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1	Spatial analysis of polybrominated diphenylethers (PBDEs) and
2	polybrominated biphenyls (PBBs) in fish collected from UK and
3	proximate marine waters
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11	Abstract: Some commonly consumed marine fish species are considered to display a
12	higher risk of bio-accumulating organic environmental contaminants such as PBDEs. As part
13	of a study to investigate the spatial distribution of these contaminants, data on polybrominated
14	diphenlyethers (PBDEs) and polybrominated biphenyls (PBBs) were collected and analysed
15	by introducing a web-based resource which enables efficient spatial, species and concentration
16	level representations. Furthermore, hierarchical cluster analyses permits correlations within the
17	data to be predicted. The data provide current information on levels of PBDE and PBB
18	occurrence, allowing identification of locations that show higher contaminant levels.
19	135 fish samples of various species were analysed from UK marine waters, but
20	encompassing the waters around Norway in the North and to the Algarve in the South. PBDEs
21	were observed in all samples with the majority of measured congeners being detected. The
22	concentrations ranged from 0.087 μ g/kg to 8.907 μ g/kg whole weight (ww) for the sum of all

23 measured PBDE congeners. PBBs occurred less frequently showing a corresponding range of 24 $< 0.02 \ \mu g/kg$ to 0.97 $\mu g/kg$ ww for the sum of seven PBB congeners. Concentrations vary depending on species and locations where landed, e.g. PBBs occurred more frequently and at 25 26 higher levels in grey mullet from French waters. The high frequency of PBDE occurrence 27 makes it prudent to continue the monitoring of these commonly consumed marine fish species. 28 The web-based resource provides a flexible and efficient tool for assessors and policy-makers to monitor and evaluate levels within caught fish species improving evidenced-based decision 29 30 processes.

31 Keywords: polybrominated diphenlyethers, polybrominated biphenyls, fish, spatial analysis,

32 visualization, hierarchical clustering

33

34 **1. Introduction:**

The bioaccumulation of environmental contaminants by various marine fish and shellfish species has been widely documented (Bruggeman et al., 1984; Magalhaes et al., 2003; El-Moselhy et al., 2014; Hashizume et al., 2014). Thus, the consumption of edible species of marine fish and shellfish has the potential to make a significant contribution to the human exposure of these contaminants.

In an effort to reduce or prevent inputs that could cause pollution, affect human health or adversely impact legitimate uses of the marine environment, the Marine Strategy Framework Directive encourages collaboration and coordination between individual EU Member States with the aim of protecting and preserving marine ecosystems. In the context of the present study, one of the targets for good environmental status under the directive is the limiting of contamination in fish and other seafood along with compliance with maximum contaminant levels established by European Commission regulation, or other relevant standards. In addressing this aim, complex and large datasets require analyses which encompasses spatial,
species and concentration levels presenting a challenge to assessors to be able to clearly
identify trends or correlations.

50 PBDEs are classified as legacy brominated flame retardants (BFRs) that were 51 manufactured for incorporation into a number of commonly used commercial materials such 52 as plastics, rubbers, textiles and electronic components. These open-ended applications allow 53 the PBDEs to diffuse out of materials into the environment, and this can occur during 54 manufacture, use and disposal of the product.

55 As an important analogue of PBDE, polybrominated biphenyls (PBBs) had been primarily produced as BFRs during the early 1970s until the Michigan incident in 1973 (Luross et al., 56 57 2002). Due to their persistence in the environment, PBBs still show (environmental, food?) 58 occurrence in the reports published recently (Luross et al., 2002; Bramwell et al., 2017; Chang et al., 2017). Both PBDEs and PBBs. could bioaccumulate through the food chain, and cause 59 60 endocrine system disruption, neurobehavioral effects and reproductive toxicity (McDonald, 61 2002). Additionally, they may be particularly harmful during a critical window of brain 62 development during pregnancy and early childhood (Fernandes A, 2012; Chiaramonte and Zota, 2016; Schrenk and Cartus, 2017). The detection and analysis of the concentration level 63 64 and geographical dispersion of these contaminants are of great significance, which will provide a useful reference for the development of related policy. Even though their occurrence in food 65 66 has been investigated (FSA, 2006; Fernandes A, 2009), a more comprehensive investigation is 67 still needed.

