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1   **Geospatial Visualisation of Food Contaminant Distributions:**  
2   **Polychlorinated Naphthalenes (PCNs), Potentially Toxic Elements**  
3   **(PTEs) and Aflatoxins**

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5   Fernandes

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10   **Abstract:** Large volume of multidimensional data can be summarised, both in terms of  
11   tabulated statistics, and as graphic geospatial visualisations. The latter approach allows rapid  
12   interpretation and communication of complex information to stake-holders such as regulators,  
13   risk assessors and policy makers. In the main study on polychlorinated naphthalene (PCN),  
14   individual samples representing different edible fish species were analysed from around the  
15   UK. PCNs were observed in all samples with nearly all of the twelve measured congeners being  
16   detected. Summed congener concentrations ranged from 0.7 ng/kg ww (turbot) to 265 ng/kg  
17   ww (sprats). The highest contamination levels were recorded for sprats and mackerel with mean  
18   summed concentrations of 67 ng/kg ww and 68 ng/kg ww respectively. Two ancillary studies,  
19   on potentially toxic elements (PTEs) in crabs from China and aflatoxin in children's blood  
20   from Tanzania, demonstrate the wide applicability of this approach. The PTE contents in crab  
21   showed strong dependence on the tested tissues and elements, and crabs from Tai and  
22   Yangcheng Lakes showed obviously higher PTE levels than the other lakes. Geospatial  
23   distribution of the aflatoxin biomarker AF-alb in children's serum from 3 locations showed

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24 how individual anthropometric or socio-economic data reveals the relationship between family  
25 size, socio-economic score and magnitude of serum aflatoxin levels. In addition to facilitating  
26 the flow of interpreted data to stakeholders, these techniques can direct the formulation of risk  
27 mitigation activities and help with the identification of data gaps. When combined with  
28 hierarchical cluster analyses, correlations within the data can also be predicted.

29 **Keywords:** PCN-TEQ, fish, spatial analysis, visualization, risk mitigation, hierarchical  
30 clustering

## 31 1. Introduction

32 The visualisation of the spatial distribution of contaminant concentrations in foods  
33 sampled across large areas allows the rapid and efficient communication of information for  
34 regulators, risk assessors and managers. In particular, this mode of presentation of geospatial  
35 intelligence is especially relevant when multidimensional data is involved, where the sheer  
36 volume of generated data makes interpretation time-consuming and difficult. When combined  
37 with the outputs from the geospatial analysis of contaminant distributions in regions or  
38 countries, visualisation provides a powerful tool for the study of risk identification and  
39 establishment. It can also provide rapid and effective representation of affected or at-risk areas  
40 which aids the development of mitigation activities. Additionally, the technique can be  
41 complemented with hierarchical cluster analyses to allow the prediction of correlations within  
42 data sets.

43 Some web-based mapping resources such as Google Maps (Google Maps 2016) provide  
44 a powerful platform for the visualisation of geographical data. These platforms can be  
45 developed to efficiently represent and explore the complex inter-relationships between food  
46 contaminant occurrences, e.g. in different edible fish species or in different crop producing  
47 regions, as demonstrated in this paper. Using an interactive webpage, specific aspects derived  
48 from occurrence data can be represented in location and magnitude by different coloured circles

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49 or radiation patterns corresponding to their GPS position and to represent different species for  
50 example. The extent of contamination or the magnitude of the occurrence can be superimposed  
51 on this data by varying the radii of circles or the offset distances for radiation patterns.

52 Polychlorinated naphthalenes (PCNs) are legacy industrial chemicals widely used as  
53 dielectrics and flame retardants in the 20<sup>th</sup> century and are now recognised as persistent organic  
54 pollutants (POPs), as evidenced by their listing in Annex A and C of the Stockholm Convention.  
55 Environmentally, they occur ubiquitously with reports of contamination of biota, sediments  
56 and air, both in temperate as well as Polar Regions (Biddleman et al., 2010; Braune and Muir  
57 2017). Biochemically, they demonstrate properties of persistence and high bioaccumulation  
58 potential, coupled with a similarity in structural configuration to polychlorinated  
59 dibenzodioxins and furans (PCDD/Fs). This latter property is related to human health concerns  
60 as many PCN congeners have been reported to contribute to dioxin-like toxicity (Villeneuve et  
61 al., 2000; Falandysz et al., 2014; Fernandes et al., 2017) and elicit different effects such as  
62 mortality, embryotoxicity, hepatotoxicity, dermal lesions, teratogenicity and carcinogenicity  
63 (Behnisch et al., 2003; Falandysz et al., 2014). A recent review of their occurrence in human  
64 tissue and foods (Fernandes et al., 2017) show higher levels in fish relative to other foods.

