.72	0.10	0.69
<i>Euglypha ciliata</i>	<i>Euglypha strigosa</i>	2.08	1.04	31	19	21.66	25.78	0.30	0.15	0.27
<i>Euglypha laevis</i>	<i>Euglypha rounda</i>	1.66	2.62	22	24	24.24	24.75	0.27	0.42	0.47
<i>Euglypha strigosa</i>	<i>Euglypha compressa</i>	1.04	0.29	19	8	25.78	26.25	0.21	0.06	0.22
<i>Heleopera petricola</i>	<i>Heleopera rosea</i>	2.47	2.82	27	22	26.90	26.04	0.29	0.33	0.52
<i>Nebela bohémica</i>	<i>Nebela collaris</i>	0.72	0.23	6	5	20.68	23.20	0.13	0.04	0.09
<i>Nebela carinata</i>	<i>Nebela marginata</i>	0.18	0.91	5	9	8.82	9.59	0.01	0.05	0.05
<i>Nebela militaris</i>	<i>Nebela collaris</i>	6.62	0.23	30	5	27.85	23.20	0.81	0.03	0.83
<i>Nebela parvula</i>	<i>Nebela tinctoria</i>	0.04	14.68	2	37	29.35	29.29	0.01	5.87	5.86
<i>Nebela penardiana</i>	<i>Nebela tubulosa</i>	0.42	0.69	8	8	19.12	16.41	0.12	0.20	0.23
<i>Phryganella acropodia</i>	<i>Diffugia globulosa</i>	2.99	0.24	28	1	28.25	3.00	1.88	0.15	2.00
<i>Sphenoderia lenta</i>	<i>Tracheleuglypha dentata</i>	0.13	0.81	5	13	17.01	23.01	0.04	0.25	0.21

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C) Turkey

Species A	Species B	% total		N		WA optima		RMSEP		
		A	B	A	B	A	B	A→B	A←B	A↔B
<i>Phryganella acropodia</i>	<i>Cyclopyxis arcelloides</i>	1.04	0.27	22	3	39.74	9.34	0.03	0.00	0.03
<i>Cyclopyxis eurystoma</i>	<i>Phryganella acropodia</i>	0.84	1.04	8	22	68.28	39.74	0.22	0.35	0.37
<i>Cyclopyxis arcelloides</i>	<i>Cyclopyxis eurystoma</i>	0.27	0.84	3	8	9.34	68.28	0.55	0.17	0.72
<i>Corythion dubium</i>	<i>Trinema lineare</i>	8.24	1.41	31	13	47.40	63.76	1.65	0.35	1.59
<i>Euglypha compressa</i>	<i>Euglypha ciliata</i>	0.12	0.49	5	15	25.39	48.87	0.01	0.13	0.12
<i>Euglypha strigosa</i>	<i>Euglypha compressa</i>	0.07	0.12	4	5	30.29	25.39	0.01	0.01	0.01
<i>Euglypha strigosa</i>	<i>Euglypha ciliata</i>	0.07	0.49	4	15	30.29	48.87	0.00	0.03	0.03
<i>Heleopera rosea</i>	<i>Heleopera petricola</i>	3.45	0.08	27	2	41.03	28.59	0.90	0.01	0.90
<i>Nebela penardiana</i>	<i>Nebela tubulosa</i>	0.03	0.03	2	2	29.63	29.46	0.00	0.00	0.00
<i>Nebela tinctoria</i>	<i>Nebela penardiana</i>	0.47	0.03	14	2	43.69	29.63	0.01	0.00	0.01
<i>Centropyxis aerophila</i> type	<i>Plagiopyxis cf. callida</i>	2.33	0.06	20	2	57.28	12.62	0.38	0.01	0.38

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D) Alaska

Taxon A	Taxon B	% total		N		WA optima		RMSE		
		A	B	A	B	A	B	A→B	A←B	A↔B
<i>Arcella arenaria</i>	<i>Arcella artocrea</i>	2.02	0.10	58	4	35.79	30.92	0.44	0.02	0.46
<i>Centropyxis eicornis</i>	<i>Centropyxis laevigata</i>	0.76	1.26	19	20	28.35	44.19	0.48	0.80	1.28
<i>Centropyxis aerophila</i>	<i>Centropyxis platystoma</i>	3.05	0.12	38	5	26.43	28.06	0.95	0.04	0.93
<i>Corythion dubium</i>	<i>Trinema</i> spp.	4.81	0.96	48	33	31.44	29.41	1.32	0.26	1.10
<i>Diffugia globulosa</i>	<i>Phryganella acropodia</i> type	0.15	6.89	3	85	19.59	34.72	0.01	0.29	0.29
<i>Euglypha ciliata</i>	<i>Euglypha compressa</i>	4.95	0.83	67	28	35.76	37.60	0.78	0.13	0.75
<i>Euglypha ciliata</i>	<i>Euglypha strigosa</i>	4.95	0.23	67	11	35.76	23.47	3.06	0.14	2.97
<i>Euglypha strigosa</i>	<i>Euglypha compressa</i>	0.23	0.83	11	28	23.47	37.60	0.18	0.64	0.82
<i>Heleopera petricola</i>	<i>Heleopera sylvatica</i>	3.84	0.31	43	12	32.45	33.42	0.57	0.05	0.58
<i>Heleopera petricola</i>	<i>Heleopera sphagni</i>	3.84	3.74	43	33	32.45	24.39	1.17	1.14	2.13
<i>Nebela penardiana</i>	<i>Nebela marginata</i>	0.06	0.33	3	6	18.27	18.35	0.02	0.10	0.09
<i>Nebela tinctoria</i>	<i>Nebela penardiana</i>	3.25	0.06	60	3	42.25	18.27	2.74	0.05	2.78
<i>Hyalosphenia elegans</i>	<i>Nebela militaris</i>	3.98	1.76	47	40	32.03	46.80	2.59	1.15	2.71
<i>Euglypha rotunda</i>	<i>Tracheleuglypha dentata</i>	1.15	0.03	32	3	31.69	14.52	0.74	0.02	0.73
<i>Tracheleuglypha dentata</i>	<i>Sphenoderia lenta</i>	0.03	0.35	3	12	14.52	20.68	0.00	0.04	0.04

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