As an example, PBDE/PBB concentration data in various fish species caught across various locations are analysed (Fernandes et al., 2004; Fernandes et al., 2008). This dataset exemplifies the problem of variation in concentration levels across congeners (orders of magnitude), species and geographical locations and we present a proposed methodology to

efficiently explore spatial distribution in species and congeners to demonstrate an efficient and
 intuitive representation method.

74 Google Maps is a web mapping service which provides a powerful platform for 75 geographical data visualization (Google Maps 2016). To have a better understanding of the 76 PBDE dispersions around UK, and also to provide a convenient method to visualize 77 geographical information for both, professionals and laypersons, this paper introduces a novel web-based resource developed from google maps to efficiently represent and explore the 78 79 complex inter-relationships between the fish species, contaminants, catch locations and 80 statistical clustering within the data. Each sample was presented with a circle or radiation 81 pattern marked at its GPS position on an interactive webpage with different colours used to 82 distinguish the species or contaminants, and radius (for circles) or offset distance (for radiation 83 pattern) to represent the contaminant concentration. The webpage also allows users to select 84 individual interested species and contaminants. Concentration levels across congeners may 85 vary by orders of magnitude, therefore a scaling factor is included to allow easy calibration of 86 the presented data. Thus, this paper aims to present a geographical data visualization method 87 to present complex concentration data sets across multiple species, congeners and locations and provides results of PBDE/PBB data obtained from a study to investigate occurrence in 88 89 commonly consumed marine fish species in UK and proximate marine waters as an example 90 application.

91 **2. Methods and Materials**

92 **2.1 Sample preparation and analysis**

135 samples were collected mostly from waters around the UK, but extending to other
North Atlantic regions around Norway, the Irish Sea, and the European coastal North Atlantic
regions, including the North Western coast of France, Biscay and the Algarve. The main species

96 targeted were sardines, sprats, sea bass, mackerel, herring, grey mullet, but other species were97 also included.

Whole fish were dissected to collect edible muscle tissue and to exclude skin, bones and organs. However, for some species such as sprats, whole fish were used. This reflects consumer practice for fish consumption. The selected tissue was minced, homogenised by blending and freeze-dried. Dry samples were stored at -18 °C and re-homogenised before analysis.

102 The method used for the preparation, extraction and analysis of samples has been reported 103 previously (Fernandes et al., 2004; Fernandes et al., 2008). In brief, samples were fortified with 104 ¹³C-labelled analogues of target compounds and exhaustively extracted using mixed organic 105 solvents. PBDEs and ortho substituted PBBs were separated from other contaminants by 106 fractionation on activated carbon. The fraction was purified using adsorption chromatography on alumina. Analytical measurement was carried out using high resolution gas 107 108 chromatography-high resolution mass spectrometry (HRGC-HRMS). The analysis is ISO 109 17025 accredited (UKAS) and includes an in-house reference material and method blanks 110 which were evaluated prior to reporting of sample data and used to determine the limits of 111 detection.

PBDE congeners analysed (IUPAC numbers 17, 28, 47, 49, 66, 71, 77, 85, 99, 100, 119,
126, 138, 153, 154, 183 and 209) include those specified in European Commission
recommendation 2014/118/EU which are given in bold font. PBB congeners included: IUPAC
numbers 49, 52, 80, 101, 153 and 209.