65 A number of trace elements, particularly heavy metals, have long been recognised as  
66 potentially toxic elements (PTEs). Nutritionally, some trace elements, such as selenium, zinc,  
67 copper, etc. are essential to health but may be toxic at high levels of dietary intake.  
68 Environmental sources are the main contributors to contamination of food which is the major  
69 source of the overall exposure of consumers to PTEs, although other routes may also be  
70 significant (e.g. oral exposure via drinking water, occupational exposure by inhalation). Certain  
71 food groups naturally accumulate some elements (e.g. fish and shellfish are known to  
72 accumulate arsenic and mercury and cereals can accumulate cadmium) and consequently  
73 contain high concentrations of these elements compared to other foods. PTEs may enter marine

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74 and aquatic environments and bio-accumulate in species at any point during growth and  
75 harvesting. Chinese mitten crab (*Eriocheir sinensis*) is widely consumed in China because of  
76 its unique aroma and umami taste, but has also been reported (Xu et al., 2016; Shou et al.,  
77 2012) to be at high risk of PTE accumulation with the tissues of crabs exhibiting diverse PTE  
78 (e.g. lead and cadmium) bioaccumulation due to feeding behaviour and environmental factors.

79 Aflatoxins are contaminants that are produced by the moulds, *Aspergillus flavus* and  
80 *Aspergillus parasiticus*. Aflatoxin B1 is recognised as a genotoxic carcinogen and is the most  
81 potent of the fourteen aflatoxins that are natural products of these fungi. Within crops such as  
82 maize and groundnuts, the growth of these moulds is promoted by poor storage conditions  
83 combined with ambient humidity and temperature. Although hepatocellular carcinogenesis is  
84 generally recognised as the most lethal biological effect, the stunting of growth in childhood  
85 (affecting 165 m. children world-wide) that occurs in some countries in the African continent,  
86 may show an association with dietary aflatoxin occurrence (Gong et al., 2002; 2004). High  
87 levels of childhood aflatoxin exposure were observed during a study on three regions of  
88 Tanzania in 2010 (Shirima et al., 2013; 2015)

89 This paper focusses on the occurrence and geospatial mapping of these contaminants to  
90 provide effective visualisation of their distribution within geographical regions. The main focus  
91 of this work is on the distribution of PCN contamination in fish, but the applicability of this  
92 approach is also demonstrated for PTEs and aflatoxins, in different geographical locations. For  
93 PCNs, occurrence is mapped for different edible marine fish species from locations in the North  
94 Atlantic Ocean. PTE contents of three types of tissues (muscle tissue, hepatopancreas and  
95 sexual organs) of Chinese mitten crabs were determined, and mapped for samples taken from  
96 5 lakes in Jiangsu province, China. For Aflatoxin B1, visualisation of human exposure as  
97 measured in blood AF-alb biomarkers in 3 rural locations in Tanzania was superimposed with  
98 related anthropometric data to examine correlations between these variables and exposure.

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99      **2. Experimental**

100     **2.1 Polychlorinated Naphthalene study**

101       In the main study on PCNs, 75 samples of Sea Bass, Herring, Mackerel, Mullet and  
102      Sardine were identified, as part of a larger study investigating other contaminants such as  
103      brominated flame retardants, PCDD/Fs, etc. Samples were collected in major fishing areas,  
104      mainly but not exclusively, from waters around the UK, Ireland and Northern France. In  
105      general, selected fish sizes approximated to those that are commercially sold for consumption.  
106      The following PCN congeners were included in this study - PCN-52/60, 53, 66/67, 68, 69,  
107      71/72, 73, 74, & 75.

108       The analytical methodology for the determination of PCN concentrations in fish species  
109      has been described elsewhere (Fernandes et al., 2010). In brief, sample aliquots were fortified  
110      with <sup>13</sup>C-labelled PCN analogues of target compounds and extracted using mixed organic  
111      solvents. PCNs were chromatographically fractionated from potential interferants such as  
112      PCBs, on an activated carbon column and further purified using adsorption chromatography  
113      on alumina. Analytical measurement was carried out using high resolution gas chromatography  
114      coupled to high resolution mass spectrometry (HRGC-HRMS). Additional control was  
115      provided by the inclusion of procedural blanks and a reference material.

116       The quality control criteria used for evaluating data, and method performance parameters  
117      have been reported before (Fernandes et al., 2010). There are no available reference materials  
118      (RMs) specific to PCNs, but the same reference material that is used for PCDD/F and PCB  
119      analysis (cod liver oil), was analysed during the course of this work with results showing good  
120      consistency and agreement with established values. The reporting limits (LOQ quoted as “<”)  
121      for all analytes incorporate the relevant procedural blank and were derived using the current  
122      guidance on LOQ estimation (European Commission, 2017).

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123     **2.2. Potentially Toxic Elements in Crabs**

124       The sampling sites included 5 major lakes in Yangtze River Delta, China, i.e., Yangcheng  
125      Lake, Tai Lake, Changdang Lake, Gucheng Lake and Hongze Lake. 8 crabs were collected  
126      from each lake, rinsed with tap water, and dissected quickly to obtain the sexual gland,  
127      hepatopancreas and muscle. The sample tissues were digested using a microwave-assisted  
128      method (López et al., 2003), then Cadmium (Cd), Lead (Pb), Mercury (Hg) and Copper (Cu)  
129      in the digested solutions were measured by inductively coupled plasma mass spectrometry  
130      (F.J.Sánchez L., 2003) (ICP-MS, XSeries II, Thermo Fisher Scientific, USA). Blanks and QC  
131      samples included and the LOQs ranged from 3-5 µg/kg.

132     **2.3 Aflatoxins in Serum**

133       The aflatoxin data is derived from a study (Shirima et al., 2015) that investigated 3 regions  
134      in Tanzania (Iringa, Tabora and Kilimanjaro). Dietary intake data and anthropometric indices-  
135      body weight, recumbent length, etc. were recorded. Validated methodology was used to  
136      measure aflatoxin–albumin adducts (AF-alb), as biomarkers in blood plasma taken from 166  
137      children. A description of this methodology and quality parameters has already been given  
138      before (Chapot., 1991) but in brief, the tests included extraction of albumin, digestion of  
139      protein, purification, and ELISA quantification of AF-alb adducts. Each batch of plasma was  
140      analysed with three positive controls and one negative control. Samples were measured in  
141      ELISA in quadruplicate on at least two occasions on separate days. The detection limit was 3  
142      pg of aflatoxin-lysine equivalents per milligram albumin.

143     **2.4 Geo-spatial mapping**

144       For the geo-spatial mapping, an interactive webpage based on Google Maps was used (Li  
145      et al., 2018) to integrate sample location (GPS), with contaminant concentration data. The  
146      webpage primarily consisted of user selective control fields such as individual or summed

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147 contaminant, species, occurrence scaling, colour legend of selected species/contaminants and  
148 a help field. Each sample was presented with a circle centred at its GPS coordinate location.  
149 The magnitude of occurrence was indicated by the radius of each circle (ng/kg for PCNs; µg/kg  
150 for PTEs; pg/mg for AF-alb), which can be simultaneously colour coded to variables such as  
151 socio-economic status. The design allows multiple species or congeners to be selected,  
152 differentiated by colour coding and line type (solid, dashed, dotted) as applied to relevant circle  
153 representations. The aflatoxin study used plasma AF-alb biomarker data and/or contaminant  
154 levels within maize/other foods at the designated sampling region together with hazard  
155 indicators such as child weight z-scores. This approach offers efficient data exploration without  
156 full statistical analyses.

157 With the volumes of data generated in this study, hierarchical clustering (HC) analysis can  
158 complement the above approach, a feature which would allow the user to investigate if certain  
159 locations or species/individuals are susceptible to concentration levels of concern. HC is a  
160 typical algorithm to analyse the similarities (or dissimilarities) of objects in variable space  
161 (Smoliński et al., 2002). In order to have a better understanding of the geographical  
162 distribution, or independence of occurrence of the congeners, HC was employed in the PCN  
163 study e.g. by dividing the fish samples into 3 clusters with the PCN congener concentration as  
164 input variables to investigate if correlations existed between the species and spatial locations.  
165 HC analysis was performed using R language with the Ward method as the amalgamation rule  
166 and Euclidean distance as metric (Smoliński et al., 2002).

### 167 **3. Results and Discussion**

168 The raw data from the PCN study is very large and has been presented in a sponsor report  
169 (Fernandes et al., 2015). Raw data from the mitten crab study is provided in the supplementary  
170 information (SI). Descriptive statistics are summarised in Tables 3.1 and 3.2, by analyte type.