116 **2.2 Software and data analysis**

117 2.2.1 Interactive webpage for data visualization

To efficiently visualize the geographical dispersion of the contaminants, an interactive webpage (<u>www.pbde.droppages.com</u>) was designed based on Google Maps which utilised sample GPS data and the sample concentrations. As shown in Fig.1, the webpage primarily 121 consists of the following control fields; species selection, contaminant selection, factor selection (for scaling), colour legend of selected species/contaminant and a help field. This 122 permits a convenient facility to explore required permutations of the available underlying data. 123 124 Where a single fish species and contaminant are selected, each sample is presented with a circle located at its reported catch location based on the Global Positioning System (GPS) 125 coordinates and where the radius of each circle reflects the level of the measured concentration 126 (ng/kg) within the sample. The concentration in each sample is represented by a variable radius 127 128 defined as in equation (1),

129

$$radius = Factor \times C$$
(1).

130 Where C is the contaminant concentration, *Factor* is the scaling factor which is set by default

131 automatically but which may be modified by the user dependent upon the contaminant

132 concentration which vary by orders of magnitude between the congeners and also allows

133 recalibration due to the map size selected.

134 Where multiple species or contaminants are selected, differentiation between the 135 species/contaminants is achieved by the use of colour codes and line type (including solid, 136 dashed and dotted) shown in the legend and which is applied to the relevant circle 137 representations. Additional data retrieval within the same species may be achieved by placing the cursor over a particular sample, in which case an information window will be presented to 138 139 display the cumulative distribution function of the sample set and which provides the ranking 140 of current sample (red star) and, if required, specific details for that sample will be displayed if the cursor is moved over the red star. 141



Fig.1 Interactive webpage for data visualization based on google maps GPS coordinates of where the sample was reported as caught. Circle radius indicates concentration level and inset indicates total sample set information. Additional fields show species selection, contaminant selection, factor selection (for scaling), colour legend of selected species/contaminant and a help field.

143

149 If multiple species and contaminants are selected simultaneously, each sample is 150 presented with a radiation pattern as shown in Fig.2. Different colours are used to distinguish 151 different species and all selected contaminants are arranged in sequence at the vertices of a k -152 polygon in a clockwise direction. The radiation pattern is centred at the sample catch location 153 and is associated with the contaminant concentration as described in equations (2) and (3)

154





$\theta_{\rm k} = 500 / \Pi \times {\rm K}$	(2)

$$\mathbf{d}_{\mathbf{k}} = \mathbf{Factor} \times \mathbf{C}_{\mathbf{k}} \tag{3}$$

156 Where *n* is the total number of the selected contaminants, *k* is the sequence number of the 157 k^{th} vertex which represents the k^{th} contaminant, θ_k is the heading angle of the k^{th} vertex 158 measured clockwise from true north (0 degrees), d_k is the distance between the vertex and the 159 sample location, which is proportional to the contaminant concentration C_k . Again, to provide 160 user control, *Factor* is the scaling factor chosen dependent upon the map size and allows 161 recalibration due to the levels of the contaminant concentration.





Fig.2 Illustration of the radiation pattern to display multiple species and contaminants simultaneously. The
 polygonal architecture is defined using equations (2) and (3).

165 2.2.2 Hierarchical Clustering analysis

Hierarchical Clustering (HC) is a typical algorithm to analyse the similarities (or 166 167 dissimilarities) of objects in the variable space (Smoliński et al., 2002; Chen et al., 2005). To have a better understanding of the geographical distribution of the PBDE congeners, HC was 168 169 employed to divide the fish samples into 3 clusters with the PBDE congener concentration as 170 input variables to investigate if correlations exist between the species and spatial locations. 171 This particular feature will enable assessors to investigate if certain locations or species are 172 likely to have certain concentrations levels. The analysis was performed using R language with the Ward method as the amalgamation rule and Euclidean distance as metric. 173

174 **3. Results & Discussion:**

175 **3.1 General comparison between different species**

PBDEs were observed in all samples with most measured congeners being detected. The concentrations ranged from 87 ng/kg to 8907 ng/kg whole weight for the sum of all measured PBDE congeners (64 ng/kg to 8635 ng/kg for the ten EU listed PBDEs). A summary of the data are presented in Table 1. There are only minor differences between the average values for both the sum of the 23 congeners (PBBs included) and the sum of the 10 congeners specified in the 181 EC recommendations, which confirms a more informed choice of congeners for the EU list.