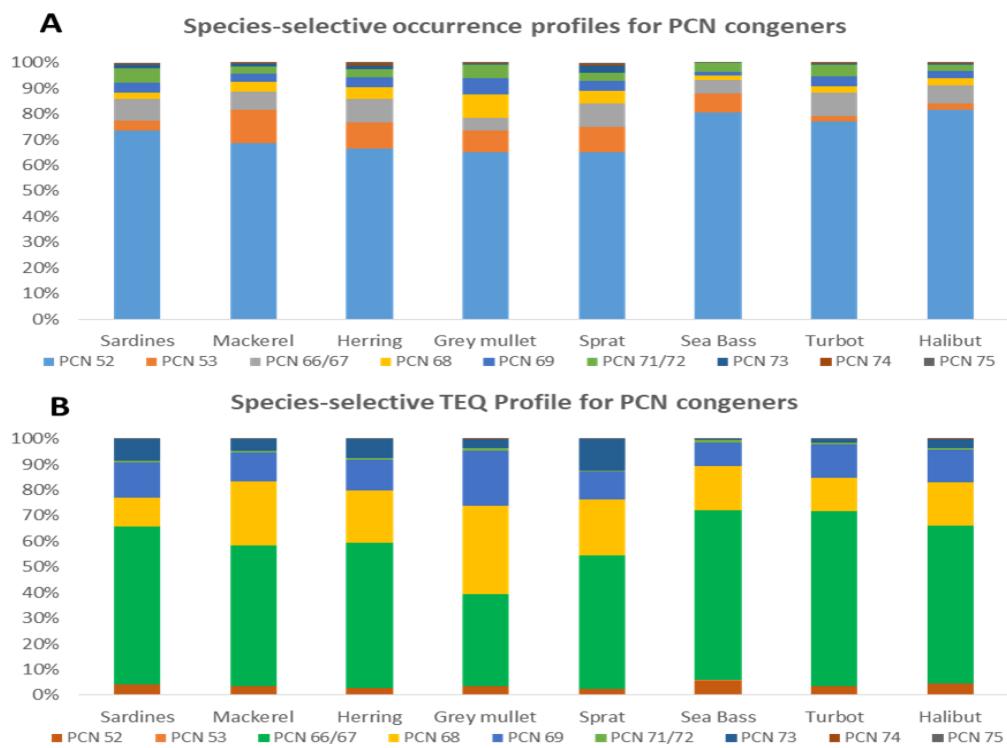
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171     **3.1 PCN**

172       PCN concentrations are reported in ng/kg wet weight (ww) as per convention. Congeners  
173       were selected for measurement based on toxicological significance as described in earlier  
174       studies (Fernandes et al 2010; Falandysz et al., 2014). From the point of view of food safety,  
175       and given the known AhR receptor active responses (Falandysz et al., 2014; Fernandes et al.,  
176       2017), these concentrations have also been converted to toxic equivalents (TEQs) using a  
177       similar approach to PCDD/Fs and dioxin-like PCBs (Van den berg et al., 2006). The TEQ  
178       values were calculated using relative potencies (REPs) as used in other studies (Fernandes et  
179       al., 2010; 2011; 2017; Jongchu et al., 2018; Lili et al., 2018).

180       PCNs were detected in all of the samples analysed with variations in patterns of  
181       occurrence and levels of concentrations depending on the species and location. Higher  
182       concentration levels were recorded for penta- and hexa-chlorinated PCNs (e.g. PCNs 52, 53  
183       and 66/67) with lowest levels occurring mainly for the octa-chlorinated PCN-75. The extent of  
184       positive detection was high, at 86% of all measurements. The patterns of occurrence are  
185       dominated by PCN-52 but the occurrence of other congeners varies widely depending on the  
186       species as shown in Figure 3.1A (average pattern for each species, except for halibut which  
187       was a single sample), whereas the absolute concentrations appeared to be more dependent on  
188       the location of the sample. Summed concentrations (sum of twelve measured congeners)  
189       showed a range extending from 0.7 ng/kg ww (turbot) to 265 ng/kg ww (sprats). Highest  
190       concentrations were recorded for sprats and mackerel with mean concentrations of 67 ng/kg  
191       ww and 68 ng/kg ww respectively. Given the large geographical range of the sampling region,  
192       these observations are indicative, although other studies (Fernandes et al., 2009) appear to show  
193       similar findings. However, other parameters such as size and the age of the fish, and seasonal  
194       physiological changes (e.g. spawning periods, etc.) will undoubtedly influence the findings.