182 PBBs occurred less frequently showing a corresponding range of $< 0.02 \ \mu$ g/kg to 0.97 183 μ g/kg ww for the sum of seven PBB congeners.

184 Fig.3 shows the dependence of mean value for the sum of the PBDEs on the fish species, and the concentration distribution of the top five congeners, including BDE-47, BDE-49, BDE-185 186 100, BDE-99 and BDE-154, in different species. In terms of the sum of the PBDEs in each sample, Herring and Bass ranked at the top level with a mean value around 2100 ng/kg, Sprat, 187 188 Mackerel and Mullet at the second with mean values around 1200 ng/kg, Sardine and Turbot 189 at the lowest level with mean values around 400 ng/kg. The change in certain congener 190 concentrations in the samples approximately agreed with the trend of the PBDEs sum, but in 191 some cases the congeners could show quite different patterns. For example, BDE-99 in Sea 192 Bass were as low as Sardine and Turbot, which suggests that the difference of the congener level in fish is also closely correlated with the congener types. 193

194



196 Fig.3 Dependence of mean value for the sum of PBDEs on fish species, and the concentration distribution



	Sardines				Mackerel (n=27)				Herring (n=14)				Grey Mullet (n=25)			
PBDE Concentrations,	(n=15)															
ng/kg whole weight	MIN	MEDIAN	MEAN	MAX	MIN	MEDIAN	MEAN	MAX	MIN	MEDIAN	MEAN	MAX	MIN	MEDIAN	MEAN	MAX
Sum measured PBDEs	159	415	535	2206	169	894	1206	3636	808	1634	2140	8906	147	577	1198	5422
*Sum PBDEs (EU list)	133	388	497	2118	141	806	1096	3381	757	1553	2039	8635	77	555	1094	5360
						<u> </u>								0.1		
	Sprat				Sea Bass				Turbot				Others			
	(n=12)				(n=25)				(n=13)				(n=4)			
	MIN	MEDIAN	MEAN	MAX	MIN	MEDIAN	MEAN	MAX	MIN	MEDIAN	MEAN	MAX	MIN	MEDIAN	MEAN	MAX
Sum measured PBDEs	383	975	1301	4620	297	1806	2033	5747	87	300	367	859	100	141	169	293

198Table 1 Summary of the detected PBDEs in the samples

199

*- Sum BDE-28, 47, 49, 99, 100, 138, 153, 154, 183 and 209 (EU recommendation 2014/118/EU)

201 Therefore, it could be expected that the geographical distribution of the congeners would 202 be affected by both the fish species and the congener types. To avoid the mixing of the two 203 variables, it is necessary to analyse the data in single species successively. However, it would 204 be unpersuasive to draw a distribution pattern if the sample numbers were too small, so, only 205 the species with more than 15 samples were analysed in Section 3.2, including Sea Bass, 206 Herring, Mackerel, Mullet and Sardine. Clearly, if more data become available, confidence in 207 the statistical calculations will increase. Therefore, as more data are accumulated they may be 208 used to augment or update the current data provided the sampling conditions remain the same 209 e.g. catch location.

210 **3.2** Analysis of the PBDEs in single fish species

211 3.2.1 Sea Bass

212 Within the sample set, 25 Sea Bass samples were collected with the sum of the 23 detected 213 congeners in individual samples varying between 297 to 5747 ng/kg with an average value of 214 2033 ng/kg as shown in Table 1. The concentration distribution of all the 23 congeners in Sea Bass is shown in Fig.4. The PBDE pattern was dominated by BDE-47, followed by BDE-100 > 215 216 BDE-49 > BDE-154 > BDE-28, and these top 5 congeners represented 94.71% of the summed 217 PBDEs. Typically, within these datasets, levels below the limit of detection (LOD) occur and 218 so in utilising tools of this form, decisions regarding the representation of these values should 219 be agreed.