195



196

197 Figure 3.1 Average pattern (A), of PCN congener occurrence and (B), TEQ, based on species

198 Table 3.1. Statistical summary of PCN occurrence in various species of edible marine fish

Species (N) & average fat	Sum PCNs	Sum PCNs	*PCN TEQ	Species (N) & average fat	Sum PCNs	Sum PCNs	*PCN TEQ		
	Lipid basis	whole weight	whole weight		Lipid basis	whole weight	whole weight		
	ng/kg fat	ng/kg w.w.			ng/kg fat	ng/kg w.w.			
Sardines (12)	MIN	22	5.4	0.004	Sprat (15)	MIN	138	29.4	0.014
	MEDIAN	172	16.6	0.007		MEDIAN	335	46	0.027
	MEAN	291	19.8	0.009		MEAN	680	66.5	0.044
	MAX	1215	63.1	0.031		MAX	2390	264.8	0.204
Mackerel (14)	MIN	112	10.1	0.002	Sea Bass (13)	MIN	302	14.2	0.004
	MEDIAN	451	50.5	0.024		MEDIAN	848	29.2	0.008
	MEAN	648	68	0.035		MEAN	999	29.4	0.01
	MAX	1654	243	0.17		MAX	3084	48.5	0.026
Herring (6)	MIN	141	18.3	0.009	Turbot (6)	MIN	132	0.7	<0.001
	MEDIAN	231	29.7	0.016		MEDIAN	246	3.5	0.002
	MEAN	431	38.7	0.024		MEAN	343	5.3	0.003
	MAX	1342	89.5	0.069		MAX	828	15.5	0.009
Grey mullet (9)	MIN	125	4.2	0.001	Halibut (1)	MIN	253	5.85	0.003
	MEDIAN	554	12.4	0.006		MEDIAN	2.3		
	MEAN	1293	14.7	0.007		MEAN			
	MAX	7572	33.5	0.014		MAX			

199 \*PCN TEQ calculated using REP values from Fernandes et al. 2010 Env. Sci. Technol., 44, 3533–3538. Sum

200 PCNs = sum of 12 congeners

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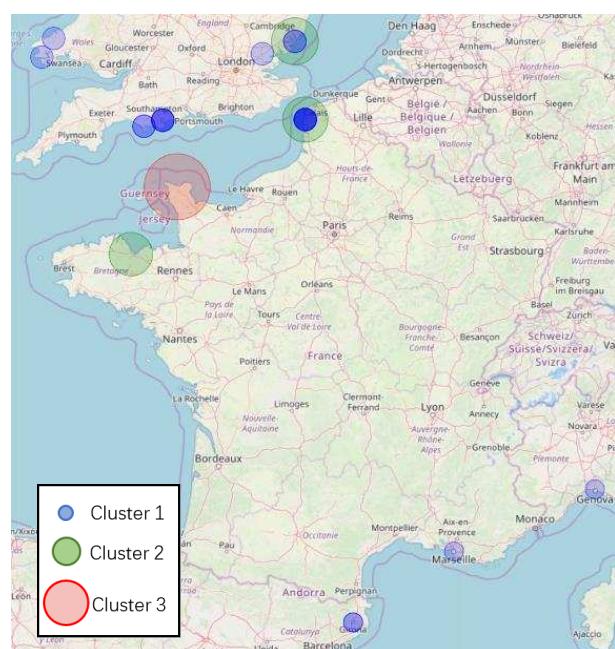
201       The variation in congener occurrence is reflected in the TEQ distribution (Figure 3.1B),  
202   although here, the higher REPs of congeners PCN-66/67, PCN-68 and PCN-69 (Falandysz et  
203   al., 2014) make them the largest contributors to the cumulative TEQ. The TEQ values for these  
204   fish samples (Table 3.1), range from <0.001 to 0.2 ng/kg ww and correlate well with the  
205   summed concentrations, with highest average TEQ concentrations for mackerel and sprats at  
206   0.035 and 0.044ng/kg ww. This range of values is substantially lower than the corresponding  
207   TEQ arising from PCDD/Fs and PCBs in fish, but it does contribute to the overall burden of  
208   dioxin-like toxicity.

209       Preliminary geospatial mapping analysis showed that the variables that would provide the  
210   most useful interpretation of the results were likely to be location, species and PCN congener  
211   (including sum of congeners). The visualisation plots of the concentrations (Figure 3.2A)  
212   revealed that although locations across the southern/eastern UK coasts and northern France  
213   showed a majority of concentrations above the average for sum PCNs (39 ng/kg ww), the  
214   highest PCN concentrations were recorded in samples of mackerel and sprats from the Irish  
215   Sea. This distribution is confirmed when plots of the indicative PCN congeners are examined  
216   e.g. PCN-52 as the most abundant congener measured in this study (Figure 3.3A), and PCN-  
217   66/67 as the largest contributor to the PCN TEQ (Figure 3.3B). In terms of species, Figure 3.2A  
218   also reveals that of the measured species, mackerel, sprats and sea-bass are the most susceptible  
219   to PCN contamination. Considering the significant difference between species, analysis of  
220   individual species was necessary to examine the spatial distribution of PCNs exposure. Using  
221   grey mullet as an example, HC analysis was performed to determine the correlations between  
222   grey mullet samples with PCN congener concentrations used as input variables. Figure 3.2B  
223   clearly shows that cluster 2 and cluster 3, present higher level of PCNs and appeared along the  
224   north coast of France.

A



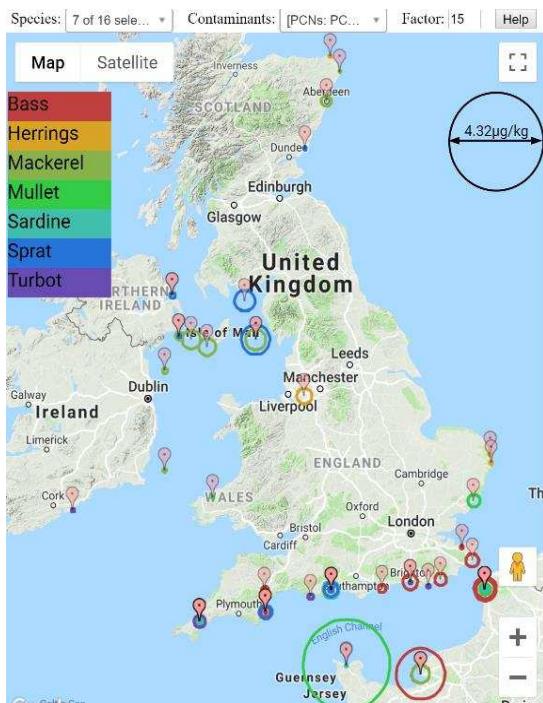
B



225

226 Figure 3.2. (A) Spatial distribution of sum PCN congeners for locations around the UK;  
227 (B) HC analysis for PCN levels of Grey mullet

A



B



228

229 Figure 3.3. Spatial distribution of A. PCN-52 and B. PCN-66/67

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230     **3.2 Analysis and visualization of PTEs in Crab**

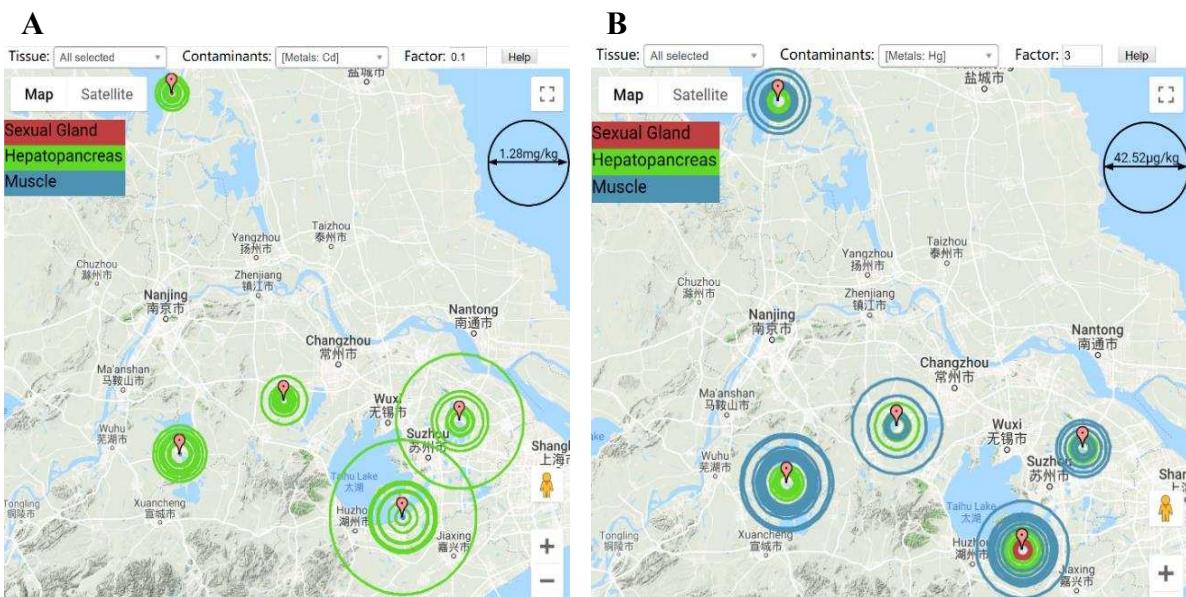
231       A statistical summary of the measured PTE concentrations in different crab tissues is  
232       presented in Table 3.2. The relatively high occurrence of Cu in all samples follows a normal  
233       distribution for fish and shellfish (De Mora et al., 2004) species but concentrations for some of  
234       the other PTEs showed a strong dependency on the type of tissue. The geometric mean  
235       concentration for Cd in the hepatopancreas was significantly higher than in the sexual gland or  
236       muscle. Conversely, the mean concentration of Hg in muscle was approximately double the  
237       concentration in the hepatopancreas and about 50-fold higher than in the sexual gland. There  
238       was no significant difference observed for Pb in different tissue in terms of the geometric mean.  
239       The maximum concentrations generally reflect the means apart from Hg in the sexual gland  
240       which shows a large difference. Such characteristics of PTE occurrence in crab tissues can be  
241       conveniently superimposed on the geographical locations of the samples using geospatial  
242       mapping. In Figure 3.4A the distribution of Cd with location as a variable, immediately displays  
243       the higher occurrence levels in the Tai and Yangcheng lakes relative to the other lakes. In this  
244       plot, only Cd in the hepatopancreas is immediately apparent, mainly due to the higher  
245       concentrations in this tissue (131 to 2382 µg/kg with an average value of 579 µg/kg, compared  
246       to e.g. <5.0 to 21 µg/kg with average value of 2.4 µg/kg in the sexual gland). Although the  
247       immediate effect is to obscure the detail for the muscle and the sexual gland, this can easily be  
248       overcome by using the scaling factor to expand the scale of the plot as shown in Figure 3.4B  
249       which shows the distribution for Hg plotted at a much higher magnification. Here the highest  
250       concentrations occur in the muscle tissue. The relatively high accumulation of Cd in  
251       hepatopancreas was also observed by Ma et al., 2008 and Wang et al., 2008, who suggested  
252       that this could be due to the hepatopancreas playing an important role in Cd detoxification in  
253       this species (Hopkin and Nott, 1980).