Fig.4 Concentration distribution of the detected congeners in Sea Bass.

222

223 For spatial or geographical analysis, BDE-47 was selected as a typical example. BDE-47 in Sea Bass, was measured at between 127 to 3722 ng/kg and exhibited significant geographical 224 225 variation. As shown in Fig.5 (a), 4 of the top 6 concentration values occur where the waters of 226 the English Channel and the North Sea meet, and the other two were located off the coasts of 227 Wales and La Rochelle in France. Ten other congeners, including BDE-17, BDE-28, BDE-49, 228 BDE-66, BDE-77, BDE-99, BDE-100, BDE-119, BDE-153 and BDE-154, exhibited a similar 229 geographical distribution trend with BDE-47. Such distribution indicates that Sea Bass from 230 the south-east marine region of the UK show higher PBDEs contamination when compared 231 with other regions.



Fig.5 (a) Geographical distribution of BDE-47 in Sea Bass and (b) the Hierarchical Clustering results.

232

To confirm the distribution of the contaminants, HC analysis (Liang et al., 2011) was performed to divide all 25 samples into 3 clusters with their concentrations as input variables. The clustering results shown in Fig.5 (b) gave good agreement with the geographical distribution of the aforementioned congeners as BDE-47. The samples belonging to Cluster 3 located at the junction area of English Channel and the North Sea, Wales and La Rochelle, which contained higher concentration of contaminants, while Cluster 1 containing least contaminant levels were mainly located at the middle part of the south coast of the UK.

Furthermore, four of the congeners, including BDE-71, BDE-99, BB-49 and BB-52
showed no obvious regularity, and the other 8 congeners were at very low levels (most < LOD).

244 3.2.2 Herring

Of the 14 herring samples collected, the sum of the 23 detected congeners in single samples varied from 808 to 8906 ng/kg with an average value of 2140 ng/kg. To observe the different content levels of the congeners, the concentration distributions of all the 23 congeners were compared (shown in Fig.6) by arranging them in descending order of their mean value. Similar with Sea Bass, BDE-47 was also at the highest level in herring, followed by BDE-49 > BDE-100 > BDE-209 > BDE-99, these top 5 congeners represented 88.27% of the summed PBDEs. Such difference of the congeners could also be intuitively observed by presenting their concentrations with rings in map at the same scale level (see Fig.7 (a)).



253

254

Fig.6 Concentration distribution of the detected congeners in 14 Herring samples.

255

Fig.7 (a) also reveals that herring samples from the central part of UK, i.e., Mersey, showed higher levels of most of the PBDE congeners. HC results as shown in Fig.7 (b) agreed well with this conclusion, only samples from Mersey were recognized as Cluster 3, which contained highest level of contaminants. All samples from the south-east part of UK belonged to Cluster 2 and samples from the north part showed no clear pattern. In Fig. 7(a) it is shown using the web page how mutiple contaminants within single species can be effectively displayed across the spatial region of interest.



Fig.7 (a) Geographical distribution of detected congeners in Herring and (b) the Hierarchical Clustering
 results

266 3.2.3 Mackerel

263

267 27 Mackerel samples were collected in the location extending to Norway in the north and 268 Algarve in the south. The sum of all detected congeners in individual samples ranged from 169 269 to 3636 ng/kg with an average value of 1206 ng/kg. As expected, BDE-47 was still the congener 270 responsible for the highest level of contamination as it showed the largest average 271 concentration. BDE-47 alone represented 38.57% of the summed contaminants, followed by 272 BDE-49, BDE-100, BDE-99 and BDE-154, and the top 5 congeners represented 80.66% of the 273 summed PBDEs.

Geographically, the north-east region of UK (mainly around Peterhead) and the English
Channel area showed higher concentration of contaminants including BDE-17, BDE-28, BDE47, BDE-49, BDE-66, BDE-99, BDE-100, BDE-119, BDE-153, BDE-154, BDE-209 and BB52. The remaining congeners showed very low concentrations (< LOD) in Mackerel. HC results
identified samples from Norway and Spain were recognized as Cluster 1, which contained the
lowest level of contaminants.