254       Besides single-factor comparison, multiple factors including tissue, location and element,

could be compared at the same time with a radiation pattern. As shown in Figure 3.5, each vertex of the radiation pattern presents an element, the colour presents tissue part. Figure 3.5 clearly shows that PTE occurrence in the crabs strongly depended on tissue type and individual PTEs, which is consistent with Table 3.2 and Figure 3.4. Spatially, crabs from Tai Lake and Yangcheng Lake showed higher level of PTE bio-accumulation than the other 3 lakes.

Table 3.2. Statistical summary of the measured PTE contents in different crab tissues

Tissue parts	Statistics	PTEs ( $\mu\text{g}/\text{kg}$ )			
		Cd	Pb	Hg	Cu
Hepatopancreas	Min.	131	14.6	<3.0	3687
	Median	461	40.9	11.5	25123
	Mean	579.	52.3	11.4	30424
	Max.	2382	210.2	25.2	79060
Sexual Gland	Min.	<5.0	12.9	<3.0	3996
	Median	<5.0	42.4	<3.0	19469
	Mean	2.4	59.0	0.4	24278
	Max.	21.0	277.1	9.2	59659
Muscle	Min.	<5.0	17.2	<3.0	7763
	Median	<5.0	32.1	23.6	15485
	Mean	3.9	50.2	21.7	16040
	Max.	42	349	47	31425



261

Figure 3.4. Spatial distribution of A. Cd and B. Hg for different tissues of crab



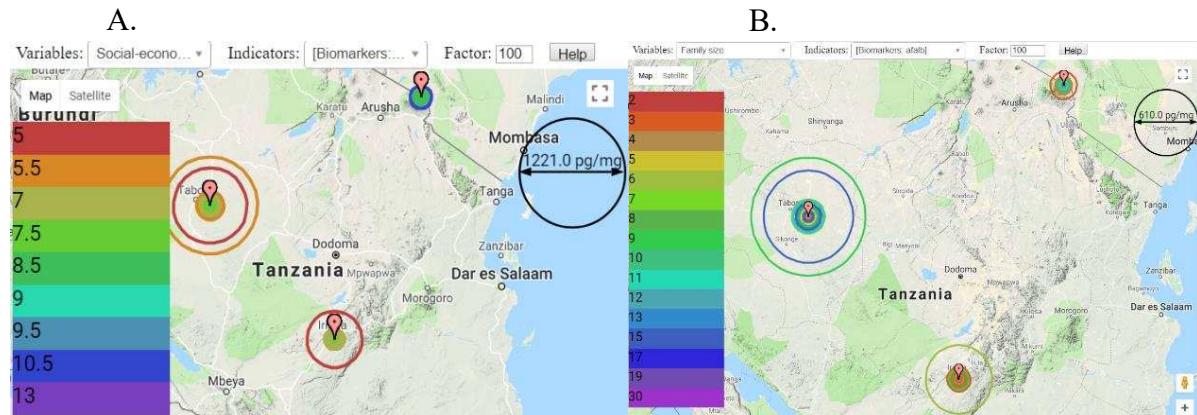
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264 Figure 3.5. Radiation pattern of Cd, Hg and Pb for different tissues