281 3.2.4 Mullet

282 The 25 mullet samples were collected mainly from the south coast of UK and surrounding coastal area of France. The sum of all detected congeners in individual samples ranged from 283 284 147 to 5422 ng/kg with an average value of 1198 ng/kg. Similar with the previously described species, the top 5 congeners were BDE-47, BDE-100, BDE-209, BDE-154 and BDE-29 (See 285 286 Fig.10), which represented 87.89% of the summed PBDEs. Eight of the congeners, including 287 BDE-71, BDE-183, BDE-77, BDE-85, BDE-126, BDE-138, BB-80 and BB-209, showed 288 concentration lower than LOD in most of the samples, which would not be considered for 289 further analysis.

Geographically, it was found that the BDE compounds and the PBB compounds exhibited quite different distribution patterns. Samples from the coastal area of UK tended to show much higher levels of BDE compounds when compared with France, whereas the PBB compounds in samples from UK all showed concentrations lower than LOD, and all the samples with a concentration higher than the LOD came from the coast of France. This confirmed a trend observed in other studies (Fernandes et al., 2008), which may reflect a higher utilisation of PBBs in France relative to the UK.

297 3.2.5 Sardine

The Sardine samples with a total number of 15 were collected mainly from the coastal areas around the English Channel. The sum of all detected congeners in individual samples ranged from 159 to 2206 ng/kg with an average value of 535 ng/kg. The top 5 congeners were BDE-47, BDE-209, BDE-49, BDE-100 and BDE-99, which represented 87.14% of the summed PBDEs. Thirteen of the congeners, including BDE-17, BB-52, BDE-71, BDE-77, BB-49, BDE-85, BDE-126, BDE-138, BDE-183, BB-80, BB-101, BB-153 and BB-209, showed 304 concentrations lower than LOD in most of the samples, which would not be considered for305 further analysis.

In terms of the geographical distribution, samples from Poole in the UK showed higher concentration of PBDEs compared with other locations. However, the other regions showed no clear distribution patterns. It should be emphasised that this approach, although offering quantitative information does not address wider issues e.g. sampling levels and behavioural feeding patterns or ranges of species. Thus, if particular trends are observed a more detailed analyses should be conducted to establish causative factors and/or minimise sources of uncertainties.

313

328

314 3.2.6 Summary

315 As seen from the above analysis, BDE-47 was the congener of highest concentration in all species, and was always followed by congeners as, BDE-49, BDE-99, BDE-100, BDE-316 317 154 and BDE-209 (orders could be different), which agrees well with previous reports 318 (Brown et al., 2006; Colombo et al., 2011). However, the BB compounds and several of the 319 BDE compounds such as BDE-71, BDE-77, BDE-85, BDE-126, BDE-183 and BDE-138 320 always showed much lower concentrations than other congeners. For the PBDEs, this finding 321 is consistent with the lower occurrence of these congeners in commercial mixtures (perhaps 322 apart for BDE-183 which forms part of the deca-BDE product) Geographically, the 323 distribution pattern of the PBDEs in different species varied significantly, even though the 324 different congeners in the same species always followed a similar pattern. 325 It can generally be concluded that BDE-47, BDE-49 and BDE-100 were of much higher 326 concentration than other congeners in most of the samples, and the coast around the English Channel shows higher PBDEs concentration levels when compared with other areas. Both these 327

conclusions had been confirmed by the analysis in Section 3.1 and 3.2, and agreed with a

previous report (Johansson et al., 2006), which reflects the influence of urban and industrialsources in these coastal areas.

331 **3.4 Conclusions**

332 This study has presented the use of an interactive webpage facility to rapidly evaluate 333 complex environmental contaminant datasets which may comprise of multiple factors, e.g. 334 congeners, species and sample locations. Qualitative assessment of correlations between the 335 location and species with congeners is demonstrated which will enable more targeted detailed 336 analyses when considering these complex multifactor datasets. As an example we have 337 characterised PBDE and PBB occurrence in five commonly consumed fish species, (Sea 338 Bass, Herring, Mackerel, Mullet and Sardine) taken from marine waters around the UK and 339 from other North Atlantic fishing areas from which retail fish in the UK is commonly 340 sourced. PBDEs were observed in all samples with the majority of measured congeners being 341 detected. The concentrations ranged from 0.087 µg/kg to 8.907 µg/kg whole weight (ww) for 342 the sum of all measured PBDE congeners. PBBs occurred less frequently showing a corresponding range of $< 0.02 \mu g/kg$ to 0.97 $\mu g/kg$ ww for the sum of seven PBB congeners. 343 344 Generally, BDE-47 was the congener of highest concentration in all species, and was always 345 followed by congeners as, BDE-49, BDE-99, BDE-100, BDE-154 and BDE-209 (orders 346 could be different), the BB compounds and several of the BDE compounds such as BDE-71, 347 BDE-77, BDE-85, BDE-126, BDE-183 and BDE-138 always showed much lower 348 concentrations than other congeners. The web-based resource provides a flexible and efficient 349 tool for assessors and policy-makers to monitor and evaluate levels within caught fish species 350 and enables evidenced-based decisions to be justified. Spatial variation, species and congener 351 correlations with clustering can be achieved offering a useful investigative resource to 352 explore the relationships within large and complex datasets. Results show that the spatial 353 distribution of the PBDEs in different species varied significantly, but the different congeners

354 in the same species always followed a similar pattern. The coast area around the English 355 Channel shows higher PBDEs concentration levels when compared with other areas. With the 356 inclusion of additional data, confidence levels of suggested trends may be improved. 357 Furthermore, this system could be readily enhanced to include consumption data and enable predictions of exposure estimates for comparison with UK or European tolerable daily intake 358 359 (TDI) levels. The levels and frequency of PBDE occurrences make it prudent to continue the monitoring of these commonly consumed marine fish species from the point of view of 360 361 public health, as well as the status of the marine environment. This approach may be more 362 readily applied to land-based contaminant datasets and enable effective intervention 363 monitoring.

364

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369 **References:**

370 Bramwell, L., Mortimer, D., Rose, M., Fernandes, A., Harrad, S., Pless-Mulloli, T., 2017.

371 UK dietary exposure to PCDD/Fs, PCBs, PBDD/Fs, PBBs and PBDEs: comparison of results

372 from 24-h duplicate diets and total diet studies. Food Addit Contam Part A Chem Anal Control

- 373 Expo Risk Assess 34, 65-77.
- Brown, F.R., Winkler, J., Visita, P., Dhaliwal, J., Petreas, M., 2006. Levels of PBDEs,
 PCDDs, PCDFs, and coplanar PCBs in edible fish from California coastal waters.
- 376 Chemosphere 64, 276-286.
- 377 Bruggeman, W., Opperhuizen, A., Wijbenga, A., Hutzinger, O., 1984. Bioaccumulation of