265 **3.3 Aflatoxin**

266 These preliminary results of aflatoxin analysis demonstrate the initial exploratory  
 267 potential of the application. Geospatial analysis was applied to the distribution of the aflatoxin  
 268 biomarker AF-alb in children's blood from 3 locations, to individual anthropometric or socio-  
 269 economic data. The examples in Figures 3.6A and 3.6B show the magnitude of AF-alb content  
 270 overlaid with socio-economic status and family size respectively and shows how potential  
 271 associations for indicators may be explored. The Tabora region (located centre of Figures 3.6A  
 272 & B) shows relatively stronger associations. Generally, the characteristic of poorer social  
 273 economic state (lower score) show higher level of aflatoxin biomarker, which means more risk  
 274 for aflatoxin exposure. Similarly, correlation to child hazard data (e.g. stunting) can be  
 275 performed to evaluate biomarker to hazard indicators.

276



277 Figure 3.6. Selected spatial distribution for the aflatoxin biomarker AF-alb in Tanzania with A.  
278 associated socio-economic status colour coding and B. family size colour-coding, respectively.

### 279 3.4 General Discussion with focus on advantages of geo-spatial visualisation

280 Figures 3.2-3.6 demonstrate rapid and effective geo-spatial visualisation of the large  
281 amount of data associated with these studies. Although spatially distributed data can also be  
282 summarised in tabular format, navigating listed numerical data can be time consuming and  
283 tedious when compared to the amount of information that is quickly apparent from the visual  
284 plots. Areas of higher contamination that may relate to relatively higher risk, or locations  
285 requiring attention, are immediately apparent from the plots and enable rapid qualitative  
286 assessment of potential associations within data. So for example, in the case of the PCN  
287 contamination of edible fish species, it is immediately apparent from the plots that fish from  
288 the northern reaches of the Irish Sea between Northern Ireland and South-West England, and  
289 the English Channel show a higher levels of PCNs. This also allows for possible geographical  
290 associations with sources to be made e.g. Commercial PCN sold as Seekay waxes were  
291 manufactured at Runcorn (Crookes and Howe, 1993) on the southern bank of the river Mersey  
292 which flows into the Irish Sea. In the case of the AF-alb data from Tanzania, the introduction  
293 of socio-economic data relating to the families of the children who provided blood samples, as  
294 a variable within the geospatial analysis, allowed potential associations with poorer socio-  
295 economic scores and larger family sizes to be correlated with higher AF-alb levels in the blood

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296 (Figure 3.6). The plot also shows that the relatively affluent area around Kilimanjaro in the  
297 north of the country was associated with smaller family sizes and lower AF-alb blood levels.

298 If sufficient data are available, more detailed analyses, e.g. HC, can be conducted or  
299 alternatively, where data are lacking, this promotes more informed planning, e.g. where  
300 additional research/data acquisition is required, and enables greater transparency across  
301 stakeholders. These techniques and the resulting visualised data also allow interpreted  
302 information to become more amenable to non-specialist stakeholders, e.g. risk assessors and  
303 managers who are then empowered to direct the formulation of risk mitigation activities and  
304 also help with the identification of data gaps. Another aspect of the techniques is the ability to  
305 retrospectively add new types of data that may become available e.g. rates of crop failure or  
306 any changes due to temporal or climatic conditions or spontaneous events.

307 **4. Conclusions**

308 The geo-spatial visualisation approach applied to the data presented here allows the rapid  
309 and effective evaluation of complex and multidimensional data sets. Three diverse datasets on  
310 PCNs, PTEs and aflatoxins were analysed demonstrating the versatility of the system. In the  
311 main study on PCNs, correlations between location and species with PCN congeners was  
312 demonstrated, showing that mackerel and sea-bass were the edible species that were most  
313 susceptible to PCN contamination, and the Northern reaches of the Irish Sea and the English  
314 Channel showed relatively higher levels of contamination than the other marine areas studied.  
315 In two ancillary studies, data on potentially toxic elements (PTEs) in crabs from 5 lakes in  
316 China and aflatoxin levels in maize from three regions of Tanzania, demonstrate the wide  
317 applicability of this visualisation approach. The PTE contents in crab showed strong  
318 dependence on the tested tissues and elements, and crabs from Tai Lake and Yangcheng Lake  
319 showed obviously higher level of PTE accumulation than the other lakes. Geospatial  
320 distribution of the aflatoxin biomarker AF-alb in children's blood from 3 locations in Kenya

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321 showed how individual anthropometric or socio-economic data reveals the relationship  
322 between family size, socio-economic score and magnitude of serum aflatoxin levels. In  
323 addition to facilitating the flow of interpreted data to stakeholders, these techniques can help  
324 direct the formulation of risk mitigation activities and also help identify gaps in data and  
325 knowledge.

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