- 378 super-lipophilic chemicals in fish. Toxicological & Environmental Chemistry 7, 173-189.
- 379 Chang, J.-W., Hung, C.-F., Hsu, Y.-C., Kao, Y.-T., Lee, C.-C., 2017. Polybrominated
- 380 diphenyl ethers (PBDES) and hexa-brominated biphenyls (Hexa-BBs) in fresh foods ingested
- in Taiwan. Environmental Pollution 220, 1180-1189.
- 382 Chen, T.-B., Zheng, Y.-M., Lei, M., Huang, Z.-C., Wu, H.-T., Chen, H., Fan, K.-K., Yu,
- K., Wu, X., Tian, Q.-Z., 2005. Assessment of heavy metal pollution in surface soils of urban
 parks in Beijing, China. Chemosphere 60, 542-551.
- Chiaramonte, A., Zota, A.R., 2016. Pregnant Women's Health Consequences following
 exposure to PBDEs.
- 387 Colombo, J.C., Cappelletti, N., Williamson, M., Migoya, M.C., Speranza, E., Sericano, J.,
- Muir, D.C., 2011. Risk ranking of multiple-POPs in detritivorous fish from the Rio de la Plata.
 Chemosphere 83, 882-889.
- El-Moselhy, K.M., Othman, A., El-Azem, H.A., El-Metwally, M., 2014. Bioaccumulation
 of heavy metals in some tissues of fish in the Red Sea, Egypt. Egyptian Journal of Basic and
 Applied Sciences 1, 97-105.
- Fernandes, A., Dicks, P., Mortimer, D., Gem, M., Smith, F., Driffield, M., White, S., Rose,
 M., 2008. Brominated and chlorinated dioxins, PCBs and brominated flame retardants in
 Scottish shellfish: methodology, occurrence and human dietary exposure. Molecular nutrition
 & food research 52, 238-249.
- Fernandes A, R.M., Smith F and Holland M, 2012. Organic Environmental contaminants
 in the 2012 Total Diet Study samples. Report to the Food Standards Agency.
- Fernandes A, S.F., Petch R, Panton S, Carr M, and Rose M, 2009. Investigation into the
 occurrence of brominated contaminants in selected foods. Report to the Food Standards
 Agency.
- 402 Fernandes, A., White, S., D'Silva, K., Rose, M., 2004. Simultaneous determination of

- 403 PCDDs, PCDFs, PCBs and PBDEs in food. Talanta 63, 1147-1155.
- 404 FSA, 2006. Brominated chemicals : UK dietary intakes. Food Standards Agency, London.
 405 GOOGLE MAPS, 2016. [online].
- 406 Hashizume, N., Tanabe, A., Inoue, Y., Sawada, T., Murakami, H., Suzuki, Y., Sumi, S.,
- 407 Tsubokura, Y., Yoshida, T., Ajimi, S., 2014. Prediction of the bioconcentration factor in
- 408 common carp (Cyprinus carpio L.) using data from the dietary exposure bioaccumulation fish
- 409 test. Environmental toxicology and chemistry 33, 1406-1414.
- Johansson, I., Héas-Moisan, K., Guiot, N., Munschy, C., Tronczyński, J., 2006.
 Polybrominated diphenyl ethers (PBDEs) in mussels from selected French coastal sites: 1981–
 2003. Chemosphere 64, 296-305.
- Liang, Y., Van Nostrand, J.D., Deng, Y., He, Z., Wu, L., Zhang, X., Li, G., Zhou, J., 2011.
 Functional gene diversity of soil microbial communities from five oil-contaminated fields in
 China. The ISME journal 5, 403-413.
- Luross, J.M., Alaee, M., Sergeant, D.B., Cannon, C.M., Michael Whittle, D., Solomon,
 K.R., Muir, D.C.G., 2002. Spatial distribution of polybrominated diphenyl ethers and
 polybrominated biphenyls in lake trout from the Laurentian Great Lakes. Chemosphere 46,
 665-672.
- Magalhaes, V.d., Marinho, M.M., Domingos, P., Oliveira, A., Costa, S.M., Azevedo,
 L.O.d., Azevedo, S.M., 2003. Microcystins (cyanobacteria hepatotoxins) bioaccumulation in
 fish and crustaceans from Sepetiba Bay (Brasil, RJ). Toxicon 42, 289-295.
- 423 McDonald, T.A., 2002. A perspective on the potential health risks of PBDEs.
 424 Chemosphere 46, 745-755.
- 425 Schrenk, D., Cartus, A., 2017. Chemical contaminants and residues in food. Woodhead426 Publishing.
- 427 Smoliński, A., Walczak, B., Einax, J.W., 2002. Hierarchical clustering extended with

- 428 visual complements of environmental data set. Chemometrics and Intelligent Laboratory
- 429 Systems 64, 45-54.